

# 1 Quantifying the trade-offs in re-operating dams for the environment 2 in the Lower Volta River

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## 9 **Abstract**

10 The construction of the Akosombo and Kpong dams in the Lower Volta River Basin in Ghana changed  
11 the downstream riverine ecosystem and affected the lives of downstream communities, particularly  
12 those who lost their traditional livelihoods. In contrast to the costs borne by those in the vicinity of  
13 the river, Ghana as a whole, has enjoyed vast economic benefits from the affordable hydropower,  
14 irrigation schemes and lake tourism that developed after construction of the dams. Herein lies the  
15 challenge; there exists a trade-off between water for river ecosystems and related services on the one  
16 hand, and anthropogenic water demands such as hydropower or irrigation on the other. In this study,  
17 an Evolutionary Multi-Objective Direct Policy Search (EMODPS) is used to explore the multi-sectoral  
18 trade-offs that exist in the Lower Volta River Basin. Three environmental flows, previously determined  
19 for the Lower Volta are incorporated separately as environmental objectives. The results highlight the  
20 dominance of hydropower production in the Lower Volta but show that there is room for providing  
21 environmental flows under current climatic and water use conditions if firm energy requirement from  
22 Akosombo Dam reduces by 12% to 38% depending on the environmental flow regime that is  
23 implemented. There is uncertainty in climate change effects on runoff in this region, however multiple  
24 scenarios are investigated. It is found that climate change leading to increased annual inflows to the  
25 Akosombo Dam reduces the trade-off between hydropower and the environment while climate  
26 change resulting in lower inflows provides the opportunity to strategically provide dry season  
27 environmental flows, that is, reduce flows sufficiently to meet low flow requirements for key  
28 ecosystem services such as the clam fishery. This study not only highlights the challenges in balancing  
29 anthropogenic water demands and environmental considerations in managing existing dams, but also  
30 identifies opportunities for compromise in the Lower Volta River.

31 **Keywords:** Environmental flows, Multi-objective evolutionary optimization, Direct policy search, Volta  
32 River, Akosombo Dam

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## 35 1 Background

36 Freshwater resources are under increasing pressure worldwide (WWF, 2018; He et al., 2019). As global  
37 population and standards of living have gone up, the capacity of many river basins to meet social,  
38 economic and environmental water demands has declined (Best, 2019; Fitzhugh and Richter, 2004;  
39 Postel and Richter, 2003). In the mid-20th century, many dams were built with ambitious goals for  
40 hydropower generation, flood control and irrigation among others and dam construction is seeing a  
41 resurgence in recent years (Grill et al., 2015; Best, 2019). This phenomenon is occurring, despite the  
42 fact that even the economic justification for many existing dams is being called into question (Ansar  
43 et al., 2014; Flyvbjerg and Bester, 2021), the life cycle emissions of some dams is above the median  
44 emissions for fossil fuel plants (Schlömer et al., 2014; Almeida et al., 2019) and the negative social and  
45 environmental impacts of dams on riparian ecosystems and communities have been established for  
46 some time (WCD, 2000; Stone, 2011; Duflo and Pande, 2007; Richter et al., 2010). Proponents of dam  
47 construction argue that in developing regions, particularly in Africa, the large energy deficit (Hafner et  
48 al., 2018) coupled with high inter-annual rainfall variability and the fact that 75% of the population  
49 live in semi-arid or arid regions (Vörösmarty et al., 2005; Smith, 2004), makes multipurpose dams  
50 important infrastructures for energy and food security. Evidently, tools are required for investigating  
51 operation policies for managing and maximising the benefits of dams and the water resources they  
52 control.

53 Multi-objective evolutionary algorithms (MOEAs) are one such tool for assessing the trade-offs  
54 between water users in a river basin. MOEAs use stochastic search tools to simultaneously find the  
55 Pareto approximate set across multiple objectives (Reed et al., 2013; Matrosov et al., 2015; Hurford  
56 et al., 2020; Zatarain Salazar et al., 2016; Kiptala et al., 2018). The Pareto approximate or non-  
57 dominated set of solutions are the suite of solutions for which increasing the water allocation to one  
58 user leads to a reduction in the benefit to others. The advantage of MOEAs is that they do not require  
59 pre-specifying preferences across objectives, thereby supporting unbiased *a posteriori* decision  
60 making (Reed et al., 2013; Hurford et al., 2014). Furthermore, MOEAs allow for heterogeneous and  
61 non-linear problem formulations with incommensurable objectives and different risk attitudes across  
62 objectives. Accordingly, non-market objectives can be evaluated alongside conventional economic  
63 objectives. This is particularly useful for including environmental flows (e-flows) and ecosystem  
64 services for which monetary valuation is often difficult and contested (Bingham et al., 1995; Costanza  
65 et al., 1997, 2014; Luisetti et al., 2011). The capability of MOEAs to find Pareto approximate strategies  
66 for a suite of water systems applications has been thoroughly assessed by Reed et al. (2013), and for  
67 multi-purpose reservoir operations by Zatarain Salazar et al. (2016). In this paper, an Evolutionary

68 Multi-Objective Direct Policy Search (EMODPS) framework is applied to map the states of a system, in  
69 this case, reservoir levels and time of the year, to actions, the release of water for different water uses  
70 (Giuliani et al., 2016; Zatarain Salazar et al., 2017). This approach has been applied to find Pareto  
71 approximate operating policies for multi-objective, multi-reservoir systems (Quinn et al., 2017; Wild  
72 et al., 2019). The motivation to use EMODPS was informed by the fact that for the selected case study,  
73 multi-objective reservoir operating policies had to be found under uncertainty. Traditional approaches  
74 for optimal control, such as stochastic dynamic programming, do not permit finding the Pareto  
75 approximate policies across multiple objectives in a single run, requiring instead that the Pareto front  
76 is constructed by testing different weights for each of the system's objectives. Such a method  
77 increases the computational burden and yields a sparse Pareto front thereby potentially missing  
78 regions of suitable policies. The use of EMODPS overcomes this challenge by generating the trade-offs  
79 across all the system's objectives simultaneously in a single algorithmic run, creating a diverse and  
80 more accurate Pareto front (Giuliani et al., 2016). This motivates the use of direct policy search, in  
81 which radial basis functions are used to find a flexible shape to map storage levels and time to release  
82 decisions for multiple objectives.

83 In many of the studies where MOEAs have been applied, the e-flow objective in the simulation  
84 component of the model either meets a minimum flow release (Zatarain Salazar et al., 2017; Gonzalez  
85 et al., 2021; Kiptala et al., 2018; Hurford et al., 2020) or minimizes the deviation of flow from the  
86 natural, unregulated flow regime (Hurford and Harou, 2014). The former objective, minimum flow  
87 releases, fails to thoroughly capture the essence of e-flows which are the "quantity, timing, and quality  
88 of freshwater flows and levels required to sustain aquatic ecosystems" (Brisbane Declaration, 2018).  
89 The latter, the objective of returning fully to the natural flow regime, is an unlikely objective in many  
90 highly modified and utilized river basins (Acreman et al., 2014; Horne et al., 2017b). In this study, a  
91 multi-objective analysis of the trade-offs between key water users and the environment in the heavily  
92 modified Lower Volta River Basin in Ghana is carried out. The environmental objectives include  
93 designer e-flows (Acreman et al., 2014) developed for different ecosystem services in the basin. In  
94 contrast to the aim of restoring a river to a near natural state, designer e-flows define and construct  
95 parts of the flow hydrograph of a river to meet certain desired ecological and social outcomes  
96 (Acreman et al., 2014; Horne et al., 2017a). Three e-flows, defined for the Lower Volta River in previous  
97 studies, are investigated and compared: one to support the Volta clam (*Galatea paradoxa*) (Owusu et  
98 al., 2022b) and the other two to support multiple ecosystem services including fisheries, aquatic weed  
99 control, flood recession agriculture and sediment transport (Mul et al., 2017). In addition, future  
100 climatic scenarios are investigated. This study highlights the challenges faced by dam operators in  
101 balancing environmental and anthropogenic water demands for existing dams in heavily modified and

102 utilized river basins, and simultaneously investigates the room for compromise in the case of the  
103 Lower Volta River under uncertainty. The main contribution of the paper is twofold: First, it explores  
104 the room for compromise in the Lower Volta by the quantifying the Pareto approximate trade-offs  
105 when e-flows previously prescribed for the basin are implemented. Secondly this paper is a new  
106 application of the EMODPS under high uncertainty where only the system goals and direction of  
107 preference are specified in the multi-objective decision problem.

