

# 1 The general formulation for mean annual runoff components 2 estimation and their change attribution

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## 8 Abstract

9 Estimating runoff components, including surface flow, baseflow and total runoff is essential for  
10 understanding precipitation partition and runoff generation and facilitating water resource  
11 management. However, a general framework to quantify and attribute runoff components is still  
12 lacking. Here, we propose a general formulation through observational data analysis and  
13 theoretical derivation based on the two-stage Ponce-Shetty model (named as the MPS model).  
14 The MPS model characterizes mean annual runoff components as a function of available water  
15 with one parameter. The model is applied over 662 catchments across China and the contiguous  
16 United States. Results demonstrate that the model well depicts the spatial variability of runoff  
17 components with  $R^2$  exceeding 0.81, 0.44 and 0.80 for fitting surface flow, baseflow and total  
18 runoff, respectively. The model effectively simulates multi-year runoff components with  $R^2$   
19 exceeding 0.97, and the proportion of runoff components relative to precipitation with  $R^2$   
20 exceeding 0.94. By using this conceptual model, we elucidate the responses of surface flow and  
21 baseflow to available water and environmental factors for the first time. The surface flow is  
22 jointly controlled by precipitation and environmental factors, while baseflow is mainly influenced  
23 by environmental factors in most catchments. The universal and concise MPS model offers a new  
24 perspective on the long-term catchment water balance, facilitating broader application in  
25 large-sample investigations without complex parameterizations and providing an efficient tool to  
26 explore future runoff variations and responses under changing climate.

27 **Key Points**

28 (1) A general and concise formulation is proposed to quantify, and attribute mean annual  
29 surface flow, baseflow and total runoff.

30 (2) The formulation characterizes runoff components as a function of available water without  
31 additional and complicated parameter calculation.

32 (3) The formulation performs well in quantifying and attributing runoff components in 662  
33 catchments.

34 **1. Introduction**

35 Runoff is the primary freshwater resource accessible for human life and plays an essential role  
36 in the water cycle (He et al., 2022; Wang et al., 2024). Based on the propagation time and  
37 hydraulic response of a catchment, total runoff ( $Q$ ) can be divided into baseflow ( $Q_b$ ) and surface  
38 flow ( $Q_s$ ) (Gnann et al., 2019; Singh et al., 2019). Baseflow, also referred to as slow flow, is  
39 defined as the flow that originates from groundwater and other delayed sources (such as wetlands,  
40 lakes, snow and ice), and generally sustains streamflow during dry periods (Gnann, 2021; Hall,  
41 1968). Baseflow is the relatively stable component of runoff, playing a vital role in aquatic  
42 ecosystems (de Graaf et al., 2019; Price et al., 2011), water quality (Ficklin et al., 2016) and  
43 sustained water supplies (Fan et al., 2013). Surface flow, also referred to as fast flow, results from  
44 rapid processes like the saturation or infiltration of excess overland flow and swift subsurface  
45 flow (Beven and Kirkby, 1979), leading to immediate water movement. Surface flow occurs more  
46 rapidly and with more drastic changes than baseflow, which is primarily responsible for flood  
47 generation (Yin et al., 2018) and soil erosion (Morgan and Nearing, 2011).

48 Most current studies focus on total runoff variability and attribution, and the relevant  
49 researches are fairly mature (Berghuijs et al., 2017; Han et al., 2023; Liu et al., 2021). However,  
50 few studies pay attention to comprehensive research on the different runoff components (Li et al.,  
51 2020; Liu et al., 2019), and the attributions of  $Q_s$  and  $Q_b$  changes are still unclear (Hellwig and  
52 Stahl, 2018). Baseflow and surface flow represent different hydrological processes, and their  
53 implications for watershed management are also not identical (Zheng and Sun, 2014). For

example, the research conducted by Ficklin et al. (2016) in the United States points out apparent spatial differences between  $Q_b$  and  $Q_s$  in different seasons. Therefore, it is necessary to quantify runoff components and distinguish their controlling factors to better understand the runoff dynamics and facilitate water resources management in the context of intensified climate change and anthropogenic disturbance.

Unlike  $Q$ , which is ascertainable through direct observation at hydrological gauges,  $Q_b$  and  $Q_s$  can only be estimated through indirect methods, including baseflow separation (Wu et al., 2019; Zhang et al., 2017), isotope tracing (Hale et al., 2022; Wallace et al., 2021) and hydrological modeling (Al-Ghobari et al., 2020; Cheng et al., 2020; Huang et al., 2007; Kaleris and Langousis, 2017). The first two methods estimate  $Q_b$  initially, and  $Q_s$  is then derived as the difference between the  $Q$  and the estimated  $Q_b$ , limiting their ability to examine the dynamic variations of each runoff component independently, and the isotope tracing method is challenging to conduct on a large and long-term scale. The hydrological modeling enables to simulate  $Q_b$  and  $Q_s$  separately, typically reflected in different modules and empirical formulations. In hydrological models,  $Q_b$  is encoded using linear or non-linear storage-discharge functions (Chen and Ruan, 2023; Cheng et al., 2020).  $Q_s$  is closely related to rainfall, but the models for estimating it are usually event-based (such as the Soil Conservation Service Curve Number method (Al-Ghobari et al., 2020; SCS, 1972; Shi et al., 2017) and very few studies explored the controls on the mean annual  $Q_s$  (Neto et al., 2020). Among various models, the Budyko framework (Budyko, 1974) in conjunction with water-energy balance method (Choudhury, 1999; Yang et al., 2008) (see the second row in Table 1), has been widely used in the analysis of mean annual  $Q$  due to its simple, universal and transparent characteristics (He et al., 2022; Roderick and Farquhar, 2011).

Recently, utilizing the extended Budyko framework to estimate  $Q_b$  and  $Q_s$  has attracted attention. Wang and Wu (2013) and Neto et al. (2020) established the regression relationship between baseflow fraction ( $BFC$ , the ratio of  $Q_b$  to precipitation ( $P$ )) and aridity index ( $\phi$ , the ratio of mean annual potential evapotranspiration ( $E_0$ ) to  $P$ ) using analytical formulation. However, Gnann et al. (2019) reported that using only the  $\phi$  struggles to delineate baseflow variability in humid catchments, where the impact of soil water storage capacity ( $S_p$ ) is as critical as that of the  $\phi$ . Thus, Cheng et al. (2021) proposed an analytical curve for describing mean annual  $Q_b$  by introducing  $S_p$  as another theoretical boundary. Results show that the developed

84 curve agrees well with the observed *BFC* ( $R^2 = 0.75$ , RMSE = 0.058) and  $Q_b$  ( $R^2 = 0.86$ ,  
 85 RMSE = 0.19 mm), outperforming the original Budyko framework. Analogously, Yao et al.  
 86 (2021) derived similar functions incorporated the  $\phi$ ,  $S_p$  and a shape parameter to model *BFC* and  
 87 baseflow index (*BFI*, the ratio of  $Q_b$  to  $Q$ ). These extended Budyko frameworks accounting for  $S_p$   
 88 have advantages in simulating  $Q_b$ . However,  $S_p$  is challenging to obtain through observations and  
 89 often requires calibration (Cheng et al., 2021) or computation (Yao et al., 2021), adding certain  
 90 uncertainties to the model. Notably, the calibration performance of  $Q_s$  in equation (1) to obtain  $W_p$   
 91 (the proxy of  $S_p$ ) in the catchments of China are not always satisfactory, especially in the northern  
 92 catchments. Moreover, the complicated forms can bring inherent uncertainties and these studies  
 93 have not validated the formulations of  $Q_s$ , which are derived by subtracting  $Q_b$  from  $Q$  or fitting  
 94 curves (Cheng et al., 2021; Neto et al., 2020), implying that they may overlook the physical  
 95 processes represented by surface flow. In the subsequent discussion, the Budyko framework and  
 96 extended Budyko equations are collectively referred to as the "Budyko-type formulations" (Table  
 97 1).

98 Many researchers have observed similar behavior of  $Q_b$  to  $Q$  (Cheng et al., 2021; Gnann et al.,  
 99 2019; Wang and Wu, 2013). Is there a similar behavior for  $Q_s$ ? In a two-stage partitioning theory  
 100 (L'vovich, 1979), runoff components are delineated based on the available water at each stage.  
 101 Therefore, is there a general framework to unify different runoff components? Although various  
 102 functional forms have been proposed for estimating runoff components in the literature, a  
 103 universal method that reveals the mechanisms of mean annual runoff components generation and  
 104 subsequent quantification and attribution is still in need.

105 **Table 1.** The Budyko-type formulations for estimating mean annual runoff components

References	Formulations	Parameters
Choudhury (1999); Yang et al. (2008)	$Q = P - \frac{P \times E_0}{(P^n + E_0^n)^{1/n}}$	$n$ calibrated
Wang and Wu (2013)	$\frac{Q_b}{P} = 1 - \left[ 1 + \left( \frac{E_0}{P} \right)^{-v} \right]^{-1/v}$	$v$ fitted
Neto et al. (2020)	$f_s(\phi) = \exp(-\phi^a + \delta_s)^b$ $f_B(\phi) = \exp(-\phi^c + \delta_B)^d$	$a, b, c, d$ $\delta_s = \ln \left( \left[ \frac{\bar{Q}_s}{\bar{P}} \right]_{max} \right)^{1/b}$

$$\delta_B = \ln \left( 1 - \left[ \frac{\bar{Q}_S}{\bar{P}} \right]_{max} \right)^{1/d}$$

fitted

$$\frac{Q_s}{P} = -\frac{E_0 + S_p}{P} + \left[ 1 + \left( \frac{E_0 + S_p}{P} \right)^{\alpha_1} \right]^{1/\alpha_1}$$

Cheng et al. (2021)

$S_p, \alpha_1, \alpha_2$

$$\frac{Q_b}{P} = \frac{S_p}{P} + \left[ 1 + \left( \frac{E_0}{P} \right)^{\alpha_2} \right]^{1/\alpha_2} - \left[ 1 + \left( \frac{E_0 + S_p}{P} \right)^{\alpha_2} \right]^{1/\alpha_2}$$

calibrated

$$Q_b = \frac{P + S_b - \sqrt{(P + S_b)^2 - 2aS_bP}}{a} \left[ \begin{array}{l} 1 \\ \\ - \frac{1 + \frac{E_0}{P} \frac{P}{S_b} - \sqrt{\left( 1 + \frac{E_0}{P} \frac{P}{S_b} \right)^2 - 2a \frac{E_0}{P} \frac{P}{S_b}}}{a} \end{array} \right]$$

$S_b$  (estimated from cumulative distribution function),  $a$  (calibrated)

Yao et al. (2021)

$$Q = P - \frac{\frac{P}{S_b} + 1 - \sqrt{\left( \frac{P}{S_b} + 1 \right)^2 - 2a \frac{P}{S_b}}}{a} * \frac{E_0 + S_b - \sqrt{(E_0 + S_b)^2 - 2aE_0S_b}}{a}$$

106 Note that  $P$  is the mean annual precipitation,  $E_0$  is the mean annual potential evapotranspiration,  $f_S(\phi)$  and  
107  $f_B(\phi)$  are the surface flow and baseflow function, respectively and  $S_p$  is the catchment storage capacity.

