# HESS Opinions: The unsustainable use of groundwater conceals a "Day Zero"

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14 Abstract. Water scarcity is a pressing global issue driven by increasing water demands and changing climateie conditions. 15 Based on novel estimates of water availability and water use in Chile, we examine the challenges and risks associated with groundwater (GW) withdrawals in the country's central-north region-focusing on the case of central north Chile (27–35°S), 16 where extreme water stress conditions prevail. As total water uses within a basin approaches the renewable freshwater 17 18 resources, the dependence on GW reserves in unsustainable ways intensifies. This overuse has consequences that extend 19 beyond mere resource depletion, manifesting into environmental degradation, societal conflict, and economic costs. We argue that the "Day Zero" scenario, often concealed by the uncertain attributes hidden nature of GW resources, calls for a 20 reconsideration of water allocation rules and a broader recognition of the long-term implications of unsustainable GW use. 21 22 Our results offer insights for regions worldwide facing similar water scarcity challenges and emphasize the importance of 23 proactive and sustainable water management strategies.



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#### 26 1 Introduction

The risk of water scarcity in a basin escalates increases as the water demand approaches the available renewable freshwater 27 28 resources, hereafter - referred to as water availability hereafter. Water availability at the basin scale is represented, understood 29 by the difference between total precipitation and natural evapotranspiration (ET), excluding land use-induced perturbations. 30 When the ratio of total water uses to water availability, known as the water stress index (WSI) – exceeds 40%, a basin is 31 considered as highly water stressed (Oki and Kanae, 2006). In extreme cases, when the curves-time series of water availability 32 and total water demands within a basin intersect or are get too close to each other or intersect (i.e., WSI = 100%), a "Day Zero"  $(D_{\theta})$  may occur. During a  $D_{\theta}$ -event, during which water cuts are applied to prioritize water access for human consumption-33 34 and where the ecological flows to maintain the well-being of ecosystems cannot be safeguarded. The concept of  $D_0$  has faced 35 criticism from the scientific community (Warner and Meissner, 2021), attributing it with a sensationalist use and lack of 36 scientific robustness, among other concerns. Nevertheless, it generates a sense of urgency that puts pressure on decision-37 makers to take actions.

38 A well-known  $D_0$  event almost happened in 2018 in South Africa in the context of a severe multi-year drought, when Cape 39 Town (South Africa) was at risk of being the first major city worldwide to run out of drinking water due to the low dam levels 40caused by a severe multi-year drought. The announcement of an imminent that the  $D_0$  would arrive on a specific date (12 April 2018, estimated based on the remaining water stored in the reservoir and the water use requirements from the city) triggered 41 42 water saving strategies that -along with the arrival of winter precipitation that interruptinged the multi-year drought - allowed 43 the city to avoid drastic water cuts (Maxmen, 2018; Burls et al., 2019). Another near D<sub>0</sub> scenario situation happened occurred in July 2023 in Montevideo, capital of-(Uruguay). Due to low water reserves in the main reservoirs that supply provide 44 45 drinking water to the city, the water supply was replaced with brackish desalinated water from Rio de la Plata. As a

46 consequence, the metropolitan area of Uruguay was-receiveding non-drinkable water from their taps. Public opinion argued

47 <u>criticized that this measure, claiming it concealed masked the a critical  $D_0$  situation by, as it has avoidinged supply cuts at the</u>

48 expense of providing non-potable water (Gudynas, 2023).

Besides some emblematic-well-known cases in metropolitan areas (Ahmadi et al., 2020), many basins around the globe are approaching or have <u>already</u> reached <u>the point where the intersection between</u>-water uses <u>intersects with or surpasses and-water</u> availability, especially in water-limited regions with intensive irrigation (Oki and Kanae, 2006). In those cases, water needs are usually met by exploiting <u>groundwater (</u>GW) reservoirs <u>at rates that exceed natural recharge, which is in</u>-unsustainable ways-in the long term, i.e., with withdrawal rates above GW natural recharges (de Graaf et al., 2019). In<u>deed-faet</u>, many societies have <u>sustained supported</u> agricultural and population growth by pumping GW storage (Bierkens and Wada, 2019).