108 In the following section, a description of the Lower Volta River Basin is given, followed by the methods  
109 section in which (i) the simulation model for the lower Volta is described, (ii) the multi-objective  
110 evolutionary optimization set up is explained, (iii) the objective functions are formulated, and (iv)  
111 relevant climate-induced effects on discharge are specified. Next is the results section where we  
112 present the trade-off analysis between e-flows and other water uses in the Lower Volta for the current  
113 baseline scenario and possible future scenarios. We conclude with a discussion on the implications of  
114 implementing e-flows in the Lower Volta and draw lessons for the application of EMODPS in other  
115 heavily modified basins under uncertain future conditions.

## 116 2 Lower Volta River Basin

117 The Lower Volta River, located in Ghana is one of four sub-basins in the Volta River Basin in West Africa  
118 (Figure 1). It is located furthest downstream, flowing into the Gulf of Guinea and covering an area of  
119 66700 km<sup>2</sup>, approximately 16% of the Volta Basin. The most important hydraulic infrastructure in the  
120 Lower Volta is the Akosombo Dam, which was built in 1965 for hydropower production with an  
121 installed capacity of 1038 MW (1,020MW Akosombo Hydro Electric Power Plant, 2021). In 1981, a  
122 smaller 160 MW run-of-the-river dam, the Kpong Dam, also began operation downstream. The lake  
123 created by the Akosombo Dam is the largest man-made lake by surface area at about 8500 km<sup>2</sup>. It has  
124 an average depth of 18.8 m and holds approximately 148 km<sup>3</sup> of water at maximum capacity  
125 (1,020MW Akosombo Hydro Electric Power Plant, 2021).



126

127 *Figure 1: The Akosombo and Kpong Dams located in the Lower Volta River Basin, which discharges into the Gulf of Guinea*

128 Construction of the Akosombo Dam led to the resettlement of over 80,000 people (Darko and Tsikata,  
 129 2019; Alhassan, 2009) and also changed the dynamic flow regime downstream from one with average  
 130 low and high flows of approximately 36 m<sup>3</sup>/s in March and 5,100 m<sup>3</sup>/s in September-October  
 131 respectively, to a steady flow of about 1,000 m<sup>3</sup>/s per month all year round with no account taken of  
 132 seasonality (Ntiamoa-Baidu et al., 2017). Consequently, the riverine ecosystem changed and so did  
 133 the lives of downstream communities. Creek fishing, floodplain agriculture and the clam fishing  
 134 industries, which together made up three-quarters of total real income of the Lower Volta riparian  
 135 population in 1954, collapsed (Moxon, 1969; De-Graft Johnson, 1999; Tsikata, 2008; Lawson, 1972).  
 136 In addition, invasive aquatic weeds proliferated, providing habitat for disease vectors including  
 137 mosquitoes and snails and thereby increasing the prevalence of waterborne and water related  
 138 diseases such as malaria and schistosomiasis (Akpabey et al., 2017; Gyau-Boakye, 2001). Other  
 139 environmental costs include changes to the sediment load leading to erosion along the coastline of  
 140 Ghana, as well as Togo and Benin (Bollen et al., 2011; Roest, 2018; Appeaning Addo et al., 2020), as  
 141 well as a reduction in salt water intrusion (Beadle, 1974; People and Rogoyska, 1969; Nyekodzi et al.,  
 142 2018). Among the population in the Lower Volta, perceptions of the Akosombo Dam and the run-of-  
 143 the-river Kpong Dam downstream, are still overwhelmingly negative: in a survey of over 400 citizens

144 older than 50 years in 2016, approximately 92% considered their socio-economic conditions to be  
145 better under pre-dam natural flows (Baah-Boateng et al., 2017).

146 The costs borne by the river ecosystem and the communities in the vicinity of the dams in the Lower  
147 Volta is in strong contrast to the vast economic benefits that Ghana as a whole has enjoyed from the  
148 relatively affordable hydropower, irrigation schemes and tourism that developed after construction  
149 of the dams (Eshun and Amoako-Tuffour, 2016; Alhassan, 2009). After construction, the Akosombo  
150 Dam provided over 70% of Ghana's electricity and is credited with powering Ghana's industrialization  
151 and making it one of the most developed countries in West Africa (Alhassan, 2009). The dam currently  
152 makes up about 20% of the installed electricity generating capacity in Ghana (Dye, 2020) and is  
153 operated by the Volta River Authority (VRA).

154 The local-national mismatch in benefits deriving from the operation of the Akosombo dam has been  
155 investigated in previous studies, notably in 2016 by Ntiamao-Baidu et al. (2017). The 2016 study  
156 adopted a simulation-based approach using the Water Evaluation and Planning (WEAP) tool to  
157 compare the current flow regime with the natural flow regime and two other scenarios for re-  
158 operating the Akosombo dam (Annor et al., 2017; Mul et al., 2017). The alternative dam operation  
159 scenarios were found to reduce power generation by 45% to 74%, and were deemed undesirable  
160 (Annor et al., 2017). This was at a time when Akosombo and Kpong dams made up about 40% of  
161 installed capacity and Ghana was experiencing power rationing due to low water levels in the Volta  
162 River and a shortfall in gas supply to other power plants (Dye, 2020). The present energy context of  
163 Ghana is very different and is characterized by an "overabundance" of electricity generation potential  
164 - almost twice the peak load demand and therefore a reduced dependence on power generation at  
165 Akosombo and Kpong dams. (Dye, 2020; Kumi, 2017). While installed capacity does not directly  
166 translate into power delivery, it is worth re-examining the trade-offs between water users in the Lower  
167 Volta under this changed situation given that it is as a result of 'take-or-pay' power purchase  
168 agreements with private power companies whereby 90% of the power made available by these  
169 companies has to be paid for irrespective of whether it is used or not (Dye, 2020). In 2018, the cost of  
170 this extra capacity was approximately 5% of the country's gross domestic product (GDP) (The World  
171 Bank, 2018; Dye, 2020).

### 172 3 Methods

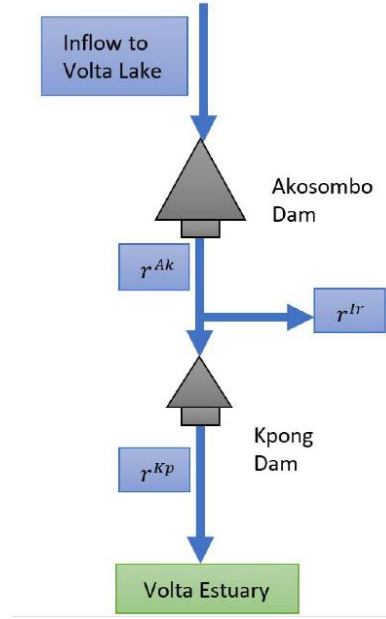
173 The simulation model for this study was developed using the mass balance of inflows, net evaporation  
174 rates and releases from the Akosombo Dam from 1981 to 2012 (data obtained from VRA). In addition  
175 to net evaporation and inflows, additional input data to the model consisted of downstream water  
176 levels and physical characteristics of the dam such as the storage-area-level relationships. The model

177 was initially set up using the current baseline dam operation policy for hydropower, flood control and  
178 irrigation over a wet (2010), dry (2006) and normal (1985) year. The choice to calibrate for years with  
179 specific conditions as against the full historical time series was a practical one to expedite the  
180 calibration phase of the study. The Nash-Sutcliffe model efficiencies of 0.89, 0.91 and 0.90 respectively  
181 were obtained when the modelled and observed reservoir volumes were compared for each of the  
182 wet, dry and normal years, thus indicating a good model fit (Figure S1-S3, Supplementary material).

