# **1** Root water uptake patterns are controlled by tree species

# 2 interactions and soil water variability

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# 17 Abstract

18 Throughfall is the largest source of water entering the soil in forests, and its spatial distribution depends on several biotic and abiotic factors. It is well documented that the distribution of throughfall results in 19 20 reoccurring higher and lower water inputs at certain locations. However, the role of horizontal root water 21 uptake patterns in understanding the effects of throughfall patterns on subsurface water dynamics remains 22 unresolved. Therefore, here we investigate root water uptake patterns by considering spatial patterns of 23 throughfall and soil water patterns in addition to soil and neighboring tree characteristics. In a beech-24 dominated mixed deciduous forest in a temperate climate, we conducted weekly intensive throughfall 25 sampling at locations paired with soil moisture sensors during the 2019 growing season. We employed a 26 linear mixed--effects model to understand controlling factors for root water uptake patterns. Our results 27 show that soil water patterns and interactions among neighbouring trees are the most significant factors 28 regulating root water uptake patterns. Temporally stable throughfall patterns did not influence root water 29 uptake patterns. Similarly, soil properties were unimportant for spatial patterns of root water uptake. We

found that wetter locations (rarely associated with throughfall hotspots) promoted greater root water
uptake. Root water uptake in monitored soil layers also increased with neighbourhood species richness.
Ultimately our findings suggest that complementarity mechanisms within the forest stand, in addition to
soil water variability and availability, govern root water uptake patterns.

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35 Key words: root water uptake, throughfall, soil water, spatial patterns, beech

# 36 1) Introduction

Vegetation intercepts and redirects precipitation into throughfall and stemflow, collectively referred to as below-canopy precipitation. <u>TMoreover, throughfall is usually-typically</u> the largest component of below canopy precipitation (Levia and Frost, 2006; Sadeghi et al., 2020). For instance, in temperate forests throughfall can account for about 70% of above canopy precipitation ends up as throughfall (Levia and Frost, 2003; Sadeghi et al., 2020). <u>This makesHence, throughfall serves as it</u> the primary source of for replenishing soil moisture-replenishment in vegetated areas.

43 Below-canopy precipitation is modified by several biotic and abiotic factors (Levia and Frost, 2006; Levia 44 et al., 2011), such as including vegetation type, and canopy architecture (Crockford and Richardson, 2000; 45 Pypker et al., 2011; Levia et al., 2017), and forest structure (Rodrigues et al., 2022), meteorological 46 elements such as wind speed (Staelens et al., 2008; Van Stan et al., 2011; Fan et al., 2015), precipitation 47 intensity and event size (Dunkerley, 2014; Magliano et al., 2019; Zhang et al., 2016; Staelens et al., 2008). 48 This implies that As a result, throughfallit inherently varies across space and time. Furthermore However, 49 previous studies showed that the spatial distribution of throughfall persists persists repeatedly over time 50 (Keim et al., 2005; Staelens et al., 2006; Guswa and Spence, 2012; Carlyle-Moses et al., 2014; Metzger 51 et al., 2017; Van Stan et al., 2020).

Throughfall patterns potentially translate their spatial variability of water inputs into soil moisture (Raat et al., 2002; Blume et al., 2009; Zimmermann et al., 2009; Zehe et al., 2010; Bachmair et al., 2012;
Rosenbaum et al., 2012; Zhang et al., 2016). A decade ago Coenders-Gerrits et al., (2013) proposed that throughfall patterns are translated into soil wetting dynamics with a model based on combined hillslope

56 topographic and throughfall data collected in a beech-dominated catchment. However, in this model, the 57 effect of throughfall patterns on soil moisture patterns rapidly ceased. Later, Metzger et al. (2017) 58 empirically confirmed that throughfall patterns barely alter soil moisture in-response to rainfall, and the 59 this limited influence rapidly disappears. Recently More recently, Zhu et al. (2021) observed that stable 60 spatial patterns of throughfall were weakly related to the spatial distribution of soil moisture since -61 However, this relationship was restricted only to relatively wet soil locations and throughfall hotspots. 62 They also showed that throughfall patterns had a weak influence on the temporal dynamics of soil water 63 content compared to soil bulk density and litter layer properties.

64 Previous studies have suggestedly proposed explanations for that soil properties (Metzger et al., 2017),

66 et al., 2019; Fischer-Bedtke et al., 2023), and litter layer processes (Raat et al., 2002), and local root water

preferential flow induced by dry antecedent soil conditions (Jost et al., 2004; Blume et al., 2009; Molina

67 <u>uptake enhanced by throughfall hotspots (Bouten et al., 1992; Coenders-Gerrits et al., 2013)</u>the may result
68 <u>in</u> weak and short-term <u>influence effects</u> of throughfall patterns on the soil moisture patterns include:-soil
69 properties (Metzger et al., 2017), preferential flow induced by dry antecedent soil conditions (Jost et al.,
70 2004; Blume et al., 2009; Molina et al., 2019; Fischer et al., 2023), litter layer (Raat et al., 2002), and
71 local root water uptake enhanced by throughfall hotspots (Bouten et al., 1992; Coenders-Gerrits et al.,

72 <del>2013).</del><u>variability.</u>

65

73 MoreoverRegardless, (Fischer-Bedtke et al., (2023)(Fischer et al., 2023) found that recurring throughfall 74 patterns leaveft a notable imprint on thesoil moisture response to rainfall. However, regardless of event size, although these patterns do not leave significant signature on the spatial variation effect on absolute 75 values of in-soil water content even-after drainage was rather weak. There, other factors such as soil 76 77 macroporosity, distance from the tree and other processes, namely fast flow, and root water uptake, more 78 strongly influenceda- soil moisture patterns. Based on a one-dimensional soil-water model, Bouten et al. 79 (1992) proposed that throughfall patterns alter and localize root water uptake -andas well as promote fast 80 drainage-via preferential flow paths. However, to the best of our knowledge, the feedback mechanism of throughfall patterns on root water uptake variation has not yet been investigated in the field. Therefore, it 81 82 is unclear how water uptake patterns play a role in translating throughfall patterns into spatio-temporal 83 variation of soil water and vice versa.

84 In addition to spatial variation of throughfall and soil moisture, soil properties are among the abiotic 85 factors that may alter root water uptake patterns (Nadezhdina et al., 2007; Kirchen et al., 2017). For a 86 given evaporative demand<del>meteorological conditions</del>, water uptake at a particular location is a function of 87 water transport resistance between root and soil in addition to the soil-water potential (Cardon and Letey, 88 1992; Shani and Dudley, 1996; Lhomme, 1998). Both characteristics depend on local soil properties and 89 soil moisture water status, and the latter in turn is affected by the local water uptake rate. Soil water 90 moisture distribution variability may shape root water uptake patterns even more than root networks 91 (Kühnhammer et al., 2020; Guderle et al., 2018). On the flip side, root water uptake can amplify or 92 homogenize soil moisture variability (Hupet and Vanclooster, 2005; Teuling and Troch, 2005; Ivanov et 93 al., 2010; Baroni et al., 2013; Martínez García et al., 2014). Soil properties control soil water redistribution (Grayson et al., 1997; Cosh et al., 2008; Jarecke et al., 2021) and water availability for root structures 94 95 (Vereecken et al., 2007: Cai et al., 2018). Thus soil properties can influence root water uptake patterns 96 (Nadezhdina et al., 2007; Kirchen et al., 2017). Moreover, variations in soil water content reflect root 97 water uptake by root systems (Hupet et al., 2002; Schume et al., 2004; Schwärzel et al., 2009; Guderle and Hildebrandt, 2015; Jackisch et al., 2020). On the flip side, root water uptake can amplify but mostlyor 98 99 homogenize soil moisture variability distribution (Hupet and Vanclooster, 2005; Teuling and Troch, 2005; 00 Ivanov et al., 2010: Baroni et al., 2013: Martínez García et al., 2014). 01 As a resTult, temporal and diurnal changes in local soil water content can be employed to quantify and 02 drive root water uptake by dissecting soil water flow and water uptake under meteorological conditions 03 that ensure transpiration demand (Guderle and Hildebrandt, 2015; Jackisch et al., 2020; Hupet et al., 04 2002). Furthermore Other methods, especially using tracers, exist to evaluate the spatial distribution of 05 root water uptake. Specifically, stable water isotopes can be used to estimate water sources for- water 06 uptake by plant individuals by comparing the isotopic composition of plant xylem water to that of potential 07 water sources using different methods including graphical inference, two-end--member mixing models, 08 multi-source linear mixing models, and physically based analytical models (Rothfuss and Javaux, 09 2017). In addition, tracking isotopically enriched water transport an assist to determine in the 10 determiningdetermination of water uptake dynamics (e.g., Zarebanadkouki et al., 2013) in the laboratory. 11 In contrast to these methods, soil water daily fluctuations in soil water (Guderle and Hildebrandt, 2016)