108 To address these questions, we derived a modified two-stage partitioning framework through  
109 observational data analysis and theoretical derivation based on the Ponce-Shetty model (Ponce  
110 and Shetty, 1995; Sivapalan et al., 2011) (namely the modified Ponce-Shetty model, MPS model)  
111 at mean annual time scale. The Ponce-Shetty model is a conceptual model with physical  
112 constraint developed at annual scale to depict how precipitation is partitioned, stored and released  
113 in the catchment (Gnann et al., 2019). It posits that annual precipitation is partitioned into  $Q_s$  and  
114 soil wetting ( $W$ ) and, subsequently, the resulting  $W$  is partitioned into  $Q_b$  and vaporization ( $V$ )  
115 (Sivapalan et al., 2011). The MPS model enables large-sample catchments research, which may  
116 lead to new understanding of mean annual water balance and allocation.

117 In general, the objectives of this study are to (1) develop a general and concise formulation to  
 118 describe runoff components variability at mean annual time scale; (2) validate and compare the  
 119 performance of the developed formulation against Budyko-type formulations; (3) attribute the  
 120 variations of runoff components induced by the changes of precipitation and other factors. Here,  
 121 we modify the Ponce-Shetty model according to some conditions and hypothesize a general  
 122 runoff components model (the MPS model), that describes  $Q_s$ ,  $Q_b$  and  $Q$  as a function of  
 123 respective available water with one parameter. The MPS model is then validated over 662  
 124 catchments across China and the contiguous United States (the CONUS) over a wide range of  
 125 hydro-meteorological circumstances. The performance of the MPS model is also compared with  
 126 the Budyko-type formulations. Section 2 introduces the derivation of the MPS model. Section 3  
 127 provides the study catchments, data and the parameter estimation technique. Section 4 shows the  
 128 results followed by a discussion in Section 5. The conclusions are summarized in Section 6.

129 **2. Derivation of the Modified Ponce-Shetty Model**

130 L'vovich (1979) proposed a conceptual theory for the two-stage catchment water balance  
 131 partition at the annual time scale according to Horton's approach (Horton, 1933). Firstly,  
 132 precipitation is partitioned into surface flow ( $Q_s$ ) and catchment wetting ( $W$ , stored water), and  
 133 then, the catchment wetting is partitioned into baseflow ( $Q_b$ ) and vaporization ( $V$ , including  
 134 interception loss, evaporation and transpiration). Ponce and Shetty (1995) conceptualized the  
 135 partition of each step as the form of a competition, and derived the formulations of runoff  
 136 components based on the proportionality hypothesis. Sivapalan et al. (2011) reintroduced the  
 137 Ponce-Shetty equations as follows:

138 In the first stage,  $P = Q_s + W$ :

$$Q_s = \begin{cases} 0, & \text{if } P \leq \lambda_s W_p \\ \frac{(P - \lambda_s W_p)^2}{P + (1 - 2\lambda_s)W_p}, & \text{if } P > \lambda_s W_p \end{cases} \quad (1)$$

$$W = \begin{cases} P, & \text{if } P \leq \lambda_s W_p \\ P - \frac{(P - \lambda_s W_p)^2}{P + (1 - 2\lambda_s)W_p}, & \text{if } P > \lambda_s W_p \end{cases} \quad (2)$$

$$P \rightarrow \infty, Q_s \rightarrow P - W_p, W \rightarrow W_p \quad (3)$$

139 In the second stage,  $W = Q_b + V$ :

$$Q_b = \begin{cases} 0, & \text{if } W \leq \lambda_b V_p \\ \frac{(W - \lambda_b V_p)^2}{W + (1 - 2\lambda_b) V_p}, & \text{if } W > \lambda_b V_p \end{cases} \quad (4)$$

$$V = \begin{cases} W, & \text{if } W \leq \lambda_b V_p \\ W - \frac{(W - \lambda_b V_p)^2}{W + (1 - 2\lambda_b) V_p}, & \text{if } W > \lambda_b V_p \end{cases} \quad (5)$$

$$W \rightarrow \infty, Q_b \rightarrow W - V_p, V \rightarrow V_p \quad (6)$$

140 where  $\lambda_s$  and  $\lambda_b$  are the surface flow and baseflow initial abstraction coefficients, respectively,  
141 which range from 0 to 1. The larger value of  $\lambda$ , the more difficult it is to generate flow.  $W_p$  and  $V_p$   
142 are catchment wetting potential and vaporization potential, respectively, which are greater than 0.  
143 The relative  $\lambda_s W_p$  and  $\lambda_b V_p$  are the surface flow and baseflow generation thresholds,  
144 respectively.

145 Note that the interannual water storage change is supposed to be negligible (Ponce and Shetty,  
146 1995). In a companion paper of Sivapalan et al. (2011), Harman et al. (2011) employed the  
147 annual Ponce-Shetty model at mean annual time scale and validated its applicability. Using the  
148 first phase as an example,  $Q_s$  can be considered a function of  $\lambda_s$ , denoted as  $f(\lambda_s)$ :

$$f(\lambda_s) = \begin{cases} 0, & \text{if } \lambda_s \geq P/W_p \\ \frac{(P - \lambda_s W_p)^2}{P + (1 - 2\lambda_s) W_p}, & \text{if } \lambda_s < P/W_p \end{cases} \quad (7)$$

149 When  $\lambda_s < P/W_p$ , the Taylor expansion of  $f(\lambda_s)$  at  $\lambda_s=0$  is:

$$f(\lambda_s) = f(0) + f'(0) * \lambda_s + \frac{f''(0)}{2!} * \lambda_s^2 + \dots + \frac{f^n(0)}{n!} * \lambda_s^n + \dots \quad (8)$$

150 Hence, we have the zeroth-order approximation:

$$f(\lambda_s) \approx \frac{P^2}{P + W_p} \quad (9)$$

151 When the remainder term is relatively small, an approximation equation can be used to  
152 estimate the multi-year  $Q_s$  as:

$$Q_s = \frac{P^2}{P + W_p} \quad (10)$$

153 In addition, the zeroth-order approximation of  $Q_b$  can be similarly obtained as:

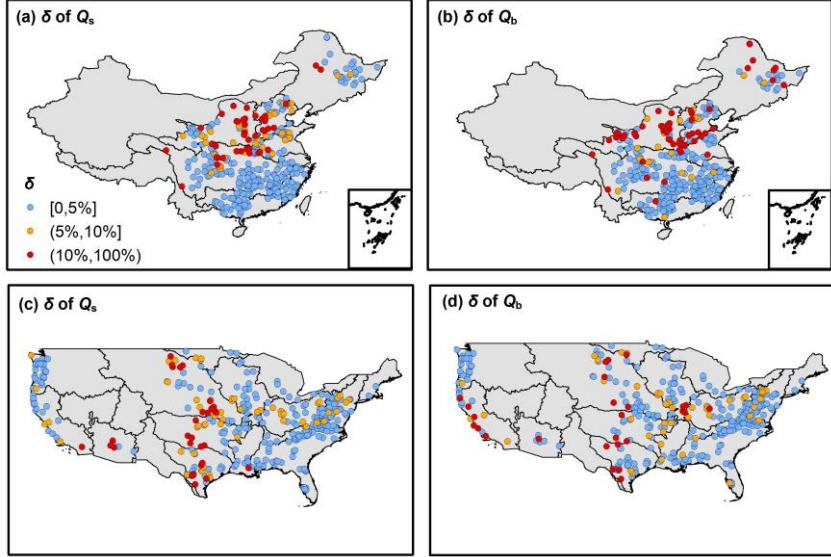
$$Q_b = \frac{W^2}{W + V_p} \quad (11)$$

154 To evaluate the impact of the remainder term, we calculate the relative bias ( $\delta$ ) of runoff  
 155 components for 312 basins in China and 350 basins in the United States using the approximate  
 156 equations (Eq (10) and Eq (11)) and the original Ponce-Shetty equations (Eq (1) and Eq (4)) (data  
 157 sources in Section 3.1). The parameters in the original Ponce-Shetty equations are calibrated  
 158 using the nonlinear least squares method. The  $\delta$  is calculated as:

$$\delta = \frac{|\widetilde{Q}_y - Q_y|}{Q_y} \quad (12)$$

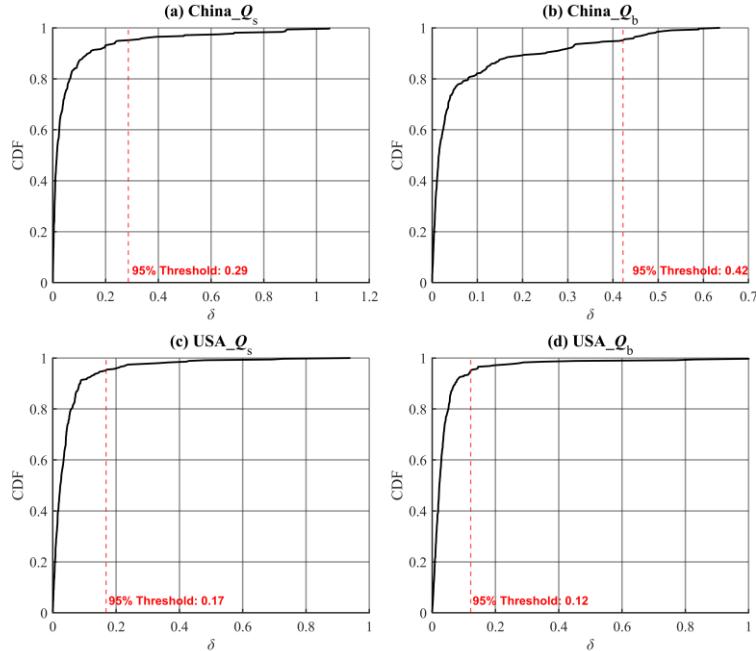
159 where  $Q_y$  represents runoff components estimated by the Ponce-Shetty equations, and  $\widetilde{Q}_y$   
 160 represents runoff components estimated by the approximate equations (Eq (10) and Eq (11)).

161 The spatial distribution of  $\delta$  and the cumulative distribution functions (CDFs) of  $\delta$  are  
 162 shown in Figure 1 and Figure 2, respectively. As shown in Figure 1, 77% of the basins have an  $\delta$   
 163 of less than 5%. The average  $\delta$  for estimating  $Q_s$  is 6.5% in China and 4.8% in the United States,  
 164 while the average  $\delta$  for estimating  $Q_b$  is 7.9% in China and 6.6% in the United States, with  
 165 larger deviations observed in arid basins. Figure 2 indicate that the  $\delta$  values for the approximate  
 166 model are within acceptable limits across both China and CONUS. The relatively low 95%  
 167 threshold values, particularly for the USA datasets, suggest that the majority of predictions fall  
 168 within a narrow error range, indicating robust model performance. This acceptability of  $\delta$  across  
 169 regions and variables highlights the approximate equations' capability to maintain prediction  
 170 accuracy under varying geographical and hydrological conditions, indicating that the Zeroth-order  
 171 approximation is representative for the original Ponce-Shetty model.



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**Figure 1.** The distribution of relative bias ( $\delta$ ) between the results by the approximate equations (Eq (10) and Eq (11)) versus the original Ponce-Shetty equations (Eq (1) and Eq (4)). The first row shows the results for 312 basins in China, and the second row shows the results for 350 basins in CONUS. The first column corresponds to surface flow ( $Q_s$ ), and the second column corresponds to baseflow ( $Q_b$ ).