183 Radial Basis Functions (RBF) were used to parameterize the control policies for mapping reservoir  
184 levels and the time into daily release decisions (Zatarain Salazar et al., 2016, 2017). RBFs are non-linear  
185 approximating networks that allow the time dependent operating decisions to depend on more than  
186 a single variable and most importantly can accommodate multiple objectives simultaneously (e.g.,  
187 hydropower, irrigation, flood prevention, and e-flows). By providing alternative Pareto approximate  
188 solutions, it is possible to visualise trade-offs between the objectives and thereby inform policy  
189 decisions. In a comparative analysis, (Giuliani et al., 2016) found that RBF solutions performed better  
190 in terms of convergence, consistency and diversity of solutions as compared to another widely used  
191 universal approximator, Artificial Neural Networks (ANN). Indeed, using such a non-linear  
192 approximating network avoids “restricting the search for the optimal policy to a subspace of the  
193 decision space that does not include the optimal solution” (Giuliani et al. 2018).

### 194 3.1 Multi-objective problem formulation for the Lower Volta system

195 When the storage volume in the reservoir is known at time  $t$ , and the decision time step is set to 1,  
196 the downstream releases can be determined for the time interval  $[t, t + 1]$ . The release from  
197 Akosombo Dam ( $r_{t+1}^{Ak}$ ) is determined by the irrigation, hydropower, environmental and flood control  
198 demands ( $r_{t+1}^{IHEF}$ ) (Eq. ( 1 )) (Figure 2). Kpong Dam is operated as a run-of-the-river hydropower system  
199 by VRA after water is diverted for irrigation (Eq. ( 2 )).



200

201 *Figure 2: Topology of reservoir system in the Lower Volta.  $r^{Ak}$ ,  $r^{Kp}$  and  $r^{Ir}$  are the flow releases from Akosombo, Kpong and*  
 202 *for irrigation respectively.*

203 The release from Kpong Dam ( $r_{t+1}^{Kp}$ ) is therefore calculated as the difference between the release from  
 204 Akosombo Dam and irrigation ( $r_{t+1}^{Ir}$ ), and represents the downstream releases for hydropower, e-  
 205 flows and floods ( $r_{t+1}^{HEF}$ ).

$$r_{i+1}^{Ak} = r_{i+1}^{IHEF} \quad (1)$$

$$r_{i+1}^{Kp} = r_{i+1}^{Ak} - r_{i+1}^{Ir} = (r_{i+1}^{HEF}) \quad (2)$$

206

207 The operating policy is commonly parametrised as a function of the reservoir storage volume at a  
 208 particular time. The parameterized operating policy  $f$  is then defined as a mapping between the  
 209 decisions  $\mathbf{u}_t$  and the policy inputs  $z_t$  comprising the time  $t$  and system state, or storage volume,  $x_t$ ,  
 210 Eq. ( 5 ), namely:

$$\mathbf{u}_t = f(z_t) \quad (3)$$

211 The  $k$ th decision variable in the vector  $\mathbf{u}_t$  (with  $k = 1, \dots, n$ ) is therefore defined as a weighted sum  
 212 of radial basis functions, as specified in Eq. ( 4 ):

$$u_t^k = \sum_{i=1}^N w_{i,k} \phi_i(z_t) \quad (4)$$

213 where  $N$  is the number of radial basis functions  $\phi_i(\cdot)$ , and  $w_{i,k}$  is the non-negative weight of the  $i$ th  
 214 radial basis function, and the  $N$  weights sum to unity. The  $i$ th Gaussian radial basis function is then  
 215 given by Eq. ( 5 ):



$$\varphi_i(z_t) = \exp \left[ - \sum_{j=1}^M \left( \frac{(z_t - c_{j,i})^2}{b_{j,i}^2} \right) \right] \quad (5)$$

216 Where  $j = 1, \dots, M$  is the number of input variables,  $z_t$  is the policy input (e.g. time  $t$ , reservoir level  
 217  $x_t$ ) and  $c_{j,i}$  and  $b_{j,i}$  are the centres and radii respectively of the  $i$ th Gaussian radial basis function for  
 218 the  $j$ th input variable. The parameter vector  $\theta$  is defined as  $\theta = (c_{i,j}, b_{i,j}, w_{i,k})$  with  $i = 1, \dots, N$ ;  $j =$   
 219  $1, \dots, M$ ;  $k = 1, \dots, n$ , where the centre and radius are normalized with  $c_{i,j} \in [-1, 1]$  and  $b_{i,j} \in (0, 1]$ .  
 220 The policy parameters  $\theta$  are determined by simulating the system over the time horizon  $H$  under the  
 221 policy  $f = \{f(t, x_t, \theta): t = 0, \dots, H - 1\}$ . In this way the inputs to the RBF policy (time index and  
 222 reservoir storage volume) are mapped to the outputs (release decisions). The policy parameters are  
 223 evaluated by solving the multi-objective problem function,  $f$ , specified in Eq. ( 6 ) in the objective space  
 224 using an informed search algorithm  $\theta^*$ . The objective functions  $J$  are the operating objectives of the  
 225 reservoir as defined in Eq.s ( 7 ) to ( 12 ) with any maximization objectives multiplied by -1 to  
 226 reformulate all the objectives as minimizations.

$$f\theta^* = \arg \min_{\theta} J(\theta) \quad (6)$$

227 The number of RBFs used in this study was four (i.e.  $n = 4$ ). Thus, the total number of parameters ( $\theta$ )  
 228 for the control policy in this study is 24. A daily decision timestep, representative of “real operations”  
 229 of the two dams by VRA was used (Annor et al., 2017). The simulation time horizon,  $H$ , of 29 years  
 230 using historical data starting from January 1984, was constrained by the low availability of data. This  
 231 period however encompasses key dry and wet periods. In 1997-2000 and 2006-2007, there was power  
 232 rationing in Ghana as water levels in the Akosombo reservoir fell to 73.01 m (July 1999) and 72.16 m  
 233 (August 2006) respectively; both lower than the minimum operating level of 73.15 m (VRA, 2021). On  
 234 the other hand, in 1991 and 2010, extremely high inflows caused the reservoir water level to rise to  
 235 83.90 m and 84.42 m, respectively, close to the maximum operating level of 84.73 m, and necessitated  
 236 the opening of the spillways. The 2010 reservoir level remains the highest point ever recorded at  
 237 Akosombo. The four objectives considered in this study are described in more detail below,

- 238 1. Annual hydropower: Maximization of the annual hydropower generated at Akosombo and Kpong  
 239 dams as defined in Eq. ( 7 ). While the annual firm energy requirement from Akosombo Dam is  
 240 4415 GWh/year, the amount of electricity generated has typically exceeded this target in the past  
 241 due to high national dependence on power generation from this dam. As such, operations at  
 242 Akosombo has generally been to maximise power considering the reservoir volume and inflows to  
 243 the dam (Annor et al., 2017). There is no firm energy target at Kpong which is a run-of the river  
 244 dam and generates power with releases from Akosombo after the diversion for irrigation.

$$J_H = \sum_{t=1}^I HP_t \quad (7)$$

245 where energy production ( $HP_t$ ) in GW is given by Eq. ( 8 ):

$$HP_t = \eta g \rho_w h_t q_t^{turb} \cdot 10^{-9} \quad (8)$$

246 where  $t \in I$  are the days in a year,  $\eta$  is the turbine efficiency (dimensionless),  $g$  is acceleration due  
 247 to gravity ( $9.81 \text{ m/s}^2$ ),  $\rho_w$  is water density ( $1000 \text{ kg/m}^3$ ),  $h_t$  is net hydraulic head (m) and  $q_t^{turb}$  is  
 248 flow through the turbines ( $\text{m}^3/\text{s}$ ). The hydropower objective at Akosombo is subject to constraints  
 249 on the minimum daily firm energy requirement of 6 GWh/day for system stability and the  
 250 maximum possible power production due to turbine capacity ( $1,603 \text{ m}^3/\text{s}$ ) at the maximum safe  
 251 operating level of 84.12 m.

