



1 **Is the reputation of *Eucalyptus* plantations for using more**
2 **water than *Pinus* plantations justified?**
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26



27 **Abstract**

28 The effect of *Eucalyptus* plantations on water balance is thought to be more severe than for commercial
29 alternatives such as *Pinus* species. Although this perception is firmly entrenched, even in the scientific
30 community, only four direct comparisons of the effect on the water balance of a *Eucalyptus* species and a
31 commercial alternative have been published. One of these, from South Africa, showed that *Eucalyptus grandis*
32 caused a larger and more rapid reduction in streamflow than *Pinus patula*. The other three, one in South Australia
33 and two in Chile, did not find any significant difference between the annual evapotranspiration of *E. globulus* and
34 *P. radiata* after canopy closure.

35

36 While direct comparisons are few, there are at least 57 published estimates of annual evapotranspiration of either
37 a *Eucalyptus* or *Pinus* species. This paper presents a meta-analysis of these published data. Zhang et al. (2004)
38 fitted a relationship between the vegetation evaporation efficiency and the climate wetness index to published
39 data from catchment studies and proposed this approach for comparing land uses. We fitted the same model to
40 the published data for *Eucalyptus* and *Pinus* and found that the single parameter of this model did not differ
41 significantly between the two genera ($p=0.48$). This implies that for a given climate wetness index the two genera
42 have similar annual water use. The residuals compared to this model were significantly correlated with soil depth
43 for *Eucalyptus*, but this was not the case for *Pinus*. For *Eucalyptus* the model overestimates the vegetation
44 evaporation efficiency on deep soils and underestimates the vegetation evaporation efficiency on shallow soils.

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48 **1. Introduction**

49 There are now more than 23 Mha of *Eucalyptus* plantations in the temperate and tropical zones of the world
50 (Keenan et al., 2015; Macdicken et al., 2016). These plantations extend from near the equator to approximately
51 43 degrees of latitude North and South and play an important and growing role in minimizing the gap between
52 global demand for wood products and the supply (Kanninen, 2010). These *Eucalyptus* plantations are mostly
53 established in seasonally dry climate zones (dry tropics, sub-tropics, and Mediterranean climate types). This and
54 the reputation of *Eucalyptus* for high rates of water use when compared to alternatives, mean that wherever large-
55 scale planting of *Eucalyptus* has occurred, it has been associated with concern, debate and often protest about the
56 effect of these plantations on the security of water supply (Albaugh et al., 2013). Afforestation with *Pinus* and
57 other genera has also resulted in concern about changes in local hydrology (Huber and Iroumé, 2001; Little et al.,
58 2009) but has not been associated with the same level of polemic or controversy as the planting of *Eucalyptus*.

59
60 In 2010, plantations managed for wood production occupied a total land area 109 Mha (Kanninen, 2010).
61 Approximately 35% of these plantations were of *Pinus* species while 10% were *Eucalyptus* (Kanninen, 2010).
62 The annual increase in production plantations between 2010 and 2015 was 1.2%. During this time the total area
63 of *Pinus* plantations remained virtually unchanged and much of the global increase was in either *Eucalyptus*
64 plantations or other short rotation options such as *Acacia* (Payn et al., 2015). The global trends in plantations are
65 towards *Eucalyptus* or species managed on short rotations to grow pulp or biomass for energy. While these global
66 trends are important, the conflict associated with the establishment of *Eucalyptus* plantations and the potential for
67 reduced water availability manifests locally. In South Africa and South Australia these concerns have resulted in
68 legislation to regulate either water use (Greenwood, 2013) or planting (Albaugh et al., 2013). The effects of
69 *Eucalyptus* on water are currently being actively debated in Chile, where Arauco SA (the largest plantation grower
70 in Chile and the second largest pulp producer in the world) plan to replace approximately 250,000 ha of *P. radiata*
71 plantations with *Eucalyptus*. In China regional governments are supporting research to investigate the water
72 benefits of mixed plantings of local species with *Eucalyptus*. It is also likely that the global goal of reduced CO₂
73 emissions will intensify debate about *Eucalyptus* water use. Given the dominance of the global plantation estates
74 by species of *Pinus* and *Eucalyptus* and the direct substitution of *Pinus* with *Eucalyptus*, a quantitative comparison
75 between the water use characteristics of these two genera is timely.

76
77 The evidence that plantations use more water than grasslands or dryland crops is very strong (Zhang et al., 2001;
78 Zhang, 2004). Similarly, there is evidence that plantations use more water, and therefore generate less streamflow,
79 than natural forest in Chile (Huber et al., 2008), Brazil (Almeida et al., 2007; Meinzer et al., 1999) and Spain
80 (Rodriguez Suarez et al., 2014). The magnitude of the difference between plantations and natural forest is less
81 than that observed between plantations and annual pastures (Zhang et al., 2004).

82
83 While there is a perception that *Eucalyptus* use more water than alternative commercial plantation options such
84 as *Pinus*, three of four published comparisons of the water use (defined as evapotranspiration) reported no
85 difference between the water use of these two genera. The evidence for higher rates of water use by *Eucalyptus*
86 is mostly from South Africa where, in a paired catchment study, Scott and Lesch (1997) showed that, at least in
87 the early stages of growth, *Eucalyptus grandis* W. Hill. used up to 92 mm more water per year than *Pinus patula*



88 Schiede ex Schldtl. et Cham. In another direct comparison of the water use of a *Pinus* and *Eucalyptus* species in
89 plantations, Benyon et al. (2006) found that the annual water use of plantations of *E. globulus* Labill. and *P.*
90 *radiata* D. Don., with or without access to shallow fresh groundwater, were not significantly different. Recent
91 stand and catchment scale comparisons of *P. radiata* and *E. globulus* in central Chile have found that not observed
92 significant differences between the average annual water use of *P. radiata* and *E. globulus* (Iroumé et al., 2021;
93 White et al., 2021). Given these equivocal results, and the trend towards more planting of *Eucalyptus*, it is
94 important to understand when and why differences might occur in the water balance of *Pinus* and *Eucalyptus*
95 plantations.

96

97 While it seems that the maximum rates of water use by *Eucalyptus* and *Pinus* can approach the energy limit, there
98 do seem to be differences between commercial *Pinus* and *Eucalyptus* in their response to soil drying. Studies in
99 Brazil (Lima et al., 1990) and Tasmania, Australia (Honeysett et al., 1996) have shown that when planted in deep
100 soils and with regular inputs of rainfall or irrigation, *Eucalyptus* plantations can use water at a rate that approaches
101 the energy limit. Similarly high rates of water use have also been observed in *P. radiata* plantations in southern
102 Australia (Benyon et al., 2006) and in Chile (Huber and Iroumé, 2001) and there are reports of rates of water use
103 close to the energy limit in both oil palm (Röll et al., 2015) and rubber plantations (Tan et al., 2011). Studies in
104 China have found that the annual rate of water use by *Eucalyptus* can be substantially less than both rainfall and
105 available energy (Lane et al., 2004; Ren et al., 2019). This occurs during the dry season and has also been observed
106 in *Pinus* species (Myers et al., 1998). Notwithstanding these similarities it has been observed that the water use
107 of *Pinus* species decreases more rapidly with the onset of water stress than is the case with commercial *Eucalyptus*
108 alternatives for the same site (Teskey and Sheriff, 1996).

