

Integration of the Vegetation Phenology Module Improves Ecohydrological Simulation by the SWAT-Carbon Model

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10 **Abstract.** Vegetation phenology and hydrological cycles are closely interacted from leaf and species levels to watershed and global scales. As one of the most sensitive biological indicators of climate change, plant phenology is essential to be simulated accurately in hydrological models. Despite the Soil and Water Assessment Tool (SWAT) has been widely used for estimating hydrological cycles, its lack of integration with the phenology module has led to substantial uncertainties. In this study, we developed a process-based vegetation phenology module and coupled it with the SWAT-Carbon model to
15 investigate the effects of vegetation dynamics on runoff in the upper reaches of Jinsha River watershed in China. The modified SWAT-Carbon model showed reasonable performance in phenology simulation, with root mean square error (RMSE) of 9.89 days for the start-of-season (SOS) and 7.51 days for the end-of-season (EOS). Simulations of both vegetation dynamics and runoff were also substantially improved compared to the original model. Specifically, the simulation of leaf area index significantly improved with the coefficient of determination (R^2) increasing by 0.62, the Nash–
20 Sutcliffe efficiency (NSE) increasing by 2.45, and the absolute percent bias (PBIAS) decreasing by 69.0 % on average. Additionally, daily runoff simulation also showed notable improvement, particularly noticeable in June and October, with R^2 rising by 0.22 and NSE rising by 0.43 on average. Our findings highlight the importance of integrating vegetation phenology into hydrological models to enhance modeling performance.

1 Introduction

25 Vegetation plays a crucial role as a link between the land and the atmosphere through its influence on carbon, water, and energy cycles (Bonan, 2008; Zhou et al., 2023). Plant phenology is one of the most sensitive bioindicators of climate change (IPCC, 2021; Fu et al., 2015; Vitasse et al., 2022) and closely interacts with the hydrological cycles (Chen et al., 2022a; Hwang et al., 2014; Kim et al., 2018). Therefore, vegetation phenological processes must be accurately simulated in hydrological models. The Soil and Water Assessment Tool (SWAT) has been widely used to model hydrological processes
30 in watersheds (Bhatta et al., 2019; Tian et al., 2020). However, the performance of its vegetation module and its simulation

of vegetation phenology remain poor, resulting in large uncertainties in the simulation of hydrological processes (Chen et al., 2023). To improve predictions of ecohydrological processes under future climate change conditions, it is essential to modify the vegetation module by considering vegetation phenology.

In recent decades, climate change has substantially shifted vegetation phenology, with an advance in the start-of-season (SOS) (Fu et al., 2014) and a delay in the end-of-season (EOS) (Piao et al., 2019). This has extended the growing season by 2–10 days per decade (Zhao et al., 2015; Garonna et al., 2016; Shen et al., 2022), closely interacting with hydrological processes from leaf and species levels to watershed and global scales (Chen et al., 2022b). At the leaf scale, earlier spring leaf-out would enhance the transpiration and increase the water requirement of plants, resulting in greater water flux from the soil to the atmosphere. At the watershed scale, the extension of the vegetation growing season would reduce runoff through increased evapotranspiration (ET) (Gaertner et al., 2019; Geng et al., 2020; Kim et al., 2018). Shifts in vegetation phenology affect both the magnitude and the timing of hydrological processes (Creed et al., 2015; Hwang et al., 2023). Therefore, developing a hydrological model capable of characterizing vegetation phenology is essential for a comprehensive understanding of hydrological processes.

The SWAT model has been extensively used to assess the impact of climate change on water processes across a variety of watershed scales. The updated SWAT-Carbon model was developed to enhance the simulation of terrestrial carbon and water cycles (Mukundan et al., 2023; Zhang et al., 2013). As SWAT, SWAT-Carbon employs a simplified version of the Environmental Policy Impact Climate (EPIC) vegetation growth module (Arnold et al., 1998), which determines SOS and EOS based on a daylength threshold calculated as a function of latitude. However, phenological processes are complex, determined not only by daylength but also by temperature and precipitation (Piao et al., 2015; Wu et al., 2023; Peñuelas et al., 2004). For example, temperature is regarded as the dominant driver of spring phenology, with spring warming largely advancing leaf-out dates by meeting the forcing requirement earlier (Ge et al., 2015; Fu et al., 2018). However, winter warming-induced reductions in chilling accumulation would increase the forcing requirement and delay the leaf-out dates (Fu et al., 2015; Wu et al., 2023). Therefore, the EPIC module fails to capture the real phenological processes (Ma et al., 2019), making it urgent to couple a specific phenology module with the SWAT-Carbon model to simulate hydrological processes accurately.

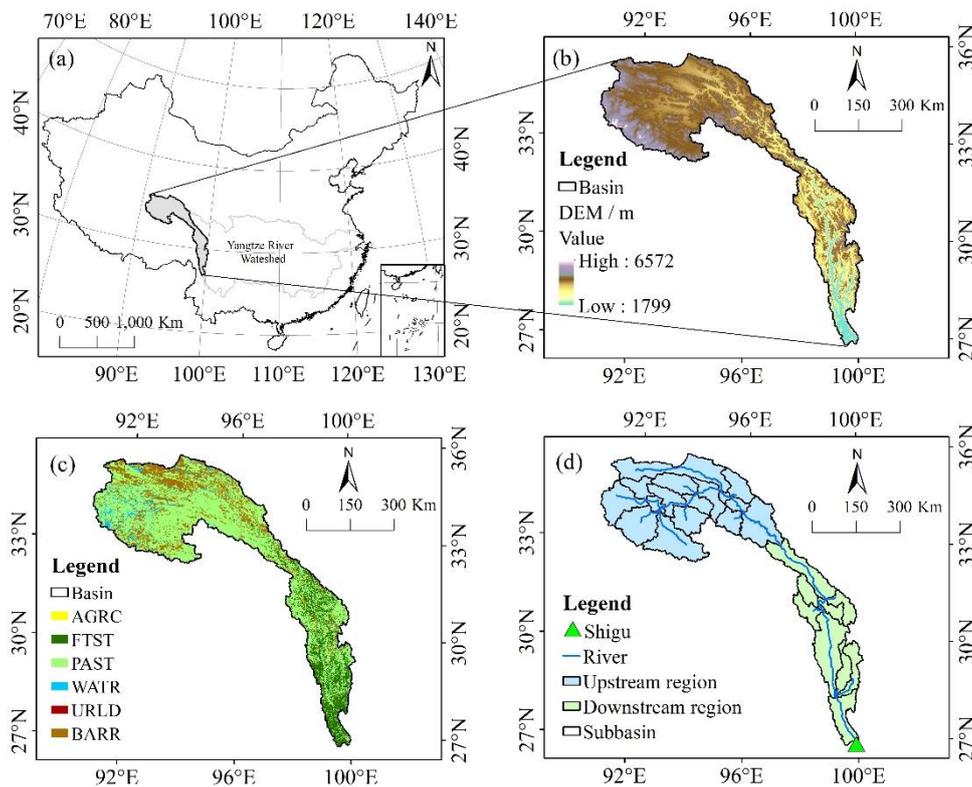
The upper reaches of the Jinsha River watershed, originating from the Tibetan Plateau and forming the upper reach of Yangtze River, is recognized as one of the ecologically fragile regions in China. In recent years, the upper reaches of the Jinsha River watershed experienced significant climate change, greatly influencing vegetation dynamics and regional water cycles (Wu et al., 2020; Jiang et al., 2024; Li et al., 2022). Enhancing our understanding and simulation of ecohydrological processes in the upper reaches of the Jinsha River watershed is critical for securing water resources and maintaining ecological stability. In this study, we first developed a process-based vegetation phenology module for the upper reaches of the Jinsha River watershed, which consists of spring and autumn phenology models that consider both temperature and

