1	Process-based three-layer synergistic optimal allocation
2	model for complex water resource systems considering
3	reclaimed water
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11	Abstract
12	The increasing water demand due to human activities has aggravated water scarcity,
13	and conflicts among stakeholders have increased the risk of unsustainable development.
14	Ignoring the effects of trade-offs leads to misguided policy recommendations. This
15	study highlights the concept of synergy among different aspects of water allocation
16	process. A process-based three-layer synergistic optimal allocation (PTSOA) model is
17	established to integrate the interests of stakeholders across sub_regions, decision levels
18	and time steps while simultaneously coupling reclaimed water to establish

environmentally friendly solutions. A synergy degree index is constructed by applying
network analysis for optimization. PTSOA is applied in Yiwu City, Southeast China,
and is shown to <u>be able to</u> improve the contradictions among different dimensionalities
in a complex system. Overall, 2.43×10⁷---3.95×10⁷ m³ of conventional water is saved,
and notable improvements in management are achieved. The application demonstrates
the efficiency and excellent performance of the PTSOA model.

Keywords Three-layer optimization, water allocation, process, synergy, reclaimed
water

27 **1. Introduction**

28 Water scarcity has become one of the major impediments to the-sustainable 29 development of cities (Yue et al., 2020). Emerging water scarcity concerns in cities are 30 associated with limited available water, severe water pollution and the relentlessly 31 growing demand for water as driven by industrial growth, population growth and higher 32 living standards; these factors have leaded to intense competition for freshwater among 33 stakeholders of interest (Dai et al., 2018; Wu et al., 2023). However, the heterogeneous 34 distribution of water resources at both spatial and temporal scales is common in many 35 cities and results in water shortage risks and conflicts, which often require the 36 optimization of water resource allocation (Friesen et al. 2017). Moreover, some 37 satisfactory alternatives for individual stakeholders may result in negative externalities on others. Nowadays, the water resources system become more and more complex, and 38

39 <u>is consisted withoften has multiple sources and users as well as water reused</u> 40 <u>infrastructure. This kind of water resources system is called complex water resources</u> 41 <u>system in the following.</u> Therefore, it is critical to develop a synergistic optimal 42 allocation model to alleviate conflicts and ensure the security, efficiency, equality, eco-43 environmental sustainability, and sustainable development of <u>complex water resources</u> 44 <u>systems water systems simultaneously.</u>;

45 As equitable access to water resources is closely related to social stability, several qualitative and indirect methods have been developed to assess water allocation 46 47 equality (D'Exelle et al. 2012). In cases with limited water resources, more water would 48 be allocated to users with better economic conditions to achieve more economic 49 benefits. Thus, stakeholders with poor economic status are ignored, resulting in 50 imbalanced development. Consequently, actions are often needed by local government 51 managers to avoid such situations. The Gini coefficient has been widely used to evaluate equality and enhance the optimization of water allocation in water use sectors 52 53 (Xu et al. 2019; Hu et al. 2016; D'Exelle et al. 2012). However, it is unable to reflect 54 the dynamic interactions among objectives, i.e., how objectives interact with each other 55 and impact the equity of a system in cases with diverse alternative decisions. While, iIn the perspective of coordinated allocaaction, multiple goals are simultaneously 56 57 considered to avoid -negative effects as much as possible. Therefore, in addition to equity, coordination should be considered in water allocation systems, and these two 58 59 concepts can be combined to promote systemic synergy. By identifying the dynamic

interactions among objectives, the internal mechanisms of a water system can be 60 clarified, and synergy can be achieved in cases with different potential decisions. It is 61 62 also helpful to identify the hurdles and opportunities associated with sustainable 63 development for cities and to establish specific action priorities for cities based on a comprehensive understanding of the interactions among objectives. To address this 64 65 knowledge gap, a correlational network approach is applied in this study, and a synergy 66 degree index is presented to consider both the equity and coordination of water systems. 67 Moreover, systemic analysis is used to assess the level of coordination of complex 68 objective interactions in city water systems.

69 Network analysis, which has been widely used in studies of complex systems (Ball 70 et al., 2000; Saavedra et al., 2011; Bond, 2017), is a holistic approach for exploring the 71 characteristics of interactions among objectives. It provides clear visualization and 72 conceptualization of the interactions among variables to fully characterize those 73 interactions (Swain and Ranganathan, 2021). An array of network metrics (for example, 74 degree centrality, betweenness centrality, eigenvector centrality, closeness centrality, 75 and community) can be applied to quantify the importance of objectives or targets in an 76 interaction network (Zhou and Moinuddin, 2017) and reveal the strongly connected 77 pairs of goals or targets in the network (Allen et al., 2019). A key network metric in 78 such analysis is connectivity, which reflects the degree of coordination among different 79 objectives in a system; in synergy networks, high connectivity indicates that many 80 objectives can be achieved simultaneously and that the negative effects of interactions

are mild (Wu et al., 2022). Thus, to facilitate the discovery of high-quality decision
alternatives, alleviate negative conflicts among multiple utilities and inform decision
making, a synergy degree evaluation index is established and applied to the network
analysis of this study.

85 Due to the negative externalities of individual decisions, conflicts occur not only 86 across different users or objectives but also across hierarchical decision levels. Water 87 use contradictions and inconsistent decision making by multiple managers inevitably 88 results in trade-offs, including positive and negative water resource feedback in cases 89 with limited water availability (Wang et al., 2022). In practice, district administrators 90 allocate water to each sector in each sub-region, and sub-region managers then make 91 use decisions based on the allocated amount of water resources (Safari et al., 2014). 92 Since each decision maker places emphasis on different targets, feedback and 93 coordination among different decision makers are of great importance. Therefore, 94 synergistic hierarchical water allocation that achieves coordination among different 95 decision makers is imperative to avoid conflicts, save water and maintain social stability. 96 To address these hierarchical problems, bi-level programming (BLP) has been widely used, wherein objectives at two hierarchical levels, namely, an upper level and 97 98 a lower level, are co-optimized (Zhang and Vesselinov, 2016; Jin et al., 2018). The 99 upper-level decision may be affected by the actions of the lower-level decision makers 100 (Arora and Gupta, 2009). Yue et al. (2020) formulated a bi-level programming (BLP) 101 framework to gain insight into the whole water allocation process with district

administrators and sub_regional farmers. Li et al. (2022) built a two-level model with the overall interests of system managers at the top and the individual interests of water supply departments at the bottom. The multi_level programming problem (MLPP) was derived from the bi_level programming problem (BLPP) and is more applicable to real world practices (Baky, 2014). However, limited studies have explored applying MLPP (more than two levels) for water resource allocation, especially in cases with unconventional water supplies.

109 To satisfy both long-term and short-term water needs and avoid unnecessary 110 administration costs and water resource use caused by a-lack of coordination among 111 different allocation steps, temporally synergistic allocation and optimization are needed 112 (Haguma and Leconte, 2018). In annual water resource planning, the monthly 113 variability of hydrologic regimes and non-stationarity of the daily water demand must 114 be considered. As an alternative example of synergistic allocation at different time steps, 115 Vicuna et al. (2010) used a monthly nonlinear programming model and an annual 116 sampling stochastic dynamic programming (SSDP) model to establish a monthly 117 operating policy. Haguma et al. (2015) proposed an optimization approach with two 118 separate time steps following the nested model approach. Haguma and Leconte (2018) 119 constructed deterministic and stochastic optimization models with two time steps (intra-120 annual and inter-annual) and two levels of inflow variability: seasonal and inter-annual. 121 The purpose of their short-time-step model was to derive aggregate performance 122 functions associated with potential long-time-step decisions in these studies. However,

short-term benefits should not be overlooked due to their appreciable impact on longterm effects. Accordingly, synergistic allocation that enhances both long-term and short-term allocations is of great importance for water resource management in cities. However, optimizing the structure of a model to achieve maximized benefits and balancing the trade-offs among time steps are tasks that have rarely been studied. The synergy among different time steps is addressed with a new innovative framework and a corresponding algorithm in our study.

130 Most of the abovementioned traditional models are based on a benefit-oriented 131 mechanism, which leads to a high degree of satisfaction in high-benefit regions and 132 large water shortages in other regions. The existence of high-benefit regions in a city 133 during the allocation process often exacerbates regional disparities and heterogeneous 134 development. Moreover, spatial factors influence allocation results, especially when 135 there is spatial hierarchical heterogeneity among water resource allocation elements (Li et al., 2022). It is thus appropriate to conceptualize water allocation problems in a 136 137 multistage framework that fully considers the interests of not only the regional authority 138 but also sub-regional managers (Yao et al., 2019). Hence, the synergy among sub-139 regions must be considered to optimally allocate water resources. Ideally, the benefits 140 of all sub-regions should be integrated equally in the model, and the weights of hyper-141 parameters should be adjusted to best support flexible policies.

142 The optimal allocation of conventional and unconventional water resources also143 significantly impacts water security and aquatic ecosystems. The reuse of reclaimed

144 water is beneficial for alleviating high water supply pressure on conventional water resources and reducing the emission of pollutants. To effectively integrate conventional 145 146 and unconventional water resources, Yang et al. (2008) and Han et al. (2008) introduced 147 unconventional water resources as critical factors in water management. Avni et al. 148 (2013) investigated the mixing of unconventional water resources with other 149 conventional water sources to meet the magnesium requirements for drinking water and 150 irrigation water. Yu et al. (2017) developed a cost-benefit analysis-based utilization 151 model for externally transferred water and desalinated water. The allocation of both 152 conventional and unconventional water has been widely studied, but there remains a 153 lack of methods to guide the synergistic allocation of conventional and unconventional water resources and embed reclaimed water supply systems in allocation schemes. The 154 155 overexploitation of conventional water resources is not conducive to the sustainable development, while the extensive use of unconventional water could ultimately result 156 157 in high economic burden. To synergistically integrate conventional and unconventional 158 water resources and guide the coordinated allocation of these two types of water 159 resources, corresponding mechanisms must be implemented. As a result, our study aims to couple the allocation of conventional water resources and unconventional water 160 161 resources to establish synergistic solutions.

In summary, as insufficient water supplies and increasing water demands intensify competition for water resources and lead to conflicts among different stakeholders in different dimensions, water allocation must be optimized in cities and regions to achieve synergistic decision-making at various levels and time steps considering the value of reclaimed water. Therefore, a new process-based three-layer synergistic optimal allocation (PTSOA) model is developed here to generate numerous candidates or Pareto solutions and identify several desirable decision alternatives. The synergy of time and space optimization is achieved in the new model to avoid waste and promote balanced spatial development. Furthermore, in the PTSOA model, reclaimed water is used to replenish conventional water resources in water-scarce areas.

The remainder of this paper is organized as follows. The mathematical model is formulated in Section 2. Section 3 gives a numerical example for Yiwu city to demonstrate the effectiveness and efficiency of the proposed methods. The results are shown in Section 4; different water allocation strategies under varying inflow conditions are explored, and <u>some</u> policy implications are discussed. Section 5 presents conclusions.