109

110 Reviews of the water use potential of *Eucalyptus* have highlighted the variability of reported rates of both
111 transpiration and evapotranspiration (Albaugh et al., 2013; Shi et al., 2012), yet there has been no systematic
112 attempt to determine if the high rate of water use observed in some studies is a characteristic of *Eucalyptus* in
113 plantations or has more to do with the conditions that prevailed in those studies. Most of the published studies of
114 water balance, with a couple of exceptions (Mendham et al., 2011; Scott and Lesch, 1997) have reported water
115 balance measurements made within a single rotation and most studies cover only a small proportion of that
116 rotation. It is likely that plantations must eventually reach a long-term equilibrium with the local climate and that,
117 except in circumstances where trees have access to water from off-site such as a regional aquifer (see O'grady et
118 al. (2011b) for a meta-analysis), these high rates of water use, often observed early in the first rotation, will not
119 be sustained. What is needed is to determine if the longer-term equilibrium water balance of catchments planted
120 to *Eucalyptus* will be associated with different levels of water storage, and therefore stream flow, from that under
121 alternative species options for wood production plantations (McDonnell, 2017).

122

123 While there are only three direct comparisons of the annual water balance of *Pinus* and *Eucalyptus*, there are many
124 studies that quantify annual water use by either a *Eucalyptus* or a *Pinus* species. These studies, and their estimates
125 of water use are very situation specific. Comparison of alternative land uses is complicated by the dominant role
126 that climate and hydrogeology play in determining the local water balance. While vegetation cover has a smaller
127 effect on catchment water balance than either climate or hydrogeology it is the part of the system that can be



128 actively managed. If studies are available for the two genera from a comparable range of annual rainfall and
129 evaporative environments, then comparison might be possible through normalizing water use (evapotranspiration)
130 with respect to potential or energy limited evaporation and plotting this as a function of the climate wetness index
131 (ratio of rainfall to potential evaporation). This approach has previously been used to compare the water use of
132 forests with dryland agriculture (Zhang et al., 2004).

133

134 In this study, we collated published annual water balance estimates for plantations with either *Eucalyptus* and/or
135 a *Pinus* species, and fitted the model described by Zhang et al. (2004) to test the null hypothesis that the
136 evaporation characteristic of commercial *Pinus* and *Eucalyptus* plantations was not significantly different. We
137 also test the hypothesis that variation from this model is determined by variation in soil depth.

138 2. Methods

139 This paper presents a meta-analysis of published measurements of the water balance of *Eucalyptus* and *Pinus*
140 plantations in tropical and temperate regions. The focus of this analysis is on post-canopy closure plantations in a
141 notional equilibrium with the site. The behaviour of plantations is quantified by comparing an index of the function
142 of the crop (the vegetation evaporation efficiency, *VEE*) with an index of climate wetness in the manner proposed
143 by Budyko (1974) and applied by Zhang et al. (2004) to compare forests with dryland agricultural systems.

144 2.1 Definitions of terms

145 The terms evapotranspiration, water-use, potential evaporation, vegetation evaporation efficiency and climate
146 wetness index have various meanings in the scientific literature and to avoid ambiguity, they are defined here as
147 they are used in this paper.

148 2.1.1. Evapotranspiration and water-use

149 Evapotranspiration (*ET*) and water-use are used in this paper to describe total evaporation from a vegetated land-
150 surface. They are the sum of transpiration of all plants (*T*, the evaporation through leaf and other plant surfaces
151 of water drawn from the soil and transported to sites of evaporation through the xylem), water intercepted by plant
152 canopies and evaporated without reaching the ground (interception, *I*) and evaporation of water directly from soil
153 and litter (often called soil evaporation, *E_s*). All these processes are affected by the choice of crop and by the
154 management of that crop and should therefore be included as part of the water-use of that vegetation.

155 2.1.2. Potential Evaporation (PET)

156 Evapotranspiration (*ET*) by any land-use is situation specific; it is affected by the climate (energy and rainfall),
157 the structure and function of the vegetation and by characteristics of the soil and the litter. In this paper, for the
158 purposes of comparison, estimates of water-use or evapotranspiration are normalized relative to measures of the
159 local water supply (rainfall) and potential evaporation, which represents the energy limited maximum rate of
160 evaporation. There are numerous measures of reference or potential evaporation including Penman Potential
161 Evaporation (Penman, 1949), FAO-56 Reference Evaporation (Allen et al., 2005), Pan Evaporation and Priestley
162 Taylor Potential Evaporation (Priestley and Taylor, 1972). They are all intended to represent the maximum



163 possible rate of evaporation by a land surface covered with vegetation. In this paper, potential evaporation (*PET*)
164 always refers to Priestley-Taylor potential evaporation (see the notes under data analysis below to see how
165 Priestley-Taylor *PET* was calculated for each site). We have used the coefficient 1.26 in the Priestley-Taylor
166 equation; this coefficient accounts for the extra roughness of forests when compared with short crops and pastures
167 (Eichinger et al., 1996). The evapotranspiration of plantations may still, of course, exceed this measure of *PET*.
168 This may be the case if there is an additional source of energy such as advection or movement of hot air into the
169 forest. This might occur at the edge of a plantation, especially if it is adjacent to an area of land from which there
170 is a large sensible heat flux. The choice of method for calculating *PET* is less important than applying the same
171 method for all calculations in this analysis.

172 2.1.3. Water- and energy limit, vegetation evaporation efficiency (*k*) and climate wetness index (*CWI*)

173 The climate imposes limits on evapotranspiration. Evapotranspiration cannot exceed the amount of water
174 available which is usually limited to rainfall but may include irrigation and soil stored water and ground water
175 (O'grady et al., 2011a). Similarly, although evapotranspiration may exceed the calculated *PET* under some
176 circumstances, it is ultimately limited by available energy.

177

178 The relationship between the ratio of actual evapotranspiration to reference evaporation) and the climate wetness
179 index (*CWI*, the ratio of rainfall to potential evaporation) (Budyko, 1974) provides a simple way of partitioning
180 rainfall between evaporation and runoff. The ratio of evapotranspiration to potential evaporation is often termed
181 the 'evaporation efficiency' of a surface (Komatsu, 2003) and a convention has developed where the surface is
182 included in the name. For example, the ratio of evaporation from a soil to the potential soil evaporation is referred
183 to as the soil evaporation efficiency (Merlin et al., 2016). In this paper, the ratio of evapotranspiration to reference
184 evaporation for commercial plantations of *Eucalyptus* and *Pinus* species is referred to as their vegetation
185 evaporation efficiency (*VEE*). A more 'evaporation efficient' plantation converts a relatively greater proportion
186 of available energy to latent rather than sensible heat.