photoperiod. Subsequently, we coupled this module with the SWAT-Carbon model. The objectives of our study are as follows: 1) to develop a process-based vegetation phenology module and integrate it into the SWAT-Carbon model; 2) to elucidate the effects of vegetation phenology and climate change on watershed ecohydrological processes; and 3) to forecast future shifts in runoff under different emission scenarios using the modified SWAT-Carbon model.

2 Materials and Methods

2.1 Study Area

Our study focused on the upper reaches of the Jinsha River watershed in China, located in the upper reach of the Yangtze River (Figure 1a). The elevation difference of the upper Jinsha River watershed is 4,773 m (Figure 1b), and the drainage area covers 216,440 km². The multiyear mean temperature was approximately -2.56 ± 4.11 °C, and the multiyear mean annual precipitation was approximately 467 ± 135 mm from 1982 to 2018. Influenced by monsoons, precipitation is mainly observed during June–September, accounting for over 75 % of the annual total, with peak precipitation mostly occurring in July. The land use types in the study area changed little during 2007–2014, with grassland (61.4 %) and forest (14.6 %) representing the dominant vegetation types (Figure 1c). Grassland areas are primarily distributed in the upstream region of the basin, whereas forests are prevalent in the downstream area.



80 **Figure 1: Study area.** (a) Location of the upper Jinsha River watershed, (b) digital elevation model (DEM), (c) distribution of land use types (AGRC-farmland, FRST-forest, PAST-grassland, WATR-water, URLD- residential area, BARR-unused land), and (d) division of subbasins and the location of the Shigu hydrological gauging station.

2.2 Dataset

The sources of the datasets used for phenology models and the SWAT-Carbon model are listed in Table 1.

2.2.1 Digital Elevation Model (DEM), Land Use, and Soil Data

85 The watershed and river network of the study area were delineated using a digital elevation model (DEM) with 90-m resolution (Figure 1b) in ArcSWAT2012 (<https://swat.tamu.edu/>). The watershed was first subdivided into 28 subbasins (Figure 1d) based on topographic characteristics and then discretized into 1,922 hydrologic response units (HRUs) based on land use, soil data, and slope characteristics.

We extracted land use data with 1-km spatial resolution from the Resources and Environmental Sciences and Data Center, 90 Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences (<https://www.resdc.cn/>). This data was reclassified to match the land use classes for HRU delineation in the SWAT-Carbon model (Figure 1c). There were six classes of land use identified in the watershed: farmland (ARGC), forest (FRST), grassland (PAST), water (WATR), residential area (URLD), and unused land (BARR).

The soil map of the upper reaches of the Jinsha River watershed was obtained from the Food and Agriculture Organization of the United Nations (<http://www.fao.org>). We used the Soil Water Atmosphere Plant model (SWAP: <http://hydrolab.arsusda.gov/soilwater>) to develop the database of soil properties for the watershed.

2.2.2 Climate Data and Runoff

Gridded climate data were processed as observed values to force both the original and the modified SWAT-Carbon models, considering of the limited distribution of meteorological stations. We used daily meteorological data from 1982 to 2018 with 100 0.1° spatial resolution, extracted from the China Meteorological Forcing Dataset (CMFD), which included precipitation, temperature, relative humidity, wind speed, and solar radiation (He et al., 2020). This dataset was developed by the Institute of Tibetan Plateau Research under the Chinese Academy of Sciences, and it represents the optimal combination of in-situ observations, satellite retrievals, and meteorological reanalysis.

CMIP6 outputs were used to predict vegetation phenology and future runoff. We selected four CMIP6 models, i.e., 105 CanESM5, FGOALS-g3, MPI-ESM1-2-HR, and MRI-ESM2-0, which were commonly used in studies of the Yangtze River (Zhang et al., 2023; He et al., 2024). The daily time series of mean temperature, maximum temperature, minimum temperature, and precipitation under three emission scenarios (SSP1-2.6, SSP2-4.5, and SSP5-8.5) were acquired from the CMIP6 website (<https://esgf-node.llnl.gov/search/cmip6/>). CMIP6 data are biased in local areas (Stevens and Bony, 2013), making it necessary to correct these errors. First, we interpolated the CMIP6 data to 0.1° spatial resolution using linear 110 interpolation. Then, the CMFD was employed as the reference to correct the bias in CMIP6 models using the empirical mapping technique (Gudmundsson et al., 2012; Wilcke et al., 2013). Daily empirical cumulative distribution functions for the calibration period were produced using a moving window of ± 15 days (Thrasher et al., 2012).

Daily runoff data (2007–2014) of the Shigu hydrological gauging station, located at the upper reaches of Jinsha River watershed outlet, were obtained from the Chinese Hydrological Data Yearbook.

115 2.2.3 Vegetation Data

We obtained the Global Inventory Modeling and Mapping Studies third generation normalized difference vegetation index (GIMMS3g NDVI) dataset from the Ecological Forecasting Lab (<https://data.tpdc.ac.cn/en/data/9775f2b4-7370-4e5e-a537-3482c9a83d88/>) for 1982–2020 (spatial resolution: 1/12°, temporal resolution: every 15 d).

Leaf area index (LAI) data from 1982 to 2018 was acquired from the Global Land Surface Satellite (GLASS) product with 120 0.05° spatial resolution and an eight-day temporal interval (<http://glass-product.bnu.edu.cn>). LAI time series for forest and grassland areas were obtained through area-weighted averaging of FRST and PAST grid values within the studied watershed (Strauch and Volk, 2013; Zhang et al., 2020).

Table 1: Details of the data used in this study.

