



1 **How useful are gridded water resources reanalysis and evapotranspiration**
2 **products for assessing water security in ungauged basins?**

3 Elias Nkiaka¹, Robert G. Bryant¹, Joshua Ntajal^{2,3}, Eliezer I. Biao⁴

4
5 ¹Department of Geography, University of Sheffield, Sheffield, S10 2TN, UK

6 ²Department of Geography, University of Bonn, 53115 Bonn, Germany

7 ³Center for Development Research, University of Bonn, 53113 Bonn, Germany

8 ⁴Laboratory of Applied Hydrology, University of Abomey-Calavi (UAC), Cotonou, Benin

9 Elias Nkiaka (Corresponding author): e.nkiaka@sheffield.ac.uk

10 Postal Address: Department of Geography, University of Sheffield, Sheffield, S10 2TN, UK

11 **Abstract**

12 Achieving water security in ungauged basins is critically hindered by a lack of in situ
13 hydrometeorological data to assess past, current and future evolution of water resources in those
14 areas. To overcome this challenge, there has been a shift toward the use of freely available satellite
15 and reanalysis hydrometeorological products. However, due to inherent bias and uncertainty, these
16 secondary sources require careful evaluation to ascertain their performance before being applied
17 in ungauged basins. The objectives of this study were to evaluate the performance of nine gridded
18 water resources reanalysis (WRR), and eight evapotranspiration (ET) products and to estimate
19 the relative uncertainties in monthly basin-scale water balance evapotranspiration (ET_{WB}) in eight
20 river basins located in Central-West Africa. Evaluation results highlight strengths and weaknesses
21 of the different WRR and ET products in simulating discharge dynamics and ET estimates
22 respectively across the basins. Analyses further revealed that the relative uncertainties in monthly
23 ET_{WB} range from 4–25 % with a significant increase in magnitude during the rainy season while
24 river discharge is the dominant source of uncertainty. Our results further revealed that the
25 performance of land surface models (LSMs) and global hydrological models (GHM) in
26 simulating river discharge is strongly influenced by the model structure, input data and spatial
27 resolution. Differences in ET estimates from the different ET products may be attributed to model
28 structure and the input data used in driving the models. Results from this study suggest that
29 gridded WRR and ET products are a useful source of data for assessing water security in
30 ungauged basins. However, given the plethora of products available, it is imperative to evaluate
31 their performance in representative gauged basins to identify products that can be applied in each
32 region.

33



34 **1. Introduction**

35 River discharge is one of the most important hydrological variables underpinning water resources
36 management, aquatic ecosystems sustainability, flood prediction, and drought warnings at
37 different scales (Mcnally et al., 2017; Couasnon et al., 2020). However, observed river discharge
38 data is often not available at the exact location where critical water management decisions need
39 to be made (Neal et al., 2009). This is especially the case in developing and semi arid/arid regions
40 where hydrometeorological gauging stations are sparse (Van De Giesen et al., 2014; Krabbenhoft
41 et al., 2022), while the number of existing stations is declining (Rodríguez et al., 2020). Despite
42 the acute shortage in observed data, developing regions are areas that are more vulnerable to
43 adverse hydroclimatological conditions (Byers et al., 2018; Kabuya et al., 2020). Furthermore,
44 achieving water security in ungauged basins in developing regions remains a critical development
45 challenge as climate change, population growth, rapid urbanization, and economic growth
46 continue to exert pressure on available water resources under hydrological uncertainty (Flörke et
47 al., 2018; Hirpa et al., 2019). This highlights the urgent need for more reliable data to better assess
48 past, current, and future evolution of water resources, and to predict extreme hydroclimatological
49 events so that better strategies can be put in place to enhance water management and mitigate the
50 impact of extreme events (Nkiaka et al., 2020; Slater et al., 2021). Water security in this study
51 refers to the availability of sufficient quantities of water for human use and ecosystem
52 sustainability.

53 Evapotranspiration (ET) is another important hydrological variable that represents the
54 linkage between water, energy and carbon cycles and ecosystem services and is the second largest
55 process in the hydrological cycle after precipitation (Zhang et al., 2019). Therefore, ET plays a
56 critical role in water availability at different scales. As such, accurate estimates of ET are also
57 crucial for water management operations such as basin-scale water balance estimation, irrigation
58 planning, estimating water footprint, and assessing the impact of climate change on water
59 availability. However, globally, in situ ET monitoring stations are also scarce while the existing
60 monitoring network cannot provide sufficient information on the temporal and spatial trends of
61 ET at large scales (Laipelt et al., 2021). ET data scarcity may therefore limit our ability to
62 understand changes in the hydrological cycle and water security in the context of environmental
63 change and hydrological uncertainty.