187

188 Zhang et al. (2004) developed a simple model that predicted vegetation evaporation efficiency (*VEE*) as a function
189 of the climate wetness index (*CWI*). This model is given by Equation 1 (equation A22 in Zhang et al. (2004)
190 below and includes the parameter *c* (an empirical catchment characteristic) which captures the effect of
191 hydrogeology and vegetation cover on the vegetation evaporation efficiency.

$$192 \quad \quad \quad VEE = 1 + CWI - (1 + CWI^c)^{\frac{1}{c}} \quad \quad \quad \text{Equation 1}$$

193

194 2.2. Meta-Analysis of Published Studies

195 While direct comparisons of the water balance of *Eucalyptus* and *Pinus* plantations are few there are a reasonable
196 number of previously published estimates of either streamflow or evapotranspiration. These data were collated
197 and used in the meta-analysis described below. The studies included are described in some detail in the
198 supplementary material and the main features are summarised in Tables 1 and 2. A list of potentially suitable
199 references were first found by conducting a series of searches of the Web of Science and Google Scholar. The
200 following searches were conducted:

201



- 202 1. Title contains (evapotranspiration or water use) and (eucalypt or eucalyptus)
203 2. Title contains (evapotranspiration or water use) and (pine or pinus)
204 3. Paper contains (evapotranspiration or water use) and (eucalypt or eucalyptus)
205 4. Paper contains (evapotranspiration or water use) and (pine or pinus)

206

207 The first two searches yielded less than 100 papers in total. The latter two found many thousands of articles. The
208 200 most relevant in each search were checked to decide their suitability. For inclusion the paper must measure
209 or estimate evapotranspiration by a *Eucalyptus* or *Pinus* species for at least one year. Only planted forests managed
210 primarily for wood production were included. Agroforestry systems were excluded as were measurements made
211 prior to canopy closure. Native forests and burned forests and plantations with access to the water table were also
212 excluded. Several of the studies covered multiple years. A single value of rainfall and evaporation was calculated
213 as the average of all the years in each study. Sometimes a paper reported multiple estimates of evapotranspiration
214 for forests in the same location and growing under the same conditions. In these cases, average values were
215 calculated for the multiple sites.

216

217 After applying these criteria to articles found in the above searches, a total of 30 *Pinus* and 27 *Eucalyptus* stands
218 were included in the meta-analysis. The location, rainfall data and evapotranspiration data are provided as
219 supplementary material. The estimates of evapotranspiration were made using one of four methods. The method
220 applied in each study is indicated in Table 1.

221 **2.2.1. Method 1 – Measurement and addition of component fluxes**

222 At the stand or plot scale evapotranspiration (water-use) is the sum of evaporation from the soil and leaf litter (E_s),
223 evaporation of rainfall intercepted by the vegetation canopy (I) and transpiration or the direct uptake of water by
224 the trees and the evaporation of this water through the leaf surface (T). Evapotranspiration can therefore be
225 calculated as the sum of the component processes.

226 **2.2.2. Method 2 – One dimensional water balance**

227 Provided there is no leakage or runoff then evapotranspiration (ET) can be calculated in stand scale studies as the
228 sum of rainfall (P) and the change in the soil water content (ΔS) between two measurements.

229

$$230 \quad ET = P + \Delta S \quad \text{Equation 2.}$$

231 **2.2.3 Method 3 – Catchment water balance**

232 For a catchment, if there is no change in the amount of water stored in the soil or the groundwater (ΔS),
233 evapotranspiration (ET) is simply the difference between rainfall and streamflow (Q). Over long time periods it
234 is often assumed that the change in storage is negligible; this is less valid as the period of the estimate is reduced
235 or if the annual total rainfall has a clear temporal trend.

236

$$237 \quad ET = Q - P + \Delta S \quad \text{Equation 3}$$



238 **2.2.4 Method 4 – Eddy covariance (flux towers)**

239 Properly located flux towers can be used to estimate the net carbon and water flux (evapotranspiration) above an
240 ecosystem. The instruments on these towers measure the total solar and net radiation and partition this to latent
241 (evapotranspiration) and sensible heat flux (air temperature change) and heat storage changes in soil and biomass.
242 The covariances of high frequency measurements of air temperature, humidity and CO₂ are used to calculate total
243 evaporation and carbon exchange between the atmosphere and the underlying vegetation (Aubinet et al., 2012).
244 Measurements are typically made on a 30-minute time interval to represent fluxes from an upwind surface area
245 or “footprint”. The area of the footprint is dependent on strength of the turbulence in the air, a function of wind
246 speed and surface roughness elements, and the height of the instruments, thus the location of land surface
247 influencing the measurements changes through time. Eddy covariance measurements give total fluxes from the
248 contributing footprint, thus are useful for total ecosystem energy, water and carbon balances. However,
249 partitioning the fluxes between different contributing vegetation and soil components requires additional
250 measurements, such as sap flow, rain throughfall and soil evaporation. Also, the measurements are unreliable
251 during periods of stable air and low turbulence, such as still cold nights but, for the purposes of the analyses in
252 this paper, these are periods typically with very low water fluxes and have only minor influence on the total system
253 water balance. There is a substantial literature describing these methods and complementary measurements, a
254 detailed description is beyond the scope of this paper but can be found in Wilson et al. (2001) where the method
255 is compared with alternatives.

256 **2.3. Variations at Two Sites**

257 A study by Scott and Lesch (1997) at Mokobulaan in South Africa reported more rapid changes in streamflow
258 after planting of *E. grandis* than after planting of *P. patula*. The soil was very deep, and it is probable, though this
259 was not measured, that evapotranspiration exceeded rainfall and that this was more pronounced in the *E. grandis*
260 than the *P. patula*. To allow for this effect we assumed a storage of 100 mm per metre of soil and a rate of root
261 extension of 2 m per year for *E. grandis* after (Dye, 1996) and 1 m per year in *P. patula*. This relative rate is
262 consistent with the observation that streamflow ceased 5 and 10 years respectively, after planting of *E. grandis*
263 and *P. patula* (Scott and Lesch, 1997).

264
265 Another study included here was made at Lewisham in Tasmania, Australia by Honeysett et al. (1996). In this
266 study the effect of drought on the water relations and water balance of *E. globulus* and *E. nitens* were investigated
267 using irrigated controls and rainfed plots. The irrigated treatments were excluded from this meta-analysis.
268 However, to avoid mortality the rainfed treatments received some supplementary irrigation. This irrigation is
269 included in the rainfall figure in Table 1 and in the supplementary material.

270 **2.4. Derived climate and vegetation indices**

271 In each of the papers included in this analysis, evapotranspiration (*ET*) was estimated from the measurement of
272 other variables by one of the four methods described above. Rainfall data was available for all the studies included
273 in this review. Time series climate data from the 0.5-degree grid point closest to each site was also downloaded
274 for the duration of each experiment (Climate Research Unit Time Series v4.03, Harris et al., 2014). Net radiation



275 was calculated for the location after Hargreaves and Samani (1985.) and then Priestley-Taylor evaporation (*PET*)
276 was calculated as:

277

$$278 \quad \lambda PET = 1.26 \left[\frac{s}{s+\gamma} \right] R_n \quad \text{Equation 4}$$

279

280 where R_n is net radiation in W m^{-2} , λ is the latent heat of vapourisation of water (2245 kJ kg^{-1}), s is the slope of
281 the relationship between saturated vapour pressure and temperature ($\text{kPa } ^\circ\text{C}^{-1}$) and γ is the psychrometric constant
282 ($\text{kPa } ^\circ\text{C}^{-1}$). These ‘constants’ are temperature dependent; s was calculated using the empirical model in Equation
283 5 (Hahn and Landeck, 1998.) and γ was calculated using Equation 6 in which T_a and P_a are average daily air
284 temperature (calculated as the average of T_{max} and T_{min}) and atmospheric pressure (assumed to be 101.3 kPa), c_p
285 is the specific heat of dry air ($1.013 \text{ kJ kg } ^\circ\text{C}^{-1}$) and ε is the ratio of the molecular weight of water to dry air
286 (0.622).