Data	Source	Characteristics
Digital elevation model (DEM)	Geospatial Data Cloud (https://www.gscloud.cn/)	SRTMDem 90M
Land use types	Center for Resources and Environmental Sciences and Data (https://www.resdc.cn/)	1-km spatial resolution
Soil map	Food and Agriculture Organization of the United Nations (http://www.fao.org)	Harmonized World Soil Database v 1.2
Observed meteorological data	China Meteorological Forcing Dataset (https://data.tpdc.ac.cn/home)	daily, 0.1° spatial resolution
Observed runoff	Chinese Hydrological Data Yearbook	Shigu hydrological gauging station, daily
Normalized difference vegetation index (NDVI)	Global Inventory Modeling and Mapping Studies third generation (GIMMS _{3g}) (https://data.tpdc.ac.cn/en/data/9775f2b4-7370-4e5e-a537-3482c9a83d88/)	15 d, 1/12° spatial resolution
Leaf area index (LAI)	Global Land Surface Satellite (GLASS) product (http://glass-product.bnu.edu.cn).	8 d, 0.05° spatial resolution

2.3 SWAT-Carbon Model Modification

125 The SWAT model is a comprehensive, process-oriented, and physically based semi-distributed hydrological model (Neitsch et al., 2011). The SWAT-Carbon model (<https://sites.google.com/view/swat-carbon/home>), as an updated version, is developed to enhance the simulation of terrestrial carbon cycles. However, the SWAT-Carbon model still performs poorly in estimating vegetation dynamics (e.g., phenology and LAI). The current SWAT-Carbon model uses daylength thresholds, that are determined only by latitude, to simulate the onset and the end of vegetation dormancy. This approach fails to accurately capture vegetation dynamics as it largely ignores the effects of other important environmental factors (i.e., temperature) (Chen et al., 2023). Incorporating accurate phenology information could enhance the simulation of LAI, thereby improving the accuracy of runoff simulated by the hydrological model. The UniChill model and the DM model, which account for the response of phenology to various environmental variations, have been widely used to simulate spring and autumn phenology (Roberts et al., 2015; Yang et al., 2012). Therefore, we modified the SWAT-Carbon model by integrating the process-based vegetation phenology module. This phenology module consists of two phenology models: the UniChill model for simulating SOS and the DM model for simulating EOS. These phenology models provide dynamic simulations of plant phenology to replace the fixed dates of the SOS and EOS in the SWAT-Carbon model without the need for adjusting management operations. Subsequently, we set the heat accumulation calculated between dynamic phenology dates, instead of the default constant, as the annual heat requirement for vegetation growth. This dynamic heat requirement was then used to calculate daily variations in LAI.

The UniChill model, proposed by Chuine (2000), hypothesizes that spring phenology is regulated both by chilling temperatures and by forcing temperatures during the dormancy period. Because of the specific range of chilling and forcing temperatures, we added temperature conditions when calculating the state of the chilling and forcing phases (Fu et al., 2012). The chilling phase starts at the onset of dormancy (t_0 , fixed to 1 September), and is active on days with mean temperature
 145 between -5 and 10 °C. The rate of chilling can be expressed as follows:

$$R_C = \begin{cases} \frac{1}{1 + e^{C_a(x-C_c)^2 + C_b(x-C_c)}} & \text{if } -5 \leq x \leq 10 \\ 0 & \text{if } x > 10 \text{ or } x < -5 \end{cases} \quad (1)$$

where x is the daily mean air temperature, and C_a , C_b , and C_c are chilling rate parameters.

The chilling phase ends with the onset of quiescence (t_1), which occurs when the state of chilling (S_C ; Eq. 2) reaches the critical state of chilling (Eq. 3):

$$150 \quad S_C = \sum_{t_0}^t R_C(x_t) \quad (2)$$

$$t = t_1 \quad \text{if } S_C > C^* \quad (3)$$

The forcing phase begins at t_1 , and is active on days with mean temperature is above 0 °C.

$$R_F = \begin{cases} \frac{1}{1 + e^{F_b(x-F_c)}} & \text{if } x > 0 \\ 0 & \text{if } x \leq 0 \end{cases} \quad (4)$$

where F_b and F_c are forcing rate parameters.

155 The forcing phase ends with the occurrence of SOS (t_2), which occurs when the state of forcing (S_F ; Eq. 5) reaches the critical state of forcing (Eq. 6):

$$S_F = \sum_{t_1}^t R_F(x_t) \quad (5)$$

$$t = t_2 \quad \text{if } S_F > F^* \quad (6)$$

The DM model (Delpierre et al., 2009) is based on the assumption that leaf senescence is driven both by the photoperiod and
 160 by low temperatures. This model represents the progress of senescence processes using a coloring state (S_{sen}) for each day (d), which is the accumulation of the daily rate of senescence (R_{sen}). Senescence starts on the date (D_{start}) when the photoperiod is

shorter than a specific threshold (P_{start}), and the EOS is recognized as when S_{sen} exceeds the threshold value Y_{crit} . The functions expressed in the process are as follows:

$$if\ P(d) < P_{start} \begin{cases} if\ T(d) < T_b & R_{sen}(d) = [T_b - T(d)]^x \times f[P(d)]^y \\ if\ T(d) \geq T_b & R_{sen}(d) = 0 \end{cases} \quad (7)$$

$$165\ if \begin{cases} P(d) \geq P_{start} & S_{sen}(d) = 0 \\ P(d) < P_{start} & S_{sen}(d) = S_{sen}(d-1) + R_{sen}(d) \end{cases} \quad (8)$$

$$f[P(d)] = \frac{P(d)}{P_{strat}} \quad or \quad f[P(d)] = 1 - \frac{P(d)}{P_{strat}} \quad (9)$$

where $P(d)$ is the photoperiod on a given day d , $T(d)$ is the daily mean temperature ($^{\circ}C$), and T_b is the critical temperature below which senescence processes are effective ($^{\circ}C$).

We calibrated and validated the two phenology models using all pixel-year SOS and EOS data extracted from the GIMMS3g
 170 NDVI dataset during 1982–2018. Five phenological extraction methods (i.e., the HANTS-Maximum, Spline-Midpoint, Gaussian-Midpoint, Timesat-SG, and Polyfit-Maximum methods) were applied to obtain the mean phenological index value, reducing the uncertainty caused by the use of only a single method (Fu et al., 2021). To match the spatial resolution of gridded meteorological data, the SOS and EOS data were resampled to 0.1° spatial resolution, and outliers were removed using the interquartile range method (Cui et al., 2017). We optimized parameters of the UniChill model (i.e., C_a , C_b , C_c , F_b ,
 175 C_s , C^* , and F^*) and the DM model (i.e., P_{start} , T_b , x , y , and Y_{crit}) for each valid pixel in upper reaches of the Jinsha River watershed, minimizing the root mean square error (RMSE) through particle swarm optimization. Phenology data from odd-numbered years were utilized for parameter optimization, and even-numbered years were used for validation.