252 2. Irrigation: Maximization of the volumetric reliability of water supply to meet irrigation demand as  
 253 described in Eq. ( 9 ).

$$J_I = \frac{1}{H} \sum_{t=1}^H \frac{q_t}{V_t} \quad (9)$$

254 subject to the constraint in Eq. ( 10 ):

$$0 \leq q_t \leq Q_t \quad (10)$$

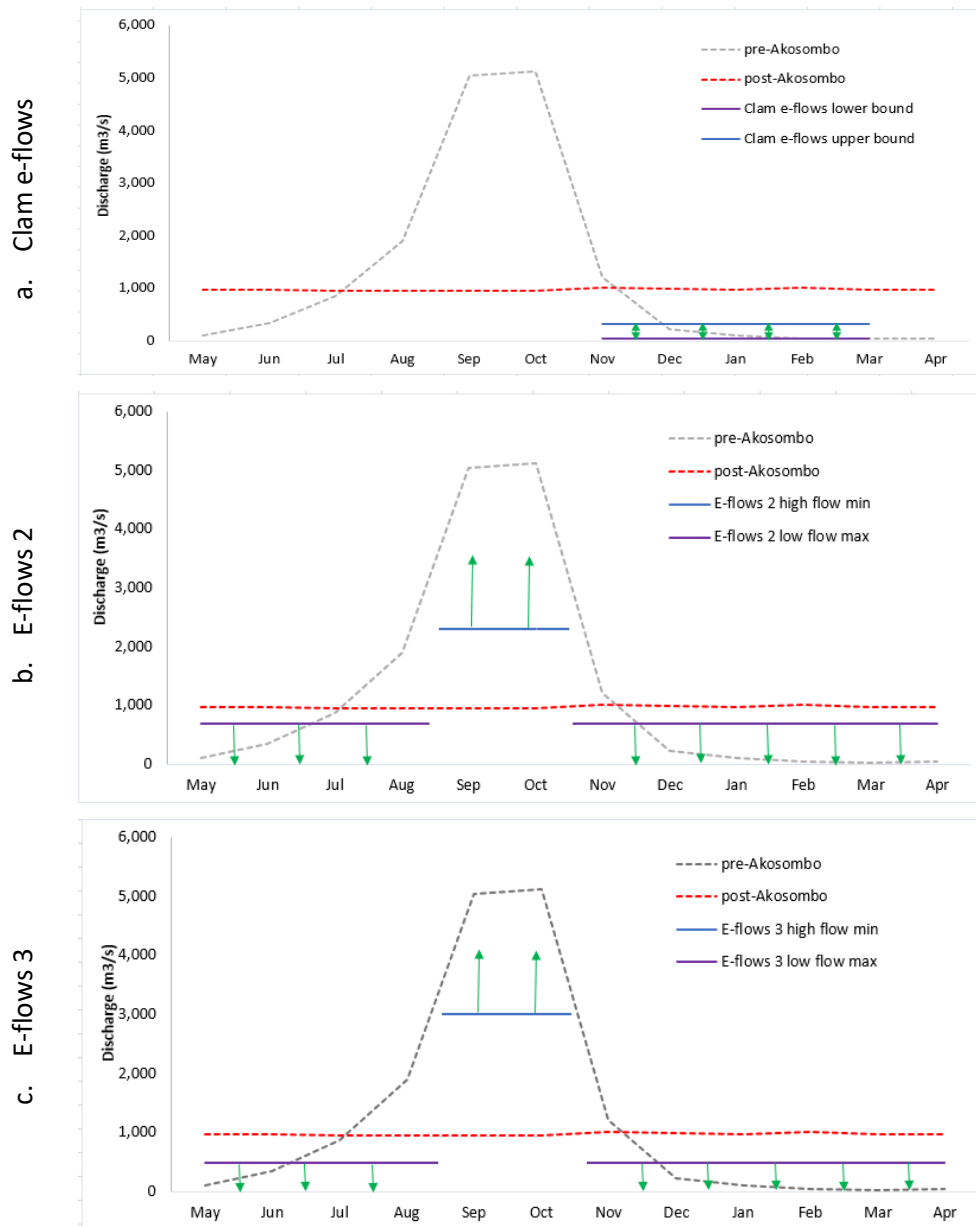
255 where  $V_t$ ,  $q_t$  and  $Q_t$  are the irrigation demand, the diverted water and the flow at diversion point  
 256 at time  $t$ , a day within the simulation horizon  $H$ . The current irrigation demand is  $10 \text{ m}^3/\text{s}$  but there  
 257 have been plans since 2009 for this to be increased to  $38 \text{ m}^3/\text{s}$  for the Accra Plain Irrigation Project  
 258 and the expansion of the Kpong left bank irrigation project (GIDA, 2009). These projects are yet to  
 259 be fully realized however, and in this study the anticipated irrigation demand of  $38 \text{ m}^3/\text{s}$  is used as  
 260 the baseline value.

261 3. Flood control: Minimization of flood occurrences defined by the average number of days where  
 262 downstream flow releases from Kpong,  $r_{t+1}^{Kp}$ , exceed  $2300 \text{ m}^3/\text{s}$ , the bank full capacity of the river  
 263 ( $Q_F$ ) (Eq. ( 11 )). Opening of the spillways of the Akosombo and Kpong dams is quite rare and has  
 264 occurred only twice, in 1991 and 2010. Consequently the riparian communities are ill-prepared for  
 265 flood releases and incur high losses whenever floods are released (Ayivor and Ofori, 2017).

$$J_F = \frac{1}{H} \sum_i^H \left( \frac{\max ( r_{i+1}^{Kp} - Q_F, 0 )}{Q_F} \right) \quad (11)$$

266 4. E-flows: The trade-off between the three objectives defined above and three different e-flows  
 267 (Figure 3) are investigated in separate runs in this study. As such, three different configurations of  
 268 the trade-off problem are investigated.

- 269 a. Clam e-flows (Figure 3a): This e-flow was designed for the Lower Volta River using the Volta  
270 clam, a stenotopic, freshwater bivalve, as an indicator species (Owusu et al., 2022b). The  
271 recommended flow is a low flow range of 50 - 330 m<sup>3</sup>/s from November to March to support  
272 the clam's veliger larvae stage, a key life stage in its lifecycle. An 80% reliability of this flow  
273 occurring in the stipulated months is an acceptable compromise for the survival of the clam  
274 veliger larvae (Owusu et al., 2022b). While only a low flow is prescribed for five months in this  
275 e-flow recommendation, this necessarily implies that flow releases at other times of the year  
276 will be higher, although the magnitude, duration and timing of this such flow is not defined  
277 and thus does not form a constraint for clam e-flows. The historical minimum for high flows in  
278 September and October under pre-dam flows was 1052 m<sup>3</sup>/s.
- 279 b. Natural flow dynamics considering bank full flows (e-flows 2) (Figure 3b): This e-flow reinstates  
280 natural flow dynamics in the Lower Volta to support multiple ecosystem services including  
281 fisheries, aquatic weed control, flood recession agriculture and sediment transport (Mul et al.,  
282 2017). The minimum discharge for the high flow period in September to October is 2,300 m<sup>3</sup>/s,  
283 (which is the bank full flow rate) to ensure that river overtopping and thus some minimum  
284 flooding of pre-dam floodplains occurs. The maximum dry season discharge for the rest of the  
285 year is 700 m<sup>3</sup>/s.
- 286 c. Natural flow dynamics considering future dry season irrigation (e-flow 3) (Figure 3c): This e-  
287 flow also re-instates natural flow dynamics of the Lower Volta while providing water for future  
288 dry season irrigation demands (Mul et al., 2017). The minimum high flow in September and  
289 October is 3000 m<sup>3</sup>/s, which inundates an area of approximately 156 km<sup>2</sup>, to support creek  
290 fishing and flood recession agriculture. The maximum dry season flow rate is 500 m<sup>3</sup>/s.