287

$$288 \quad s = 0.04145e^{0.06088T_a} \quad \text{Equation 5}$$

289

$$290 \quad \gamma = \frac{c_p P_a}{\lambda \varepsilon} \quad \text{Equation 6}$$

291

292 For each measurement year at each study location the vegetation evaporation efficiency (*VEE*) and the climate
293 wetness index were also calculated using equations 7 and 8 respectively.

294

$$295 \quad VEE = \frac{ET}{PET} \quad \text{Equation 7}$$

296

$$297 \quad CWI = \frac{P}{PET} \quad \text{Equation 8}$$

298

299 **2.5. Meta-Analysis**

300 The values of the vegetation evaporation efficiency estimated from each of the published studies were plotted as
301 a function of the climate wetness index. The model described in Equation 1 was then fitted to the data using the
302 Nonlin function in *R* and the parameter c and the coefficients of determination, r^2 , value were calculated for each
303 genus separately and for the pooled data (R-Core-Team, 2013). Analysis of variance was also completed to test
304 for a significant difference between *Pinus* and *Eucalyptus* in the parameter c (R-Core-Team, 2013). The residuals
305 (predicted minus observed) were plotted against soil depth for the sites where this data was available. Linear
306 regression was used to explore the relationship between annual transpiration and annual evapotranspiration.
307 Simple t-tests for non-paired observations were used to test for differences between genera in annual
308 evapotranspiration and the ratio of evapotranspiration to rainfall.



309 **3. Results**

310 **3.1. Rainfall Limited Plantations**

311 Twenty-seven *Eucalyptus* and 30 *Pinus* sites were included in the meta-analysis. The details of these sites are
312 summarized in three tables. The most detailed information is in the supplementary material together with the
313 measured and calculated climatic data, estimated evapotranspiration, and the detailed results of the data analysis.
314 The papers from which the data were taken are listed in Table 1 with the rainfall data, species studied, and the
315 method used to estimate evapotranspiration. Table 2 summarises the range of climatic conditions and evaporation
316 rates by species and indicates the number of studies for each species by country or continent.

317

318 The analysis included sites from tropical, dry tropical, sub-tropical, warm temperate, cool temperate,
319 Mediterranean, and montane climates with both genera represented in all but one climate type and in most
320 locations. There is a bias of *Pinus* studies to the United States and of *Eucalyptus* to Australia (Table 2). Species
321 of *Eucalyptus* represented in order of decreasing number of estimates were *E. globulus* (10), *E. nitens* (H. Deane
322 & Maiden) Maiden (7), *E. urophylla* S.T. Blake (3), *E. grandis* (2), *E. urophylla* x *grandis* (2), *E. urophylla* x
323 *globulus* (2) and *E. saligna* Sm. (1) (Table 1). Similarly estimates for species of *Pinus* were made for *P. radiata*
324 (18), *P. taeda* L. (5), *P. patula* (2), a mixed stand of *P. taeda* and *P. palustris* Miller (1), mixed stand of *P. elliottii*
325 Engel. and *P. palustris* (1), *P. elliottii* (1), *P. caribaea* var *hondurensis* W.H. Barrett and Golfari (1) and *P. strobus*
326 L. (1) (Table 1). Thus, each genus is represented by species from tropical, sub-tropical and temperate
327 environments.

328 **3.1.1. Annual Rainfall and Evapotranspiration**

329 The annual rainfall at the 24 *Eucalyptus* sites ranged from 489 mm at one of the South Australian sites to 2088 mm
330 at a site in the Rio Grande du Sol in Southern Brazil. The range of rainfall was similar for the 27 *Pinus* sites and
331 varied from 600 mm, at a South Australian site to 2081 mm at a site near Valdivia in south central Chile.
332 Interestingly, both the low rainfall site in South Australia and the high rainfall site in Chile were planted to *P.*
333 *radiata*. The situation was similar for average annual potential evaporation which ranged from 1005 to 2008 mm
334 at the *Eucalyptus* sites and from 1021 to 2004 mm at the *Pinus* sites (supplementary material). The median annual
335 rainfall for the *Eucalyptus* and *Pinus* sites respectively was 940 mm and 927 mm while average potential
336 evaporation was 1480 mm and 1551 mm (Table 2). Thus, the range and median conditions covered by the sites
337 included in this meta-analysis was very similar for both genera.

338

339 Annual evapotranspiration increased as a function of rainfall before plateauing in the same manner as reported by
340 Zhang et al. (2001). Annual rates of evapotranspiration reported for *Eucalyptus* species were between 488 mm at
341 a low rainfall site in South Australia planted to *E. globulus* (Benyon et al., 2006) and 1345 mm at a site in Brazil
342 planted to *E. urophylla* x *E. grandis* (Soares and Almeida, 2001). The lowest and highest annual evapotranspiration
343 for *Pinus* species were 355 mm for *P. radiata* at Jonkershoek in the Western Cape of South Africa (Lesch and
344 Scott, 1997) and 1291 mm for *P. strobus* in North Carolina (Ford et al., 2007).

345

346 The minimum, mean, median and maximum rates of evapotranspiration were all slightly greater for the *Eucalyptus*
347 sites than for the *Pinus* sites (Figure 1). This, albeit non-significant ($p=0.24$), difference was associated with the



348 *Eucalyptus* sites generally being on slightly wetter sites. When evapotranspiration was divided by rainfall the
349 median values of the ratio for the two genera were nearly identical at 0.77 and 0.76 (Figure 2). The ratio of
350 evapotranspiration to rainfall varied from 0.45 to 1.31 in *Eucalyptus* and from 0.44 to 1.2 in *Pinus* species. At one
351 site in South Africa (Lesch and Scott, 1997) the rate of evapotranspiration by *E. grandis* exceeded rainfall by 31%
352 (Figure 2). At the same site, evapotranspiration by *P. patula* exceeded rainfall by 19% (Figure 2).

353 3.1.2. Vegetation evaporation efficiency as a function of the climate wetness index (*Eucalyptus* and *Pinus*)

354 In Figure 3 the vegetation evaporation efficiency for each study site is plotted as a function of the climate wetness
355 index. For both the *Eucalyptus* and *Pinus* sites there is a strong, positive correlation between the vegetation
356 evaporation efficiency and the climate wetness index. For the *Eucalyptus* sites the model of Zhang et al. (2004)
357 (Equation 1) explained 66 % of the variation in the vegetation evaporation efficiency while for *Pinus* this
358 decreased to 63 %. The parameter c in the model described by Equation 1 fitted to the data was 2.84 for *Eucalyptus*
359 and 2.64 for *Pinus*. While this may be an important difference it was not statistically significant ($p=0.50$) and the
360 value for c when the relationship was fitted to the pooled data was 2.74 and the r^2 was 0.69. Figure 4 shows the
361 ratio of the predicted vegetation evaporation efficiency for *Eucalyptus* to the predicted vegetation evaporation
362 efficiency for *Pinus* as a function of climate wetness index. The maximum proportional effect of genus on the
363 vegetation evaporation efficiency of 3.5% is predicted to occur where the Climate Wetness Index is 1.