64 To enhance water security in ungauged basins, there has been a progressive shift toward
65 the use of gridded data derived from satellite and reanalysis (Odusanya et al., 2019; Nkiaka,
66 2022). This is because gridded data products can provide high spatial resolution and long-term
67 homogeneous data for previously unmonitored areas at scales that are suitable for studying



68 changes in the hydrological cycle and for water management applications (Sheffield et al., 2018).
69 Several gridded data products with global coverage have been produced in recent decades.
70 Examples of reanalysis products include Watch Forcing Data applied to ERA-Interim (Weedon
71 et al., 2014) and Climate Forecast System Reanalysis (Saha et al., 2014). There is also a plethora
72 of satellite products for different hydrometeorological variables such as precipitation,
73 temperature, soil moisture, and ET. For satellite derived ET estimates, it is worth noting that this
74 variable cannot be directly measured by satellites, but rather derived from physical variables
75 observed by satellites from space such as radiation flux. As such, satellite derived ET estimates
76 could rather be referred to as model outputs constrained by satellite data. Considering the way
77 gridded ET products are derived, they tend to suffer from large biases (Weerasinghe et al., 2020;
78 Mcnamara et al., 2021) and therefore need to be validated before use. In fact, it is argued that
79 validating gridded ET products is an essential step in understanding their applicability and
80 usefulness in water management operations (Blatchford et al., 2020).

81 Previously, much attention in the development of gridded environmental data was focused
82 on hydrometeorological variables such as precipitation and temperature. However, rapid
83 advancement in computer technology has led to the development of gridded water resources
84 reanalysis (WRR) with quasi global coverage using both land surface models (LSMs) and Global
85 Hydrological Models (GHMs) driven by satellite and reanalysis data. Examples of WRR products
86 include the Global Land Data Assimilation System [GLDAS] (Rodell et al., 2004), “The Global
87 Earth Observation for Integrated Water Resources Assessment” [earthH2Observe] (Schellekens et
88 al., 2017), and the Global Flood Awareness System [GloFAS-ERA5] (Harrigan et al., 2020).
89 Several studies have demonstrated that model-based gridded WRR products can be used as an
90 alternative to observe river discharge in ungauged basins to: (1) understand hydrological
91 processes (Koukoulou et al., 2020), (2) support transboundary water management (Sikder et al.,
92 2019), (3) identify flood events (Gründemann et al., 2018; López et al., 2020), and (4) support
93 national water policies (Rodríguez et al., 2020). These examples demonstrate that WRR products
94 have great potential for addressing water security challenges in ungauged basins. Despite their
95 numerous advantages, model outputs from WRR are also fraught with uncertainties resulting
96 from errors in the forcing data, model structure, and the parameterisation of the physical processes
97 in the model scheme (Koukoulou et al., 2020). Therefore, it is necessary to evaluate the
98 performance of these products against observed river discharge where available.

99 Whilst the use of outputs from WRR in water management has gained significant attention
100 in many ungauged areas such as Asia and Latin America (López et al., 2020; Rodríguez et al.,
101 2020; Sikder et al., 2019), they remain largely under-utilized in Africa. For example, there are



102 only a few case studies reporting on the use of these products in the Upper Blue Nile River basin
103 (Koukoulou et al., 2020; Lakew et al., 2020) and the Zambezi River basin (Gründemann et al.,
104 2018). Considering the scale of water insecurity in Africa -compounded by acute data scarcity
105 (Nkiaka et al., 2021), we feel that evaluating the performance of gridded WRR products in Africa
106 may enhance their adoption in water management in ungauged basins in the region. On the other
107 hand, several studies evaluating the performance of gridded hydrometeorological variables in
108 Africa have focused mostly on precipitation (Dinku et al., 2018; Satgé et al., 2020) while a few
109 studies that have evaluated gridded ET products focused on large basins, (Blatchford et al., 2020;
110 Weerasinghe et al., 2020; Mcnamara et al., 2021) and mostly adopting an annual timescale. This
111 may be attributed to the large scale of the basins which is ideal for the application of satellite data
112 and the coarse spatial resolution of some of the ET products. The availability of high spatial and
113 temporal resolution ET products means that it now possible to evaluate these products in small-
114 to medium-size basins and at a higher temporal resolution. Lastly, considering that the water
115 balance concept has been used widely to evaluate gridded ET products, most studies did not
116 account for uncertainties in basin-wide water balance evapotranspiration (ET_{WB}) even though
117 such uncertainties could be large (Baker et al., 2021). These are the key knowledge gaps that this
118 study will seek to address.