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**Figure 2.** Cumulative distribution functions (CDFs) of the relative bias ( $\delta$ ) for each dataset, represented by four subplots corresponding to different regions and variables: (a) China\_  $Q_s$ , (b) China\_  $Q_b$ , (c) USA\_  $Q_s$ , and (d) USA\_  $Q_b$ . Each subplot includes a red dashed line indicating the 95%  $\delta$  threshold

183 Therefore, we can approximately consider that on a multi-year scale,  $Q_s$  and  $Q_b$  can be  
184 estimated using the zeroth-order approximation in Eq (10) and Eq (11). We subsequently assume  
185 a similar formulation of mean annual  $Q$ :

$$Q = \frac{P^2}{P + U_p} \quad (13)$$

186 where  $U_p$  is the parameter representing the upper limit of the portion remaining after precipitation  
187 is allocated to runoff, hereafter we refer to  $U_p$  as evapotranspiration potential.

188 Integrating equations (10), (11) and (13), we conclude a general formulation to depict  
189 multi-year variability of runoff components and their quantification, hereafter referred to as the  
190 modified Ponce-Shetty model (the MPS model):

$$Q_y = \frac{X^2}{X + M} \quad (14)$$

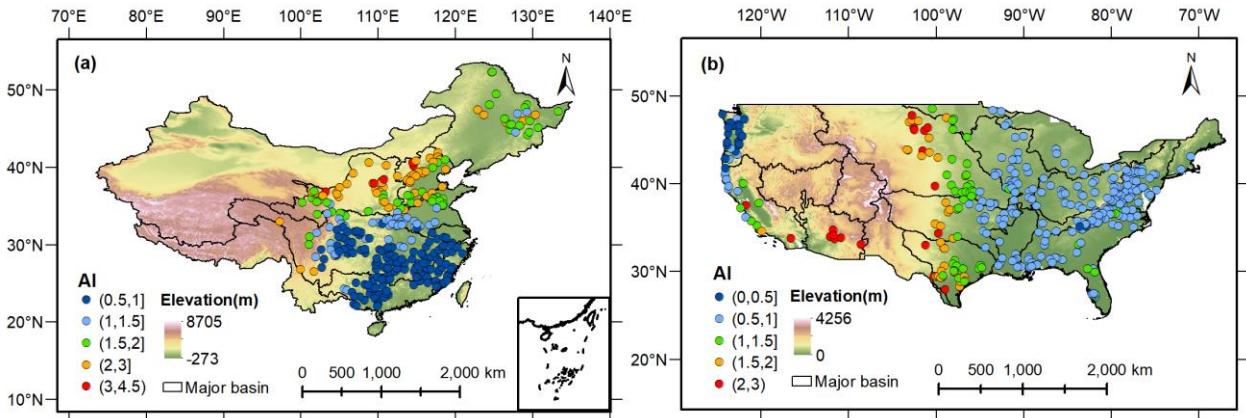
191 where  $Q_y$  represents runoff components (i.e.,  $Q$ ,  $Q_s$ ,  $Q_b$ ),  $X$  corresponds to the available water of  
192 each runoff component, i.e.,  $P$  is the available water of  $Q$  and  $Q_s$ , and  $W$  the available water of  $Q_b$ .  
193  $M$  is an integrated parameter, representing the comprehensive effects of catchment characteristics  
194 and atmospheric water and energy demand.

195 The MPS model encodes runoff components as a function of available water with only one  
196 parameter, which not only considers processes of runoff generation with physical constraints, but  
197 also, compared to the Budyko-type formulations and the original Ponce-Shetty model, is more  
198 concise in form and requires fewer parameters. Therefore, it is possible to estimate the long-term  
199 runoff components when only long-term variables are known.

### 200 **3. Data and Methodology**

#### 201 **3.1. Data**

202 To validate the reliability of the MPS model, daily hydrological and meteorological data from  
203 312 catchments in China (Li et al., 2024) and 350 catchments in the CONUS are collected. The  
204 criteria for catchments screening can refer to He et al. (2025). The location of all the catchments  
205 hydrological stations is shown in Figure 3.



206  
207 **Figure 3.** Location of hydrological stations for the (a) 312 catchments in China and (b) 350  
208 catchments in the CONUS, colored by the value of aridity index ( $\phi$ , namely  $E_0/P$ ).

209 In China, precipitation data at  $0.25^\circ$  spatial resolution are obtained from the China  
210 Gauge-based Daily Precipitation Analysis (CGDPA) (Shen and Xiong, 2016). Other  
211 meteorological data, including wind speed, sunshine hours, relative humidity, and air temperature,  
212 are from about 736 stations of the China Meteorological Data Service Center  
213 (<http://data.cma.cn/en>, last access: 11 November 2023). The in-site meteorological data are  
214 interpolated into a 10-km grid using the inverse-distance weighted method (Yang et al., 2014).  
215 We use the Penman equation (Penman, 1948) to estimate  $E_0$  of each grid using standard  
216 meteorological inputs (e.g., radiation, humidity, wind, temperature). The Penman equation is  
217 widely recommended to estimate  $E_0$  at catchment scale due to its physical basis (Pimentel et al.,  
218 2023; Wang et al., 2025), and it provides a consistent reference for our annual, large-sample  
219 analyses. The aridity index  $\phi$  is subsequently calculated as  $E_0/P$ . All grid data are aggregated  
220 and lumped for individual catchments. The discharge data are collected from the Hydrological  
221 Bureau of the Ministry of Water Resources of China (<https://www.mwr.gov.cn/english/>, last  
222 access: 20 December 2023) and are selected based on the length of records exceeding 35 years  
223 with less than 5% missing data. The time range for all data is 1960-2000.

224 In the CONUS, we use data set from CAMELS (Addor et al., 2017; Newman et al., 2015). The  
225 CAMELS data set provides 662 catchments with daily time series of precipitation and observed  
226 runoff along with aridity index, and most catchments contain 35 years of continuous runoff from  
227 1980 to 2014. The criteria for excluding catchments are referred to Gnann et al. (2019), and  
228 finally 350 catchments remained.

229 We use the one-parameter Lyne-Hollick digital filter (Lyne and Hollick, 1979) to separate  
230 daily  $Q_s$  and  $Q_b$  from daily  $Q$ . The Lyne-Hollick method is applied forward, backward, and  
231 forward again with a filter parameter of 0.925 and has manifested to be reliable to obtain runoff  
232 components (Lee and Ajami, 2023). We use the separated  $Q_s$  and  $Q_b$  as the reference. Although  
233 there are other baseflow separation algorithms, according to Troch et al. (2009), the choice of  
234 baseflow separation algorithm is not a significant determinant of the water balance at the annual  
235 scale.

236 All the hydrological and meteorological data are aggregated to the annual and mean annual  
237 time scales for further analysis.

### 238 **3.2. Calibration and Validation**

239 Spatially, to verify the MPS model's ability to characterize the variability of runoff components  
240 between catchments, we utilize the least squares fitting algorithm to estimate parameters, i.e.,  $W_p$ ,  
241  $V_p$  and  $U_p$ . The three parameters are restricted to being between 0 mm and 50, 000 mm, which is  
242 considered high enough to not affect the parameter estimation (Gnann et al., 2019).

243 In terms of time, we split all data into two periods for parameter calibration and validation of  
244 Eq. (14) for individual catchments. In China, the data ranges from 1960 to 2000, so we use the  
245 first 31 years (1960-1990) as the calibration period and the remaining 5-10 years (1991-2000) as  
246 the validation period. In the CONUS, the calibration period is chosen as 1980-2000, and the  
247 validation period is from 2001 to 2014. When we know mean annual  $Q_s$ ,  $Q_b$ ,  $Q$ ,  $P$  and  $W$  of the  
248 first period, the parameters, i.e.,  $W_p$ ,  $V_p$  and  $U_p$ , can be derived from Eq. (14). Postulating the  
249 parameters remain unchanged during two periods, we consequently can estimate the mean annual  
250  $Q_s$ ,  $Q_b$  and  $Q$  of the second period using Eq. (14). Note that the catchment wetting  $W$  is calculated  
251 as the difference of the  $P$  and estimated  $Q_s$ .

252 The surface flow fraction (*SFC*, the ratio of surface flow to precipitation) and baseflow fraction  
253 (*BFC*, the ratio between baseflow and precipitation) represent the proportion of rainfall becoming  
254 different runoff components, which are commonly used to quantify surface flow and baseflow  
255 (Wang and Wu, 2013). Therefore, we evaluate the simulation of *SFC* and *BFC* as well as the  
256 volume of runoff components.

257 The performance of the MPS model is evaluated by the coefficient of determination ( $R^2$ ) and

258 the root mean square error (RMSE):

$$R^2 = \left( \frac{\sum_{i=1}^N (X_{sim,i} - \bar{X}_{sim})(X_{obs,i} - \bar{X}_{obs})}{\sqrt{\sum_{i=1}^N (X_{sim,i} - \bar{X}_{sim})^2 \sum_{i=1}^N (X_{obs,i} - \bar{X}_{obs})^2}} \right)^2 \quad (15)$$

$$RMSE = \sqrt{\frac{1}{N} \sum_{i=1}^N (X_{sim,i} - X_{obs,i})^2} \quad (16)$$

259 where  $X$  represents the evaluated variable, i.e., mean annual  $Q$ ,  $Q_s$  and  $Q_b$ ,  $SFC$  and  $BFC$  in this  
260 study. The subscript  $obs$  and  $sim$  represents the observed and simulated value, respectively.  
261 Higher  $R^2$  and lower RMSE indicate good model performance.

262 **3.3. Attribution analysis**

263 We split the data into the first period (1960-1990 in China and 1980-2000 in the CONUS) and  
264 the second period (1991-2000 in China and 2001-2014 in the CONUS) to attribute runoff  
265 components variation between two periods. Note that the attribution of  $\Delta Q$  is only conducted in  
266 China because the  $E_0$  in CAMELS dataset is a constant in each catchment. In the MPS model, we  
267 consider that the runoff changes between two long-term periods are caused by available water and  
268 other environmental and anthropogenic factors (such as land cover/use change and  
269 evapotranspiration variation) encoded by parameters. For the changes of surface flow ( $\Delta Q_s$ ) and  
270 total runoff ( $\Delta Q$ ), postulating that each variable is independent in the MPS model, the first-order  
271 approximation of the  $\Delta Q_s$  and  $\Delta Q$  from the second period to the first period can be expressed as  
272 (Milly and Dunne, 2002):

$$\Delta Q_s = \frac{\partial Q_s}{\partial P} \Delta P + \frac{\partial Q_s}{\partial W_p} \Delta W_p \quad (17a)$$

$$\Delta Q = \frac{\partial Q}{\partial P} \Delta P + \frac{\partial Q}{\partial U_p} \Delta U_p \quad (17b)$$

273 where the two terms on the right side of equation (17a) respectively represent changes in  $Q_s$   
274 caused by changes in  $P$  ( $\Delta Q_{s-P}$ ) and other factors ( $\Delta Q_{s-W_p}$ ), and the two terms on the right side  
275 of equation (17b) respectively represent changes in  $Q$  caused by changes in  $P$  ( $\Delta Q_P$ ) and other  
276 factors ( $\Delta Q_{W_p}$ ). For convenience, we refer partial derivative coefficient  $\frac{\partial Q_s}{\partial P}$ ,  $\frac{\partial Q_s}{\partial W_p}$ ,  $\frac{\partial Q}{\partial P}$  and  $\frac{\partial Q}{\partial U_p}$  in