364 3.1.3. The effect of soil depth

365 While the relationships in Figure 3 are significant for both genera there is nonetheless substantial scatter. The soil
366 depth was not provided in all the papers included in this analysis. When the residuals (observed minus predicted)
367 were plotted as a function of the soil depth the relationship was significant for the *Eucalyptus* sites (Figure 5) but
368 not for the *Pinus* sites (data not shown). A linear relationship with soil depth explained 57% of the error for
369 *Eucalyptus* and indicated that the model shown in Figure 3, for c of 3.1, overestimated the vegetation evaporation
370 efficiency in shallow soils and underestimated it in deep soils (Figure 5), with the model having zero residual with
371 a soil depth around 10 m.

372 3.1.4. Transpiration as a proportion of evapotranspiration

373 A subset of the studies, again indicated in the supplementary material, also provided estimates of transpiration
374 made using sapflow sensors. For both *Eucalyptus* and *Pinus* there was a strong linear relationship between
375 transpiration and evapotranspiration with an approximate slope of 0.5 (Figure 6).

376 4. Discussion

377 The results of the meta-analysis of published records of evapotranspiration for *Eucalyptus* and *Pinus* species in
378 this paper suggest that for a given climate wetness index the water use of *Eucalyptus* and *Pinus* plantations is not
379 significantly different ($p=0.50$). This does not mean that there are not circumstances, or periods within a rotation,
380 when *Eucalyptus* will use more water than the alternatives. The water balance of plantations and alternatives is
381 very situation specific, and our focus should be on understanding the sources of variation rather than generalizing
382 about one land use option. The work of Scott and Lesch (1997) and the results of White et al. (2009) from three



383 *E. globulus* plantations established in south-western Australia highlight the potential of *Eucalyptus* plantations to
384 exceed the water limit early in the rotation on deep soils. This is an issue that warrants deeper understanding and
385 the development of management strategies. The results of the meta-analysis suggest that the average annual water
386 use by the two genera will be similar over large areas and long time periods (decades). They do not, however,
387 preclude periods of high-water use by *Eucalyptus*.

388

389 The range of annual rainfall, climate wetness indices and annual evapotranspiration in the published studies was
390 similar for the 27 *Eucalyptus* and 30 *Pinus* sites included in meta-analysis (Table 1, Table 2 and supplementary
391 material). Only a few sites had climate wetness indices more than 1.5. These were Jijou and Hetou in China,
392 Huape and Valdivia in central Chile and Coweeta in North Carolina. In the case of the Chinese sites, Lane et al.
393 (2004) and Ren et al. (2019) concluded that plantations of *Eucalyptus* would not have an important effect on water
394 resources nor on water security in this part of China. Notwithstanding this conclusion there is still a lot of
395 investment made to quantify to water use of *Eucalyptus* in these regions. Wherever the climate wetness index
396 exceeds 1.5 then the amount of streamflow will always be substantial, even in lower rainfall years (White et al.,
397 2016). Thus, rather than annual water balance, the focus should be on water quality and dry season flow to better
398 understand the effect of land use change, including the planting of *Eucalyptus*, on water security.

399

400 For the published *Eucalyptus* and *Pinus* studies analysed here, there was a strong positive correlation between
401 evapotranspiration and rainfall and therefore between the vegetation evaporation efficiency and the climate
402 wetness index (Figure 3). The coefficient, or 'catchment characteristic', c was greater in *Eucalyptus* (2.84) than
403 in *Pinus* (2.64) but the difference between the two genera was not statistically significant ($p=0.50$). When this
404 result was discussed with colleagues in the forestry sector or with people in the forest research community it met
405 with responses ranging from mild surprise to disbelief. The belief that *Eucalyptus* uses more water than any of
406 the alternative crops is very firmly entrenched even though it does not seem to have a firm scientific foundation.
407 Given that the meta-analysis presented in this paper produced a result that was counter to the prevailing view it is
408 very important to consider the direct and corroborative evidence that either support or oppose this observation.
409 The following paragraphs attempt to provide a mechanistic basis for the observation that, while under some
410 circumstances *Eucalyptus* can use water much more rapidly than *Pinus* (Scott and Lesch, 1997), the average
411 behaviour of the two genera appears similar (Benyon and Doody, 2015), Figure 3). This mechanistic basis is then
412 used to indicate under which circumstances the effects of plantations of *Pinus* or *Eucalyptus* species on water
413 resources should be evaluated and actively managed.

414

415 The key to understanding the patterns of water use in *Eucalyptus* and *Pinus* plantations lies in the hydraulic
416 architecture of the two genera and in the way that this affects the relationship between water use and carbon gain.
417 There are some consistent differences between the group of *Eucalyptus* and *Pinus* species that are grown in
418 commercial plantations. First, and very importantly, *Pinus* species are gymnosperms and their water conducting
419 elements are tracheids while in *Eucalyptus* water is transported in vessels. The maximum hydraulic conductivity
420 of angiosperms exceeds that of conifers with almost no overlap in the ranges (Sperry et al., 2006). It is the diameter
421 of the vessels that afford angiosperms greater maximum hydraulic conductance (Sperry et al., 2006). It is also
422 known that in the *Eucalyptus* genus vessel size, and maximum hydraulic conductivity of the xylem, is correlated



423 with climate wetness (Pfausch et al., 2016) so that the major plantation species can have hydraulic conductivities
424 among the highest in the plant kingdom. Leaf conductance and maximum photosynthetic capacity scale directly
425 with the hydraulic conductivity of the xylem (Hubbard et al., 2001; Tyree, 2003).

426

427 Thus, plantation *Eucalyptus* species, the most important of which are from the Symphyomyrtus subgenus and
428 grow naturally in the wetter fringes of the Australian continent, have higher maximum hydraulic conductivity,
429 water use and photosynthetic capacity than commercially grown *Pinus* species (Whitehead and Beadle, 2004). In
430 the early growth phase, Symphyomyrtus *Eucalyptus* species also have a much higher specific leaf area (ratio of
431 leaf area to mass) than *Pinus* and this results in more rapid canopy development and the potential for faster early
432 growth and water use such as observed by Scott and Lesch, (1997). This can of course only happen if there is
433 water available to support this growth and canopy development and this can be supplied by rainfall throughout
434 the year or by additional sources of water stored in deep soil (Dye and Olbrich, 1992; Scott and Lesch, 1997;
435 White et al., 2014), shallow groundwater (Benyon et al., 2006; Brooksbank et al., 2011; Eamus et al., 2000;
436 O'grady et al., 2011b) or applied as irrigation (Honeysett et al., 1996). If *Eucalyptus* plantations are grown on
437 deep soils and in regions where the climate wetness index is much less than one (potential evaporation exceeds
438 rainfall) then, by virtue of their hydraulic architecture, they have the potential to affect the water balance more
439 than alternatives.

