



1 **Soil water sources and its implications on vegetation restoration**
2 **in the Three-river Headwaters Region during different ablation**
3 **periods**

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12 **Abstract:** Under climate warming, effective restoration and protection of
13 the ecological environment could happen by timely supplementing soil
14 water. So it is crucial to understand the spatial-temporal changes in soil
15 water sources. Two thousand six hundred samples of soil water,
16 precipitation, river water, ground ice, supra-permafrost water, and glacier
17 snow meltwater have been collected from June, August, and September
18 2020 to quantify the soil water sources in the Three-River Headwater
19 Region under different ablation periods. Results indicated that
20 precipitation, ground ice, and snow meltwater accounted for approximately
21 72%, 20%, and 8% of soil water during the early ablation period. Snow is



22 completely melted in the heavy and the end of the ablation period, and
23 precipitation contributed to about 90% and 94% of soil water, respectively.
24 These recharges also vary markedly with altitude and vegetation type.
25 Various factors influence soil water sources, including temperature,
26 precipitation, vegetation, evapotranspiration, and the freeze-thaw cycle.
27 However, soil water loss will further exacerbate vegetation degradation
28 and pose a significant threat to the ecological security of the “Chinese
29 water tower.” So there is an urgent need to monitor soil water, warn of
30 vegetation degradation associated with soil moisture loss, and identify
31 reasonable water-soil conservation and vegetation restoration patterns.

32 **Keywords:** soil water sources, precipitation, ground ice, three-River
33 Headwater Region

34

35 **1. Introduction**

36 Soil water is a vital water resource, a link between precipitation, surface
37 water, soil water, and groundwater, which is an essential component in the
38 formation, transformation, and consumption of water resources with
39 spatial-temporal scales. It substantially impacts regional water resource
40 distribution patterns, the ecological environment, and river runoff as the
41 key factors in terrestrial hydrological cycles and environmental succession
42 (Gao et al., 2017; Sazibet al., 2020). Soil water plays a fundamental role in
43 controlling the exchange of water and heat between the land surface and



44 atmosphere, which has been widely applied to study the regional
45 microclimate, energy and material balance, and global climate change
46 (Spennemann et al., 2017; Sprenger, Tetzlaff, & Soulsby, 2017). Moreover,
47 soil water is directly involved in physiological activities and promotes
48 productivity and carbon sequestration capacity. It was sensitive to the
49 interaction between soil and vegetation that altered soil physicochemical
50 properties, internal structures, and material composition (Marchionni et al.,
51 2021). Consequently, soil water sources can be affected by many factors,
52 such as climate, vegetation, soil type, and topography (Martinez Garcia et
53 al., 2014). Understanding the spatial-temporal changes in soil water
54 sources is essential for better projection of the water and ecology. So,
55 studying soil water sources is a hot topic in international hydrology and
56 soil science.

57 Research on soil water has progressed in a series of studies related to
58 the hydrometeorological, hydro-climatological, ecological, and
59 biogeochemical processes. Permafrost existence can affect inter-annual
60 changes in soil water, and its degradation, including the increasing active
61 layer thickness and disappearance, would decrease ecosystem resilience
62 (Liu et al., 2021). At high latitudes, active-layer deepening associated with
63 soil water changes occurred over less than 8% of the current permafrost
64 area under climate warming (Zachary et al., 2013). Soil water modulates
65 regional climate from sub-seasonal to seasonal timescales. Zhang et al.



66 (2020) found that drier soil led to a more significant increase in the upper
67 quantile of summer heatwaves frequency than in the lower quantile. Liu et
68 al. (2014) thought wet (dry) initial soil water anomalies reduce (amplify)
69 the drought extremes, diminish (reinforce) the hot extremes, and enhance
70 (reduce) the cold extremes over areas of strong soil water-atmosphere
71 coupling. Qi et al. (2020) found that soil water could decrease by at least
72 20% in March-May if there is no snow. Alexander (2021) also found that
73 despite the high humidity in autumn and the high snow reserves
74 accumulated during winter, soil water decreased after the snow had melted.
75 The soil water movement is an important carrier of the material cycle and
76 energy flow. The horizontal flow weakens in the freezing period due to the
77 terrain slope and the freezing-thawing cycle, whereas the vertical migration
78 of soil water moves and strengthens (Cao et al., 2017). Zhang et al. (2021)
79 investigated the water movement in reconstructed soil and evaluated the
80 effects of mining waste rock on plant growth in an arid-cold region. The
81 interaction between soil water and the ground thaw was more dependent at
82 wetter sites, and the interactive soil water and thaw depth behavior on hill
83 slopes changed with location (Guan et al., 2010). Inter-annual anomalies
84 of soil water and vegetation due to rainfall during a given summer were
85 maintained through the freezing winter to the spring, acting as an initial
86 condition for subsequent summer land-surface and rainfall conditions
87 (Masato and Banzragch, 2011). Based on the observation in the central



88 Tibetan Plateau, the four GLDAS models tend to systematically
89 underestimate the surface soil water (0-5 cm) while well simulated the soil
90 water for the 20-40 cm layer, especially during the soil thawing period
91 (Chen et al., 2013; Li et al., 2019). The vegetation effect and the freezing-
92 thawing cycle may be the significant factors that led to an unsatisfactory
93 performance of the Soil Moisture Active Passive (SMAP) mission
94 (Wagner et al., 2003; Ma et al., 2017). As mentioned above, the
95 quantification of soil water sources is relatively insufficient.

96 Soil water has also been deeply concerned in TRHR. Cao and Jin
97 (2021) analyzed the distribution characteristics of soil water and its
98 relationship with temperature and precipitation in TRHR. The influence of
99 precipitation on soil water in the alpine steppe was greater than that in an
100 alpine meadow, especially in lower-altitude areas (Li et al., 2022). Chen et
101 al. (2021) constructed the spatial-temporal changes in soil water and its
102 influencing factors from 2003 to 2020. Huang et al. (2022) studied the
103 variation of surface soil water in an alpine meadow with different
104 degradation degrees in the study region. Xing et al. (2016) analyzed the
105 groundwater storage changes and their influence on soil water in the TRHR.
106 Guo et al. (2022) thought the main factors influencing soil water changes
107 were NDVI and precipitation, followed by air temperature and wind speed
108 in the sources region of the Yellow river. Land degradation significantly
109 reduced soil water by 4.5-6.1% at a depth of 0-100 cm and increased the



110 annual mean soil surface temperature by 0.8 °C under climate warming in
111 the sources region of the Yangtze river (Xue et al., 2017). Soil water and
112 temperature showed decreasing trends from 0-80 cm and an increasing
113 trend from 80-100 cm (Li et al.,2020). The change of soil water resulted in
114 vegetation degeneration, soil desertification, and leanness in the source
115 regions of the Yangtze River (Wang et al., 2012), and it also had a positive
116 correlation with the average thickness of wind deposition (Song et al.,
117 2019). The TRHR is undergoing a glacier retreat, permafrost degradation,
118 precipitation increase, snowfall decrease, water conservation decrease, and
119 soil erosion intensification under climate warming (Li et al., 2021). These
120 changes have caused large fluctuations of soil water, bringing great
121 uncertainty to vegetation growth and causing challenges in vegetation
122 restoration. So there is an urgent need to quantify the soil water sources to
123 improve the effectiveness of ecological restoration in permafrost regions.

124 However, the field observations are too sparse to satisfy the need for
125 quantifying soil water sources in TRHR. As the natural tracers, stable
126 isotopes can be applied in water cycle studies to trace precipitation, soil
127 water, groundwater, and plant water (Zhang et al., 2017; Wang et al., 2018;
128 Yang et al., 2019; Li et al., 2022). Monitoring the stable isotope
129 characteristics of soil water could provide information about water sources,
130 changes in soil water, and moisture cycling (Sprenger et al., 2017). So
131 based on 2600 samples of soil water, precipitation, river water, ground ice,



132 supra-permafrost water, and glacier snow meltwater collected from June,
133 August, and September 2020, this study (a) analyzes the spatiotemporal
134 distribution of $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in soil water at different ablation stages; (b)
135 discusses the hydrological processes of soil water and its differences; (c)
136 quantifies the major sources and its contributions to soil water; (d)
137 confirms the corresponding implications for ecological protection. The
138 result presents new observational evidence of soil water sources in the
139 “Chinese Water Tower.” It provides a scientific basis for establishing a
140 complex interplay between soil water and vegetation as a theoretical basis
141 for developing water-soil conservation and vegetation restoration patterns
142 in cold regions, especially in the permafrost region.