178 2. Modelling

With water resources becoming increasingly scarce, multi-dimensional synergistic optimal allocation in a hierarchical system is crucial for ensuring sustainable development in water-scarce cities. There are three dimensions of synergy in the established allocation model, as shown in Fig. 1: process synergy, decision-level synergy and time-scale synergy. The synergy of the process refers to synergistic water allocation among the three stages throughout the whole allocation process to reduce

waste in bridging processes, which has rarely been considered. In the three stages, first, 185 the original water is released from reservoirs or diverted from external water transfer 186 projects to water works; then, the water stored in water works infrastructure is supplied 187 188 to different departments that need different types of water, including both conventional 189 and reclaimed water. Finally, the-water is supplied to different users. Decision-level 190 synergy refers to synergistic water allocation considering the interests of decision makers at different levels, namely, the city, water department and regional levels, to 191 192 coordinate solutions and avoid conflicts among decision makers. The city level 193 represents the overall interests of a city from the perspective of government, the water 194 supply department level represents the interests of water supply corporations, and the regional level focuses on the comprehensive benefits of each region in the city and 195 196 mitigate development imbalance among regions. Optimal decision making at the department level is constrained by the allocation results at the city level, and so on, and 197 198 the final solution should satisfy the needs of decision makers across all levels. The time 199 scale synergy involves the coordination of the daily configuration goal with the monthly 200 goal, the monthly goal with the yearly goal, and so on. Synergistic temporal allocation 201 can largely alleviate time conflicts during configuration operations, ensuring that all 202 configuration periods serve the same final configuration objectives to save water 203 resources and improve efficiency. However, time scale synergy mainly depends on 204 artificial operations rather than automated intelligent operations in practice. In-depth 205 exploration has yet to be demonstrated. Consequently, the PTSOA model is constructed

here to fully consider these three dimensionalities of synergy. The dimensionalities are
coupled in this model to achieve the efficient maximization of comprehensive benefits
at all levels under the premise of saving water resources. In Fig. 1, the grey boxes
indicate the three different allocation dimensions, the green boxes indicates the three
different decision levels coupled with spatial scales, the bright yellow boxes indicates
every key nodes in the whole allocation process and the buff boxes indicates nested
time scale.



Fig. 1. Conceptual map of the PTSOA model

217 In water-scarce cities, using reclaimed water as an alternative water source is 218 proved to be useful a proven approach to efficiently improve the environment by 219 reducing sewage discharge. The quality of inland tributaries has deteriorated in many 220 water-scarce cities due to limited consideration of the water environment and the large-221 scale emission of pollutants. Transferring reclaimed water and main river water to urban 222 inland tributaries for ecological water replenishment is a promising approach for improving the quality of urban water environments and areas with water shortages. 223 224 However, there has been a lack of studies on the integration of reclaimed water reuse 225 systems and inland water distribution systems in allocation modelling. Therefore, in 226 addition to saving water resources and improving efficiency through multi-dimensional 227 synergistic allocation, the model encompasses reclaimed water reuse systems and 228 ecological water distribution systems for inland tributaries. 229 Finally, the PTSOA model is constructed to solve the multi-dimensional synergistic 230 allocation problem involving complex water resource networks that couple reclaimed 231 water reuse systems and inland ecological water distribution systems with multiple 232 sources, processes and regions to guarantee the sustainable development of water-

scarce cities. To select the most synergistic solution of the PTSOA model, a new evaluation index named the total synergy index (*TSI*) is proposed to assess the synergy degree of different decision alternatives. System entropy (H(S)) can describe the evolution direction of a water resource system and was used to promote the coordination of water supply departments in a water resource allocation system(Li et

238 al., 2022). So, it-was is used for comparison to evaluate the validity of this proposed 239 index. Furthermore, the network analysis method is applied for the first time to analyse 240 dynamic interactions in water optimal allocation. This method visually depicts the 241 dynamic interactions and conflicts among different subareas in a city, which is helpful for system managers to realize how the water allocation scheme in one region 242 243 influences that of other areas; consequently, more reasonable and flexible measures are 244 established based on dynamic regional development targets. The detailed framework 245 developed in this study is shown in Fig. 2. In Fig. 2 this figure, there are three layers in 246 the framework and each layer has two parts: multi-objective optimal water resources 247 allocation and collaborative water resources allocation for objectives. In the multiobjective optimal water resources allocation, sub-layers contain key nodes in the 248 249 allocation process and relevant objectives and constraints. In the collaborative water resources allocation for objectives, sub-layers contain optimization algorithm and 250 251 decision selection method.

252



253

254

Fig. 2. Framework of the PTSOA model

255 2.1 First layer of the PTSOA decision-making process



260 satisfy the overall development goals of the city, the first-layer processes involve city-261 level decision-making. The city manager focuses on the overall goal of the complex 262 water resources systemwater resource system in the city, which is the first and most important phase of the decision-making process. The established allocation scheme 263 264 highly influences decision makers at other levels, and optimal allocation schemes at 265 other levels must align with this overall goal. Additionally, since water resource planning in most Chinese cities is based on an annual planning period and monthly 266 planning unit, the time step of the first layer is set as months. Finally, the monthly 267 268 decision alternatives for the volume of water allocated from reservoirs to water works 269 is obtained at the city decision level.

270 2.1.1 Objective functions of the first layer

271 Social objective function: <u>Minimization minimization</u> of total water supply 272 shortages

The social objective function is established by the city manager to minimize the total water supply shortages in a water system. The objective is established to sufficiently meet the water demands of users in a water <u>rAesourceesresources</u> system. The water deficit is considered, and this objective can reflect the ability of the water supply to meet the water demand, as shown in Eqs. (1-3):

278
$$\min f_{11}(x) = D - S$$
 (1)

279
$$D = \sum_{r=1}^{R} \sum_{t=1}^{T} D_{r}^{t}$$
(2)

280
$$S = \sum_{t=1}^{T} \sum_{i=1}^{I} \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij} + \sum_{t=1}^{T} \sum_{e=1}^{E} \sum_{j=1}^{J} x_{ej}^{t} \beta_{ej}$$
(3)

where $D(10^4 \text{ m}^3)$ is the total water demand of the system, $D_r^t(10^4 \text{ m}^3)$ is the water 281 282 demand of the rth sub-region at th time step, r = 1, 2, ..., R, R is the total number of subregions in the area, t=1,2,...,T, T is the total number of months in the period, S (10⁴ m³) 283 is the total water supply of the water system for all waterworks, x_{ij}^{t} (10⁴ m³) is the 284 amount of water supplied from *i*th reservoir to the *j*th waterworks in the *t*th month of 285 the configuration period, i=1,2,...I, I refers to the total number of reservoirs, j=1,2,...J, 286 J is the number of total water works, x_{i}^{t} (10⁴ m³) is the amount of water supplied from 287 288 the eth external transfer water source to the *j*th water works in the *t*th month of the configuration period, e=1,2,...,E, E is the total number of external transfer water 289 290 sources in the city, α_{ii} is the water supply relationship coefficient between the *i*th 291 reservoir and the *j*th water works, where 0 indicates no supply and 1 indicates a water supply, and β_{ej} is the water supply relationship coefficient between the *e*th external 292 transfer water source and the *j*th water works, where 0 indicates no supply and 1 293 294 indicates a water supply.

Economic objective function: <u>Maximization maximization</u> of the total water supply benefit

297 A city manager operates a water allocation system to maximize the overall economic

298 benefit by establishing an economic objective function, as shown in Eqs. (4-7):

299
$$\max f_{12}(x) = B - C_{rs} - C_{es}$$
(4)

300
$$C_{rs} = k \times \sum_{t=1}^{T} \sum_{i=1}^{I} \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij} + \sum_{t=1}^{T} \sum_{i=1}^{J} \sum_{j=1}^{J} \left(x_{ij}^{t} \alpha_{ij} \times c_{i} \right)$$
(5)

301
$$C_{es} = m \times \sum_{t=1}^{T} \sum_{j=1}^{J} \sum_{e=1}^{E} x_{ej}^{t} \beta_{ej} + \sum_{t=1}^{T} \sum_{e=1}^{E} \left(n_{e} \times \sum_{j=1}^{J} x_{ej}^{t} \beta_{ej} \right)$$
(6)

302
$$B = \sum_{j=1}^{J} b_j \times \left(\sum_{t=1}^{T} \sum_{i=1}^{I} x_{ij}^t \alpha_{ij} + \sum_{t=1}^{T} \sum_{e=1}^{E} x_{ej}^t \beta_{ej} \right)$$
(7)

303 The overall economic benefit is the difference between the total benefit and total 304 cost at the city level. In the equations, B (Chinese yuan Yuan, shortened to yuan in the 305 following text) is the total direct water supply benefit (mainly considering the income 306 from water charges for the city). The total water supply cost consists of the reservoir water supply cost C_{rs} and the external water supply cost C_{es} ; k (yuanChinese 307 <u>Yuan/m³</u>) denotes the water resources fees paid to the government; c_i (yuanChinese 308 309 Yuan/m³) denotes the water fees paid to the *i*th reservoir authority; m (yuanChinese 310 Yuan/m³) is the charge to an external administrative district per unit of externally transferred water; n_e (yuanChinese Yuan/m³) is the charge associated with the *e*th 311 external water source per unit of transferred water; and b_j (yuanChinese Yuan/m³) is 312 313 the unit price of water supply revenue for the *j*th user.

Sustainable development objective function: <u>Maximization maximization</u> of the total amount of reserved water in reservoirs

In water-scarce cities, the problem of water scarcity is a serious challenge that preventssustainable allocation of water resources. A prominent feature of most water-scarce

cities is that water inflows are limited, and the fluctuations in water availability are large. Therefore, to reduce the risk that the inflows in the next configuration period are too short to meet the basic demand of the city, such that a sustainable development objective function is developed. The sustainable development objective function seeks to maximize the amount of water remaining in the reservoir at the end of a configuration period to hedge against drought risk and guarantee water use in the next period, as shown in Eqs. (8-10):

325
$$\max f_{13}(x) = \sum_{i=1}^{N} (V_i^{\max} - V_i) \times p(V_i^{\max} - V_i)$$
(8)

326
$$p(V_i^{\max} - V_i) = \begin{cases} 2 \times V_i / V_i^{\max} & 0 < V_i < V_i^{\max} / 2\\ -2 \times V_i / V_i^{\max} + 2 & V_i \ge V_i^{\max} / 2 \end{cases}$$
(9)

327
$$V_{i} = \sum_{t=1}^{T} \left(R_{i,initial} + I_{i}^{t} + P_{i}^{t} - A_{i}^{t} - E_{i}^{t} - EP_{i}^{t} \right) - \sum_{t=1}^{T} \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij}$$
(10)

where $V_i^{\text{max}}(10^4 \text{ m}^3)$ is the maximum allowable storage capacity of the *i*th water source, 328 which is expressed based on the limited storage capacity of a reservoir in the flood 329 season, and V_i (10⁴ m³) is the water storage capacity of the *i*th water source at the end 330 of the scheduling period. As much water as possible but less than V_i^{max} is reserved. 331 332 However, a reserved water volume in the reservoir that is too high at the end of the scheduling period may lead to considerable pressure on reservoirs to urgently release 333 334 water if a flood event is forecasted. The reserved water volume should be neither too 335 large nor too small. Thus, the benefits of reservoir retainment reserve stock must be 336 thoroughly evaluated. Based on the characteristic of the benefit of residual water, we

propose a boundary benefit function $p(V_i^{\max} - V_i)$ for different reserved water 337 338 volumes in a reservoir. The benefit function is a piecewise function, and when V_i is less than $V_i^{\text{max}}/2$, p increases as V_i increases. When V_i is equal to or greater than 339 $V_i^{\text{max}}/2$, p decreases as V_i increases. When $V_i = V_i^{\text{max}}$, p decreases to 0. $R_{i,initial}$ (10⁴) 340 m³) is the initial storage of the *i*th water source, $I_i^t (10^4 \text{ m}^3)$ is the inflow of the *i*th water 341 source at the *t*th time step, P_i^t (10⁴ m³) is the precipitation associated with the *i*th water 342 source at the *t*th time step, A_i^t (10⁴ m³) and E_i^t (10⁴ m³) are the agricultural and 343 ecological water supplies associated with the *i*th water source at the *t*th time step, 344 respectively, and EP_i^t (10⁴ m³) is the evaporation from the *i*th water source at the *t*th 345 346 time step.