- allow for estimating the spatial distribution of ecosystem evapotranspiration using standard measurements
   of soil water content (Guderle and Hildebrandt, 2016) without the need for additional infrastructure.
- Root networks can also regulate soil moisture distribution by transporting water from wetter places to drier locations, which has been observed in a variety of ecosystems (e.g., Emerman and Dawson, 1996; Katul and Sigueira, 2010; Yu and D'Odorico, 2015; Privadarshini et al., 2016; Hafner et al., 2017).
- 17 In addition, tree size, age and tree species richness affects the dynamics of spatio-temporal variation in root 18 water uptake (Volkmann et al., 2016; Spanner et al., 2022; Kostner et al., 2002; Dawson, 1996; 19 Brinkmann et al., 2019; Gaines et al., 2016). Neighboring different tree species withutilize different hydraulic strategies, such as may extracting water from different soil depths-regions (Silvertown et al., 20 21 2015; Guo et al., 2018; Brum et al., 2019), and therefore more diverse forest stands can be more resilient under drought stress (Pretzsch et al., 2013). However, soil water scarcity during droughts can initiate or 22 23 enhance competition mechanisms for water among different tree species (González de Andrés et al., 2018; 24 Vitali et al., 2018; Magh et al., 2020). MoreoverFurthermore, studies conducted in temperate forest 25 ecosystems have demonstrated that the relationship between tree species richness and water uptake 26 mechanisms competition-varies (Krämer and Hölscher, 2010; Kunert et al., 2012; Meißner et al., 2012; 27 Forrester, 2014; Lübbe et al., 2016).
- 28 BrieflyTaken together, throughfall and soil water variability, soil properties, and root water uptake 129 patterns form complex and intertwined interactions in the terrestrial hydrological cycle. It has not yet been 30 shown empirically how root water uptake patterns are affected by throughfall and spatial distribution of 31 soil water content. In line with previous suggestions based on modelling modeling approaches results 32 (Bouten et al., 1992; Coenders-Gerrits et al., 2013) we hypothesize that throughfall hotspots maximize enhance water availability for root structures at certain locations that canthanelevate root water uptake. 33 34 Further we investigate the role of <u>and</u>-soil water variation in combination with soil properties and neighboring tree characteristics on root water uptake patterns.- (Bouten et al., 1992; Coenders-Gerrits et 35 al., 2013) Therefore, here we investigate the role of throughfall patterns and We pose the following 36 37 questions to test the main hypothesis and- guide the investigation:
- i) How do throughfall patterns influence root water uptake patterns?

- ii) How do<u>es soil moisturewater and its variation, along with and soil properties</u>, control variation
   in root water uptake?
- What is the role of biotic factors, namely tree size, distance, number, and species richness of
   neighbouring trees, on root water uptake patterns?

Here, we address these questions by employing <u>a</u> linear mixed effects model based on weekly throughfall sampling at locations paired with intensive soil moisture measurements in a beech-dominated unmanaged forest. We estimate root water uptake using a water balance method applied at soil moisture measurement points. In addition, we incorporate data on field capacity, bulk density, and neighboring tree characteristics <u>namely size and species</u>.

# 148 **2) Materials and Methods**

### 149 **2.1) Research Site and Field Sampling**

#### 150 **2.1.1) Research Site**

The research site is located in the forested upper hill region of the Hainich low mountain range in Thuringia, Germany, as a part of the Hainich Critical Zone Exploratory (CZE) (Küsel et al., 2016). The altitude in the research site ranges from 362 m to 368 m a.s.l. Mean annual air temperature varies between 7.5 and 9.5 °C, and the mean annual precipitation ranges from less than 600 to 1000 mm in the CZE (Küsel et al., 2016).

156 In the study area, thin-bedded alternations of limestones and marlstones of carbonate rock (Middle

Triassic) form the bedrock overlain by <u>a</u> shallow Pleistocene loess layer with cambisols and luvisols as

dominant soil types (IUSS Working Group, 2006; Metzger et al., 2021). The median soil depth above the

159 weathered bedrock is 37 cm, with soil depths ranging from 15 cm to a maximum depth of 87 cm (Metzger

160 et al., 2017).

161 In 2019, the tree community in the research site consisted of 574 individuals of various ages (diameter at

162 breast height  $\geq$  5cm). The dominant species is European beech (*Fagus sylvatica* L.), which makes up 70%

163 of the tree community, followed by sycamore maple (Acer pseudoplatanus L.) with 21 %, and European

164 ash (Fraxinus excelsior L.) with 4%. These dominant species are accompanied by Large-leaved linden

165 (Tilia platyphyllos Scop.), European hornbeam (Carpinus betulus L.), Norway maple (Acer platanoides

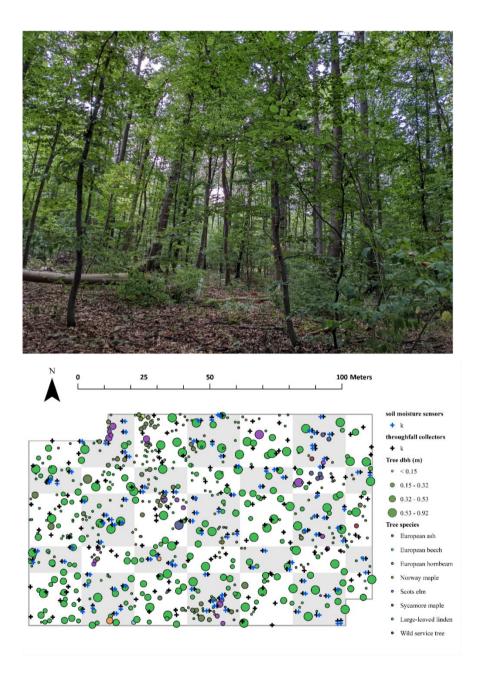
166 L.), Scots elm (*Ulmus glabra* L.), and Wild service tree (*Sorbus torminalis* (L.) Crantz). The stand has a 167 total basal area of  $40 \text{ m}^2 \text{ ha}^{-1}$  and has been unmanaged since 1997 (Kohlhepp et al., 2017).

### 168 **2.1.2) Soil moisture monitoring and soil properties**

169 The forest site (1 ha) was equipped with a soil moisture monitoring network (SoilNet; Bogena et al., 2010) 170 consisting of SMT100 frequency domain sensors (Treuebner GmbH, Neustadt, Germany). Metzger et al. 171 (2017) first described the soil moisture monitoring setup. Briefly, the observation platform (Figure 1) was 172 divided into 100 subplots (10 m  $\times$  10 m), and 49 subplots were equipped with soil moisture sensors at 173 two random measuring points each, for a total of 98 locations. At each measuring point, sensors were 174 placed at two different depths, 7.5 cm (top sensors) and 27.5 cm (bottom sensors). The soil moisture 175 network is maintained through a regular bi-weekly routine to avoid potential failures such as depleted 176 sensors batteries, hardware problems, etc.

177 Undisturbed soil samples were collected during the sensor installation in 2014 and 2015 to estimate bulk 178 density and water content at field capacity. In addition, we collected additional disturbed soil samples (n 179 = 40) near sensor locations in 2019. Bulk density was determined from oven-dried (24h, 105°C) soil mass 180 weight and water content at field capacity by applying 60 hPa pressure to the saturated undisturbed sample 181 for 72 h.

Soil properties vary slightly from top to subsoil at the research site. While silty loam is the dominant soil texture in both layers, the clay content is higher in the subsoil (Metzger et al., 2021). The median volumetric water content at the field capacity is 44% in the topsoil and 42% in the subsoil. Moreover, the water content at the field capacity varies from 27% to 60% and from 31% to 62% in the topsoil and subsoil, respectively. The average bulk density (d<sub>bulk</sub>) of the topsoil is 1.16 g cm<sup>-3</sup>, with a range of 0.73 to 1.5 g cm<sup>-3</sup>. In the subsoil, the average bulk density (d<sub>bulk</sub>) is slightly higher at 1.37 g cm<sup>-3</sup> but has a similar range (0.7 - 1.6 g cm<sup>-3</sup>) (See supplement for details).



189

190 Figure 1 (above) The photo of the site. (below) the field monitoring setup of stratified randomly distributed throughfall 191 collectors and soil moisture sensors together with the trees which are sized according to the diameter at breast height (dbh) 192 and coloured according to the species. Throughfall collectors are paired with soil moisture sensors at 98 locations (n=182) in 193 the grey shaded subplots. White coloured subplots are equipped with only throughfall collectors.

### 194 **2.1.3**) Gross precipitation and throughfall sampling

Five gross precipitation funnels were placed 1.5 m above ground level in an adjacent open grassland (ca. 250 m distance to the research site). As described in Metzger et al. (2017) and Demir et al. (2022), the precipitation funnels were made of a circular plastic funnel (12 cm in diameter) and sampling bottle (2 L in volume), and ping pong balls were placed in the funnel orifice to prevent evaporation losses.