143

144 **2. Data and methods**

145 **2.1 Study region**

146 Three-River Headwater Region (TRHR) ($31^{\circ}39'-36^{\circ}12'N$, $89^{\circ}45'-$
147 $102^{\circ}23'E$, 2610-6920m a.s.l.) is the source region of Yangtze (YZR),
148 Yellow (YLR), and Lancangjiang Rivers (LCR), which is significant to
149 freshwater resources in China and Asia (Fig.1). The TRHR is $36.3 \times 10^4 \text{ km}^2$
150 and accounts approximately 50.4% of the total area of the Qinghai
151 Province. The region has a plateau continental climate with an annual
152 average temperature of $-5.38-4.14^{\circ}\text{C}$ and annual precipitation of 262.2-
153 772.8 mm (Cao and Pan, 2014). The radiation is abundant, with total



154 annual sunlight as high as 2300-2900 h due to the high altitude. The
155 permafrost is extensively developed and is well distributed in the YZR with
156 a depth averaging between 50 and 120 m, whereas permafrost was
157 discontinuous and sporadic with a depth below 50 m in the YLR and LCR
158 (Zhang et al., 2001b). The YLR, YZR, and LCR cover 167,000 km²,
159 159,000 km², and 37,000 km², accounting for 46%, 44%, and 10% of the
160 total area of TRHR, respectively. The YLR, YZR, and LCR contribute
161 approximately 49%, 25%, and 15% of the total runoff and supply up to 600
162 × 10⁸ m³/a freshwater resources. Additionally, more than 180 rivers, 1800
163 lakes, 2000 × 10⁸ m³ of glaciers, and 73,300 km² of wetlands are present
164 in the TRHR. Protecting the ecology of the TRHR, maintaining and
165 improving its function of water-soil conservation, and water containment
166 are of vital importance to the stable supply of water resources, as well as
167 to climate stability, ecological security, and sustainable economic and
168 social development in Asia. The first largest national park, the Three-River
169 Headwaters National Park, has been built, a restorative practice region for
170 constructing ecological civilization and beautifying China.

171 Grasslands are the main ecosystems and comprise approximately 70%
172 of the regional vegetation area. The grasses are typical for alpine meadows
173 and alpine steppes, dominated by *Kobresia capillifolia*, *Kobresia humilis*,
174 *Stipa purpurea*, *Elymus dahuricus*, etc. Other vegetation types are
175 temperate steppe and alpine desert with small distributions, dominated by



176 *Stipa spp.*, *Achnatherum splendens*, *Carex spp.*, and *Thylacospermum*
177 *caespitosum*, *Androsace tapete*, *Oxytropis sp.*, *Saussurea subulata*,
178 respectively (Fan et al., 2010). Moreover, the ecosystems in the TRHR are
179 characterized by diversity, fragility, sensitivity, weak carrying capacity, and
180 restoration capacity. Most of the soils are thin in thickness and coarse in
181 texture. From high altitude to low altitude, the soil types are alpine desert
182 soil, alpine meadow soil, alpine steppe soil, mountain meadow soil, grey-
183 cinnamon soil, castanozems, and mountain forest soil, respectively. The
184 alpine meadow soil is the primary soil type in the region, and other
185 intrazonal soils are also commonly developed.

186 **2. 2 Data and methods**

187 **2.2.1 Samples: collection and preparation**

188 Primary data was collected through fieldwork in June, August, and
189 September 2020. It was used to explore the seasonal pattern and its
190 influence on soil-water sources. A scientific understanding of vegetation
191 restoration in the “Chinese Water Tower” (Fig.2) was developed from these
192 soil-water sources. A total collection of 2600 samples includes soil water,
193 ground ice, precipitation, river water, supra-permafrost water, and glacier
194 snow meltwater in the ~~Three-river-Headwaters-Region~~, with spatial and
195 temporal frequency (Fig.3). The sampling details are described in the
196 following sections.

197 **Soil samples:** The soil profile was excavated, and its thickness can be



198 determined based on the actual thickness of the soil layer, and the samples
199 were collected at 20 cm intervals from 79, 70, and 93 sampling sites in
200 June, August, and September, respectively (Fig.3). Meanwhile, soil
201 temperature was measured in °C, and the test range was from -40 °C to
202 100°C, with a $\pm 0.5^\circ\text{C}$ accuracy. Soil moisture was measured as a %
203 (m^3/m^3), with a test range of 0 to 100% and a response time of less than 2s.
204 Three parallel samples were collected from each layer for soil water stable
205 isotope analysis. The samples were collected from 2 cm below the surface
206 to avoid the soil samples being influenced by the free atmosphere. ~~Seven~~
207 ~~hundred forty one~~ soil samples were collected and stored in HDPE bottles
208 and sealed with parafilm.

209 **Precipitation samples:** At Zhimenda (92.26°E, 34.14°N, 3540 m),
210 Tuotuohe (34.22°N, 92.24°E, 4533 m), Zaduo (32.53°N, 95.17°E, 4066.4
211 m), Dari (33.45°N, 99.39°E, 3967 m) and Maduo (34.55°N, 98.13°E,
212 4272.3 m) stations, a total of 375 precipitation event-scale samples were
213 collected during from June 2019 to July 2020 in TRHR (Fig.3). All
214 precipitation occurring from 20:00 on the first day to 20:00 the next day
215 was collected from sampling the precipitation event. During sample
216 collection, precipitation, air temperature, wind speed, and relative humidity
217 were recorded at the corresponding national meteorological stations. In
218 order to avoid evaporation, the sample was collected immediately after the



219 event.

220 **Ground ice:** In order to collect the ground ice samples, a 1 m deep soil
221 profile of the active permafrost layer was dug at each of the sampling sites,
222 to look for permafrost ground ice (Fig.3). ~~The~~ 66, 40, and 37 ground ice
223 samples have been obtained on June, August, and September, respectively,
224 in the TRHR, which were preserved in pre-cleaned HDPE bottles sealed
225 with parafilm and kept frozen. The outer layer of each ice sample was
226 chipped off to avoid contamination from the soil.

227 **River water:** The river water has also been collected in TRHR,
228 including 259, 231, and 186 samples in June, August, and September,
229 respectively, to analyze the spatial and temporal relationship between soil
230 and river water. River water samples were collected 20 cm below the river
231 surface and stored in HDPE bottles sealed with parafilm.

232 **Supra-permafrost water:** Supra-permafrost water is mainly stored
233 in the active permafrost layer (Li et al., 2020). To study the hydraulic
234 connection between supra-permafrost water and soil water, 125, 161 and
235 130 samples were collected at different altitudes during June, August, and
236 September of 2020, respectively. First, a 1-m deep profile of the active
237 permafrost layer was manually dug at each sampling site. Second, the
238 collected water samples were immediately filtered with a 0.45- μ m
239 millipore filtration membrane at the bottom of each profile, and then stored
240 in HDPE bottles and sealed with parafilm.



241 **Glaciers snow meltwater:** At Jianggudiru Glacier (91°E,33.45°N,
242 5281 m), Dongkemadi Glacier (92°E, 33°N, 5423 m), and Yuzhufeng
243 Glacier (94.22°E, 35.63°N, 5180 m) in the sources region of Yangtze river
244 (Fig. 1), and Halong glacier (99.78°E, 34.62°N, 5050 m) in the sources
245 region of Yellow river, and Yangzigou glacier (94.85°E, 33.46°N, 5260 m)
246 in the sources region of Lancangjiang river, 27, 32 and 41 samples were
247 collected from streams flowing out of the glacier front during June, August
248 and September of 2020, respectively, and were then stored in HDPE bottles
249 and sealed with parafilm.

250 Before analysis, all samples were stored at 4 °C in a refrigerator
251 without evaporation. Soil water had to be extracted from the soil. We used
252 a cryogenic freezing vacuum extraction system (LI-2000, Beijing Liga
253 United Technology Co., Ltd., China) to extract soil water, as it can achieve
254 complete extraction and has a high precision (Li et al., 2016). The test tubes
255 containing soil samples were installed on the extraction line and frozen
256 with liquid nitrogen. After 10 min, the line was checked to ensure no leaks.
257 After it was completely sealed, the larger test tube was heated using a
258 heating sleeve at 95 °C, and the smaller test tube was frozen with liquid
259 nitrogen (-196 °C). Due to temperature gradients, water vapor moved from
260 the larger test tube to the smaller one and condensed into ice. The
261 extraction process took 2 hr and had an efficiency above 98%. Water



262 samples were analyzed for $\delta^{18}\text{O}$ and ^2H through laser absorption
263 spectroscopy (liquid water isotope analyzer, Los Gatos Research DEL-100,
264 USA) at the Key Laboratory of Ecohydrology of Inland River Basin,
265 Northwest Institute of Eco-Environment and Resources, CAS. The results
266 are reported relative to the Vienna Standard Mean Ocean Water (VSMOW).
267 Measurement precisions for $\delta^{18}\text{O}$ and $\delta^2\text{H}$ were better than 0.5‰ and 0.2 ‰,
268 respectively.