347 2.1.2 Constraints of the first layer

348 The layer includes six main constraints: the reservoir water supply constraint, water 349 demand constraint, reservoir storage constraint, water balance constraint, external water 350 transfer constraint, and nonnegative constraint.

351 Reservoir water supply constraint

The maximum water available to supply from an individual reservoir is determined by the difference between the total input and total reservoir output. The inputs include inflow and precipitation, and the outputs mainly involve agricultural and environmental water supplies, evaporation, water supplied for waterworks and reservoir leakage loss. All these factors directly affect the decision-making process and are incorporated into 357 the model building process as shown in Eqs. (11-15):

$$V_i^t \le V_{i,\max}^t \tag{11}$$

359
$$V_i^t = \sum_{j=1}^J x_{ij}^t \alpha_{ij}$$
 (12)

360
$$V_{i,\max}^{t} = \sum_{t=1}^{t-1} \left(R_{i,initial} + I_{i}^{t} + P_{i}^{t} - A_{i}^{t} - E_{i}^{t} - EP_{i}^{t} - \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij} - L_{i}^{t} \right) - V_{i,d}$$
(13)

$$EP_i^t = ep_i^t \times s_i^t / 1000 \tag{14}$$

362
$$V_i^t = \xi_i^t \times \left(R_i^{t-1} + R_i^t\right)$$
(15)

where V_i^t (10⁴ m³) denotes the total water supply from the *i*th reservoir at the *t*th time 363 step; $V_{i,\text{max}}^{t}$ (10⁴ m³) is the maximum water available to be supplied from the *i*th 364 reservoir at the *t*th time step; ep_i^t (mm) is the water surface evaporation from the *i*th 365 366 reservoir in the *t*th month; s_i^t (m²) is the monthly average surface area of the *i*th reservoir in the *t*th month; $V_{i,d}$ (10⁴ m³) is the dead storage of the *i*th reservoir; L_i^t (10⁴ 367 m³) is the reservoir leakage loss from the *i*th reservoir at the *t*th time step; R_i^{t-1} (10⁴ m³) 368 is the storage of the *i*th reservoir at the *t*-1th time step; R_i^t (10⁴ m³) is the storage of the 369 *i*th reservoir at the *t*th time step; and ξ_i^t is the *t*th monthly leakage coefficient for the 370 *ith* reservoir. 371

372 Water demand constraint

The high-quality water demand of each subarea in a city should be satisfied in the water allocation process. High-quality water in this model refers to water that satisfies the relevant primary (surface water can be used for drinking after simple purification treatment, such as filtration and disinfection) and secondary water quality requirements
(water is slightly polluted and can be used for drinking after routine purification
treatment, such as flocculation, precipitation, filtration, disinfection, and other
processes) according to the Chinese Standard (GB5749), as shown in Eq. (16):

380

381
$$0.8 \times D_{r} \leq \sum_{t=1}^{T} \sum_{i=1}^{J} \sum_{j=1}^{Jr} x_{ij}^{t} \alpha_{ij} + \sum_{t=1}^{T} \sum_{e=1}^{F} \sum_{j=1}^{Jr} x_{ej}^{t} \beta_{ej} \leq 1.2 \times D_{r}, r = 1, 2, ..., R$$
(16)

where D_r (10⁴ m³) is the high-quality water demand in the *r*th sub_region and there are a total of *R* sub_regions in the city. *Jr* is the number of waterworks in the *r*th sub_ region. To ensure that the water supply guarantee in each area is greater than 80%, the total water supplied to every subarea is greater than 80% of its demand.

386 **Reservoir storage constraint**

$$R_i^T \le V_{i,f} \tag{17}$$

388
$$R_{i}^{T} = \sum_{t=1}^{T} \left(R_{i,initial} + I_{i}^{t} + P_{i}^{t} - A_{i}^{t} - E_{i}^{t} - EP_{i}^{t1} - \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij} - V_{i}^{t} \right)$$
(18)

389 where R_i^T (10⁴ m³) is the storage of the *i*th reservoir at the end of the configuration 390 period and $V_{i,f}$ (10⁴ m³) is the flood-limit storage capacity.

391 Water balance constraint

392
$$R_i^{t+1} = R_i^t + I_i^t + P_i^t - A_i^t - EP_i^t - E_i^t - V_i^{t-1} - \sum_{j=1}^J x_{ij}^t$$
(19)

393 External transfer water constraint

394
$$\sum_{t=1}^{T} \sum_{j=1}^{J} x_{ej}^{t} \beta_{ej} \leq E_{e,\max}$$
(20)

395 where $E_{e,\max}$ refers to the maximum water supply capacity of *an* external water source 396 over the whole configuration period.

397 Nonnegative constraint

$$x_{ii} \ge 0 \tag{21}$$

399 2.2 Second layer of the PTSOA decision-making model

400 Similarly, the second layer of the PTSOA model fuses all three dimensions of 401 synergistic water resource allocation mentioned previously. The second stage of the process (the water stored in water works is supplied to different departments needing 402 403 water volumes of different quality) is optimized in the second layer. After city-level 404 decision-making, a conflict of interest inevitably occurs between traditional water 405 supply departments and unconventional water supply departments. Because 406 conventional and unconventional water supply departments compete for limited water demand market shares, the stability of the water allocation system may be jeopardized 407 408 if excessive competition is not controlled. Thus, the second layer is implemented at the 409 department level. Decision-making at the department level seeks to guide the two water supply departments to partake in benign competition and avoid conflicts to realize 410 411 synergy. In this case, the decision plan of the first layer in the hierarchy is followed. Temporally, short-term allocation changes are needed as mentioned above; hence, the 412 413 time scale of the second layer is daily. Thus, the daily decision alternatives for the 414 volume of water allocated from water works to different water departments are obtained

415 to make relevant decisions.

416 2.2.1 Objective functions of the second layer

417 Conventional water supply department objective function: Minimization 418 minimization of the total amount of water retained in water works

The managers of conventional water supply departments strive to operate conventional water systems efficiently and achieve the most equitable water share possible. The amount of conventional water (of high quality) retained in a water works system is a crucial factor affecting the efficiency and benefits of conventional water supply departments. Therefore, the benefit of conventional water departments is established by minimizing the total amount of water retained in water works at the end of a configuration period, as shown in Eq. (22):

426
$$\min f_{21}(x) = W_L = \sum_{t=1}^T \sum_{i=1}^J \sum_{j=1}^J x_{ij}^t \alpha_{ij} + \sum_{t=1}^T \sum_{e=1}^L \sum_{j=1}^J x_{ej}^t \beta_{ej} - \sum_{t=1}^T \sum_{m=1}^M \sum_{j=1}^J \sum_{z=1}^Z q_{jz}^{t,m} \chi_{jz}$$
(22)

where W_L (10⁴ m1) is the total amount of water retained in a water works system at the end of a configuration period; $q_{jz}^{t,m}$ (10⁴ m³) is the water supply from the *j*th water works system to the *z*th water user on the *m*th day in the *t*th month in the period of configuration; m=1,2,...,M; and M is the total number of days in the *t*th month (28, 29, 30 or 31). Additionally, z=1,2,...Z, and Z is the total number of water users. χ_{jz} is the water supply relationship coefficient between the *j*th water work and the *z*th water user, where 0 indicates no supply and 1 indicates **a**-water supply. 434 Unconventional water supply objective function: <u>Maximization-maximization</u> of

435 the amount of unconventional water supplied

436 The reclaimed water reuse system and ecological water distribution system for inland tributaries are incorporated into the PTSOA model which are associated with 437 unconventional water supply departments. The managers of unconventional water 438 supply departments seek to supply as much unconventional water as possible to 439 promote their interests. Thus, the objective of unconventional water departments is 440 established to maximize the amount of unconventional water supplied. Unconventional 441 442 water mainly includes reclaimed water and river water, which is of low quality (i.e., not meeting the quality standard mentioned in Section 2.1.2) and is mainly used for 443 444 industrial production, ecological water replenishment for inland rivers and municipal 445 road sprinkling.

446 Unconventional water departments operate reclaimed water reuse systems and 447 ecological water distribution systems to supply unconventional water, and the 448 associated equations are as follows in Eqs. (23-26):

449
$$\max f_{22}(x) = W_r + EW_r$$
 (23)

450
$$W_{r} = \sum_{t=1}^{T} \sum_{n=1}^{N} \sum_{j=1}^{J} r_{nj}^{t} p(b_{c}, b_{u}) \theta_{nj}$$
(24)

451
$$p(b_c, b_u) = \frac{1}{3} \times \frac{b_c}{b_u} - \frac{2}{3}$$
(25)

452
$$EW_r = \sum_{t=1}^T \sum_{n=1}^N \sum_{z=1}^Z r_{nz} \theta_{nz}$$
(26)

453 where W_r (10⁴ m³) is the total amount of reclaimed water supplied for all water users; 454 EW_r (10⁴ m³) is the total amount of river water supplied to maintain ecological flows in inland tributaries; r_{ij}^{t} (10⁴ m³) is the amount of water supplied to the *j*th user from 455 456 the *n*th reclaimed water source at the *t*th time step; n = 1, ..., N; N is the total number of reclaimed water sources; $p(b_c, b_u)$ is a function expressing the willingness of residents 457 to use reclaimed water, b_c (yuanChinese Yuan/10⁴ m³) is the price per unit of 458 conventional water; b_{μ} (yuan<u>Chinese Yuan</u>/10⁴ m³) is the price per unit of 459 unconventional water; and θ_{nj} is the water supply relationship between the *n*th 460 reclaimed water source and the *j*th user. In this case, $\theta_{nj} = 1$ indicates a water supply 461 relationship, and $\theta_{nj} = 0$ indicates no water supply relationship. r_{nz} (10⁴ m³) is the 462 463 amount of water supplied from the *n*th reclaimed water source to the *z*th inland tributary; z = 1, 2, ..., Z; Z is the total number of inland tributaries requiring ecological flow 464 compensation; and θ_{nz} is the water supply relationship between the *n*th reclaimed 465 water source and the zth inland tributary. 466

467 2.2.2 Constraints of the second layer

468 Conventional water supply constraint

According to conservation of mass, the total daily amount of conventional water allocated in the second layer should be less than the total monthly amount of conventional water allocated in the first layer, as described in Eq. (27):