277 equation (17) as  $\zeta_{Qs-P}$ ,  $\zeta_{Qs-Wp}$ ,  $\zeta_{Q-P}$  and  $\zeta_{Q-Wp}$ , which can be calculated as:

$$\zeta_{Qs-P} = \frac{P^2 + 2PW_p}{(P + W_p)^2} \quad (18a)$$

$$\zeta_{Qs-Wp} = \frac{-P^2}{(P + W_p)^2} \quad (18b)$$

$$\zeta_{Q-P} = \frac{P^2 + 2PU_p}{(P + U_p)^2} \quad (18c)$$

$$\zeta_{Q-Wp} = \frac{-P^2}{(P + U_p)^2} \quad (18d)$$

278 The changes of baseflow ( $\Delta Q_b$ ) is induced by the variations of the  $W$  and  $V_p$ . However, we  
 279 focus more on the impact of  $P$  in application. Therefore, we combine equation (10), (11) and  $W =$   
 280  $P - Q_s$ , so the  $Q_b$  can be calculated as :

$$Q_b = \frac{P^2 W_p^2}{(P + W_p)(PW_p + PV_p + W_p V_p)} \quad (19)$$

281 The  $\Delta Q_b$  can be attributed as the variations of  $P$ ,  $W_p$  and  $V_p$ :

$$\Delta Q_b = \frac{\partial Q_b}{\partial P} \Delta P + \frac{\partial Q_b}{\partial W_p} \Delta W_p + \frac{\partial Q_b}{\partial V_p} \Delta V_p \quad (20)$$

282 where the three terms on the right side of equation (20) respectively represent changes in  $Q_b$   
 283 caused by changes in  $P$  ( $\Delta Q_{b-P}$ ),  $W_p$  ( $\Delta Q_{b-Wp}$ ) and  $V_p$  ( $\Delta Q_{b-Vp}$ ). The partial derivative  
 284 coefficient  $\frac{\partial Q_b}{\partial P}$  ( $\zeta_{Qb-P}$ ),  $\frac{\partial Q_b}{\partial W_p}$  ( $\zeta_{Qb-Wp}$ ) and  $\frac{\partial Q_b}{\partial V_p}$  ( $\zeta_{Qb-Vp}$ ) can be calculated as:

$$\zeta_{Qb-P} = \frac{2P^2 W_p^3 V_p + P^2 W_p^4 + 2P W_p^4 V_p}{(P + W_p)^2 (PW_p + PV_p + W_p V_p)^2} \quad (21a)$$

$$\zeta_{Qb-Wp} = \frac{P^4 W_p^2 + 2P^4 W_p V_p + 2P^3 W_p^2 V_p}{(P + W_p)^2 (PW_p + PV_p + W_p V_p)^2} \quad (21b)$$

$$\zeta_{Qb-Vp} = \frac{-P^2 W_p^2}{(P + W_p)^2 (PW_p + PV_p + W_p V_p)^2} \quad (21c)$$

285 To verify the applicability of the MPS model for runoff components attribution, we compare  
 286 the calculated  $\Delta Q_s$ ,  $\Delta Q_b$  and  $\Delta Q$  using equation (17) and (20) with the observed  $\Delta Q_s$ ,  $\Delta Q_b$   
 287 and  $\Delta Q$  between two periods. The evaluation metrics are  $R^2$  and RMSE.

288 The relative contribution ratios of  $P$  and other factors to runoff components change are  
289 calculated as:

$$\eta_P = \frac{|\Delta Q_{y-P}|}{|\Delta Q_{y-P}| + |\Delta Q_{y-Wp}| + |\Delta Q_{y-Vp}|} \times 100\% \quad (22a)$$

$$\eta_{Wp} = \frac{|\Delta Q_{y-Wp}|}{|\Delta Q_{y-P}| + |\Delta Q_{y-Wp}| + |\Delta Q_{y-Vp}|} \times 100\% \quad (22b)$$

$$\eta_{Vp} = \frac{|\Delta Q_{y-Vp}|}{|\Delta Q_{y-P}| + |\Delta Q_{y-Wp}| + |\Delta Q_{y-Vp}|} \times 100\% \quad (22c)$$

290 where  $\eta_P$ ,  $\eta_{Wp}$  and  $\eta_{Vp}$  are the relative contribution ratios of  $P$ ,  $W_p$  and  $V_p$  to runoff  
291 components, respectively. We subsequently use the absolute values of  $\eta$  to identify the dominant  
292 factor impacting runoff components.

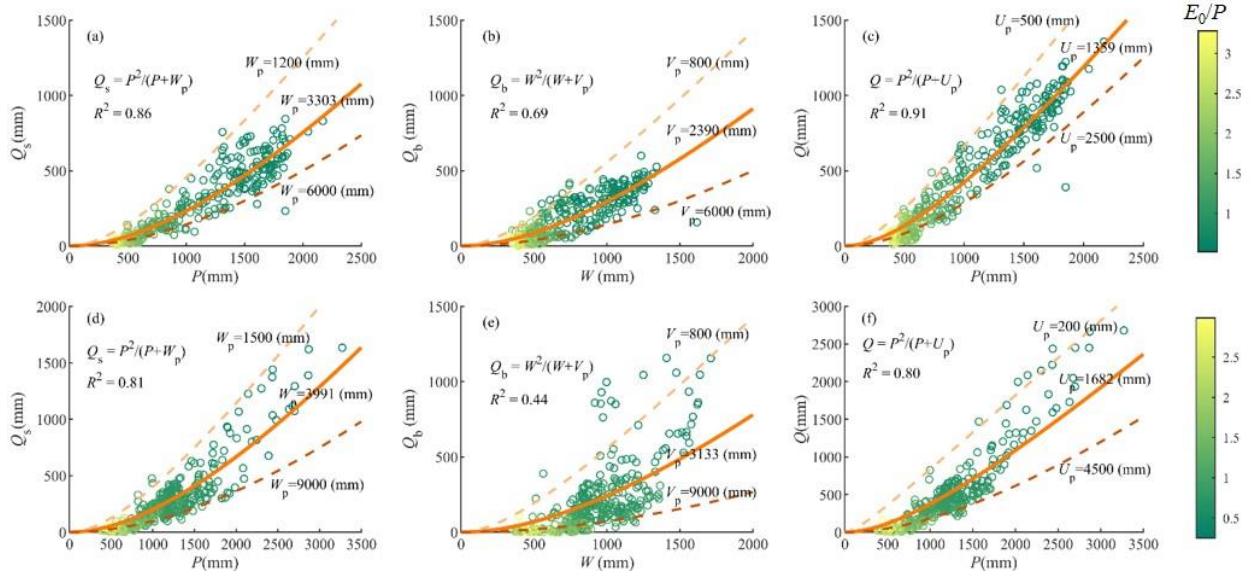
## 293 4. Results

### 294 4.1. Inter-Catchment Variability of Runoff Components

295 We employ the MPS model to fit the relationship between mean annual available water and  
296 runoff components. In China, as shown in Figure 4(a-c), the MPS model performs well in  
297 describing runoff components variability between catchments, with  $R^2$  values of 0.86, 0.68 and  
298 0.91 for fitting  $Q_s$ ,  $Q_b$  and  $Q$ , respectively. The solid lines are the best-fitted MPS curves derived  
299 using the least squares fitting algorithm, implying the median values of different parameters. We  
300 also give the potential upper and lower limits of  $W_p$ ,  $V_p$  and  $U_p$  across catchments. Similarly,  
301 Figure 4(d-f) illustrates that the MPS model achieves good fitting in the CONUS, with  $R^2$  of 0.81,  
302 0.44 and 0.80 for fitting  $Q_s$ ,  $Q_b$  and  $Q$ , respectively. The fitted parameters in the CONUS are  
303 smaller than those in China, while they have more comprehensive ranges between catchments,  
304 meaning a more significant heterogeneity in climate and underlying surface.

305 Figures 4 demonstrates that the MPS model can effectively reproduce the spatial variability of  
306 different runoff components along with the aridity index ( $E_0/P$ ), which are primarily controlled by  
307 the available water of the corresponding partition stage. The performance of MPS model to fit  $Q_s$   
308 and  $Q$  is better than that of  $Q_b$ , indicating that the factors controlling  $Q_b$  are more complicated and

309 not fully reflected in the model. With catchment properties and other factors (integrated by the  
 310 parameters in the MPS model) remaining unchanged, the more the available water, the higher the  
 311 runoff generated. Conversely, smaller parameter values are associated with greater runoff for a  
 312 given amount of available water.



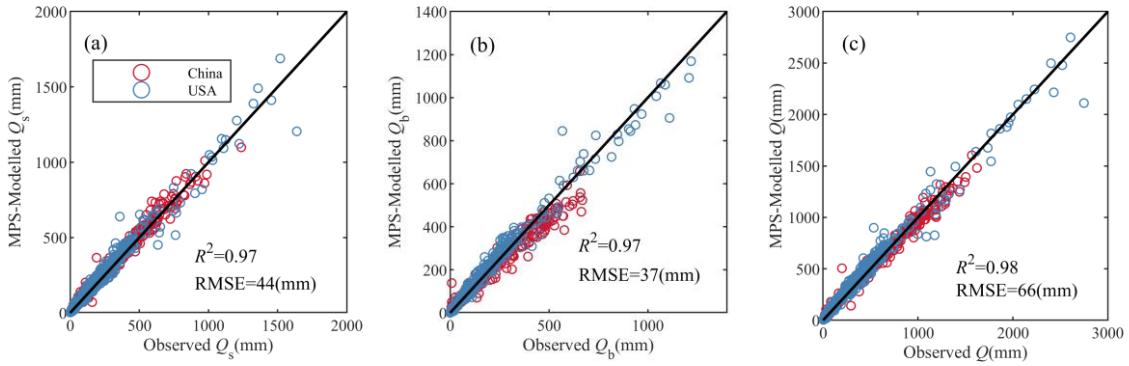
313 **Figure 4.** The MPS model relating (a)  $P$  versus  $Q_s$ , (b)  $W$  versus  $Q_b$  and (c)  $P$  versus  $Q$  in China  
 314 and (d)  $P$  versus  $Q_s$ , (e)  $W$  versus  $Q_b$  and (f)  $P$  versus  $Q$  in the CONUS. The lines are the fitted  
 315 MPS curves with best fitting (solid line) and potential upper limit and lower limit (dashed lines)  
 316 parameters.

#### 318 4.2. Validation of Runoff Components Estimation

319 Figure 5 shows the estimated mean annual  $Q_s$ ,  $Q_b$  and  $Q$  in validation periods using the MPS  
 320 model with inverted parameters in equation (14) in China and the CONUS. The simulated runoff  
 321 components match very well with the observed, with  $R^2$  greater than 0.97 and RMSE less than 66  
 322 mm. There is no significant difference in the performance in simulating  $Q_s$ ,  $Q_b$ , and  $Q$ , except for  
 323 a slight underestimation in simulating  $Q_b$  of catchments in China and some in the CONUS.

324 In panels (a), (b), and (c), we observe that the scatter points for both China (red circles) and the  
 325 CONUS (blue circles) are closely aligned with the 1:1 line, further underscoring the strong  
 326 correlation between modeled and observed values. Specifically, the results show that the MPS  
 327 model effectively captures surface flow ( $Q_s$ ), baseflow ( $Q_b$ ), and total runoff ( $Q$ ) for both regions.  
 328 Despite the generally good performance, a slight underestimation of  $Q_b$  is evident in a subset of  
 329 catchments in China and, to a lesser extent, in the CONUS. However, these discrepancies are

330 minimal and do not significantly detract from the model's overall accuracy.