269 In addition, air temperature, precipitation, evaporation, and ground
270 temperature in the TRHR were mainly obtained from the China
271 Meteorological Data Network (<http://data.cma.cn/>). The normalized
272 Vegetation Index (NDVI) is derived from MODIS data, downloaded from
273 the NASA website (<https://search.earthdata.nasa.gov/>), with a spatial
274 resolution of 0.05° and a temporal resolution of 16d, and the data are in
275 HDF format.

276 **2.2.2 Methods**

277 The end member mixing analysis (EMMA) tracer approach has been
278 widely used for analyzing potential soil water sources (Hooper et al., 1990;
279 Hooper, 2003; Gibson et al., 2005; Peng et al., 2012; Li et al., 2014; 2020).
280 The EMMA tracer method assumes that i) the tracer concentration in a
281 potential water source varies significantly in time and space, ii) the
282 chemical properties of the selected tracer are stable, and iii) changes occur
283 as a result of water mixing. Tracer techniques involve graphical analyses



284 in which chemical and isotopic parameters represent the designated end
285 members. Essentially, the changing composition of the studied water likely
286 results from the intersections during its passage through each landscape.
287 Tracers can be used to determine the sources and flow paths. Both the two-
288 and three-component methods can be described by a uniform equation:

$$289 \quad Q_t = \sum_{m=1}^n Q_m, \quad Q_t C_t^j = \sum_{m=1}^n Q_m C_m^j, \quad j = 1, \dots, k \quad (1)$$

290 where Q_t is the total runoff discharge, Q_m is the discharge of component m ,
291 and C_m^j is the tracer j incorporated in the component m . In addition, the
292 global meteoric water line (GMWL), local meteoric water lines (LMWLs),
293 and evaporation line (LEL) have been used to analyze the relationship
294 between soil water and other waters in the TRHR.

295 **3. Results**

296 **3.1 $\delta^{18}\text{O}$ and $\delta^2\text{H}$ of soil water in different ablation periods**

297 Soil water stable isotopes show significant changes in the early ablation
298 period (June), the substantial ablation period (August), and the end of
299 ablation (September). The average value of $\delta^{18}\text{O}$ and δD is relatively higher
300 in June and negative in August. Again it becomes higher in September,
301 while it exhibits an opposite trend for d-excess (Table.1). Two reasons can
302 explain this variation: (1) precipitation gradually increases from June,
303 reaches a maximum in August, and then decreases; (2) the effect of
304 evapotranspiration on soil water also shows seasonal variations. Soil water



305 stable isotopes in different ablation periods show apparent regional
306 differences. This reflects that precipitation is the main source of soil water,
307 and the differences in precipitation stable isotopes are reflected in that of
308 soil water. The temporal variation of stable isotope in 20–80 cm, similar to
309 the TRHR, is progressively negative in the surface soil (0–20 cm). This is
310 due to its high susceptibility to perturbation, and environmental changes
311 (Table.1). Soil water stable isotopes on the eastern slope were gradually
312 negative from June to September, while the other slope directions were
313 consistent with the TRHR (Table.1). Moreover, the strong ablation period
314 of soil water isotopes in meadow and grassland areas was gradually
315 negative from the beginning to the end of ablation, while it was
316 continuously negative in forest areas (Table.1). These facts show the
317 stochastic nature of soil water changes as the indicators of environmental
318 changes.

319 As Fig. 3 shows, the slope and intercept for LEL are the lowest in the
320 heavy ablation period and then in the early ablation period and the end
321 ablation period, which reflects the seasonal variation of the influence from
322 evaporation or non-equilibrium dynamic fractionation. In contrast, these
323 values gradually increase from June to September in the TRHR. The slope
324 and intercept of LEL for the 0–40 cm layer were the lowest during the
325 heavy ablation period, while they were relatively high at the beginning
326 and end of ablation, whereas the slope and intercept of the 40–80 cm layer



327 gradually increased (Fig.4). This reflects that the soil layer above 40 cm is
328 greatly disturbed by the environment. Its variation is more sensitive to
329 environmental changes, while the deeper soil layer is relatively stable. For
330 different altitudes, the slope and intercept of LEL increased continuously
331 from the beginning to the end of ablation at 3000–3500 m and 5100 m,
332 while at 3500–4500 m the heavy ablation period was the lowest and the
333 beginning and end of ablation were relatively high (Table.2). In the
334 grassland, forest and scrub areas, the slope and intercept of LEL are higher
335 during the heavy ablation period and lower at the beginning and end of
336 ablation, while the opposite is evident in the meadow areas (Table.2).
337 More interestingly, the slope and intercept of LEL on the northern and
338 eastern slopes are lower during the heavy ablation period and higher at the
339 beginning and end of the ablation period; on the southern and western
340 slopes, they gradually increase and reach the maximum at the end of
341 ablation period (Table.2). These changes again reflect the multiplicity and
342 complexity of factors influencing soil water, and suggest that conducting
343 soil water source should be predicated on continuous systematic sampling
344 at the regional scale.

345 **3.2 Relationship between soil water and surface waters in different** 346 **ablation periods**

347 In the study region, the LMWL was $\delta^2\text{H} = 7.89\delta^{18}\text{O} + 12.43$ ($R^2 = 0.97$; N
348 $= 375$) based on event-level precipitation. As Fig.5 shows, soil water was



349 primarily located on the LWML, suggesting that precipitation was the
350 major soil water source, and some soil water plotted below the LWML
351 owing to high evaporation. The $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values varied among
352 precipitation, ground ice, and snow meltwater in the early ablation period.
353 This suggests that in June, as the permafrost and snow melt, ground ice
354 meltwater, snow meltwater, and precipitation combine to recharge soil
355 water, and that snow meltwater recharge is mainly in the area above 4000
356 m based on sampling during the expedition. In the heavy ablation period,
357 soil water is located on the LWML in August, with some sampling sites
358 below it because of stronger evaporation (Fig.5). At this time of year, the
359 snowpack has melted away, and the ground ice in the active layer is also
360 melting rapidly, with precipitation and ground ice meltwater recharging the
361 soil water. Soil water lies above the LWML, and the high slope reflects the
362 relatively low influence of evaporation in the end ablation period, while
363 the absence of snow meltwater and the ground ice in the soil has also
364 melted in areas below 4000 m, so precipitation dominates the soil water
365 source (Fig.5). These variations reflect seasonal variability in the soil water
366 sources. They suggest that freeze-thaw cycles are a key influence on soil
367 water variability.

368 Interestingly, soil water, supra-permafrost, and river water show a
369 clustered distribution at all ablation stages in the TRHR, reflecting a close
370 hydraulic connection. Precipitation first recharges soil water due to



371 permafrost distribution, while some soil water transforms into supra-
372 permafrost water. Then some soil and supra-permafrost water recharge the
373 runoff, reflecting the uniqueness of the hydrological process in cold
374 regions. These facts show that various recharge sources with significant
375 seasonal variations influence soil water sources. The relationship between
376 soil water and the LWML varied significantly at different altitudes; **the**
377 **higher the altitude, the lower the left-hand side of LWML above 4000 m**
378 **in the end ablation period, and vice versa, below 4000 m(Fig.5). Reflecting**
379 **the variation of soil water sources at different altitudes in the end ablation**
380 **period, soil water is mainly recharged by precipitation in areas below 4000**
381 **m, while it is also recharged by ground ice meltwater which is strongly**
382 **influenced by evaporation, resulting in a relatively positive soil water**
383 **stable isotope. In the early ablation period, the order of altitude is close to**
384 **the LWML, and as follows: from 3500–4000 m, 4000–4500 m, above 4500**
385 **m, and below 3500 m, confirming the variability of soil water sources in**
386 **different altitudes (Fig.5).** On the one hand, precipitation in the area below
387 4500 m is primarily liquid, while above the area it is mostly snow, which
388 is strongly affected by evaporation when the snow melts, resulting in a
389 relatively positive soil water stable isotope and lower recharge to soil water;
390 on the other hand, precipitation in June is relatively low, while the
391 temperature in the lower altitude area rises faster, and evaporation is strong,
392 which leads to a positive soil water stable isotope. In the heavy ablation



393 period, the distance between the LEL of soil water and the LWML is
394 comparable at different altitudes, being slightly closer below 3500 m and
395 slightly further from the 4000–4500 m, reflecting less altitudinal variability
396 in soil water sources at this time of year, with abundant precipitation
397 dominating the soil water sources and intense evaporation becoming an
398 important factor influencing soil water dynamics (Fig.5).