472
$$\sum_{t=1}^{T} \sum_{i=1}^{J} \sum_{j=1}^{J} x_{ij}^{t} \alpha_{ij} + \sum_{t=1}^{T} \sum_{e=1}^{E} \sum_{j=1}^{J} x_{ej}^{t} \beta_{ej} \ge \sum_{t=1}^{T} \sum_{m=1}^{M} \sum_{j=1}^{J} \sum_{z=1}^{Z} q_{jz}^{t,m} \chi_{jz}, t = 1, ..., T$$
(27)

473 Unconventional water constraints

The two types of unconventional water have separate constraints. For reclaimed water supplied to water users, the amount should satisfy the relevant water recycling standard. The ecological water used to replenish inland tributaries is mainly pumped from reclaimed water works and main rivers. Therefore, this replenished volume is limited by the pumping capacity. The constraints for unconventional water are shown in Eqs. (28)-(29):

480
$$\sum_{t=1}^{T} \sum_{n=1}^{N} \sum_{j=1}^{J} r_{nj}^{t} \theta_{nj} + \sum_{t=1}^{T} \sum_{n=1}^{N} \sum_{z=1}^{Z} r_{nz} \theta_{nz} = \sum_{t=1}^{T} \sum_{i=1}^{J} \sum_{j=1}^{J} x_{ij}^{t} \delta_{ij} \eta_{ij} + PU$$
(28)

481
$$PU = \sum_{t=1}^{T} \sum_{m=1}^{M} \sum_{p=1}^{P} Q_{t,m}^{p,s} / 10^4$$
(29)

where δ_{ij} is the sewage discharge coefficient, which is the proportion of the water supplied from sewage discharge; η_{ij} is the sewage water reuse rate, which is the proportion of reused water in the total volume of sewage water; *PU* (10⁴ m³) is the amount of water pumped from the main river; and $Q_{t,m}^{p,s}$ (t/d) is the flow through the sth pumping station on the *m*th day at time step *t*.

487 **Pumping constraints**

$$Q_{t,s}^{p} \le Q_{\max,s}^{p} \tag{30}$$

489
$$Q_t^P = \sum_{s=1}^{N_P} r_{t,s}^P$$
(31)

490 where $Q_{\max,s}^{p}$ (t/d) denotes the upper flow boundary of the *s*th pumping station; r_{t}^{s} (t/d) 491 is the power of the *p*th pump installed at the *s*th pump station; and *Np* is the number of 492 pumps stalled at the *s*th pump station.

493 Water quality constraint

To control the impacts of various point and nonpoint sources on receiving water bodies in cities, water authorities impose water quality standards for the management of river basins. These standards seek to maintain the water quality at a desired target level by defining discharge limits for conventional, specific, or priority pollutants. To satisfy the relevant standards, the following water quality constraint is established:

499
$$\sum_{t=1}^{T}\sum_{i=1}^{N}\sum_{j=1}^{M} \left(x_{ij}^{t}\delta_{ij}\psi_{ij}h_{j}^{u} - x_{ij}^{t}\delta_{ij}\eta_{ij}h_{j}^{u}\right) \times 10 \le H^{u}$$
(32)

where ψ_{ij} is the sewage water treatment rate, which is the proportion of sewage water that is treated; h_{j}^{u} (mg/L) is the concentration of the *u*th contaminant per unit treated water required by the *j*th user; and H^{u} (kg) is the upper limit of the *u*th contaminant allowed to be discharged in the study area.

504 **2.3 Third layer of the PTSOA decision-making model**

After obtaining the results for the former two stages of the allocation process and the two levels of decision-making, the third model layer is constructed to achieve regional synergy synergy. It refers to the collaborative allocation of water resources in different sub-regions of a city, and it is intended to balance and maximize the interests of each

509 sub-region as much as possible. Additionally, the needs of different kinds of water users 510 in different sub-regions can be met to the greatest extent possible with this approach. 511 Therefore, the three dimensions of synergy are also fused in this layer. The third stage 512 of the process (the water in different departments is supplied to different kinds of water 513 users, namely, residential users, industrial users and municipal users, in different sub-514 regions) is optimized in this layer. After department-level decision-making, conflicts of 515 interest inevitably occur among various water users in different sub-regions of a city. 516 Therefore, the third layer considers regional-level decision-making to to-coordinate 517 water needs and avoid conflicts of sub-regions in the city. Moreover, the various 518 development priorities of sub-regions are emphasized by adjusting certain hyper-519 parameters in the third layer. This layer is established based on the allocation scheme 520 obtained in the second layer of the hierarchy, and the time scale of this layer is the same 521 as that of the second.

522 Although water pollutants are controlled in the second layer, the detailed spatial 523 distribution of pollutants remains unknown. If one of the sub-regions emits a greater 524 pollution load than others such that the river pollution limit is exceeded, it constrains 525 sustainable development and undermines the fairness of the allocation. To ensure the 526 coordination of water quality among regions, the representative pollutant concentration 527 of the main reach in each sub-region after configuration should meet the relevant 528 environmental capacity requirements. If these requirements are not met, then the objective function for this sub-region will call for a punishment, and more 529

environmentally friendly plans will be searched. After sewage with pollutants is 530 transported from outlets to water bodies, advective transport, longitudinal dispersion 531 532 and transverse mixing will occur. At the same time, physical, chemical and biological 533 interactions will occur in the water body. To objectively describe the degradation of pollutants in water, it is necessary to use mathematical models to simulate physical 534 dynamics. Due to the heterogeneity of pollutants entering water bodies and the 535 uncertainty of hydrological processes, it is usually of little practical significance to 536 537 calculate the change in river water capacity over time. A steady-state model is therefore 538 used to calculate the water capacity of the target water body (Cetintas et al., 2010; 539 Zhang et al., 2019). When water quality changes are studied at the annual scale and 540 complete mixing is assumed, the following equation can be used to describe the water 541 quality change, as shown in Eq. (33):

542
$$\frac{Vdc}{dt} = Q(Ce-C) + Sc + r(c)V$$
(33)

where V (m³) is the volume of water; Q (m³/a) is the flow in and out of the system at equilibrium; Ce (g/m³) is the contamination concentration in the inflow (g/m³); C is the pollutant concentration; Sc -denotes other external pollution sources (m³/a); and r(c) is the reaction rate of pollutants in water (g/m³/a). The above equation can be defined as the basic mass balance of a water body in a completely mixed system. Because the pollutants are evenly mixed in each small interval, the horizontal and vertical concentration gradients of pollutants can be neglected. Therefore, the model of water quality in mixed rivers under steady-state design conditions is adopted (Yue etal., 2021):

552
$$W_{c} = 31.54 * \left[C_{s} \cdot \left(Q_{p} + Q_{E} + Q_{s} \right) - Q_{p} \cdot C_{p} \right]$$
(34)

where W_c represents the water environmental capacity (t/a); Q_p is the flow in the 553 reach (m³/s); C_P is the pollutant concentration in the river (mg/L); Q_E is the sewage 554 discharge (m³/s); Q_s is the total flow of nonpoint sources into the reach above the 555 control section (m³/s); and C_s is the target concentration of river pollutants (mg/L). 556 557 The result calculated based on the total hydrological capacity standard is often 558 relatively large, which is generally referred to as nonconservative. To conform to real 559 conditions, the concept of a nonuniformity coefficient is introduced for correction: $W_c = \alpha \cdot W_c = 0.6 \cdot W_c$ 560 (35)

This coefficient is used to assign a punishment if the water quality exceeds the relevant value in a given sub-region. Based on the coefficient value, the objective functions and constraints are adjusted accordingly. Finally, the daily decision alternatives for water allocation from water departments to water users are obtained at the regional decision level.

566 2.3.1 Objective function of the third layer

567 Regional objective function: <u>Maximization maximization</u> of the comprehensive
 568 benefits of each sub_region

569
$$\max f_3(x) = \sum_{t=1}^T \sum_{i=1}^I \sum_{j=1}^{J_r} x_{ij}^t b_{ij} - \sum_{j=1}^{J_r} \left(D_j - \sum_{t=1}^T \sum_{i=1}^I x_{ij}^t \alpha_{ij} \right) \times \omega_j - P_r(r_{nz}) - G_r(x_{ij}^t) \times q \quad (36)$$

570
$$P_r(r_{nz}) = e_i \times \sum_{p=1}^{\Pr} P_p^{pump} \times \nabla t_r + x_{ij}^t \delta_{ij} \psi_{ij} \omega_{ij}$$
(37)

571
$$\nabla t_{r} = \sum_{t=1}^{T} \sum_{n=1}^{N} \sum_{z=1}^{Zr} \left\{ \left(l_{nz} + \left(r_{nz}^{t} \theta_{nz} / CAS_{nz} \right) \right) / \left(Q_{nz}^{\max} / CAS_{nz} \right) \right\} / 3600$$
(38)

572
$$G_r(x_{ij}) = \sum_{z=1}^{Zr} \sum_{u=1}^{U} \left(Q_{z,u,r}^{final} - Q_{z,u,r}^0 \right)$$
(39)

where b_{ij} (yuanChinese Yuan/m³) is benefit per unit of water supply for the *j*th user; 573 ω_i is the penalty coefficient per unit of water deficiency; j=1,2,...,Jr; Jr is the number 574 of water users in *r*th sub_region; r=1,2,...,R; $P_r(r_{nz})$ is the penalty function for cost in 575 the *r*th sub-region; e_i (yuanChinese Yuan/kW·h) is the unit electricity fee; P_p^{pump} 576 577 $(kW \cdot h)$ is the electrical power consumed by the *p*th pump at a pump house in each hour; 578 p ranges from 1 to Pr; Pr is the total number of pumps in the rth sub_region; ∇t_r (h) 579 is the time required for water transfer to provide support for the inland river flow in the 580 *r*th sub_region; ω_{ij} (yuanChinese Yuan) denotes to the fee paid for sewage treatment; l_{nz} (m) is the length of a water diversion pipe from reclaimed water source *n* to the *z*th 581 582 inland river; z ranges from 1 to Zr; Zr denotes the number of inland rivers in the rth sub-region; CAS_{nz} (m²) is the cross-sectional area of a pipe from the *n*th reclaimed 583 water source to the zth inland river; Q_{nz}^{max} (m³) is the maximum overflow capacity of 584 the diversion pipe from the *n*th reclaimed water source to the *z*th inland river; $G_r(x_{ij})$ 585 is the penalty function for substandard water quality in the *r*th sub_region; $Q_{z,u,r}^{final}$ (mg/L) 586 587 is the final concentration of the uth pollutant in the control section of the zth inland river

in the *r*th subregion after optimal configuration; $Q_{z,u,r}^0$ (mg/L) is the initial concentration of the *u*th pollutant in the *z*th inland river in the *r*th sub_region; and *q* is the penalty coefficient for substandard water quality in the *r*th sub_region. The number of objective functions in this layer depends on the number of sub_regions divided in the city, which is based on local conditions.