The warming-up, calibration, and validation periods for the original and modified SWAT-Carbon models were set at 2005–2006, 2007–2011, and 2012–2014, respectively. The calibration of the modified SWAT-Carbon model involved the
 180 following two steps, whereas that of the original SWAT-Carbon model employed only the second step. 1) We optimized the eight parameters (Table S1) that control the shape, magnitude, and temporal dynamics of the LAI for forest and grassland through particle swarm optimization, to align the simulated LAI curves with the observed dynamic LAI. Ten particles with different initial velocities were generated through the Latin hypercube sampling method. The Nash–Sutcliffe efficiency (NSE) was used as the objective function to continuously update the parameters. Optimal parameter sets were determined
 185 when the objective function remained unchanged after ten iterations. 2) We adjusted those parameters that control the hydrological processes using SWAT-CUP 2019 (<https://swat.tamu.edu/software/swat-cup/>) (Abbaspour, 2008). The Sequential Uncertainty Fitting (Sufi-2) algorithm in SWAT-CUP was used for sensitivity analysis, uncertainty analysis, and parameter calibration (Arnold et al., 2012). Overall, 18 parameters (Table S2) were used for model calibration. These

parameters were derived by combining the 14 most sensitive parameters selected from the pool of 22 parameters for each original and modified model.

2.4 Model Evaluation

The widely used coefficient of determination (R^2), NSE, and percent bias (PBIAS) were selected as metrics to evaluate the performance of the model in simulating LAI and runoff in our study area (Moriassi et al., 2007):

$$R^2 = \frac{\left(\sum_{i=1}^n (O_i - \bar{O})(P_i - \bar{P}) \right)^2}{\sum_{i=1}^n (O_i - \bar{O})^2 \sum_{i=1}^n (P_i - \bar{P})^2} \quad (10)$$

$$NSE = 1 - \frac{\sum_{i=1}^n (O_i - P_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (11)$$

$$PBIAS = \frac{\bar{P} - \bar{O}}{\bar{O}} \times 100\% \quad (12)$$

where n is the number of observations; O_i and P_i are the observed and simulated values, respectively, at time i ; and \bar{O} and \bar{P} represent the average of the observed and simulated values, respectively.

3 Results

3.1 Spatiotemporal Variations in Climate and Vegetation Phenology

There is notable spatiotemporal heterogeneity in the multiyear mean annual temperature and accumulated precipitation across the upper reaches of the Jinsha River watershed (Figure 2a and 2b). The mean annual temperature exhibits a gradual increase from the upstream to the downstream areas. The annual accumulated precipitation is higher in the downstream region than in the upstream. In addition, the mean annual temperature and accumulated precipitation showed clear temporal trends during 1982-2018. The mean annual temperature significantly increased by 0.06 °C per year on average. The annual accumulated precipitation across the entire watershed increased by 2.64 mm per year. However, downstream regions exhibited a declining trend (Figure 2b). Over the past four decades, the upper reaches of the Jinsha River watershed showed a “warming–wetting” trend in the upstream area and a “warming–drying” trend in some downstream areas.

During 1982–2018, the multiyear mean SOS and EOS in the upper reaches of the Jinsha River watershed were 164 ± 12 and 298 ± 11 , (Day of year, DoY), respectively (Figure 2c and 2d). The mean SOS initiated in early May upstream and early July

downstream. The mean EOS occurred in early October upstream, one month earlier than downstream. The SOS advanced in 60.5% of the study area over the past four decades, primarily in the middle and upper reaches of the watershed (-0.05 ± 0.13 days per year), while being delayed by 0.09 ± 0.30 days per year in the downstream area. During the study period, the EOS was delayed by 0.08 ± 0.27 days per year in the downstream area but advanced by -0.03 ± 0.14 days per year in the upstream area. This spatiotemporal heterogeneity of vegetation phenology further highlights the importance of incorporating vegetation dynamics into hydrological models.

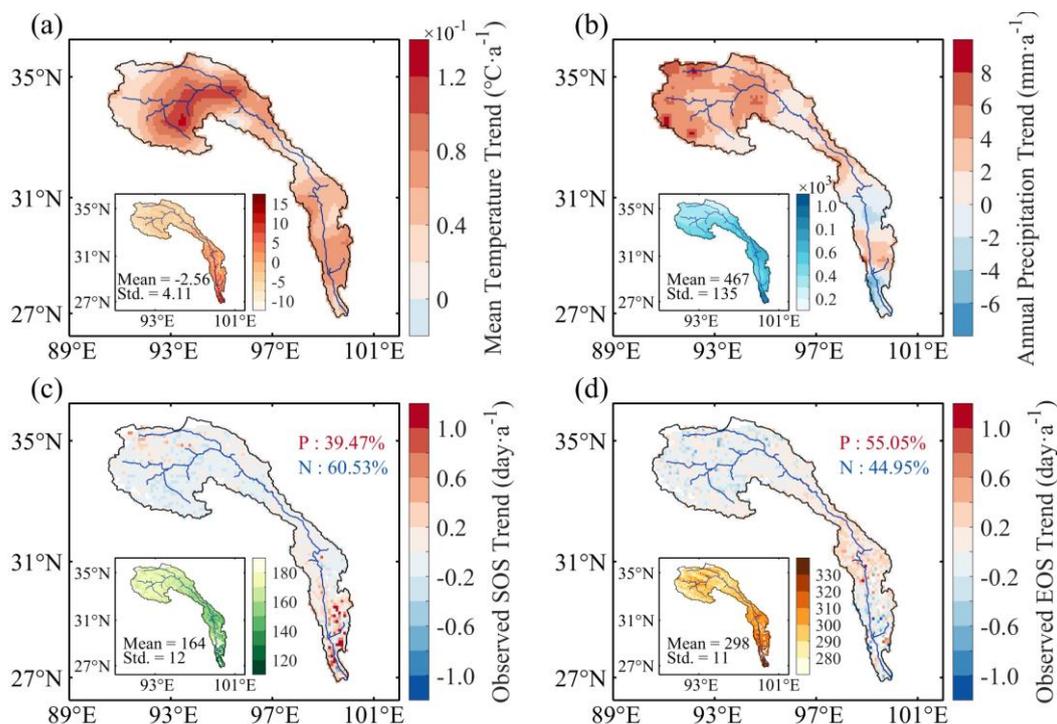
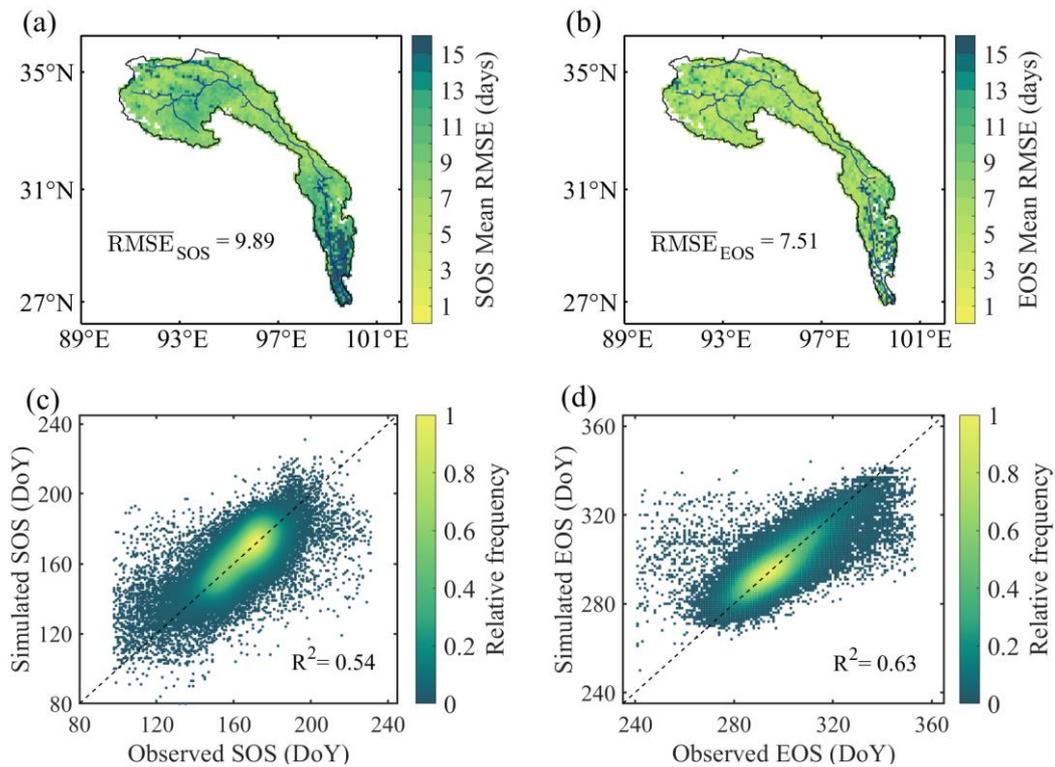