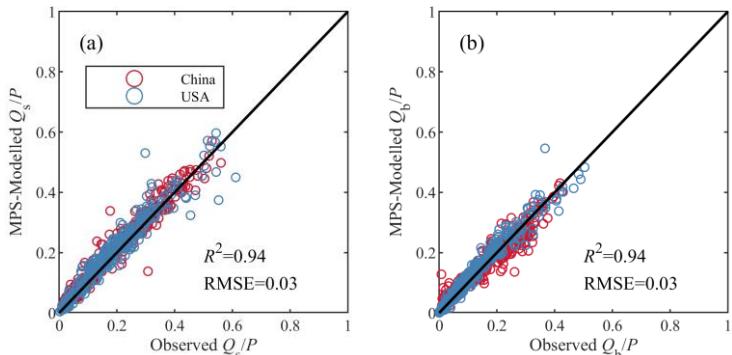


331

332 **Figure 5.** The observed and simulated mean annual (a) surface flow, (b) baseflow and (c) total  
333 runoff by the MPS model in China (red circles) and the CONUS (blue circles).

334 Figure 6 presents the estimation of *SFC* and *BFC* in validation periods using the MPS model.  
335 Similar to the simulation of  $Q_s$ , the two methods also show highly consistent estimation of *SFC*  
336 (panel (a)), with  $R^2$  of 0.94 and RMSE of 0.03. This demonstrates the MPS model's robust  
337 capability to estimate the surface flow fraction in China and the CONUS, closely aligning with  
338 the observed data. Panel (b) presents the estimation of *BFC*, where the MPS model achieves  
339 significant accuracy, reflected by the same  $R^2$  and RMSE values (0.94 and 0.03, respectively).  
340 This strong performance indicates that the MPS model is highly effective in simulating *SFC* and  
341 *BFC* across various catchments.

342



343

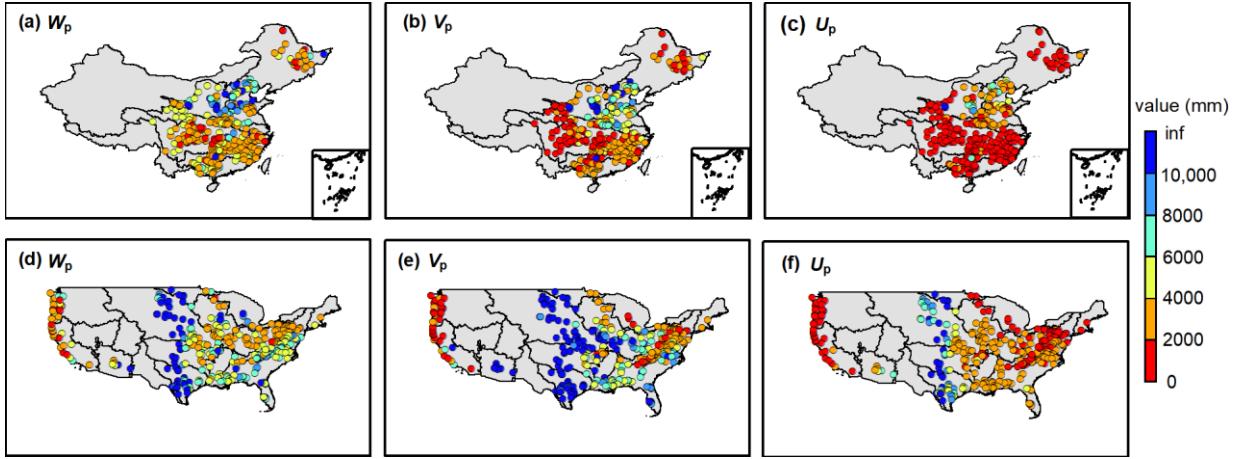
344 **Figure 6.** The observed and simulated (a) surface flow fraction ( $Q_s/P$ ) and (b) baseflow fraction  
( $Q_b/P$ ) by the MPS model in China (red circles) and the CONUS (blue circles).

345 Figure 5 and Figure 6 document that the MPS model can effectively estimate the multi-year  
346 average of all runoff components and the proportions of precipitation allocated to runoff.

347 The good validation performance of the MPS model verified our hypothesis that the parameters

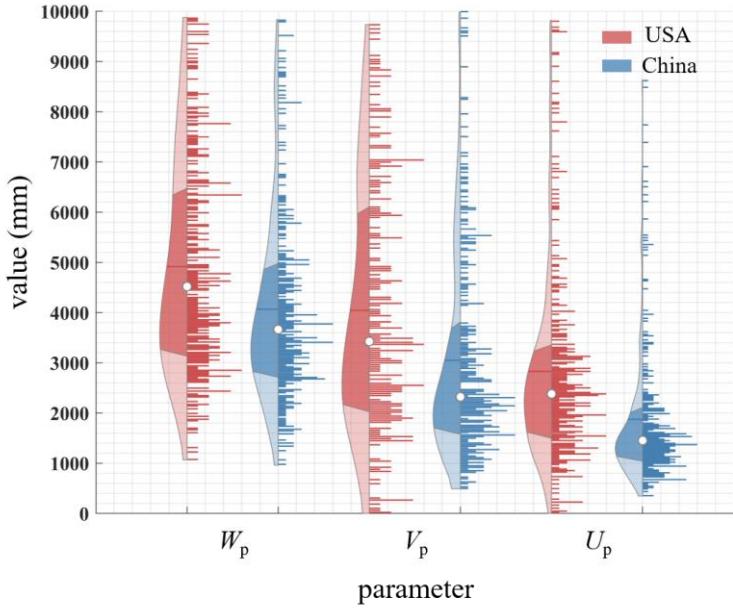
348 in the general formulations remain stable at the mean annual time scale. The parameters reflect  
349 the comprehensive impact of climate and catchment characteristics, i.e., catchment wetting  
350 potential ( $W_p$ ), vaporization potential ( $V_p$ ) and the upper limit of the portion remaining after  
351 precipitation is allocated to runoff ( $U_p$ ). As shown in Figure 7(a-c), the spatial distribution of the  
352 parameters across China exhibits pronounced divergence between the northern and southern  
353 catchments, as well as the eastern and the western. The  $W_p$ ,  $V_p$  and  $U_p$  exhibit similar spatial  
354 patterns, which can be approximately divided into two tiers from north to south. In the catchments  
355 of the Songliao River Basin in the northeast, the Yangtze River Basin and Pearl River Basins in  
356 the south, the parameters are relatively small, with  $W_p$  and  $U_p$  ranging from 0 to 2000 mm, and  $V_p$   
357 from 0 to 4000 mm, resulting large flow. In the catchments of the Yellow River Basin, Huaihe  
358 River Basin and Haihe River Basin in the north, the parameters are quite large and usually more  
359 than 5000 mm and even 8000 mm, leading to small flow. From west to east,  $W_p$  exhibits higher  
360 values in the Yangtze and Yellow Rivers Basin sources, whereas  $V_p$  and  $U_p$  are smaller in the  
361 source regions. This disparity may reflect variations in the two-stage partition of precipitation,  
362 contributing to spatial differences in total runoff. According to Figure 7(c), we can deduce that  
363 the spatial distribution of higher total runoff in south and lower in north across China, aligning  
364 with previous observational studies (He et al., 2021; He et al., 2022; Yang et al., 2019).

365 Figure 7(d-f) shows an evident west-east discrepancy of the three parameters across the  
366 CONUS. Typically,  $W_p$ ,  $V_p$  and  $U_p$  of the catchments in the west coast and eastern regions are less  
367 than 5000 mm, while parameters in the central United States are extensive with values more than  
368 8000 mm. This indicates relatively low flow in the central regions. Notably, the parameters upper  
369 limits in the catchments of the CONUS are significantly higher than those in China. The  
370 extremely large values may be associated with significant parameter uncertainty (Gnann et al.,  
371 2019). Figure 7 demonstrates that the values of the three parameters are larger in arid catchments  
372 and their spatial patterns are similar to that of climate zoning, which provides insights for  
373 parameterization.



374  
375 **Figure 7.** The (a) wetting potential ( $W_p$ ), (b) vaporization potential ( $V_p$ ) and (c)  
376 evapotranspiration potential ( $U_p$ ) of the catchments in China and (d) wetting potential ( $W_p$ ), (e)  
377 vaporization potential ( $V_p$ ) and (f) evapotranspiration potential ( $U_p$ ) of the catchments in the  
378 CONUS.

379 Figure 8 shows the violin plots of the parameters in the catchments of China and the CONUS.  
380 The median values of  $W_p$ ,  $V_p$ , and  $U_p$  in China are 3659 mm, 2220 mm and 1453 mm,  
381 respectively. The median values of  $W_p$ ,  $V_p$ , and  $U_p$  in the CONUS are 4531 mm, 3424 mm and  
382 2385 mm, respectively. Overall parameters in China are smaller and denser than those in the  
383 CONUS, implying a smaller variability of runoff components in China. Furthermore, the  $C_v$  value  
384 of  $V_p$  (1.6 in China and 6.8 in the CONUS) is the largest, followed by  $U_p$  (0.9 in China and 1.6 in  
385 the CONUS), and the smallest for  $W_p$  (0.6 in China and 1.5 in the CONUS). This indicates that  
386 the parameter dispersion controlling the second partition stage of rainfall is the greatest, which  
387 could partly account for the challenges in accurately estimating  $Q_b$ .

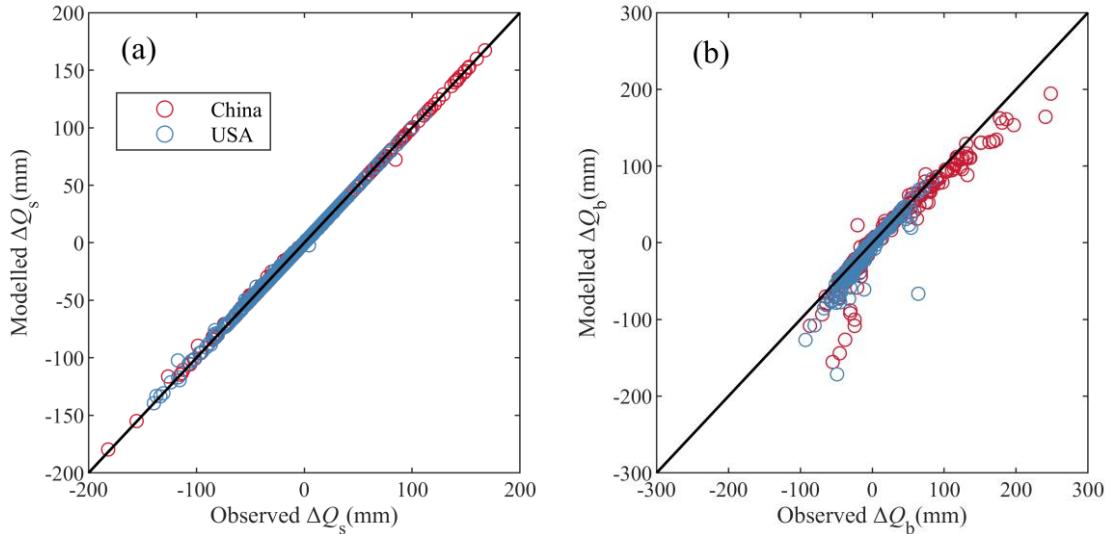


388

389 **Figure 8.** Violin plots of the parameters in the catchments of China and the CONUS. In each  
 390 violin plot, the left side represents the distribution, with the shaded area indicating the box plot,  
 391 the dot representing the mean, and the right side showing the histogram. The length of the  
 392 histogram represents the number of catchments (values larger than 10,000 are not shown).