119 Focusing on eight basins of different sizes in Africa, the objectives of this paper were to:
120 (1) evaluate the performance of earth2Observe Tier 1 and other WRR products in simulating
121 discharge in the basins, (2) evaluate the performance of eight gridded ET estimates across the
122 basins and (3) estimate the relative uncertainties in ET_{WB} in the basins. Considering that only a
123 few studies have attempted to evaluate gridded WRR and ET products over Africa, this paper
124 contributes to the contemporary debate on the performance of these products and how there can
125 be used to assess water security in ungauged basins.

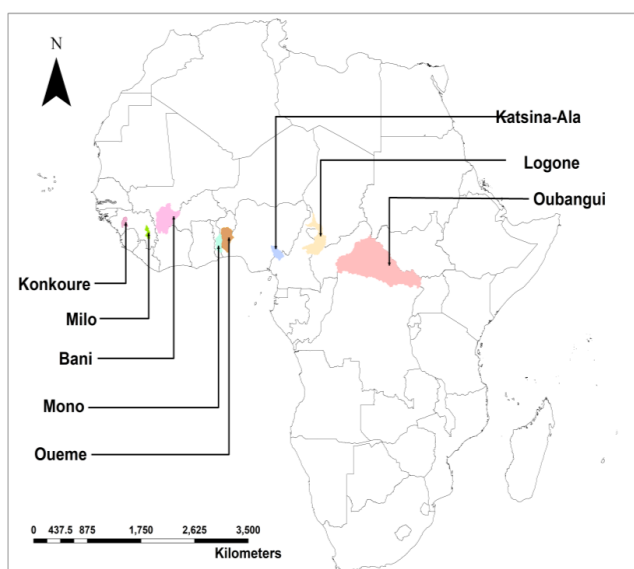
126 2. Materials and methods

127 2.1. Study area

128 The selected basins are located in Central-West Africa ranging in size from 9,000 km² to 499,000
129 km² (Figure 1). Rainfall in the region is mostly controlled by the north-south movement of the
130 intertropical convergence zone (ITCZ). The main criteria for selecting the basins were: (1)
131 availability of observed river discharge data and (2) for the period of the available discharge data
132 to coincide with the period when gridded WRR and ET data are also available. Additionally,
133 some of the selected basins are facing substantial water security challenges caused by population
134 displacement from conflicts in the Sahel and Lake Chad regions (Kamta et al., 2021; Nagabhatla
135 et al., 2021). The evaluation timestep was determined by the timestep of river discharge data.



136 Shapefiles for all the basins were obtained from HydroSHEDS, locations of the discharge gauging
 137 stations were obtained from the respective data sources while the area of each basin was
 138 calculated from the basin shapefiles. HydroSHEDS drainage network offers the unique
 139 opportunity to generate watershed boundaries for GRDC gauging stations using a proofed dataset
 140 and applying a consistent methodology. Table 1 shows that some of the basins are transboundary
 141 in nature.



142
 143 **Figure 1:** Locations of the eight river basins where the performance of WRR and gridded ET
 144 products were evaluated

145 **Table 1:** Characteristics of river basins and sources of river discharge data

| Basin | Total area (km ²) | Transboundary (Yes or No) Country(ies) | Population (thousands) | Source of river discharge data |
|-------------|-------------------------------|--|------------------------|--------------------------------|
| Bani | 101,600 | (Yes) Ivory Coast, Mali, and Burkina Faso | 63,766 | GRDC |
| Katsina-Ala | 22,963 | (Yes) Cameroon and Nigeria | 219,875 | NHSA |
| Konkoure | 10,250 | (No) Guinea-Conakry | 13,053 | GRDC |
| Logone | 87,953 | (Yes) Cameroon, Chad, and Central Africa Republic | 44272 | LCBC |
| Milo | 9,620 | (No) Guinea-Conakry | 13,053 | GRDC |
| Mono | 21,575 | (Yes) Togo, Benin | 21,479 | Co-author |
| Oubangui | 499,000 | (Yes) Central Africa Republic and the Democratic Republic of Congo | 88,742 | GRDC |
| Oueme | 46,990 | (No) Benin | 11,488 | Co-author |