291  
 292 *Figure 3: E-flow configurations considered in this study with pre-dam (natural) and current post dam flow regime in the Lower*  
 293 *Volta provided for comparison (using monthly average flow data from Volta River Authority, Ghana) for a hydrological year*  
 294 *which starts in May in the Lower Volta. For clam e-flows the green arrows show the range of low flows recommended from*  
 295 *November to March. For e-flows 2 and 3, the prescribed flow for September and October are a minimum threshold while for*  
 296 *the other ten months, this flow is a maximum threshold. The green arrows begin at these thresholds and point in the direction*  
 297 *of where the flows should be per the e-flow recommendation.*

298 The three alternative e-flow objectives (clam e-flows, e-flows 2 and e-flows 3) were modelled as a  
 299 maximization of the reliability of the recommended flow rates occurring (Eq. ( 12 )):

$$J_E = 1 - \frac{n_E}{n_T} \quad (12)$$

300 where  $n_E$  is number of days when downstream flow falls outside the e-flow range and  $n_T$  is the  
 301 total number of days when the recommended e-flows are required.

## 302 3.2 Future climate scenarios

303 In addition to a baseline scenario optimizing the trade-offs between hydropower, irrigation, flooding  
304 and e-flows under the present climate, future scenarios representing different climate futures were  
305 analysed. The recent Sixth Assessment Report of Working Group I of the Intergovernmental Panel on  
306 Climate Change (IPCC) projects that mean temperature in West Africa will increase 1.5 °C by 2040 and  
307 projects with high confidence that monsoon rainfall over West Africa will increase in the mid (2041-  
308 2060) to long term (2081-2100) but have a delayed start (IPCC, 2021). Future projections made with  
309 medium confidence relate to the delayed retreat of the monsoon rains and an increase in the  
310 frequency and duration of droughts in the latter part of the 21<sup>st</sup> century (IPCC, 2021).

311 These latest climate projections draw a mixed picture of future climate in the West African sub-region.  
312 A further review of the anticipated impacts of climate change specifically on runoff in the Lower Volta  
313 was carried out with the goal of identifying studies that focussed on the entire Volta basin, either as a  
314 whole or all sub-basins. A search was conducted in Scopus using the search string: TITLE-ABS-KEY  
315 (Volta AND climate AND (change OR impact), AND (flow OR discharge OR water) AND (availability OR  
316 resources)). From the 60 papers returned, a review by (Roudier et al., 2014) on climate change impacts  
317 on runoff in West African Rivers provided the first point of reference. From this review by (Roudier et  
318 al., 2014), four studies meeting the search criteria were identified. An additional four papers from the  
319 Scopus search results, not reviewed by (Roudier et al., 2014), were also retained for analysis.

320 Four papers (Kunstmann and Jung, 2005; Aerts et al., 2006; Jung et al., 2012; Abubakari, 2021) found  
321 that annual runoff in the Volta will increase (by 4% to 65%), and of these papers, three (with the  
322 exclusion of Aerts et al. (2006)) also presented monthly trends which generally showed an increase in  
323 wet season flow from June to October and a decrease in dry season flow. The findings on only monthly  
324 trends of (Jin et al., 2018) are also in line with these predictions. (McCartney et al., 2012) and (Sood et  
325 al., 2013), in contrast, find that there will be a decrease in annual runoff (ranging from -13% to -45%)  
326 while (Amisigo et al., 2015) find that the results across the various scenarios are inconsistent. Table S1  
327 in the supplementary materials provides further details on the papers, the climate change scenarios  
328 considered and the models used. It is important to note that using a combination of models, i.e.: global  
329 and regional climatic models and then hydrological models, introduces uncertainty in the findings of  
330 climate projections for runoff (McCartney et al., 2012) and the wide-ranging results from the reviewed  
331 papers show that particularly in the case of the Volta the direction of change in runoff is still unclear.

332 The climate-runoff studies, just as the latest IPCC report, present a mixed picture for the Lower Volta.  
333 Therefore, bearing the high level of uncertainty in mind, five scenarios indicative of the range of  
334 climate-induced changes predicted for the Volta discharge for the mid to long term are investigated

335 (Table 1). These include both increases and decreases in annual runoff as well as seasonal variations  
 336 in runoff into the Lake Volta.

337 *Table 1: Design of future scenarios encompassing climate-induced changes in the Lower Volta discharge*

Scenario	Annual decrease	Annual increase		Seasonality		
	Decrease -45%	Increase +12%	Increase +65%	Dry season decrease (Nov to May) -10%	Wet season increase (Jun to Oct) +10%	Wet season increase (Jun to Oct) +55%
<b>1</b>	<b>x</b>	-	-	-	-	-
<b>2</b>	-	<b>x</b>	-	-	-	-
<b>3</b>	-	-	<b>x</b>	-	-	-
<b>4</b>	-	-	-	<b>x</b>	<b>x</b>	-
<b>5</b>	-	-	-	<b>x</b>	-	<b>x</b>

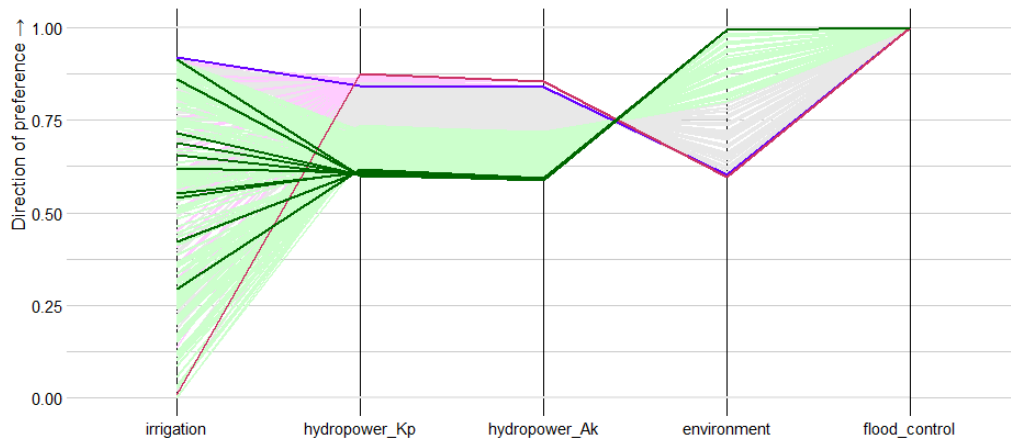
## 338 4 Results

339 The relationship between different water users in the Lower Volta is presented using parallel axis plots  
 340 (Figure 4 and Figure 5). Every line crossing the axes is a Pareto approximate (non-dominated) solution  
 341 and shows the performance of each water user under an alternative dam operation policy for the  
 342 Lower Volta system. The range of values for each water user has been normalized using its maximum  
 343 and minimum values with the best performance featuring at the top of each axis. For irrigation and  
 344 the environmental objectives, the highest value is interpreted as a dam operation policy whereby  
 345 100% of irrigation demand and e-flows are provided, while for the flood control objective, this is  
 346 interpreted as there being no downstream flow releases above the flooding flow threshold of 2300  
 347 m<sup>3</sup>/s over the simulation horizon. For hydropower at Akosombo and Kpong dams, the maximum and  
 348 minimum values used in the normalization encompass the maximum and minimum annual  
 349 hydropower generated across all the scenarios considered: at Akosombo Dam these are 5,100 and  
 350 845 GWh/year respectively and at Kpong Dam, 1,000 and 130 GWh/year respectively. The trade-offs  
 351 between the three e-flow objectives and other water users are shown with the ‘best’ or highest  
 352 performing operation policy for each objective highlighted. Additionally, room for compromise,  
 353 characterised in this study as operation policies meeting the firm hydropower demand of 4,415  
 354 GWh/year for Akosombo Dam (“fair hydropower”) and e-flow demands at least 80% of the time (‘fair  
 355 environment’), have also been highlighted. It is important to note that, the terms ‘best’ and ‘fair’ as  
 356 used here, are not qualifying adjectives but solution descriptors in the Pareto approximate space with  
 357 the latter used to denote ‘reasonable’ or ‘satisfactory’ rather than ‘equitable’ solutions. Cumulative  
 358 distribution graphs showing the function values for the baseline and future scenarios are presented in  
 359 Figure S4 in the supplementary materials.