440

441 The capacity of *Eucalyptus* for high instantaneous sap velocities that are associated with elevated photosynthetic
442 capacity also affects the seasonal patterns of water use in *Eucalyptus* compared to *Pinus*. Transpiration of
443 *Eucalyptus* species increases rapidly in spring associated with high maximum stem and leaf conductivity (White
444 et al., 1999). The relative behaviour of *E. globulus* and *P. radiata* is well understood making them good exemplars.
445 They are also two plantation species of great global importance that are grown in similar areas including in central
446 Chile. In Chile and Australia, *P. radiata* is known to be capable of surviving more severe droughts than *E. globulus*
447 and plantations of the species therefore extend into drier areas than *E. globulus* both in Chile and in Australia.
448 The greater drought tolerance of *P. radiata* than *E. globulus* is mediated by a much stronger stomatal response to
449 soil drying (Mitchell et al., 2014). In situations where the amount of soil water storage imposes an upper limit on
450 annual use then, while this store of water will be completely depleted by both species, it will be used earlier in the
451 growing season by *Eucalyptus*. Thus, the period of peak physiological activity and growth in *Eucalyptus* is
452 associated with lower average temperatures and more moderate air saturation deficits. This pattern of water use
453 biased towards spring and early summer can result in very efficient water-use growth and wood production (White
454 et al., 2015). This behaviour of the *Eucalyptus* is closer to a mimic of the seasonal water use pattern of an annual
455 species. This mechanism underlies the greater water use efficiency of *Eucalyptus* species than of the *Pinus* but is
456 also associated with an increased risk of mortality (White et al., 2003; White et al., 2009) if the soil water runs
457 out. It also underlies the high rates of water use sometimes observed on deep soils (Scott and Lesch, 1997).

458

459 At equilibrium *Eucalyptus* and *Pinus* species generally have different seasonal patterns of water use. Nonetheless,
460 the average annual water use does not differ significantly between the two genera amongst the published studies
461 presented in Figure 3. This observation is entirely consistent with the observed hydraulic architecture of these two
462 genera in the field. Radiation interception and absorption, and therefore productivity and evapotranspiration in



463 forests, including plantations, are strongly correlated with leaf area index. Battaglia et al. (1998) proposed that
464 after the canopy closes, plantations will arrive at an 'equilibrium' leaf area index that maximises the net primary
465 productivity. They further demonstrated that the value of this optimum leaf area index is strongly influenced by
466 the climate wetness; higher optimum values of leaf area index were observed in wetter situations. The value of
467 this 'optimum leaf area index' tends to be higher for a given climate wetness in *Pinus* species than in *Eucalyptus*
468 species. For those experiments included in this analysis that reported leaf area index, the average value for *Pinus*
469 was approximately 4, nearly a full unit greater than the average value for the *Eucalyptus* plantations.

470

471 In comparing *Eucalyptus* and *Pinus* in commercial plantations it is important to compare at least one and possibly
472 more, full crop rotations. *Pinus* is generally managed for solid wood production and therefore on a longer rotation
473 than *Eucalyptus* which is usually, but not exclusively, grown for pulpwood production. Around the world the time
474 from planting to harvest of *Pinus* species is between two and three times that of the *Eucalyptus* in the same
475 location. In Chile, for example, *Eucalyptus* is harvested after about 12 years while *Pinus* is grown for about 25
476 years. *Pinus* is usually grown for solid wood or veneer production and is therefore thinned at least once and is
477 often pruned to produce clear wood. After the harvesting of the first *Eucalyptus* crop, a *Pinus* plantation on the
478 same location would remain standing and operating at, or near, the water limit. For a period of between two and
479 three years after the *Eucalyptus* harvest the evapotranspiration of the *Pinus* will therefore exceed that of the
480 *Eucalyptus*. This is evident in the results of Scott and Lesch (1997) who compared *E. grandis* with *P. patula*. The
481 frequency of harvest of *Eucalyptus* will be a key factor affecting the comparative water balance of *Pinus* and
482 *Eucalyptus* plantations. Paradoxically, more frequent harvests will increase the average streamflow from
483 *Eucalyptus* plantations relative to *Pinus*. It has been demonstrated that the effects of thinning on the water balance
484 are transient, lasting for a maximum of one year in both *Pinus* and *Eucalyptus* (Scott and Lesch, 1997; White et
485 al., 2014).

486

487 The proportion of evapotranspiration that occurs as transpiration was approximately 0.5 for both *Pinus* and
488 *Eucalyptus* across a wide range of climate wetness indices (Figure 6). This means that the annual partitioning of
489 evapotranspiration to fluxes other than transpiration is similar for these two genera. The partitioning of these other
490 fluxes to understorey transpiration, soil evaporation and interception may have important implications for
491 ecosystem productivity and efficiency. The water use efficiency of wood production is directly correlated with
492 the ratio of transpiration to other fluxes (White et al., 2015). In a study that compared *E. globulus* and *P. radiata*
493 Benyon and Doody (2015) observed that interception was more than half the non-transpirational fluxes in *P.*
494 *radiata* and less than half in *E. globulus*. This variation in partitioning is a direct consequence of the previously
495 noted tendency for *Pinus* to have a higher leaf area index than *Eucalyptus* and the greater canopy storage per unit
496 leaf area in *Pinus* than in broadleaved species (Iida et al., 2005). A weakness of this analysis and of the literature
497 on water balance is the exclusion of stemflow from most water-balance studies. It is likely that stemflow will
498 contribute more to throughfall in *Eucalyptus* (7% of rainfall) than in *Pinus* (2 to 5%) (Crockford and Richardson,
499 1990). This difference is approximately equivalent in magnitude to the observed, albeit non-significant, difference
500 between the genera in this analysis.

501

502 **5. Conclusion**



503 Water use by vegetation is very situation specific. The comparison between *Eucalyptus* and *Pinus* depends on the
504 age of the plantation, the length of the rotation, the seasonality of rainfall and the depth of the soil. In this paper a
505 meta-analysis of published estimates of evapotranspiration by *Pinus* and *Eucalyptus* species in commercial
506 plantations did not find a significant difference between the genera. Specifically, while there was a small, but
507 systematic difference of about 3% in water use between the genera (see Figures 5 and 6), this analysis finds that
508 for a given climate wetness index the evapotranspiration by *Pinus* and *Eucalyptus* was statistically the same.
509 Moreover, our understanding of the hydraulic architecture and stomatal physiology of pines and eucalypts
510 suggests that, although the long-term average behaviour may be similar, there will be differences in their temporal
511 pattern of water use both within and between years. *Eucalyptus* will use more water than *Pinus* early in the
512 growing season and in the early years of the rotation. On deep soils this may result in lasting differences but under
513 most circumstances the total effect on water balance will be similar. The reputation of much higher water use by
514 *Eucalyptus* may stem partly from the observation of vigorous early growth of *Eucalyptus* and the many studies
515 on young plantation stands.

516 **Competing Interests**

517 From July 2015 to April 2020, Drs White and Silberstein were paid to provide advice to Bioforest SA on
518 Ecohydrology and Ecophysiology. Bioforest SA are an R and D company owned by Arauco, the largest plantation
519 grower in central Chile. In the course of this work Dr White has also received some financial support from the
520 Guangxi Forestry Research Institute in China.

