



1 Developing Functional Recharge Systems to Repel Saltwater 2 Intrusion via Integrating Physical, Numerical, and Decision-Making 3 Models for Coastal Aquifer Sustainability

4 Yehia Miky¹, Usama Hamed Issa^{2,3}, Wael Elham Mahmood^{3,4}

5 ¹Department of Geomatics, Faculty of Architecture and planning, King Abdulaziz University, Jeddah, Saudi Arabia;
6 yhassan@kau.edu.sa

7 ²Department of civil engineering, Faculty of Engineering, Minia University, Egypt; usama.issa@mu.edu.eg

8 ³Department of civil engineering, college of Engineering, Taif University, Saudi Arabia; wemahmod@tu.edu.sa;
9 +966543656209

10 ⁴Civil Engineering Department, Faculty of Engineering, Assiut University, 71515 Assiut, Egypt; wdpp2006@aun.edu.eg; +20-
11 01006328492,
12 orcid.org/0000-0003-4340-0525

13 *Correspondence to:* Wael Elham Mahmood (wdpp2006@aun.edu.eg)

14 **Abstract.** Controlling the hydraulic heads along the coastal aquifer may help to effectively manage saltwater intrusion,
15 improve the conventional barrier's countermeasure, and ensure the coastal aquifer's long-term viability. This study proposed a
16 framework that utilizes a decision-making model (DMM) by incorporating the results of two other models (physical and
17 numerical) to determine proper countermeasure components. The physical model is developed to analyze the behavior of
18 saltwater intrusion in unconfined coastal aquifers by conducting two experiments: one for the base case and one for the
19 traditional vertical barrier. MODFLOW is used to create a numerical model for the same aquifer, and experimental data is
20 used to calibrate and validate it. Three countermeasure combinations, including vertical barrier, surface, and subsurface
21 recharges, are numerically investigated using three model case categories. Category (a) model cases investigate the hydraulic
22 head's variation along the aquifer to determine the best recharge location. Under categories (b) and (c), the effects of surface
23 and subsurface recharges are studied separately or in conjunction with a vertical barrier. As a pre-set of the DMM, evaluation
24 and classification ratios are created from the physical and numerical models, respectively. The evaluation ratios are used to
25 characterize the model cases results, while the classification ratios are used to classify each model case as best or worst. An
26 analytic hierarchy process (AHP) as DMM is built using the classification ratios of hydraulic head (HHR), salt line (SLR),
27 intrusion (IR), repulsion (Rr), wedge area (WAR), and recharge (RER) as selection criteria to select the overall best model
28 case. The optimal recharging location, according to the results, is in the length ratio (LR) range from 0.45 to 0.55. Furthermore,
29 the DMM supports case3b (vertical barrier + surface recharge) as the best model case to use, with a support percentage of
30 47.93%, implying that this case has a good numerical model classification with a minimum IR of 67.9%, a maximum Rr of
31 29.4%, and an acceptable WAR of 1.25. The proposed framework could be used in various case studies under different
32 conditions to assist decision-makers in evaluating and controlling saltwater intrusion in coastal aquifers.

33 **Keywords:** Saltwater intrusion · Hydraulic heads · Unconfined coastal aquifer · Vertical barrier · Surface recharge · subsurface
34 recharge · Decision making · AHP

35 1 Introduction

36 Due to the natural effects of long-term climate change such as sea level change and tidal intensity fluctuations, seawater flows
37 toward the freshwater aquifers. In addition, increased water demands accompanied by anthropogenic activities such as
38 excessive pumping of freshwater in coastal areas cause the lowering of water tables as well as saltwater intrusion (Abd-Elaty
39 et al. 2019; Sutar and Rotte 2022). Saltwater intrusion lowers the potentiality and quality of freshwater in coastal regions, as
40 reported at many locations all over the world (Qi and Qiu 2011; Shi and Jiao 2014; Anders et al. 2014; Cary et al. 2015;



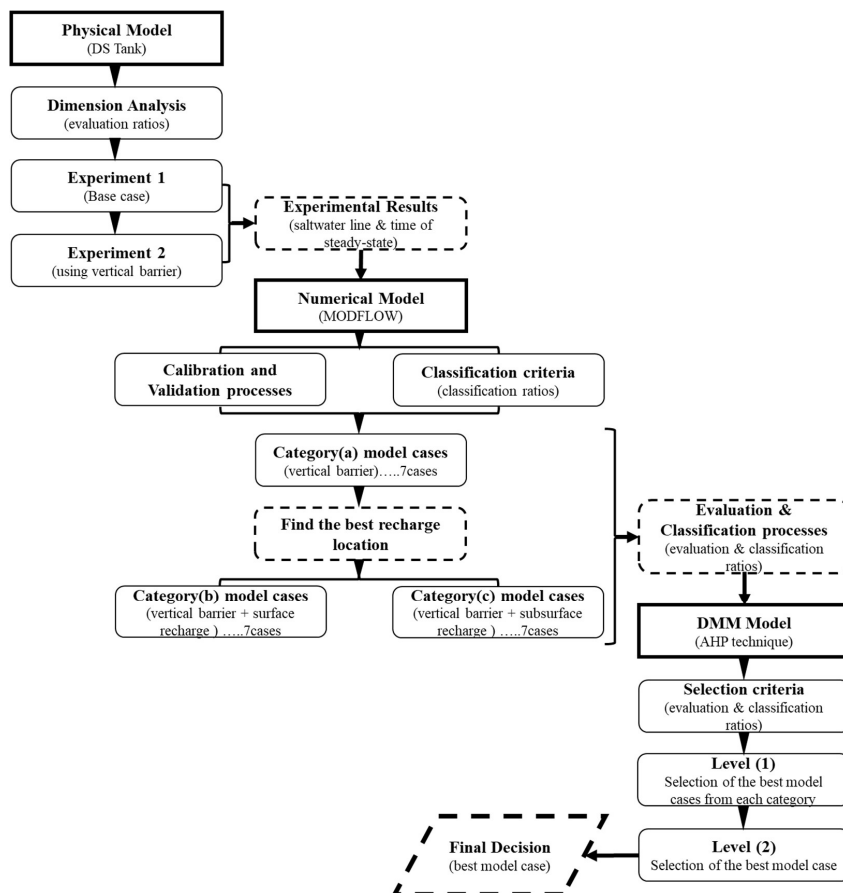
41 Srinivasamoorthy 2015; Abd-Elhamid 2016; Eissa et al. 2018; Abd-Elhamid et al. 2019; Pramada et al. 2021). Therefore, it is
42 important to control saltwater intrusion with efficient countermeasures to achieve sustainable freshwater sources.
43 Traditional methods for controlling saltwater intrusion include reducing pumping rates, relocating pumping wells, changing
44 pumping patterns, constructing physical subsurface barriers, and saltwater abstraction (Abd-Elhamid and Javadi 2011;
45 Kallioras et al. 2013; Cai et al. 2015; Huang and Chiu 2018; Abd-Elhamid et al. 2019; Hussain et al. 2019). The limitations
46 and high costs of the aforementioned methods are substantial challenges to their implementation..
47 Artificial recharge techniques, such as surface and subsurface recharge systems, are critical for establishing hydraulic barriers
48 and mitigating the effects of saltwater intrusion. These techniques have several advantages compared to traditional methods,
49 including low cost, no inundation storage space, less water evaporation, and improved water quality. Although artificial
50 recharge has numerous advantages, it also has disadvantages, including groundwater contamination from surface water,
51 difficulty in implementation due to a lack of understanding of aquifer hydrogeological properties, the potential for
52 environmental damage and soil disturbance, and high maintenance costs. Surface recharge systems include ditches and
53 furrows, recharge basins, stream augmentation, and runoff conservation structures (terracing, contour bunds, percolation tanks,
54 gully plugs, Nalah bunds, and check dams) (Maliva 2020b, c; ASCE 2001). On the other hand, subsurface recharge systems
55 include subsurface injection wells, borewells, and recharging pits and shafts (Maliva 2020a, d; ASCE 2001). Combining
56 traditional and artificial recharge techniques is one way to overcome the disadvantages of both. Although many studies
57 investigate saltwater intrusion in coastal aquifers, only a limited number study the control methods of saltwater intrusion.
58 Physical and numerical models have not only proven to be more effective and economic tools for selecting the optimum
59 solutions for repelling saltwater intrusion but can be used to reduce the high cost of hydrogeological and environmental
60 investigations before constructing a full-scale project (Mantoglou 2003; Zhou, et al. 2003; Abarca et al. 2006; Singh 2015;
61 Abd-Elaty et al. 2019; Guo et al. 2019; M Armanuos et al. 2019).
62 Although physical and numerical models are effective economic tools for selecting the best solutions for repelling saltwater
63 intrusion, deficiencies in the acquisition of appropriate evidence to support the final decision are discovered. It is necessary to
64 use decision models in conjunction with physical and numerical models to guide stakeholders toward sustainable resource
65 management based on a set of criteria. The analytical hierarchy process (AHP) is a decision-making method that has been used
66 alone or in conjunction with other techniques such as GIS and fuzzy logic in a variety of groundwater-related fields. Based on
67 a broader set of criteria, this technique is used to guide stakeholders involved in groundwater development and sustainable
68 resource management (Vaidya and Kumar 2006; Alwetaishi et al. 2017). The applications of AHP in the field of groundwater
69 include assessing groundwater vulnerability by developing indices based on hydrogeological parameters and mapping
70 groundwater potential zones (Arunbose et al. 2021; Osiakwan et al. 2022; Ahmadi et al. 2021; Castillo et al. 2022; Achu et al.
71 2020; Sajil Kumar et al. 2022; Nithya et al. 2019; Phin et al., 2022; Zghibi et al., 2020; Mallick et al., 2019; Shao et al., 2020).
72 In the field of saltwater intrusion, a GIS-based AHP weighted index overlay analysis technique has been demonstrated to
73 determine the distribution of groundwater vulnerability (Gangadharan, Nila, et al. 2016; Güllü and Kavurmacı 2023). A fuzzy-
74 AHP evaluation model is developed for analyzing the level of seawater intrusion in long-term monitoring data from multiple
75 river basins (Yang et al., 2022). The AHP is also used to compute weights for the GALDIT parameters, which are used to
76 assess the vulnerability of coastal aquifers to saltwater intrusion (Pham et al., 2022).
77 According to the preceding overview, both traditional and artificial techniques of repelling seawater intrusion have limitations,
78 and using physical, numerical, and decision-making models is crucial. The unconfined coastal aquifer is investigated in this
79 work, and the surface and subsurface recharge methods, either alone or in conjunction with typical vertical barriers, are
80 analyzed by integrating physical, numerical, and decision-making models. On the other hand, the behaviors of saltwater
81 intrusion, groundwater flow, and hydraulic head are numerically investigated using three categories of model cases: category
82 (a), (b), and (c). The aims of this study are: (i) to examine experimentally the behavior of saltwater intrusion via coastal
83 unconfined aquifers with and without vertical barrier countermeasures; (ii) to develop a validated numerical model regarding



84 the experimental findings of transitory saltwater intrusion; (iii) to numerically analyze the flow behavior of saltwater and
85 freshwater, therefore identifying the saltwater-freshwater interaction via porous media; (iv) to identify the optimal recharging
86 location utilizing the location of the minimum hydraulic head; (v) to determine the optimal vertical barrier depth for saltwater
87 intrusion management; (vi) to identify the components of an effective countermeasure system, such as a vertical barrier, surface
88 recharge, and subsurface recharge, either alone or in combination; (vii) to develop a DMM model to aid decision makers in
89 the selection among several saltwater countermeasures and picking the most appropriate one depending on various demanding
90 scenarios.

91 **2 Materials and Methodologies**

92 Saltwater intrusion is investigated experimentally in part of this study by developing a laboratory physical model of an
93 unconfined coastal aquifer. Two experiments are carried out in this part, and ratios are formed using the dimension analysis
94 method, namely as evaluation ratios. These evaluation ratios are used to analyze and characterize the saltwater line and
95 hydraulic head variations of the numerical model cases, as forthcoming later. A numerical finite difference model is created,
96 and the validation and calibration processes are carried out using the experimental results. Following that, the numerical
97 repelling of saltwater intrusion is investigated, taking into account the combined effect of using vertical barriers with surface
98 or subsurface recharging systems, as depicted by model cases divided into three categories (a, b, and c) (seven cases in each
99 category). The results of the category (a) model cases reveal the location of the minimal hydraulic heads, which are expected
100 to be the locations of the indicated artificial recharge systems. A classification process is then implemented to classify model
101 cases in each category as best or worst model case using a developed set of ratios, namely classification ratios. Because each
102 model case is expected to have benefits and drawbacks, as well as several criteria governing the model cases, the benefits and
103 drawbacks of each model case should be quantified in order to identify the most effective one. Following that, the most
104 effective model case is decided on using a new DMM model based on the AHP technique. To make the final decision, two
105 selection levels (levels 1 and 2) are considered. **Figure 1** illustrates a flow chart for the framework of the study.