During the early growing season of 2019, we placed throughfall collectors in soil moisture monitoring subplots at 98 locations. We paired these throughfall collectors with the soil moisture sensors by placing them within 1 m of each other. The paired collectors were placed down-slope to avoid interference with soil moisture measurements. For the rest of the research site, in 51 other subplots, we adopted a separate independent stratified random design from Metzger et al. (2017). Briefly, we placed two throughfall collectors in each subplot that was not equipped with soil moisture sensors. All throughfall collectors were placed roughly 37 cm above the ground.

We conducted <u>mostly</u> weekly manual measurement of throughfall and gross precipitation during the 2019 growing season (April to August). <u>Sampling was only-conducted</u>. <u>Gross precipitation and throughfall</u> <u>was read out on rainfreerain free days only-without rain.</u> We measured gross precipitation and throughfall on rainless days <u>Thus</u>, the sampling interval therefore, in some of the sampling weeks, the interval between field measurements ranged between six and eight day<u>s.s.-depending on the occurrence of rain</u> <u>events.</u>.

We used the paired throughfall collectors (n = 98) to identify the drivers of root water uptake patterns, as we derived root water uptake values based on soil water content measurements (see below). However, we used all randomly placed throughfall collectors (n = 200) to describe the spatio-temporal variation of throughfall within the research site.

### 216 **2.2) Estimation of potential evapotranspiration**

We calculated the daily potential evapotranspiration by applying the concept of thermodynamic limits of convection <del>(</del>(Kleidon and Renner, 2013; Kleidon et al., 2014):

219 
$$E_{\text{pot}} = \frac{1}{\lambda} \frac{s}{s+\gamma} \frac{R_{\text{sn}}}{2}$$
(1)

Where  $R_{sn}$  is absorbed solar radiation (W m<sup>-2</sup>),  $\lambda$  is the latent heat of vaporization (2.5×10<sup>6</sup> Jkg<sup>-1</sup>),  $\gamma$  is the psychrometric constant (65 PaK<sup>-1</sup>), and *s* is the slope of the saturation vapor pressure curve (PaK<sup>-1</sup>). Here, we acquired solar radiation, air temperature, and precipitation data for the throughfall sampling period from a nearby weather station ("Reckenbuel") which is located approximately 1.4 km northeast of the research site and provides data in 10 minutes intervals. The site-specific albedo for the summer period was adopted from Otto et al. (2014).

In addition, wWe used the precipitation data measured at the weather station to define rain events and dry
 periods, as described below.

### 228 2.3) Data analysis

#### 229 2.3.1) Quality control of soil water content data

230 We systematically reviewed the six-minute soil water content data for quality control in two steps: 1) 231 identification of problems (such as jumps to extremely low and high values, duplicated time stamps of 232 different values, long discontinuities in the measurements, and lack of temporal variation in the time series 233 despite rain events), 2) classification and removal of detected outliers and irregularities. We visually 234 identified and removed unrealistic measurements such as extremely low (< 5 vol-%) and high values far 235 beyond the field capacity (>75 vol-%) and long plateaus of repeated values despite rain events. We also 236 excluded anythe time series that exhibited long-term discontinuities that prevented us from calculating 237 root water uptake. During the visual inspection, we eliminated values with duplicated time stamps that 238 violated the actual temporal trend. Next, we scanned the data using the Hampel filter function of the 239 'pracma' R package (Borchers, 2021) with customized moving window length and Pearson's rule threshold 240 value (Pearson, 1999) to flag possible outliers.

Despite regular maintenance, many sensors failed to <u>provide data that metmeets</u> the quality criteria in during the growing season (March-August) in 2019. Only 56 sensor locations (out of 98) <u>simultaneously</u> provided <u>high-quality</u> data from both top and bottom sensors <u>that met the qualification criteria described</u> <u>above</u> with <u>different-varying datetime</u> intervals throughout the <u>growing</u> season. Of these, only 34 sensor locations <u>provided were data-used for to estimate the</u>-root water uptake <u>estimationas they simultaneously</u> <u>provided data from both top and bottom sensors within the dry periods.-</u>

### 247 2.3.2) Soil water calculation

248 We estimated soil water (S) at measurement locations for the monitored soil layer based on volumetric

soil water content measured by top and bottom sensors.

250 
$$\mathbf{S}_{i,d} = \sum \mathbf{z}_t \theta_{i,d}^t + \mathbf{z}_b \theta_{i,d}^b$$
(2)

251 We similarly integrated the soil water at field capacity  $(S_{FC,i})$ 

252 
$$S_{FC,i} = \sum z_t \theta_{FC,i}^t + z_b \theta_{FC,i}^b$$
(3)

where  $z_t$  is the depth of the soil column monitored by the top sensor and  $z_b$  is the depth of soil represented by the bottom sensor, and  $\theta_{i,d}$  is  $\oplus$  volumetric soil water content at location *i* on date *d*, and  $\theta_{FC,i}$  the soil water content at the field capacity.

256 We calculated bulk density at the sensors' locations for the monitored soil layer.

257 
$$\overline{d_{bulk,i}} = \frac{\sum z_t d_{bulk,i}^t + z_b d_{bulk,i}^b}{\sum z_t + z_b}$$
(4)

258 where  $d_{bulk,i}^{t}$  and  $d_{bulk,i}^{b}$  are the bulk density of the topsoil and subsoil, respectively, at location *i*.

### 259 2.3.3) Descriptive Statistics

We calculated the coefficient of quartile variation (CQV) and the interquartile range to describe spatial variation of throughfall, volumetric soil water content, and root water uptake. Also, we estimated octile skewness (OS<sub>8</sub>) of throughfall based on the first and seventh octile and standard deviation (SD) of the estimated daily root water uptake.

264 
$$CQV = \frac{Q_3 - Q_1}{Q_3 + Q_1}$$
 (5)

265 
$$OS_8 = \frac{(Q_7 - median) - (median - Q_1)}{Q_7 - Q_1}$$
 (6)

We characterized spatial patterns of daily root water uptake ( $E_t$ ) by calculating the spatial deviation from the mean ( $\delta E_{t,i,d}$ , Equation 7) (Vachaud et al., 1985).

$$268 \qquad \delta E_{t i,d} = \frac{E_{t, i,d} - \overline{E_{t,d}}}{\overline{E_{t,d}}}$$
(7)

where  $E_{t, i, d}$  is daily root water uptake estimated at *i* sensor location on date *d* and  $\overline{E}_{t, d}$  is spatial average of daily root water uptake on date *d*. 271 Similarly, we calculated the spatial deviation of soil water and throughfall to identify their spatial patterns.

## 272 2.4) Root water uptake estimation

We estimated root water uptake using the multi-step, multi-layer regression method (MSML), which <u>is a</u>
<u>water--balance method and</u> derives evapotranspiration from diurnal differences in soil water content
(Guderle and Hildebrandt, 2015; Guderle et al., 2018). This approach does not require prior information
on root structure but relies on high temporal and spatial resolution data on multiple soil layers. <u>This</u>
<u>method has previously</u>Previous studies <u>been applied</u> using additional measurements such as sap-flow
and lysimeters demonstrated that the MSML method successfully estimates <u>to estimate transpiration in</u>
<u>both forest and grassland ecosystems (Guderle et al., 2018; Jackisch et al., 2020).</u>

As described in Guderle and Hildebrandt (2015), the MSML derives root water uptake from distinct differences in the day and night portions of soil moisture time series. The main assumption is that, in the absence of rainfall-driven rapid vertical soil water flow, evapotranspiration occurs only during the day, while soil water flow occurs both during the day and at night. As a result, soil moisture time series reflect a distinct day/night signal under dry weather conditions. This method has previously been applied to estimate transpiration in both forest and grassland ecosystems (Guderle et al., 2018; Jackisch et al., 2020).

286

Therefore<u>In applying this method to our study</u>, we first excluded potential periods of fast vertical flow periods from the time series due to previous rainfall events and identified periods for estimating daily root water uptake. We considered <u>an 8</u> h buffer period to include canopy dripping and 48 h for the cessation of rainfall influence on soil water. Thus, a total of 56 h was the time interval used to define the <u>start of</u> <u>the</u> water uptake estimation period. The period when the root water uptake is estimated is hereafter referred to as the dry period.