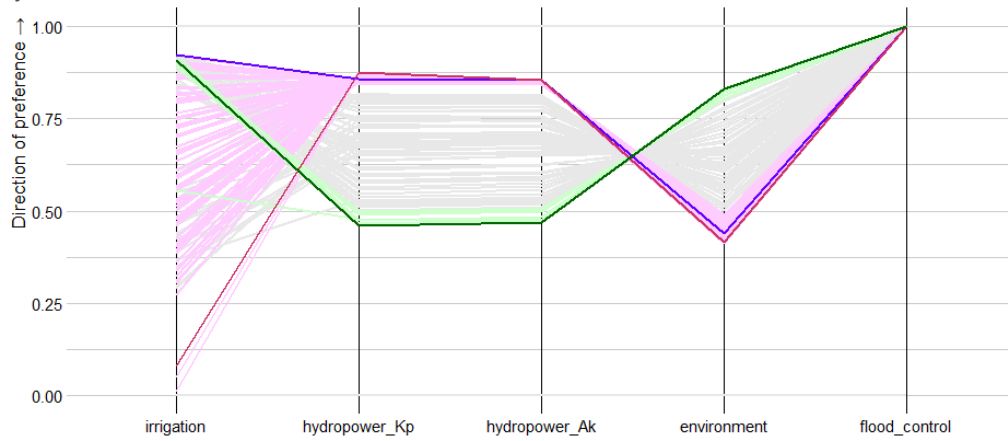
#### 360 4.1 Baseline scenario

361 For the baseline scenario (Figure 4), the highest performing dam operation policies for hydropower  
362 trade-off sharply with the provision of e-flows (all configurations) in the Lower Volta such that there  
363 is no overlap even among fair solutions for either objective for all e-flow configurations considered in  
364 this study. To meet the current firm energy requirement of 4,415 GWh/year at Akosombo, e-flows can  
365 only be released about 60%, 49% and 47% of the required time for the clam e-flows, e-flows 2 and e-  
366 flows 3, respectively. From the environmental perspective, the results show that hydropower demand  
367 from Akosombo and Kpong have to fall to approximately 3,903 GWh/year and 760 GWh/year  
368 respectively for the release of clam e-flows; or approximately 3000 GWh/year and 563 GWh/year for  
369 the release of e-flows 2; or alternatively approximately 2711 GWh/year and 508 GWh/year for the  
370 release of e-flows 3 to become possible 80% of the recommended time under current climatic  
371 conditions. The solutions for clam e-flows generally lead to higher hydropower generation compared  
372 to the other e-flows because for seven months of the year there is no constraint on water releases in  
373 the Lower Volta for the environment, hence hydropower generation can be maximised in these  
374 months. Comparing the dynamic e-flow configurations, e-flows 2 yields higher hydropower generation  
375 because its dry season flow recommendation is higher at 700 m<sup>3</sup>/s as compared to that of e-flow 3  
376 which is 500 m<sup>3</sup>/s thus allowing for higher power generation in the dry season.

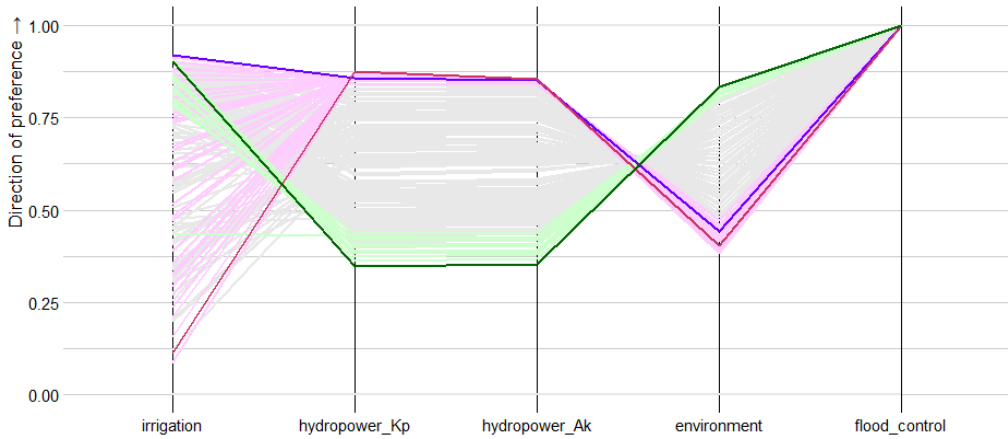
377 Considering the relatively low water demand for irrigation as compared to hydropower, marginal  
378 increases in hydropower generation at Kpong lead to significant reduction in the amount of irrigation  
379 demand that is met in the baseline scenario. The solutions for all e-flow configurations perform well  
380 for the flood control objective even though e-flow 2 and 3 prescribe flood releases for two months of  
381 the year. As such, comparing clam e-flows to e-flows 2 and 3, there is a reduction (0.99 for clam e-  
382 flows vs. 0.83 for e-flows 2 and 3) in the performance of the 'best environment' solution for the latter  
383 two, as expected, showing that the requirement for floods for two months in a year in those e-flow  
384 configurations are not met.



a: Clam e-flows



b: E-flows 2



c: E-flows 3

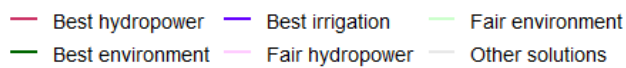


Figure 4: Full set of non-dominated solutions in the baseline scenario with the 'best' solution for each water user and 'fair' solutions for hydropower and environment highlighted. The objective values for each water user have been normalised using the minimum and maximum values over all simulations with the highest performance for each objective placed at the top of the axes. The order in which water users are presented has been chosen to highlight trade-offs between them.



385 4.2 Future scenarios

386 The effects of different climate futures on the Pareto approximate solutions for the Volta basin are  
387 presented in Figure 5. Scenario 1, where there is 45% decrease in annual inflows to Akosombo dam  
388 stands out as the system becomes water stressed so that the best performing operation policies even  
389 for hydropower lead to only about 2776 GWh and 550 GWh annual power generation at Akosombo  
390 and Kpong respectively for all e-flow configurations. The best operating policies for the environment  
391 however improve slightly from 0.99 to 1 for clam e-flows and remain unchanged at 0.83 for e-flows 2  
392 and 3 relative to the baseline. This is because these solutions, even in the baseline scenario are those  
393 for which only dry season low flows are released. In contrast to Scenario 1, under Scenario 3 and to a  
394 lesser extent Scenario 2, where annual flows to the Akosombo Dam increase by 65% and 12%  
395 respectively, the solutions move up on the two hydropower axes and some fair solutions for the  
396 environment lead to higher annual hydropower production of up to 4,242; 3,392 and 2,926 GWh/year  
397 for clam e-flows, e-flows 2 and e-flows 3 respectively at Akosombo.

