



Root water uptake patterns are controlled by tree species 1

interactions and soil water variability 2

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Abstract 17

18 Throughfall is the largest source of water entering the soil in forests, and its spatial distribution depends 19 on several biotic and abiotic factors. It is well documented that the distribution of throughfall results in 20 reoccurring higher and lower water inputs at certain locations. However, the role of horizontal root water uptake patterns in understanding the effects of throughfall patterns on subsurface water dynamics remains 22 unresolved. Therefore, here we investigate root water uptake patterns by considering spatial patterns of throughfall and soil water patterns in addition to soil and neighboring tree characteristics. In a beechdominated mixed deciduous forest in a temperate climate, we conducted weekly intensive throughfall sampling at locations paired with soil moisture sensors during the 2019 growing season. We employed a linear mixed effects model to understand controlling factors for root water uptake patterns. Our results show that soil water patterns and interactions among neighbouring trees are the most significant factors regulating root water uptake patterns. Temporally stable throughfall patterns did not influence root water uptake patterns. Similarly, soil properties were unimportant for spatial patterns of root water uptake. We



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- 30 found that wetter locations (rarely associated with throughfall hotspots) promoted greater root water
- 31 uptake. Root water uptake in monitored soil layers also increased with neighbourhood species richness.
- 32 Ultimately our findings suggest that complementarity mechanisms within the forest stand, in addition to
- 33 soil water variability and availability, govern root water uptake patterns.

Key words: root water uptake, throughfall, soil water, spatial patterns, beech

1) Introduction

- 37 Vegetation intercepts and redirects precipitation into throughfall and stemflow, collectively referred to as
- 38 below-canopy precipitation. Moreover, throughfall is usually the largest component of below canopy
- 39 precipitation (Levia and Frost, 2006; Sadeghi et al., 2020). For instance, in temperate forests throughfall
- 40 can account for about 70% of above canopy precipitation (Levia and Frost, 2003; Sadeghi et al., 2020).
- 41 This makes it the primary source of soil moisture replenishment in vegetated areas.
- 42 Below-canopy precipitation is modified by several biotic and abiotic factors (Levia and Frost, 2006; Levia
- 43 et al., 2011), such as vegetation type and canopy architecture (Crockford and Richardson, 2000; Pypker
- 44 et al., 2011; Levia et al., 2017), forest structure (Rodrigues et al., 2022), meteorological elements such as
- wind speed (Staelens et al., 2008; Van Stan et al., 2011; Fan et al., 2015), precipitation intensity and event
- size (Dunkerley, 2014; Magliano et al., 2019; Zhang et al., 2016; Staelens et al., 2008). This implies that
- 47 it inherently varies across space and time. Furthermore, previous studies showed that the spatial
- distribution of throughfall persists repeatedly over time (Keim et al., 2005; Staelens et al., 2006; Guswa
- 49 and Spence, 2012; Carlyle-Moses et al., 2014; Metzger et al., 2017; Van Stan et al., 2020).
- 50 Throughfall patterns potentially translate their spatial variability into soil moisture (Raat et al., 2002;
- 51 Blume et al., 2009; Zimmermann et al., 2009; Zehe et al., 2010; Bachmair et al., 2012; Rosenbaum et al.,
- 52 2012; Zhang et al., 2016). A decade ago Coenders-Gerrits et al., (2013) proposed that throughfall patterns
- are translated into soil wetting dynamics with a model based on combined hillslope topographic and
- 54 throughfall data collected in a beech-dominated catchment. However, in this model, the effect of
- 55 throughfall patterns on soil moisture patterns rapidly ceased. Later, Metzger et al. (2017) empirically



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confirmed that throughfall patterns barely alter soil moisture in response to rainfall and the limited influence rapidly disappears. Recently, Zhu et al. (2021) observed that stable spatial patterns of throughfall were related to the spatial distribution of soil moisture. However, this relationship was restricted only to relatively wet soil locations and throughfall hotspots. They also showed that throughfall patterns had a weak influence on the temporal dynamics of soil water content compared to soil bulk density and litter layer properties. Previously proposed explanations for the weak and short-term influence of throughfall patterns on the soil moisture patterns include: soil properties (Metzger et al., 2017), preferential flow induced by dry antecedent soil conditions (Jost et al., 2004; Blume et al., 2009; Molina et al., 2019; Fischer et al., 2023), litter layer (Raat et al., 2002), and local root water uptake enhanced by throughfall hotspots (Bouten et al., 1992; Coenders-Gerrits et al., 2013). Based on a one-dimensional soil water model, Bouten et al. (1992) proposed that throughfall patterns alter and localize root water uptake and promote fast drainage via preferential flow paths. However, to the best of our knowledge, the feedback mechanism of throughfall patterns on root water uptake variation has not yet been investigated in the field. Therefore, it is unclear how water uptake patterns play a role in translating throughfall patterns into spatio-temporal variation of soil water and vice versa. Soil water distribution may shape root water uptake patterns even more than root networks (Kühnhammer et al., 2020). Soil properties control soil water redistribution (Grayson et al., 1997; Cosh et al., 2008; Jarecke et al., 2021) and water availability (Vereecken et al., 2007; Cai et al., 2018). Thus soil properties can influence root water uptake patterns (Nadezhdina et al., 2007; Kirchen et al., 2017). Moreover, variations in soil water content reflect water uptake by root systems (Hupet et al., 2002; Schume et al., 2004; Schwärzel et al., 2009; Guderle and Hildebrandt, 2015; Jackisch et al., 2020). On the flip side, root water uptake can amplify but mostly homogenize soil moisture distribution (Hupet and Vanclooster, 2005; Teuling and Troch, 2005; Ivanov et al., 2010; Baroni et al., 2013; Martínez García et al., 2014). Root networks can also regulate soil moisture distribution by transporting water from wetter places to drier locations, which has been observed in a variety of ecosystems (e.g., Emerman and Dawson, 1996;

Katul and Siqueira, 2010; Yu and D'Odorico, 2015; Priyadarshini et al., 2016; Hafner et al., 2017).





- 83 In addition, tree species richness affects the dynamics of root water uptake (e.g., Volkmann et al., 2016;
- 84 Spanner et al., 2022). Neighboring different tree species utilize different hydraulic strategies, such as
- extracting water from different soil depths (Silvertown et al., 2015; Guo et al., 2018; Brum et al., 2019).
- 86 However, soil water scarcity can initiate or enhance competition mechanisms for water among tree species
- 87 (González de Andrés et al., 2018; Vitali et al., 2018; Magh et al., 2020). Moreover, studies conducted in
- 88 temperate forest ecosystems demonstrate that the relationship between tree species richness and water
- 89 uptake mechanisms varies (Krämer and Hölscher, 2010; Kunert et al., 2012; Meißner et al., 2012;
- 90 Forrester, 2014; Lübbe et al., 2016).
- 91 Briefly, throughfall and soil water variability, soil properties, and root water uptake patterns form complex
- and intertwined interactions in the terrestrial hydrological cycle. It has not yet been shown empirically
- how root water uptake patterns are affected by throughfall and soil water variation in combination with
- soil properties and neighboring tree characteristics. Therefore, here we investigate the role of throughfall
- patterns and pose the following questions to guide the investigation:
 - i) How do throughfall patterns influence root water uptake patterns?
 - ii) How do soil water and its variation and soil properties control variation in root water uptake?
- What is the role of biotic factors, namely tree size, distance, number, and species richness, on root water uptake patterns?
- Here, we address these questions by employing linear mixed effects model based on weekly throughfall
- sampling at locations paired with intensive soil moisture measurements in a beech-dominated unmanaged
- 102 forest. In addition, we incorporate data on field capacity, bulk density, and neighboring tree
- 103 characteristics.