393 **4.3. The Changes Attribution of Runoff Components**

394 The metrics to evaluate the attribution results between the changes of the observed and  
 395 simulated runoff components are shown in Table 2. We use the MPS model to estimate the  
 396 changes of  $Q_s$  ( $\Delta Q_s$ ),  $Q_b$  ( $\Delta Q_b$ ) and  $Q$  ( $\Delta Q$ ) from two long-term periods by equation (17) and  
 397 (20), and for comparison, we use the Budyko framework to estimate  $\Delta Q$ , which is considered as  
 398 the changes induced by  $P$ ,  $E_0$ , and parameter  $n$  (the calculation formulations can refer Xu et al.  
 399 (2014)). The estimated and observed runoff components variations exhibit high consistency  
 400 (Figure 9), with an  $R^2$  of 0.99 and RMSE of 1.6 mm of  $\Delta Q_s$  attribution and  $R^2$  of 0.88 and RMSE  
 401 of 18 mm of  $\Delta Q_b$  attribution, respectively. As for  $\Delta Q$ , both the MPS model and the Budyko  
 402 framework can attain satisfactory performance, while the MPS model has a higher  $R^2$  (0.91) than  
 403 the Budyko framework (0.89). Table 2 demonstrates that the MPS model can accurately quantify  
 404 changes in runoff components over two periods. Subsequently, we quantify the contribution of  
 405 precipitation and other factors (encoded by parameter  $W_p$  and  $V_p$ ) to  $\Delta Q_s$  and  $\Delta Q_b$ .



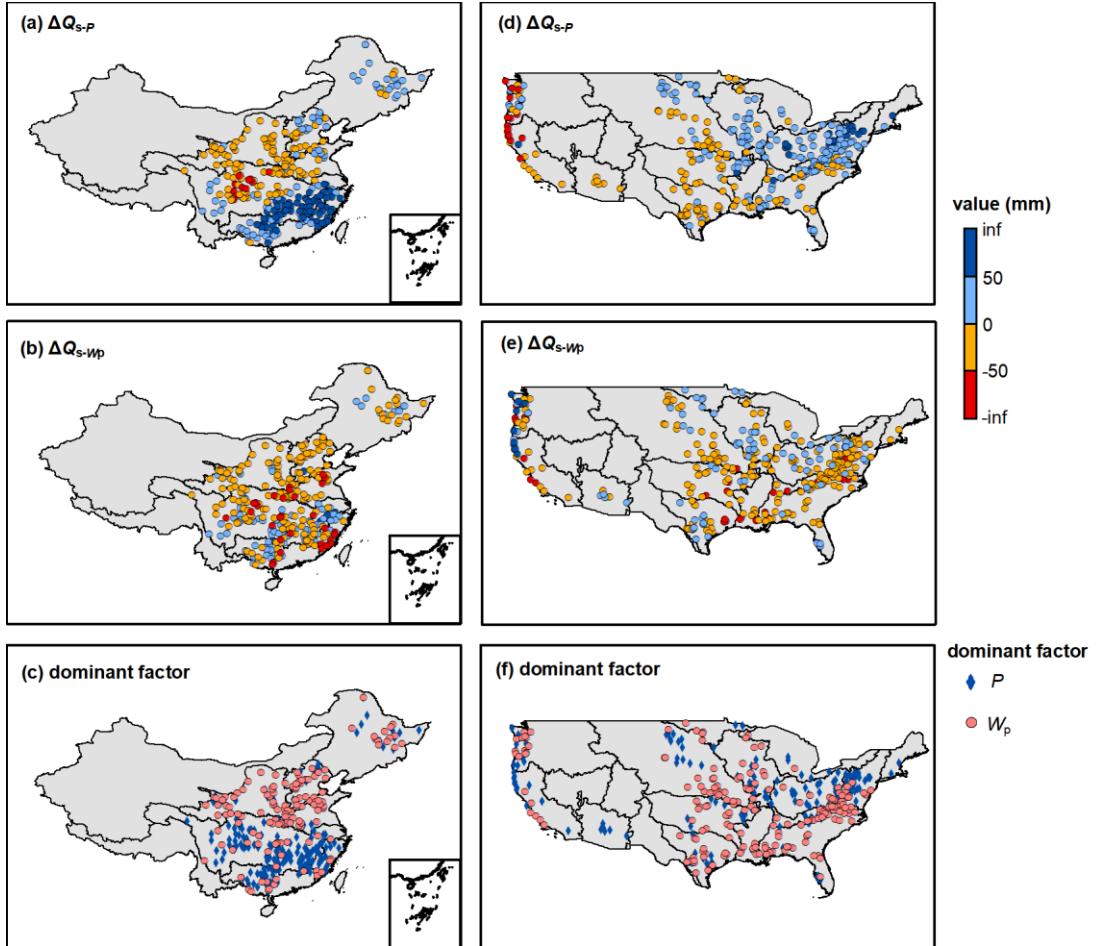
**Figure 9.** The observed and modelled (a) surface flow and (b) baseflow by the MPS model.

**Table 2.** The metrics of the attribution validation

Variables	$R^2$	RMSE (mm)
$\Delta Q_s$	0.99	1.6
$\Delta Q_b$	0.90	16
$\Delta Q$ (the MPS model)	0.91	42
$\Delta Q$ (the Budyko framework)	0.89	41

Figure 10 shows the  $\Delta Q_s$  induced by  $P$  ( $\Delta Q_{s-P}$ ) and other factors ( $\Delta Q_{s-W_p}$ ) along with the dominant factor in the catchments of China and the CONUS. From 1960-1990 to 1991-2000 in China, the multi-year variation in  $P$  has resulted in  $Q_s$  change ranging from -105 to 344 mm, mainly increasing  $Q_s$  in the catchments of the Songliao River Basin, the middle and lower Yangtze River Basin, the Southeast River Basin and Pearl River Basin, and decreasing  $Q_s$  in the catchments of the Yellow River Basin and the upper Yangtze River Basin (Figure 10(a)). The variations of other factors, such as land use/cover change and human activities, have resulted in  $Q_s$  change ranging from -186 to 124 mm, primarily decreases  $Q_s$  in 70% catchments (Figure 10(b)).  $P$  and other  $W_p$  are the dominant factor altering  $Q_s$  in southern and northern China, respectively (Figure 10(c)). From 1980-2000 to 2000-2014 in the CONUS, variation in  $P$  has resulted in  $Q_s$  change ranging from -469 to 149 mm, mainly increasing  $Q_s$  in the catchments of Interior Plains (except Great Plains), Coastal Plain, Interior highlands and Appalachian Plain, and decreasing  $Q_s$  in the catchments of the Great Plains and Pacific Mountains (the physiographic

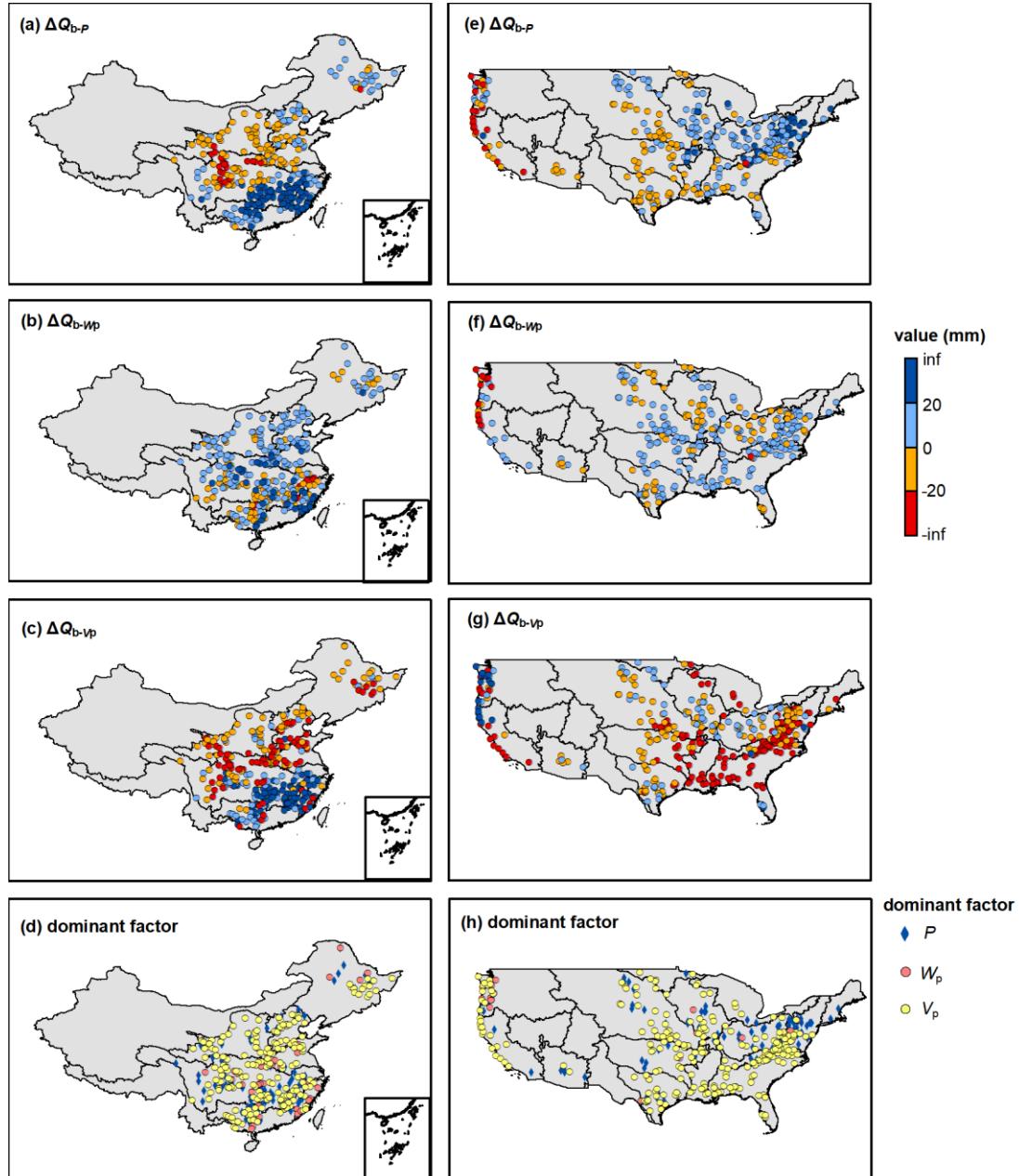
422 divisions are referred to Wu et al. (2021)) (Figure 10(d)). The variations of other factors have  
 423 resulted in  $Q_s$  change ranging from -230 to 467 mm, primarily decreases  $Q_s$  in 75% catchments  
 424 (Figure 10(e)). The catchments in the CONUS dominated by  $P$  and  $W_p$  account for 43% and 57%,  
 425 respectively (Figure 10(f)).



**Figure 10.** The surface flow change induced by precipitation and wetting potential ( $W_p$ ) along  
 with the dominant controlling factor.

429 Figure 11 shows the  $\Delta Q_b$  induced by  $P$  ( $\Delta Q_{b-P}$ ),  $W_p$  ( $\Delta Q_{b-Wp}$ ) and  $V_p$  ( $\Delta Q_{b-Vp}$ ) in the  
 430 catchments of China and the CONUS. The spatial pattern of the effect of  $P$  on  $Q_b$  is similar to that  
 431 of the  $Q_s$ , resulting in  $Q_b$  change from -38 to 79 mm in China (Figure 11(a)) and -129 to 92 mm in  
 432 the CONUS (Figure 11(e)), respectively. Catchment wetting potential has a positive effect on  $Q_b$   
 433 in 70% and 75% catchments of China and the CONUS, respectively (Figure 11(b)and (f)), mainly  
 434 in the northern China and the Interior Highlands, Coastal Plain and Appalachian Highlands of the  
 435 CONUS. Vaporization potential has a negative effect on  $Q_b$  in 56% and 77% catchments of China  
 436 and the CONUS, respectively, mainly in the upper Yangze River Basin and northern China and

437 the central and southeastern CONUS (Figure 11(c)and (g)). Although  $V_p$  is the dominant factor  
 438 controlling  $Q_b$  variation in most catchments in both China (62%) and the CONUS (71%) (Figure  
 439 11(d)and (h)), the contributions of the  $P$ ,  $W_p$  and  $V_p$  are not significantly discrepant in terms of  
 440 magnitude.