Next, we split each soil moisture time series into a day (transpiration active period) and a night branch, as <u>explained by</u> Guderle and Hildebrandt (2015) <u>explained</u>. We defined the transpiration period (starts 2 h after sunrise and ends 2 h before sunset) based on local sunrise and sunset time. Sunrise and sunset times were obtained from the R package 'suncal' (Thieurmel and Elmarhraoui, 2022). We fit linear models to each split branch of the time series and derived the slopes. The difference between the slope of the day branch ( $m_{tot}$ ) and the average slope of the antecedent and preceding night ( $\overline{m_{flow,i}}$ ) gives the rate of water uptake. Thus, we estimated daily evapotranspiration at each soil water content location *i* (Equation 8, 9) by accounting for soil layer thickness and slope difference-

301

302 
$$E_{t,msml,i}^{t,b} = (m_{tot,i}^{t,b} - \overline{m_{flow,i}^{t,b}}) d_{z,i_{z,i}}^{t,b}$$
 (8)

303 
$$E_{t,i} = \sum (E_{t,msml,i}^t + E_{t,msml,i}^b)$$
 (9)

304

### 305 2.5) Linear Mixed Effects Model

We employed a linear mixed effects model to investigate the driving factors for root water uptake patterns. 306 307 A linear mixed effects model is a multivariate statistical tool that. It describes the relationship between a 308 dependent variable and explanatory variables (fixed effects) while controlling for dependencies in the 309 data that may arise due to repeated sampling with certain designs (random effects). Fixed effects are 310 informative, repeatable levels of explanatory and quantified variables that can influence the mean of the dependent variable, and they can be tested. In addition, in a linear mixed-effects model, how the 311 312 relationship between the dependent variable and one predictor how dependsings on the level of another 313 predictor can be represented via-by the interaction term. 314 Random factors are uninformative levels of predictor variables but can explain parts of the residual of the fixed effects model by calculating different intercepts for different category levels. They are included in 315 316 mixed effects models to account for qualitative information from repeated sampling with respect to

317 individuals, time stamps, or treatments. For the model, we considered Here, Because of repeated

318 observations at the measurement locations, soil moisture sensor pointslocation and time stampdry periods,

319 i.e. date. drv periods. (i.e., the root water uptake estimated time interval), were considered are taken as

320 <u>random effects because of repeated observations at the measurement locations.</u>

321 <u>To ensure spatial co-location of water input, soil water and root wate uptake we used only paired</u>

\$22 throughfall and soil moisture measurement locations where both top and bottom sensors provided data

323 <u>within the dry periods.</u>

324

325 For the model, we used only paired throughfall and soil moisture measurement locations where both top 326 and bottom sensors provided data during within the dry periods. All All considered potential controlling 327 explanatory driversfactors, which are included as fixed factors in the model-so can be tested, for root 328 water uptake patterns are listed in Table 1. -These factors include abiotic and biotic variables that possibly 329 influence relative local root water uptake: They are daily spatial average soil water storage, the spatial 330 deviation of soil water from the mean, soil water at field capacity and bulk density of the monitored soil 331 layer to represent spatial variability in how and where water stored in soil together with soil properties. 332 Moreover, we To account for spatial variability in water input throughfall, wWe guantified calculated the 333 spatial variabilitdeviation from the mean v ofby using throughfall as the difference between the 334 throughfall measured at a given location and the spatial mean, normalized by the spatial meanEquation 7 335 to account spatial variability in water input. Here we considered this variable at a two-different time 336 scales: the sampling week(s) prior to root water uptake estimation period, and the median of over the entire 337 measurementhroughfall sampling t-period.

Further, as biotic factors, we included, -number of trees, and number of species within a 5 m radius of each soil moisture location, and inverse\_-distanced-weighted -basal area (BA) within 5 m radius of each soil moisture location, Basal area was calculated as follows:

$$341 \qquad BA_i = \frac{\sum_{R=1}^{R} W_R A_{tree}}{(10)}$$

342 with  $W_R = \frac{(x_i - x_R)^2}{\sum_R (x_i - x_R)^2}$  (11)

where i is the soil moisture sensor located at  $x_i$ , R is the tree index located at  $x_R$ , and A<sub>tree</sub> is the individual basal area of the corresponding tree, A is the area around the soil moisture sensor i with 5 m in radius. Even though our research plot is a beech-dominated forest, in some spots, two to four species were present

346 within a 5 m radius of the soil moisture sensors.

347 Moreover, we quantified the spatial variability of throughfall as the difference between the throughfall 348 measured at a given location and the spatial mean normalized by the spatial mean. Here we considered 349 this variable at a two-time scales: the week(s) prior to root water uptake estimation period, and the median 350 of the entire measurement period. We also included interaction terms (Table 1) as fixed factors in the

- model to capture complex and non-linear relationships among the biotic and abiotic factors to estimate
   relative root water uptake.
- 353 . Because of repeated observations at the measurement locations, soil moisture sensor points and dry
   354 periods, (i.e., the root water uptake estimated time interval), were considered as random effects.
- We conducted all analyses with the R statistical software (R Core Team, 2022) and used the *lmer* function in the 'lme4' package (Bates et al., 2015) for the model development. We visually checked the model assumptions using the 'check model' function of the 'performance' package (Lüdecke et al., 2021).
- In addition, we calculated both conditional and marginal  $R^2$  of the model with the 'MuMIn' package (Bartoń, 2020). While the conditional  $R^2$  includes the variance of the entire model, the marginal  $R^2$ subsumes only the fixed effects (Bartoń, 2020). Before fitting the linear mixed effects model, we tested for co-linearity of the considered variables and scaled the data with a Z-transformation by using the 'scale' function in base R (R Core Team, 2022), which allowed us to evaluate the individual effect of fixed effects by comparing slopes and significance levels.
- 364 We developed the optimal model by applying a systematic model selection procedure based on Akaike's 365 Information Criterion (AIC) comparison in combination with the examination of the factors. Model selection began with the beyond-optimal model, which included all possible fixed and random effects. 366 367 We stepwise evaluated each fixed effect based on its respective significance (p value comparision) by fitting the model the maximum likelihood (ML) to be able to compare AIC values (Zuur et al., 2009). In 368 369 each step, starting with interaction terms, we identified the least significant effect and formulated a model 370 without it. We compared the AIC values of the model before and after removing the effect, discarding it 371 in case the AIC was unaffected or decreased. We followed the procedure with the next equally detected 372 effect, and repeated it until only significant fixed effects remained, and the model with the lowest AIC 373 (the optimal model) was obtained.
- As a final step, the best model was refitted with restricted maximum likelihood (REML) (Zuur et al.,2009).
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- 377
- 378

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**Table 1** List of fixed and random factors considered for estimating the root water uptake patterns through linear mixed effects model. Interaction is shown with 'x'.

Fixed Factors	
Single Factors	Interaction Factors
Spatial average of soil water storage in the monitored soil layer ( $\overline{S}$ )	$\bar{S} \times S_{FC}$
Spatial deviation of soil water storage from the mean ( $\delta S$ )	$\delta S \times S_{FC}$
Field capacity of the monitored soil layer $(S_{FC})$	$\delta S \times BA$
Bulk density capacity of the monitored soil layer (d <sub>bulk</sub> )	$\overline{S} \times BA$
Spatial deviation of throughfall of events measured in sampling week previous to the corresponding dry period ( $\delta P_{TF_{last ev.}}$ )	$\delta S \times n_{tree}$
The median of spatial deviation of throughfall measured within the whole sampling period $(\widetilde{\delta P_{TF}})$	$\overline{S} \times n_{tree}$
Number of trees (n <sub>tree</sub> )	$\delta P_{TF_{last ev.}} \times S_{FC}$
Basal area (BA)	$\delta P_{TF_{temp. stable.}} \times S_{FC}$
Number of species (n <sub>sp,tree</sub> )	$\delta P_{TF_{last ev.}} \times d_{bulk}$
	$\delta P_{TF_{temp. stable.}} \times d_{bulk}$
	$n_{sp,tree} \times WA_{int}$
Random factors	
Soil moisture sensor location	
Dry period	

Dry period

# **383 3) Results**

# 384 3.1) Spatio-temporal distribution of throughfall and soil water content

In 12 out of the 16 sampling weeks, the weekly gross precipitation was more than half of the total potential
evapotranspiration. Table 2-further shows the distribution of throughfall sampled in 2019 (April-August)

at 200 collectors and the 98 collectors that were paired with soil moisture sensors. <u>WThe-weekly</u> throughfall increased with <u>anthe</u> increase in rain-events. <u>Additionally, T</u>the coefficient of quartile variation (CQV) of throughfall was generally lower for larger cumulative weekly rains. On average, the <u>paired</u>-collectors <u>paired with soil-moisture sensors</u> received similar amounts of throughfall to all collectors (Table 2). The CQV of data from the paired collectors ranged from 0.27 to 0.6, which is similar to the CQV of throughfall sampled at all collectors. The octile skew (OS<sub>8</sub>) of paired and <del>of</del> all collectors was also similar.

As the growing season progressed in 2019, the average soil water content decreased in both the topsoil and subsoil. In April and early May, the average volumetric soil water content in the topsoil was above 30%, <u>andwhich</u> dropped to below 10% by the end of August. In the subsoil, the volumetric soil water content similarly declined from above 40 % to below 20 % over the sampling period (Figure 2). On average, soil water changed from 52.5mm to 17.5 mm in the topsoil and from 80 mm to 40mm in the subsoil.

400 We derived root water uptake for four periods (a total of 19 days) under different soil wetness conditions 401 that captured the seasonal variation of soil water content, including late spring when the soil water content 402 was higher and drier periods during the summer, following re-wetting-wetted soil conditions with late summer rains. As listed in Table 3 and shown in Figure 2, two periods were in late May and early June, 403 404 and each lasted two days. The third period began in late June and lasted 11 days; the last was four days 405 in late July. From the start of the first dry period to the end of the last <del>During these periods</del>, the average 406 soil water content declined from 33 to 15 % in the topsoil and from 43 to 27% in the subsoil. Table 3 407 additionally shows that within the dry periods, the coefficient of quartile variation (CQV) of soil water 408 content was between 0.09 -0.14 and 0.08 to 0.16 in the topsoil and subsoil, respectively. During the dry 409 periods, the spatial heterogeneity of soil water content in the subsoil increased systematically. In contrast, 410 the spatial variation of topsoil soil water content in the topsoil didwas not correlated with soil dryness.