593 2.3.2 Constraints of the third layer

594 Water quality constraints

602

Mathematical models are often developed to help satisfy the water quality standards at monitoring points (Zhang et al., 2019; Pourshahabi et al., 2020; Friesen et al., 2017). However, for some cities with very few monitoring points, such approaches may lead to good water quality in the monitored sections and poor water quality in other sections. In these circumstances, the quality of water bodies in each sub_region of a city is not simultaneously maintained. To maintain the water quality in all sub_regions of a city at the desired target level, the water quality constraint in Eq. (40) is established:

$$Q_{z,u,r}^{final} \le Q_{z,u,r}^{control} \tag{40}$$

603 where $Q_{z,u,r}^{control}$ (mg/L) denotes the control standard for the *u*th pollutant in the control 604 section of the *z*th inland river in the *r*th sub_region.

605 **2.4 Model solution**

606 2.4.1 Synergy degree evaluation

Enhancing the understanding of the synergy among water allocation alternatives to 607 608 achieve broad coordination and equilibrium is crucial. The evaluation of the synergy of 609 a complex water resources systemwater system is strongly related to multiple complex 610 interactions, such as the interactions among different processes, users, and regions. 611 However, these interactions have rarely been explicitly captured in prior evaluations of 612 water allocation. One of the key network metrics used in network analysis, connectivity, 613 is a promising measure of the degree of coordination among different objectives in complex systems (Weitz et al., 2018). Connectivity reflects the connectedness of a 614 615 given link to all possible links in the network, and the strength of each link is weighted, 616 reflecting the number and strength of correlations (Felipe-Lucia et al., 2020). In this 617 study, connectivity is used to embody coordination in the context of synergy, as shown 618 in Eq. (26). Due to the limited supply of water resources, competition among different 619 objectives is unavoidable, and the objectives cannot be fully optimized to equal extents, 620 i.e.,-, an increase in one target output may decrease another output. Therefore, 621 equilibrium is integrated as another vital part of the synergy devoted to maintaining a 622 balance among the satisfaction of each goal in a system. The equilibrium based on the principle of information entropy (Gao et al., 2013; Zivieri, 2022) is shown in Eq. (27). 623 624 Information entropy_-is a measure of the uncertainty associated with a random variable 625 and is used to quantify the information contained in a message, usually in bits or bits/symbols; furthermore, it has been widely used to represent the fairness or 626 equilibrium of a system (Chen et al., 2022; Zhao et al., 2022). When H is low, the level 627 628 of equilibrium in the system is high. This factor is also used to be compared with the 629 proposed index. By combining the quantification of coordination and equilibrium, the 630 synergy degree is appropriately determined (Eq. (29)). Notably, the total synergy index (TSI) of a system is used for both generating candidate management alternatives in the 631 632 generation phases of PTSOA and performing assessments of the associated level of 633 synergy, as shown in Eqs. (41-44).

634
$$SSI_{ob_i} = \frac{\sum_{j=1}^{N} c_{ij} \times (ob_i + ob_j)}{\sum_{j=1}^{N} (ob_i + ob_j)}, i \neq j$$
(41)

635
$$H(S) = -\sum_{i=1}^{N} \frac{(1 - u_{ob_i})}{N} \log \frac{(1 - u_{ob_i})}{N}$$
(42)

636
$$u_{ob_i} = \frac{ob_i - ob_{i,\min}}{ob_{i,\max} - ob_{i,\min}}$$
(43)

637
$$TSI = \frac{\sum_{i=1}^{N} SSI_{ob_i}}{H(S)}$$
(44)

638 where SSI_{ob_i} is the connectivity of the *i*th object; c_{ij} is the Pearson correlation 639 between the *i*th object and *j*th object; ob_i and ob_j are the values of the *i*th and *j*th 640 objective functions, respectively; TSI-is the synergy index of the system; H(S) is 641 the overall equilibrium of all objects based on the principle of information entropy, and

it is abbreviated as H in the following; u_{ob} is the standardized value of the *i*th object; 642 N is the total number of objects in the system; $ob_{i,min}$ and $ob_{i,max}$ are the minimum 643 644 and maximum critical thresholds of the parameter ob_i , respectively. SSI is ranged from 645 0~N, and higher SSI indicates higher connectivity of the objects in the system which 646 means they are easier to promote each other, and lower SSI indicates lower connectivity 647 which means the promotion is hard to realize and obstacles to each other may occur. H 648 is ranged from $0 \sim N^{1}\log(1/N)$, and lower H indicates better overall equilibrium and 649 higher H indicates worse overall equilibrium from objective perspective. TSI is greater 650 than 0. When a water resource system's TSI value is higher, the degree of synergy is 651 higher; conversely, when a water resource system's degree of synergy is lower, the TSI 652 value is lower. In our application, based on actual evaluation, we define when $TSI \ge 5$, the degree of synergy is considered satisfactory. $5 > TSI \ge 3$ is defined as moderate and 653 654 3>TSI is defined as low. In our application, based on actual evaluation, the criteria as divided. When $TSI \ge 5$, the degree of synergy is considered satisfactory. Additionally, 655 $5 > TSI \ge 3$ is moderate synergy degree and 3 > TSI is low degree. 656

657 2.4.2 Hierarchical optimal algorithm design for the PTSAO PTSOA 658 model

Based on the algorithm design with a hierarchical objective function proposed by Li et al. (2022), a new level is added to the original two levels of the algorithm, and the alternative generation phase is improved for better synergy. In this algorithm, the objective functions in the upper decision level is first satisfied, and then the lower-level objective function provides an optimal result based on the results of optimal allocation in the upper level. To provide as comprehensive solutions as possible, the decision
alternatives need to be classified into different sets for further selection. In addition, the
synergy degree of the result of each layer is calculated to select optimal decisions
among all Pareto front solutions. The detailed steps of the hierarchical optimal
algorithm are as follows:

In the first level, calculate the objective function (city level) values for the social,
economic and sustainable development components, and sort the results with
NSGA-III (Pourshahabi et al., 2020; Chen et al., 2017) to obtain each Pareto
front F₁, F₂, ..., F_i.

673II.Classify the Pareto fronts into K (K is determined based on the diversity of674policies) elements with the K-means algorithm (Liu et al., 2022), which is used675to partition a data set into K distinct and non_overlapping clusters. To perform676K-means clustering, we first specify the desired number of clusters K. Then, the677K-means algorithm is used to assign each observation to exactly one of the K678clusters.

679 III. Calculate the synergy degree of each individual in the front, and select the
680 solution that yields the greatest synergy in each cluster. *K* solutions are obtained
681 in the first layer.

682 IV. Use the selected K solutions in the first layer to establish constraints in the 683 second layer. Solve the objective function of the second layer with NSGA-III.

684 V. Calculate the synergy degree of each individual in the front and select the

36
solution that yields the greatest synergy as well as the two solutions that
maximize the conventional and unconventional water supply department
objective functions in all Pareto fronts with *K* preconditions.

- 688 VI. The three selected solutions in the second layer are used to establish constraints
 689 in the third layer. Solve the objective function of the third layer with NSGA-III
 690 under the three preconditions.
- VII. The synergy degree of each individual in the front is calculated, and the solution that yields the greatest synergy in the third layer is selected. Three solutions are obtained considering the synergy in the former two layers. Finally, the synergistic configurations optimal for all stages in the whole process are identified considering the synergy among decision levels, processes and time scales.

697 **3. Application**

698 3.1 Study area

Yiwu city is selected as a case study to validate the applicability of the PTSOA model.
Yiwu This city is in Southeast China, located from 119°49 'E-120 °17' E and 29°02 '13
"N-29 °33' 40" N and. The city covers an area of 1105 km². The area is characterized
by a scarcity of water resources, and the conventional water supply is under severe
stress. The regional water consumption depends heavily on transported water and

704	external water transfer. The per capita water resources is in total 622 m ³ , which is only
705	22.6% of the provincial average and 19.1% of the national average. Moreover, the
706	problem of water pollution has become a bottleneck constraint for the development of
707	Yiwu city. Therefore, it represents a typical water-scarce city with limited conventional
708	water. Notably, water quality in Yiwu has been subjected to significant environmental
709	stress because of the negative effects of wastewater discharge with the rapid
710	development of industry. The current water quality is poor, with Class V-water, and the
711	main pollutant concentrations exceed the corresponding standards (Zhejiang Natural
712	Resources and Statistical Yearbook on Environment, 2020). As shown in Fig. 3, the
713	Yiwu River crosses the city from northeast to southeast. Additionally, there are six
714	ecological water compensation outlets in six main tributaries in the Yiwu River. In Fig.
715	3, the white labels indicate five sub-regions in the city, the black labels near the
716	reservoirs are their names, the black labels named O1~O6 indicates the name of the
717	water distribution outlets and the labels near the lifting pump station are their names.
718	



3.2 Generalization of the <u>complex water resources system</u>water system

725 An initial multi-source, complementary and mutually regulated system has been developed for Yiwu, and this system spans the entire urban water cycle (water source-726 water supply-water use-drainage-drainage collection-recycling and reuse). To apply the 727 728 optimal water allocation model to the complex real-world water system, all 729 stakeholders in the complex water resources systemwater system should be 730 schematized into a topological system, as shown in Fig. 4. The diagram comprises five 731 modules: water sources, water supply, water use, water drainage and external discharge 732 for all stakeholders.

733 The first module includes seven main reservoirs, two water diversion projects, the Central Sewage Treatment Plant and the Yiwu River. The seven reservoirs and two 734 735 water diversion projects (as shown in Table 1) supply high-quality water. There are 736 complex connections between the first and second modules. For example, two reservoirs supply water to one waterworks or one reservoir feeds two or three 737 738 waterworks simultaneously. The reservoirs also supply some of the agricultural and 739 ecological waters to subareas of the city. The Yiwu River, with a total length of 38.39 740 km and 21 first-class tributaries in the city, and the Central Sewage Treatment Plant, as 741 shown in Table 2, are low-quality water sources. Additionally, excluding water from 742 reservoirs, most agricultural irrigation water is supplied from surface water stored in hundreds of small reservoirs and mountain ponds. Since tThere are no data available
for agricultural irrigation water, which accounts for only a small portion of the total
water demand in the area, and most agricultural irrigation water is supplied from surface
water stored in hundreds of small reservoirs and mountain ponds (2020-Yiwu
Ecological Environment Status Bulletin, 2020). So, this water volume is ignored in the
model.

749 For the second module, high-quality water piped from reservoirs is transported to 750 nine urban and rural centralized waterworks (as shown in Table 2). The Yiwu River 751 distributes low-quality water to the Yijishan Ecological Water Plant and Sufu Industrial 752 Water Plant through the Yijishan and Baisha Water Pump Stations, respectively. The 753 water discharged at the Central Sewage Treatment Plant is transferred to the Choujiang 754 Industrial Water Plant. Based on the water supply project distribution and the economic 755 as well as social development levels-, Yiwu is divided into five districts, as shown in 756 Table 3: the Central District, Yidong District, Yibei District, Yinan District and Yixi 757 District. The third module comprises both high-quality water users (high-quality water 758 users consist of urban and rural domestic water users and some industrial water users 759 in the water supply network of urban and rural public water plants) and low-quality 760 water users (low-quality water users include other industrial water users, municipal 761 water users and ecological water replenishment for inland rivers) in each district. There 762 are nine sewage treatment plants in the fourth module (which focuses on the drainage stage), as shown in Table 2. The unreused water from sewage treatment plants is 763

discharged to the external environment. Reuse processes are also considered in thesystem.