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2) Materials and Methods

2.1) Research Site and Field Sampling

2.1.1) Research Site

- The research site is located in the forested upper hill region of the Hainich low mountain range in
- Thuringia, Germany, as a part of the Hainich Critical Zone Exploratory (CZE) (Küsel et al., 2016). The



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109 altitude in the research site ranges from 362 m to 368 m a.s.l. Mean annual air temperature varies between 110 7.5 and 9.5 °C, and the mean annual precipitation ranges from less than 600 to 1000 mm in the CZE 111 (Küsel et al., 2016). 112 In the study area, thin-bedded alternations of limestones and marlstones of carbonate rock (Middle 113 Triassic) form the bedrock overlain by shallow Pleistocene loess layer with cambisols and luvisols as 114 dominant soil types (IUSS Working Group, 2006; Metzger et al., 2021). The median soil depth above the 115 weathered bedrock is 37 cm, with soil depths ranging from 15 cm to a maximum depth of 87 cm (Metzger 116 et al., 2017). 117 In 2019, the tree community in the research site consisted of 574 individuals of various ages (diameter at 118 breast height \geq 5cm). The dominant species is European beech (Fagus sylvatica L.), which makes up 70% 119 of the tree community, followed by sycamore maple (Acer pseudoplatanus L.) with 21 %, and European 120 ash (Fraxinus excelsior L.) with 4%. These dominant species are accompanied by Large-leaved linden 121 (Tilia platyphyllos Scop.), European hornbeam (Carpinus betulus L.), Norway maple (Acer platanoides

2.1.2) Soil moisture monitoring and soil properties

The forest site (1 ha) was equipped with a soil moisture monitoring network (SoilNet; Bogena et al., 2010) consisting of SMT100 frequency domain sensors (Treuebner GmbH, Neustadt, Germany). Metzger et al. (2017) first described the soil moisture monitoring setup. Briefly, the observation platform (Figure 1) was divided into 100 subplots (10 m × 10 m), and 49 subplots were equipped with soil moisture sensors at two random measuring points each, for a total of 98 locations. At each measuring point, sensors were placed at two different depths, 7.5 cm (top sensors) and 27.5 cm (bottom sensors). The soil moisture network is maintained through a regular bi-weekly routine to avoid potential failures such as depleted sensors batteries, hardware problems, etc.

Undisturbed soil samples were collected during the sensor installation in 2014 and 2015 to estimate bulk

L.), Scots elm (*Ulmus glabra* L.), and Wild service tree (*Sorbus torminalis* (L.) Crantz). The stand has a

total basal area of 40 m² ha⁻¹ and has been unmanaged since 1997 (Kohlhepp et al., 2017).

density and water content at field capacity. In addition, we collected additional disturbed soil samples (n

= 40) near sensor locations in 2019. Bulk density was determined from oven-dried (24h, 105°C) soil mass





136 weight and water content at field capacity by applying 60 hPa pressure to the saturated undisturbed sample 137 for 72 h. 138 Soil properties vary slightly from top to subsoil at the research site. While silty loam is the dominant soil texture in both layers, the clay content is higher in the subsoil (Metzger et al., 2021). The median 139 140 volumetric water content at the field capacity is 44% in the topsoil and 42% in the subsoil. Moreover, the water content at the field capacity varies from 27% to 60% and from 31% to 62% in the topsoil and 141 subsoil, respectively. The average bulk density (d_{bulk}) of the topsoil is 1.16 g cm⁻³, with a range of 0.73 to 142 1.5 g cm⁻³. In the subsoil, the average bulk density (d_{bulk}) is slightly higher at 1.37 g cm⁻³ but has a similar 143 range (0.7 - 1.6 g cm⁻³) (See supplement for details). 144



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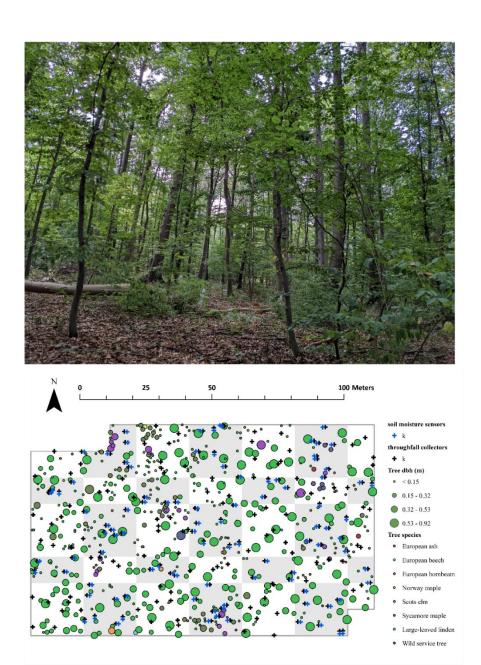


Figure 1 (above) The photo of the site. (below) the field monitoring setup of stratified randomly distributed throughfall collectors and soil moisture sensors together with the trees which are sized according to the diameter at breast height (dbh) and coloured according to the species. Throughfall collectors are paired with soil moisture sensors at 98 locations (n=182) in the grey shaded subplots. White coloured subplots are equipped with only throughfall collectors.





2.1.3) Gross precipitation and throughfall sampling

- 151 Five gross precipitation funnels were placed 1.5 m above ground level in an adjacent open grassland (ca.
- 152 250 m distance to the research site). As described in Metzger et al. (2017) and Demir et al. (2022), the
- precipitation funnels were made of a circular plastic funnel (12 cm in diameter) and sampling bottle (2 L
- in volume), and ping pong balls were placed in the funnel orifice to prevent evaporation losses.
- During the early growing season of 2019, we placed throughfall collectors in soil moisture monitoring
- subplots at 98 locations. We paired these throughfall collectors with the soil moisture sensors by placing
- them within 1 m of each other. The paired collectors were placed down-slope to avoid interference with
- soil moisture measurements. For the rest of the research site, in 51 other subplots, we adopted a separate
- independent stratified random design from Metzger et al. (2017). Briefly, we placed two throughfall
- 160 collectors in each subplot that was not equipped with soil moisture sensors. All throughfall collectors
- were placed roughly 37 cm above the ground.
- We conducted weekly manual measurement of throughfall and gross precipitation during the 2019
- growing season (April to August). We measured gross precipitation and throughfall on rainless days
- therefore, in some of the sampling weeks, the interval between field measurements ranged between six
- and eight days.

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- We used the paired throughfall collectors (n = 98) to identify the drivers of root water uptake patterns as
- we derived root water uptake values based on soil water content measurements (see below). However, we
- used all randomly placed throughfall collectors (n = 200) to describe the spatio-temporal variation of
- throughfall within the research site.