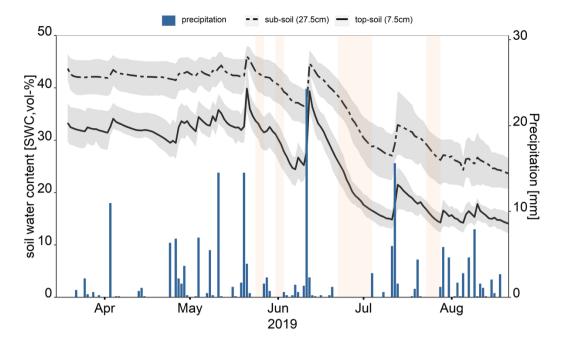




Figure 2 Soil moisture temporal variation in top and subsoil together with the daily precipitation measured at the nearby Reckenbühl station (approximately 1.4 km to the Northeast). The solid and dashed lines are spatial mean of soil water content estimated based on top (7.5 cm) and bottom (27.5 cm) sensors, and grey shaded areas show first and third quartiles. The reddish shaded areas show defined dry periods within the throughfall sampling when root water uptake could be estimated.

ve potent 1, spatial	mm, interquartile range (LQK), coefficient of quartile variation (CQV) and octile skewness (USs) of both all and paired through tail lectors during the sampling week. The values are ordered according to the cumulated gross precipitation size.
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Date	$\mathbf{E}_{\mathrm{pot,cum}}$	$P_g$	$P_g/\mathrm{E}_{\mathrm{pot}}$	$\overline{P_{\mathrm{TF}}}$	IQR P <sub>TF</sub>	P <sub>TF</sub>	$\frac{OS_8}{P_{TF}}$	$\overline{\mathbf{P}_{\mathrm{TF}}}_{\mathrm{paired}}$	IQR PTF <sub>naired</sub>	$\frac{CQV}{P_{TF}_{naired}}$	$\frac{\mathrm{OS}_8}{\mathrm{P}^{\mathrm{TF}}\mathrm{naired}}$
04-06-2019	13.55	0.76	0.06	0.35	0.18	0.25	0.46	0.34	0.16	0.24	0.49
26-06-2019	20.87	1.73	0.08	0.97	0.44	0.24	0.16	0.98	0.53	0.27	0.27
17-04-2019	5.62	2.42	0.43	1.72	0.27	0.08	0.23	1.72	0.33	0.09	0.09
18-06-2019	9.46	4.00	0.42	2.58	0.62	0.12	-0.03	2.57	0.53	0.10	-0.08
29-05-2019	10.15	6.27	0.62	3.77	1.24	0.17	-0.52	3.63	1.50	0.21	-0.42
24-07-2019	13.52	7.80	0.58	4.61	1.06	0.12	-0.34	4.48	0.88	0.10	-0.63
21-08-2019	8.94	8.54	0.96	5.19	1.06	0.10	-0.47	5.17	0.97	0.10	-0.44
30-07-2019	12.68	10.73	0.85	7.81	2.25	0.15	-1.51	7.58	2.28	0.15	-1.17
07-05-2019	6.65	12.56	1.89	9.21	1.33	0.07	-0.75	9.21	1.99	0.11	-1.05
14-08-2019	8.51	13.79	1.62	11.19	2.65	0.12	-1.40	10.99	2.98	0.13	-1.13
08-08-2019	13.91	23.87	1.72	16.60	2.65	0.08	-1.10	16.52	2.65	0.08	-1.17
30-04-2019	5.93	24.47	4.13	18.44	3.09	0.08	-1.63	18.30	2.65	0.07	-1.23
17-07-2019	8.28	29.27	3.54	24.22	3.54	0.07	-2.08	24.39	3.54	0.07	-2.59
15-05-2019	7.42	29.53	3.98	22.10	3.54	0.08	-2.11	22.21	3.54	0.08	-2.11
22-05-2019	6.74	41.82	6.20	30.94	3.54	0.06	-3.04	30.54	3.54	0.06	-3.46
13-06-2019	14.47	71.84	4.96	57.77	8.51	0.07	-5.82	57.99	7.29	0.06	-6.52

**Table 3** The spatial average of daily volumetric soil water content ( $\overline{\theta_{top-soil}}$ , vol-%) in topsoil (0-17.5 cm), and ( $\overline{\theta_{subsoil}}$ , vol-%) 417

418 in subsoil (17.5 - 37.5 cm) during the defined dry periods. The inter quartile range (IQR), and coefficient of quartile variation 41

419	(CQV) of daily	v volumetric soil	water content in l	both layers	during the dry periods.
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Date	$\overline{oldsymbol{ heta}_{top-soil}}$ (vol-%)	IQR $\theta_{top-soil}$ (vol-%)	CQV $\theta_{top-soil}$ (vol-%)	$\overline{oldsymbol{ heta}_{sub-soul}}_{( ext{vol-\%})}$	$\begin{array}{c} IQR  \theta_{subsoil} \\ (vol-\%) \end{array}$	$\begin{array}{c} CQV  \theta_{subsoil} \\ (vol-\%) \end{array}$	Dry Period
25 -05-2019	33.17	5.72	0.09	42.82	6.72	0.08	1
26-05-2019	32.12	6.62	0.10	42.46	6.67	0.08	1
01-06-2019	30.23	6.87	0.12	40.61	6.9	0.09	2
02-06-2019	29.22	7.23	0.13	40.11	6.85	0.09	2
23-06-2019	25.01	6.69	0.14	37.80	6.38	0.08	3
24-06-2019	24.04	6.45	0.14	36.94	6.22	0.08	3
25-06-2019	22.52	5.43	0.12	36.13	6.54	0.09	3
26-06-2019	21.48	5.07	0.12	35.24	6.71	0.10	3
27-06-2019	20.20	4.25	0.11	33.98	7.75	0.12	3
28-06-2019	19.45	3.85	0.10	33.31	8.08	0.12	3
29-06-2019	18.98	3.83	0.10	32.36	8.05	0.12	3
30-06-2019	18.44	3.52	0.09	31.37	8.15	0.13	3
01-07-2019	17.67	3.62	0.10	30.45	8.18	0.13	3
02-07-2019	17.29	4.18	0.12	29.84	8.87	0.15	3
03-07-2019	16.89	3.72	0.11	29.26	8.98	0.15	3
24-07-2019	16.15	3.48	0.11	28.56	8.7	0.16	4
25-07-2019	15.51	3.47	0.11	27.85	8.67	0.16	4
26-07-2019	14.98	3.57	0.12	27.21	8.49	0.16	4
27-07-2019	14.57	3.65	0.13	26.65	8.63	0.16	4

420

#### 3.2) Soil water storage, potential evapotranspiration, and root water uptake 421

422 The integrated field capacity of the monitored soil depth was 160 mm on average at the research site. 423 Table 4 shows that soil water storage was much lower than the field capacity during the dry periods, and 424 the mean soil water storage dropped below 42 mm in late July. In addition, Table 4 demonstrates that the average root water uptake ( $\overline{E}_t$ ) ranged from 0.94 mm d<sup>-1</sup> to 3 mm d<sup>-1</sup> while potential evapotranspiration 425 (E<sub>pot</sub>) ranged from 1.75 mm d<sup>-1</sup> to 3.12 mm d<sup>-1</sup>. The discrepancy between average root water uptake and 426 427 the potential evapotranspiration increased as soil water storage assessed by the soil sensors progressively 428 decreased, especially during the longest dry period (Table 4). Root water uptake showed greater spatial 429 variation than water input and soil wetness. The coefficient of quartile variation (CQV) of root water 430 uptake ranged from 0.15 to 0.28, which was higher than the CQV of throughfall and volumetric soil water 431 content in both soil layers.