399 The relationship between soil water and the LWML also varied
400 significantly by vegetation, with grassland being farthest from the LWML,
401 followed by meadows and frosts at the early and end ablation period. In
402 contrast, it is farthest for meadows, followed by grassland and forests in a
403 heavy ablation period (Fig.5). These variations indicate that: (1) forests
404 have relatively little effect on shallow soil water due to the predominant
405 use of groundwater and the lower effect of evapotranspiration under the
406 shade of the trees; (2) under the relatively low precipitation, the low soil
407 water in grassland, combined with the effect of evapotranspiration, results
408 in relatively positive soil water stable isotopes; (3) soil water stable
409 isotopes are positive when the meadow is growing, and evapotranspiration
410 is intense under the abundant precipitation season. Evapotranspiration
411 mainly dominates the influence of vegetation on soil water sources. These
412 changes indicate the stochastic nature of the soil water sources and the
413 multiplicity of influencing factors.

414 3.3 Soil water sources in different ablation periods



415 Based on the EMMA model, there were significant differences in the d-
416 excess and $\delta^{18}\text{O}$ concentrations of ground ice, precipitation, snow
417 meltwater, and soil water during different ablation periods (Fig.6).
418 Accordingly, these $\delta^{18}\text{O}$ and d-excess data were selected for analysis
419 because they could effectively characterize the sources. There were large
420 spatiotemporal variations in the $\delta^{18}\text{O}$ and d-excess concentrations. Soil
421 water was plotted on a triangle spanning the three end members, suggesting
422 that soil water was a mixture of them in the early ablation period (Fig. 6).
423 Therefore, precipitation was considered as the first end member, and
424 ground ice as the second end member, and snow meltwater was considered
425 as the third end member. Whereas soil water was plotted on a straight line
426 spanning the two end members, suggesting that soil water was a mixture
427 of precipitation and ground ice in the heavy and end ablation periods (Fig.
428 6). ~~The intersection between the LWML and the LEL is considered to be
429 the isotopic value of the initial water body that recharges the soil water.~~
430 The intersection between the LWML and the LEL is considered to be the
431 isotopic value of the initial water body that recharges the soil water, and
432 the corresponding $\delta^{18}\text{O}$ and $\delta^2\text{H}$ are -17.63‰ and -127.61‰, -18.81‰
433 and -136.94‰, -23.04‰ and -170.36‰ during the early, heavy and end
434 ablation period in the TRRH, respectively. These values are extremely
435 close to the corresponding mean monthly precipitation values, reflecting



436 that precipitation is the main source of soil water.

437 Based on the calculation, precipitation, ground ice, and snow
438 meltwater account for approximately 72%, 20%, and 8% of soil water,
439 respectively (Fig.7). Moreover, the recharge pattern shows a clear
440 ~~difference in altitude~~, with no snow meltwater recharge below 4000 m due
441 to snow melting depletion and a higher snow meltwater recharge at higher
442 elevations. The maximum of ground ice meltwater recharge occurs at
443 3500–4000 m and decreases with increasing altitude. This reflects that the
444 change in altitude of snow and ground ice meltwater is a key factor
445 affecting the source of soil water during the early ablation period.
446 Regarding different vegetation types, ground ice meltwater is higher in
447 meadow areas. In contrast, snow meltwater recharge is relatively high in
448 grassland areas and mainly in precipitation recharge in forest areas. Ground
449 ice and snow meltwater recharge is significantly higher on shaded slopes
450 than on sunny slopes (Fig.7).

451 In the heavy ablation period, precipitation and ground ice accounted
452 for approximately 90% and 10% of soil water in the TRHR, respectively.
453 Snow is completely melted at this time of year, and the recharge of soil
454 water by precipitation decreases with increasing altitude, while ground ice
455 meltwater gradually increases, ~~with all below 3500 m being recharged by~~
456 ~~precipitation~~. The forested area is fully recharged by precipitation, while
457 the meadow area is recharged by ground ice meltwater higher than the



458 grassland area, and the shaded slope is also larger than the sunny slope
459 (Fig.7).

460 According to the EMMA, precipitation and ground ice accounted for
461 approximately 94% and 6% of soil water in the TRHR, respectively, during
462 the end ablation period. All ground ice in soils below 4000 m at this time
463 of year is ablated away, and all soil water is recharged by precipitation,
464 with a small amount of ground ice water recharge occurring in the higher
465 altitude areas. There is only a small amount of recharge from ground ice
466 meltwater on shady slopes, which is still higher in meadow areas than in
467 grassland areas (Fig.7).

468 **4. Discussion**

469 **4.1 Influencing factors on soil water sources in different ablation** 470 **periods**

471 Various factors influence soil water sources, including temperature,
472 precipitation, vegetation, evapotranspiration, and the freeze-thaw cycle. As
473 mentioned above, soil water is mainly recharged by precipitation and
474 ground ice meltwater, whereas the amount of ground ice is challenging to
475 observe, so it is reflected by high or low ground temperature. As
476 **supplemental Fig.1** shows, spatial correlations of soil moisture with air and
477 ground temperatures were analyzed during the sampling period.
478 Interestingly, there was a positive correlation in the early ablation period
479 because the active layer of permafrost is in the melting process. The higher



480 the ground temperature, the faster the ground ice melts, causing an increase
481 in soil water, especially at lower altitudes. The liquid water produced by
482 ground ice melting and the snow meltwater on the surface would move
483 down to the upper limit of permafrost, and the precipitation will also move
484 downward when the active layer completely melted, which would increase
485 the soil water in the active layer (Jiao et al., 2014). Liquid soil water
486 increased in the cold months under the increasing soil temperature and
487 enhancing ground ice melting, while changes in the warm months were the
488 results of competition between positive precipitation and adverse soil
489 temperature effects in permafrost regions (Lan et al., 2015). The active
490 permafrost layer melted slowly at higher altitude regions, and ~~the higher~~
491 ~~the ground temperature, the more evaporation occurred,~~ causing a decrease
492 in soil water. Wen et al. (2020) also indicated that temperature increases
493 reduced the shallow soil water in cold regions. In the heavy ablation period,
494 soil water exhibits a clear negative correlation with ground temperatures,
495 with the end of thawing the active permafrost layer and the weakening
496 effect of permafrost ground ice on soil water, and the higher the
497 temperature, the stronger the evaporation and lower the soil water. Most
498 regions display a clear positive correlation in September, with only a few
499 lower altitude areas showing a negative correlation. Two reasons can
500 account for it: (1) the top layer of soil at higher altitudes starts to freeze at
501 night and thaws during the day, thus increasing soil water; (2) soil water at



502 lower altitudes is affected by evaporation and decreases again. These facts
503 also indicate that changes in freeze-thaw processes are an important
504 influence on the evolution of soil water. During the thawing phase of the
505 active permafrost layer, the increase in precipitation or soil water led to an
506 increase in the thawing rate of frozen soil, accompanied by an increase in
507 water infiltration as the frozen soil continued to thaw, leading to an increase
508 in deep soil water and a decrease in surface soil water (Ma et al., 2021).
509 Under freeze-thaw cycles, the adequate soil water in the root layers of
510 different alpine meadows was ranked as follows: non-degraded meadow >
511 moderately-degraded meadow > seriously degraded meadow (Lv et al.,
512 2022). Xue et al. (2017) found that permafrost degradation significantly
513 reduced soil water by 4.5–6.1% at a depth of 0–100 cm and increased the
514 annual mean surface soil temperature by 0.8 °C in the source region of the
515 Yangtze River.