55 Previous An increasing number of studies are studies have raisinged the alarm about on the unsustainable use of GW use worldwide, highlighting potential threats to water security and food security highlighting the challenges and risks of this 56 practice (e.g., Gleeson et al., 2016; Schwartz et al., 2020). These studies recognize that accessing GW savings is crucial for 57 58 addressing water scarcity, especially during droughts. However, this depending on GW savingsituation becomes precarious risky when over-extraction renders it a scarce or depleted<del>GW ceases to be an accessible</del> resource-due to unsustainable 59 60 extraction. Although assessing the limits beyond which GW extraction becomes unsustainable remains an ongoing research and technical challenge, we know that pumping rates exceeding the natural GW recharge leads to long-term decline in reserves 61 beyond those expected from climate variability (Bierkens and Wada, 2019). From that perspective, the evidence of a long-62

63 term decline in water table can be used to infer or anticipate an unsustainable GW use in a basin.

Here, we build upon previous the above evidence and propose emphasize that a  $D_{\theta}$  scenario, –which typically triggers alarms 64 and immediate management responses when is associated to surface reservoir depletions, may be concealed by the-an 65 66 unsustainable use of GW. To illustrate this, we focus on the central-north region of Chile, where GW overexploitation and its potential consequences have recently been reported by Taucare et al. (2024) and Jódar et al. (2023). By following different 67 approaches, those studies assessed the effects of climate variability and GW withdrawals on GW level changes over the last 68 69 four decades (1979-2020), and drew similar conclusions regarding the dominant role of water withdrawals in GW declines. 70 Water withdrawals in those studies were computed based on allocated water use rights -the legal entitlements that allow users 71 to extract water-, which is a common approach adopted in water resources assessment studies in Chile due to the lack of actual 72 water use estimates. A caveat of this approach is that it may introduce large uncertainties as water use rights do not necessarily 73 represent actual water uses but rather the fixed monthly or annual flows that are legally permitted for use (see Sect. 5).

- 74 Our study complements these recent findings by making use of novel estimates of water use and availability developed by the
- 75 Center for Climate and Resilience Research (Alvarez-Garreton et al., 2023b), covering a large period (1960-2020). We reflect
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76 on how this-the unsustainable GW usesituation carries risks usually unforeseen or neglected, even before a  $D_0$  is reached, and

77 discuss how this may poses an intergenerational justice dilemma. Finally, we identify key water management limitations, from

78 general perceptions of GW to specific management instruments, that should be addressed to mitigate the impacts of GW

79 overexploitation. While the argument is framed using Chile as a case studyn example, the approach and conclusions can be

80 applied to any region experiencing significant water stress and unsustainable groundwater <u>GW</u> usage.

## 81 <u>2 A novel dataset to illustrate the case of Chile</u>

- 82 The data products used here were developed by the Center for Climate and Resilience Research (Alvarez-Garreton et al.,
- 83 2023b) to evaluate the long-term evolution of water availability, land use and water use across continental Chile, and are
- 84 available for exploration at www.seguridadhidrica.cr2.cl. Time series of water availability were computed for each basin in
- 85 Chile as the difference between catchment-scale annual precipitation and near-natural ET. Precipitation comes from the
- 86 <u>CR2MET meteorological dataset, which provides daily time series in a regular 1/20° grid (~ 5-km) for the period 1960-2020</u>
- 87 (Boisier, 2023). The near natural ET, with the same temporal and spatial resolutions, is estimated from a simple 'bucket' water
- 88 balance model. This model is driven by CR2MET meteorological inputs and a static land cover representative of the year 1950,

89 assuming no irrigation. The model includes parameters related to the soil water holding capacity, root depth, canopy rainfall

90 interception, maximum ratio of actual to potential evapotranspiration (akin to crop coefficients), and phenological cycle (leaf

- 91 area index) across 61 natural and anthropic land cover classes.
- 92 The water use dataset provides annual time series for the period 1960-2020 of consumptive and non-consumptive sectorial 93 water uses. For the LULUCF sector (Land Use, Land Use Change and Forestry), water uses were computed over the same grid 94 and ET model previously described, but taking into account the changes in land cover and irrigation practices over the study 95 period. Water consumption from industrial sectors and drinking water were estimated at the level of communal units, considering specific drivers (e.g., population, energy, or mining production) and water use coefficients, following a 96 97 methodology similar to DGA (2017). Some of the key novelties of this product is that it is based on actual socioeconomic 98 activities, including the evolution of LULUCF, industrial production and population, and that land use-related water 99 consumption estimates are consistent with the climate data. It is important to note that these water use estimates are independent 100 from the allocated water use rights, thereby addressing an important information gap in Chile and enabling a proper contrast 101 of climate variability, water availability, and water uses. This, in turn, facilitates the assessment of the historical evolution of 102 water stress and groundwater overexploitation, including an examination of its causes and consequences.