2.2) Estimation of potential evapotranspiration

- We calculated the daily potential evapotranspiration by applying the concept of thermodynamic limits of
- 172 convection (Kleidon and Renner, 2013):

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$$E_{pot} = \frac{1}{\lambda} \frac{s}{s+v} \frac{R_{sn}}{2}$$
 (1)

- Where R_{sn} is absorbed solar radiation (W m⁻²), λ is the latent heat of vaporization, γ is the psychrometric
- 175 constant, and s is the slope of the saturation vapor pressure curve.





- Here, we acquired solar radiation, air temperature, and precipitation data for the throughfall sampling period from a nearby weather station ("Reckenbuel") which is located approximately 1.4 km northeast of the research site and provides data in 10 minutes intervals. The site-specific albedo for the summer period was adopted from Otto et al. (2014).
- In addition, we used the precipitation data measured at the weather station to define rain events and dry periods, as described below.

2.3) Data analysis

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2.3.1) Quality control of soil water content data

- We systematically reviewed the six-minute soil water content data for quality control in two steps: 1) identification of problems (such as jumps to extremely low and high values, duplicated time stamps of different values, long discontinuities in the measurements, and lack of temporal variation in the time series despite rain events), 2) classification and removal of detected outliers and irregularities. We visually identified and removed unrealistic measurements such as extremely low (< 5 vol-%) and high values far beyond the field capacity (> 75 vol-%) and long plateaus of repeated values despite rain events. We also excluded the time series that exhibited long-term discontinuities that prevented us from calculating root water uptake. During the visual inspection, we eliminated values with duplicated time stamps that violated the actual temporal trend. Next, we scanned the data using the Hampel filter function of the 'pracma' R package (Borchers, 2021) with customized moving window length and Pearson's rule threshold value (Pearson, 1999) to flag possible outliers.
- Despite regular maintenance, many sensors failed to meet the quality criteria in the growing season
- 196 (March-August) in 2019. Only 56 sensor locations (out of 98) simultaneously provided high-quality data
- 197 from both top and bottom sensors with different time intervals throughout the season. Of these, only 34
- sensor locations provided data for the root water uptake estimation.

2.3.2) Soil water calculation

- We estimated soil water (S) at measurement locations for the monitored soil layer based on volumetric
- soil water content measured by top and bottom sensors.





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$$S_{i,d} = \sum z_t \theta_{i,d}^t + z_b \theta_{i,d}^b$$
 (2)

We similarly integrated the soil water at field capacity $(S_{FC,i})$

$$S_{FC,i} = \sum z_t \theta_{FC,i}^t + z_b \theta_{FC,i}^b$$
(3)

- where z_t is the depth of the soil column monitored by the top sensor and z_b is the depth of soil represented
- by the bottom sensor, and $\theta_{i,d}$ is a volumetric soil water content at location i on date d, and $\theta_{FC,i}$ the soil
- water content at the field capacity.
- We calculated bulk density at the sensors' locations for the monitored soil layer.

$$209 \quad \overline{d_{bulk,i}} = \frac{\sum z_t d_{bulk,i}^t + z_b d_{bulk,i}^b}{\sum z_t + z_b}$$

$$(4)$$

where $d_{\text{bulk},i}^{\text{t}}$ and $d_{\text{bulk},i}^{\text{b}}$ are the bulk density of the topsoil and subsoil, respectively, at location *i*.

211 **2.3.3**) **Descriptive Statistics**

- We calculated the coefficient of quartile variation (CQV) and the interquartile range to describe spatial
- variation of throughfall, volumetric soil water content, and root water uptake. Also, we estimated octile
- skewness (OS₈) of throughfall based on the first and seventh octile and standard deviation (SD) of the
- estimated daily root water uptake.

$$216 CQV = \frac{Q_3 - Q_1}{Q_3 + Q_1} (5)$$

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$$OS_8 = \frac{(Q_7 - median) - (median - Q_1)}{Q_7 - Q_1}$$
 (6)

- We characterized spatial patterns of daily root water uptake (E_t) by calculating the spatial deviation from
- the mean ($\delta E_{t,i,d}$, Equation 7) (Vachaud et al., 1985).

$$220 \qquad \delta E_{t i, d} = \frac{E_{t, i, d} - \overline{E_{t, d}}}{\overline{E_{t, d}}} \tag{7}$$

- where $E_{t, i,d}$ is daily root water uptake estimated at *i* sensor location on date *d* and $\overline{E}_{t,d}$ is spatial average
- of daily root water uptake on date d.
- 223 Similarly, we calculated the spatial deviation of soil water and throughfall to identify their spatial patterns.



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2.4) Root water uptake estimation

We estimated root water uptake using the multi-step, multi-layer regression method (MSML), which

derives evapotranspiration from diurnal differences in soil water content (Guderle and Hildebrandt, 2015;

Guderle et al., 2018). This approach does not require prior information on root structure but relies on high

temporal and spatial resolution data on multiple soil layers.

229 As described in Guderle and Hildebrandt (2015), the MSML derives root water uptake from distinct

differences in the day and night portions of soil moisture time series. The main assumption is that in the

absence of rainfall-driven rapid vertical soil water flow, evapotranspiration occurs only during the day,

while soil water flow occurs both during the day and at night. As a result, soil moisture time series reflect

a distinct day/night signal under dry weather conditions. This method has previously been applied to

estimate transpiration in both forest and grassland ecosystems (Guderle et al., 2018; Jackisch et al., 2020).

235 Therefore, we first excluded potential periods of fast vertical flow periods from the time series due to

previous rainfall events and identified periods for estimating daily root water uptake. We considered 8 h

buffer period to include canopy dripping and 48 h for the cessation of rainfall influence on soil water.

238 Thus, a total of 56 h was the time interval used to define the water uptake estimation period. The period

when the root water uptake is estimated is hereafter referred to as the dry period.

Next, we split each soil moisture time series into a day (transpiration active period) and a night branch,

as Guderle and Hildebrandt (2015) explained. We defined the transpiration period (starts 2 h after sunrise

and ends 2 h before sunset) based on local sunrise and sunset time. Sunrise and sunset times were obtained

from the R package 'suncal' (Thieurmel and Elmarhraoui, 2022). We fit linear models to each split branch

of the time series and derived the slopes. The difference between the slope of the day branch (m_{tot}) and

the average slope of the antecedent and preceding night $(\overline{m_{flow,l}})$ gives the rate of water uptake. Thus, we

estimated daily evapotranspiration at each soil water content location i (Equation 8, 9) by accounting for

soil layer thickness and slope difference-

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$$E_{t,msml,i}^{t,b} = (m_{tot,i}^{t,b} - \overline{m_{flow,i}^{t,b}}) d_{z,i}^{t,b}$$
 (8)

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$$E_{t,i} = \sum (E_{t,msml,i}^t + E_{t,msml,i}^b)$$
 (9)



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2.5) Linear Mixed Effects Model