398 Seasonal climate change effects on the Lower Volta system under scenarios 4 and 5 are comparable  
399 to annual climate change effects. As a result, the solutions for Scenario 4 are similar to Scenario 2  
400 while those for Scenario 5 are similar to Scenario 3 save for the slightly lower hydropower generation  
401 values for Scenario 5 and hence fewer 'fair hydropower' solutions in line with its relatively lower  
402 inflows (+65% annual inflows for Scenario 3 vs 55% wet season inflows for Scenario 5). This is due to  
403 the high residence time of water in the Lake Volta (3.9 years) and the fact that the Lower Volta has  
404 highly seasonal inflows naturally so that an annual inflow increase applied to inflows across the year,  
405 as applied in Scenario 2 and 3, results in a minimal increase in the absolute values of inflows to the  
406 dam in the dry season but a relatively significant increase in the absolute values of wet season inflows,  
407 thus amplifying seasonality.

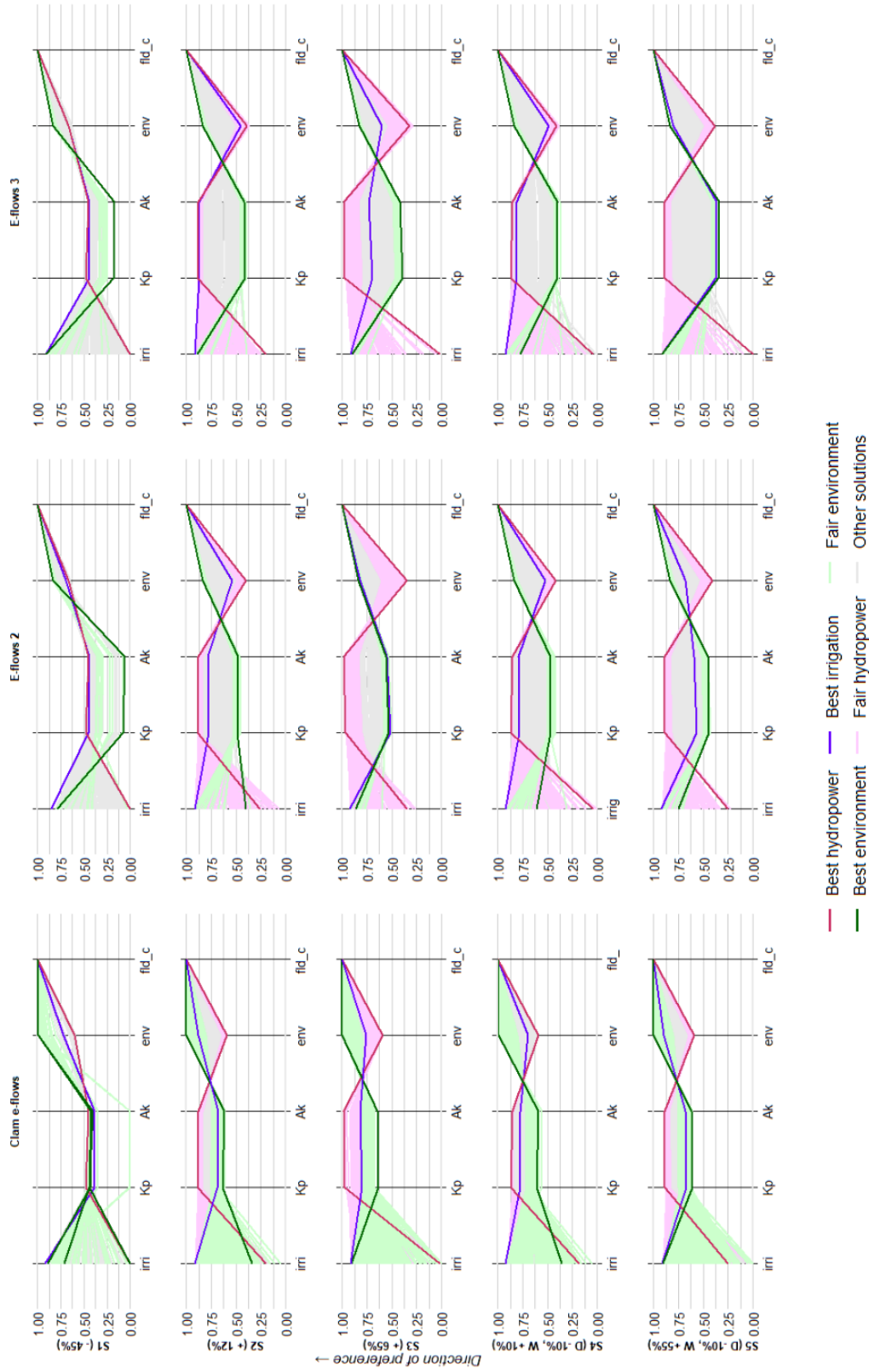


Figure 5: Full set of non-dominated solutions in the future scenarios with the 'best' solution for each water user and 'fair' solutions for hydropower and environment highlighted. The objective values for each water user have been normalised using the minimum and maximum values over all simulations with the highest performance for each objective placed at the top of the axes. The order in which water users are presented has been chosen to accentuate the trade-offs. Notation- water users: Irri- irrigation, Kp- hydropower from Kpong Dam, Ak- hydropower from Akosombo Dam, env- environment (clam e-flows, e-flows 2 or e-flows 3), fld\_c- flood control. Notation- scenarios: S1 to S5- Scenario 1 to 5, D- dry season flow, W-wet season flow.

## 409 5 Discussion

410 The Lower Volta River System is characterized by the dominance of hydropower generation for Ghana  
411 and its neighbouring countries. This has come at a high cost to downstream ecosystem services and  
412 communities (Tsikata, 2008; Lawson, 1972; Ntiamoah-Baidu et al., 2017). The results from this study  
413 show that likewise, some cost to hydropower production would have to be borne for e-flows  
414 implementation and the restoration of some of these ecosystem services under current climatic  
415 conditions. For the implementation of clam e-flows 80% of the time, i.e., a fair environmental solution,  
416 the country would forfeit at least 11.6% of the firm annual power demand from Akosombo Dam. For  
417 the implementation of only the dry season flow recommendations of e-flows 2 and 3, about 32% and  
418 38% respectively of current firm energy requirement would have to be supplemented with power  
419 generation from other sources. The release of floods as recommended in the dynamic e-flow  
420 configurations (e-flows 2 and 3) is not a Pareto approximate operation policy within the current  
421 operating constraints of the Akosombo Dam because in addition to flooding pre-dam floodplains  
422 which are now permanently inhabited (Ayivor and Ofori, 2017), releasing flows above the maximum  
423 turbine capacity of 1,603 m<sup>3</sup>/s at Akosombo means that these water volumes are lost to power  
424 generation. Flood releases also far exceed irrigation water demands.

425 While the majority of climate predictions for the Volta River generally point to an increase in annual  
426 water availability (Kunstmann and Jung, 2005; Aerts et al., 2006; Jung et al., 2012; Abubakari, 2021;  
427 Jin et al., 2018; Sylla et al., 2018), based on this study, an argument can be made that both an increase  
428 or a decrease in inflows to the Lower Volta enhance the potential for e-flows implementation  
429 compared to the current baseline. On the one hand, an increase in inflows to the Akosombo dam as  
430 applied in scenario 3, reduces the amount of the firm energy requirement that would have to be  
431 supplemented by other sources for the implementation of 'fair environmental solutions' to about  
432 3.9% (vs 11.6%) for clam e-flows, and then 23.2% (vs 32%) for e-flows 2 and 33.7% (vs 38%) for e-flows  
433 3. On the other hand, a decrease in inflows to the Akosombo Dam, whereby at best only 2,774  
434 GWh/year of hydropower can be generated, provides opportunity to strategically release  
435 recommended dry season e-flows to reap some environmental benefits out of a 'bad' situation where  
436 annual flow releases from the dam will be low anyway. This operation policy under dry climate  
437 scenarios could also be adopted in dry years, in essence modelling the Episodic E-flows  
438 Implementation approach, which is an opportunistic approach to dam re-operation that takes  
439 advantage of prevailing hydrological conditions (Warner et al., 2014; Yang and Yang, 2014; Owusu et  
440 al., 2021). This contrasts with the alternative approaches, Adaptive Management and Blanket

441 Operation which represent more structural inclusion of e-flows in the dam operation policy (Warner  
442 et al., 2014).