521

522 **Code / Data Availability**

523 Provided as Supplementary Material

524

525 **Author Contributions**

526 Don A White – Conceptualization, Data Curation, Formal Analysis, Methodology, Validation, Original Draft
527 Preparation, Review and Editing

528 Shiqi Ren – Conceptualisation, Funding acquisition, Supervision

529 Daniel Mendham – Conceptualisation, Data Curation, Formal Analysis, Review and Editing

530 Francisco Balocchi-Contreras - Conceptualisation, Review and Editing

531 Richard Silberstein – Conceptualisation, Review and Editing

532 Andrés Iroumé – Conceptualisation, Validation

533 Pablo Ramirez de Arellano – Conceptualisation, Methodology, Project Administration, Supervision

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Table 1. Brief description of all the papers and the associated studies included in the meta-analysis. See the supplementary material for a full summary of the data used in the analysis. Data are sorted by Region and Annual Rainfall. The annual rainfall data provided here are measurements from the cited paper unless noted otherwise.

| Species | Region | Number of Sites | Rainfall Range (mm) | Climate Type | Number of Years Data | Method Used to Estimate ET | Reference |
|---------------------------------------|---------------------------------|-----------------|---------------------|----------------|----------------------|----------------------------|----------------------------|
| <i>E. urophylla</i> x <i>globulus</i> | Terra Dura, Brazil | 2 | 1433 - 1626 | Sub-Tropical | 12* | Method 3 | (Almeida et al., 2016) |
| <i>E. globulus</i> | Green Triangle, Australia | 3 | 489-701 | Cool Temperate | 3 to 4+ | Method 1 | (Benyon et al., 2006) |
| <i>E. globulus</i> | Portugal | 2 | 788 | Mediterranean | 9# | Method 3 | (David et al., 1994) |
| <i>E. globulus</i> | Tasmania, Australia | 1 | 975 | Cool Temperate | 4# | Method 2 | (Honeysett et al., 1996) |
| <i>E. nitens</i> | Tasmania, Australia | 1 | 960 | Cool Temperate | 4# | Method 2 | (Honeysett et al., 1996) |
| <i>E. urophylla</i> | Leizhou Peninsula, China | 2 | 1620-1920 | Tropical | 2+ | Method 1 | (Lane et al., 2004) |
| <i>E. grandis</i> | Northern Province, South Africa | 1 | 756 | Sub-Tropical | 9+ | Method 3 | (Lesch and Scott, 1997) |
| <i>E. urophylla</i> x <i>grandis</i> | Grao Mogol, Brazil | 1 | 1121 | Tropical | 2+ | Method 2 | (Lima et al., 1990) |
| <i>E. saligna</i> | Rio Grande do Sul, Brazil | 1 | 2088 | Sub-Tropical | 1+ | Method 3 | (Reichert et al., 2017) |
| <i>E. urophylla</i> | Guangxi, China | 1 | 1294 | Sub-Tropical | 1# | Method 1 | (Ren et al., 2019) |
| <i>E. nitens</i> | Tasmania, Australia | 4 | 1222-1259 | Cool Temperate | 1-3# | Method 1 | (Roberts et al., 2015) |
| <i>E. globulus</i> | South India | 1 | 1568 | Montane | 9* | Method 3 | (Samraj et al., 1988) |
| <i>E. grandis</i> | South Africa | 1 | 1163 | Sub-Tropical | 10# | Method 3 | (Scott and Lesch, 1997) |
| <i>E. urophylla</i> x <i>grandis</i> | Aracruz, Brazil | 1 | 1396 | Tropical | 1+ | Method 2 | (Soares and Almeida, 2001) |
| <i>E. globulus</i> | Arauco, Chile | 1 | 1395 | Mediterranean | 3 | Method 1 | (White et al., 2021) |
| <i>E. nitens</i> | Curanilahue, Chile | 2 | 1845 | Mediterranean | 3 | Method 2 | (Balocchi et al., 2020) |
| <i>E. globulus</i> | Nascimento, Chile | 2 | 1272 | Mediterranean | 8 | Method 1 | (Iroumé et al., 2021) |



| | | | | | | | |
|--|------------------------------------|---|-----------|----------------|------|----------|------------------------------|
| <i>P. taeda</i> and <i>P. palustris</i> | South Carolina, USA | 1 | 1319 | Sub-Tropical | 20+ | Method 3 | (Amatya et al., 2006) |
| <i>P. radiata</i> | New Zealand | 1 | 1554 | Cool Temperate | 27* | Method 3 | (Beets and Oliver, 2006) |
| <i>P. radiata</i> | Green Triangle, Australia | 4 | 600-724 | Cool Temperate | 4+ | Method 1 | (Benyon et al., 2006) |
| <i>P. radiata</i> | NE Victoria, Australia | 1 | 1400 | Cool Temperate | 1+ | Method 3 | (Bren and Hopmans, 2007) |
| <i>P. elliotii</i> | SE Queensland, Australia | 1 | 1284 | Sub-Tropical | 10+ | Method 3 | (Bubb and Croton, 2002) |
| <i>P. strobus</i> | North Carolina, USA | 1 | 2240 | Sub-Tropical | 2+ | Method 1 | (Ford et al., 2007) |
| <i>P. taeda</i> | Florida, USA | 2 | 1098-1175 | Tropical | 2-4# | Method 4 | (Gholz and Clark, 2002) |
| <i>P. radiata</i> | Central Chile | 4 | 1084-2081 | Mediterranean | 2-3+ | Method 1 | (Huber and Iroumé, 2001) |
| <i>P. radiata</i> | Western Cape, South Africa | 1 | 642 | Mediterranean | 11# | Method 3 | (Lesch and Scott, 1997) |
| <i>P. patula</i> | Natal, South Africa | 1 | 886 | Sub-Tropical | 11# | Method 3 | (Lesch and Scott, 1997) |
| <i>P. caribea</i> var <i>hondurensis</i> | Grao Mogol, Brazil | 1 | 1121 | Tropical | 3 | Method 2 | (Lima et al., 1990) |
| <i>P. elliotii</i> and <i>P. palustris</i> | North Carolina, USA | 2 | 883-1033 | Sub-Tropical | 4 | Method 4 | (Powell et al., 2005) |
| <i>P. radiata</i> | Central Tablelands, NSW, Australia | 1 | 738 | Cool Temperate | 16 | Method 3 | (Putuhena and Cordery, 2000) |
| <i>P. patula</i> | Northern Province, South Africa | 1 | 756 | Sub-Tropical | 17 | Method 3 | (Scott and Lesch, 1997) |
| <i>P. taeda</i> | North Carolina, USA | 1 | 1091 | Sub-Tropical | 4 | Method 4 | (Stoy et al., 2006) |
| <i>P. taeda</i> | North Carolina, USA | 1 | 1238 | Sub-Tropical | 4 | Method 4 | (Sun et al., 2010) |
| <i>P. radiata</i> | Constitucion, Chile | 1 | 1016 | Mediterranean | 1 | Method 1 | (White et al., 2021) |
| <i>P. radiata</i> | Arauco, Chile | 1 | 1395 | Mediterranean | 3 | Method 1 | (White et al., 2021) |
| <i>P. radiata</i> | Valdivia, Chile | 2 | 2210 | Mediterranean | 8 | Method 2 | (Balocchi et al., 2020) |



| | | | | | | | |
|-------------------|----------------------|---|------|---------------|---|----------|-------------------------|
| <i>P. radiata</i> | Nascimento, Chile | 2 | 1272 | Mediterranean | 8 | Method 1 | Iroumé et al. (2021) |
|-------------------|----------------------|---|------|---------------|---|----------|-------------------------|