768

Fig. 4. Schematic diagram of Yiwu city

769 3.3 Parameter determination

770 According to the flow duration curve of the annual natural inflow data for 51 years

771	(1963-2014), three years with exceedance probabilities of 50%, 75% and 90% are
772	selected to represent normal (1984.1–1985.1, annual mean inflow: $1.33 \times 10^{8} \text{ m}^{3}$), dry
773	(2008.1–2009.1, annual mean inflow: 1.11 \times 10 8 m³), and extremely dry (1971.1–
774	1972.1, annual mean inflow: 0.63 \times 10 8 m^3) scenarios, respectively. In addition to
775	inflow, the data used in the PTSOA model mainly include the data for the parameters
776	in each layer. Water demand values were calculated using the Yiwu City Water
777	Resources Comprehensive Plan 2020, as shown in Table 1.

778

Table 1 Water demands of various regions in Yiwu in 2020 (10⁴ m³)

Subregion	Yibei	Yidong	Zhucheng	Yixi	Yinan	
water demand	1695	572	11813	2198	2045	

779

The water resources fees paid to the government<u>are in</u> total 0.3 <u>yuanChinese</u> 781 <u>Yuan/m³</u>. The parameters of the reservoirs and external water division projects in Yiwu 782 city are listed in Table 2.

783

Table 2 Parameters of the reservoirs and external water division projects

				Flood	
				limit	
	Water Fee	Initial	Dead	storage	Absolute
Reservoirs &	yuan(Chines	storage	storage	capacity	storage capacity
External sources	<u>e Yuan</u> /m ³)	<u>(</u> 10 ⁴ m ³)	<u>(</u> 10 ⁴ m ³)	(10 ⁴ m ³)	(10^4 m^3)
Badu	0.99	1359	49	2688	2639
Qiaoxi	1.30	1505	77	2933	2856
Weixin	0.37	500	17	483	466
Baifeng	1.05	1013	15	2010	1995
Fengkeng	1.15	778	55	1501	1446
Yankou	1.49	1820	499	3140	2641
Changyan	0.70	491	41	940	899
Pujiang Project	1.00	0	0	3000	3000

Dongyang Project	1.00	0	0	5000	5000	
						-

784

794

The Tennant method is applied to calculate the ecological water demand. In this method, the relationship between the annual average discharge and habitat quality is considered, and the percentage of the annual average natural runoff is used as the recommended value of the ecological water demand for a given river channel. According to the recommended values, the percentage of runoff required for the fish spawning period from April to September is 30% and the percentage runoff in the general water consumption period (October to March) is 10%.

Based on observations obtained with the F601 evaporator (a standard evaporationinstrument widely used in China), evaporation is calculated as:

$$EP = E \times k \tag{45}$$

795 where EP(mm) is the evaporation of a reservoir; E(mm) is the observed evaporation; 796 and k is a reduction coefficient. According to observations, the difference of 797 this coefficient is quite slight within a small watershed (Zhao, 2014). Thus, k is simplified to the same value 0.88 is the same for every reservoir and varies throughout 798 799 the year according to expert experience(Zhao, 2014) (Zhao, 2014). The prices of 800 conventional water and reclaimed water are 1.7 and 2.6 yuanChinese Yuan/m³, respectively. In our application of the model, this precipitation component associated 801 802 with the water sources were calculated by the Thiessen polygon method (Liu et al., 803 2014) based on the measured data of seven rainfall stations (Shi Caotou, Suxi, Yiwu, Fotang, Baifeng, Fengkeng, Changfu) in the basin in normal (1984.1–1985.1), dry
(2008.1–2009.1), and extremely dry (1971.1–1972.1) scenarios, relatively.

806 The monthly mean monitoring data for effluent pollutant concentrations and the 807 daily maximum processing capacities of sewage treatment plants were obtained from 808 the monitoring systems of the sewage treatment plants. For example, the concentrations of COD, NH3-N, TN, and TP in the sewage of the Jiangdong Sewage Treatment Plant 809 are 13.80 (mg/L), 0.22 (mg/L), 6.02 (mg/L), and 0.13 (mg/L), respectively. The daily 810 811 maximum processing capacity of Jiangdong Sewage Treatment Work is 12 (10^4 t/d). 812 The effluent quality of sewage treatment works satisfies the Class A Standard used in 813 China. The maximum capacities of the Baisha pump station, Yijiashan pump station, 814 Choujiang pump station and water treatment centre pump station are 13 t/d, 13.5 t/d, 10 815 t/d, and 4.5 t/d, respectively.

816 Additionally, the environmental capacities of the six tributaries that are replenished 817 with ecological water are calculated according to Eqs. (33)-(35), and the results are 818 listed in Table 3. COD, TP and TN are major pollutants in Yiwu City (Yiwu Ecological 819 Environment Status Bulletin, 2020), and they are also major controlled pollutants of all 820 the monitoring sections. So, these thereey were selected as representative pollutants in 821 the tributaries to guarantee the water environmental quality of inland rivers. The water 822 quality goals for the tributaries must conform to the Class III standard according to GB 823 5749-2006 in China. The unit electricity price of pump stations in Zhejiang Province is 0.41 yuanChinese Yuan/kW h. GB50014-2006 (2014 edition) stipulates that the 824

825	comprehensive urban domestic sewage quota should be 8090%, and the urban
826	comprehensive domestic sewage quota should be 90% in areas with extensive drainage
827	facilities. There are plenty of many influencing factors in the model, and the most
828	important ones among them are the value of water demand, the value of available water
829	and some key hyper-parameter. According to the "Yiwu-Jinhua Water Resources
830	Bulletin 2020(2020)", the urban comprehensive domestic sewage quota is set to 90%,
831	and the sewage treatment rate is set to 100%. The benefits per unit water supply for
832	different users in different subregions are determined from the Yiwu Water Price
833	Adjustment Plan (2020)-2020.

834

835 **Table 3** Area and environmental capacity of tributaries

Name of tributory	$\Delta real (1rm^2)$	Class III		
Name of tributary	Area (km ²)	COD (t/a)	TN (t/a)	TP (t/a)
Chengdong	3.4	188.1	4.7	0.4
Chengzhong	8.7	432.7	31.5	3.8
Chengxi	6.3	302.5	9.5	2.3
Chenganan	7.1	318.8	0	3.6
Hongxi	12.5	778.8	138.8	7.9
Dongqingxi	38	1271.4	221.5	12.7

836 4. Results and discussion

By solving the PTSOA model for Yiwu city, synergistic optimal water allocation results
for different layers (across different decision levels, water use sectors, and sub_regions)
are obtained under normal, dry and extremely dry conditions. Pareto sets are obtained
across 500 runs and 1000 iterations (<u>Fin most cases</u>) of the PTSOA model with the

proposed hierarchical optimization algorithm. <u>If Tthe feasible solutions could not be</u>
found in some cases, the number of iteration would be increased. It took approximately
mean-34 h of CPU time on a computer with 32 GB memory and intel corei7@3.4 GHz
of CPU. Therefore, in this study, each iteration for a single trial solution takes 0.24 s of
CPU time on the computer with the named specifications.

4.1 Results of the The fFirst layer of the PTSOA model for synergistic optimal water allocation

To demonstrate the relationship among conflicting objectives, sets of Pareto solutions 848 849 for the first layer under normal, dry and extremely dry conditions are shown in Fig. 5. 850 The optimization using the Pareto concept allows the operator to choose an appropriate 851 solution depending on the prevailing circumstances and analyse the trade-off among 852 the conflicting objectives. In each of the figures, the total water supply shortage, total water supply benefit and total amount of water retained in reservoirs in Yiwu city are 853 854 plotted. The colour of the markers indicates the classification of the solutions of the Kmeans method, as described in Section 2.4.2. All of the decision alternatives are 855 856 classified into six groups marked in different colours for broad-scale decision-making. The names of the classes are marked in the figure in red (for example, K1-1 represents 857 858 the first class of solutions in the normal scenario, and K3-2 represents the second class 859 of solutions in the extremely dry scenario). The red arrows indicate optimization directions. The ideal solution is located at the top-right corner (low total water supply 860

861	shortage, high total water supply benefit, and relatively high total amount of reserved
862	water in reservoirs) of the plot. The geometries of the trade_offss vary significantly
863	across the applications, as is expected given the different hydrological conditions.
864	Generally, the total water supply shortage and the total amount of water retained in
865	reservoirs show an inverse relationship. In contrast, the total water supply benefit shows
866	a direct and positive influence on the total water supply shortage. The water shortage
867	varies in the range of -1.2×10 ⁶ ~0.8×10 ⁵ m ³ , -0.5×10 ⁵ ~2.0×10 ⁶ m ³ , 0~3.5×10 ⁶ m ³ in
868	normal, dry and extremely dry scenarios respectively. The average water demand is
869	around 1.8×10 ⁸ m ³ , and water shortage of the selected decision alternatives are all less
870	<u>9×10⁶ m³ So, The-the</u> water supply reliability of the selected decision alternatives is
871	greater than 95% under normal, dry and extremely dry conditions with the consideration
872	of water demand. The total amount of reserved water in reservoirs under normal
873	scenarios varies in the range of 2.91×10^7 m ³ to 6.14×10^7 m ³ , which is much higher than
874	that under the extremely dry scenario, with a value of 1.44×10^7 m ³ to 2.93×107 m ³ .
875	This finding demonstrates that the optimal allocation is able to reconcile the present
876	demand and future needs, even in extremely dry scenarios. The total water supply
877	shortage in all scenarios is less than 5% of the water demand, which indicates that the
878	guaranteed water supply is greater than 95%.



893	PTSOA solutions under three scenarios, as shown in Fig. 6. In the PTSOA model, the
894	Pareto solutions with the best TSI values are input to the second layer for further
895	optimization. Thus, the red points in Fig. 6 represent the selected schemes for all classes.
896	We observe that the variation in the TSI is consistent with that in the SSI in some, but
897	not all cases. In some cases, differences are mainly caused by the influence of H, which
898	influences the optimal hydrological equilibrium, especially in dry conditions.
899	Although normal conditions are most conducive to achieving equilibrium, the better H
900	value in extremely dry conditions than in dry conditions seems nonintuitive. However,
901	these These results suggest that when water is very limited, equally limited water is
902	supplied to all users, thus enhancing the overall equilibrium. We note that the SSI value
903	is higher in the normal scenario than in the other two scenarios. We attribute this to
904	relatively abundant water being useful for stakeholders to achieve synergy due to the
905	reduced competition compared to other cases. The TSI values reach maximums of 5.36,
906	7.37 and 10.82 under normal, dry and extremely dry conditions, respectively.
907	In Fig.6, the value of TSI are significantly diverse among different scenarios as
908	well as different solutions. H is widely used to evaluate the equality of different
909	solutions (Gao et al., 2013;)(Li et al., 2022). As a contrast, the value of H, which is used
910	for comparison and construction of TSI, show slight differences among solutions and
911	even are the same in some classes. Therefore, it is difficult for decision makers to select
912	the best solution among all candidates if we only use H for evaluation and selection in
913	the decision process. Compared to <i>H</i> , <i>TSI</i> introduces <i>SSI</i> into evaluation and the 51

914 <u>difference of coordination relationship between different schemes is distinguished by</u>
915 <u>SSI. But H only pays attention to the equity among the stakeholders. So, TSI is more</u>
916 <u>effective and validity than H in some extent. Additionally, Since since the</u> TSI is used
917 to illustrate the synergy of allocation plans under certain conditions, the three kinds of
918 TSI values are not comparable.