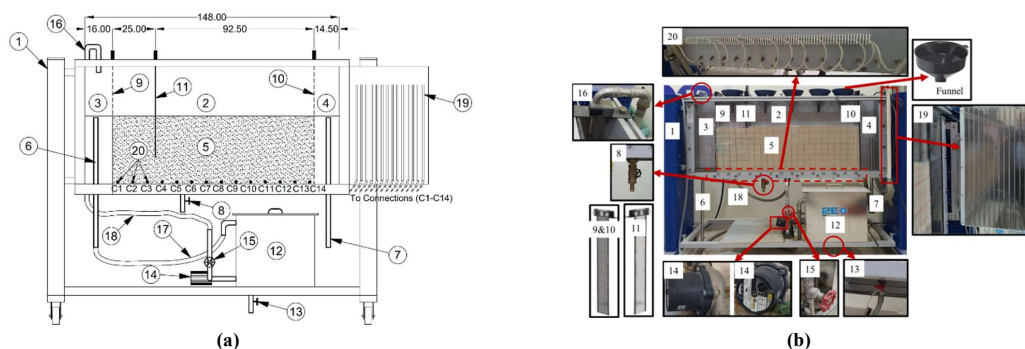
106

107 **Figure 1: Flow chart shows the proposed framework to identify most effective model case using the physical, numerical, and**
 108 **decision-making models**

109 **2.1 Experimental Setup**

110 **2.1.1 Drainage and Seepage Tank (DS Tank)**

111 The DS Tank is used in this study to visualize groundwater flow through permeable porous media. The model of the DS Tank
 112 that is used in the current study is HM 169 GUNT HAMBURG. The DS Tank consists of a porous media container, a lower
 113 water tank as a water source, a pump for the water flow, a valve to adjust the water supply, and measuring connections in the
 114 experiment section, which are connected to 14 glass tube manometers to display and measure hydraulic heads along the DS
 115 Tank. The sand container consists of an aluminum rectangular tank with a transparent front side (methacrylate material) to
 116 visualize groundwater flow and optimize observation of the experiments through the porous media. In the DS Tank, two fine
 117 mesh screens are used to create feed and discharge chambers and to separate the experimental section from these two chambers.
 118 There are two adjustable overflow pipes in the DS Tank for adjusting the water levels in the mentioned chambers and
 119 measuring the water flow. To prevent seawater intrusion, an aluminum sheet pile is used as a vertical barrier. As a result, the
 120 DS Tank has a closed water circuit with a storage tank and pump. The DS Tank and its components are depicted in **Figure 2**
 121 **and Table 1.**



122
 123

124 **Figure 2: DS Tank and its components: (a) Drawing of details, (b) Photo**

125 **Table 1: DS Tank components and descriptions**

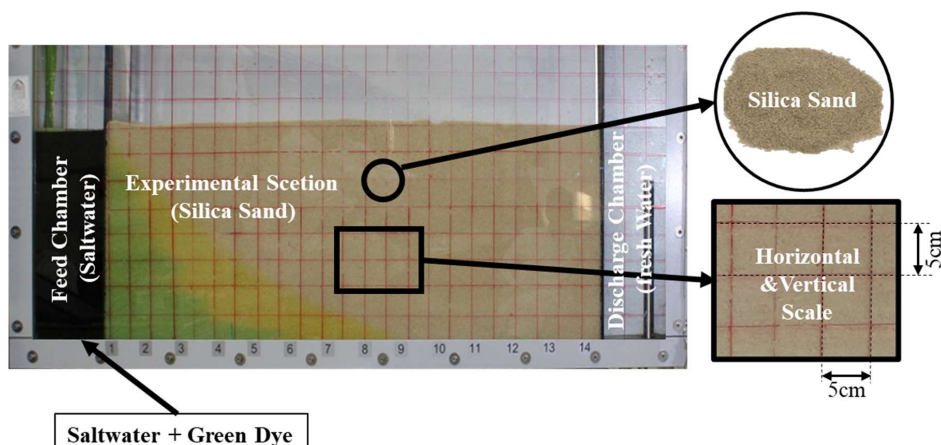
No.	Component Name	Description	No.	Component Name	Description
1	Steel frame	The DS Tank's frame	11	Vertical aluminum sheet pile	Vertical barrier to repel saltwater intrusion
2	Experimental section	Tank with porous media for monitoring saltwater intrusion	12	Storage tank	The primary source of seawater
3	Feed Chamber	Source of saltwater	13	Draining pipe2	Before the next experiment, drain the saltwater from the storage tank.
4	Discharge Chamber	Source of freshwater	14	Pump	Pumping saltwater to the feed chamber
5	Porous media	Silica sand (0.71-1.18mm)	15	Pump valve	Pump flow rate adjustment
6	Outflow pipe1	Changing the saltwater level in the feed chamber	16	Saltwater inflow pipe	Connecting with a pump to allow saltwater to flow from the pump to the feed chamber
7	Outflow pipe2	Changing the level of freshwater in the discharge chamber	17	Hose1	Connecting the outflow pipe 1 to the storage tank
8	Draining pipe1	Before beginning a new experiment, drain the water from the experimental section.	18	Hose2	Linking the saltwater inflow pipe to the pump
9	Vertical screen1	Separating the feed chamber from the experimental section	19	14 glass manometer tubes	Hydraulic head monitoring along the experimental section
10	Vertical screen2	Separating the discharge chamber from the experimental section	20	Measuring connections	linked to the 14 glass manometer tubes

126 **2.1.2 Configuration and Experimental Set**

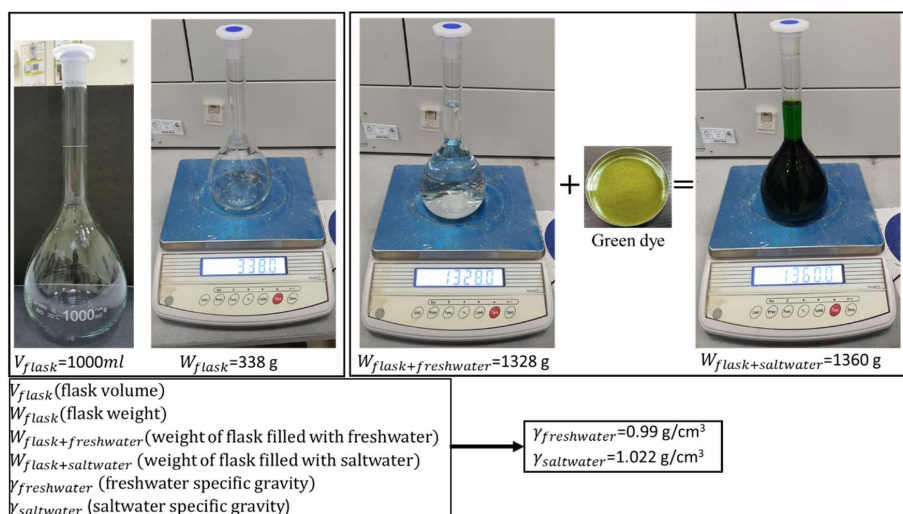
127 The DS Tank and the associated materials, including saltwater, freshwater, and porous media, are pre-set for the experiments.
 128 A horizontal and vertical scale of 5cm x 5cm is drawn on the transparent front side of the DS Tank, as shown in **Figure 3**. The
 129 left chamber is configured as a saltwater feed chamber with a width of 16cm. The right chamber is configured as a freshwater
 130 discharge chamber with a width of 14.5cm. Vertical screen barriers separate the experimental section of the DS Tank (length
 131 117.5cm) from the feed and discharge chambers. The experimental section is filled to a depth of 40cm with porous media soil
 132 (graded silica sand with grain sizes ranging from 0.71-1.18mm (see **Figure 3**). The filling process is done in layers of 5cm
 133 each, with a falling height of 50 cm for each layer, to ensure a homogeneous hydrogeological property of the media sand. In
 134 the filling process, funnels are used, which are distributed along the experimental section as shown in **Figure 1b**.
 135 The seawater used in the experiments is collected from the Red Sea, and its density, as well as that of the freshwater, is
 136 calibrated using a sensitive scale and a standard flask (see **Figure 4**). According to the calibration, the densities of saltwater



137 and freshwater are 0.99 and 1.022 m³/sec, respectively. In saltwater, a 0.15 g/L concentration of green food dye is used to
 138 easily visualize the saltwater line and measure the intrusion distance inside the media sand (see **Figure 3**).



139
 140 **Figure 3: DS Tank pre-set for experimental procedures**



141
 142
 143 **Figure 4: Saltwater and freshwater calibration**

144 **2.1.3 Experimental Procedures**

145 The experiment procedures include the following five steps:

146 1-Freshwater saturation of the media sand: at the start of the experiment, the outflow pipes 1 and 2 for both the feed and
 147 discharge chambers are set to be at the same level as the media sand surface (40 cm from the DS Tank bed). Following that,
 148 fresh water is discharged at a constant rate into both chambers until the media sand in the experimental section is saturated.
 149 The hydraulic heads along the experimental section are monitored by the 14 glass tube manometers until the water level reaches
 150 the sand surface in all the manometers to verify the saturation condition.

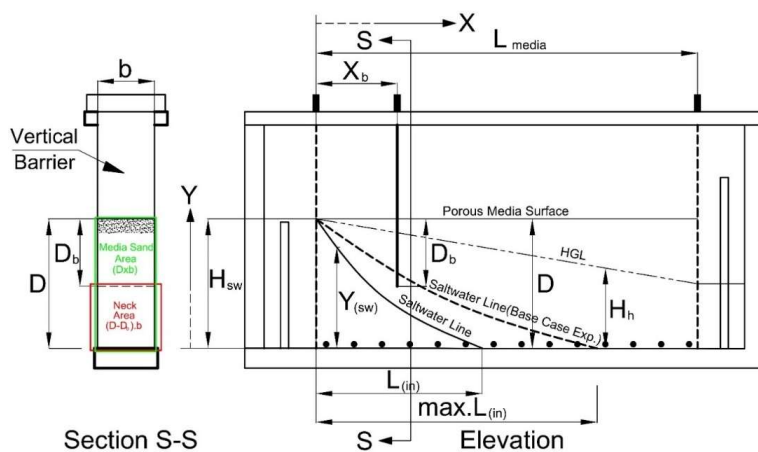
151 2-Feeding the experiment with colored saltwater: in the feed chamber, an aluminum sheet pile is used to block water seepage
 152 through the experimental section. Following that, the feed chamber's outflow pipe 1 is moved to the DS Tank bed level to
 153 empty it of freshwater. The outflow pipe is then returned to its previous level (media sand surface level), and the storage tank
 154 is subsequently emptied and filled with the green-dyed saltwater. When the pump is turned on and the pump valve is opened,



155 saltwater begins to fill the feed chamber all the way to the top of the outflow pipe1. Following that, the pump valve is manually
 156 adjusted to maintain the saltwater level at the surface of the media sand.
 157 3-Adjusting the water levels in the feed and discharge chambers: the first step in this process is to remove the aluminum sheet
 158 pile from the feed chamber. Furthermore, to achieve a suitable flow through the media sand, the difference in water levels
 159 between the feed and discharge chambers is tested several times and finally adjusted to 10 cm, resulting in a hydraulic gradient
 160 of 0.085. To accomplish this, the outflow pipe 2 for the discharge chamber is adjusted to be 10 cm below the media sand
 161 surface.
 162 4-Monitoring of saltwater intrusion: in the experimental section, saltwater begins to infiltrate through the media sand and can
 163 be observed through the transparent front side of the DS Tank. The temporal saltwater intrusion could be measured using the
 164 horizontal and vertical scales drawn on the transparent front side. The saltwater intrusion is measured at 30-minute intervals.
 165 Photos for each time interval are taken with a high-resolution digital camera and used to validate the observed saltwater lines
 166 with AutoCAD software. During the experiment, the freshwater level inside the discharge chamber rises until it reaches its
 167 maximum level by adjusting the outflow pipe2 level above the media sand surface level until it reaches a steady state.
 168 This experimental part of the study considers two experiments:
 169 Experiment 1 (Base Case): this is the case in which the saltwater intrusion through the media sand is studied without any
 170 countermeasures. In this case, the procedures from steps 1 to 4 are carried out.
 171 Experiment 2 (using a vertical barrier): through this experiment, the media sand is removed from the experimental section.
 172 Then, the vertical aluminum sheet pile (vertical barrier) is used as a countermeasure against saltwater intrusion and placed at
 173 the experimental section, 25cm from the feed chamber. Hereafter, media sand is refilled in the experimental section. The
 174 penetration depth of the vertical aluminum sheet pile is set below the silica sand surface by a depth of 30cm. Then, the steps
 175 from 1 to 4 are implemented.

176 **2.2 Dimension Analysis and Evaluation Ratios**

177 **Figure 5** illustrates the geometric properties of the experiments described in the previous section. **Table 2** summarizes and
 178 defines several variables, parameters, and constants that influence saltwater intrusion.



179 **Figure 5: Geometric characteristics of the experiments**
 180

181 **Table 2: Definition of the geometric characteristics of the experiments**

No.	Quantity	Type			Definition
		Constant	Parameter	Variable	
1	H_{sw}	✓			Head of the saltwater boundary



2	D	√			Sand media depth
3	X_b	√			Vertical barrier location
4	L_{media}	√			Sand media length (experimental section length)
5	$max.L_{(in)}$	√			Maximum length of saltwater intrusion (attained for experiment 1 (base case))
6	A	√			The cross sectional area of the entire media sand
7	a		√		Neck region (cross section area of media sand below the vertical barrier)
8	D_b		√		Aluminum sheet pile depth
9	X			√	Horizontal distance from the saltwater boundary measured from any embedded point in the media sand
10	Y			√	Vertical distance measured from the experimental section bed for any embedded point in the media sand
11	$Y_{(sw)}$			√	Observed saltwater intrusion depth at any X distance at a specific time (t).
12	H_h			√	Observed hydraulic head at any X distance at a specific time (t).
13	$L_{(in)}$			√	The observed length of saltwater intrusion at a specific time (t)

182
 183 Excluding the constant quantities (H_{sw} , D , X_b , L_{media} , $max.L_{(in)}$, A), the function of dimensionless hydraulic and geometric
 184 parameters is represented in **Eq. 1**.

185 $L_{(in)}/max.L_{(in)} = f(Y_{(sw)}/H_{sw}, D_b/H_{sw}, H_h/H_{sw}, a/A, X/L_{media}, Y/D)$ (1)

186 **Table 3** represents the dimensionless quantities which will be considered in this study as evaluation ratios.

187 **Table 3: Evaluation Ratios**

Evaluation Ratios	Definition
$L_{(in)}/max.L_{(in)}$	Intrusion Ratio (IR) : the ratio of observed intrusion length at a time (t) to maximum saltwater intrusion length (base case).
$Y_{(sw)}/H_{sw}$	Salt Line Ratio (SLR) : the ratio of the depth of the saltwater line at a specific distance X to the depth of the saltwater boundary.
D_b/H_{sw}	Barrier Depth Ratio (BDR) : the ratio of barrier depth to the depth of the saltwater boundary.
H_h/H_{sw}	Hydraulic Head Ratio (HHR) : the ratio of the depth of the hydraulic grade line at a specific distance X to the depth of the saltwater boundary..
a/A	Neck Area Ratio (NAR) : the ratio of neck area to media sand cross section area.
X/L_{media}	Length Ratio (LR) : the ratio of horizontal distance along the experimental section to the length of media sand.
Y/D	Depth Ratio (DR) : the ratio of the vertical distance Y measured from the bed of the experimental section to the total media sand depth.