146 Global River Discharge Centre [GRDC], Nigeria Hydrological Services Agency [NIHSA], Lake Chad Basin
 147 Commission [LCBC].

148 **2.2. Input data**

149 **2.2.1. Water resources reanalysis [WRR]**

150 The WRR product evaluated in this study include “The Global Earth Observation for Integrated
 151 Water Resources Assessment” (earth2Observe), Famine Early Warning Systems Network



152 [FEWS NET] Land Data Assimilation System (FLDAS), and TerraClimate. The earth2Observe
 153 Tier 1 product consists of a multi-model ensemble of ten global models at a spatial resolution of
 154 $0.5^\circ \times 0.5^\circ$ spanning from 1979 to 2012 and driven by Watch Forcing Data methodology applied
 155 to ERA-Interim reanalysis (WFDEI) data (Schellekens et al., 2017). The WRR from the
 156 earth2Observe project are freely available through the project data portal
 157 (<https://wci.earth2observe.eu/portal/>). Model evaluation here omits the Joint UK Land
 158 Environment Simulator (JULES), Simple Water Balance Model (SWBM), and the simple
 159 conceptual HBV hydrological model (HBV-SIMREG) as data from the models was not available
 160 from the data portal for the selected basins at the time of writing. As such, seven models and
 161 model ensemble were included in this study. Although there is an available Tier 2 product with a
 162 higher spatial resolution (0.25°), this study did not utilise these data as selected basins were not
 163 included at the time of conducting this research. We also evaluated the NOAH model from
 164 FLDAS with spatial resolution of 0.1° and runoff data from TerraClimate reanalysis with a spatial
 165 resolution of 0.041° . Table 2 provides a brief summary of the different models used in this study.

166 **Table 2:** Water resources reanalysis (WRR) evaluated

| Model provider | Model name | Model type | Routing scheme | Reference |
|--|---|------------|-------------------------------------|-------------------------------|
| CNRS (Centre National de la Recherche Scientifique) | ORCHIDEE (Organizing Carbon and Hydrology in Dynamic Ecosystems) | LSM | Cascade of linear reservoirs | (Krinner et al., 2005) |
| CSIRO (Commonwealth Scientific and Industrial Research Organization) | AWRA-L (Australian Water Resources Assessment) | GHM | Cascade of linear reservoirs | (Van Dijk et al., 2014) |
| ECMWF (European Centre for Medium-Range Weather Forecasts) | HTESSSEL (Hydrology Tiled ECMWF Scheme for Surface Exchanges over Land) | LSM | CaMa-Flood | (Balsamo et al., 2009) |
| JRC (Joint Research Centre) | LISFLOOD | GHM | Double kinematic wave | (Van Der Knijff et al., 2010) |
| UniUt (Universiteit Utrecht) | PCR-GLOBWB | GHM | Travel time | (Van Beek et al., 2011) |
| MeteoFr (Meteo France) | SURFEX | LSM | TRIP with stream | (Decharme et al., 2010) |
| UniK (Universitat Kassel) | WaterGAP | GHM | Manning–Strickler | (Wada et al., 2014) |
| NASA | NOAH | LSM | Soil-layer water and energy balance | (Menally et al., 2017) |
| University of California Merced | Water- Balance Model | GHM | Bucket type model | (Abatzoglou et al., 2018) |

167

168 2.2.2. Evapotranspiration products

169 The gridded ET products evaluated in this study include FLDAS, GLEAM3.5a & 3.5b,
 170 MODIS16A2, PMLV1, PMLV2, SSEBop, and TerraClimate (see Table 3). Data from the ET



171 products are freely available with a global coverage except for FLDAS, which covers only the
172 African domain. Although the gridded ET products all have different spatial resolutions, we did
173 not resample the data to the same resolution because a previous study has shown that resampling
174 does not have any significant impact on the results (Weerasinghe et al., 2020). We also leveraged
175 on the power of cloud computing by downloading data for some ET products using the climate
176 engine research App. (www.climateengine.com). Table 3 provides a summary of all ET products
177 evaluated in this study.