- We employed a linear mixed effects model to investigate the driving factors for root water uptake patterns.
- A linear mixed effects model is a multivariate statistical tool. It describes the relationship between a
- dependent variable and explanatory variables (fixed effects) while controlling for dependencies in the
- data that may arise due to repeated sampling with certain designs (random effects).
- For the model, we used only paired throughfall and soil moisture measurement locations where both top
- and bottom sensors provided data within the dry periods. All considered potential controlling factors for
- 259 root water uptake patterns are listed in Table 1. These are daily spatial average soil water storage, the
- spatial deviation of soil water from the mean, soil water at field capacity and bulk density of the monitored
- soil layer, number of trees, and number of species within a 5 m radius of each soil moisture location, and
- 262 inverse distanced basal area (BA) within 5 m radius of each soil moisture location. Basal area was
- 263 calculated as follows:

$$264 BA_i = \frac{\sum_{R=1}^R W_R A_{tree}}{A} (10)$$

265 with
$$W_R = \frac{(x_i - x_R)^2}{\sum_R (x_i - x_R)^2}$$
 (11)

- where i is the soil moisture sensor located at x_i , R is the tree index located at x_R , and A_{tree} is the individual
- basal area of the corresponding tree, A is the area around the soil moisture sensor i with 5 m in radius.
- Even though our research plot is a beech-dominated forest, in some spots, two to four species were present
- within a 5 m radius of the soil moisture sensors.
- Moreover, we quantified the spatial variability of throughfall as the difference between the throughfall
- 271 measured at a given location and the spatial mean normalized by the spatial mean. Here we considered
- 272 this variable at a two-time scales: the week(s) prior to root water uptake estimation period, and the median
- of the entire measurement period. We also included interaction terms (Table 1) as fixed factors in the
- 274 model. Because of repeated observations at the measurement locations, soil moisture sensor points and
- dry periods, (i.e., the root water uptake estimated time interval), were considered as random effects.



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We conducted all analyses with the R statistical software (R Core Team, 2022) and used the *lmer* function in the 'lme4' package (Bates et al., 2015) for the model development. We visually checked the model assumptions using the 'check model' function of the 'performance' package (Lüdecke et al., 2021). In addition, we calculated both conditional and marginal R² of the model with the 'MuMIn' package (Barton, 2020). While the conditional R² includes the variance of the entire model, the marginal R² subsumes only the fixed effects (Bartoń, 2020). Before fitting the linear mixed effects model, we tested for co-linearity of the considered variables and scaled the data with a Z-transformation by using the 'scale' function in base R (R Core Team, 2022), which allowed us to evaluate the individual effect of fixed effects by comparing slopes and significance levels. We developed the optimal model by applying a systematic model selection procedure based on Akaike's Information Criterion (AIC) comparison in combination with the examination of the factors. Model selection began with the beyond-optimal model, which included all possible fixed and random effects. We stepwise evaluated each fixed effect based on its respective significance (p value comparision) by fitting the model the maximum likelihood (ML) to be able to compare AIC values (Zuur et al., 2009). In each step, starting with interaction terms, we identified the least significant effect and formulated a model without it. We compared the AIC values of the model before and after removing the effect, discarding it in case the AIC was unaffected or decreased. We followed the procedure with the next equally detected effect, and repeated it until only significant fixed effects remained, and the model with the lowest AIC (the optimal model) was obtained. As a final step, the best model was refitted with restricted maximum likelihood (REML) (Zuur et al., 2009).



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Table 1 List of fixed and random factors considered for estimating the root water uptake patterns through linear mixed effects model. Interaction is shown with 'x'.

Fixed Factors	
Single Factors	Interaction Factors
Spatial average of soil water storage in the monitored soil layer (\overline{S})	$\bar{S} \times S_{FC}$
Spatial deviation of soil water storage from the mean (δS)	$\delta S \times S_{FC}$
Field capacity of the monitored soil layer (S_{FC})	$\delta S \times BA$
Bulk density capacity of the monitored soil layer (d_{bulk})	$\bar{S} \times BA$
Spatial deviation of throughfall of events measured in sampling week previous to the corresponding dry period ($\delta P_{TF_{last~ev.}}$)	$\delta S \times n_{tree}$
The median of spatial deviation of throughfall measured within the whole sampling period $(\widetilde{\delta P}_{TF})$	$\bar{S} \times n_{\text{tree}}$
Number of trees (n _{tree})	$\delta P_{TF_{last ev.}} \times S_{FC}$
Basal area (BA)	$\delta P_{TF_{\text{temp. stable.}}} \times S_{FC}$
Number of species (n _{sp,tree})	$\delta P_{TF_{last ev.}} \times d_{bulk}$
	$\delta P_{TF_{\text{temp. stable.}}} \times d_{\text{bulk}}$
	$n_{sp,tree} \times WA_{int}$
Random factors	
Soil moisture sensor location	
Dry period	

3) Results

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3.1) Spatio-temporal distribution of throughfall and soil water content

In 12 out of the 16 sampling weeks, the weekly gross precipitation was more than half of the total potential evapotranspiration. Table 2 further shows the distribution of throughfall sampled in 2019 (April-August) at 200 collectors and the 98 collectors that were paired with soil moisture sensors. The weekly throughfall increased with the increase in rain events. Additionally, the coefficient of quartile variation (CQV) of

water content did not correlate with soil dryness.



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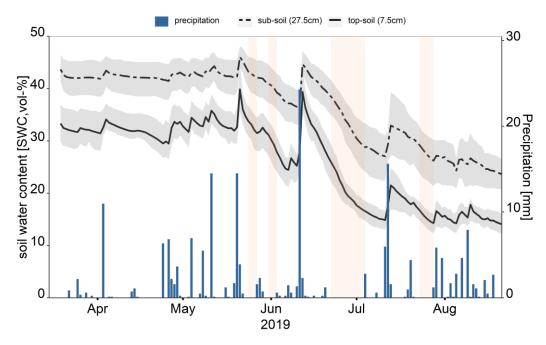
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throughfall was generally lower for larger cumulative weekly rains. On average, the paired collectors received similar amounts of throughfall to all collectors (Table 2). The CQV of data from the paired collectors ranged from 0.27 to 0.6, which is similar to the CQV of throughfall sampled at all collectors. The octile skew (OS₈) of paired and of all collectors was also similar. As the growing season progressed in 2019, the average soil water content decreased in both the topsoil and subsoil. In April and early May, the average volumetric soil water content in the top soil was above 30%, which dropped to below 10% by the end of August. In the subsoil, the volumetric soil water content similarly declined from above 40 % to below 20 % over the sampling period (Figure 2). On average, soil water changed from 52.5mm to 17.5 mm in the topsoil and from 80 mm to 40mm in the subsoil. We derived root water uptake for four periods (19 days) under different soil wetness conditions that captured the seasonal variation of soil water content, including late spring when the soil water content was higher, following re-wetting with late summer rains. As listed in Table 3 and shown in Figure 2, two periods were in late May and early June, and each lasted two days. The third period began in late June and lasted 11 days; the last was four days in late July. During these periods, the average soil water content declined from 33 to 15 % in the topsoil and from 43 to 27% in the subsoil. Table 3 additionally shows that within the dry periods, the coefficient of quartile variation (CQV) of soil water content was between 0.09 -0.14 and 0.08 to 0.16 in the topsoil and subsoil. During the dry periods, the spatial heterogeneity of soil water content in the subsoil increased systematically. In contrast, the spatial variation of topsoil soil







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Figure 2 Soil moisture temporal variation in top and subsoil together with the daily precipitation measured at the nearby Reckenbühl station (approximately 1.4 km to the Northeast). The solid and dashed lines are spatial mean of soil water content estimated based on top (7.5 cm) and bottom (27.5 cm) sensors, and grey shaded areas show first and third quartiles. The reddish shaded areas show defined dry periods within the throughfall sampling when root water uptake could be estimated.