Figure 2: Spatial and temporal variations of climatic variables and vegetation phenology during 1982–2018. Temporal variations of mean annual temperature (a), accumulated annual precipitation (b), start-of-season (SOS, c), and end-of-season (EOS, d). The inner plot of each subfigure depicts the spatial pattern of the multi-year means of each variable. P, percentage of pixels showing a positive trend, indicates increased temperature and precipitation, or delayed SOS and EOS; N, percentage of pixels showing a negative trend, indicates decreased temperature and precipitation, or advanced SOS and EOS.

3.2 Performance of Phenology Module and Modified SWAT-Carbon Model

The mean RMSE for the spring and autumn phenology models are 9.89 and 7.51 days, with calibration and validation RMSEs of 8.71 and 10.82 days for SOS, and 7.03 and 7.77 days for EOS, respectively (Fig.3a and b). Within the study area, 67.1% and 85.8% of pixels exhibited an RMSE of less than 10 days for SOS and EOS, respectively. The correlation coefficients between observed and simulated SOS and EOS are 0.54 and 0.63, respectively.



230 **Figure 3: Performance of phenology module.** Spatial patterns of the RMSE for start-of-season (SOS, a) and end-of-season (EOS, b). Scatter plots depicting the relationship between simulated and observed start-of-season (SOS, c) and end-of-season (EOS, d). The overlined RMSE represents the average RMSE. The colored points on the scatter plots represent kernel density, with lighter colors indicating a higher density distribution. The dashed line represents the 1:1 ratio line.

In the modified SWAT-Carbon model, the fixed phenology dates were replaced by the outcomes of phenology module, and the simulated LAI before the start of the growing season was changed to a non-zero value for both forest and grassland. After these modifications, the calibrated and validated LAI growth curves largely improved compared to the original model (Figure 4a and 4b, Table 2). Specifically, the average R^2 and NSE values improved from 0.31 and -1.42 to 0.79 and 0.78 for forest, and from 0.13 and -1.84 to 0.88 and 0.87 for grassland, respectively. Additionally, the absolute PBIAS decreased by 41.0% for forest and 97.0% for grassland. To further exclude the impact of non-growing season LAI, we also adjusted the non-growing season LAI to non-zero values in the original SWAT-Carbon model. The results indicated that, compared to the default settings, there were only slight improvements in the LAI growth curves in the original model (Figure S1). Specifically, the average NSE value improved from -1.42 to -0.29 for forest and from -1.84 to -0.70 for grassland, which is much smaller than NSE for LAI in the modified model. Furthermore, the average R^2 decreased by 0.17 for forest and by 0.01 for grassland, respectively.