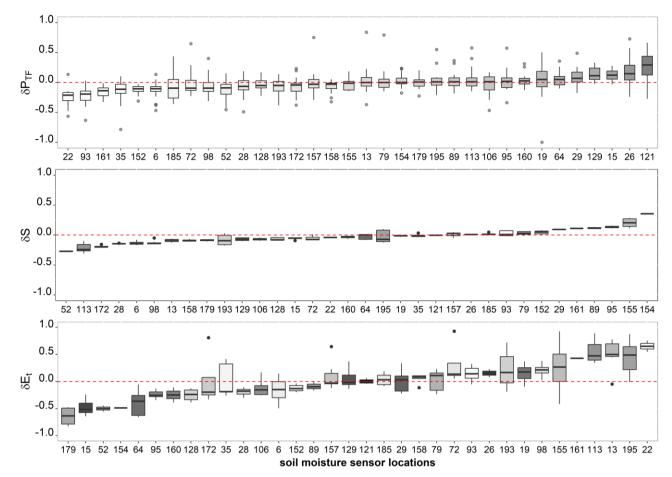
432 **Table 4** The daily average air temperature ( $T_{air}$ , °C), potential evapotranspiration ( $\underline{E}_{pot}$ , mm), mean soil water storage ( $\overline{S}$ , mm)

433 in monitored soil layer (0 - 37.5 cm), and spatial mean of daily root water uptake ( $\overline{E_t}$ , mm) based on all soil moisture sensors, 434 and the ratio of the root water uptake to the potential evapotranspiration together with and standard deviation (SD) and 435 coefficient of quartile variation (CQV) of the daily root water uptake during the defined dry periods

Date	T <sub>air</sub> (°C)	E <sub>pot</sub> (mm)	<u>5</u> (mm)	$\frac{\overline{E_t}}{(\mathbf{mm})}$	$\overline{E_t} / E_{\text{pot}}$	SD $\overline{E_t}$	$\operatorname{CQV}\overline{E_t}$	Dry Period
25-05-2019	12.74	1.80	71.94	1.09	60.56	0.38	0.28	1
26-05-2019	14.43	1.90	70.57	1.30	68.42	0.48	0.25	1
01-06-2019	18.42	2.59	67.16	2.26	87.26	0.98	0.27	2
02-06-2019	21.38	2.77	65.79	2.50	90.25	1.12	0.18	2
23-06-2019	19.45	2.79	59.81	2.83	101.43	0.90	0.19	3
24-06-2019	20.22	2.82	58.16	2.62	92.91	0.76	0.17	3
25-06-2019	22.52	2.89	55.96	2.67	92.39	0.78	0.16	3
26-06-2019	25.73	2.96	54.13	3.00	101.35	0.88	0.15	3
27-06-2019	18.83	2.75	51.91	2.28	82.91	0.55	0.16	3
28-06-2019	16.07	2.58	50.55	1.53	59.30	0.40	0.20	3
29-06-2019	19.59	2.85	49.55	2.11	74.04	0.60	0.20	3
30-06-2019	25.54	3.12	48.26	2.57	82.37	0.86	0.18	3
01-07-2019	20.63	2.30	46.69	1.59	69.13	0.53	0.18	3
02-07-2019	14.88	1.75	45.81	1.08	61.71	0.42	0.24	3
03-07-2019	13.77	1.91	44.95	0.94	49.21	0.30	0.23	3
24-07-2019	24.39	2.76	43.61	1.88	68.12	0.64	0.19	4
25-07-2019	25.33	2.82	42.31	1.77	62.77	0.60	0.24	4
2019-07-26	23.27	2.64	41.18	1.40	53.03	0.55	0.18	4
2019-07-27	21.29	2.68	40.23	1.21	45.15	0.47	0.19	4

### 436 **3.3**) Soil water, throughfall, and root water uptake patterns

437 At soil moisture measurement points where daily root water uptake was determined (n = 34), we 438 calculated the spatial deviation from the median of throughfall, soil water storage, and root water uptake 439 to illustrate the spatial patterns. Figure 3 separately shows that some locations received repeatedly less 440 (or more) throughfall than average ( $\delta P_{TF} < 0$ ) throughout the sampling season. Similarly, some locations, 441 either stored less water in the soil, i.e., were repeatedly wetter or drier ( $\delta S < 0$ ), and some places regularly 442 had lower or higher root water uptake ( $\delta E_t$ ) or higher than average water uptake throughout the sampling 443 period. However, these locations were not related to each other. In fact, Figure 3 demonstrates that neither 444 throughfall nor soil water patterns are directly correlated with the root water uptake patterns. For example, 445 the locations with higher water uptake were not coupled with elevated throughfall input (locations colored 446 dark) or higher soil water storage. In addition, soil water storage patterns were not correlated with 447 throughfall patterns.



448

Figure 3 Temporal stability of throughfall patterns which is estimated by the spatial deviation from the mean ( $\delta P_{TF}$ ) throughout the sampling period in 2019 (April-August), soil water ( $\delta S$ ) and root water uptake ( $\delta E_t$ ) based on the spatial deviation from the mean during the defined dry periods. Soil moisture sensor locations colored according to throughfall input.

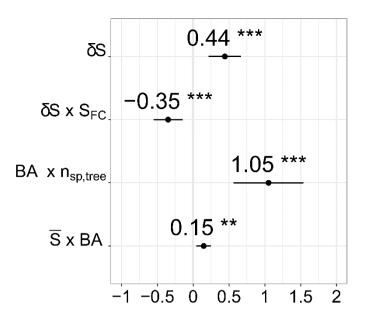
### 452 **3.5**) Fixed factors regulating root water uptake patterns

We used a linear mixed effects model to disentangle the effects of throughfall, soil water, soil properties, and the neighbouring tree characteristics on root water uptake patterns. The fixed and random effects contributed almost equally to the model. The  $R^2$  of the model was 0.77, and the contribution of the fixed

- 456 effect to the  $R^2$  was 0.39 (See the supplement for more details on the optimal model).
- 457 Figure 4 shows only the significant fixed effects for root water uptake patterns. Spatial deviation of soil

458 water from the mean (i.e., soil water patterns) was the only single and the most significant factor positively

- 459 related to the spatial deviation of root water uptake. Thus, water uptake was elevated at locations where
- 460 the most water was retained in the soil at the given time, i.e., greater soil water storage.



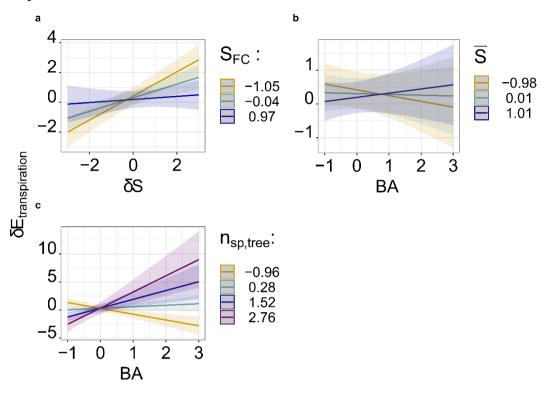
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**Figure 4** The significant fixed factors of the best model to estimate root water uptake patterns ( $\delta E_t$ ). Values on the x-axis indicate the slope of the relations. All variables were scaled by Z-transformation. Interaction is shown with 'x'. Here  $\delta S$  is the spatial deviation of soil water,  $S_{FC}$  is the field capacity,  $n_{sp,tree}$  is the number of species, BA is the basal area, and  $\overline{S}$  is soil water storage. Significance codes are \*\*\*  $\cong 0, ** \cong 0.001$ . (the details on the model can be found in the supplement)

466

Field capacity by itself was not a significant factor affecting local root water uptake. However, it strongly influenced how local soil water\_-controlled root water uptake as a part of the significant interaction term. Figure 5a illustrates how to root water uptake was more dependent on local soil water when field capacity was low (i.e., higher macroporosity). In contrast, soil bulk density and therefore total porosity was not part of the final model.

Although the spatial average of soil water storage, e.g., the state of wetness, was not an important factor for local root water uptake by itself, it moderated the impact of basal area (BA) on the spatial distribution of water uptake. We found that as the plot dries, uptake shifts from places with higher to places with lower basal area (Figure 5b). Furthermore, the statistical model revealed that water uptake increased with the higher basal area at locations where multiple species co-existed (Figure 5c). However, the number of species and the basal area were individually not significant fixed effects. Lastly, throughfall patterns were not significant predictors of local root water uptake. Only the median of the spatial deviation of 479 throughfall, which represents temporally stable patterns within the sampling period ( $\delta P_{TF}$ ), marginally 480 improved the final model.



481

**Figure 5** Visualisation of the significant relations shown in Figure 4, representing the significant drivers of root water uptake patterns during the defined dry periods. Relation to (a) interactive relation of the spatial deviation of soil water storage and field capacity ( $S_{FC}$ ), (b) the interactive relation of basal area (BA) and the spatial average of soil water storage ( $\overline{S}$ ), (c) the interactive relation of number of species ( $n_{sp,tree}$ ) and basal area (BA).