516 Precipitation infiltration is considered the primary source of soil water
517 in the active permafrost layer during the freeze-thaw action, which is
518 considered a major factor and imposes limitations (Cao et al.,2018). In June,
519 the spatial variation of soil water and precipitation in most regions,
520 especially at high altitudes, shows a negative correlation; while only a few
521 low-altitude regions show a positive correlation (supplemental Fig.2). Two
522 reasons can account for it: 1) on the one hand, this indicates that
523 precipitation in high altitude regions is mainly in the form of snowfall,



524 which is difficult to recharge soil water directly, and the active permafrost
525 layer melts slowly, and there is also the phenomenon of alternating
526 between melting and freezing. So the more the precipitation there is, the
527 less the soil water changes; 2) on the other hand, all the permafrost in low
528 altitude regions melts during the **season**, and soil water is mainly recharged
529 by precipitation, and the more precipitation, the higher the soil water. The
530 correlation between soil water and precipitation is low during the warm
531 season in permafrost areas and high in seasonal frozen areas because
532 permafrost may help maintain soil water stability. In contrast, the
533 permafrost degradation would reduce the regulating capacity of soil water,
534 affecting the Tibetan Plateau ecosystem and hydrological cycle (Wu et al.,
535 2021).

536 **Soil water changes in August exhibit a significant negative**
537 **correlation with precipitation. During this period, the active layer of**
538 **permafrost melted. However, the source of soil water was mainly**
539 **precipitation. More precipitation results in a higher quantity of soil water**
540 **(supplemental Fig.2).** Most areas show a positive correlation in September.
541 Only a few high-altitude areas display a negative correlation; due to the
542 temperature drop, precipitation in high-altitude areas is mainly snowfall,
543 which has less effect on the recharge of soil water, while the lower altitude
544 areas still show a positive correlation with rainfall, which directly
545 recharges soil water. Deng et al. (2020) also indicated that soil water



546 increased with precipitation in most regions of TRHR. Based on the
547 observation in TRHR, the soil water at 10 cm, 20 cm, and 30 cm increased
548 by 0.47% /mm, 0.46% /mm, and 0.41% /mm, **when the precipitation**
549 **increased by 1 mm**, while the soil water at 10 cm, 20 cm and 30 cm
550 decreased by 3.8%/d, 3.3% /d and 2.3% /d when the number of days
551 without precipitation increased by 1d, respectively (Li et al., 2022). The
552 average soil water during 2003–2020 was 20%, increasing at a rate of
553 0.5%/10a, and its changes were influenced by precipitation and
554 temperature in TRHR (Chen et al., 2021). In addition, the effect of snow
555 cover on the soil water thawing up was greater than that of freezing, and
556 the effect on shallow swamp soils was greater than that of shallow meadow
557 soils (Chang et al., 2012).

558 Evapotranspiration is the reverse process of soil water recharge. Soil
559 water, in general, shows a significant negative correlation with
560 evapotranspiration in June, August, and September in TRHR, indicating
561 that stronger evapotranspiration will result in less soil water (supplemental
562 Fig.3). Based on **observation under the simulated warming conditions** in
563 the Chengduo station in TRHR, the soil temperature increased by 2.50 °C
564 and 1.36 °C at the soil depth of 0–15 cm and 15–30 cm, respectively, while
565 the soil water decreased by 0.07% and 0.09% at the soil depth of 0–15 cm
566 and 15–30 cm, respectively (Yao et al., 2019). Cao and Jin (2021) also
567 concluded that soil water is negatively correlated with air temperature and



568 positively correlated with precipitation.

569 **4.2 Soil water sources and implications for vegetation restoration**

570 As the limiting factor determining ecosystem stability in cold regions, there
571 may be complex feedback relationships between vegetation and soil water.
572 This is of great significance for improving the understanding of the
573 hydrological process, soil and water conservation, and water resource
574 utilization. As supplemental Fig.4 shows, the correlation between soil
575 water and vegetation index in June is positive, and ~~the higher the altitude,~~
576 ~~the more significant the correlation.~~ On the one hand, the vegetation has
577 just resumed growth during this period, and the growth is slow with the
578 relatively weak evapotranspiration, and because the soil is dry after a
579 freezing period, the vegetation has a higher capacity to hold water; on the
580 other hand, the active permafrost layer is still in the melting process, and
581 the melting of ground ice increases the soil water accompanied by the
582 continuous vegetation growth. In the early stage of vegetation growth, the
583 upper soil layer has a high water-holding capacity, and the infiltration rate
584 of precipitation was slow through the surface layer to the deeper layer of
585 the soil, and there was a more uniform spatial distribution of soil water
586 with the obvious water-holding function (Wang et al., 2003). Liu et al.
587 (2021) also thought that the thawing of frozen soil increased the soil water
588 in the root zone, regulated root respiration, and brought the vegetation into
589 the growing season. Wei et al. (2022) also indicated that NDVI and surface



590 soil water were positively correlated in the Loess Plateau with a more
591 significant mutual feedback relationship.

592 Soil water displayed a negative correlation with vegetation index in most
593 areas in August, reflecting better vegetation growth and stronger
594 evapotranspiration, and **lower soil water**, as the active permafrost layer had
595 all melted by that time of year. Vegetation was in an active growth phase.
596 **It** showed a negative correlation with the vegetation index at most lower
597 elevation areas in September, reflecting better vegetation growth, stronger
598 evapotranspiration, and lower soil water; some higher elevations showed a
599 positive correlation, reflecting the effects of the freeze-thaw cycle.

600 Soil water is, therefore, the basis of vegetation. The vegetation indices
601 were closely associated with soil water, which played a key role in the
602 active layer thickness-vegetation relationship, especially at depths of 30–
603 40 cm in the northeastern Qinghai-Tibet Plateau (Jin et al., 2020). Thus,
604 precipitation and vegetation were the main factors that caused soil moisture
605 variation in summer and autumn, while the soil freeze-thaw cycle was the
606 main contributing factor in spring (Ma et al., 2021). Based on the
607 observation in the permafrost region, the mean surface soil water in the
608 alpine meadow was higher than that in the alpine steppe, while soil water
609 variability in the cold alpine steppe was larger than that in the alpine
610 meadow, which decreased with depths (Yang et al., 2011). The soil water
611 has reduced rapidly after the vegetation degeneration, especially in the soil



612 depth of 0~30 cm, and so climate warming and permafrost degradation
613 tend to decrease topsoil water in Tibetan Plateau (Wang et al., 2012).
614 The soil water in the alpine steppe and temperate steppe were mainly
615 affected by air temperature, and the influence factors for alpine meadow
616 and shrub were precipitation and NDVI (Zhang et al., 2022). The effect of
617 different vegetation types on the surface soil water varied widely, and the
618 higher the vegetation cover, the greater increase in soil water (Ma, 2016).
619 The surface soil water appeared to be significantly reduced with vegetation
620 degradation. The more serious the degradation, the more the water loss (up
621 to 38.6%) (Wang et al., 2010). In Qilian Mountains, the water loss has a
622 clear positive relationship with soil water and a negative relationship with
623 soil temperature for shrubland, grassland, and spruce forest (Hu et al.,
624 2019). Lu et al. (2020) also thought community cover was sensitive to
625 surface soil water and increased as a function of soil water from 1.1%–10.0%
626 and gradually tended to saturate. There was a significant positive
627 correlation between summer NPP and soil water in the watershed, but their
628 interactions manifested spatial heterogeneity (Yue et al., 2021). Thus, the
629 high soil water could support more plants with varied vegetation types
630 (Jiao et al., 2020).

631 As mentioned above, the variability of soil water has further increased
632 under warming, becoming the most critical factor affecting vegetation
633 growth, especially as soil water loss will further exacerbate vegetation



634 degradation and pose a great threat to the ecological security in the
635 “Chinese water tower.” Yue et al. (2022) also found that, under future
636 climate change, only timely supplementing soil water could promote net
637 primary productivity growth, improve vegetation productivity, and
638 effectively restore and protect the ecological environment. Therefore,
639 active measures should be taken in the following four aspects:

640 (1) Relying on ground-based meteorological observation stations,
641 hydrological stations, and field observation stations, combined with remote
642 sensing monitoring, to build a real-time observation network for soil water
643 variation to provide data support for the formulation of scientific and
644 reasonable water-soil conservation and vegetation restoration measures.

645 (2) In-depth research on the influence mechanism of soil water on
646 vegetation growth, construction of vegetation growth model based on soil
647 water and vegetation carrying capacity, the establishment of the soil water-
648 vegetation change-vegetation restoration early warning platform, real-time
649 early warning of vegetation degradation, and provision of the scientific
650 basis for the restoration of degraded vegetation.

651 (3) The melting of the active permafrost layer and the change of soil
652 water show seasonality, so the most suitable time for vegetation restoration
653 should also be determined scientifically. In lower altitude regions,
654 vegetation enters the growing season in June, and the restoration work
655 should be implemented from the end of May to promote seed germination.