# 103 The variations in GW levels are obtained from the observation wells network maintained by the Water Directorate, which is 104 publicly accessible from https://snia.mop.gob.cl/BNAConsultas/reportes.

105 To illustrate the relationship between water uses, water availability (with their corresponding WSI) and a concealed  $D_0$ 106 situation, we focus on 10 major basins in the central-north region of Chile (Fig. 1), and present the case of the country's where region experienceing a number of water conflicts have been experienced due to rising demand and reduced water availability 107 108 related to climate change variability (e.g., Barría et al., 2021b; Muñoz et al., 2020), and where GW overexploitation has been 109 recently reported (Taucare et al., 2024; Duran-Llacer et al., 2020; Jódar et al., 2023). We first discuss the situation-case of of the Maipo basin, which encompasses houses the extensive metropolitan large urban area of Santiago, and subsequently expand 110 111 then move to the broadera larger region in of central-north Chile to illustrate GW-the argumentuse from a wider perspective. 112 We relate the GW overuse to water management practices and provide recommendations to improve them.



114Figure 1: Location of the Ten-10 major basins in central-north Chile. T-and the land cover obtained corresponds to from (Zhao et115al., 2016). The maps show the GW observations wells that have at least 10 years of observations since 1960. Each observation well is116colored by their mean depth computed for their complete period of record.

#### 117 2-3 A concealed "Day Zero" scenario in Santiago

118 The metropolitan area of Santiago, located in the Maipo River basin in central Chile (Fig. 1), houses home to nearly six million inhabitants (<u>concentrates-30%</u> of the Chilean population). The city is located in the Maipo basin in central Chile (Fig. 1). 119 According to the water use dataset (Sect. 2) recently developed by the Center for Climate and Resilience Research (available 120 at https://seguridadhidrica.cr2.el), the Maipo basin's currently has a total water consumption is approximately of around 75 m<sup>3</sup> 121 122  $+s^{-1}$ , accounting representing for 15% of Chilethe country's total consumptive water useage. About 60% of the basin's water 123 consumption comes-is attributed tofrom the irrigated agricultural sector, while 35% is allocated for sanitation and drinking 124 water. 125 Since the mid-20<sup>th</sup> century, <del>Over the last decades, total water uses hashave</del> continuously increased and, in the last decade, has

approached water availability in the Maipo basin (Fig. 2). These extreme water stress conditions (WSI close to 100%) emerge in <u>thea</u> context of a protracted drought spanning more than a decade since 2010, with precipitation deficits ranging between 20 to 70% (Garreaud et al., 2019, 2017). The so-called megadrought in central Chile is partially controlled by natural climate variability, but also weighted by a long-term drying trend affecting the subtropical South Pacific region (Boisier et al., 2016, 2018). <u>This trend, asAs</u> an anthropogenic climate change signal, <u>this trend</u> is <u>projected expected</u> to <u>persist in the</u> <u>comingeontinue over the next</u> decades, <u>in line along</u> with global climate <u>pathways projections</u> (IPCC, 2022).

Yet, the causes of the extreme water stress in the Maipo basin are not driven-solely <u>attributable to by</u>-climate variability and droughts. <u>The water demand, d</u>Driven by the expansion of irrigated agriculture, <u>water usage in central Chile</u>-has been maintained or <u>even increased in central Chile continued to rise</u> over <u>recentthe last</u> decades, despite the diminishing water availability. To a large extent, water use in this region has been sustained at the expense of GW resources, as <u>shown-indicated</u> by the increasing <u>water use rights</u> allocation for <u>of water use rights for GW abstractionswithdrawals (Fig. 2a). Larger GW</u> extractions and lower recharge rates are <u>, and</u> further evidenced by the <u>sustained</u>-ongoing depletion of GW levels (Fig. 2b).



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139 Figure 2: Panel a) shows the time series of water availability, water uses and water use rights (WUR) for the Maipo basin. Surface 140and underground WUR were obtained from CAMELS-CL dataset (Alvarez-Garreton et al., 2018). The water uses were obtained 141 upon request from the CR2WU product available at https://seguridadhidrica.cr2.cl. Panel b) shows the GW level anomalies of 89 142 observation wells located in the Maipo basin, computed as the difference between GW levels and the mean level for the 1980-2010 143 (baseline adopted to avoid including the megadrought period). The median (solid lines) and standard deviation (shaded area) of the 144 GW level anomalies are plotted when at least five observations were available. Following well construction guidelines (INDAP, 2010), 145 oObservation wells were classified as shallow and or deep wells if their mean annual GW levels (shown in Fig. 1) were above or below 15 m, respectively. The GW observation data were obtained from the Water Directorate website 146 147 https://snia.mop.gob.cl/BNAConsultas/reportes.