*Full Rotation

+Post Canopy Closure Only

#Includes Pre and Post Canopy Closure

Table 2. Summary of the studies included in the meta-analysis (see Appendix for more details, and references for each study). This table indicates the number of studies included by country or continent, species, and climate zone.

| | | Eucalyptus | Pinus | Total |
|--|---------------------------|-------------------|--------------|--------------|
| Country/Continent | Australia and New Zealand | 9 | 8 | 17 |
| | United States | 0 | 8 | 8 |
| | South America | 10 | 11 | 21 |
| | South Africa | 2 | 3 | 5 |
| | China | 3 | 0 | 3 |
| | Europe | 2 | 0 | 2 |
| | India | 1 | 0 | 1 |
| | Total | 27 | 30 | 57 |
| Rainfall (mm) and Evapotranspiration (mm) | Min | 489 | 600 | |
| | Annual Rain | | | |
| | Median Annual Rain | 1259 | 1152 | |
| | Max Annual Rain | 2088 | 2240 | |
| | Min Annual ET | 488 | 355 | |
| | Median Annual ET | 940 | 927 | |
| | Max Annual ET | 1345 | 1291 | |

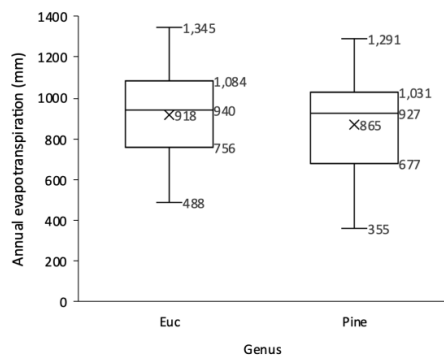


Figure 1. Box and whisker plots of annual evapotranspiration for the *Eucalyptus* and the *Pinus* sites. The three horizontal lines in the box show the median, 25th and 75th percentile values. The whiskers show the minimum and maximum values and the x indicates the mean. The associated labels indicate the actual values.

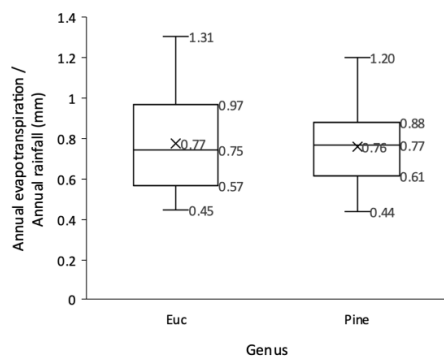


Figure 2. Box and whisker plots of the ratio of the evapotranspiration to rainfall for the *Eucalyptus* and the *Pinus* sites. The three horizontal lines in the box show the median, 25th and 75th percentile values. The whiskers show the minimum and maximum values, and the x indicates the mean values. The associated labels indicate the actual values. The mean ratio was 0.81 for *Eucalyptus* and 0.79 for *Pinus* while the medians for the same two genera were 0.77.

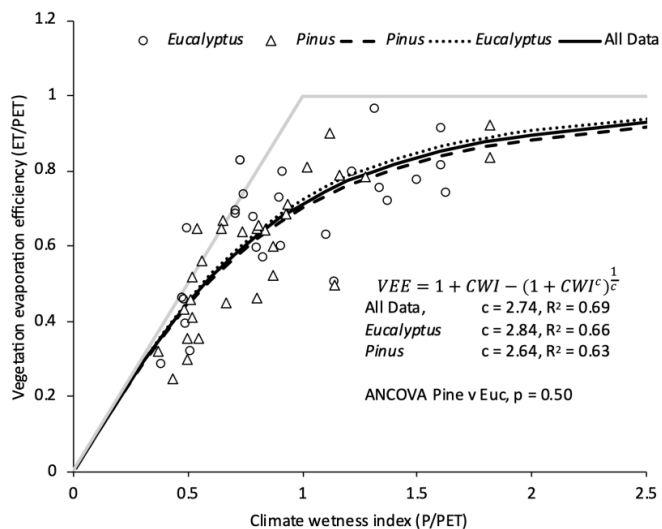


Figure 3. The vegetation evaporation efficiency as a function of the climate wetness index (a Budyko plot) for 57 (27 *Eucalyptus* and 30 *Pinus*) published studies. The solid grey lines are the water limit (evapotranspiration is equal to rainfall) and the energy limit (evapotranspiration is equal to potential evaporation). The dotted and dashed lines are for Equation 1 fitted separately to the data for *Eucalyptus* and *Pinus*.

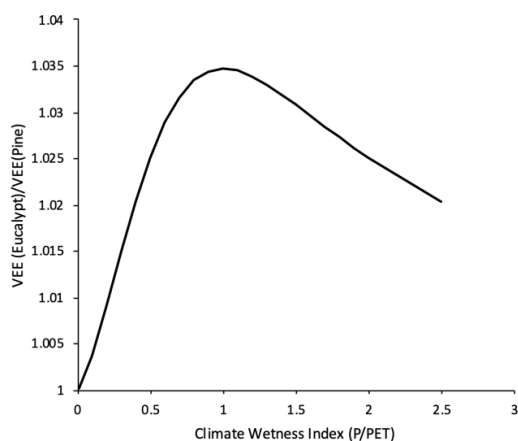


Figure 4. The ratio of the vegetation evaporation efficiency (VEE) for *Eucalyptus* to the vegetation evaporation efficiency for *Pinus* plotted as a function of the Climate Wetness Index. The vegetation evaporation efficiency was predicted using the separate relationships for the two genera in Figure 3.

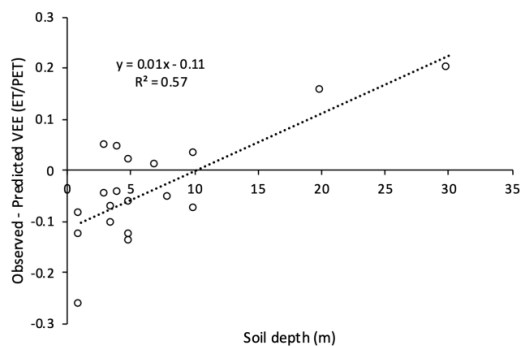


Figure 5. The residuals from Figure 4 for the *Eucalyptus* sites plotted as a function of soil depth. The model in Figure 4 with a value for c of 3.1 overestimates the observed value of k in shallow soils and underestimates k in deep soils.

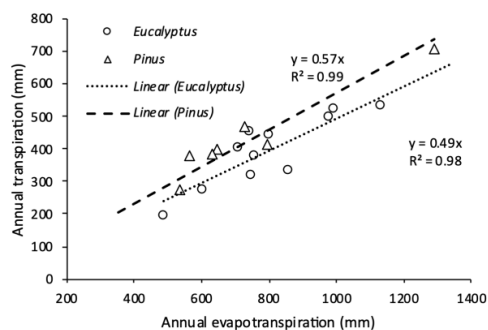


Figure 6. The relationship between annual transpiration and annual evapotranspiration for the subset of sites where transpiration was measured using sapflow sensors.