Hydrology and Sarth System

evapotranspiration, spatial mean of throughfall based on all collectors $(\overline{P_{TF}})$, spatial mean of throughfall based paired collectors $(\overline{P_{TF}})$ in mm, interquartile range (IQR), coefficient of quartile variation (CQV) and octile skewness (OS₈) of both all and paired throughfall **Table 2** Cumulative potential evapotranspiration in mm (E_{pot,cum}), gross precipitation (Pg), the ratio of total precipitation to the potential collectors during the sampling week. The values are ordered according to the cumulated gross precipitation size.

Preparation of the Pere Pere Pere Pere Pere Pere Pere Pe	0.18 0.27 0.27 0.62 1.24 1.06 1.06 1.33 2.65	Pre Pre 0.25 0.46 0.24 0.16 0.08 0.23 0.12 -0.03 0.17 -0.52 0.17 -0.52 0.17 -0.52 0.17 -0.54 0.10 -0.47 0.15 -1.51 0.07 -0.75	0.34 0.98 1.72 2.57 3.63 4.48 5.17	P.Tr paired 0.16 0.53 0.53 0.53 1.50 0.88 0.97	P.Tr.paired 0.24 0.27 0.09 0.10 0.21 0.10	Prpaired 0.49 0.27 0.09 -0.08 -0.42 -0.63
0.35 0.97 1.72 2.58 3.77 4.61 5.19 7.81 9.21 11.19 16.60 18.44 24.22	0.18 0.44 0.27 0.62 1.24 1.06 1.06 1.33 2.65			0.16 0.53 0.33 0.53 1.50 0.88 0.97	0.24 0.27 0.09 0.10 0.21 0.10	0.49 0.27 0.09 -0.08 -0.42 -0.43
0.97 1.72 2.58 3.77 4.61 5.19 7.81 9.21 11.19 16.60 18.44 24.22	0.44 0.27 0.62 1.24 1.06 1.06 1.33 2.65			0.53 0.53 1.50 0.88 0.97	0.27 0.09 0.10 0.21 0.10	0.27 0.09 -0.08 -0.42 -0.63
2.58 2.58 3.77 4.61 5.19 7.81 9.21 11.19 16.60 18.44 24.22	0.27 0.62 1.24 1.06 1.06 2.25 1.33			0.33 0.53 1.50 0.88 0.97	0.09 0.10 0.21 0.10 0.10	0.09 -0.08 -0.42 -0.63 -0.44
	0.62 1.24 1.06 1.06 2.25 1.33 2.65			0.53 1.50 0.88 0.97	0.10 0.21 0.10 0.10	-0.08 -0.42 -0.63
	1.24 1.06 1.06 2.25 1.33 2.65			1.50 0.88 0.97	0.21 0.10 0.10	-0.42 -0.63 -0.44
	1.06 1.06 2.25 1.33 2.65		•	0.88 0.97	0.10	-0.63 -0.44
	1.06 2.25 1.33 2.65			0.97	0.10	-0.44
	2.25 1.33 2.65			300		
	1.33	•	`	7.70	0.15	-1.17
	2.65		-	1.99	0.11	-1.05
	27 0			2.98	0.13	-1.13
	7.02	•		2.65	80.0	-1.17
	3.09	•		2.65	0.07	-1.23
	3.54	•		3.54	0.07	-2.59
	3.54			3.54	80.0	-2.11
	3.54		` .	3.54	90.0	-3.46
	8.51			7.29	0.06	-6.52





Table 3 The spatial average of daily volumetric soil water content ($\overline{\theta}_{top-soil}$, vol-%) in topsoil (0-17.5 cm), and ($\overline{\theta}_{subsoil}$, vol-%) in subsoil (17.5 – 37.5 cm) during the defined dry periods. The inter quartile range (IQR), and coefficient of quartile variation (CQV) of daily volumetric soil water content in both layers during the dry periods.

Date	$\overline{\theta_{\text{top-soil}}}$ (vol-%)	$\begin{array}{c} IQR \theta_{top\text{-soil}} \\ (\text{vol-}\%) \end{array}$	$\begin{array}{c} CQV \; \theta_{\text{top-soil}} \\ (\text{vol-}\%) \end{array}$	$\overline{m{ heta_{sub-soil}}}$ (vol-%)	$\begin{array}{c} IQR \theta_{subsoil} \\ (vol\text{-}\%) \end{array}$	$\begin{array}{c} CQV \;\; \theta_{subsoil} \\ (vol\text{-}\%) \end{array}$	Dry Period
25 -05-2019	33.17	5.72	0.09	42.82	6.72	0.08	1
26-05-2019	32.12	6.62	0.10	42.46	6.67	0.08	1
01-06-2019	30.23	6.87	0.12	40.61	6.9	0.09	2
02-06-2019	29.22	7.23	0.13	40.11	6.85	0.09	2
23-06-2019	25.01	6.69	0.14	37.80	6.38	0.08	3
24-06-2019	24.04	6.45	0.14	36.94	6.22	0.08	3
25-06-2019	22.52	5.43	0.12	36.13	6.54	0.09	3
26-06-2019	21.48	5.07	0.12	35.24	6.71	0.10	3
27-06-2019	20.20	4.25	0.11	33.98	7.75	0.12	3
28-06-2019	19.45	3.85	0.10	33.31	8.08	0.12	3
29-06-2019	18.98	3.83	0.10	32.36	8.05	0.12	3
30-06-2019	18.44	3.52	0.09	31.37	8.15	0.13	3
01-07-2019	17.67	3.62	0.10	30.45	8.18	0.13	3
02-07-2019	17.29	4.18	0.12	29.84	8.87	0.15	3
03-07-2019	16.89	3.72	0.11	29.26	8.98	0.15	3
24-07-2019	16.15	3.48	0.11	28.56	8.7	0.16	4
25-07-2019	15.51	3.47	0.11	27.85	8.67	0.16	4
26-07-2019	14.98	3.57	0.12	27.21	8.49	0.16	4
27-07-2019	14.57	3.65	0.13	26.65	8.63	0.16	4

3.2) Soil water storage, potential evapotranspiration, and root water uptake

The integrated field capacity of the monitored soil depth was 160 mm on average at the research site. Table 4 shows that soil water storage was much lower than the field capacity during the dry periods, and the mean soil water storage dropped below 42 mm in late July. In addition, Table 4 demonstrates that the average root water uptake (\bar{E}_t) ranged from 0.94 mm d⁻¹ to 3 mm d⁻¹ while potential evapotranspiration (E_{pot}) ranged from 1.75 mm d⁻¹ to 3.12 mm d⁻¹. The discrepancy between average root water uptake and the potential evapotranspiration increased as soil water storage assessed by the soil sensors progressively decreased, especially during the longest dry period (Table 4). Root water uptake showed greater spatial variation than water input and soil wetness. The coefficient of quartile variation (CQV) of root water uptake ranged from 0.15 to 0.28, which was higher than the CQV of throughfall and volumetric soil water content in both soil layers.