148 Unlike estimating the time until a surface reservoir runs out, such as in the Cape Town and Montevideo cases, there is large

149 significant uncertainty in estimating the time remaining before when a GW reserve is will be exhausted (or reduced to a state

150 that is brought to a practical unrecoverable state within human time frames), leading to and an absolute  $D_0$  scenariois reached.

151 Determining that time frame requires a precise quantification of the water volume remaining in aquifers, GW recharges, and

152 GW extraction rates. <u>However, a tentative estimate of the absolute D<sub>0</sub> can be derived as:</u>

153 Eq. 1 
$$D_o = \frac{W}{U_{GW} - R_N}$$

where *W* represents the water volume of the saturated aquifer,  $U_{GW}$  the groundwater extraction rate, and  $R_N$  the net GW recharge rate (water recharges minus springs). Previous studies have estimated a volume of water of about 30 km<sup>3</sup> for the main aquifer in the Maipo basin (Araneda et al., 2010; DGA, 2021) with net GW recharge rates (including springs) in the range of 10 to  $230 \text{ m}^3 \text{ s}^{-1}/\text{s}$  (Döll and Fiedler, 2008; Riffo, 2022). If we consider present water uses of 75 m<sup>3</sup> s<sup>-1</sup>/s and a ratio of underground to total water uses ranging between 350 to 65% (upper bound corresponds to the ratio of GW to total water use rights in the Maipo basin, Fig. 2), the absolute  $D_0$  time frame would range between 50 to 200 years, depending on the values considered for recharge and underground GW to total water use ratio.

161 These Consistent with this estimated range, another study for the Maipo basin forecasts a 33% reduction of GW storage for 162 the 2020-2050 period in comparison to its 1990-2020 value are rough estimates computed from variables that are challenging 163 to accurately estimate and monitor, making it difficult to assess the risk of overconsumption. However, they provide an order 164 of magnitude of several decades to a few centuries to deplete the <u>underground GW</u> sources in the Maipo basin and reaching 165 an absolute  $D_0$  in the capital of Chile.

In contrast to the short time frames associated to a  $D_{\theta}$  caused by depleted surface reservoirs in some major cities (e.g., Cape Town, Montevideo), several decades may <u>sound seem</u> like plenty of time to prepare for an absolute  $D_{\theta}$  in Chile. However, such absolute  $D_{\theta}$  is far more critical than running out of surface reserves since GW reservoirs <u>may taketake a much</u> significantly-longer time to <u>replenishrecover</u>.

In addition to the risks discussed in sections 3 and 4, tThis situation represents-poses an intergenerational justice dilemma, as since we are depleting the waterusing savings from of previous generations and spending them in short-term economic activities whose benefits may not be perceived perceived by future by next generations (Hiskes, 2009). Conversely, limiting the current generation's use of GW to sustainable levels entails, brings about costs for the today's societypresent generation that cannot be easily offset by the benefits that this sustainable use would have in future \_\_\_\_\_and potentially wealthier\_\_\_\_\_\_ generations (e.g., Andersen et al., 2020).

The use of fossil water (i.e., resources that entered the aquifers centuries or millennia ago) would be a <u>paragon</u> example of this intergenerational dilemma. The quantification of fossil water in arid and semi-arid Chile has not been fully addressed. Still, there is evidence indicating that a part of GW reserves in several basins derive from late glacial climate conditions (Gayo et al., 2012; Moran et al., 2019; Viguier et al., 2018). <u>Available evidence from a headwater section of the Maipo basin</u> (Mapocho basin) have reported deep wells (water table below 112 m depth) characterized by low tritium concentrations (<0.8) and <sup>14</sup>C contents (70 pmc), indicating that such resources were likely recharged about 3,000 years ago (Iriarte et al., 2009).

These GW age tracers indicate that GW with modern and old fractions could also occur in the downstream area of the Maipo
 basin.