188 **2.3 Numerical Model**

189 MODFLOW-2005, in conjunction with the SWI2 package, is used in this study for numerical modeling of saltwater intrusion.
 190 SWI2 is a software package used to analyze three-dimensional groundwater flow, model saltwater intrusion, and calculate
 191 hydraulic heads. The main advantage of using the SWI2 package is that it requires fewer cells for the simulation process than
 192 variable-density groundwater flow packages like SEAWAT. The ability of SWI2 to represent each aquifer as a single layer of
 193 cells results in significant model run-time savings.

194 Saltwater intrusion is investigated using either a traditional vertical barrier or artificial recharge methods. Various penetration
 195 depths are simulated to control saltwater intrusion using a vertical barrier. As artificial recharge methods, surface and
 196 subsurface recharge systems are used. The above management options are evaluated separately and in combination to
 197 determine their effectiveness in addressing the saltwater intrusion problem. **Table 4** displays the numerical model cases that
 198 are investigated.

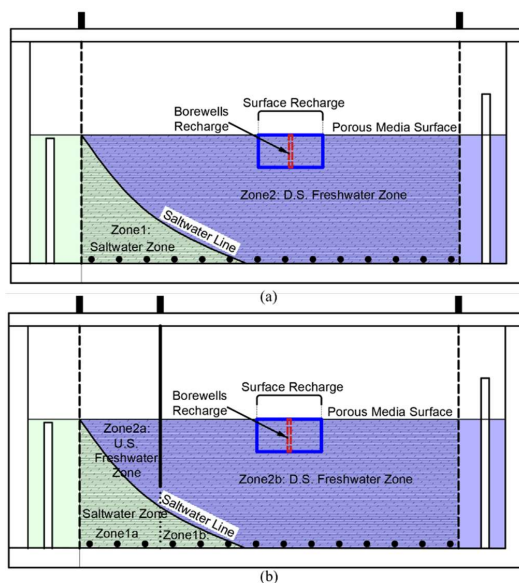


199 MODPATH is a post-processing package for particle tracking that computes and displays three-dimensional pathlines based
 200 on MODFLOW output. The MODPATH package is used to visualize the flow behavior of both freshwater and saltwater
 201 through the sand media by visualizing the expected transport trajectories coming from the saltwater boundary, the freshwater
 202 boundary, and the flow path from the recharge area for the cases defined in **Table 4**. When there is no vertical barrier, two
 203 water zones can be identified: zone 1 (saltwater zone) and zone 2 (freshwater zone) (see **Figure 6a**). Zones 1 and 2 are further
 204 subdivided into two zones after using a vertical barrier: zone 1a and zone 1b for saltwater and zone 2a and zone 2b for
 205 freshwater, as shown in **Figure 6b**.

206 **Table 4: The studied cases using numerical simulation**

Model Cases	Description
Category (a): using vertical barrier	
Case1a	Base Case (Verification of experiment1)
Case2a	Vertical Barrier ($X_b=25\text{cm}$, $D_b=35\text{cm}$, $NAR=0.125$, $BDR=0.875$)
Case3a	Vertical Barrier (Verification of experiment2) ($X_b=25\text{cm}$, $D_b=30\text{cm}$, $NAR=0.25$, $BDR=0.75$)
Case4a	Vertical Barrier ($X_b=25\text{cm}$, $D_b=25\text{cm}$, $NAR=0.375$, $BDR=0.625$)
Case5a	Vertical Barrier ($X_b=25\text{cm}$, $D_b=20\text{cm}$, $NAR=0.50$, $BDR=0.50$)
Case6a	Vertical Barrier ($X_b=25\text{cm}$, $D_b=15\text{cm}$, $NAR=0.625$, $BDR=0.375$)
Case7a	Vertical Barrier ($X_b=25\text{cm}$, $D_b=10\text{cm}$, $NAR=0.875$, $BDR=0.125$)
Category (b): using vertical barrier and surface recharge	
Case1b	Base Case + Surface Recharge
Case2b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=35\text{cm}$, $NAR=0.125$, $BDR=0.875$) + Surface Recharge
Case3b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=30\text{cm}$, $NAR=0.25$, $BDR=0.75$) + Surface Recharge
Case4b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=25\text{cm}$, $NAR=0.375$, $BDR=0.625$) + Surface Recharge
Case5b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=20\text{cm}$, $NAR=0.50$, $BDR=0.50$) + Surface Recharge
Case6b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=15\text{cm}$, $NAR=0.625$, $BDR=0.375$) + Surface Recharge
Case7b	Vertical Barrier ($X_b=25\text{cm}$, $D_b=10\text{cm}$, $NAR=0.875$, $BDR=0.125$) + Surface Recharge
Category (c): using vertical barrier and subsurface recharge	
Case1c	Base Case + Subsurface Recharge
Case2c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=35\text{cm}$, $NAR=0.125$, $BDR=0.875$) + borewells Recharge
Case3c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=30\text{cm}$, $NAR=0.25$, $BDR=0.75$) + borewells Recharge
Case4c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=25\text{cm}$, $NAR=0.375$, $BDR=0.625$) + borewells Recharge
Case5c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=20\text{cm}$, $NAR=0.50$, $BDR=0.50$) + borewells Recharge
Case6c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=15\text{cm}$, $NAR=0.625$, $BDR=0.375$) + borewells Recharge
Case7c	Vertical Barrier ($X_b=25\text{cm}$, $D_b=10\text{cm}$, $NAR=0.875$, $BDR=0.125$) + borewells Recharge

207



208
 209 **Figure 6: Saltwater intrusion conceptual model: (a) freshwater and saltwater zones without barrier, (b) freshwater and saltwater**
 210 **zones with barrier**

211 **2.3.1 Calibration and Verification Processes**

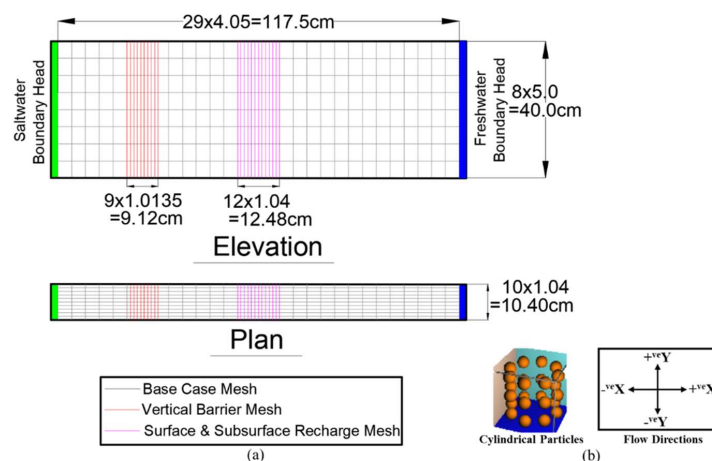
212 Many factors contribute to groundwater model inconsistency, including hydrogeological properties, discretization, potentially
 213 spatial discretization, time step, and solver parameters. Using the experimental results, many trials are carried out to calibrate
 214 the model using various hydrogeological properties, with reference to the values listed in **Table 5**. Furthermore, various
 215 discretization schemes are tested to provide an accurate diagnosis of differences in head drawdowns and water balances. **Figure**
 216 **7a** illustrates that 2320 cell discretization are used in this study. The transient stress period, on the other hand, is assigned to
 217 be 120 minutes based on the experiment's monitored results, and the suitable equal interval time step is selected to be 30
 218 minutes. The impact on the heads on the cells and the accumulated volume water balance are evaluated. Following that, a
 219 verification procedure is implemented for:

- 220 1- Confirming the time when a steady-state condition occurs in based on the results of experiment 1.
- 221 2- Fitting the observed saltwater line in experiments 1 and 2 for the transient and steady-state conditions.

222 The particle tracking in the MODPATH package is simulated in the forward tracking direction using cylinder particle
 223 placement, as shown in **Figure 7b**. in this study, the flow direction will be defined as $+^{ve}Y$, $-^{ve}Y$, $+^{ve}X$, and $-^{ve}X$, as illustrated
 224 in **Figure 7b**.

225 **Table 5: Hydrogeological values of sand (Domenico et al. 1998; Rotz 2021)**

Soil Type	Effective Porosity	Specific Yield	Hydraulic Conductivity "K" (m/s)
Coarse Sand	from 0.18 to 0.43	from 0.18 to 0.43	from 9×10^{-7} to 6×10^{-3}
Medium Sand	from 0.16 to 0.46	from 0.16 to 0.46	from 9×10^{-7} to 5×10^{-4}
Fine Sand	from 0.01 to 0.46	from 0.01 to 0.46	from 2×10^{-7} to 2×10^{-4}



226
 227 **Figure 7: Structure of the numerical model: (a) discretization and boundary conditions, (b) particle tracking and flow directions**

228 2.3.2 Classification Ratios

229 As a pre-set to select the best model case for repelling saltwater intrusion, four ratios are proposed to analyze the results of the
 230 model cases included in categories (a), (b), and (c). Each ratio is calculated for each model case and then classified by its value
 231 into best or worst. These ratios are the previously defined intrusion ratio (IR), as well as three new ratios: repulsion ratio (Rr),
 232 wedge area ratio (WAR), and recharge ratio (RER). The last three ratios are computed using the **equations 2, 3, and 4**,
 233 respectively with the RER ratio computed only for cases in categories (b) and (c). The criteria for classifying the best model
 234 cases are that it has low values of IR, WAR, and RER, as well as the maximum value of Rr. on the other hand, cases with high
 235 values of IR, WAR, and RER, as well as the lowest value of Rr, are classified as worst model case and is not recommended
 236 for repelling saltwater intrusion. Because of the difficulty of having one of the model cases have all the best or worst values
 237 of classification ratios to be classified as the best or worst model case (unclassified model case), it is important to use the
 238 DMM models to use the values of these classification ratios to make this final decision.

$$239 R_r = IR_{case1a} - IR_{case1} \quad (2)$$

$$240 WAR = \frac{Wedge\ Area_{case(i)}}{Wedge\ Area_{case1}} \quad (3)$$

$$241 RER = \frac{Recharge_{case(j)}}{Saltwater\ boundary\ Recharge_{case(j)}} \quad (4)$$

242 Where case(i) is any case included at any category (a, b, and c), and case(j) is the cases included at category (b) and category
 243 (c).

244 2.4 Decision-Making Model (AHP technique)

245 The AHP technique is commonly employed in decision-making systems designed to aid in decision-making and rate options
 246 (Saaty, 1986). Actual metrics such as pricing, headcount, or subjective opinions are used as inputs into a numerical matrix in
 247 AHP. Ratio scales and consistency indices derived from eigenvalues and eigenvectors are among the results. The AHP model
 248 is a decision-making framework that assumes decision levels have a unidirectional hierarchical relationship (Presley, 2006).
 249 AHP can study the interrelationships among all criteria using the hierarchical approach (Singh et al., 2007).

250 In this study, an AHP-based model is proposed to be employed on two levels to determine the best model case by comparing
 251 18 model cases using numerous ratios as a selection criteria. Level (1) involves the model dealing with three categories (a, b,
 252 and c) in order to choose the best model case among the six cases in each category. There are four criteria in case a (Rr, SLR,
 253 HHR, and WAR), while there are five in cases b and c. (Rr, SLR, HHR, WAR, and RER). The top three model cases from
 254 each of the three categories that emerged from level (1) can be used to create the final choice for the best model case at level

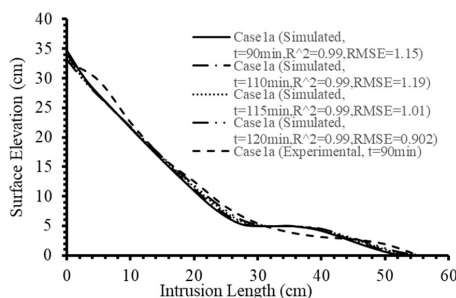


255 (2). Pairwise comparisons with other criteria aid in determining the relative importance of each criterion in the hierarchical
256 structuring of the problem. The model's first level consists of one matrix (5x5) reflecting the relative weights of the criteria
257 and five matrices (6x6) showing the relative weight among the alternatives in the case of each criterion. The model, on the
258 other hand, takes the same matrix for criteria weights and five matrices, each of which is (3x3) and expresses the relative
259 weight among the final three alternatives for each criterion in its second level.
260 The evaluations are carried out using a preference scale ranging from 1 (which represents "equally important") to 9 (which
261 represents "extremely important") (Saaty, 1986). The consistency ratio (CR), which is a function of the consistency index (CI)
262 and the relative importance (RI), is a measure of cognitive effort in the decision, and its value should not exceed 10%
263 (Abdelwahab et al., 2021).

264 3 Results and Discussions

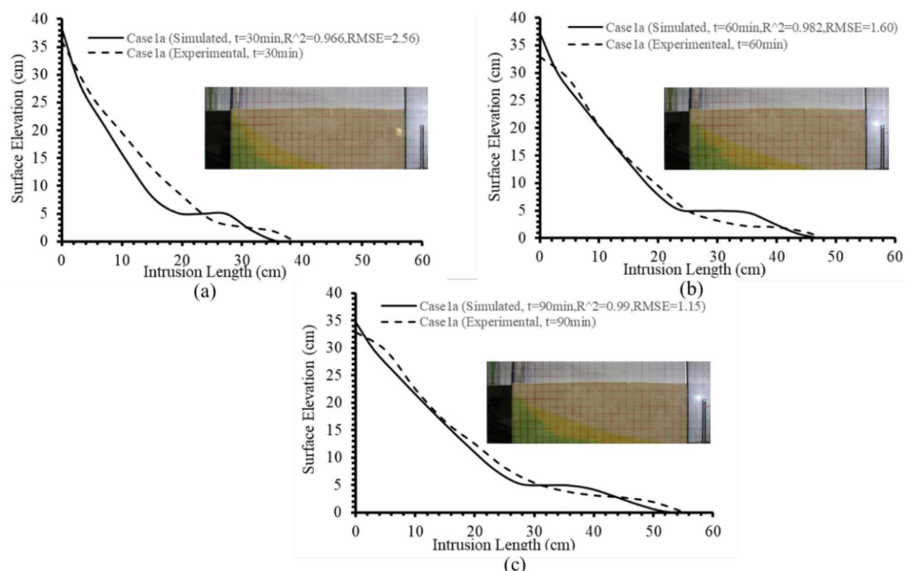
265 3.1 Calibration and Verification of Numerical Model

266 The steady-state condition in experiment 1 (the base case) results in saltwater intrusion 90 minutes after the experiment begins.
267 As a time validation of the numerical model steady state simulation, **Figure 8** shows the observed and simulated saltwater
268 lines for various simulation times greater than 90 minutes. The figure shows that the simulated saltwater line closely matches
269 the observed one, with RMSE values ranging from 0.90 to 1.19 for time ranging from 90 to 120 minutes.