Table 4 The daily average air temperature (T_{air} , °C), potential evapotranspiration (E_{pot} , mm), mean soil water storage (\overline{S} , mm) in monitored soil layer (0 - 37.5 cm), and spatial mean of daily root water uptake (\overline{E}_t , mm) based on all soil moisture sensors, and the ratio of the root water uptake to the potential evapotranspiration together with and standard deviation (SD) and coefficient of quartile variation (CQV) of the daily root water uptake during the defined dry periods

Date	T_{air} (°C)	E _{pot} (mm)	<u>\$</u> (mm)	$\overline{E_t}$ (mm)	$\overline{E_t}/E_{ m pot}$	$SD \overline{E_t}$	$\operatorname{CQV} \overline{E_t}$	Dry Period
25-05-2019	12.74	1.80	71.94	1.09	60.56	0.38	0.28	1
26-05-2019	14.43	1.90	70.57	1.30	68.42	0.48	0.25	1
01-06-2019	18.42	2.59	67.16	2.26	87.26	0.98	0.27	2
02-06-2019	21.38	2.77	65.79	2.50	90.25	1.12	0.18	2
23-06-2019	19.45	2.79	59.81	2.83	101.43	0.90	0.19	3
24-06-2019	20.22	2.82	58.16	2.62	92.91	0.76	0.17	3
25-06-2019	22.52	2.89	55.96	2.67	92.39	0.78	0.16	3
26-06-2019	25.73	2.96	54.13	3.00	101.35	0.88	0.15	3
27-06-2019	18.83	2.75	51.91	2.28	82.91	0.55	0.16	3
28-06-2019	16.07	2.58	50.55	1.53	59.30	0.40	0.20	3
29-06-2019	19.59	2.85	49.55	2.11	74.04	0.60	0.20	3
30-06-2019	25.54	3.12	48.26	2.57	82.37	0.86	0.18	3
01-07-2019	20.63	2.30	46.69	1.59	69.13	0.53	0.18	3
02-07-2019	14.88	1.75	45.81	1.08	61.71	0.42	0.24	3
03-07-2019	13.77	1.91	44.95	0.94	49.21	0.30	0.23	3
24-07-2019	24.39	2.76	43.61	1.88	68.12	0.64	0.19	4
25-07-2019	25.33	2.82	42.31	1.77	62.77	0.60	0.24	4
2019-07-26	23.27	2.64	41.18	1.40	53.03	0.55	0.18	4
2019-07-27	21.29	2.68	40.23	1.21	45.15	0.47	0.19	4

3.3) Soil water, throughfall, and root water uptake patterns

At soil moisture measurement points where daily root water uptake was determined (n = 34), we calculated the spatial deviation from the median of throughfall, soil water storage, and root water uptake to illustrate the spatial patterns. Figure 3 separately shows that some locations received repeatedly less (or more) throughfall than average ($\delta P_{TF} < 0$) throughout the sampling season. Similarly, some locations, either stored less water in the soil, i.e., were drier ($\delta S < 0$), and some places had lower root water uptake (δE_t) or higher than average water uptake throughout the sampling period. However, these locations were not related to each other. In fact, Figure 3 demonstrates that neither throughfall nor soil water patterns are directly correlated with the root water uptake patterns. For example, the locations with higher water uptake were not coupled with elevated throughfall input (locations colored dark) or higher soil water storage. In addition, soil water storage patterns were not correlated with throughfall patterns.

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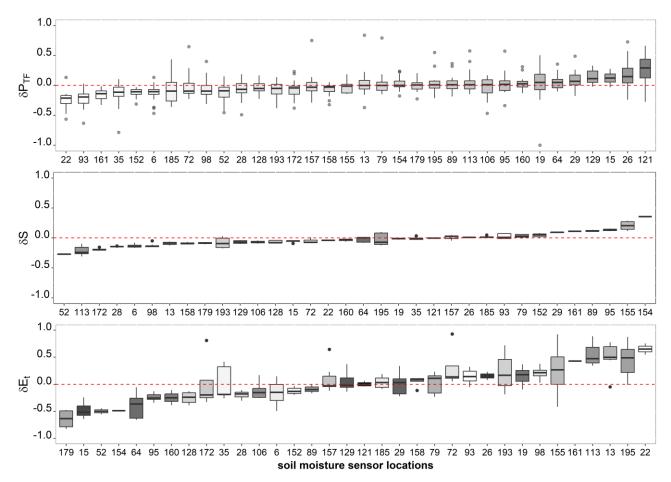


Figure 3 Temporal stability of throughfall patterns which is estimated by the spatial deviation from the mean (δP_{TF}) throughout the sampling period in 2019 (April-August), soil water (δS) and root water uptake (δE_t) based on the spatial deviation from the mean during the defined dry periods. Soil moisture sensor locations colored according to throughfall input.

3.5) Fixed factors regulating root water uptake patterns

We used a linear mixed effects model to disentangle the effects of throughfall, soil water, soil properties, and the neighbouring tree characteristics on root water uptake patterns. The fixed and random effects contributed almost equally to the model. The R^2 of the model was 0.77, and the contribution of the fixed effect to the R^2 was 0.39 (See the supplement for more details on the optimal model).

Figure 4 shows only the significant fixed effects for root water uptake patterns. Spatial deviation of soil water from the mean (i.e., soil water patterns) was the only single and the most significant factor positively related to the spatial deviation of root water uptake. Thus, water uptake was elevated at locations where the most water was retained in the soil at the given time, i.e., greater soil water storage.



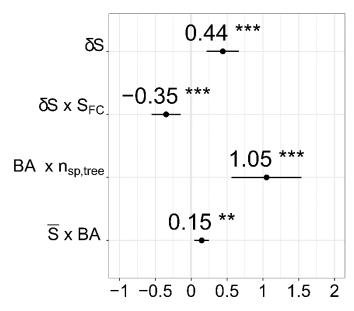


Figure 4 The significant fixed factors of the best model to estimate root water uptake patterns (δE_t). Values on the x-axis indicate the slope of the relations. All variables were scaled by Z-transformation. Interaction is shown with 'x'. Here δS is the spatial deviation of soil water, S_{FC} is the field capacity, $n_{sp,tree}$ is the number of species, BA is the basal area, and $\overline{\textbf{S}}$ is soil water storage. Significance codes are *** $\cong 0$, ** $\cong 0.001$. (the details on the model can be found in the supplement)

Field capacity by itself was not a significant factor affecting local root water uptake. However, it strongly

influenced how local soil water-controlled root water uptake as a part of the significant interaction term. Figure 5a illustrates how to root water uptake was more dependent on local soil water when field capacity was low (i.e., higher macroporosity). In contrast, soil bulk density and therefore total porosity was not part of the final model.

Although the spatial average of soil water storage, e.g., the state of wetness, was not an important factor for local root water uptake by itself, it moderated the impact of basal area (BA) on the spatial distribution of water uptake. We found that as the plot dries, uptake shifts from places with higher to places with lower basal area (Figure 5b). Furthermore, the statistical model revealed that water uptake increased with the higher basal area at locations where multiple species co-existed (Figure 5c). However, the number of species and the basal area were individually not significant fixed effects. Lastly, throughfall patterns were not significant predictors of local root water uptake. Only the median of the spatial deviation of





throughfall, which represents temporally stable patterns within the sampling period $(\widetilde{\delta P}_{TF})$, marginally improved the final model.