178 **Table 3: Summary of the characteristics of the different ET products**

| ET product | Core equation | Temporal resolution | Spatial resolution | References |
|---------------------|-----------------------------|---------------------|--------------------|------------------------------------|
| FLDAS | Penman–Monteith | Daily | 0.1° x 0.1° | (Mcnally et al., 2017) |
| GLEAM3.5a & 3.5b | Priestley–Taylor | Monthly | 0.25° x 0.25° | (Martens et al., 2017) |
| MODIS16A2 | Penman–Monteith | 8-day | 1/48°x1/48° | (Mu et al., 2007; Mu et al., 2011) |
| PMLV1 | Penman–Monteith– Leuning | Monthly | 0.5° x 0.5° | (Zhang et al., 2016) |
| PMLV2 | Penman–Monteith– Leuning | 8-day | 1/192°x1/192° | (Zhang et al., 2019) |
| SSEBop | Surface Energy Balance | Monthly | 1/96° x 1/96° | (Senay et al., 2013) |
| TerraClimate | Penman–Monteith | Monthly | 1/24° x 1/24° | (Abatzoglou et al., 2018) |

179

180 **2.3. Evaluation data**

181 **2.3.1. River discharge**

182 Observed river discharge data were used to evaluate the performance of WRR models and to
183 estimate basin-wide water balance evapotranspiration (ET_{WB}) using the water balance concept.
184 The source of the river discharge data is available in Table 1. Gaps in the discharge data were
185 filled using Self-Organizing Maps which have been shown to be a robust method for infilling
186 missing gaps in hydrometeorological time series (Nkiaka et al., 2016).

187 **2.3.2. Precipitation**

188 Climate Hazards Group InfraRed Precipitation with Station data (CHIRPS) was used in this study
189 to estimate ET_{WB} . CHIRPS has a quasi-global coverage at a spatial resolution of 0.05° x 0.05°,
190 spanning the period from 1981 to the present at a daily timescale (Funk et al., 2015). The dataset
191 was explicitly designed taking into consideration the weaknesses of existing products (Sulugodu
192 et al., 2019). As such, CHIRPS blends gauge and satellite precipitation covering most global land
193 regions, it has low latency, high resolution, low bias, and long period of record (Funk et al., 2015).
194 CHIRPS has extensively been validated (Dinku et al., 2018; Satgé et al., 2020) and used in several
195 studies in Africa (Larbi et al., 2021; Nkiaka, 2022). The data was downloaded as the spatial
196 average for each basin using the climate engine App and used to estimate ET_{WB}



197 **2.3.3. GRACE**

198 GRACE data are monthly anomalies of terrestrial water storage changes (TWSC) used to quantify
199 changes in terrestrial water storage. The dataset has a global coverage spanning the period 2003–
200 2017 (Tapley et al., 2019). The data was derived from Jet Propulsion Laboratory (JPL) RL06M
201 Version 2.0 GRACE mascon solution at a spatial resolution of $0.5^\circ \times 0.5^\circ$. The data has a coastline
202 resolution improvement (CRI) filter to reduce leakage errors across coastlines and land-grids,
203 using scaling factors derived from the community land model (Wiese et al., 2016). GRACE data
204 has recently been re-processed to reduce measurement errors and represents a new generation of
205 gravity solutions that do not require empirical post-processing to remove correlated errors, as
206 such, the present data is better than the previous GRACE version that was based on spherical
207 harmonic gravity solution (Wiese et al., 2016). GRACE data was used in this study to estimate
208 ET_{WB} following the approach used in several other studies e.g., (Andam-Akorful et al., 2015; Liu,
209 2018; Xie et al., 2022).