184 Furthermore, Regardless of time frame for D<sub>0</sub>, the depletion of GW levels, as observed in both shallow and deep wells (Fig. 185 2.b), cause environmental degradation, social conflict and economical. For example, wetlands may become disconnected from 186 their groundwater source and deeper wells are required to access the water table, among other negative consequences. In 187 Santiago, a recent adaptive measure taken by governmental and water supply agencies to face address water scarcity has been 188 the construction of deep pumping wells (reaching up to 300 m depth in Santiago) for to supplying drinking water for human 189 consumption. While this measure is timely in the short term, helping to mitigate the risks of supply shortages in a highly 190 populated region-with nearly six million inhabitants, it also encourages promotes the use of GW sources that might be, at best. 191 very hard-difficult to recoverreplenish within human timescales. This, in turn, will likely exacerbate GW depletion.

#### 192 **3-4** The non-sustainable GW use in <u>central-north</u> Chile

193 Similar-Liketo the Maipo basin, consumptive water usease in most basins located in central-north Chile (27-35°S, Fig. 1) has 194 are nearing reached or exceedinged their water availability (Fig. 3 and Fig. A1). The definition of water availability entails the 195 maximum potential of the integrated long-term fresh-water flow incoming to a basin, which could potentially be much higher 196 than the long-term GW recharge. Therefore, the intersection of water uses and availability curves reveals a structural 197 imbalance: permanent water uses rely more on depleting GW storage rather than on renewable sources. This situation will 198 likely persist or worsen in the medium term due to the expected reduction in precipitation and GW recharge. Current climate models project a reduction recharge resulting from the negative precipitation trends projected in this regionClick or tap here 199 200 to enter text. where an average of 10 to 30% in annual precipitation by the end of the 21st century under less annual 201 precipitation is projected by the end of the 21<sup>st</sup> century from climate models with high (SSP1-RCP2.6) and low (SSP3-RCP7.0) 202 greenhouse gas emission mitigation scenarios, respectively (Alvarez-Garreton et al., 2023b; IPCC, 2022). At the same time, 203 water useage will likely increase with population and economic growth (Meza et al., 2014).



Figure 3: Panel a) shows the WSI of ten major basins of central-north Chile, computed as water use to availability ratio for the periods 1980-2020 and 2000-2020. The basins are sorted from north to south. The two most arid ones (Copiapó and Huasco) have WSI values above the upper WSI limit, so their values are written beside the bars. The categories of water stress were obtained from Oki and Kanae, (2006). Panel b) shows the trend of GW levels of shallow and deep observations wells computed for the period 1980-2020.

eExcept for the most arid <u>one-basin (Copiapó)</u>, Before year 2000, all basins from in central-north Chile-Fig. 1, maintained a moderate level of water stress <u>during the 1980-2000 period (Fig. 3a</u>). However, during the 2000-2020 period, <u>nine 9</u> out of the ten <u>10</u> basins shown in Fig. 1 moved towards a high or extreme water stress (Fig. 3a). The Rapel basin is the sole exception, maintaining a moderate stress level over the past two decades. This basin, as well as the Maipo basin, exhibit a Mediterraneantype elimate and boast has the largest water availability compared to the other eight basins (refer to Fig. A1), which – allows for sustaining large water uses (similar to those of the Maipo basin) at lower water stress levels. However, unlike the Maipo basin, Furthermore, the Rapel basin has a lower total water consumption, and its water supply in the Rapel basin relies less on
 GW sources, as indicated by the lower allocation of GW use rights in the basin, which also explains its lower GW declines in

219 <u>Fig. 3b</u>.

For the rest of the basins, Tthe large WSI values displayed in Fig. 3a align with the GW levels-declines observed in these basins over the past few-four decades (Fig. 3b), and the growing trend in the allocation of water use rights from underground GW sources (Fig. A1). The GW declines in this region have been reported in previous studies (Pizarro et al., 2022; Valois et al., 2020; Duran-Llacer et al., 2020; Jódar et al., 2023; Taucare et al., 2024). Within the study areaHere, the most significant case is the mean reduction of up to 6 m per/decade in the Aconcagua basin (Fig. 3b).

225 Despite the current water stress levels observed in Fig. 3a, the decreased water availability driven by the megadrought over 226 the last decade has brought a relatively low direct cost to the Chilean economy (Fernández et al., 2023). This couldmay be 227 partly explained by the extensive explained by the intense and widespread use of GW in the region and, as well as the country's robust infrastructure-capacity. However, Wwhen shifting from a national to a local scale, the direct impacts of declining GW 228 229 levels become more evident. In particular For example, the necessity need for deeper wells to reach the water table likely has 230 intensified worsens social inequalities, particularly. -I in rural areas, where people reliance on shallow pumping wells is 231 common. Here, -the GW levels decline have has led to interruptions in the water supply for basic needs and small-scale 232 agriculture activities, as reported by previous studies in Petorca and Ligua basins (Duran-Llacer et al., 2020; Muñoz et al., 233 2020). Although this does not involve water cuts in a megacity, it This represents a  $D_0$  condition for those communities.