920 Fig. 6. Comparison of *TSI* (total synergy index), *SSI* (total connectivity) and *H*



922	the (K1) normal, (K2) dry, and (K3) extremely dry scenarios. (K1-n, K2-n and K3-n
923	represents the nth class of solutions in the normal, dry and extremely dry scenario
924	separately, n=1~6.)
925	Fig. 6. Comparison of TSI (total synergy index), SSI (total connectivity) and H-
926	(overall equilibrium) values among various Pareto solutions in different classes for-
927	the (K1) normal, (K2) dry, and (K3) extremely dry scenarios.
928	As an example, Fig. 7 provides the specific water supply decision alternatives for
929	the first layer that maximize synergy in each cluster under normal conditions. The water
930	allocation plans for the seven main reservoirs and two external water diversion projects
931	in every month of the configuration period are displayed. All reservoirs and water works
932	are represented by abbreviations based on their full names in Fig. 7. For example, QX-
933	CB is the label for the water supplied from Qiaoxi Reservoir to Chengbei Water Works.
934	The water volumes supplied by Qiaoxi Reservoir to Chengbei Water Works (ranging
935	from 1.78×10^7 m ³ to 3950×10^4 m ³) and from the Pujiang External Water Division
936	Project to Chengbei Water Works (ranging from $2.57 \times 10^7 \text{ m}^3$ to $3 \times 10^7 \text{ m}^3$) are relatively
937	high in all clusters. This result is consistent with the fact that Chengbei Water Works is
938	one of the main conventional water sources for the central city area, a region that
939	accounts for more than 50% of the total water demand of Yiwu city. The water supplied
940	by the two external water diversion projects from August to December is higher than
941	that in other months. The mean monthly precipitation in these months is only 58-74%
942	of the mean annual precipitation in Yiwu, so more external water is supplied for 53



944 their two connected waterworks.



953 The 6×3 decision alternatives selected in the six clusters of the optimal first-layer results 954 in the normal, dry and extremely dry scenarios are inputs into the second layer for 955 further optimization. As shown in Fig. 8, the total amount of water retained in water works and the amount of unconventional water supplied show a negative correlation. 956 In the alternative generation phase of game bargaining between the two objectives, the 957 958 greater the total amount of water retained in water works is, the greater the amount of 959 unconventional water supplied will be will be supplied, which indicates that more conventional water will be saved when more unconventional water is supplied. 960 Conversely, the amount of unconventional water supplied is affected by the total 961 amount of water retained in water works. 962

963 In the second layer, three alternatives in each scenario are selected as prior 964 conditions for further optimization. In addition to the two individual extrema of the two 965 objectives, the alternative that yields the best synergy is also identified, and it is similar 966 to that in the first layer. In the normal scenario, the TSI values are -0.90, -1.02 and -0.88 967 in the cases with the optimal conventional water supply, unconventional water supply 968 and synergy, respectively. The most synergistic approach includes only 7.08×10⁴ m³ 969 more conventional water retained than that in the conventional water supply cases and 970 only 9.72×10^4 m³ more than that in the optimal unconventional water supply case. 971 Therefore, not only is the best TSI value obtained, but the requirements of both 972 conventional and unconventional water supply departments are met. The TSI of the 973 most synergistic solution is the highest under dry conditions, with a value of -0.79. 974 Overall, the total amount of water retained in the water works ranges from 3.95×10^7

975 m^3 to $5.75 \times 10^7 m^3$, $3.12 \times 10^7 m^3$ to $5.31 \times 10^7 m^3$, and $2.43 \times 10^7 m^3$ to $3.96 \times 10^7 m^3$ for the

976	three types of conditions. The total amount of unconventional water supplied ranges
977	from 5.95×10^7 m ³ to 7.48×10^7 m ³ , 6.34×10^7 m ³ to 7.56×10^7 m ³ , and 6.28×10^7 m ³ to
978	7.37×10^7 m ³ in thenormal, dry and extremely dry scenarios, respectively. <u>Moreover</u> ,
979	by selecting the solution with the highest <i>TSI</i> , 7.35×10^7 m ³ , 7.56×10^7 m ³ , and 7.37×10^7
980	m ³ of unconventional water would be supplied as an effective supplement to
981	conventional water. In the other word, conventional water would be saved by our
982	proposed model and index in the three scenarios. It is notable that the drier the
983	conditions are, the lower the amount of water retained in water works and the greater
984	the amount of unconventional water supplied. Thus, This this approach is useful for
985	cities to mitigate the risk of drought. Additionally, based on the constraints regarding
986	the contaminants allowed to be discharged, more than 1272.21 t and 48.81 t of COD
987	and ammonia nitrogen emissions are avoided per year. In other words, the balancing of
988	the two objectives is beneficial for managers to determine an equilibrium solution that
989	satisfies the relevant demand and successfully avoids surplus conventional or
990	unconventional water supply in terms of sustainable development.







1000 with the hierarchical optimal algorithm in the normal, dry and extremely dryscenarios. F1 represents the total amount of water retained in water works (10⁴ m³), 1002 and F2 represents the amount of unconventional water supplied (10⁴ m³). The-1003 direction of optimization is from the top-right corner to the bottom-left corner.

4.3 Results of tThe tThird layer of the PTSOA model for 1004 synergistic optimal water allocation 1005

1006 After selecting the three scenarios that yield the best synergy and the two best objective 1007 functions for characterizing all Pareto fronts of the second layer in each scenario, these 1008 3×3 solutions are input to the third layer for further optimization. Fig. 9 shows the trade-1009 offs among the five objectives in the third layer of the PTSOA model for the (S1) normal, 1010 (S2) dry, and (S3) extremely dry scenarios (these abbreviations are used to distinguish 1011 these results from those of the above two layers). The number following the '-1012 ' represents the selected solution from the second layer. For example, S1-1 represents 1013 the normal scenario with the minimum total amount of water retained in water works, 1014 S1-2 represents the normal scenario with the maximum unconventional water supply 1015 and S1-3 represents the normal scenario with the maximum synergy degree in the 1016 second layer. In each of these plots, the abscissa denotes the identifier for the objective 1017 functions, which ranges from 1 to 5, and the ordinate gives the objective values in the 1018 Pareto fronts (10⁴ yuanChinese Yuan). The five dimensions include the comprehensive 1019 benefits of the Yibei (1.0 dimension), Yidong (2.0 dimension), Yixi (3.0 dimension),

1020	Yinan (4.0 dimension) and central city (5.0 dimension) sub_regions. As shown in the
1021	figure, the central city achieves the most comprehensive benefit among the five
1022	citiessub-regions. This is primarily attributed to the large population and intensive
1023	industry in this area. However, the benefits in the other four sub-regions are also high
1024	compared to recent levels and those achieved with traditional allocation methods, as
1025	shown in Table 9. Interestingly, the comprehensive benefits in the sub-regions are
1026	greater in the scenario with the maximum synergy degree under normal conditions than
1027	in the other two scenarios. Technically, tThe total comprehensive benefits in the five
1028	sub_regions in this scenario are approximately 2.3×10 ⁸ 5.1×10 ⁸ yuanChinese Yuan
1029	higher than those in other cases, which indicates that the solution with the highest
1030	synergy degree in the second layer is the best choice for managers in normal years.
1031	However, the various subregions obtain the greatest benefits when maximizing the
1032	unconventional water supply in dry and extreme scenarios. This result indicates that
1033	increasing the use of unconventional water in dry and extremely dry years would
1034	significantly increase the potential benefits.





1038	PTSPOA model attained across all runs for the (S1) normal, (S2) dry, and (S3)
1039	extremely dry scenarios (S1-1 represents the normal scenario with the minimum total
1040	amount of water retained in water works, S1-2 represents the normal scenario with the
1041	maximum unconventional water supply and S1-3 represents the normal scenario with
1042	the maximum synergy degree in the second layer)
1043	Fig. 9. Illustration of parallel-reference Pareto sets from the third layer in the
1044	PTSPOA model attained across all runs for the (S1) normal, (S2) dry, and (S3)-
1045	extremely dry scenarios
1046	
1047	Fig. 10 presents the optimal comprehensive benefit in each sub-region. In all
1048	scenarios, the central city is associated with the highest comprehensive benefit,
1049	followed by Yixi and Yinan, and the comprehensive benefit in Yidong is relatively low.
1050	This result may be related to <u>Yidong which has this subregion having</u> the smallest area
1051	(72.2 km ²) and the smallest population (7.7×10 ⁴ people). The comprehensive benefits
1052	vary among different solutions and scenarios. Among the three normal decision
1053	alternatives, F1, F2 and F5 are highest in S1-3, with values of 3.03×10 ⁹ yuanChinese
1054	Yuan, 9.90×10 ⁸ yuanChinese Yuan and 1.12×10 ¹⁰ yuanChinese Yuan, respectively. This
1055	indicates that considering the synergy degree could increase the comprehensive benefit
1056	in most sub-regions in the normal scenario. Among the alternatives in the dry and
1057	extremely dry scenarios (excluding F4 and F5), other objectives are highest in S2-2,
1058	with values of 2.84×10^9 yuan <u>Chinese Yuan</u> , 9.63×10^8 yuan <u>Chinese Yuan</u> and 2.67×10^8

 $\frac{1059}{1060}$ yuan<u>Chinese Yuan</u>, respectively. It suggests that maximizing the unconventional water supply is beneficial for the system in dry conditions. Additionally, F4 is highest, with a value of 2.29×10^9 yuan<u>Chinese Yuan</u>, in S2-3 among the three solutions in the dry scenario, and F5 is highest, with a value of 9.17×10^9 yuan<u>Chinese Yuan</u>, in S3-1 in the extremely dry scenario.