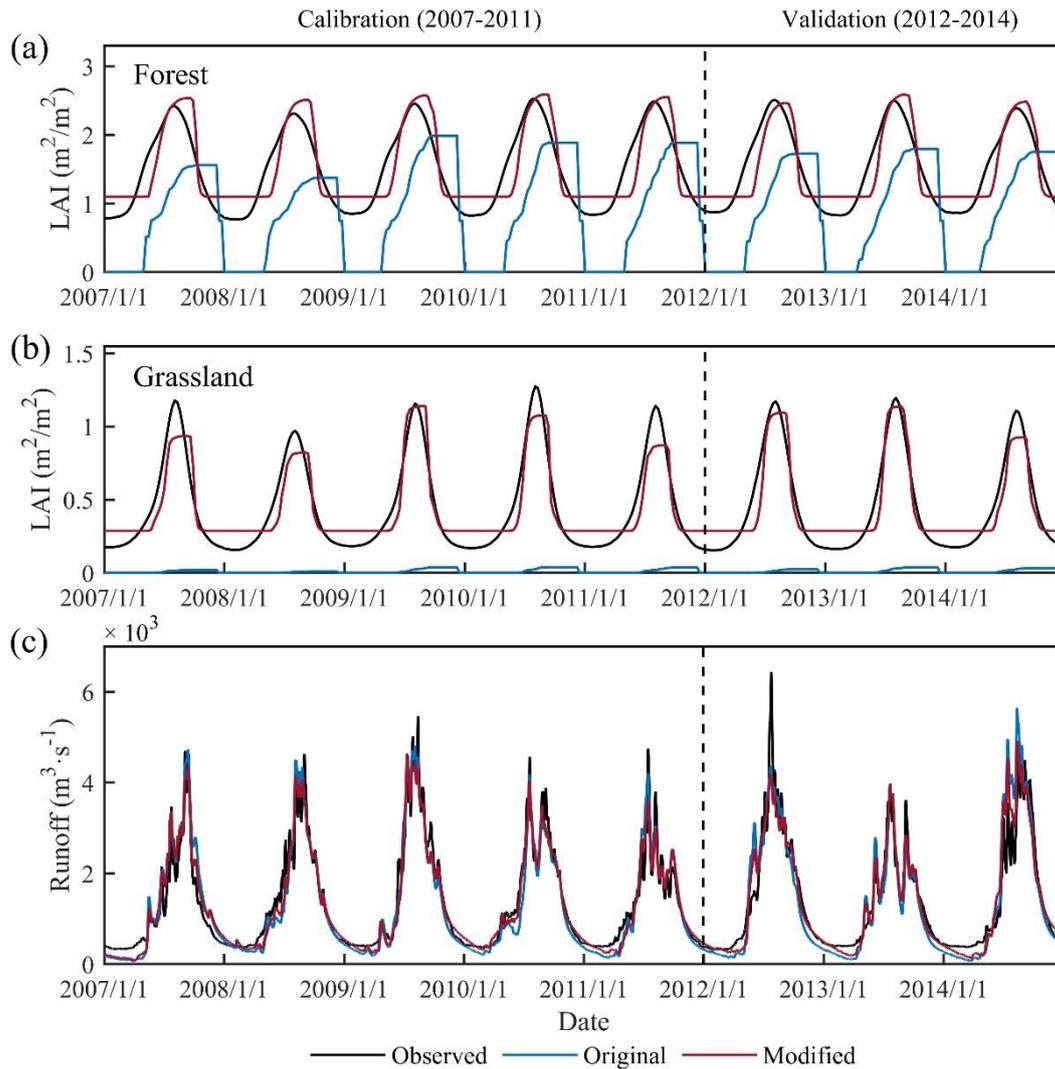
245 Incorporating the phenology module into the SWAT-Carbon model enhanced the performance of runoff simulation. The modified SWAT-Carbon model produced runoff simulation with a 'very good' level of performance ($0.80 < R^2 \leq 1$ and 0.75

250 < $NSE \leq 1$, Figure 4c and Table 2). Furthermore, the performance of runoff simulation showed improvement in various months, particularly during the vegetation greening period (June), the senescence period (October), and the non-growing season (Table S3). Specifically, the R^2 and NSE values for runoff increased by 0.13 and 0.39 in June, respectively, by 0.30 and 0.46 in October. The NSE value for runoff from November of the previous year to May increased by 0.35.

Table 2: Evaluation indices of the original and modified SWAT-Carbon model in simulating LAI and runoff.

Variable	Evaluation indices	Calibration (2007–2011)		Validation (2012–2014)	
		Original	Modified	Original	Modified
LAI (forest)	R^2	0.28	0.77	0.33	0.80
	NSE	-1.47	0.76	-1.36	0.79
	PBIAS (%)	42.77	-1.50	41.75	1.12
LAI (grassland)	R^2	0.11	0.87	0.15	0.89
	NSE	-1.87	0.86	-1.80	0.88
	PBIAS (%)	97.47	0.65	97.15	-0.01
Runoff	R^2	0.93	0.95	0.89	0.92
	NSE	0.92	0.95	0.86	0.92
	PBIAS (%)	4.80	1.75	5.04	1.22

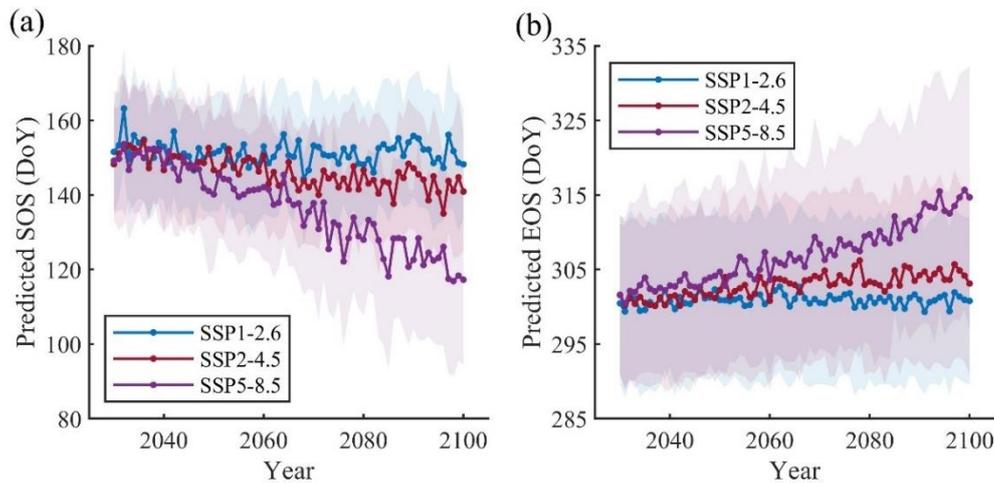
R^2 , coefficient of determination; NSE, Nash–Sutcliffe efficiency; PBIAS, percent bias.



255 **Figure 4: Temporal variability of the LAI and runoff during the calibration (2007–2011) and validation (2012–2014) periods.** 8-d LAI time series observed by satellite and simulated by the original and modified SWAT-Carbon models for Forest (a) and grassland (b). Observed and simulated daily runoff by the original and modified SWAT-Carbon models (c).

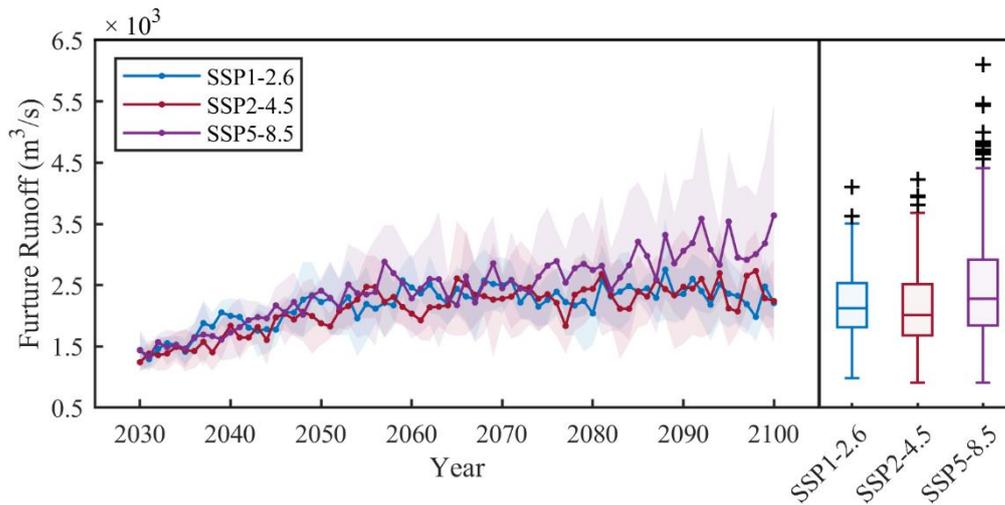
3.3 Future Shifts in Vegetation Phenology and Runoff