441  
 442 **Figure 11.** The baseflow change induced by precipitation, wetting potential ( $W_p$ ) and  
 443 vaporization potential ( $V_p$ ) along with the dominant controlling factor.

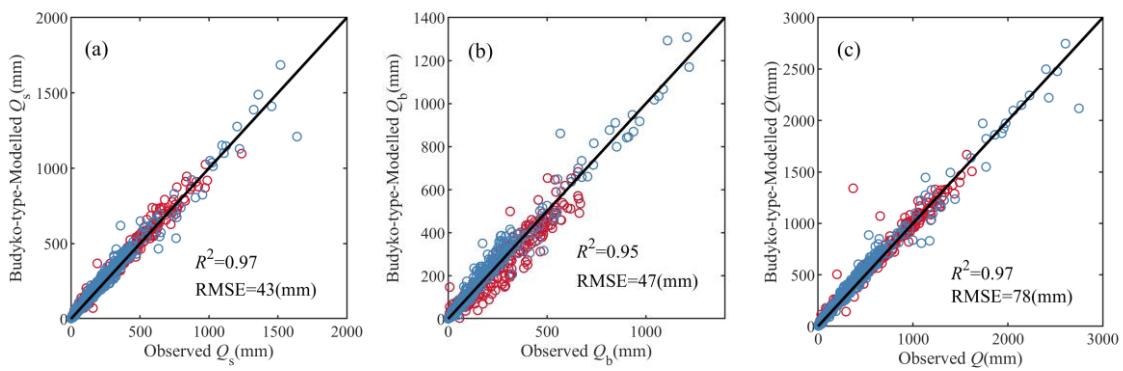
444 Overall, Figure 10 and 11 illustrate that the variation of  $Q_s$  is jointly controlled by  $P$  and other  
 445 factors, while the variation of  $Q_b$  is mainly influenced by  $V_p$ . This demonstrates that  $Q_s$  is closely  
 446 related to rainfall and soil storage capacity, while  $Q_b$  is more affected by catchment attributes,

447 atmospheric water and energy demand, etc. In regions where runoff components are reduced,  
 448 focus should be given to the risks of drought and river discontinuity; conversely, in areas  
 449 experiencing runoff components increase, there is a need to guard against the risk of flooding.

450 **5. Discussion**

451 **5.1. Superiorities of the MPS Model**

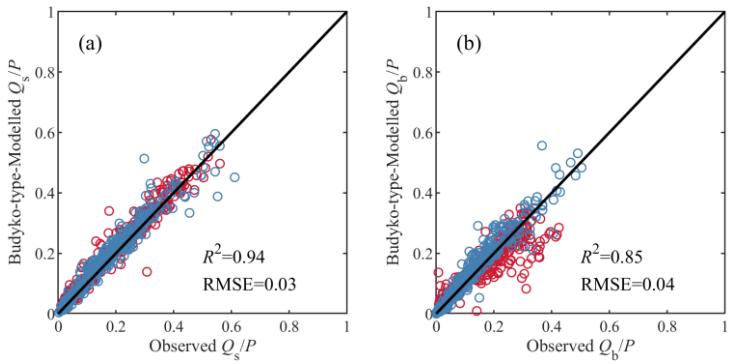
452 The researches about long-term runoff components quantification and attribution are currently  
 453 fragmented and region-specific (Beck et al., 2013; Gnann, 2021). This study has developed a  
 454 general formulation (the MPS model) through observational data analysis and theoretical  
 455 derivation based on the Ponce-Shetty model, unveiling the patterns of variability in different  
 456 runoff components at mean annual time scale. Compared to the commonly used Budyko-type  
 457 formulations, it can not only estimate mean annual  $Q$  and  $Q_b$ , but also can depict the variability of  
 458  $Q_s$ . Figure 12 shows the estimated mean annual runoff components by the Budyko-type  
 459 formulations (equations in the second and fifth rows of Table 1 in this paper). The Budyko-type  
 460 formulations also achieve good validation performance, with  $R^2$  greater than 0.95 and RMSE less  
 461 than 78 mm. Although the MPS model and the Budyko-type formulations are comparable in  
 462 terms of  $R^2$ , especially with almost equal simulation results of  $Q_s$ , the MPS model reduced the  
 463 RMSE values by 10 mm and 12 mm for estimating  $Q_b$ , respectively.



464  
 465 **Figure 12.** The observed and simulated mean annual (a) surface flow, (b) baseflow and (c) total  
 466 runoff by the Budyko-type formulations in China (red circles) and the CONUS (blue circles).

467 Figure 13 presents the estimation of  $SFC$  and  $BFC$  in validation periods using the Budyko-type  
 468 formulations. The two methods also show highly consistent estimation of  $SFC$ , with  $R^2$  of 0.94

469 and RMSE of 0.03. However, the Budyko-type formulations underestimate the *BFC* of most  
 470 catchments in China, while the MPS model greatly improves the simulation accuracy of *BFC*.



471

472 **Figure 13.** The observed and simulated (a) surface flow fraction ( $Q_s/P$ ) and (b) baseflow fraction  
 473 ( $Q_b/P$ ) by the MPS model in China (red circles) and the CONUS (blue circles).

474 In conclusion, the MPS model has comparable capability in simulating  $Q_s$  and *SFC* to that of  
 475 Budyko-type formulations. Moreover, it outperforms Budyko-type formulations in estimating  $Q_b$   
 476 and  $Q$ , and reveals superiority in estimating *BFC*. By characterizing runoff components as  
 477 functions of available water at corresponding stages with a composite parameter, the MPS model  
 478 is more concise in form and eliminates additional and complex parameter computations, thereby  
 479 facilitating broader application in large-sample investigations.

480 In addition to precisely quantifying runoff components and the allocation of precipitation, this  
 481 model has innovatively attributed the contributions of different factors on the changes of  $Q_s$  and  
 482  $Q_b$ . Our results show that the variation of  $Q_s$  is jointly controlled by  $P$  and other factors.  $P$  plays  
 483 an dominant role in the variation of  $Q_s$  in the catchments of the Yangtze River Basin, Southeast  
 484 Basin and Pearl River Basin of China and the west coast of the CONUS, where precipitation has  
 485 been reported to have undergone significant changes (Li et al., 2021; Mallakpour and Villarini,  
 486 2017; Massoud et al., 2020; Xu et al., 2022). This is possibly due to more extreme precipitation  
 487 events and summer rainfall in the middle-lower Yangtze River Basin (Ye et al., 2018) and an  
 488 increasing trend in the frequency of heavy precipitation over large areas of the CONUS  
 489 (Mallakpour and Villarini, 2017). Previous studies reported that the variation of  $Q$  in these  
 490 regions are dominated by  $P$  (He et al., 2022; Huang et al., 2016). Now it seems that  $P$  mainly  
 491 affects the first allocation stage ( $Q_s$ ) and consequently change total runoff. The variation of  $Q_b$  is  
 492 mainly influenced by  $V_p$ , indicating that we should pay more attention to the changes of

493 catchment attributes, atmospheric water and energy demand in most catchments when  
494 investigating  $Q_b$ .

495 Overall, this conceptual model extracted from observed rainfall-runoff data provides a concise,  
496 general and effective tool for predicting runoff components, and evaluating their responses to  
497 climate and environment under global change.

498 **5.2. Parameter Interpretation**

499 In the MPS model, each runoff component is associated with a parameter that can be  
500 interpreted as the upper limit of the remaining portion of available water after it has been  
501 partitioned into runoff at each stage. For instance, in the first stage, precipitation is allocated to  
502 surface flow and catchment wetting, with  $W_p$  representing the upper limit of catchment wetting,  
503 which describes the catchment's storage capacity related to soil, topography and so on (Cheng et  
504 al., 2022).  $W_p$  is influenced by soil properties and available storage capacity, determining the  
505 fraction of precipitation that rapidly becomes surface runoff versus what is stored. For the second  
506 stage, the available water comes from catchment wetting, which is then allocated to baseflow and  
507 vaporization. The parameter  $V_p$  is the upper limit of the fraction of wetting returned to the  
508 atmosphere as water vapor (Ponce and Shetty, 1995), and is likely responds to subsurface  
509 characteristics such as aquifer permeability and geological layering. For instance, in highly  
510 heterogeneous aquifers with well-developed preferential pathways (e.g., fractured rock or karst  
511 systems), water is rapidly drained toward the stream, leading to a higher efficiency of baseflow  
512 production and thus a lower  $V_p$  value (as less water is retained for evaporation). Conversely, in  
513 catchments with more homogeneous, porous media (e.g., sandy aquifers), water movement is  
514 slower and more diffuse, potentially allowing for a greater fraction of stored water to be  
515 evaporated, resulting in a higher  $V_p$ . For the total runoff, we consider precipitation as the available  
516 water competing with evapotranspiration, whose upper limit is represented by the parameter  $U_p$ .  
517 Similar to  $V_p$  in the second stage,  $U_p$  can be regarded as a sort of atmospheric water and energy  
518 limit (somewhat analogous to potential evapotranspiration) and emerges from the interaction of  
519 the available energy, vegetation and other catchment characteristics. To some extent, the MPS  
520 model links  $Q_s$  and  $Q_b$  with  $Q$  using  $P$  in the first trade-off and  $V_p$  in the second trade-off, so that  
521 the forms of different runoff components can be unified.

522     Additionally, we compared the distribution of the parameters in the MPS model with that in  
523     Gnann (Gnann et al., 2019) and Siva's work (Sivapalan et al., 2011), which did not omit the  
524     initial abstraction coefficients  $\lambda_s$  and  $\lambda_b$ . There is a very similar spatial pattern of  $W_p$  and  $V_p$  in  
525     the CONUS. Specifically, high  $W_p$  can be seen in the middle of the United States (Great Plains)  
526     and the east (southern parts of the Appalachians) (Figure 7(d)), and high  $V_p$  can be seen in the  
527     middle of the United States (Great Plains) and all southern regions (Figure 7(e)). This, to some  
528     extent, illustrates the rationality of the simplification of the original Ponce-Shetty model in  
529     describing the spatial variability of runoff components. According to Ponce and Shetty (1995)  
530     and Sivapalan et al. (2011), the products  $\lambda_s W_p$  and  $\lambda_b V_p$  are viewed as the initial abstraction to  
531     generate runoff. This definition is reasonable for short-term scales, such as event and annual  
532     scales. However, on the multi-annual scale, the catchment maintains a state of water balance and  
533     water losses can be disregarded (Han et al., 2020). Hence, simplifying  $\lambda$  to zero is rational to  
534     quantify and attribute runoff components and offer a new perspective on the long-term catchment  
535     water balance.