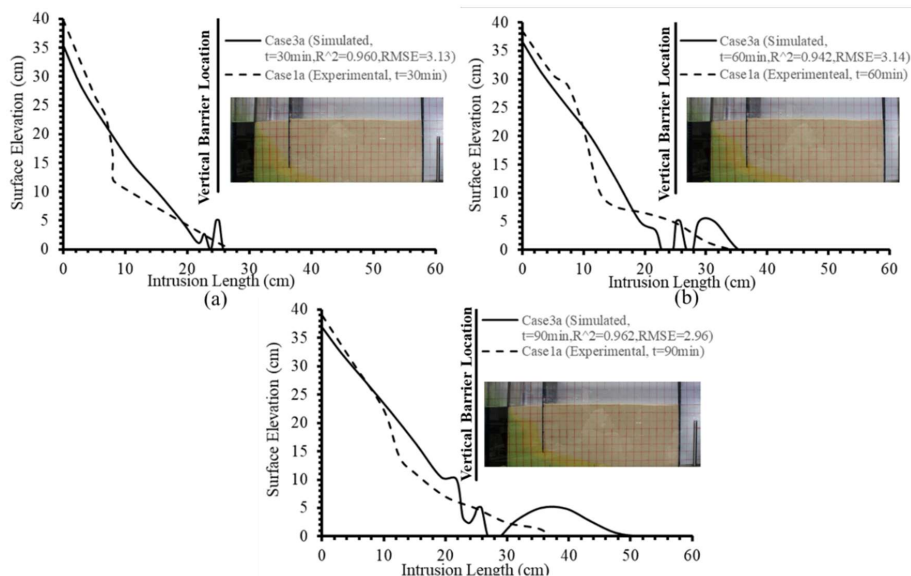


270
271 **Figure 8: Observed and simulated saltwater lines for experiment 1 (Case1a) under steady-state conditions at intervals longer than**
272 **90 minutes.**

273 For transient results, the saltwater line for experiments 1 and 2 for simulation time of 30, 60, and 90 minutes is used to verify
274 the corresponding results of numerical model, as shown in **Figure 9** and **Figure 10**. Both figures show that the model produces
275 reasonable simulated results for the saltwater lines (case3a) when compared to the observed ones. Table 6 also shows the
276 calibrated hydrogeological properties of the verified numerical model, including hydraulic conductivities in X, Y, and Z
277 directions (k_x , k_y , k_z), specific yield (S_y), specific storage (S_s), and effective porosity (η).



278
 279 **Figure 9: Observed and simulated saltwater lines for experiment1 (Case1a) for transient state condition: (a) 30min, (b) 60min, (c)**
 280 **90min**



281
 282 **Figure 10: Observed and simulated saltwater lines for experiment2 (Case3a) for transient state condition: (a) 30min, (b) 60min, (c)**
 283 **90min**

284 **Table 6: Calibrated values of the hydrogeological properties**

Hydrogeological Properties	k_x	k_y	k_z	S_y	S_s	η
Values	0.0069	0.0069	0.03	0.04	0.0619	0.0428

285



286 **3.2 Behavior evaluation of saltwater intrusion, flow, and hydraulic heads for categories (a), (b), and (c) model cases**

287 **3.2.1 Saltwater intrusion and flow behaviors in category(a) model cases**

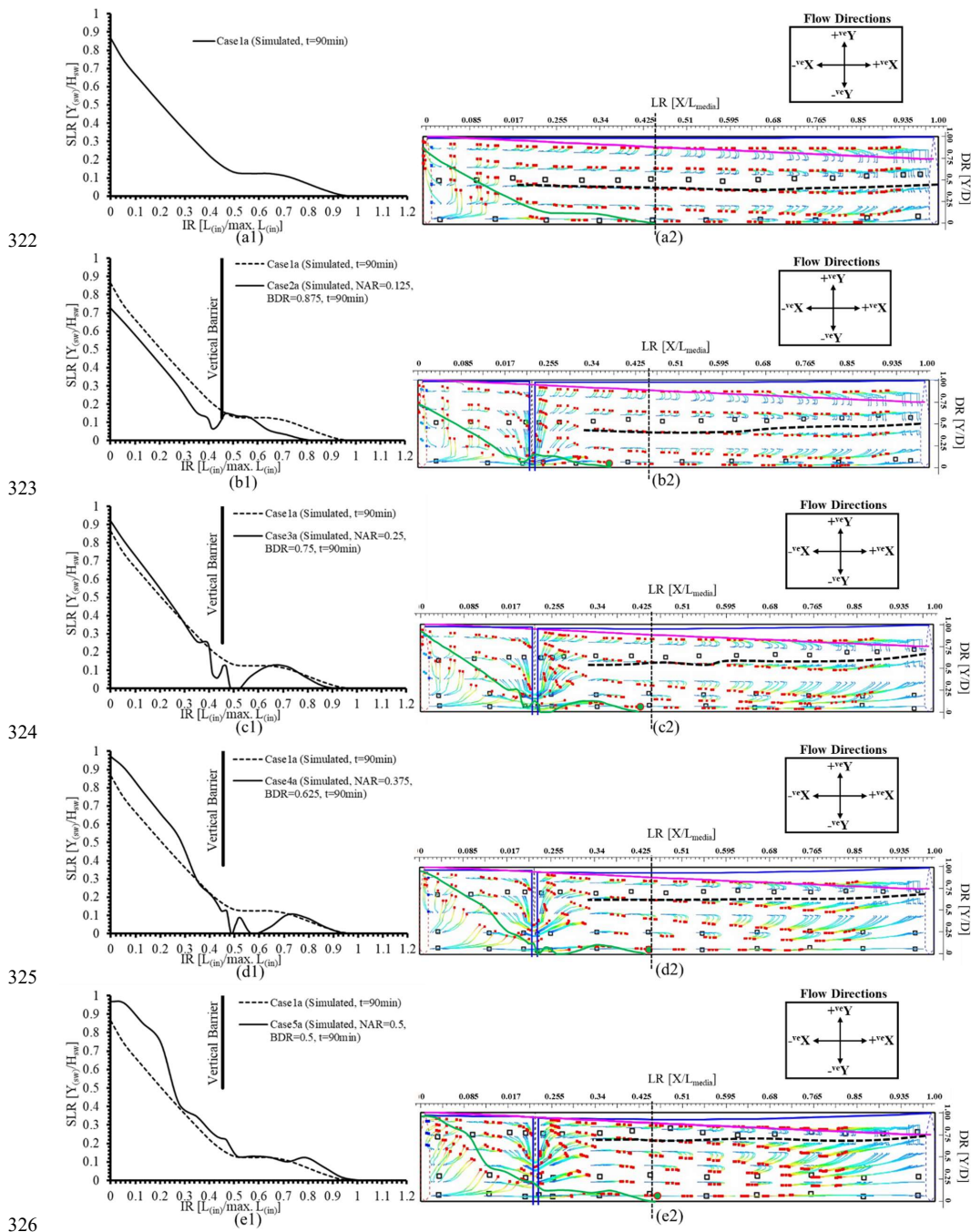
288 The modeling results of saltwater intrusion and the accompanying flow behavior for the cases in category (a) will be discussed
289 in this section, considering the evaluation ratios as summarized in **Table 3**. For case1a, **Figure 11a1** and **Figure 11a2** show
290 that the simulated saltwater line has an IR value of 0.97 and a corresponding LR value of 0.45. **Figure a2** illustrates the flow
291 behavior of this case, which shows that the flow in zone1 takes two directional flows: $+^{ve}Y$ and $+^{ve}X$. The $+^{ve}Y$ flow causes
292 the hydraulic heads near the saltwater boundary to be conserved at the media sand level. Furthermore, the $+^{ve}X$ flow force the
293 freshwater above the saltwater line to follow the same direction. Freshwater flow directions in zone 2 are towards $-^{ve}X$, $+^{ve}Y$,
294 and $-^{ve}Y$. Because of the $+^{ve}Y$ and $-^{ve}Y$ flows, a separation line between both flows could be identified with a DR ratio in the
295 range from 0.37 to 0.45, as shown in **Figure 11a2**. The $+^{ve}Y$ flow direction conserves hydraulic head along zone2.

296 After adding the vertical barrier (Case2a), the IR and LR values of the saltwater line are reduced to 0.83 and 0.39, respectively,
297 compared to those of Case1a (see **Figure 11b1** and **Figure 11b2**). Furthermore, **Figure 11b1** shows that the value of SLR
298 along the saltwater line of zone1a decreases more than that of case1a, with a sudden drop and rise just before and after the
299 vertical barrier. **Figure 11b2** explains this behavior by demonstrating that the high BDR (0.875) and low NAR values (0.125)
300 impede freshwater flows from zone2a to zone2b, causing overlaying pressure on zone1a. However, when compared to case1a,
301 the DR value of the separation line increases to be in the range from 0.40 to 0.50.

302 As shown in **Figure 11c1** and **Figure 11c2**, the IR and LR ratios for case3a are 0.90 and 0.42, respectively. **Figure 11c1**
303 shows that the SLR value at zone1a slightly increases when compared to case1a. It has also been observed that the saltwater
304 line fluctuates beneath the vertical barrier. This flow behavior is caused by the flow from zone2a to zone2b, which begins after
305 a decrease in BDR (0.75) and an increase in NAR (0.25) compared to case2a. This explanation is supported by the flow path
306 lines depicted in **Figure 11c2**. Because this partially flows between zones 2a and 2b, the saltwater line fluctuates and the
307 overlaying pressure on zone 1a decreases, causing the SLR value in this zone to rise. The separation line, on the other hand,
308 rises with DR values in the range from 0.50 to 0.68, implying that the majority of the freshwater flow is in the $-^{ve}Y$ flow
309 direction, resulting in hydraulic head reduction in zone2b.

310 By continuing to decrease the BDR value while increasing the NAR value, as in cases 4a, 5a, 6a, and 7a, the flow behavior for
311 these cases follows the same pattern as in case 3a (see **Figure 11d**, **11e**, **11f**, and **11g**, respectively). In these cases, the IR and
312 LR ratios are gradually increased from 0.97 and 0.46, respectively, in case4a to the maximum values of 1.05 and 0.49,
313 respectively, in case7a. Similarly, SLR values rise along zone1a until they reach the same levels as in case1a (see **Figure 11g**).
314 Moreover, the DR value continued to increase until it reached the range from 0.75 to 0.95 in case7a resulting in a hydraulic
315 head reduction in zone2b (see **Figure 11g2**).

316 Based on the above results, it is possible to conclude that using vertical barriers with small BDR and large NAR increases SLR
317 at zone1a as well as IR and LR values. Furthermore, the separation line with high DR values reduces the hydraulic heads along
318 zone2b, causing an excess increase in IR and LR. Conclusively, controlling both the SLR of zone 1a and the DR of the
319 separation line could effectively manage the saltwater intrusion represented by the IR and LR ratios. Management of saltwater
320 intrusion will be discussed in this study by controlling the DR of the separation line as well as the hydraulic heads along zone2b
321 using groundwater artificial recharge associated with the use of a vertical barrier.