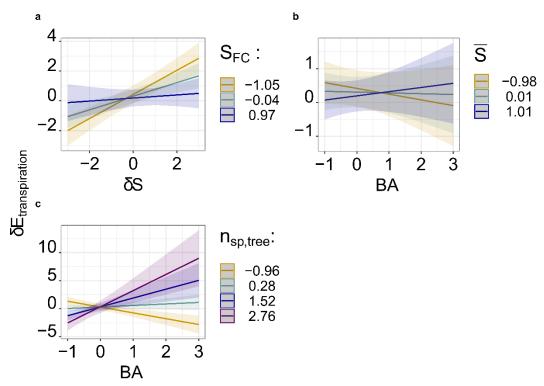


Figure 5 Visualisation of the significant relations shown in Figure 4, representing the significant drivers of root water uptake patterns during the defined dry periods. Relation to (a) interactive relation of the spatial deviation of soil water storage and field capacity (S_{FC}), (b) the interactive relation of basal area (BA) and the spatial average of soil water storage (\overline{S}), (c) the interactive relation of number of species ($n_{sp,tree}$) and basal area (BA).

4) Discussion

4.1) Spatial variation in throughfall does not affect root water uptake patterns

We adequately captured the spatial distribution and temporal stability of throughfall at locations where local root water uptake was derived. Consistent with previous observations in temperate forests (e.g., Whelan and Anderson, 1996; Staelens et al., 2006; Metzger et al., 2017), the amount of weekly rainfall significantly altered the spatial distribution of throughfall such that more rainfall, and thus more throughfall, resulted in less spatial variability. Previous studies repeatedly showed that throughfall patterns exhibit temporal stability in forest ecosystems (e.g., Keim et al., 2005; Staelens et al., 2006;





411 Wullaert et al., 2009; Rodrigues et al., 2022). At the same research site, using event-based sampling, 412 Metzger et al., (2017) and Fischer et al., (2023) demonstrated that throughfall patterns persist over time, 413 which was not different in our weekly sampling in 2019. With canopy cover being the key driver of 414 throughfall (Fischer et al., 2023), it is not surprising that weekly cumulative events resulted in a localized 415 high and low throughfall input. 416 Contrary to expectations (Bouten et al., 1992; Guswa and Spence, 2012; Coenders-Gerrits et al., 2013; 417 Fischer et al., 2023), our results showed that throughfall hotspots do not increase or facilitate greater root 418 water uptake. In addition, the linear mixed effects model results confirmed that throughfall patterns do 419 not drive the variation in root water uptake. We attributed the absence of this to two reasons: (1) decoupled 420 soil water and throughfall patterns, (2) non-water limited conditions. 421 Regarding (1), we confirmed that the temporally stable throughfall patterns do not correspond to the post-422 event soil water and root water uptake patterns. We paired the measurements of throughfall and soil water 423 content measurements – and thus the estimates of root water uptake- within a distance of 1 m. The spatial 424 correlation length of soil water content and throughfall is on the order of 6-10 m in natural temperate 425 forests (Keim et al., 2005; Gerrits et al., 2010; Zehe et al., 2010). In the same study site with the spatially extended throughfall sampling, Fischer et al., (2023) found that the throughfall correlation length 426 427 increased with decreasing event size, varying from 6.2 m to 9.5 m depending on the size of the rain events. 428 Thus, the paired sampling design in our study likely provided co-located throughfall and soil moisture 429 measurements. Nevertheless, only locations that stored more water than average rarely corresponded with the elevated throughfall input without a significant correlation. Hence, variation in soil water storage was 430 not related to throughfall patterns despite temporally persistent local high and low throughfall inputs. 431 432 On the one hand, some studies, mostly conducted in the arid regions and coniferous forests, reported that 433 soil wetting patterns were not or only partly linked to throughfall variation, despite recurrent throughfall 434 patterns (Raat et al., 2002; Shachnovich et al., 2008; Zhu et al., 2021). Forest floor thickness, horizontal 435 water flow, and soil properties were suggested as reasons for the decoupled patterns. On the other hand, 436 some modeling and field studies conducted in temperate deciduous forests found that throughfall patterns 437 influenced soil moisture response rather than soil water storage variability (Coenders-Gerrits et al., 2013;

Metzger et al., 2017; Fischer et al., 2023). In those studies, possible reasons were attributed to local



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processes such as preferential flow due to soil water repellency, the soil pore structure, or elevated root 440 water uptake. Our results support that it is not root water uptake but preferential flow paths that likely decouples the throughfall and soil water patterns. In fact, Fischer et al., (2023) using independent throughfall and soil water content sampling designs, demonstrated that the signature of throughfall patterns dissipated in the post-event soil water variation. However, they detected the stronger influence 444 of throughfall patterns in the soil moisture response to rainfall in the 2015 and 2016 growing seasons. The temporal variation in soil water content in the 2019 growing season was similar to the seasonal decline in soil water content in 2015 (Metzger et al., 2017). Dry soil conditions can lead to rapid drainage due to 446 reduced water holding capability (Jost et al., 2004; Blume et al., 2009; Wiekenkamp et al., 2016; Demand et al., 2019; Molina et al., 2019) regardless of throughfall amount and its variation. Therefore, our findings 448 449 support that the localized throughfall input likely enhances preferential flow because of low soil retention (Fischer et al., 2023) rather than local root water uptake. 450 As a result, the fast flow processes likely dominate how water is stored and transported at our site, erasing 452 the throughfall distribution signature in soil water and root water uptake patterns. Our results also support that the spatial variation of throughfall affects drainage and subsurface flow (Keim et al., 2006; Blume et 454 al., 2009; Guswa and Spence, 2012), and root activity does not alter canopy-attributed heterogeneity in 455 drainage pathways (Guswa, 2012). 456 The second reason (2) is related to water-limitation conditions. In central Europe, 2019 was the second consecutive extremely dry summer (Boergens et al., 2020), which damaged beech forests (Obladen et al., 2021). On average, however, the potential evapotranspiration demand was met at the study site despite the low soil water storage. The ratio of root water uptake to potential evapotranspiration was mostly above 459 460 65%, which is within the expected range even in the absence of shallow groundwater storage (Nie et al., 2021). Hence, local biotic and abiotic factors determined the spatial variation of root water uptake during growing season rather than throughfall patterns. However, the discrepancy between daily potential 462 evapotranspiration and root water uptake only increased as the soil in the sampled layers dried out, possibly due to a potential shift in the water uptake depth (see below).