Furthermore, Eenvironmental impacts may also emerge well before reaching an absolute  $D_{\theta}$  condition. Declining GW levels have the potential to directly impact the ecological integrity of groundwater-dependent ecosystems and may result-cause ain the disconnection between surface and underground-GW water sources (Jódar et al., 2023), which potentiallycan leading-to the drying out of desiccation of rivers, wetlands and lakes. This, as has been reported observed in the Ligua and Petorca (Duran-Llacer et al., 2022, 2020; Muñoz et al., 2020) as well as basins and in the Maipo basin (Barría et al., 2021b).

#### 239 4-<u>5</u> Caveats in water management

Water management in Chile is primarily regulated by the Water Code, promulgated in 1981, with several amendments since then (Congreso Nacional de Chile, 1981, 2022, 2005). This Code defines water as a national asset for public use and regulates its allocation by means of water use rights, which are legal entitlements specifying the quantity and timing of water that can be used by private and public users. While these titles are granted free of charge by the State, once allocated, they represent

244 property assets that can be traded by their owners, with no water market regulations to their buying and selling transactions.

<u>Further details regarding the water management system can be found in</u> Alvarez-Garreton et al. (2023a), Taucare et al. (2024),
Barría et al. (2021a) and Alvez et al. (2020)

247 The estimates of total water uses at present time shown in Fig. 2 and Fig. A1 indicate that, for most of the study basins water

248 uses are within the legal limits -i.e., they do not exceed (i.e., below-the allocated water use rights. However, this evidence

249 aggregated at the catchment scale does not imply the absence of illegal water uses within the basins. In fact, previous studies

250 have reported unauthorized GW extraction in basins in central Chile, such as Maipo (Venegas-Quiñones, 2022), Petorca

251 (Muñoz et al., 2020) and La Ligua (Budds, 2008).

252 While no disregarding the potential occurrence of unauthorized water withdrawals at specific locations, our findings imply ),

253 which indicates that the problem at the basin scale is not due to illegal overuse, but rather to the current water allocation scheme

is being inadequate failing to prevent water stress conditions, rather than being an issue of illegal overuse. To elaborate on this,
 we highlight some aspects outlined in the Water Code that help us to understand the <u>causes of the current</u> overallocation of
 water resources:

- Since its amendment in 2005 (Congreso Nacional de Chile, 2005), the Water Code has-stipulatesd that the allocation of surface water use rights must consider the protection of ecological flows. HoweverNonetheless, the specified safeguarded streamflow-value defined by the law is insufficient, as it, sets at a maximum n upper limit of 20% of the mean annual streamflow, is not sufficient to be safeguarded as ecological flow. This implies total water usage exceeding 80% of the water availability (i.e., WSI > 80%), which is associated to an extreme water stress condition (Alvarez-Garreton et al., 2023a).
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   2. The allocation of GW use rights does not consider the interactions between the GW and surface systems, nor does it
   264 consider the pre-existing surface water use rights within a given basin.
- Surface and <u>underground <u>GW</u>water use rights are allocated as fixed absolute flows values and do not account for
   long-term, climate-driven changes in water availability (Barría et al., 2021a).
  </u>
- 4. Before the last Water Code modification in 2022, water use rights were allocated in perpetuity. After 2022, new allocations may expire after 30 years only if the central authorities demonstrate they are not being used or are causing water scarcity problems (Congreso Nacional de Chile, 2022). This modification does not apply to water rights granted before 2022, which account for the mostmajority of the water rights allocated in Chile.
- The above dispositions of the Water Code entail an <u>The</u>-inadequate protection of ecological flows, <u>the a</u> disconnected
   management of surface and GW resources, and the failure to account for long term <u>and virtually irreversible</u>-climate change.
   <u>This -resultsallows for-in</u> a tight balance between water uses and availability <u>that can become or a structural condition of</u>
   permanent overuse in a basin. In combination to the drier climatic conditions experienced in central-north Chile over the last

<u>decade, these water management caveats have led, which in turn could lead</u> to unsustainable GW withdrawals (Jódar et al.,
2023; Taucare et al., 2024).-.