656 In higher altitude regions, vegetation enters the growing season around the
657 end of June, and in order to improve the survival rate of vegetation, rapid
658 seed germination and breeding techniques should be developed. Mature
659 seedlings should be directly transplanted to make full use of the short
660 growing season to improve the effectiveness of vegetation restoration.

661 (4) There are significant differences in soil water and its environmental
662 effects in grasslands, and restoration measures should be taken according
663 to the different degradation degrees. The integrated pattern of winter rodent
664 eradication + growing season grazing ban + fertilizer application
665 technology is for the lightly degraded grasslands, which can significantly
666 improve the vegetation cover and height of grasses and maintain a stable
667 increase in soil water. The integrated pattern of winter rodent control +
668 growing season grazing ban + fertilizer application + no-till replanting is
669 for the moderately degraded grassland, which not only significantly
670 increases the cover, height, amount of good forage and above-ground
671 vegetation of grassland but also promotes the water-holding capacity. The
672 integrated pattern of winter rodent eradication + growing season grazing
673 ban + fertilizer application + re-vegetation technology is used in heavily
674 degraded grasslands to restore vegetation and to ensure that soil water is
675 stable enough to support vegetation growth.

676 (5) For areas of severe vegetation degradation, the focus is on the
677 adequate compensation of precipitation in time and space. Changing the



678 micro-topography to collect rainwater in the form of runoff or artificially
679 produced flow, including fish-scale pit and horizontal ditch technologies,
680 to achieve the objectives of water storage and moisture conservation,
681 increase the survival rate of vegetation and improve ecological water use.

682 **5. Conclusions**

683 Based on 2600 samples of soil and surface water collected in the Three-
684 River Headwater Region, the results indicated that $\delta^{18}\text{O}$ and δD are
685 relatively higher in June, then relatively negative in August, and then
686 higher in September, while it exhibits an opposite trend for d-excess. The
687 LEL and the relationship between soil water and surface waters in different
688 ablation periods can confirm this. The intersection between the LWML and
689 the LEL is considered to be the isotopic value of the initial water body that
690 recharges the soil water, and the corresponding $\delta^{18}\text{O}$ and $\delta^2\text{H}$ are -17.63‰
691 and -127.61‰ , -18.81‰ and -136.94‰ , -23.04‰ and -170.36‰
692 during the early, heavy and end ablation period in the TRRH, respectively.
693 These are close to the mean monthly precipitation values. Precipitation,
694 ground ice, and snow meltwater accounted for approximately 72%, 20%,
695 and 8% of soil water during the early ablation period, respectively, and it
696 is with no snow meltwater recharge below 4000 m due to snow melting
697 depletion. In the heavy ablation period, precipitation and ground ice
698 contributed to 90% and 10% of soil water, respectively. The precipitation



699 recharge decreases with increasing altitude, while ground ice gradually
700 increases. According to the EMMA, it accounted for about 94% and 6% of
701 soil water from precipitation and ground ice, respectively, during the end
702 ablation period, and the small amount of recharge of ground ice mainly
703 occurred in the area above 4000 m.

704 In the early ablation period, the higher temperature, the faster the
705 ground ice melts, causing an increase in soil water, especially at lower
706 altitudes. With higher temperature, more is the evaporation and lower is
707 the soil water in the heavy ablation period. Most regions display a clear
708 positive correlation in September owing to the freeze-thaw cycle. In June,
709 soil water and precipitation show a negative correlation because there is
710 mostly snowfall in high-altitude regions, whereas more the precipitation,
711 the higher the soil water in August and September. However, heavy
712 evapotranspiration will result in less soil water. It positively correlates with
713 vegetation in June owing to the melting permafrost active layer, but the
714 better vegetation growth, the lower soil water in August and September.
715 However, soil water loss will further exacerbate vegetation degradation
716 under warming and pose a significant threat to the ecological security of
717 the “Chinese water tower.” So, it is urgent to build the real-time soil water
718 observation network, construct the soil water-vegetation change-
719 vegetation restoration early warning platform, determine the most suitable
720 time for vegetation restoration, and apply the different soil water



721 conservation and vegetation recovery patterns.

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746 **Code/Data availability**

747 The raw/processed data required to reproduce these findings cannot be shared at
748 this time as the data also forms part of an ongoing study. We will not share our data
749 until all relevant results are completed.

750

751 **Author Contributions**

752 Zongxing Li led the write-up of the manuscript with significant contribution.
753 Zongxing Li developed the research and designed the experiments. Juan Gui and
754 Baijuan Zhang collected the water samples and analysed the data. All authors discussed
755 the results and contributed to the preparation of the manuscript.

756

757 **Competing interests**

758

759 This manuscript has not been published or presented elsewhere in part or in
760 entirety and is not under consideration by another journal. We have read and understood
761 your journal's policies, and we believe that neither the manuscript nor the study violates
762 any of these. There are no conflicts of interest to declare.

763

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1035 **Tables**

1036 Table.1 The average values of stable isotopes and relationship between

1037 $\delta^{18}\text{O}$ and d-excess for soil waters in TRHR

	Relationship between $\delta^{18}\text{O}$ and d-excess/ R^2	average values for: $\delta^{18}\text{O}$, $\delta^2\text{H}$ and d-excess in June	average values for: $\delta^{18}\text{O}$, $\delta^2\text{H}$ and d-excess in August	average values for: $\delta^{18}\text{O}$, $\delta^2\text{H}$ and d-excess in September
All soil water samples	$Y=-0.16x+3.87$, $R^2=0.0065$	-12.00; -89.78; 6.30	-13.26; -100.0; 8.58	-13.04; -98.11; 6.24
0-20cm	$Y=-0.43x+0.98$, $R^2=0.065$	-11.91; -90.07; 5.18	-13.24; -101.44; 8.87	-14.23; -108.14; 5.71
20-40cm	$Y=-0.4564x+0.7948$, $R^2=0.0392$	-12.07; -90.74; 5.84	-12.96; -99.01; 11.23	-12.42; -92.72; 6.61
40-60cm	$Y=-1.05x-7.33$, $R^2=0.1667$	-12.38; -90.38; 8.68	-13.63; -101.46; 5.67	-12.33; -92.06; 6.55
60-80cm	$Y=-0.32x+2.5781$, $R^2=0.0167$	-11.36; -83.77; 7.09	-13.32; -98.51; 4.17	-12.42; -92.88; 6.45
Northern slope	$Y=-1.1944x-7.3393$, $R^2=0.1584$	-12.33; -90.61; 7.99	-13.07; -98.34; 12.45	-12.05; -91.64; 4.75
Eastern slope	$Y=-0.7x-2.2479$, $R^2=0.0956$	-11.31; -85.49; 5.028	-13.77; -103.422; 6.16	-12.17; -89.9; 7.47
Western slope	$Y=-0.4337x+0.8866$, $R^2=0.0543$	-12.62; -93.63; 7.36	-12.92; -96.89; 11.99	-12.2; -91.5; 6.15
Grassland	$Y=-0.4921x-0.5722$, $R^2=0.0715$	-10.39; -77.66; 5.45	-12.13; -89.28; 27.06	-9.62; -71.87; 5.13
Meadow	$Y=-0.6067x+0.8133$, $R^2=0.0615$	-12.15; -90.36; 6.87	-13.45; -101.94; 5.25	-12.82; -96.56; 6.02



Forest	$Y = -1.4013x - 12.706$, $R^2 = 0.2283$	-13.6; 5.1	-103.66;	-13.66; - 103.16; 5.24	-15.82; - 118.98; 7.60
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1039 Table.2 The LEL for soil waters in study region