260 The Unichill and DM phenology models were used to forecast future shifts in SOS and EOS during 2030–2100 under different emission scenarios. The SOS in the upper reaches of the Jinsha River watershed was projected to advance under all scenarios (Figure 5a), with the most significant advancement observed under the high-emission scenario (0.49 days per year, $p < 0.01$). In addition, a delaying trend in EOS was observed under all scenarios (Figure 5b), with the largest delay occurring under the high-emission scenario (0.18 days per year, $p < 0.01$).



265 **Figure 5: Projection of future spring and autumn phenology during 2030–2100.** Future shifts in start-of season (SOS, a) and end-of-
 season (EOS, b) under three Shared Socioeconomic Pathway (SSP) scenarios based on the Coupled Model Intercomparison Project Phase
 6 (CMIP6) multi-model ensemble. SSP1-2.6, SSP2-4.5, and SSP5-8.5 refer to low, moderate, and high emissions, respectively. Colored
 lines and shading represent the mean and one standard deviation, across the four CMIP6 models under each emission scenario.

The runoff under SSP5-8.5 scenario exhibited greater fluctuation with a continuous increase from 2030 to 2100. Specifically,
 runoff was projected to increase from approximately $1,437 \text{ m}^3 \text{ s}^{-1}$ in 2030 to $3,638 \text{ m}^3 \text{ s}^{-1}$ in 2100. However, the runoff under
 270 SSP1-2.6 and SSP2-4.5 scenarios is projected to initially increase and then stabilize after 2060 (Figure 6). The average
 runoff under SSP5-8.5 ($2,459 \text{ m}^3 \text{ s}^{-1}$) is greater than that under SSP1-2.6 ($2,180 \text{ m}^3 \text{ s}^{-1}$) and SSP2-4.5 ($2,120 \text{ m}^3 \text{ s}^{-1}$). Results
 from the original SWAT-Carbon model showed a smaller increase in future runoff compared to those estimated by the
 modified SWAT-Carbon model (Figure S2). Monthly total runoff outputs were obtained from CMIP6 models and
 aggregated to annual runoff from 2030 to 2100. We calculated the watershed-scale future runoff from CMIP6 using the area-
 275 weighted mean method and found that the future runoff from CMIP6 models was larger than that simulated by the modified
 SWAT-Carbon model (Figure S3).



280 **Figure 6: Projection of future runoff during 2030–2100 using the modified SWAT-Carbon model.** Colored lines and shading in the right subplot represent the mean and one standard deviation across the four Coupled Model Intercomparison Project Phase 6 (CMIP6) models. The scenarios SSP1-2.6, SSP2-4.5, and SSP5-5.8 refer to low emission, moderate, and high emissions, respectively. The cross symbol represents the outlier.

4 Discussion

4.1 Improvement of Runoff Simulation by Modified Hydrological Model

Vegetation phenology has strong controls on seasonal and annual hydrological processes (Kim et al., 2018; Yang et al., 285 2023). However, the vegetation phenology in most hydrological models, such as the SWAT-Carbon model, is based on simple thermal and daylength thresholds (Arnold et al., 1998). This approach results in large uncertainty in simulating vegetation dynamics and thereby in predicting hydrological processes, especially under climate change (Zhang et al., 2009; Wang et al., 2022; Ma et al., 2019). In this study, we integrated process-based spring and autumn phenology models into the SWAT-Carbon model, which substantially improved the simulation of vegetation dynamics and ecohydrological processes.

290 The simulation of the LAI curve, particularly in terms of seasonality and magnitude, exhibited significant improvement in the modified SWAT-Carbon model (Figure 4a and 4b, Table 2). The inaccuracy of LAI simulation in the original model was attributed to the initiation and end of vegetation dormancy being determined only by a latitude-based daylength threshold. This approach leads to mismatches in the timing of phenological stages in the LAI curve compared to observed data (Figure 4a, 4b, and Figure S1). Based on the widely used processed phenological models (Fu et al., 2020), we incorporated the

295 phenology module into the SWAT-Carbon model, which provides a method for accurately simulating the timing of vegetation growth stages and the runoff simulation. In addition, the highest and lowest simulated LAI values in the original model showed significant underestimation. This is attributed to the inability to meet the default constant value of heat accumulation requirements, owing to low daily temperatures in the upper reach of Jinsha River watershed. In the modified model, we replaced the constant value with a dynamic heat accumulation requirement and optimized vegetation growth

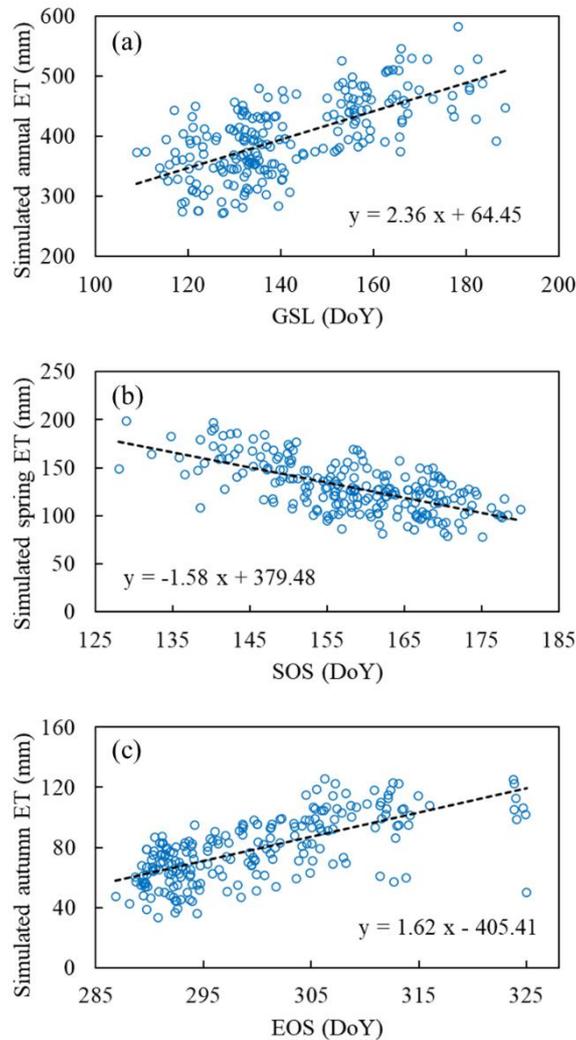
300 parameters, which finally enhanced the simulation accuracy of LAI values and thereby improved runoff simulation. Therefore, we highlight the importance of integrating the phenology module into hydrological models.