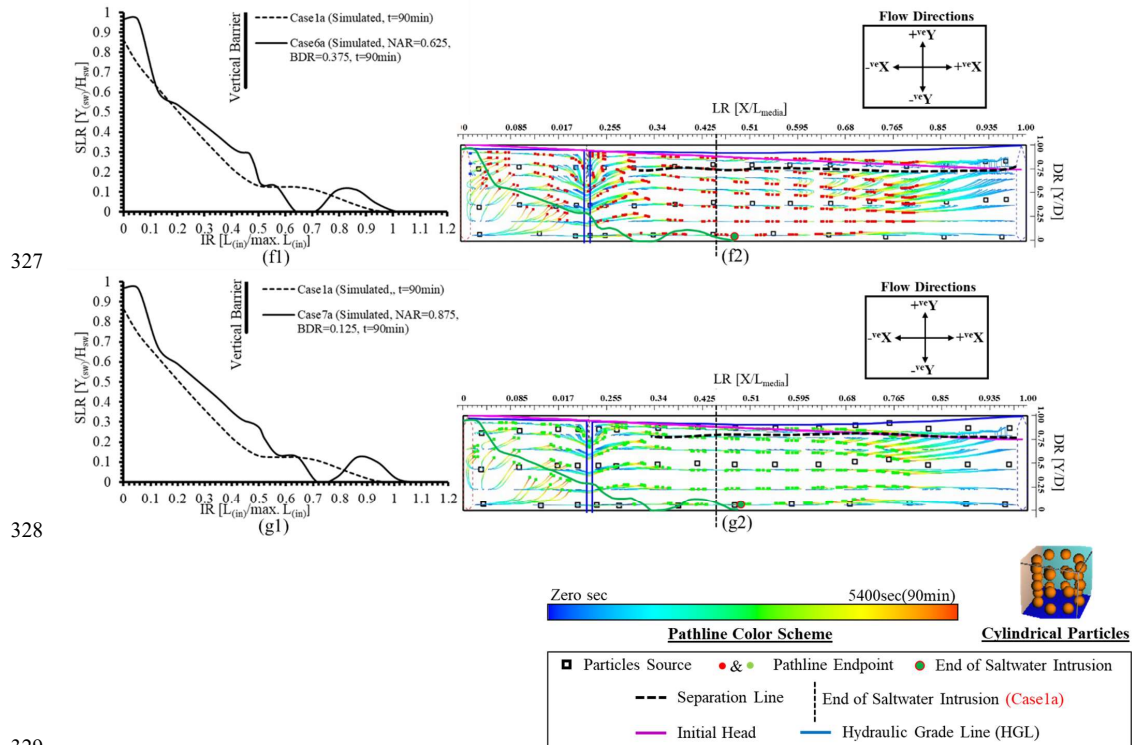
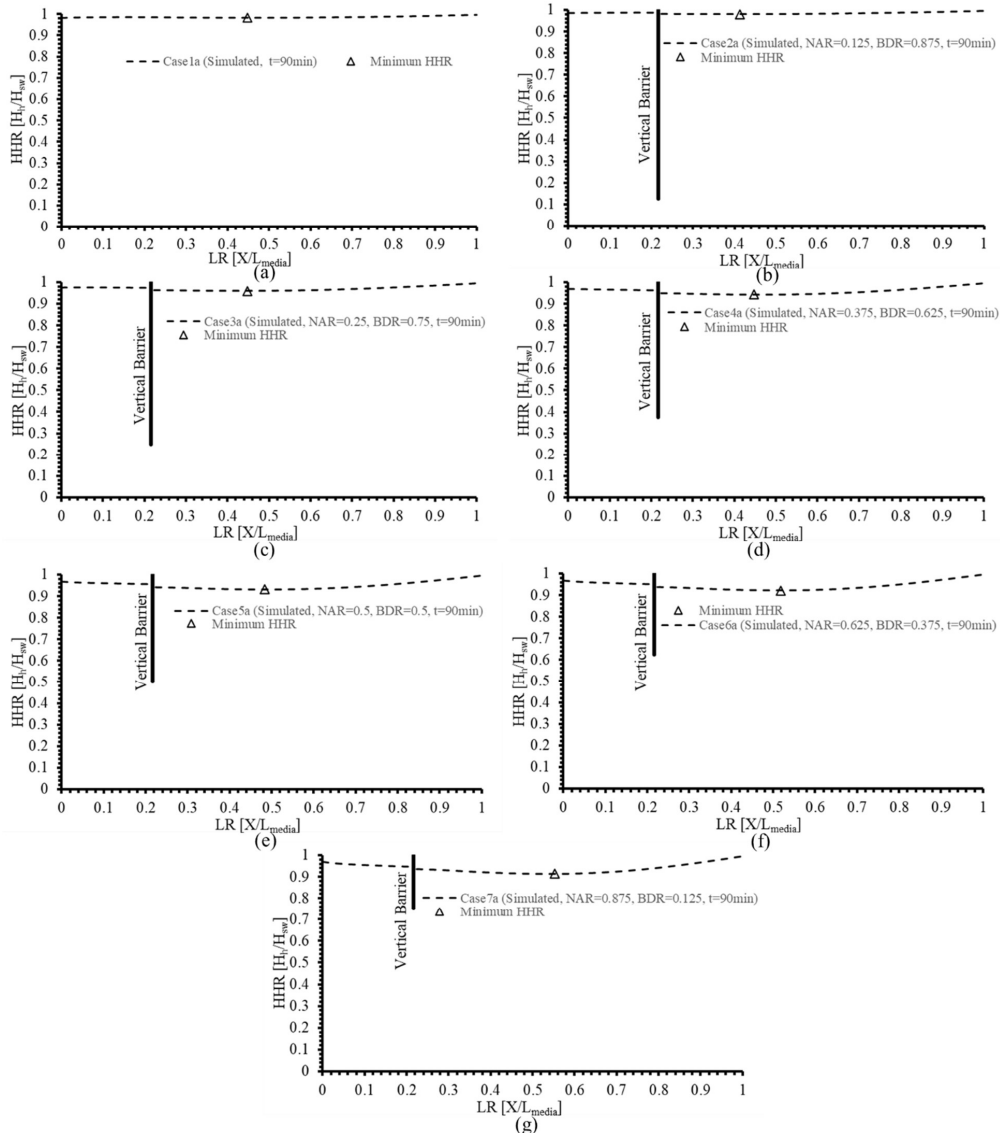


Figure 11: Simulated saltwater lines and groundwater flow behavior of the category (a) model cases: (a) case1a, (b) case2a, (c) case 3a, (d) case4a, (e) case5a, (f) case6a, (g) case7a

3.2.2 Hydraulic head variations in category(a) model cases

The hydraulic head variations are analyzed for the category(a) modeling cases, as shown in **Figure 12**. This figure illustrates the relationship between HHR and LR ratios with displaying the minimum HHR values and their locations along the aquifer. **Figure 12** shows that the hydraulic head of case7a has the lowest HHR value of 0.91 compared with the other cases (cases 1a-6a) located at a LR value of 0.55 (see **Figure 12g**). On the other hand, Case1a, has the highest value of the minimum HHR (0.98), and a location has a LR of 0.44, as shown in **Figure 12a**.



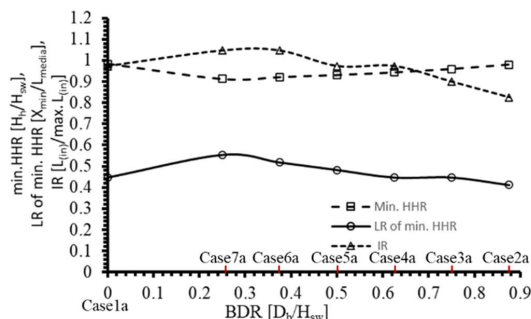
338

339

340 **Figure 12: Values and locations of the minimum HHR of the category (a) model cases: (a) case1a, (b) case2a, (c) case 3a, (d) case4a,**
 341 **(e) case5a, (f) case6a, (g) case7a**

342 The above results can be summarized as illustrated in **Figure 13**, which depicts the effect of BDR on the location (LR) and
 343 value of the minimum HHR and the IR ratios. The minimum hydraulic head is located at zone2b for all study cases (case1a-
 344 case7a), with LR values ranging from 0.45 to 0.55 and corresponding minimum HHR values ranging from 0.91 to 0.98. On
 345 the other hand, the maximum IR, occurs for both cases 6a and 7a with a value of 1.05 when using a BDR in the value range
 346 from 0.25 to 0.38. Given these findings, increasing the hydraulic head represented by HHR could effectively repel saltwater
 347 intrusion when combined with the vertical barrier countermeasure. For this purpose, using groundwater artificial recharge,
 348 whether by surface or subsurface recharge, at the location of the minimum HHR value (LR in the range from 0.45 to 0.55),
 349 combined with the use of a vertical barrier, could be used to repel saltwater intrusion, as will be discussed in the following
 350 sections of this study.

351



352
 353 **Figure 13: Effect of BDR on the IR and minimum HHR values and locations**

354 **3.2.3 Saltwater intrusion and flow behaviors in categories (b) and (c) model cases**

355 Groundwater artificial recharge is used to repel saltwater intrusion in zone2b along the LR range (from 0.45 to 0.55), which
 356 has a minimum value of HHR for preserving its value at the unity value. Surface and subsurface recharge are numerically
 357 discussed, either separately or in conjunction with the vertical barrier, as shown in **Table 4** for category (b) and category (c)
 358 study cases. The recharge is applied along the whole range of LR values from 0.45 to 0.55 for surface recharge. In contrast,
 359 for subsurface recharge, the recharge is applied as a line of wells at the midpoint of the same LR range at a value of 0.5. The
 360 results of category (b) and category (c) study cases will be compared with the base case results (case1a) and the corresponding
 361 cases of category (a) in the following discussions.

362 **Figure 14a1** and **Figure 14a2** show that the saltwater line in case1b has IR and LR values of 1.00 and 0.47, respectively,
 363 which are slightly higher than those in case1a (0.97 and 0.46, respectively). When comparing case 1c to case 1a, the IR and
 364 LR values increased to 1.05 and 0.49, respectively. For both model cases 1b and 1c, it is observed that the value of SLR
 365 increased more in both cases than in case 1a. This salt line behavior could be explained by the fact that, in case1b, the surface
 366 recharge works as a hydraulic barrier that prevents saltwater from flow in the ^{+ve}X direction as well as forces it to flow
 367 intensively in the ^{+ve}Y direction. This behavior causes an increase in the SLR (increase in saltwater head), which leads to an
 368 increase in the IR and LR values, and the majority of recharged freshwater is forced to take a ^{+ve}X direction (see **Figure 14a2**).
 369 The flow behavior in case1c is the same as in case1b (see **Figure 14a3**); however, the countermeasure effect of subsurface
 370 recharge, which is a line of wells, is less than that of surface recharge, which is a water mass. This could explain the higher IR
 371 and LR values in case1c compared to case1b, with the same SLR value.

372 The combined effect of the vertical barrier and surface recharge repels saltwater intrusion significantly in case2b, with IR and
 373 LR values of 0.75 and 0.35, respectively, compared to corresponding values of 0.83 and 0.39 in case2a (see **Figure 14b1** and
 374 **Figure 14b2**). In contrast to case 2a, although the value of SLR increases in zones 1a and 1b, the ^{-ve}X flow direction of surface
 375 recharge towards the neck area under the vertical barrier forces the saltwater line to repulse (**Figure 14b2**). Because of the
 376 lower effect of well recharge than that of surface recharge, the IR and LR ratios are higher in case 2c than in case 2b, with
 377 values of 0.82 and 0.38, respectively. Conclusively, the IR, LR, and SLR values for Case 2b are the lowest compared with
 378 those of Case 2c and the previous cases (Case 1b and Case 1c) confirming the efficient combination of the vertical barrier and
 379 surface recharge at the location of the minimum HHR (LR in the range from 0.45 to 0.55).

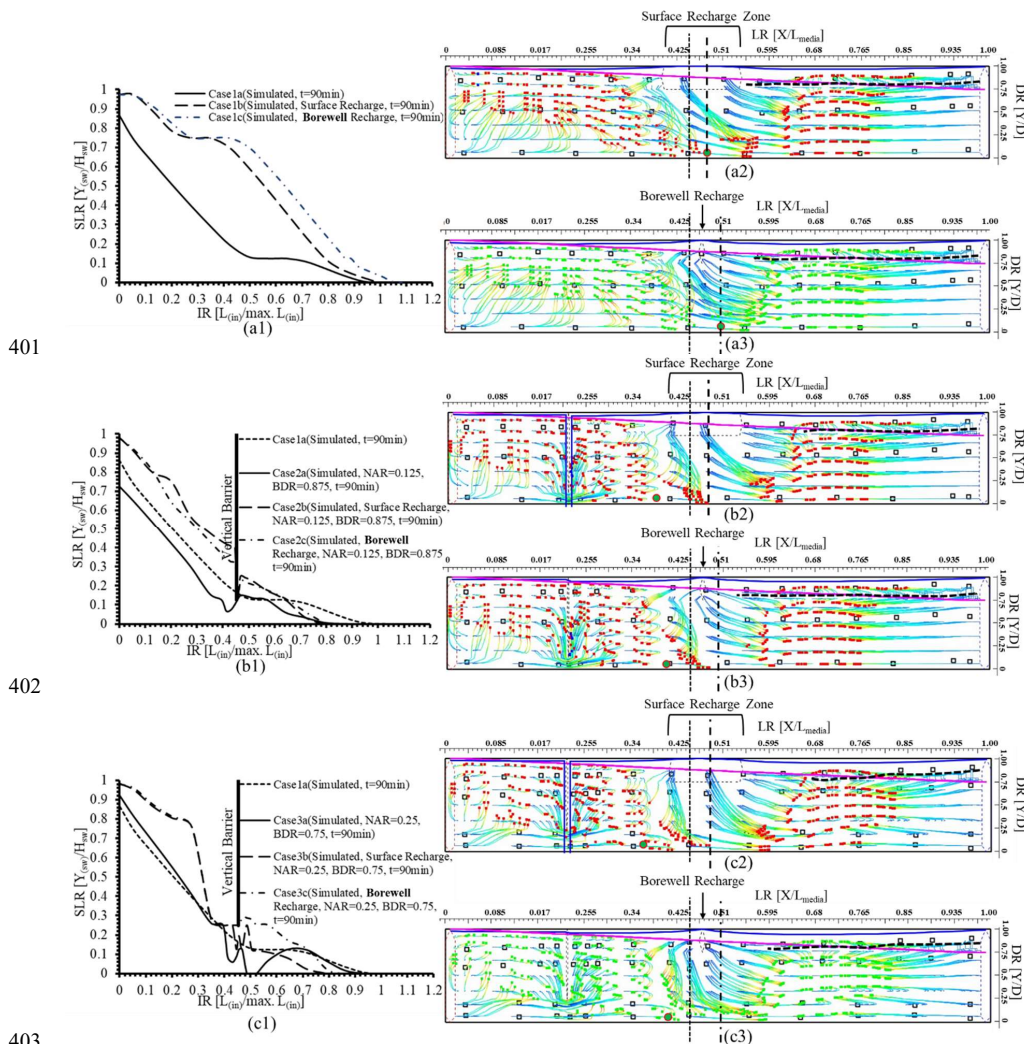
380 **Figure 14c1** and **Figure 14c2** show that the IR and LR values for case3b are significantly lower, at 0.68 and 0.32, respectively,
 381 when compared to the corresponding values for case2b and case3a. On the contrary, the SLR value increases in comparison
 382 with the same cases, however, it drops dramatically just behind the vertical barrier of zone 1a, as shown in **Figure 14c1**.
 383 According to one interpretation of this behavior, because of the high NAR value (0.25) compared to cases 2b and 3a, freshwater
 384 flows intensively from zone 2a to zone 2b, causing a decrease in SLR value in the adjacent area to the vertical barrier as well
 385 as fluctuations beneath the barrier (see **Figure 14c2**). The IR and LR ratios for case3c are 0.82 and 0.38, respectively, which



386 is quite large when compared to case3b. Furthermore, due to the weak effect of well recharge, the SLR value of zone1b is
 387 greater than that of case3b (see **Figure 14c1** and **Figure 14c3**).