277 In addition to the risks discussed in sections 2 and 3, this situation represents an intergenerational justice dilemma since we 278 might be using savings from previous generations and spending them in short-term economic activities whose benefits may 279 not be perceived by next generations. These water management limitations highlighted here promote water overuse, as the only 280 main "signal" that current users have (besides natural water scarcity by drought) is not the price of water extraction (owners 281 of water use rights do not pay for water) but rather the amount of water provided by their the allocated water use rights they own. Despite the evidentthe elear shortcomings of the water allocation system that make it prone to overallocation of water 282 283 rights with respect to natural water availability, addressing these issues is far from straightforward. Segments of public opinion 284 in Chile argue that revising already allocated water use rights, such as limiting their perpetuity condition or adjusting their 285 allocated volumes considering current and future water availability, may introduce legal uncertainties that might harm the economy (Libertad v Desarrollo, 2019). The argument is that the owners of water use rights need to plan their investments and 286 287 revenues based on certain water volumes. Indeed, like private property, the nature of water rights regulation allows owners to 288 mortgage their water entitlements to obtain a loan from the bank (Muchnik et al., 1997). Regardless of the various perspectives 289 on this matter, it is important to note that such legal certainty become physically unrealistic if the basin is not able to provide 290 the amount of water stipulated by the allocated water use rights.

#### 291 <u>6 Water management recommendations</u>

292 <u>Climate projections for central Chile suggest that droughts such the one of the 2010 decade will become more frequent. With</u> 293 <u>current extraction rates (a conservative scenario), GW levels will likely continue to decrease, causing socio-economic and</u> 294 <u>environmental impacts, and bringing a major urban area such as Santiago closer to an absolute  $D_0$ . The large uncertainty 295 regarding  $D_0$  estimates as those shown here (50 to 200 years) highlights the urgent need to improve the estimations of GW 296 volume and recharge rates and to account for its uncertainty in decision making.</u>

297 <u>Furthermore, long-term water management plans should consider the risks of GW overexploitation in order to achieve</u>
 298 <u>sustainable development goals and water security. This involves revisiting the Water Code to address specific limitations, such</u>
 299 <u>as those highlighted here. In particular, we recommend to consider the following recommendations:</u>

To updated the formulae used to compute ecological flows and remove its upper threshold of 20% of mean annual
 streamflow.

- 302
   To integrate the allocation of surface and GW use rights in such a way that the total flow granted within a basin does

   303
   not exceed the water uses associated to an extreme water stress level (i.e., WSI < 70%), while accounting for the</td>
- 304 decreased precipitation expected for Chile.

### 305 5-7 Final remarksConclusions

306 This paper argues that Aaccessin tog GW savings is crucial for addressing coping with water scarcity, particularly during 307 periods of droughts. However, when water withdrawals are steadily greater than recharges, GW storage inevitably declines 308 over time. The partial or total complete depletion of GW depletion has present-day detrimental impacts on society and 309 environment, plus potential negative effects extending-that extend beyond the generational-time frame of this generationes, 310 concealing risks for water security that are often underestimated or disregarded. These risks are analogous to those that would exist if water uses relied on a melting glacier or a depleting surface reservoir. However,, but unlike surface resources, designing 311 312 strategies to prevent an absolute  $D_0$  are is a challenging tasks due to the "hidden" nature of underground GW resources. The 313 road towards the absolute  $D_{\theta}$  poses an intergenerational justice dilemma, while crossing several tipping points beyond which 314 social, economic and environmental impacts may become irreversible (Gleeson et al., 2020).

The capacity to access GW allows for tapping into large water volumes, often seen as an additional water source to the one available on the surface, but <u>the maintenance of GW savings is reliant on this volume is constrained by</u> recharge rates <u>from</u> the <u>surface</u>. Given this constraint, water consumption rates that approach or exceed fresh water availability will not be sustainable in the long term, whether the access is <u>from GWunderground</u>, <u>rivers</u>, <u>channels</u> <u>surface</u> <u>based</u>, or <u>through</u> reservoirs.

To move towards a common perspective about the sustainable use of water resources, we recommend revisiting the definition of water availability to explicitly include the sustainable use of GW. It is crucial to <u>cease stop regarding considering potentially</u> non-renewable-GW savings <u>—which may not be renewable</u> within <u>generational-time</u> frames <u>of the generation that benefits from</u> its use\_ as an additional <u>and unlimited source of water-source</u>. This likely implies a change in regulations and the adoption of a set of rewards and sanctions that maintain the system far from overuse tipping points (Castilla-Rho et al, 2017).