Our study indicated that the modified SWAT-Carbon model improved the simulation of vegetation and runoff. However, some uncertainties from dataset and model persist in the modified model and should be addressed. The meteorological input data used in this study were obtained from a gridded source, which may differ from actual conditions. In addition, climate-
305 sensitive ecosystem structures, such as species composition, introduce uncertainties in assessing interactions between vegetation phenology and hydrological processes in the modified SWAT-Carbon model (Chuine, 2010; Huang et al., 2019). Therefore, coupling advanced land surface dynamic vegetation models, such as LPJ-GUESS (Sitch et al., 2003), with hydrological models could further improve our understanding of future vegetation dynamics and their effects on the carbon and water cycles.

310 **4.2 Interaction between Vegetation Phenology and Hydrological Processes**

Our results indicated significant improvement in the performance of runoff simulation, particularly during the vegetation greening period (June) and the senescence period (October) (Table 2 and Table S3), which is attributed to the accurate phenology prediction. Vegetation phenology and hydrological processes are closely intertwined through biotic and abiotic pathways (Buermann et al., 2018; Lian et al., 2020). In the upper reaches of Jinsha River watershed, the multiyear mean SOS
315 and EOS occurred in June and October (Figure 2), respectively. During the start and end of the growing season, rapid changes in vegetation physiological properties, such as stomatal conductance and LAI, influence the timing and amount of water resource allocation (Hwang et al., 2023; Zhang et al., 2021).

In addition, the vegetation dynamics and their impact on the underlying land surface properties affect watershed evapotranspiration (Chen et al., 2022b). To explore the response of ET to vegetation phenology in the upper reaches of
320 Jinsha River watershed, we analyzed the relationship between the phenological variations and sub-basin scale ET using the modified SWAT-Carbon model. Consistent with previous studies in the Northern Hemisphere, our study revealed a positive correlation between ET and growing season length (Figure 7), potentially attributed to the prolonged period of water movement from the soil to the atmosphere (Geng et al., 2020; Yang et al., 2023). This was further verified by the observation that earlier leaf unfolding and delayed leaf senescence increased spring and autumn ET. Vegetation phenology influences the
325 development and decay of the canopy, altering the water demands of vegetation and affecting the hydrological cycle (Hwang et al., 2018; Lian et al., 2020). However, our study found that both spring and autumn ET increased by similar magnitudes, with values of 1.58 and 1.62 mm per day, respectively. This finding is inconsistent with the asymmetrical increase reported in previous research (Kim et al., 2018), which may be attributable to the abundant soil moisture during summer in our study area. This moisture abundance does not impose constraints on vegetation growth and transpiration.



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Figure 7: Relationship between the phenological variations and sub-basin scales evapotranspiration (ET) using modified SWAT-Carbon model. GSL, growing season length; SOS, start-of-season; EOS, end-of-season; ET, evapotranspiration.

4.3 Hydrological Response to Future Climate Change and Vegetation Dynamics

We predicted future runoff in the upper reaches of the Jinsha River watershed and found that the runoff would largely increase under future climate change conditions. Specifically, under the SSP5-8.5 scenario, runoff exhibited the most pronounced upward trend, primarily attributed to increased precipitation largely surpassing that of the SSP1-2.6 and SSP2-4.5 scenarios (Figure 6 and Figure S4). In addition to precipitation, the early SOS and delayed EOS under global warming (Figure 5), which were predicted in the upper reaches of Jinsha River watershed, also play an important role in altering the water cycle. Despite a substantial increase in precipitation under SSP2-4.5 compared to SSP1-2.6, the projected runoff under SSP2-4.5 is marginally smaller than that under SSP1-2.6. This phenomenon is likely attributed to the extension of the

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growing season under global warming, which significantly increases evapotranspiration under the moderate emission pathway compared to the low emission pathway (Lu et al., 2021; Yang et al., 2023). The increased precipitation under SSP2-4.5 is insufficient to fully counteract the increase in evapotranspiration caused by the prolonged growing season (Figure S4).

345 In this study, future runoff simulation by the original SWAT-Carbon model is smaller than that by the modified model, especially under the high emission scenario, which is also observed in the historical runoff (Figure 4c). The underestimation of historical runoff simulation is attributed to lower baseflow compared to observed data, particularly during the non-rainy season. This discrepancy is potentially influenced by underestimated LAI simulations and unreasonable parameters in the original SWAT-Carbon model (Tang et al., 2022). Vegetation degradation is reported to lead to decreased flow during the dry season (Paiva et al., 2023). In addition, although calibration algorithms enhance the original model's ability to produce
350 accurate runoff simulations, the unrealistic representation of vegetation phenological processes necessitates remediation through additional ecohydrological processes (Luan et al., 2022; Haas et al., 2022), thereby compromising the model's fidelity. Therefore, improving the simulation accuracy of vegetation dynamics could enhance the capability of hydrological models to accurately depict ecohydrological processes, especially in the dry season, thereby facilitating more effective water resource conservation strategies under future global warming.

355 **5 Conclusions**

In summary, our study integrated a process-based vegetation phenology module into the SWAT-Carbon model, substantially improving the simulation of vegetation dynamics and hydrological processes in the upper reaches of the Jinsha River watershed. The runoff is predicted to increase over 2030-2100 under different future emission scenarios, primarily due to greater precipitation outweighing the increase in evapotranspiration induced by the prolonged growing season. Our study
360 highlights the importance of integrating phenological shifts into hydrological models. In the future, it will be crucial to consider and couple plant functional types with hydrological models to further enhance the performance of ecohydrological process simulations. Furthermore, the phenology module should be incorporated into other hydrological models and applied in various watersheds to enhance our understanding of ecohydrological processes and to support future water resources management under climate change conditions.

365 **Data availability**

Datasets for driving and evaluating the phenology models and the SWAT-Carbon model are available from the sources listed in Table 1.

Author contributions

Conceptualization: YF; Data curation: NW, JZ; Formal Analysis: ML, SC; Funding acquisition: YF, FH; Methodology: ML,
370 SC, YX, ZW; Software: ML, SC; Supervision: YF; Validation: SC; Visualization: ML, JZ; Writing – original draft: ML;
Writing – review & editing: SC, FH, NW, ZW, YX, JZ, YZ, YF.

Competing interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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