	EL/ R ² in June	EL/ R ² in August	EL/ R ² in September
2900-3500	$\delta^2H = 5.7\delta^{18}O - 21.18$ R ² =0.90	$\delta^2H = 6.8\delta^{18}O - 7.83$ R ² =0.95	$\delta^2H = 7.43\delta^{18}O - 2.59$ R ² =0.98
3500-4000	$\delta^2H = 7.58\delta^{18}O - 1.34$ R ² =0.83	$\delta^2H = 6.48\delta^{18}O - 16.54$ R ² =0.9	$\delta^2H = 7.67\delta^{18}O + 3.1$ R ² =0.97
4000-4500	$\delta^2H = 7.27\delta^{18}O - 3.46$ R ² =0.88	$\delta^2H = 6.5\delta^{18}O - 15.09$ R ² =0.93	$\delta^2H = 7.04\delta^{18}O - 6.8$ R ² =0.96
nibanshi4500-5100	$\delta^2H = 6.05\delta^{18}O - 12.4$ R ² =0.85	$\delta^2H = 6.69\delta^{18}O - 8.68$ R ² =0.93	$\delta^2H = 6.9\delta^{18}O - 6.6$ R ² =0.87
grassland	$\delta^2H = 6.4\delta^{18}O - 11.07$ R ² =0.83	$\delta^2H = 6.62\delta^{18}O - 9.07$ R ² =0.96	$\delta^2H = 6.44\delta^{18}O - 9.91$ R ² =0.92
meadow	$\delta^2H = 6.55\delta^{18}O - 10.67$ R ² =0.84	$\delta^2H = 6.4\delta^{18}O - 15.83$ R ² =0.90	$\delta^2H = 7.14\delta^{18}O - 5.05$ R ² =0.95
forest	$\delta^2H = 6.97\delta^{18}O - 8.9$ R ² =0.73	$\delta^2H = 7.61\delta^{18}O + 0.85$ R ² =0.97	$\delta^2H = 7.46\delta^{18}O - 0.97$ R ² =0.97
Northern slope	$\delta^2H = 7.33\delta^{18}O - 0.22$ R ² =0.84	$\delta^2H = 6.8\delta^{18}O - 9.46$ R ² =0.91	$\delta^2H = 6.86\delta^{18}O - 8.95$ R ² =0.90
Eastern slope	$\delta^2H = 6.92\delta^{18}O - 8.38$ R ² =0.88	$\delta^2H = 6.33\delta^{18}O - 16.9$ R ² =0.89	$\delta^2H = 6.78\delta^{18}O - 14.253$ R ² =0.93
Southern slope	$\delta^2H = 6.44\delta^{18}O - 13.22$ R ² =0.81	$\delta^2H = 6.84\delta^{18}O - 9.28$ R ² =0.96	$\delta^2H = 6.8\delta^{18}O - 7.0$ R ² =0.93
Western slope	$\delta^2H = 6.14\delta^{18}O - 16.14$ R ² =0.91	$\delta^2H = 6.46\delta^{18}O - 13.4$ R ² =0.92	$\delta^2H = 7.33\delta^{18}O - 2.07$ R ² =0.98

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1050 Figures

1051 Fig.1 The location of Three -River Headwater Region in ecological barriers
1052 of China(a); distribution map of permafrost and seasonal frozen soil (b),
1053 soil types (c) and vegetation types (d) in the study region

1054 Fig.2 the work photo for sampling glacier snow meltwater(a), soil in
1055 grassland(b), soil in meadow (c), supre-permafrost water (d), river water
1056 (e), tributary water (f), vegetation (g) and soil in forests (h)

1057 Fig.3 Distribution of sampling sites for soils and waters in June (a), August
1058 (b) and September (c)

1059 Fig.4 Plot of δD versus $\delta^{18}O$ and LEL for soil water at different soil layers
1060 on June (a), August (b) and September (c)

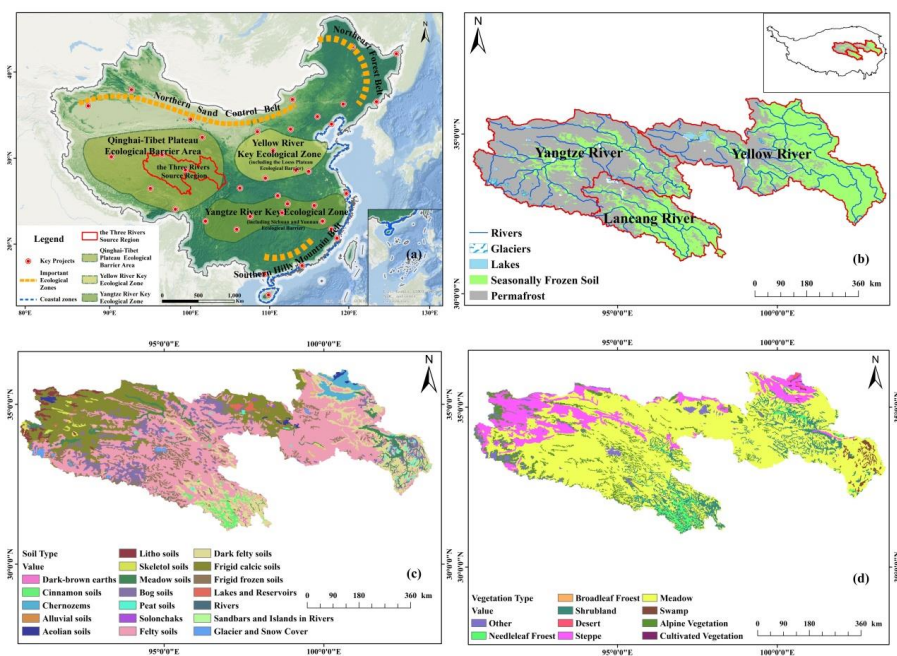
1061 Fig.5 Hydraulic connections between soil water and other waters for all
1062 samples (a), different altitudes (b) and vegetation (c) in June, all
1063 samples (d), different altitudes (e) and vegetation (f) in August, all
1064 samples (g), different altitudes (h) and vegetation (i) in September

1065 Fig.6 Three end element diagram in June (a) and Two end element diagram
1066 in August (b) and September (c) using the mean values of $\delta^{18}O$ and d-
1067 excess for soil water

1068 Fig.7 Concept map for contribution from precipitation, snow meltwater
1069 and ground ice to soil water in the whole study region, different
1070 altitudes and different vegetation in June (a), August (b) and
1071 September (c)



1072 Fig.8 Concept diagram for real-time monitoring network for soil water, soil
1073 water-vegetation change-vegetation restoration early warning platform and
1074 the different soil water conservation and vegetation restoration patterns in
1075 Three-River Headwater Region
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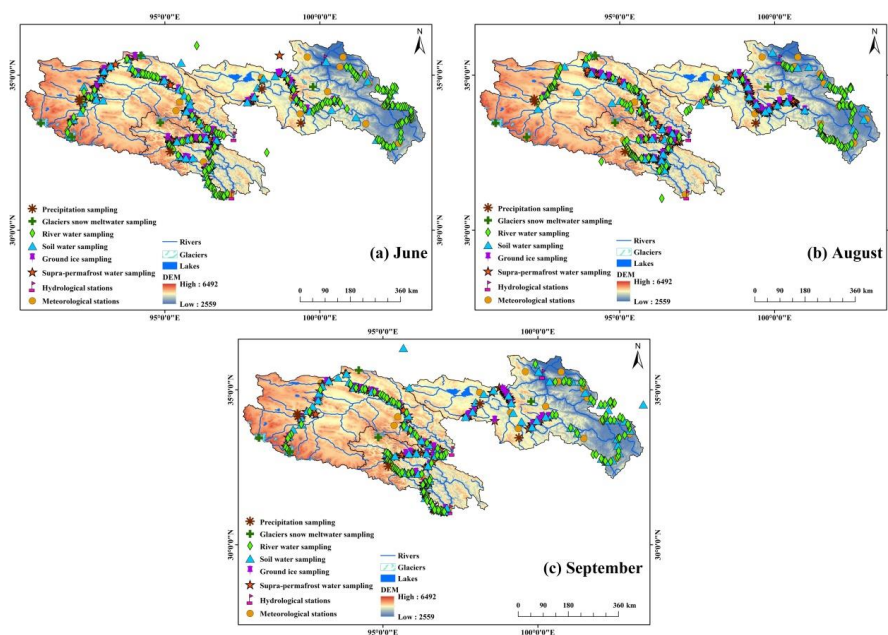
Fig.1



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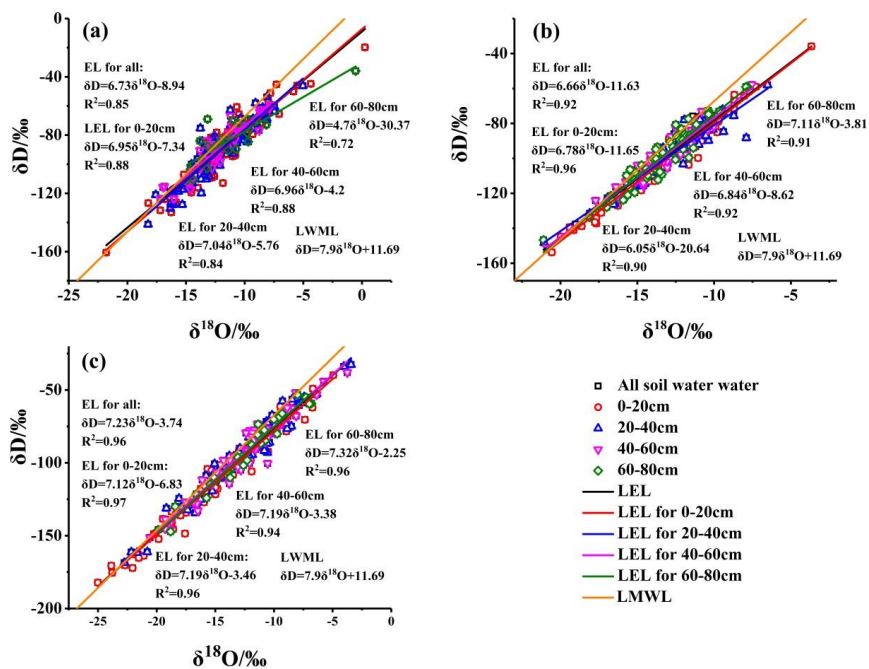
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Fig.2