210 **2.4. Evaluating gridded WRR**

211 WRR models were evaluated following a multi-objective approach commonly used in evaluating
212 the performance of hydrological models, including the Nash-Sutcliffe efficiency (NSE), Kling-
213 Gupta efficiency (KGE), and the percent bias (PBIAS). NSE scores range from $-\infty$ to 1, with 1
214 indicating a perfect representation of observed discharge. NSE scores ≥ 0.50 can be considered
215 acceptable whereas NSE scores ≤ 0.0 indicate poor model performance (Moriassi et al., 2007).
216 Similar to NSE, the KGE is a dimensionless metric that can be decomposed into three components
217 that are crucial for evaluating hydrological model performance accounting for temporal dynamics
218 (correlation), bias errors (observed vs simulated volumes), and variability errors (relative
219 dispersion between observations and simulations) (Gupta et al., 2009). KGE scores also range
220 from $-\infty$ to 1, with 1 considered the ideal value. Next, PBIAS is used to measure the tendency of
221 the simulated discharge to be larger or smaller than their observed counterparts (Gupta et al.,
222 2009). PBIAS is expected to be 0.0, with low magnitude values indicating accurate simulations,
223 positive values indicate underestimation, negative values indicate overestimation (Moriassi et al.,
224 2007). According to Moriassi et al. (2007), a hydrological model with PBIAS values in the range
225 $\pm 25\%$ can be considered to be acceptable. Furthermore, a temporal evaluation of flow
226 hydrographs was carried out by plotting the monthly simulated vs observed discharge to ascertain
227 visually if the models were able to capture the magnitude, seasonality, and interannual variability
228 of discharge.

229



230 **Table 4: Contingency table for 80th percentile river discharge**

| | Observed discharge | | |
|---------------------|--------------------|------------|-------------------|
| | | Yes | No |
| Simulated discharge | Yes | Hits (H) | False Alarms (FA) |
| | No | Misses (M) | Correct Negatives |

231

232 Lastly, we evaluated the models ability to predict discharge above specific thresholds. This
 233 evaluation step is of critical importance when considering operational water management
 234 requirements such as water allocation and reservoir operation which rely on monthly river
 235 discharge. To achieve this, we adopted the Critical Success Index (CSI) as the metric to evaluate
 236 the ability of each model to simulate discharge exceeding the 20th and 80th percentiles. CSI is
 237 calculated from a two-dimensional contingency table defining the events in which observed and
 238 simulated discharges exceed a given threshold (Thiemig et al., 2015). We used the 20th and 80th
 239 percentiles to assess the ability of the models to simulate both low and high flows respectively.
 240 The contingency table (Table 4) is a performance measure used in summarizing all possible
 241 forecast-observation combinations such as hits (H; event forecasted and observed), misses (M;
 242 event observed but not forecasted), false alarms (FA; event forecasted but not observed) and
 243 correct negatives (CN; event neither forecasted nor observed). The ideal value for CSI is 100%
 244 and the metric is calculated as follows:

245
$$CSI = \frac{H}{H + M + FA} \times 100 \quad (1)$$

246 **2.5. Evaluating gridded ET**

247 We also adopted a multi-step approach to evaluate the performance of ET products by assessing
 248 the annual ET–precipitation ratio, evaluating the statistical performance of ET products against
 249 long-term ET_{WB} and the ability of the products to capture monthly ET variability.

250 In the first step, the annual ET–precipitation ratio was calculated to compare with ratio
 251 obtained from ET_{WB} . The ET–precipitation ratio can also provide an estimate of the amount of
 252 water available in each basin after evapotranspiration losses. In the second step, different
 253 statistical metrics were used to assess the performance of the ET products using the monthly
 254 ET_{WB} as a reference (Andam-Akorful et al., 2015; Burnett et al., 2020; Koukoulou et al., 2020).
 255 The monthly ET_{BW} was calculated using the basin water balance equation as follows:

256
$$ET_{WB} = P - Q - \Delta S \quad (2)$$

257 Where P is average monthly precipitation over the basin (mm), Q is river discharge (mm) and ΔS
 258 is the terrestrial water storage change [TWSC] (mm). Unlike several studies that have evaluated



259 ET products on an annual timescale, this study adopts a monthly sample. As such, the TWSC
260 component (ΔS) in equation 2 that is often neglected when estimating ET_{WB} over several years
261 (≥ 10 years) could not be overlooked. Due to the likely impact of anthropogenic activities such as
262 reservoir operation, water withdrawal, and monthly rainfall variability on TWSC, values derived
263 at monthly timescales are important. In this case, TWSC data used in this study were obtained
264 from GRACE.