1074 F10. Comprehensive benefit in each area after the regional 1075 collaborative allocation of water resources

1076 **4.4 Discussion**

1077 To assist policymakers in understanding the complex and systemic nature of complex water resources systems water systems and reveal the dynamic interactions among 1078 1079 objectives, network analysis and optimization was-were applied. Complex network 1080 analysis helps reveal the interactions among three layers with different dimensions. We 1081 determine the level of synergy in complicated water systems, identify the challenges 1082 and opportunities for sustainable development of water systems in cities with various 1083 subregions, and provide valuable insights and specific action priorities for these 1084 regions. By revealing the interactions among different objectives, we determine the 1085 level of synergy in complicated water systems, identify the challenges and opportunities 1086 for sustainable development of water systems in cities with various subregions, and 1087 provide valuable insights and specific action priorities for these regions. In the networks 1088 shown in Fig. 11, each node represents an individual objective (F1, F2, F3, F4, and F5 1089 represent the comprehensive benefits in Yibei, Yidong, Yixi, Yinan and the central city, 1090 respectively), and pairwise objectives that are significantly (P < 0.05) correlated are connected by a link, where the strength of each link is related to the Pearson correlation 1091 1092 coefficient. The obtained networks with 5-five nodes were weighted and undirected 1093 (directionality can be estimated only if the direction of causality is known). The size of 1094 the circles in the figure indicates the connectivity of each objective. We considered 1095 trade-offs (i.e., negative correlations wherein one objective improves while the other 1096 worsens) among the objectives. In most scenarios, F5 was the relatively dominant 1097 objective, signifying that other objectives disproportionately deteriorated as progress 1098 was made towards the benefit of the central city, as shown in Fig. 11. It is evident that 1099 the trade-offs are more balanced in the scenarios with the highest degrees of synergy 1100 (S1-3, S2-3, and S3-3), which indicates that the trade-offs and competitions among the 1101 objectives are alleviated when synergy is considered. The links show that the conflicts 1102 of interest between F4 and F5 in scenarios S1-1 and S2-2 are extremely notable, 1103 suggesting that the comprehensive benefits in Yinan and the central city correspond to 1104 strong negative interactions in these cases. The connectivity of most objectives was 1105 relatively low in the trade-off network in the extremely dry scenario, but F5 played a 1106 dominant role in terms of negative interactions among objectives., although the 1107 connectivity of F5 was lower than other connectivities in most normal and dry scenarios. 1108 Moreover, as the scenario varied from normal to extremely dry, the impact of individual 1109 regional targets on the whole system diminished.



1121 SPEA-II, ε-MOEA, IBEA, and MOEA/D and Borg MOEA, to solve cases with 3+2+5 1122 mathematical objectives (3 objectives in the first layer, 2 objectives in the second layer 1123 and 5 objectives in the third layer) with the same constraints given previously for Yiwu 1124 city under normal, dry and extremely dry conditions. The constraints and common 1125 parameters, such as the maximum number of model simulations and the simulated 1126 binary crossover (SBX) rate, are set to those used in the PTSOA model. However, it is 1127 difficult to determine feasible decision alternatives with MOEAs, even though the 1128 number of iterations is increased to 20000 (which is far beyond that considered in the 1129 previous modelling) because the complexity of the system overshadows the 1130 optimization capabilities of these traditional models. These results reconfirm the 1131 superiority, efficiency and decoupling capability of the proposed model for optimal 1132 allocation cases involving complex water resources system complex water systems with 1133 multiple stakeholders, multiple sources, multiple decision-makers and embodied reused 1134 systems. By embedding the targets into hierarchical layers, the excessive abandonment 1135 of some promising alternatives is avoided, and optimal allocation is progressively 1136 achieved. In general, the hierarchical structure of the PTSOA model can simulate 1137 complicated systems with multiple complex objectives and constraints. 1138 In addition, the five six MOEAs were used to solve the equations in the third layer 1139 of the PTSOA model, and the overall targets in the first layer were determined based 1140 on these solutions. The necessary parameters and hyper-parameters were consistent

1141 with those used in the third layer of the PTSOA model. Additionally, the benefits in the

1142	current case with no optimization calculated based on the actual water supply are given
1143	for comparison. The current situation was categorized as a normal scenario, and other
1144	models were established with the same conditions to facilitate further comparison and
1145	analysis. There were distinct decision alternatives generated by each model, and the
1146	relevant results are listed based on their value ranges. As shown in Table 4, although
1147	NSGA-II and ε -MOEA yield slightly higher F2 values than PTSOA and F3 generated
1148	by IBEA (4.8×10^{8} 7.2×10 ⁸ yuan <u>Chinese Yuan</u>) is higher than obtained with PTSOA,
1149	PTSOA performs better than other models in most cases. The PTSOA model is shown
1150	to be the best model for obtaining comprehensive benefits for the sub-regions in Yiwu
1151	in the normal scenario, demonstrating that the PTSOA model offers advantages
1152	including identifying the best alternatives and achieving greater sub-regional benefits
1153	than the other models. The proposed model yields $a-1.76 \times 10^9$ - -15.67×10^9 yuanChinese
1154	<u>Yuan</u> total comprehensive benefit improvement and can save approximately 3.2×10^7 -
1155	\sim 4.7×10 ⁷ (m ³) of conventional water compared to the current values. It is also evident
1156	that the proposed model yields the highest TSI values, reflecting the improvement
1157	achieved by considering the synergy of the system. In terms of the targets in the first
1158	layer, except MOEA/D, other traditional models fail to retain enough water (water
1159	requirements for living under extreme drought conditions of the next configuration
1160	period) in the reservoirs to meet future basic needs. For MOEA/D, although it generates
1161	a slightly higher total water supply benefit, with a value of 2.81×10^8 3.12×10 ⁸ , the
1162	total water supply shortage and the total amount of reserved water in the reservoirs are

1163	worse than the amounts obtained with the proposed model. Borg MOEA, as an efficient
1164	and robust many-objective optimization tool,-It -is characterized by its use of auto-
1165	adaptive multi-operator search and other adaptive features (Reed et al., 2013). The TSI
1166	value of Borg MOEA is lower than PTSOA. Therefore, in the TSI dimension, it-iss
1167	performance is slightly worse than the PTSOA model. However, it is noticed that the
1168	Borg MOEA algorithm could save around one-fifth of the computing time of the model
1169	(around 7h). In the future, it would be interesting to figure out how to couplinge the
1170	Borg MOEA algorithm with our PTSOA model in a more efficient and synergetic way.
1171	In this study, our main focus is to find the most synergetic solution through optimization
1172	in a complex system. Thus, PTSOA has accomplished superior performance level in
1173	this respect. PTSOA-It trades some economic benefits for enhanced water supply
1174	reliability and sustainable development, resulting in a decrease in the water supply from
1175	conventional water plants.

1176 However, tThe consideration of reclaimed water in the proposed model effectively 1177 reduces the use of traditional water and improves the quality of the water environment 1178 by reducing sewage discharge, and other benefits are also achieved (such as meeting 1179 the quality standards for river water and guaranteeing that the ecological water demand 1180 of inland rivers is met). The results obtained by the PTSOA model may help guide both 1181 the government and general public. Our proposed model is superior to traditional 1182 models. It can not only optimize water resource utilization and secure water supplies but also enhance the synergy and environmental quality of water systems. Considering 1183

synergy across various time scales, the proposed model ensures the synergistic allocation of water resources at yearly, monthly and daily scales while securing both present and future water supplies.

1187**Table 4** Comparison of the comprehensive benefits in of the five regions objectives1188(F1, F2, F3, F4, and F5) and the *TSI* values in the current situation and obtained using

NSGA-II, SPEA-II, ɛ-MOEA, IBEA, MOEA/D, Borg MOEA and PTSOA in the

1189

normal scenario

	Composison	Comprehensive benefits (109 yuanChinese Yuan)					TCI
	Comparison	F1	F2	F3	F4	F5	151
	NSGA-II	2.72~2.86	0.91~1.03	2.57~2.60	3.21~3.37	7.38~9.95	-3.13~-2.82
	SPEA-II	2.84~2.97	0.93~0.99	2.58~3.15	3.02~3.68	8.22~9.99	-2.39~-2.46
	ε-MOEA	2.47~2.33	0.85~1.12	2.21~2.32	3.05~3.18	9.23~9.91	-3.41~-3.06
	IBEA	2.57~2.88	0.87~0.92	3.05~3.11	3.20~3.32	5.27~8.28	-3.28~-3.11
	MOEA/D	2.55~2.90	0.99~1.02	3.15~3.20	3.34~3.36	9.82~10.11	-2.37~-1.54
	Borg MOEA	<u>2.95~3.56</u>	<u>0.80~0.98</u>	<u>1.19~2.23</u>	<u>3.11~3.82</u>	12.88~13.90	<u>-2.51~-1.67</u>
	Current situation	2.05	0.83	2.49	3.11	9.87	-3.20
	PTSOA	2.63~3.03	0.95~0.99	2.39~2.67	3.84~4.11	10.30~11.22	-1.66~-0.89

1191 **5. Conclusions**

Applying optimal water allocation models to simultaneously enable economic benefits, water preferences and environmental demands at different decision levels, time scales and regions is a challenge. In this study, a new process-based three-layer synergistic optimal allocation model (PTSOA) is was developed and applied to a real and complex water allocation system to figure it out. The The objective functions model were was divided into three layers to coordinate conflicts of interest among decision makers at different levels and time scales. Furthermore, the allocation of reclaimed water was
embedded in the proposed model for synergistic optimal allocation of both conventional
and unconventional water. A synergistic index based on network analysis was-<u>put</u>
<u>forward introduced</u> to reduce competitions among different stakeholders and facilitate
the positive effects of stakeholder interactions. A hierarchical optimal algorithm was
designed to solve the <u>PTSAO-PTSOA</u> model.

1204 The proposed model was applied to a representative-typical city in Southeast China 1205 with scarce water resources and a developed industry. Achieving the optimal allocation 1206 of water resources in this kind of water-scarhighly developed area-city offers a valuable 1207 reference for other counties in China. The key findings of this study are as follows. Key 1208 advantages of PTSOAfindings can be concluded from these results, as follows. Firstly, 1209 the results demonstrated that the PTSOA model achieved synergistic allocation among 1210 hierarchical decision-makers across various time scales and in different regions, 1211 yielding the highest TSI (-1.66 to -0.89) among the contrast models-evaluated. Secondly, 1212 with a synergistic approach, a reasonable amount of conventional water is retained for 1213 future use in cases with potentially high risk, with volumes of 3.95×10^7 m³, 3.12×10^7 1214 m³, and 2.43×10^7 m³ retained in normal, dry and extremely dry scenarios, respectively. 1215 Moreover, 7.35×10^7 m³, 7.56×10^7 m³, and 7.37×10^7 m³ of conventional water is can be 1216 saved in the three scenarios. Thirdly, considering both reclaimed water and 1217 conventional water in the optimization process efficiently improves the quality of municipal water, and more than 1272.21 t/year and 48.81 t/year of COD and ammonia 1218

1219 nitrogen emissions are mitigated compared to those in the current situation. Lastly, 1220 Distinct distinct from previous models, the proposed optimal model was implemented 1221 with the consideration of spatial dimensions, which are important but often neglected. 1222 The results show that spatial allocation yields an improvement of 4-~95% for the 1223 comprehensive benefits in different sub-regions compared to the benefits achieved with 1224 traditional models, and the total comprehensive benefit increases by 1.76×109- $\sim 15.67 \times 10^9$ yuanChinese Yuan compared to that in the current situation. The synergy 1225 1226 index established based on network analysis is used to alleviate the competition among 1227 regions and facilitate water supply improvements.

These results and conclusions provide valuable references for the evaluations of other complicated water allocation systems. The optimal allocation scheme is can be determined for a complex water resources system complex water system upon consideration of stakeholder synergy and various hierarchical decision levels, time scales and regions. More in-depth studies of synergistic optimal water allocation are needed in the future.

1234

1235 *Data availability.* The data used to support the findings of this study are available from
1236 the corresponding author upon request.

1237

1238 Author contributions. JL and YPX designed all the experiments. JL and WZ collected

1239 and preprocessed the data. JL and WZ conducted all the experiments and analysed the
1240	results. JL wrote the first draft of the manuscript with contributions from SW and SC.
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