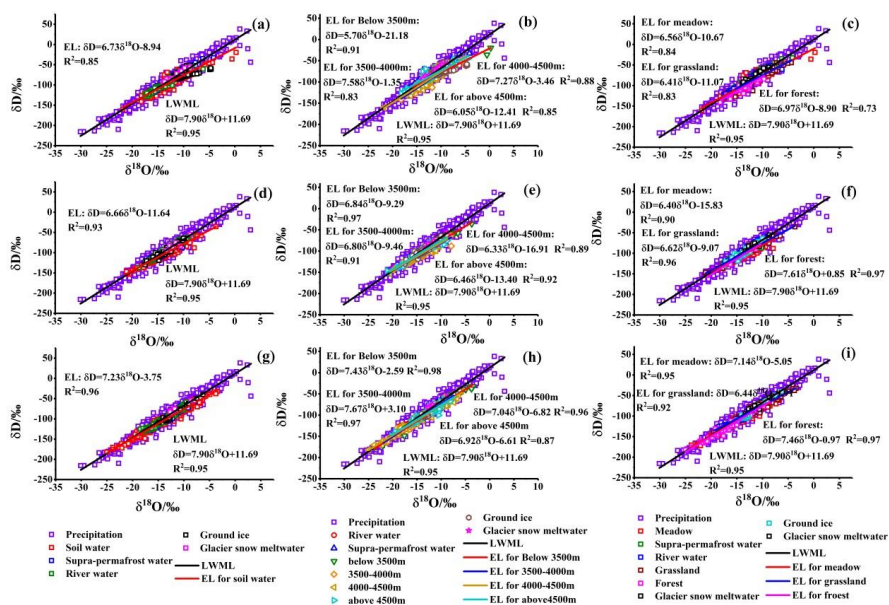
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Fig.3



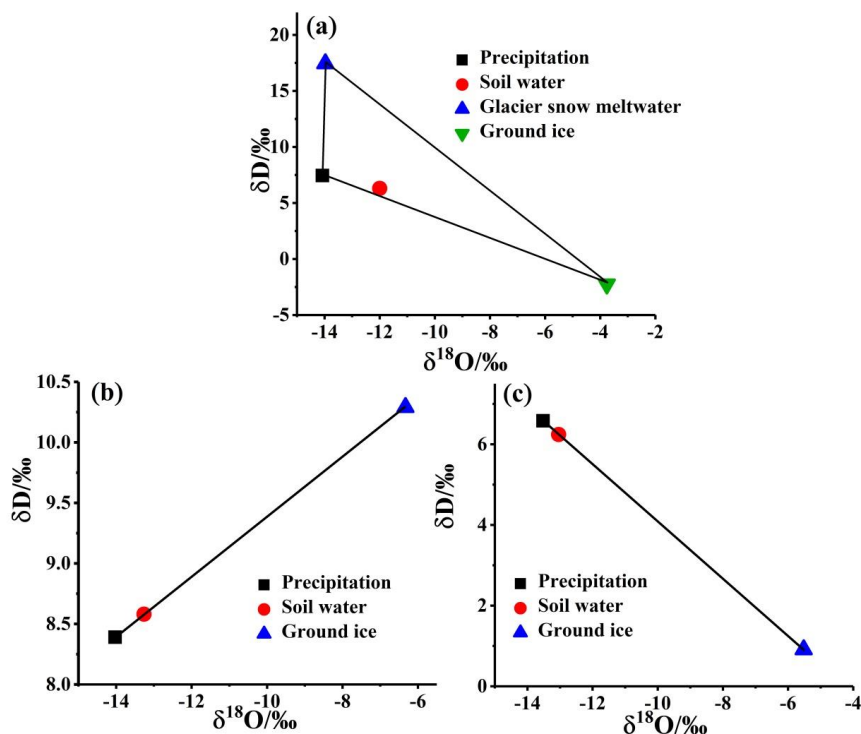
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Fig.4



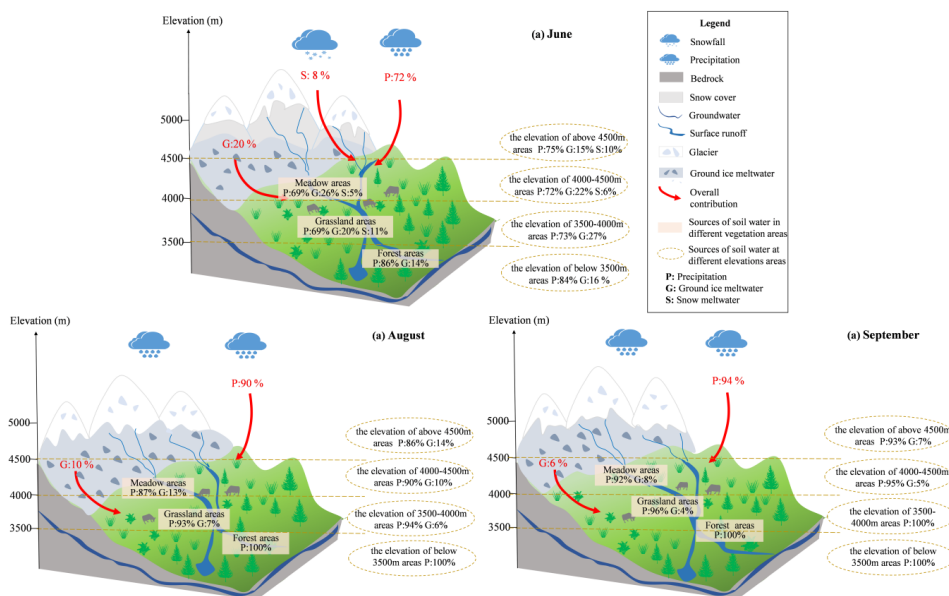
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Fig.5



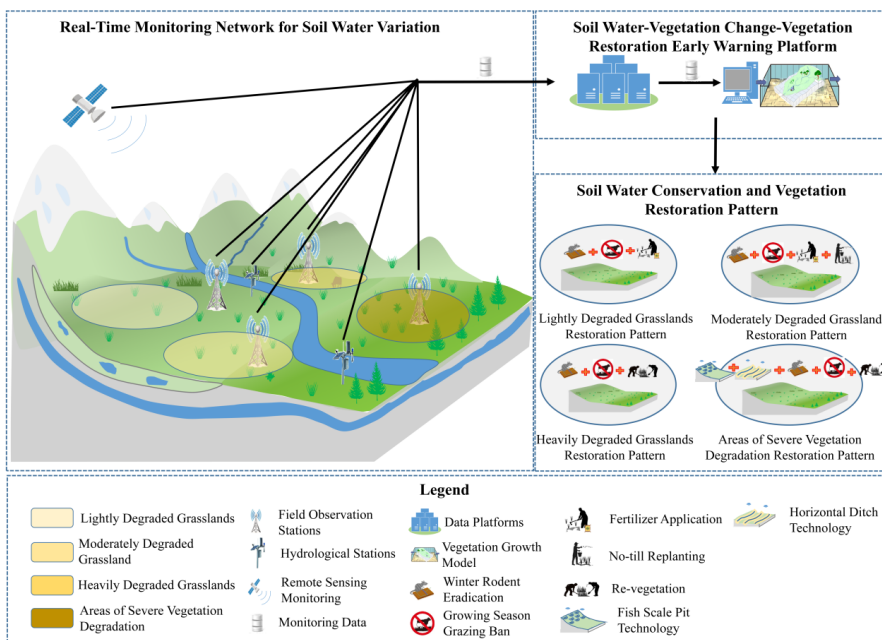
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Fig.6



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Fig.7



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Fig.8