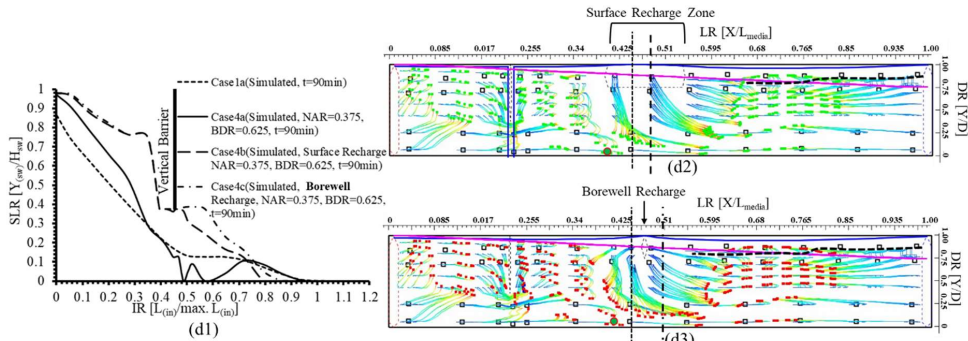
388 By continuing to decrease BDR while increasing NAR, IR and LR values increase significantly compared to the previous case
 389 (case3b) to have the same values of 0.81 and 0.38 for cases 4b, 5b, 6b, and 7b. Similarly, IR and LR values increase
 390 significantly when compared to case 3c, reaching 0.85 and 0.4, respectively, for cases 4c, 5c, and 6c, and 0.75 for case 7c
 391 (**Figure 14d1, e1, f1, and g1**). This behavior is due to the freshwater's +^{ve}X flow direction from zone 2a to zone 2b, which
 392 reduces the effect of surface and well recharge, as shown in **Figure 14d2, d3, e2, e3, f2, f3, g2, and g3**. In comparison to the
 393 corresponding IR and LR values for cases 4a, 5a, 6a, and 7a, which are in the range from 0.97 to 1.04 and from 0.46 to 0.49,
 394 respectively, the corresponding IR and LR values for category(b) and category(c) are smaller, as shown in **Figure 14d1, e1,**
 395 **f1, and g1**. The hydraulic heads along the experiment section are unchanged in all cases of categories (b) and (c), and the DR
 396 ratio of the separation line is nearly the same with a value range from 0.75 to 0.90.

397 Based on the findings, it is possible to conclude that artificial aquifer recharging along the LR values from 0.45 to 0.55, which
 398 has a minimum value of HHR ratio to conserve its value, as well as the unity accompanied by using a vertical barrier, has a
 399 significant effect on saltwater line repulsing. Furthermore, because of its body mass, surface recharge is more efficient than
 400 well recharge.

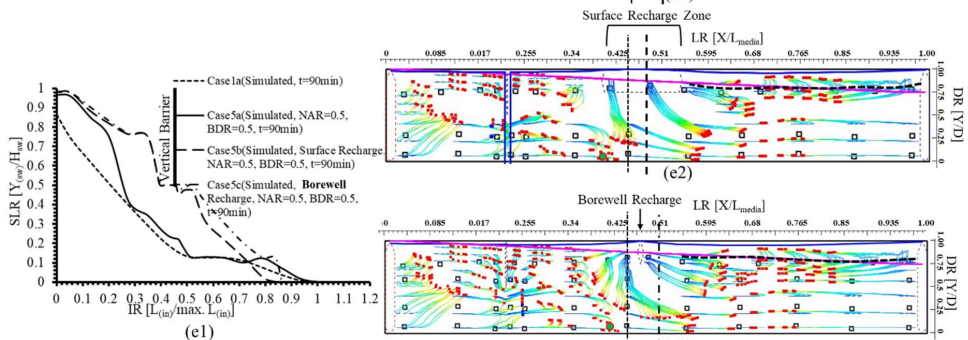




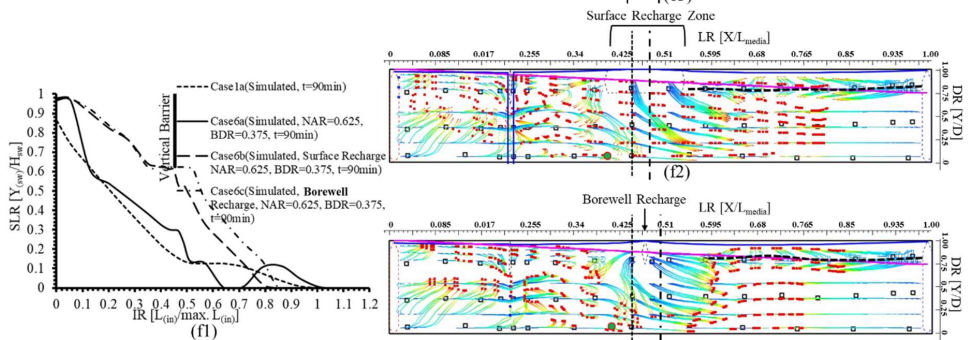
404



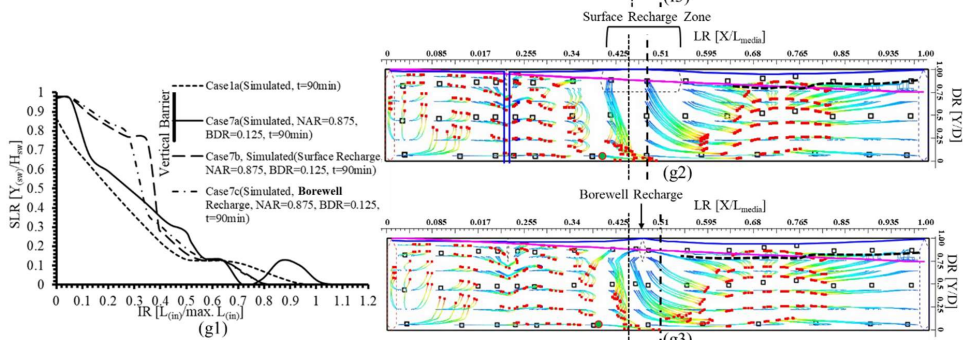
405



406

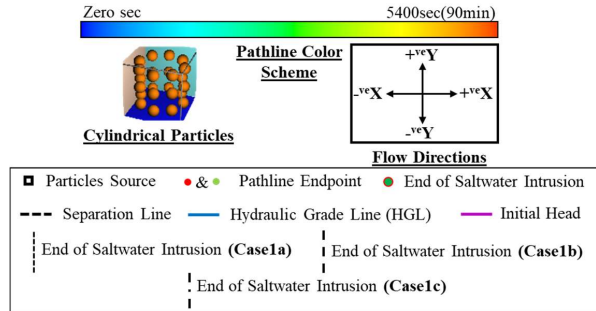


407



408

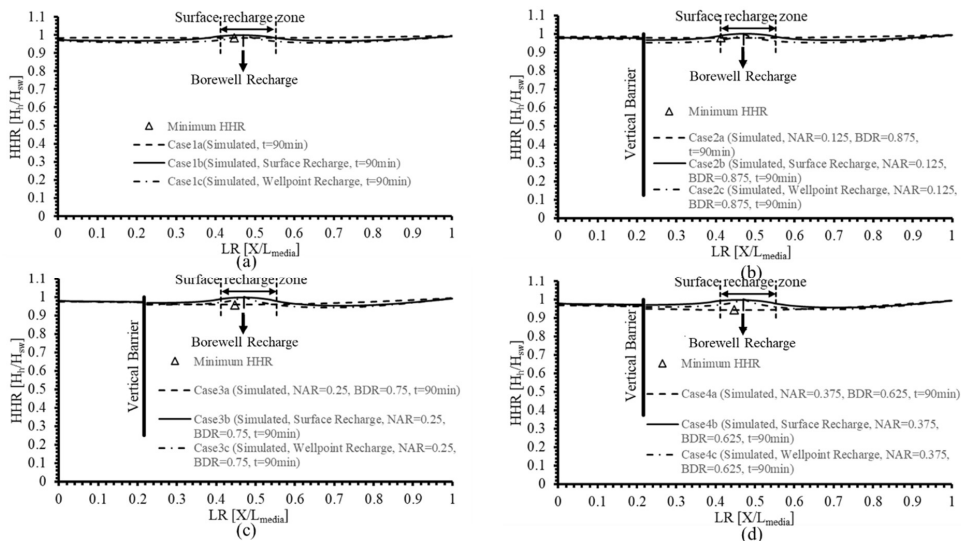
409



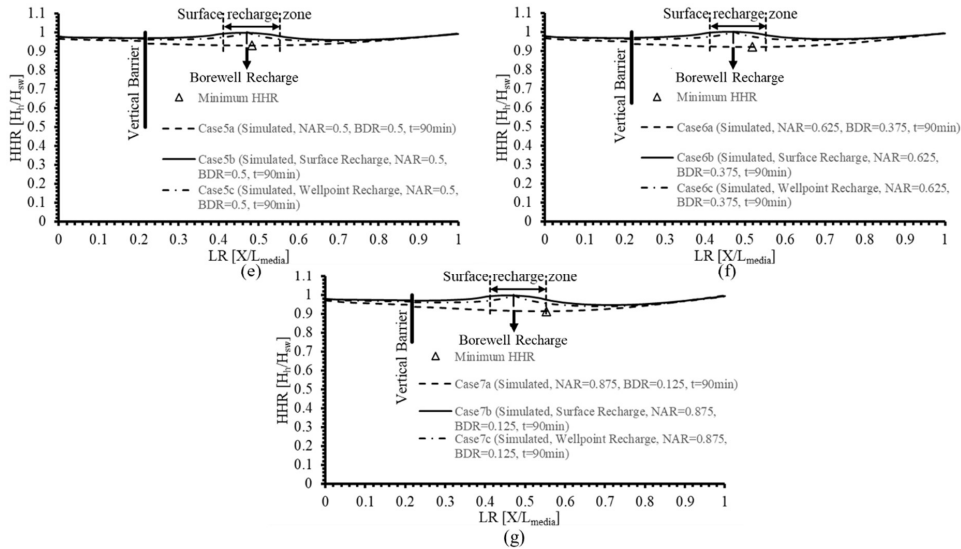
410
 411 **Figure 14: Simulated saltwater lines and groundwater flow behavior of the category (b) and (c) model cases: (a) case1b & 1c, (b)**
 412 **case2b&2c, (c) case 3b&3c, (d) case4b&4c, (e) case5b&5c, (f) case6b&6c, (g) case7b&7c**

413 **3.2.4 Hydraulic head variations in categories (b) and (c) model cases**

414 **Figure 15a to Figure 15g** depict the hydraulic heads along the aquifer as represented by the HHR ratio for cases in categories
 415 (b) and (c) compared to category (a). The hydraulic heads for all cases have been conserved along the LR ratio from 0.45 to
 416 0.55, which has the minimum value of HHR to have the unity value, and the losses through the vertical barrier are greatly
 417 reduced when compared to those of category (a).



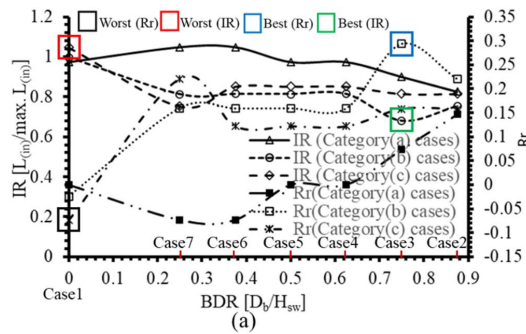
418



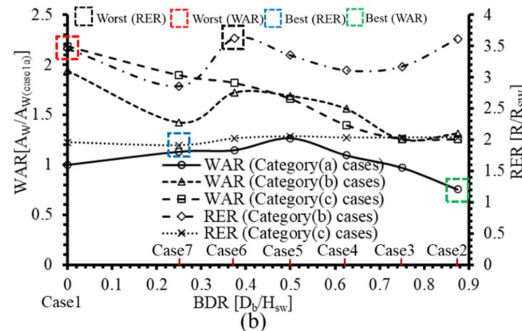
419
 420 **Figure 15: Hydraulic head variation along the aquifer for categories (b) and (c) compared with those of category (a) model cases:**
 421 **(a) case1a&1b&1c, (b) case2b&2b&2c, (c) case 3b&3b&3c, (d) case4b&4b&4c, (e) case5b&5b&5c, (f) case6b&6b&6c, (g)**
 422 **case7b&7b&7c**

423 **3.3 Classification of model cases**

424 The classification ratios described in Section "2.3.2 Classification Ratios" are summarized and classified in **Figure 16 and**
 425 **Table 7. Figure 16a** presents the IR and Rr values for each model case, whereas **Figure 16b** depicts the WAR and RER
 426 values. **Figure 16 and Table 7** show that case3b has the best IR and Rr values of 0.68 and 0.29, respectively. Case2a, on the
 427 other hand, has the highest WAR value of 0.76. Furthermore, case7c has the highest RER value of 1.91. **Figure 16 and Table**
 428 **7** show that case1c has the worst IR, Rr, and WAR values of 1.05, -0.07, and 2.18, respectively. Furthermore, case6b has the
 429 lowest RER value of 3.62. The remaining model cases are categorised as unclassified model cases. Based on the findings, it is
 430 unable to determine which model case is the most successful scenario to implement as a saltwater intrusion countermeasure.
 431 As a result, the DMM model is critical for determining the most effective model case.



432



433
 434 **Figure 16: Classifications of model cases included in categories (a), (b), and (c): (a) IR and R_r values, (b) WAR and RER values**

435 **Table 7: Model cases classification according to values of classification ratios**

Classification Ratio	Best	Worst
IR	Case3b (0.68)	Case1c (1.05)
R _r	Case3b (0.29)	Case1c (-0.07)
WAR	Case2a (0.76)	Case1c (2.18)
RER	Case7c (1.91)	Case6b (3.62)

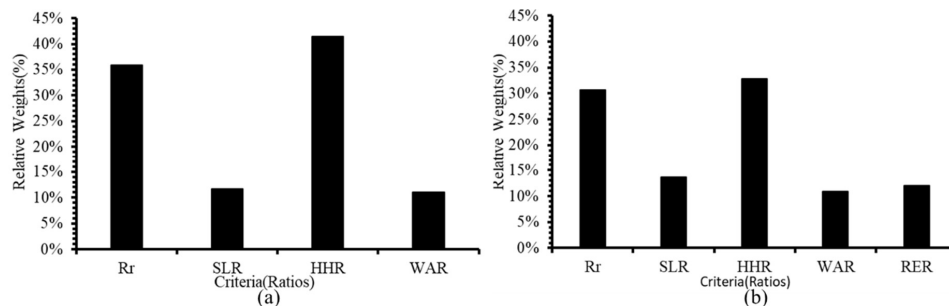
436

437 **3.4 Selecting the Most Effective Model Case (AHP application results)**

438 As previously stated, the AHP model is applied to the numerical model results at two different levels of selection (levels (1)
 439 and (2)). Model cases are referred to as alternatives at this stage, and selected ratios among the evaluation and classification
 440 ratios are referred to as criteria. Level (1) needs for determining the best alternative (model case) in each category. Furthermore,
 441 level (2) is for deciding the best model case. For category (a) model cases (alternatives), the criteria values used include SLR,
 442 HHR, R_r, and WAR, and the RER is added over these ratios for categories (b) and (c).