### 536     **5.3. Uncertainties and Future Improvements**

537     It is important to acknowledge several uncertainties in this study. First, the definition of  
538     “baseflow” itself introduces uncertainty. Although widely used as a collective term for delayed  
539     streamflow components, baseflow encompasses contributions from hydrologically distinct  
540     sources such as groundwater drainage, hyporehic exchange, snowmelt, and deeper subsurface  
541     leakage-each with distinct origins, timescales, and sensitivities to environmental factors. For  
542     instance, groundwater flow and deep leakage are strongly controlled by geological heterogeneity,  
543     including the distribution of rock types, porosity, permeability, faults, and fractures (Schiavo,  
544     2023). In contrast, snowmelt baseflow, on the other hand, is mainly driven by temperature  
545     variations within interannual to decadal climate cycles.

546     The definition of baseflow directly influences the selection of catchment areas. Guided by this  
547     macro-scale definition-viewing baseflow as the relatively stable portion of total runoff-we  
548     included large catchments in our analysis. While this inclusion may be a source of error, it does  
549     not affect the key finding that the MPS model effectively captures the variability of mean annual  
550     runoff components across catchments. A sensitivity analysis of the model's performance under

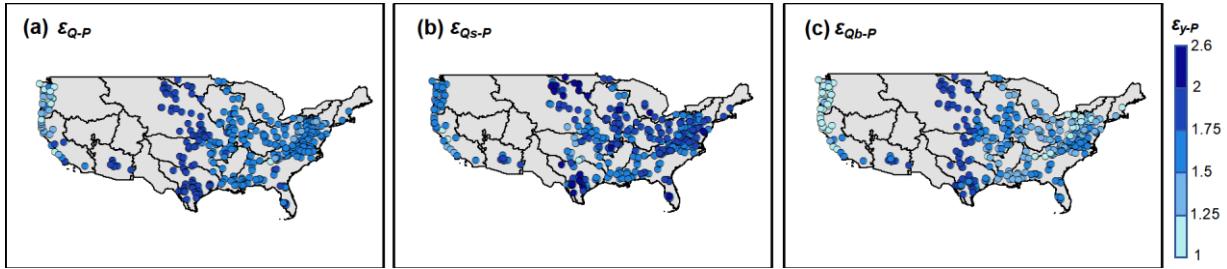
551 different area thresholds is provided in Appendix Table A1. Future studies could combine isotope  
552 tracing with hydrological modeling to better quantify the contributions of these different sources.

553 Second, methodological uncertainty arises from the digital filter method (i.e., the Lyne –  
554 Hollick algorithm) for baseflow separation. While practical and widely applied, this approach is  
555 deterministic and does not explicitly account for uncertainties related to aquifer heterogeneity,  
556 such as spatial variability in hydraulic conductivity, preferential flow paths, or geologic structures.  
557 Future work could adopt stochastic frameworks such as Monte Carlo simulation by generating  
558 multiple realistic realizations of aquifer heterogeneity to obtain more robust and probabilistic  
559 baseflow estimates (Schiavo, 2023). Additionally, our study did not take into account the spatial  
560 heterogeneity of groundwater flow, particularly its preferential pathways through fractures,  
561 macropores, or highly permeable sedimentary layers. Event-scale analyses indicate that  
562 stormflow volumes and hysteresis patterns covary with subsurface connectivity and its timing.  
563 For example, Zuecco et al. (2019) who used graph-theory metrics to quantify connectivity in  
564 headwater catchments and linked maximum connectivity to stormflow. While our study operates  
565 at mean-annual scales, these findings are consistent with our interpretation that geological  
566 heterogeneity and preferential pathways (fractures, karst, macropores) modulate the  $V_p$  dispersion  
567 and, in turn, the aggregate baseflow fraction. Future work could employ numerical models or  
568 distributed hydrological models that explicitly represent geological structures to better capture the  
569 effects of preferential flow paths at smaller scales.

570 The sensitivity of runoff to changes in climatic and environmental factors has always been  
571 highly anticipated. Schaake (1990) first introduced the concept of climate elasticity coefficients to  
572 quantify it, defined as the ratio of the relative change in mean annual runoff to the relative change  
573 in climatic factors. Various expressions have been widely applied in evaluating the hydrological  
574 response to multi-annual average climate change (Sun et al., 2014; Xu et al., 2014). The only  
575 climatic factor in the MPS model is  $P$ , so we primarily focuses on the elasticity of runoff  
576 components to  $P$  ( $\varepsilon$ ), which can be expressed as  $\varepsilon_{y-P} = \frac{\partial Q_y}{\partial P} / \frac{Q_y}{P}$ , quantifying the percentage of  
577 runoff components change caused by 1% change in  $P$ .

578 Figure 14 shows elasticities of  $Q$ ,  $Q_s$  and  $Q_b$  to  $P$  derived from the MPS model in the CONUS.  
579 We compare the elasticity distribution of the work conducted by Harman et al. (2011), who did

580 not omit the initial abstraction coefficients  $\lambda$ . In humid catchments with the aridity index of less  
 581 than 1 (such as the west coast and eastern regions of the CONUS), the results from both studies  
 582 are very close, with elasticity values from 1 to 2. However, the MPS model noticeably  
 583 underestimates the runoff sensitivity to  $P$  in semi-arid and arid catchments (such as the Great  
 584 Plains). This may be due to the error caused by the assumption that  $\lambda$  is a constant when deriving  
 585 the MPS model.



586  
 587 **Figure 14.** The elasticity of (a) total runoff, (b) surface flow and (c) baseflow to precipitation  
 588 derived the MPS model.

589 Additionally, the secondary rainfall processes, such as initial abstraction to generate runoff,  
 590 precipitation intensity and seasonality should be considered in these regions, which have been  
 591 proven to have a significant impact in attribution analysis (He et al., 2022; Ning et al., 2022;  
 592 Zhang, 2015). Moreover, the potential evapotranspiration ( $E_0$ ), which indicates the impact of  
 593 energy constraints (Huang et al., 2019; Wu et al., 2020), is quite significant in arid and semi-arid  
 594 catchments and should be taken into account.

595 In this paper, we interpret the parameters (i.e.,  $W_p$ ,  $V_p$  and  $U_p$ ) as a potential upper limit of each  
 596 partition stage competing with corresponding runoff components following the annual  
 597 Ponce-Shetty model. It is intriguing to discuss whether the connotation of the parameters has  
 598 changed from annual to mean annual time scale. On a long-term scale, the initial abstraction  
 599 coefficient (i.e.,  $\lambda_p$  and  $\lambda_W$ ) can be simplified as zero, indicating the loss for generating runoff is  
 600 negligible. However, to what extent the initial abstraction coefficient affect precipitation partition  
 601 at shorter time scales is still under-determined. The physical and theoretical interpretation of  
 602 parameters and their impacts at different time scales are temporarily outside the scope of this  
 603 study. However, it is valuable to further research in future work. In addition, the seasonality of  
 604 rainfall measures the concentration of precipitation within a year. The more concentrated the  
 605 precipitation, the more likely it is to generate surface runoff, resulting in greater intra-annual

606 fluctuations in the BFI and a lower annual BFI. In contrast, in catchments with evenly distributed  
607 precipitation, soil water and groundwater are replenished consistently and gradually, leading to  
608 relatively stable intra-annual BFI and a higher annual BFI.

609 The MPS model has only one parameter for controlling each runoff component, which is  
610 arguably simplified but dependent on calibration, and their physical meaning needs further  
611 explanation. We still need to explain the parameters in terms of regional patterns of climatic  
612 and/or catchment attributes, meaning that currently this model can only be applied to gauged  
613 catchments with runoff observations and challenging to transfer to ungauged basins. Cheng et al.  
614 (2022) proposed two machine learning methods to characterize the parameter of the Budyko  
615 framework and further employed them in estimating global runoff partition. Results show that  
616 parameters related to vegetation (such as root zone storage capacity, water use efficiency and  
617 vegetation coverage) and climate (such as precipitation depth and climate seasonality) are the  
618 primary controlling factors of the parameter. Similar work can be referred to (Chen and Ruan,  
619 2023). These investigations provide priori knowledge for quantitatively linking the parameters of  
620 the MPS model to climate forcing and catchment attributes in future work.

## 621 **6. Conclusion**

622 We developed a general formulation (the MPS model) to estimate mean annual runoff  
623 components as a function of available water with a synthetic parameter based on a two-stage  
624 partition theory, and validated it over 662 catchments across China and the CONUS with further  
625 attribution analysis. The concise MPS model provides more accurate runoff components  
626 estimation and innovative attribution, offering new insights to long-term water balance and giving  
627 additional superiorities toward making predictions of runoff variation under global change. The  
628 main conclusions are as follows:

629 (1) The investigated catchments fit well with the MPS model, with  $R^2$  of 0.86, 0.68 and 0.91 for  
630 fitting  $Q_s$ ,  $Q_b$  and  $Q$  in China and  $R^2$  of 0.81, 0.44 and 0.80 for fitting  $Q_s$ ,  $Q_b$  and  $Q$  in the CONUS,  
631 implying the MPS model can well reproduce the spatial variability of different runoff  
632 components.

633 (2) The MPS model effectively simulates multi-year runoff components with  $R^2$  exceeding 0.97,

634 and the proportion of runoff components relative to precipitation with  $R^2$  exceeding 0.94. The  
635 spatial distribution of the parameters across China and the CONUS is related to that of climate  
636 zoning.

637 (3) The MPS model has proved effective in quantifying the variations of runoff components  
638 induced by precipitation and environmental factors. The estimated and observed  $\Delta Q_s$ ,  $\Delta Q_b$  and  
639  $\Delta Q$  exhibit high consistency, with an  $R^2$  of 0.99 and RMSE of 1.6 mm of  $\Delta Q_s$  attribution,  $R^2$  of  
640 0.90 and RMSE of 16 mm of  $\Delta Q_b$  attribution and  $R^2$  of 0.91 and RMSE of 42 mm of  $\Delta Q$   
641 attribution, respectively. The variation of  $Q_s$  is jointly controlled by  $P$  and environmental factors,  
642 while the variation of  $Q_b$  is mainly influenced by  $V_p$  in most catchments.

643 In general, this study proposes a general formulation for effectively estimating and attributing  
644 the mean annual runoff, surface flow and baseflow. The structure is simple with few parameters  
645 and clear physical significance. Its reliability has been authenticated, providing new insights for  
646 analyzing watershed water resources in changing environments.

647

648 **Author Contribution**

649 Y.H: conceptualization; model development/theoretical derivation; investigation; calculation;  
650 formal analysis; visualization; writing original draft.

651 H.Y: conceptualization; model development/theoretical derivation; data curation; writing review  
652 & editing; supervision; funding acquisition.

653 C.L.: conceptualization; data analysis; visualization; writing review & editing.

654 **Competing interests**

655 The authors declare that they have no conflict of interest.

656

657 **Appendix**658 Table A1 The coefficient of determination ( $R^2$ ) and model parameters for the MPS curve fittings under different  
659 area thresholds for selecting catchments in China

Area thresholds (km <sup>2</sup> )	Number of catchments	$R^2$			Parameters (mm)		
		$Q_s$	$Q_b$	$Q$	$W_p$	$V_p$	$U_p$
2,000	67	0.85	0.62	0.89	3220	2794	1439
5,000	135	0.84	0.63	0.89	3004	2651	1356
10,000	180	0.84	0.69	0.90	3098	2614	1375
20,000	219	0.85	0.68	0.90	3138	2585	1376
80,000	257	0.85	0.69	0.90	3207	2487	1364
500,000	295	0.85	0.69	0.91	3278	2428	1362

660

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665 **Data and code Availability**

666 The CAMELS data set is available at <https://ral.ucar.edu/solutions/products/camels>. The  
667 hydro-meteorological data of the catchments across China can be obtained from the Zenodo  
668 repository via <https://zenodo.org/records/11058118> (Li et al., 2024).

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