# 486 **4) Discussion**

We investigated the role of throughfall, soil water patterns, and soil and tree characteristics on the spatial variation of root water uptake. In the following sections we discuss three main findings, which are: (1) Contrary to our hypothesis, throughfall patterns do not play a role not in root water uptake patterns despite the recurrence of distinctly localized greater and lesser throughfall inputs. (2) How and where water is stored in the soil, which, which is strongly determined by water holding capacitysoil hydraulic properties, dominates water uptake patterns. (3) The size and species of neighbouring trees regulate relative local

493 <u>water uptake such that more diverse locations surrounded by more diverse neighbourhoods are subject to</u>
 494 <u>greater water uptake.</u>

495

### 496 **4.1) Spatial variation in throughfall does not affect root water uptake patterns**

497 We adequately captured the spatial distribution and temporal stability of throughfall at locations where 498 local root water uptake was derived. Consistent with previous observations in temperate forests (e.g., 499 Whelan and Anderson, 1996; Staelens et al., 2006; Metzger et al., 2017), the amount of weekly rainfall 500 significantly altered the spatial distribution of throughfall such that more rainfall, and thus more 501 throughfall, resulted in less spatial variability. Previous studies repeatedly showed that throughfall 502 patterns exhibit temporal stability in forest ecosystems (e.g., Keim et al., 2005; Staelens et al., 2006; 503 Wullaert et al., 2009; Rodrigues et al., 2022). At ourthe same research site, using event-based sampling, 504Metzger et al., (2017) and (Fischer-Bedtke et al., (2023)Fischer et al., (2023) demonstrated that 505 throughfall patterns persist over time, which was not different inalso true for our weekly sampling in 506 2019. With canopy cover being the key driver of throughfall -(Fischer-Bedtke et al., 2023)(Fischer et al., 507  $\frac{2023}{1000}$ , it is not surprising that weekly cumulative events resulted in a localized high and low through fall 508 input.

509 Contrary to expectations (Bouten et al., 1992; Guswa and Spence, 2012; Coenders-Gerrits et al., 2013; 510 Fischer-Bedtke et al., 2023), our results showed that throughfall hotspots do not increase or facilitate 511 greater root water uptake. In addition, the linear mixed effects model results confirmed that throughfall 512 patterns do not drive the variation in root water uptake. We attributed the absence of this to two reasons: 513 (1) decoupled soil water and throughfall patterns, (2) non-water limited conditions.

Regarding (1), we confirmed that the temporally stable throughfall patterns do not correspond to the postevent soil water and root water uptake patterns. We paired the measurements of throughfall and soil water content measurements – and thus the estimates of root water uptake- within a distance of 1 m. The spatial correlation length of soil water content and throughfall is on the order of 6-10 m in natural temperate forests (Keim et al., 2005; Gerrits et al., 2010; Zehe et al., 2010). In the same study site with the spatially extended throughfall sampling, (Fischer-Bedtke et al., (2023) found that the throughfall correlation length 520 increased with decreasing event size, varying from 6.2 m to 9.5 m depending on the size of the rain events. 521 Thus, the paired sampling design in our study likely provided co-located throughfall and soil moisture 522 measurements. Nevertheless, only locations that stored more water than average rarely corresponded with 523 the elevated throughfall input without a significant correlation. <u>HoweverHowever</u>, Hence, variation in soil 524 water storage was not related to throughfall patterns despite temporally persistent local high and low 525 throughfall inputs.

526 S<del>On the one hand, s</del>ome studies, mostly conducted in the arid regions and coniferous forests, reported 527 that soil wetting patterns were not or only partly linked to throughfall variation, despite recurrent 528 throughfall patterns (Raat et al., 2002; Shachnovich et al., 2008; Zhu et al., 2021). Forest floor thickness, 529 horizontal water flow, and soil properties were suggested as reasons for the decoupled patterns. Other<del>On</del> 530 the other hand, some modeling modelling and field studies conducted in temperate deciduous forests found 531 that throughfall patterns influenced soil moisture response to rain event rather than post-event soil water 532 storage variability (Coenders-Gerrits et al., 2013; Metzger et al., 2017; Fischer et al., 2023). In those 533 studies, possible reasons were These studies attributed possible reasons to local processes such as 534 preferential flow due to soil water repellency, the soil pore structure, or elevated root water uptake. Our 535 results support that it is not root water uptake but preferential flow paths that are likely to decouples the 536 throughfall and soil water patterns. In fact, (Fischer-Bedtke et al., (2023) using independent throughfall 537 and soil water content sampling designs, demonstrated that the signature of throughfall patterns dissipated 538 in the post-event soil water variation. However, they However, they detected the stronger influence of 539 throughfall patterns in the soil moisture response to rainfall in the 2015 and 2016 growing seasons. – The 540 temporal variation in soil water content in the 2019 growing season was similar to the seasonal decline in 541 soil water content in 2015 (Metzger et al., 2017). Dry soil conditions can lead to rapid drainage due to 542 reduced water holding capability (Jost et al., 2004; Blume et al., 2009; Wiekenkamp et al., 2016; Demand 543 et al., 2019; Molina et al., 2019) regardless of throughfall amount and its variation. Therefore, our findings 544 support that the localized throughfall input likely-potentially enhances preferential flow because of low 545 soil retention (Fischer-Bedtke et al., 2023) rather than local root water uptake.

546 As a result, the fast flow processes likely dominate how water is stored and transported at our site, erasing 547 the throughfall distribution signature in soil water and root water uptake patterns. Moreover, any shortterm response of uptake to throughfall could not be captured as water uptake was calculated only after 56 hours had elapsed since the last rain event, yet we showed that temporally stable hotspots are not associated with elevated water uptake.- Our Hence, our results also supportare consistent with previous propositions stating that the spatial variation of throughfall affects drainage and subsurface flow (Keim et al., 2006; Blume et al., 2009; Guswa and Spence, 2012), and while root activity activities such as water uptake, root compensation and hydraulic redistribution does not alter canopy-attributed heterogeneity in drainage pathways (Guswa, 2012).

555 The second reason (2) is related to water-limitation conditions. In central Europe, 2019 was the second 556 consecutive extremely dry summer (Boergens et al., 2020), which damaged beech forests (Obladen et al., 557 2021). On average, however, the potential evapotranspiration demand was met at the study site despite 558 the low soil water storage. The ratio of root water uptake to potential evapotranspiration was mostly above 559 65%, which is within the expected range even in the absence of shallow groundwater storage (Nie et al., 560 2021). Hence, local biotic and soil tied abiotic factors determined the spatial variation of root water uptake 561 during growing season rather than throughfall -water input- patterns. However, the discrepancy between 562 daily potential evapotranspiration and root water uptake only increased as the soil in the sampled layers 563 dried out, possibly due to a potential shift in the water uptake depth (see below).

### 564 4. 2) Relative and average soil wetness shapes root water uptake patterns

We found that spatial variation in soil water storage strongly regulates local water uptake such that wetter locations enhance root water uptake. This finding is <u>in lineconsistent</u> with expectations as transpiration rate relies on soil water availability and distribution (Couvreur et al., 2014; Klein et al., 2014; Hildebrandt et al., 2016). Here, <u>our resultswe</u> provide further support that root water uptake is likely to reduce the spatial variability in soil water storage as has been previously suggested (Hopmans and Bristow, 2002; Ivanov et al., 2010; Neumann and Cardon, 2012).

- 571 Trees take up more water in locations where water is not subject to throughfall-driven rapid drainage (see
- 572 above), as a result root water uptake patterns are determined by where water is heldretained longer in the
- 573 soil. Our results support previous studies suggesting that tree transpiration demand is met by water with

574 longer residence time in the soil matrix - passive storage - while groundwater recharge is fed by rapid
575 <u>flow - active storage -(e.g.</u> Evaristo et al., 2019; Sprenger et al., 2019).

576 For a given meteorological condition, root-water uptake at a particular location is a function of water 577 transport resistance between root and soil in addition to the soil-water potential (Cardon and Letey, 1992; 578 Shani and Dudley, 1996; Lhomme, 1998). Both characteristics depend on local soil texture and soil 579 moisture, and the latter in turn is affected by the local rate of uptake. As soil properties we In our statistical 580 analyses, we investigated the soil properties of employed bulk density and field capacity, which are 581 strongly dependent on other soil properties that control aggregation and soil structure. -Although -bulk 582 density is attributed to strongly determined by related to texture, porosity, and soil organic carbon content 583 and strongly related to physical properties like porous pore space, texture, organic matter content and eventually, all of which also affect water retention (Zacharias and Wessolek, 2007; Looy et al., 2017), 584 585 surprisingly the soil bulk density did was not retained as a predictive variable in the optimal model. This 586 indicatesing that not only the bulk density but also those soil physical properties mentioned above that 587 are strongly correlated to it does not influence the relative local water uptake compared to other 588 variables, we surprisingly found that the bulk density of the monitored soil layer did not affect local water uptake. In contrast, the interaction term including field capacity and local soil water storage wasereas 589 590 significant in the model with a negative relationship with relative water uptake, showing that- the combination of higher field capacity (fewer macropores) and low soil water probably hinderedrs the local 591 592 water uptake due to because lower water more being is more strongly bound in the soil water retention. 593 Differences in local soil properties regulate the matric potential at a certain soil wetness. Thus, our result 594 indicates that wetter locations may not always do not necessarily -correspond to the same degree of matric 595 potential and those of easierteasiere root water uptake due to the local differences in the soil water retention characteristics (Vereecken et al., 2007; Cai et al., 2018) for which field capacity serves as a 596 proxy.- -However, our findings suggest that solely soil properties alone were less important (smaller 597 598 effects size of the interaction term including field capacity) than other tested variables factors despite their 599 control on the spatial distribution of soil moisture (Vereecken et al., 2022). and water accessibility for transpiration-(Vereecken et al., 2007; Cai et al., 2018). 600

In addition, the spatial mean of soil water <u>- a measure of overall wetness of the stand - affected influenced</u> root water uptake patterns, yet the effect depended on the basal area, i.e., the size of neighboring trees. We found that as the study site dries out, local water uptake increased in locations with smaller basal areas. Conversely, wetter site conditions facilitate greater water uptake at locations with higher basal areas, i.e., dense clusters <u>of orof</u> large trees. We interpret this as a sign that larger trees are likely to shift their water uptake to deeper soil layers to meet transpiration demands, beyond the monitored soil depth (37 cm), as follows:

Higher basal area <u>is likely to increases transpiration demand and enhances water uptake as long as water</u>
is available<u>. At the same time, Moreover, locations with higher basal area exhaust-the water storage</u>
fastermore rapidly as these locations host larger root structure and root biomass (Le Goff and Ottorini,

611 2001). but At the same time, larger sized trees are able to can shift uptake to deeper layers (Gaines et al.,

612 2016). where soil water content is not measured in our monitoring setup.