443 **3.4.1 Level (1) Results**

444 For the alternatives in each category, the model is applied at level (1) using the weights chosen among criteria as shown in
 445 **Figure 17**. According to **Figure 17a**, HHR has the largest weight for category (a) alternatives, followed by R_r. Also, WAR
 446 has the lowest weight. Similarly, for category (b) alternatives (see **Figure 17b**), the same rating is observed for HHR (the
 447 highest weight), followed by R_r, while WAR has the lowest weight.

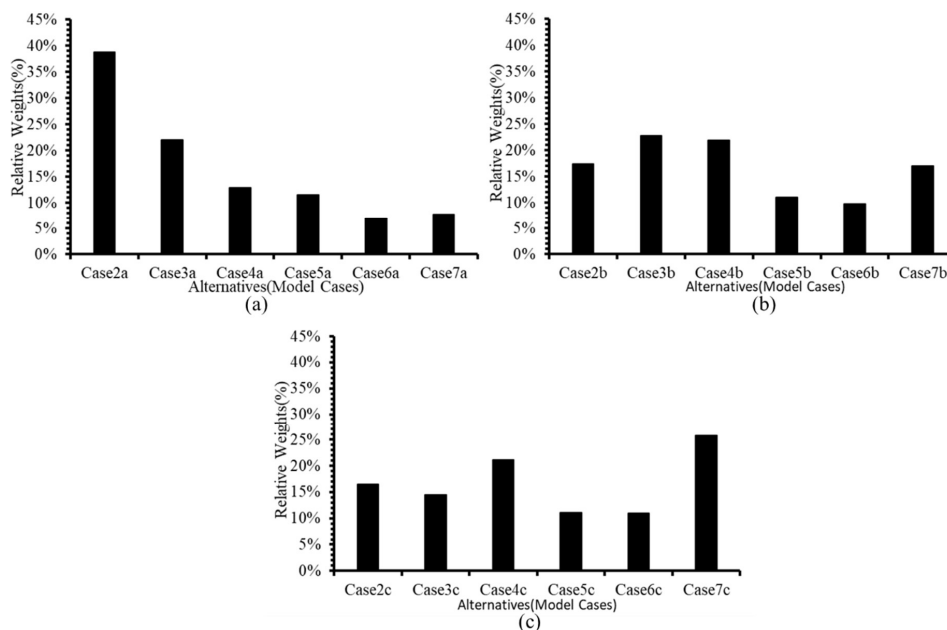


448
 449 **Figure 17: Level (1) criteria relative weights for different categories: (a) category (a), (b) category (b), and (c)**

450 **Figure 18** illustrates the results of the relative weights in the three categories for each alternative. It is evident that case 2a
 451 ranks first, with a relative weight of 38.72%, followed by case 3a, while case 6a ranks last in this category (see **Figure 18a**).
 452 Case 3b is the best alternative in category (b), followed by Case 4b, which has a weight difference of 3.5% with Case 3b. and
 453 Case 6b is the worst alternative in this category (see **Figure 18b**). Case 7c is the recommended alternative in category (c), with



454 a weight of 25.83% ahead of the rest of the alternatives, followed by Case 4c, while Case 6c placed in last place in this category
455 (see **Figure 18c**).



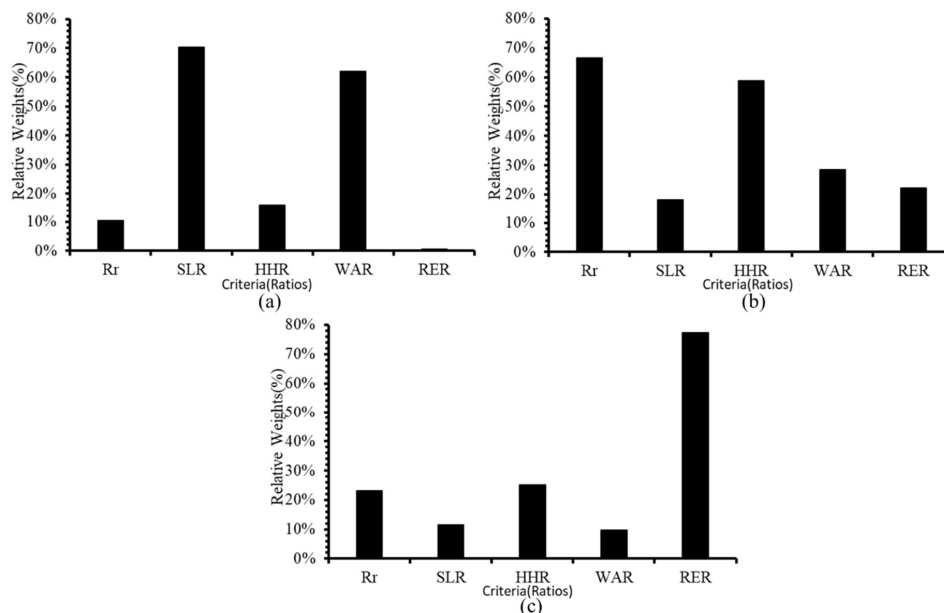
456
457 **Figure 18:** Level (1) relative weights among alternatives: (a) category (a), (b) category (b), (c) category (c)

458 3.4.2 Level (2) Results

459 The model's level (1) findings are summarized in the three best case model alternatives (cases 2a, 3b, and 7c), as shown in
460 **Figure 18**. **Figure 19** summarizes the relative weights for each criterion in relation to the three alternatives. In case 2a, SLR
461 is the most effective criterion, followed by WAR, while RER has a negligible effect (**Figure 19a**). on the other hand, Rris is the
462 most essential parameter influencing case 3b, followed by HHR (**Figure 19b**). Case 7c is clearly influenced primarily by RER
463 (**Figure 19c**).

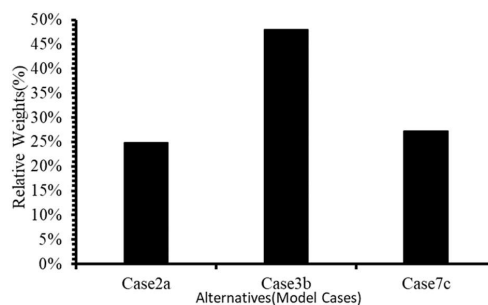
464

465



466
 467 **Figure 19: Level(2) relative weights for each criterion of the alternatives for final decision: (a) case2a, (b) case3b, (c) case7c**

468 As a result of the preceding findings, **Figure 20** illustrates the weight values of the alternatives as a final decision, which
 469 clearly supports Case 3b by a percentage of 47.93% over a percentage of 27.30% for Case 7c and 24.85% for Case 2a. Based
 470 on the findings, it could be conclude that the components of case3b (combining the vertical barrier with surface recharge along
 471 the LR ratio from 0.45 to 0.55) could be classified as best model case for use as a saltwater repellent countermeasure.
 472 Furthermore, the vertical barrier has a greater effect when combined with surface recharge than when combined with well
 473 recharge. On the other hand, surface recharge necessitates a high recharge rate (about 1.25 times the borewell recharge).



474
 475 **Figure 20: Level (2) relative weights for the three alternatives for final decision**

476 Conclusion

477 Seawater intrusion is a common environmental issue that degrades the quality of fresh groundwater in the coastal aquifer.
 478 Because of the hydraulic connection between the coastal aquifer and the sea, using conventional physical vertical barriers
 479 could reform the groundwater's hydraulic gradient, disrupt the hydrodynamic balance between the two fluids, affect the
 480 potentiometric surface of the coastal aquifer, and increase saltwater intrusion. In this study, saltwater intrusion is managed by
 481 controlling hydraulic heads along the coastal aquifer using surface or subsurface recharges in conjunction with the traditional
 482 vertical barrier countermeasure. A physical model is created to investigate the saltwater line behavior with a vertical barrier
 483 (experiment1) and without a vertical barrier (experiment2). The experimental results are used to validate a MODFLOW created
 484 numerical model. Following that, three categories of model cases ((a), (b), and (c)), each with seven numerically proposed



485 model cases, are numerically proposed for: analyzing the saltwater-freshwater interaction through porous media; selecting the
486 best location of the recharge; determining the best depth of the vertical barrier; and selecting the components of the efficient
487 countermeasure system, including the vertical barrier, surface recharge, and subsurface recharge. The dimension analysis
488 technique is used to generate evaluation ratios in order to analyze and characterize the numerical model cases' saltwater line
489 and hydraulic head variations. According to category (a) simulation results, the minimum hydraulic head occurs through length
490 ratio (LR) values ranging from 0.45 to 0.55 with corresponding values of hydraulic head ratio (HHR) ranging from 0.91 to
491 0.98. On the other hand, surface and subsurface recharge, are implemented through categories (b) and (c) to investigate
492 saltwater repulsion by maintaining the HHR value of unity within the concluded LR range. As a preset for finding the best
493 model case, classification ratios are proposed to classify the model cases included in the three mentioned categories as the best
494 or worst model case. Using the calculated classification ratio values, an analytic hierarchy process (AHP) decision-making
495 model (DMM) is used to select the best model case that is recommended for saltwater repulsion using two selection levels.
496 The first selection level concluded that the HHR has the highest relative weight in all categories, while the WAR has the
497 lowest. Cases 2a, 3b, and 7c are rated as the best model cases in categories (a), (b), and (c), respectively, and are most affected
498 by SLR, Rr, and RER, respectively. In the second selection level, the final decision is made that case 3b is the overall best
499 model case, which has a reasonable WAR of 1.25 and a minimum IR and maximum Rr of 0.68 and 0.29, respectively.
500 Moreover, the findings indicate that countermeasure systems (combining the vertical barrier with surface recharge) are the
501 best choice to be used in this case.

502

503 **Data availability:** The authors support the data availability of this research, which includes experimental measurements,
504 numerical model output files, numerical model post-processed results, and AHP output results. The data source can be found
505 at <https://doi.org/10.4211/hs.8c31e2e9f8ab459ab99c61ccc110ab08> (Mahmod, 2023).

506 **Author contribution:** "Wael Elham Mahmod designed the experiments and carried them out. He also built the numerical
507 model using ModelMuse and performed the simulations and analysed the results. With contributions from all co-authors, Wael
508 Elham Mahmod prepared the manuscript. Usama Hamed Issa developed the AHP model and performed the decision-making
509 simulations.

510 **Financial Support:** This research work was funded by Institutional Fund Projects under Grant No. IFPIP-1252-137-1443
511 provided by the Ministry of Education and King Abdulaziz University, DSR, Jeddah, Saudi Arabia.

512 **Acknowledgements:** The authors gratefully acknowledge the technical and financial support provided by the Ministry of
513 Education and King Abdulaziz University, DSR, Jeddah, Saudi Arabia.

514 **Competing interests:** The contact author has declared that none of the authors has any competing interests.

515 **References**

516 Abarca, E., Vázquez-Suñé, E., Carrera, J., Capino, B., Gámez, D., and Batlle, F.: Optimal design of measures to correct
517 seawater intrusion, *Water Resour. Res.*, 42, <https://doi.org/https://doi.org/10.1029/2005WR004524>, 2006.

518 Abd-Elaty, I., Abd-Elhamid, H. F., and Negm, A. M.: Investigation of Saltwater Intrusion in Coastal Aquifers, in: *Groundwater*
519 *in the Nile Delta*, edited by: Negm, A. M., Springer International Publishing, Cham, 329–353,
520 https://doi.org/10.1007/698_2017_190, 2019a.

521 Abd-Elaty, I., Abd-Elhamid, H., and Mousavi Nezhad, M.: Numerical analysis of physical barriers systems efficiency in
522 controlling saltwater intrusion in coastal aquifers, *Environ. Sci. Pollut. Res.*, 2019b.

523 Abd-Elhamid, H. F.: Investigation and control of seawater intrusion in the Eastern Nile Delta aquifer considering climate
524 change, *Water Supply*, 17, 311–323, <https://doi.org/10.2166/ws.2016.129>, 2016.

525 Abd-Elhamid, H. F. and Javadi, A. A.: A Cost-Effective Method to Control Seawater Intrusion in Coastal Aquifers, *Water*
526 *Resour. Manag.*, 25, 2755–2780, <https://doi.org/10.1007/s11269-011-9837-7>, 2011.