443 Only future climate scenarios were modelled in this study; however, inferences can also be made on  
444 the effect of simple energy and water demand futures on the Lower Volta system. For instance, an  
445 increase in irrigation demand will trade-off against hydropower production at Kpong Dam and an  
446 increase in the firm energy requirement or the continuation of the *de-facto* policy of hydropower  
447 maximisation at Akosombo Dam, despite the availability of alternative power generation sources  
448 (Dye, 2020; Kumi, 2017), will weaken the potential for re-operation of the dam for the riverine  
449 environment. Changes in upstream water consumption as well as the construction of new dams such  
450 as the Pwalugu Dam in northern Ghana will also affect inflows to the Akosombo Dam. Gonzalez et al.  
451 (2021), however, show that practical coordination of the operation of major infrastructure in the Volta  
452 Basin, as compared to the current approach whereby dam operators fail to consider downstream built  
453 infrastructure, reduces the impact on inflows to the Akosombo Dam in particular, and also maximises  
454 basin-wide benefits. Undoubtedly this coordination should extend beyond the Volta Basin to include  
455 the entire electricity generation portfolio of Ghana and neighbouring countries to further reduce the  
456 impact of e-flows implementation in the Lower Volta on power supply.

457 Expanding on the current electricity generation portfolio of Ghana, the contribution of other  
458 renewable energy sources besides hydropower to the power mix has remained under 1% since 2000,  
459 despite an on-grid target of 10% by 2020 (now extended to 2030) (Acheampong et al., 2021; Energy  
460 Commission Ghana, 2022). The alternative sources of electricity in Ghana use carbon fuels for thermal  
461 power generation, accounting for approximately 65% of the electricity generation portfolio in 2020  
462 (Dye, 2020; Acheampong et al., 2021; Energy Commission Ghana, 2022). It is expected that these  
463 alternative carbon-based power sources contribute more to climate change compared to power  
464 generation from Akosombo and Kpong dams due to the fact greenhouse gas emissions from  
465 hydropower dams is negatively correlated with dam age and even the more recent dam, Kpong, has  
466 been in operation for over 35 years (dos Santos et al., 2006; Barros et al., 2011). As such, dam re-  
467 operation in Ghana may have the long-term environmental and economic consequences of higher  
468 greenhouse gases emissions if it results in a higher reliance on the existing carbon-based power  
469 generation options, rather than other renewables like solar and wind power. Furthermore, in Ghana,  
470 hydropower has traditionally been a cheaper source of electricity compared to fossil fuel-based power  
471 generation and as Ghana has increased its reliance on the latter, electricity generation costs have  
472 increased resulting in higher tariffs for consumers (Energy Commission Ghana, 2022; Public Utilities  
473 Regulatory Commission, 2015). Finally, any reduction in hydropower production from Akosombo and  
474 Kpong dams due to re-operation may result in reduced overall electricity supply in Ghana as

475 experienced during periods of drought in the past (Dye, 2020). It is estimated that the negative  
476 economic impacts of power shortages and load shedding, such as decreased productivity in industries,  
477 loss of revenue for businesses, and increased costs for backup power sources led to a GDP reduction  
478 of about 1.8-2% during the 2014-2016 power crisis (Acheampong et al., 2021). Considering these  
479 potential adverse impacts of dam re-operation, it is recommended that future studies encompass a  
480 deeper analysis of the energy landscape of Ghana and investigate carbon emissions and the path to  
481 greener energy in the country, as well as energy pricing and economic implications.

482 By exploring the room for compromise in the Lower Volta with respect to e-flows implementation this  
483 research has taken a first step towards a comprehensive assessment of the trade-offs involved at a  
484 national and local level. The potential re-operation of the Akosombo and Kpong dams can also benefit  
485 from (i) the groundwork laid by research on the pre- and post-dam river system (Lawson, 1972;  
486 Tsikata, 2008; De-Graft Johnson, 1999; Nyekodzi et al., 2018; Obirikorang et al., 2013; Adjei-Boateng  
487 et al., 2012; Owusu et al., 2022b), (ii) insights deriving from interviews and extensive stakeholder  
488 engagement (Ayivor and Ofori, 2017; Ohemeng et al., 2017; Nukpezah et al., 2017), and (iii) existing  
489 supporting legislation for e-flows implementation (L.I.1692 Water Use Regulations, Ghana, 2001).  
490 Indeed, research on successful and stalled cases of dam re-operation indicates that stakeholder  
491 engagement and supporting legislation enhance the chances of successful e-flows implementation  
492 (Owusu et al., 2022a, 2021) .

493 Finally, the successful application of the EMODPS framework in exploring trade-offs inherent to e-  
494 flows implementation in a heavily modified river under uncertainty holds promise for similar  
495 applications elsewhere. In order to find a policy for multiple objectives in such cases, a flexible  
496 structure to map states to actions is needed. With traditional control optimization techniques, the  
497 uncertainties need to be modelled explicitly, which creates a high computational burden and limits  
498 the ability to evaluate a large set of uncertainties (Giuliani et al, 2016). EMODPS overcomes this  
499 challenge by directly conditioning the decisions to exogenous information without requiring an explicit  
500 probabilistic model. With EMODPS, only the goals and direction of preference are required in setting  
501 up the multi-objective decision problem, making the use of this method feasible even in data scarce  
502 conditions. This study therefore concurs with Herman et al. (2020) who argue that direct policy search  
503 methods are a promising technique to enable adaptivity in water resources assessment by allowing  
504 the flexible integration of new information about the system into management decision making.

## 505 6 Conclusion

506 A dam is designed with future uses in mind – this provides the justification for its construction. The  
507 future, however, can turn out differently from that envisaged in the dam design. Therefore, re-

508 operation of the dam to meet changing demands is a likely necessity. This study investigates current  
509 and future trade-offs between water users in the Lower Volta River Basin and specifically explores the  
510 potential to deliver environmental flows to support various ecosystem services that have been  
511 negatively impacted by the current operation of the Akosombo and Kpong dams. The results highlight  
512 the dominance of hydropower production in the Lower Volta; if this relaxes there is more opportunity  
513 for restoration of the riverine environment under current climate and water use conditions and even  
514 more so under future scenarios where inflows to the Lake Volta increase. In future scenarios whereby  
515 inflows to the Lake Volta decrease, it is still possible to strategically manage and time water releases  
516 to provide dry season low flows which will support the clam fishery and help control aquatic weeds  
517 and some water borne diseases in the Lower Volta. This study applied advanced optimisation  
518 techniques to identify and analyse dam operation policies for e-flows under discrete climatic  
519 scenarios. Future studies should focus on the robustness and limits of these policies under  
520 multitudinous future climatic and water use scenarios. Such robustness studies, together with flow  
521 experimentation, will reveal dam operation policies that may be adopted with some confidence  
522 presently. It will also build on groundwork already laid, through e-flows legislation and extensive  
523 collaborative scientific studies, for the successful re-operation of the Akosombo and Kpong dams for  
524 the environment and other water users.

525

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#### 532 **Data availability**

533 The data associated with this manuscript will be made available upon consultation with the national  
534 organisation, Volta River Authority (VRA), who owns the data

#### 535 **Declaration of interests**

536 Some authors are members of the editorial board of Hydrology and Earth Systems Sciences journal.  
537 The authors declare that they have no other known competing financial interests or personal  
538 relationships that could have influenced the work reported in this paper.

539 **Author contribution**

540 AO: Conceptualization, Methodology, Data collection, Formal analysis, Interpretation, Writing-  
541 original draft preparation.

542 JZS: Methodology, Formal analysis, Writing-review and editing.

543 MM: Conceptualization, Data collection, Writing-review and editing, Supervision, Project  
544 administration.

545 PvdZ: Conceptualization, Writing-review and editing, Supervision, Funding acquisition.

546 JS: Conceptualization, Methodology, Writing-review and editing, Supervision, Funding acquisition

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