613 Beech trees have extensive root systems at shallower depths similar to other temperate tree species, such 614 as European ash and sycamore maple (Kreuzwieser and Gessler, 2010; Brinkmann et al., 2019) -Despite 615 their shallower root system (Leuschner, 2020) in However, in response to declining soil water content in 616 the topsoil, temperate tree species can tap water from the deeper soil layers (Brinkmann et al., 2019; Agee 617 et al., 2021; Seeger and Weiler, 2021) despite their shallow root system (Leuschner, 2020). Recently, 618 Agee et al. (2021) used a three-dimensional water uptake model based on observations in temperate 619 mixed-deciduous forest to show that water uptake is shifted to the deeper soil layers as soil moisture 620 depletes, which is consistent with the field observations. Also Moreover, Krämer and Hölscher (2010) 621 observed in beech and mixed deciduous stands that roots can extract water at depths down to 70 cm soil 622 depth. Similarly, to our site, theirs had a shallow soil layer underlain by weathered limestone, but the soil 623 depth varied between 50 and 120 cm. Later(Brinkmann et al., (2019) also observed similar depth range 624 for beech-trees in a mixed forest by tracing stable water isotopes of soil and xylem water.

Further tree age and size can affect both individual and stand level transpiration because of the different
physiological characteristics and biometrics of trees associated with them (Kostner et al., 2002; Tsuruta
et al., 2023). Within the same species, the larger -presumably older- trees have an advantage in accessing

628 the deeper water storages because of their larger root biomass (Le Goff and Ottorini, 2001) and root

629 plasticity may be able to shift the depth of water uptake while younger trees rely on shallower soil water 630 storages (Dawson, 1996). Our results can be interpreted as tree size, which can be attributed to tree age, 631 affecting root water uptake patterns through differential root biomass development. Furthermore, in the 632 Hainich the coexisting species most likely represent highly coherent rooting depth distribution among 633 trees (Gebauer et al., 2012; Meinen et al., 2009) vet adopt different water uptake strategies (see below). 634 Hence consistent with previous studies focusing on temperate tree species, the linear mixed effect model results indicate that also at our site trees of different sizes response to declining soil water content by 635 636 shifting water uptake depth.

#### **4.3)** Tree species richness regulates root water uptake patterns

638 In addition to the basal area, we included the number of species and number of tree individuals in the 639 linear mixed effects analysis to explore further explore the biotic drivers of root water uptake patterns. 640 While the number of trees was unimportant, the number of species and the basal area, showed a significant 641 interaction effect on the local water uptake. The result indicates that an increase in species richness leads 642 to greater root water uptake, depending on the size and/or density of the neighboring trees: Higher basal 643 area, combined with more species, elevates water uptake. In other words, the interactions among 644 neighboring tree species strongly determine root water uptake patterns, and at-for the same basal area, 645 more water can be taken up in a diverse neighborhood compared tothan in a less diverse 646 neighborhoodlocations.

647 In temperate forests, transpiration has been observed to change with tree species richness at the stand 648 level (Krämer and Hölscher, 2010; Gebauer et al., 2012; Kunert et al., 2012; Meißner et al., 2012; 649 Forrester, 2014). Although some studies indicate a positive relationships between tree diversity and water 650 uptake rate (Forrester et al., 2010; Krämer and Hölscher, 2010; Kunert et al., 2012), tree species diversity 651 is not always positively related to water uptake. While Krämer and Hölscher (2010) observed a positive 652 correlation between water uptake and species richness of the plots in the upper soil layers during soil 653 drying in 2006 at the same study site, Meißner et al. (2012) found no relationship between tree diversity 654 and root water uptake in 2009. They attributed this finding to wetter soil conditions. In contrast, Lübbe et 655 al. (2016) observed a weak effect of diversity on transpiration in wetter soil conditions but not in drier

conditions compared to previous studies (e.g., Pretzsch et al., 2013; del Río et al., 2014). Shortage of 656 657 water can inflate competition mechanisms for water among tree species (González de Andrés et al., 2018; 658 Vitali et al., 2018; Magh et al., 2020). Our results indicate can be used to show that competition between 659 neighboring tree species increases water uptake capacity at more diverse spots (Wambsganss et al., 2021). 660 In addition, different co-existing tree species can facilitate resource uptake or reduce competition, 661 depending on the temporal and spatial availability of the sources, which is often defined as 662 complementarity (Forrester and Bauhus, 2016). As reviewed and listed by Silvertown et al. (2015), 663 several studies suggest that co-existing tree species reduce competition for subsurface water sources by 664 adopting different vertical root water uptake strategies, referred to as hydrological niche partitioning. In 665 addition, trees can transport water from moist wet to dry parts of the soil layers through their roots 666 (Neumann and Cardon, 2012). The mechanism is called hydraulic redistribution or hydraulic lift, which 667 can provide water availability to the shallow roots in drier layers (Burgess et al., 1998; Jonard et al., 668 2011; Hafner et al., 2017; Lee et al., 2018; Rodríguez-Robles et al., 2020; Hafner et al., 2021). In an 669 experiment with six temperate tree species, including the European beech, Hafner et al. (2021) found-in 670 an experiment with six temperate tree species, including the European beech, that the neighboring tree 671 species diversity may not be important for exploiting water uptake through hydraulic redistribution. Both 672 hydraulic niche partitioning and redistribution have been observed vertically, whereas horizontal patterns 673 are largely unexplored in the context of niche partitioning (Hildebrandt, 2020). Our results do not provide 674 direct evidence for either hydraulic redistribution or horizontal niche partitioning. However, they indicate 675 that horizontal root water uptake patterns are regulated by species richness and interactions among 676 neighbouring trees. Thus, we emphasize here the complex interplay between tree species diversity, 677 complementary mechanisms, and water uptake patterns, which is consistent not only with the above-678 mentioned plot-scale studies, but also with larger-scale studies. For instance (Knighton et al., (2019) using 679 the Budyko framework across more than one hundred catchments found that transpiration losses in 680 catchments with deep rooted and mixed species forests differed from those in monoculture catchments. 681 In other words, both plot and catchment scale studies support our results showing that interactions among 682 different coexisting species play a significant role in the spatio-temporal variation of root water uptake. 683

## 684 **5) Conclusion**

685 We investigated the factors that influence the spatial patterns of root water uptake by considering 686 heterogeneity in throughfall and soil water. To that end, we acquired a comprehensive data set based on 687 throughfall measurements paired with soil moisture sensors in a mixed deciduous forest. Soil and 688 neighboring tree characteristics were also included in the linear mixed effects model. We found that 689 variation in root water uptake did not correspond to throughfall consequently rejecting our hypothesis 690 that variation in throughfall is imprinted in water uptake patterns. Wetter soil locations, also poorly 691 associated with higher throughfall, increased local root water uptake. In contrast, how average soil water 692 conditions modified root water uptake depended on the neighborhood basal area. As the site dried out, 693 large trees likely took up water in deeper layers to meet transpiration demands. Furthermore, an increase 694 in species diversity promoted root water uptake, similarly depending on the size of neighboring trees, 695 suggesting active complementarity mechanisms in the forest stand. In conclusion, our results suggest 696 manifest that soil water distribution and neighboring tree characteristics regulate root water uptake 697 patterns more than soil properties and throughfall variation.

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710 Data availability

711 The dataset is currently being prepared for publication in an official repository. The DOI will be published

- 712 with the data at the latest when the data are published.
- 713

## 714 Author contributions

GD and AH designed the throughfall measurement setup, AH and JCM designed soil moisture measurement. GD conducted the field sampling with assistance from JF and the students listed in the Acknowledgments. GD analyzed the data, developed the linear mixed effects model, and analyzed the results with AH and AG. GD prepared the first version of the manuscript, and all authors contributed to discussions and the final version of the manuscript.

### 720 **Competing interests**

721 Anke Hildebrandt is part of the editorial board of HESS. The peer-review process was guided by an

722 independent editor, and the authors have also no other competing interests to declare.

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