Natural water reserves in aquifers or in the form of snow and glaciers, along with artificial savings in reservoirs, primarily contributes to water availability through temporal regulation. From an infrastructure perspective, there are other ways to increasinge water availability can also be achieved though, such as water transfer between basins and desalination of seawater, both of which. These infrastructure solutions have socio-environmental benefits and costs that should be considered in their evaluation.

- 329 The sustained declines of GW reported here, as well as those reported in several regions worldwide (de Graaf et al., 2019),
- 330 represent a critical situation that should trigger effective responses, just as they are triggered when the decline of a surface
- 331 <u>reservoir indicates the arrival of a *D*<sub>0</sub>.</u>

332 For the case of Chile discussed here, climate projections indicate that drought conditions such the one of the 2010s decade will 333 be more frequent. With current extraction rates (a conservative scenario). GW levels will likely continue to decrease, causing 334 socio economic and environmental impacts, and bringing Santiago closer to an absolute but concealed D<sub>0</sub>. The large 335 uncertainty regarding  $D_{\theta}$  estimates as those shown here (50 to 200 years) highlights the urgent need to improve the estimations 336 of GW volume and recharge rates in central Chile and to account for its uncertainty in decision making. The aim of this opinion 337 piece is not to fine-tune this calculation but to underscore that Chile should invest in advancing towards incorporating these 338 principles in policy making. Also, we argue that, even before being able to tackle these challenges and forecast the arrival of 339 an absolute  $D_{\theta_{1}}$  the declining of GW levels will likely have impacts on society, local economy and environment well before reaching an absolute Da, which calls for the implementation of measures to reach a sustainable use of water resources. 340

341 In addition to short term strategies to secure water access, long term water management plans should consider these risks in 342 order to achieve water security goals. This includes revising the Water Code to address specific limitations, such as inadequate 343 protection of ecological flows, the disconnection in managing surface and groundwater resources, and the failure to account 344 for changing water availability over time in the water allocation scheme, all of which contribute to water searcity and overuse. 345 Thus, research and monitoring efforts should focus on advancing our understanding of these systems.

#### 346 Data availability

The water use and availability data was obtained upon request from the Center for Climate and Resilience Research website at <u>https://seguridadhidrica.cr2.cl</u> (last access: 20 September 2023). The GW levels data were obtained from DGA website <u>https://snia.mop.gob.cl/BNAConsultas/reportes</u> (last access: 20 January 2023). The water use rights were obtained from CAMELS-CL dataset (Alvarez-Garreton et al., 2018), available at the Center for Climate and Resilience Research website at

351 <u>https://camels.cr2.cl</u> (last access: 20 September 2023).

#### 352 Author contributions

CAG and JPB conceived the idea, perform the analyses and wrote an early draft of the manuscript. All the authors revised
 early manuscript drafts and wrote the final paper.

#### 355 Competing interests

356 The contact author has declared that none of the authors has any competing interests.

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- Figure A1: Panel a) shows the time series of water availability, water uses and water use rights (WUR) for the ten-10 major basins
- 499 in central-northern Chile. The water availability is computed as the difference between catchment-scale precipitation and total
- 500 evapotranspiration from non-anthropogenic land cover obtained from CR2MET product (Boisier, 2023). Surface and underground
- 501 WUR were obtained from CAMELS-CL dataset (Alvarez-Garreton et al., 2018). The water uses were obtained upon request from the 502 CR2WU product available at https://seguridadhidrica.cr2.cl. Panel b) shows the GW level anomalies of the observation wells located
- 502 CR2WU product available at <u>https://seguridadhidrica.cr2.cl</u>. Panel b) shows the GW level anomalies of the observation wells located 503 in each basin, computed as the difference between GW levels and the mean level for the 1980-2010. The median (solid lines) and
- standard deviation (shaded area) of the GW level anomalies are plotted when at least five observations were available. Following
- standard deviation (shaded area) of the GW level anomales are plotted when at least five observations were available. <u>Following</u> 505 well construction guidelines (INDAP, 2010), Oobservation wells were classified as shallow and deep wells if their mean annual GW
- 506 levels (shown in Fig. 1) were above or below 15 m, respectively. The GW observation data were obtained from the Water Directorate
- 507 website https://snia.mop.gob.cl/BNAConsultas/reportes.
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- 509