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4. 2) Relative and average soil wetness shapes root water uptake patterns

We found that spatial variation in soil water storage strongly regulates local water uptake such that wetter

467 locations enhance root water uptake. This finding is in line with expectations as transpiration rate relies on soil water availability and distribution (Couvreur et al., 2014; Klein et al., 2014; Hildebrandt et al., 468 469 2016). Here, our results provide further support that root water uptake is likely to reduce the spatial 470 variability in soil water storage as has been previously suggested (Hopmans and Bristow, 2002; Ivanov 471 et al., 2010; Neumann and Cardon, 2012). 472 For a given meteorological condition, root-water uptake at a particular location is a function of water 473 transport resistance between root and soil in addition to the soil-water potential (Cardon and Letey, 1992; 474 Shani and Dudley, 1996; Lhomme, 1998). Both characteristics depend on local soil texture and soil 475 moisture, and the latter in turn is affected by the local rate of uptake. Although bulk density is attributed 476 to porous space and eventually water retention (Zacharias and Wessolek, 2007; Looy et al., 2017), we 477 surprisingly found that the bulk density of the monitored soil layer did not affect local water uptake. In 478 contrast, the combination of higher field capacity and low soil water probably hindered the local water 479 uptake due to lower soil water retention. Differences in local soil properties regulate matric potential at a 480 certain soil wetness. Thus, our result indicates that wetter locations may not always correspond to the 481 same degree of matric potential and ease root water uptake due to the local field capacity. However, our 482 findings suggest that solely soil properties were less important than other tested variables despite their 483 control on the spatial distribution of soil moisture (Vereecken et al., 2022) and water accessibility for 484 transpiration (Vereecken et al., 2007; Cai et al., 2018). 485 In addition, the spatial mean of soil water affected root water uptake patterns, yet the effect depended on 486 the basal area, i.e., the size of neighboring trees. We found that as the study site dries out, local water 487 uptake increased in locations with smaller basal areas. Conversely, wetter site conditions facilitate greater 488 water uptake at locations with higher basal areas, i.e., dense clusters of or large trees. We interpret this as 489 a sign that larger trees are likely to shift their water uptake to deeper soil layers to meet transpiration 490 demands, beyond the monitored soil depth (37 cm), as follows: Higher basal area likely increases 491 transpiration demand and enhances water uptake as long as water is available. At the same time, locations 492 with higher basal area exhaust the water storage faster but are able to shift uptake to deeper layers where



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493 soil water content is not measured in our monitoring setup. Beech trees have extensive root systems at shallower depths similar to other temperate tree species, such as European ash and sycamore maple 494 495 (Kreuzwieser and Gessler, 2010; Brinkmann et al., 2019). However, in response to declining soil water 496 content in the topsoil, temperate tree species can tap water from the deeper soil layers (Brinkmann et al., 497 2019; Agee et al., 2021; Seeger and Weiler, 2021) despite their shallow root system (Leuschner, 2020). 498 Recently, Agee et al. (2021) used a three-dimensional water uptake model based on observations in 499 temperate mixed-deciduous forest to show that water uptake is shifted to the deeper soil layers as soil 500 moisture depletes, which is consistent with the field observations. Also, Krämer and Hölscher (2010) 501 observed in beech and mixed deciduous stands that roots can extract water at depths down to 70 cm soil 502 depth. Similarly, to our site, theirs had a shallow soil layer underlain by weathered limestone.

In addition to the basal area, we included the number of species and number of tree individuals in the

4.3) Tree species richness regulates root water uptake patterns

505 linear mixed effects analysis to explore further the biotic drivers of root water uptake patterns. While the 506 number of trees was unimportant, the number of species and the basal area, showed a significant interaction effect on the local water uptake. The result indicates that an increase in species richness leads 507 to greater root water uptake, depending on the size and/or density of the neighboring trees: Higher basal 508 509 area, combined with more species, elevates water uptake. In other words, the interactions among neighboring tree species strongly determine root water uptake patterns, and at the same basal area more 510 511 water can be taken up in a diverse compared to a less diverse neighborhood. 512 In temperate forests, transpiration has been observed to change with tree species richness at the stand level (Krämer and Hölscher, 2010; Gebauer et al., 2012; Kunert et al., 2012; Meißner et al., 2012; 513 Forrester, 2014). Although some studies indicate a positive relationships between tree diversity and water 514 515 uptake rate (Forrester et al., 2010; Krämer and Hölscher, 2010; Kunert et al., 2012), tree species diversity 516 is not always positively related to water uptake. While Krämer and Hölscher (2010) observed a positive 517 correlation between water uptake and species richness of the plots in the upper soil layers during soil drying in 2006 at the same study site, Meißner et al. (2012) found no relationship between tree diversity 518 519 and root water uptake in 2009. They attributed this finding to wetter soil conditions. In contrast, Lübbe et



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al. (2016) observed a weak effect of diversity on transpiration in wetter soil conditions but not in drier conditions compared to previous studies (e.g., Pretzsch et al., 2013; del Río et al., 2014). Shortage of water can inflate competition mechanisms for water among tree species (González de Andrés et al., 2018; Vitali et al., 2018; Magh et al., 2020). Our results can be used to show that competition between neighboring tree species increases water uptake capacity at more diverse spots (Wambsganss et al., 2021). In addition, different co-existing tree species can facilitate resource uptake or reduce competition, depending on the temporal and spatial availability of the sources, which is often defined as complementarity (Forrester and Bauhus, 2016). As reviewed and listed by Silvertown et al. (2015), several studies suggest that co-existing tree species reduce competition for subsurface water sources by adopting different vertical root water uptake strategies, referred to as hydrological niche partitioning. In addition, trees can transport water from moist to dry parts of the soil layers through their roots (Neumann and Cardon, 2012). The mechanism is called hydraulic redistribution or hydraulic lift, which can provide water availability to the shallow roots in drier layers (Burgess et al., 1998; Jonard et al., 2011; Hafner et al., 2017; Lee et al., 2018; Rodríguez-Robles et al., 2020; Hafner et al., 2021). Hafner et al. (2021) found in an experiment with six temperate tree species, including the European beech, that the neighboring tree species diversity may not be important for exploiting water uptake through hydraulic redistribution. Both hydraulic niche partitioning and redistribution have been observed vertically, whereas horizontal patterns are largely unexplored in the context of niche partitioning (Hildebrandt, 2020). Our results do not provide direct evidence for either hydraulic redistribution or horizontal niche partitioning. However, they indicate that horizontal root water uptake patterns are regulated by species richness.

5) Conclusion

We investigated the factors that influence the spatial patterns of root water uptake by considering heterogeneity in throughfall and soil water. To that end, we acquired a comprehensive data set based on throughfall measurements paired with soil moisture sensors in a mixed deciduous forest. Soil and neighboring tree characteristics were also included in the linear mixed effects model. We found that variation in root water uptake did not correspond to throughfall. Wetter soil locations, poorly associated





with higher throughfall, increased local root water uptake. In contrast, how average soil water conditions modified root water uptake depended on the neighborhood basal area. As the site dried out, large trees likely took up water in deeper layers to meet transpiration demands. Furthermore, an increase in species diversity promoted root water uptake, similarly depending on the size of neighboring trees, suggesting active complementarity mechanisms in the forest stand. In conclusion, our results suggest that soil water distribution and neighboring tree characteristics regulate root water uptake patterns more than soil properties and throughfall variation.

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Data availability

- 565 The dataset is currently being prepared for publication in an official repository. The DOI will be published
- with the data at the latest when the data are published.

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Author contributions

- 569 GD and AH designed the throughfall measurement setup, AH and JCM designed soil moisture
- 570 measurement. GD conducted the field sampling with assistance from JF and the students listed in the
- 571 Acknowledgments. GD analyzed the data, developed the linear mixed effects model, and analyzed the
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574 Competing interests

- 575 Anke Hildebrandt is part of the editorial board of HESS. The peer-review process was guided by an
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956