527 Abd-Elhamid, H. F., Abd-Elaty, I., and Negm, A. M.: Control of Saltwater Intrusion in Coastal Aquifers, in: *Groundwater in*



- 528 the Nile Delta, edited by: Negm, A. M., Springer International Publishing, Cham, 355–384,
529 https://doi.org/10.1007/698_2017_138, 2019.
- 530 Abdelwahab, S. F., Issa, U. H., and Ashour, H. M.: A Novel Vaccine Selection Decision-Making Model (VSDMM) for
531 COVID-19, *Vaccines*, 9, <https://doi.org/10.3390/vaccines9070718>, 2021.
- 532 Achu, A. L., Thomas, J., and Reghunath, R.: Multi-criteria decision analysis for delineation of groundwater potential zones in
533 a tropical river basin using remote sensing, GIS and analytical hierarchy process (AHP), *Groundw. Sustain. Dev.*, 10, 100365,
534 <https://doi.org/https://doi.org/10.1016/j.gsd.2020.100365>, 2020.
- 535 Ahmadi, H., Kaya, O. A., Babadagi, E., Savas, T., and Pekkan, E.: GIS-Based Groundwater Potentiality Mapping Using AHP
536 and FR Models in Central Antalya, Turkey, <https://doi.org/10.3390/IECG2020-08741>, 2021.
- 537 Alwetaishi, M., Gadi, M., and Issa, U. H.: Reliance of building energy in various climatic regions using multi criteria, *Int. J.*
538 *Sustain. Built Environ.*, 6, 555–564, <https://doi.org/10.1016/j.ijsbe.2017.12.002>, 2017.
- 539 Anders, R., Mendez, G. O., Futa, K., and Danskin, W. R.: A geochemical approach to determine sources and movement of
540 saline groundwater in a coastal aquifer., *Ground Water*, 52, 756–768, <https://doi.org/10.1111/gwat.12108>, 2014.
- 541 Arunbose, S., Srinivas, Y., Rajkumar, S., Nair, N. C., and Kaliraj, S.: Remote sensing, GIS and AHP techniques based
542 investigation of groundwater potential zones in the Karumeniyar river basin, Tamil Nadu, southern India, *Groundw. Sustain.*
543 *Dev.*, 14, 100586, <https://doi.org/https://doi.org/10.1016/j.gsd.2021.100586>, 2021.
- 544 ASCE: Standard Guidelines for Artificial Recharge of Ground Water, EWRI/ASCE., American Society of Civil Engineers,
545 <https://doi.org/10.1061/9780784405482>, 2001.
- 546 Bobba, A. G.: Mathematical models for saltwater intrusion in coastal aquifers, *Water Resour. Manag.*, 7, 3–37,
547 <https://doi.org/10.1007/BF00872240>, 1993.
- 548 Cai, J., Taute, T., and Schneider, M.: Recommendations of Controlling Saltwater Intrusion in an Inland Aquifer for Drinking-
549 Water Supply at a Certain Waterworks Site in Berlin (Germany), *Water Resour. Manag.*, 29, 2221–2232,
550 <https://doi.org/10.1007/s11269-015-0937-7>, 2015.
- 551 Cary, L., Petelet-Giraud, E., Bertrand, G., Kloppmann, W., Aquilina, L., Martins, V., Hirata, R., Montenegro, S., Pauwels, H.,
552 Chatton, E., Franzen, M., Arouet, A., Lasseur, E., Picot, G., Guerrot, C., Fléhoc, C., Labasque, T., Santos, J. G., Paiva, A.,
553 Braibant, G., and Pierre, D.: Origins and processes of groundwater salinization in the urban coastal aquifers of Recife
554 (Pernambuco, Brazil): A multi-isotope approach, *Sci. Total Environ.*, 530–531, 411–429,
555 <https://doi.org/https://doi.org/10.1016/j.scitotenv.2015.05.015>, 2015.
- 556 Castillo, J. L., Martínez Cruz, D. A., Ramos Leal, J. A., Tuxpan Vargas, J., Rodríguez Tapia, S. A., and Marín Celestino, A.
557 E.: Delineation of Groundwater Potential Zones (GWPZs) in a Semi-Arid Basin through Remote Sensing, GIS, and AHP
558 Approaches, <https://doi.org/10.3390/w14132138>, 2022.
- 559 Domenico, P. A., Schwartz, F. W., and SCHWARTZ, F. A.: *Physical and Chemical Hydrogeology*, Wiley, 1998.
- 560 Eissa, M. A., de Dreuzy, J.-R., and Parker, B.: Integrative management of saltwater intrusion in poorly-constrained semi-arid
561 coastal aquifer at Ras El-Hekma, Northwestern Coast, Egypt, *Groundw. Sustain. Dev.*, 6, 57–70,
562 <https://doi.org/https://doi.org/10.1016/j.gsd.2017.10.002>, 2018.
- 563 Gangadharan, R., Nila, R., and Vinoth, S.: Assessment of groundwater vulnerability mapping using AHP method in coastal
564 watershed of shrimp farming area, *Arab. J. Geosci.*, 9, 1–14, <https://doi.org/10.1007/s12517-015-2230-8>, 2016.
- 565 Güllü, Ö. and Kavurmacı, M.: Investigation of temporal variation of groundwater salinity potential using AHP-based index,
566 *Environ. Monit. Assess.*, 195, 365, <https://doi.org/10.1007/s10661-023-10993-5>, 2023.
- 567 Guo, Q., Huang, J., Zhou, Z., and Wang, J.: Experiment and Numerical Simulation of Seawater Intrusion under the Influences
568 of Tidal Fluctuation and Groundwater Exploitation in Coastal Multilayered Aquifers, *Geofluids*, 2019, 2316271,
569 <https://doi.org/10.1155/2019/2316271>, 2019.
- 570 Hamed, Y., Hadji, R., Redhaouia, B., Zighmi, K., Bâali, F., and El Gayar, A.: Climate impact on surface and groundwater in
571 North Africa: a global synthesis of findings and recommendations, *Euro-Mediterranean J. Environ. Integr.*, 3, 25,
572 <https://doi.org/10.1007/s41207-018-0067-8>, 2018.
- 573 Huang, P.-S. and Chiu, Y.-C.: A Simulation-Optimization Model for Seawater Intrusion Management at Pingtung Coastal
574 Area, Taiwan, *Water*, 10, <https://doi.org/10.3390/w10030251>, 2018.
- 575 Hussain, M. S., Abd-Elhamid, H. F., Javadi, A. A., and Sherif, M. M.: Management of Seawater Intrusion in Coastal Aquifers:
576 A Review, *Water*, 11, <https://doi.org/10.3390/w1122467>, 2019.
- 577 Kallioras, A., Pliakas, F.-K., Schüth, C., and Rausch, R.: Methods to Countermeasure the Intrusion of Seawater into Coastal
578 Aquifer Systems, in: *Wastewater Reuse and Management*, 470–490, https://doi.org/10.1007/978-94-007-4942-9_17, 2013.



- 579 M Armanuos, A., Gamal Eldin Ibrahim, M., Mahmod, W., Takemura, J., and Yoshimura, C.: Analysing the Combined Effect
580 of Barrier Wall and Freshwater Injection Countermeasures on Controlling Saltwater Intrusion in Unconfined Coastal Aquifer
581 Systems, *Water Resour. Manag.*, 33, <https://doi.org/10.1007/s11269-019-2184-9>, 2019.
- 582 Mahmod, W. E. (2023). Supplement materials for the physical, numerical, and decision-making models for developing
583 functional recharge systems to repel saltwater intrusion for coastal aquifer sustainability, *HydroShare*,
584 <https://doi.org/10.4211/hs.8c31e2e9f8ab459ab99c61ccc110ab08>
- 585 Maliva, R. G.: ASR and Aquifer Recharge Using Wells, in: *Anthropogenic Aquifer Recharge: WSP Methods in Water*
586 *Resources Evaluation Series No. 5*, Springer International Publishing, Cham, 381–436, https://doi.org/10.1007/978-3-030-11084-0_13, 2020a.
- 588 Maliva, R. G.: Surface-Spreading AAR Systems (Non-basin), in: *Anthropogenic Aquifer Recharge: WSP Methods in Water*
589 *Resources Evaluation Series No. 5*, Springer International Publishing, Cham, 517–565, https://doi.org/10.1007/978-3-030-11084-0_16, 2020b.
- 591 Mallick, J., Khan, R. A., Ahmed, M., Alqadhi, S. D., Alsuhbi, M., Falqi, I., and Hasan, M. A.: Modeling Groundwater Potential
592 Zone in a Semi-Arid Region of Aseer Using Fuzzy-AHP and Geoinformation Techniques, <https://doi.org/10.3390/w11122656>,
593 2019.
- 594 Mantoglou, A.: Pumping management of coastal aquifers using analytical models of saltwater intrusion, *Water Resour. Res.*,
595 39, <https://doi.org/https://doi.org/10.1029/2002WR001891>, 2003.
- 596 Miller, J.: Review of Water Resources and Desalination Technologies, <https://doi.org/10.2172/809106>, 2003.
- 597 Nithya, C. N., Srinivas, Y., Magesh, N. S., and Kaliraj, S.: Assessment of groundwater potential zones in Chittar basin,
598 Southern India using GIS based AHP technique, *Remote Sens. Appl. Soc. Environ.*, 15, 100248,
599 <https://doi.org/https://doi.org/10.1016/j.rsase.2019.100248>, 2019.
- 600 Osiakwan, G. M., Gibrilla, A., Kabo-Bah, A. T., Appiah-Adjei, E. K., and Anornu, G.: Delineation of groundwater potential
601 zones in the Central Region of Ghana using GIS and fuzzy analytic hierarchy process, *Model. Earth Syst. Environ.*, 8, 5305–
602 5326, <https://doi.org/10.1007/s40808-022-01380-z>, 2022.
- 603 Oude Essink, G. H. P.: Improving fresh groundwater supply—problems and solutions, *Ocean Coast. Manag.*, 44, 429–449,
604 [https://doi.org/https://doi.org/10.1016/S0964-5691\(01\)00057-6](https://doi.org/https://doi.org/10.1016/S0964-5691(01)00057-6), 2001.
- 605 Pham, Q. N., Ta, T. T., Le Tran, T., Pham, T. T., and Nguyen, T. C.: Assessment of Saltwater Intrusion Vulnerability in the
606 Coastal Aquifers in Ninh Thuan, Vietnam BT - Global Changes and Sustainable Development in Asian Emerging Market
607 Economies Vol. 2: Proceedings of EDESUS 2019, edited by: Nguyen, A. T. and Hens, L., Springer International Publishing,
608 Cham, 703–712, https://doi.org/10.1007/978-3-030-81443-4_45, 2022.
- 609 Phin, T. T., Hoa, D. T. B., Trong, T. D., Hai, D. T., and Que, P. T. N.: Mapping vulnerability water supply in Rach Gia city
610 due to saline intrusion on using analytical hierarchy process, *Sustain. Water Resour. Manag.*, 8, 137,
611 <https://doi.org/10.1007/s40899-022-00712-2>, 2022.
- 612 Pramada, S. K., Minnu, K. P., and Roshni, T.: Insight into sea water intrusion due to pumping: a case study of Ernakulam
613 coast, India, *ISH J. Hydraul. Eng.*, 27, 442–451, <https://doi.org/10.1080/09715010.2018.1553642>, 2021.
- 614 Presley, A.: ERP investment analysis using the strategic alignment model, *Manag. Res. News*, 29, 273–284, 2006.
- 615 Qi, S.-Z. and Qiu, Q.-L.: Environmental hazard from saltwater intrusion in the Laizhou Gulf, Shandong Province of China,
616 *Nat. Hazards*, 56, 563–566, <https://doi.org/10.1007/s11069-010-9686-3>, 2011.
- 617 Reichard, E. G. and Johnson, T. A.: Assessment of regional management strategies for controlling seawater intrusion, *J. Water*
618 *Resour. Plan. Manag.*, 131, 280–291, [https://doi.org/10.1061/\(ASCE\)0733-9496\(2005\)131:4\(280\)](https://doi.org/10.1061/(ASCE)0733-9496(2005)131:4(280)), 2005.
- 619 Rotz, R.: Hydrogeologic Properties of Earth Materials and Principles of Groundwater Flow, *Groundwater*, 59,
620 <https://doi.org/10.1111/gwat.13085>, 2021.
- 621 Saaty, T. L.: Axiomatic foundation of the analytic hierarchy process, *Manag. Sci.*, 32, 841–855., 1986.
- 622 Sajil Kumar, P. J., Elango, L., and Schneider, M.: GIS and AHP Based Groundwater Potential Zones Delineation in Chennai
623 River Basin (CRB), India, *Sustainability*, 14, <https://doi.org/10.3390/su14031830>, 2022.
- 624 Shao, Z., Huq, M. E., Cai, B., Altan, O., and Li, Y.: Integrated remote sensing and GIS approach using Fuzzy-AHP to delineate
625 and identify groundwater potential zones in semi-arid Shanxi Province, China, *Environ. Model. Softw.*, 134, 104868,
626 <https://doi.org/https://doi.org/10.1016/j.envsoft.2020.104868>, 2020.
- 627 Shi, L. and Jiao, J. J.: Seawater intrusion and coastal aquifer management in China: a review, *Environ. Earth Sci.*, 72, 2811–
628 2819, <https://doi.org/10.1007/s12665-014-3186-9>, 2014.



- 629 Singh, Murty, H. R., Gupta, S. K., and Dikshit, A. K.: Development of composite sustainability performance index for steel
630 industry, *Ecol. Indic.*, 7, 565–588, <https://doi.org/https://doi.org/10.1016/j.ecolind.2006.06.004>, 2007.
- 631 Singh, A.: Managing the environmental problem of seawater intrusion in coastal aquifers through simulation–optimization
632 modeling, *Ecol. Indic.*, 48, 498–504, <https://doi.org/https://doi.org/10.1016/j.ecolind.2014.09.011>, 2015.
- 633 Srinivasamoorthy, S. G. A.-S. G. A.-K.: Application of Geophysical and Hydrogeochemical Tracers to Investigate Salinisation
634 Sources in Nagapatinam and Karaikal Coastal Aquifers, South India, *Aquat. Procedia*, v. 4, 65-71–2015 v.4,
635 <https://doi.org/10.1016/j.aqpro.2015.02.010>, 2015.
- 636 Sutar, A. and Rotte, V.: Prevention of Saltwater Intrusion: A Laboratory-Scale Study on Electrokinetic Remediation, 389–400,
637 https://doi.org/10.1007/978-981-16-5501-2_31, 2022.
- 638 Vaidya, O. S. and Kumar, S.: Analytic hierarchy process: An overview of applications, *Eur. J. Oper. Res.*, 169, 1–29,
639 <https://doi.org/https://doi.org/10.1016/j.ejor.2004.04.028>, 2006.
- 640 Yang, H., Jia, C., Li, X., Yang, F., Wang, C., and Yang, X.: Evaluation of seawater intrusion and water quality prediction in
641 Dagu River of North China based on fuzzy analytic hierarchy process exponential smoothing method, *Environ. Sci. Pollut.*
642 *Res.*, 29, 66160–66176, <https://doi.org/10.1007/s11356-022-19871-y>, 2022.
- 643 Zghibi, A., Mirchi, A., Msaddek, M. H., Merzougui, A., Zouhri, L., Taupin, J.-D., Chekirbane, A., Chenini, I., and Tarhouni,
644 J.: Using Analytical Hierarchy Process and Multi-Influencing Factors to Map Groundwater Recharge Zones in a Semi-Arid
645 Mediterranean Coastal Aquifer, *Water*, 12, <https://doi.org/10.3390/w12092525>, 2020.
- 646 Zhou, X., Chen, M., and Liang, C.: Optimal schemes of groundwater exploitation for prevention of seawater intrusion in the
647 Leizhou Peninsula in southern China, *Environ. Geol.*, 43, 978–985, <https://doi.org/10.1007/s00254-002-0722-9>, 2003.
- 648
- 649