265 Due to the coarse spatial resolution of GRACE, it has been argued that GRACE is not
266 sensitive at detecting changes in monthly TWSC in small-size basins $\leq 150,000$ km² (Rodell et
267 al., 2011). Based on this claim, it might be argued that GRACE data may not be applicable in this
268 study considering that most of the basins are below this threshold except the Oubangui (499,000
269 km²). However, several studies (Liu, 2018; Biancamaria et al., 2019; Oussou et al., 2022; Xie et
270 al., 2022), have demonstrated that GRACE can provide acceptable TWSC estimates for basins
271 that are smaller than this threshold. Encouraging results from these and other studies do therefore
272 suggest that GRACE data can be used in this study; albeit with the expectation of considerable
273 uncertainties in TWSC estimates. For this study, GRACE data for each basin were obtained by
274 averaging the timeseries of all coincident GRACE grid cells. To estimate changes in monthly
275 TWSC, we calculated the difference between consecutive GRACE measurements for each basin,
276 divided by the time between measurements, using the following equation:

$$277 \quad \Delta S = (S_{[n]} - S_{[n-1]})/dt \quad (3)$$

278 where ΔS represents the TWSC (mm), n is the measurement number, and dt is the time difference
279 between two consecutive GRACE measurements (months).

280 Lastly, temporal evaluation of the products was carried out by plotting the time series of
281 all ET products against ET_{WB} to visually establish if the gridded ET products were able to capture
282 the magnitude, seasonality, and interannual variability of ET across the basins.

283 **2.6. Estimating relative uncertainty in basin-scale water balance ET (ET_{WB})**

284 To estimate the relative uncertainty in monthly ET_{WB} , we first calculated the absolute uncertainty
285 in monthly ET_{WB} by propagating errors through each of the components in equation 2 (Rodell et
286 al., 2011), as follows:

$$287 \quad \sigma_{ET} = \sqrt{\sigma_P^2 + \sigma_Q^2 + \sigma_{\Delta S}^2} \quad (4)$$

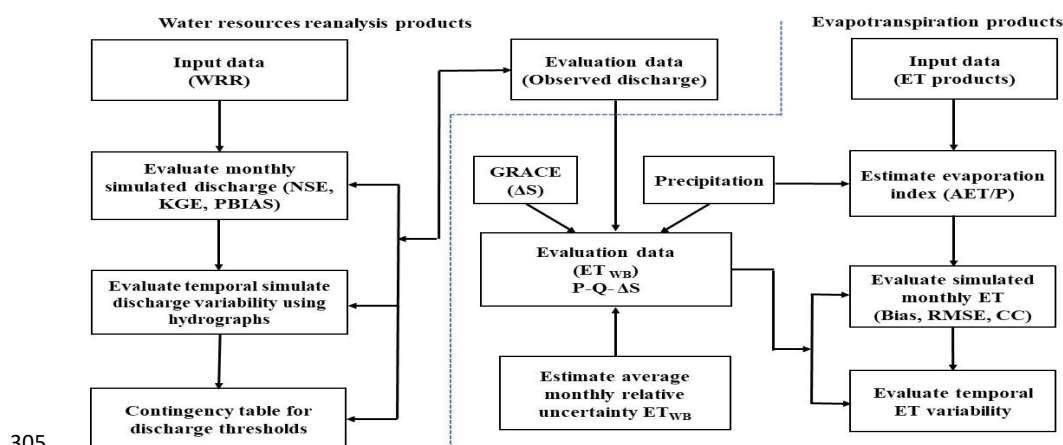
288 Where σ_P , σ_Q and $\sigma_{\Delta S}$ represent the absolute uncertainties in basin precipitation, observed river
289 discharge, and TWSC respectively. Uncertainty in precipitation was estimated as systematic



290 errors (bias). For this, we used a value of 2 % estimated for CHIRPS data at monthly timescale
 291 from 1981–2016 over Africa from a validation study using the Global Precipitation Climatology
 292 Centre (Shen et al., 2020). Uncertainty in TWSC was determined using the gridded fields of
 293 measurement and leakage errors (residual errors after filtering and rescaling) that are provided
 294 with the GRACE data. The uncertainty for each basin was calculated by averaging the values of
 295 all GRACE grid cells within each basin. To account for month-to-variation in equation 3, the
 296 TWSC error values were multiplied by $\sqrt{2}$ to obtain $\sigma_{\Delta S}$ (Andam-Akorful et al., 2015). Because
 297 no uncertainty estimates were provided with the river discharge data, we adopted a value of 20
 298 % which has been used in a recent study in the region (Burnett et al., 2020). After calculating the
 299 absolute uncertainty in monthly ET_{WB} , the relative monthly uncertainty was calculated using
 300 equation 4 (Baker et al., 2021) as follows:

$$301 \quad vET = \frac{\sigma ET}{ET_{WB}} \times 100 \quad (5)$$

302 Where vET is the monthly relative uncertainty (%), σET is the absolute monthly uncertainty
 303 (mm), and monthly ET_{WB} (mm). Figure 2 shows a flowchart WRR and ET products evaluation
 304 steps.



305
 306 **Figure 2:** Flowchart outlining the steps used in evaluating the WRR and ET products (The blue
 307 dotted line in the flow chart separates evaluation of WRR from ET products)

308 3. Results

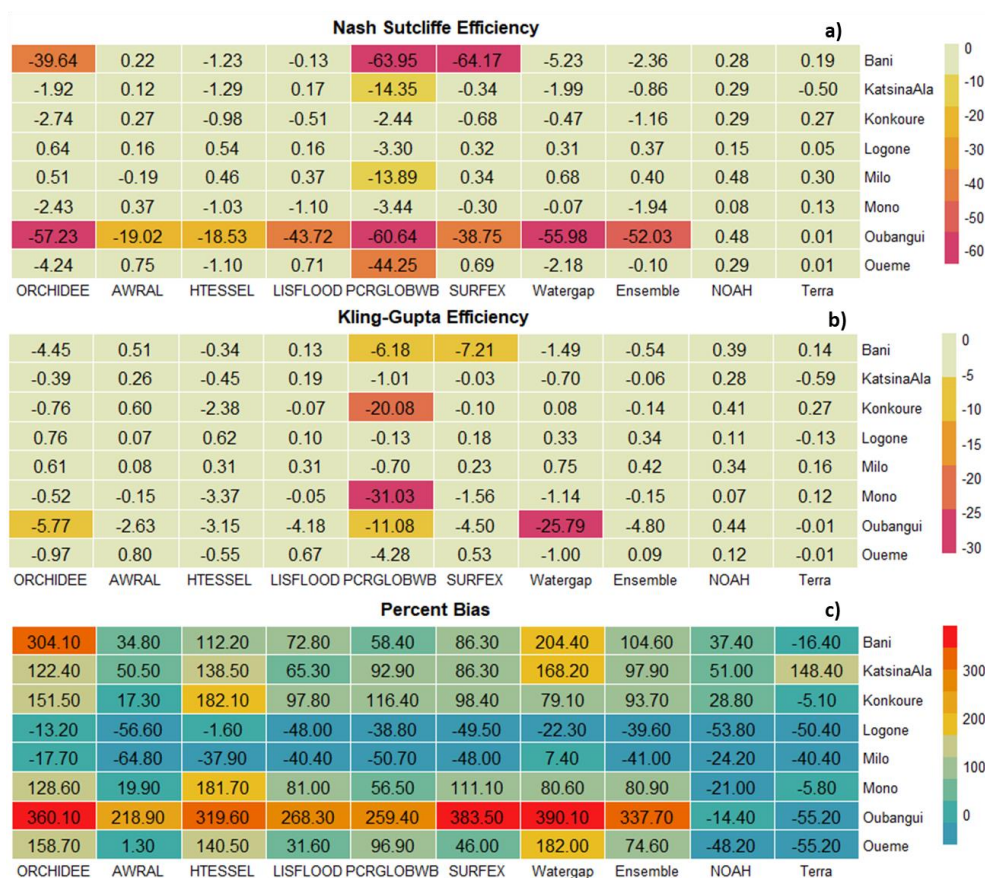
309 3.1. Water resources reanalysis products

310 3.1.1. Hydrological performance

311 A multi-objective approach using different statistical metrics (NSE, KGE and PBIAS) was used
 312 to evaluate the models in WRR Tier 1. The performance of the models in simulating river



313 discharge is shown in Figure 3. Using the NSE as a performance metric, results show that NOAH
 314 produced positive scores in all the basins (0.15–0.48). Terra, AWRAL and Lisflood models
 315 produced positive scores (0.01–0.75) in seven, six and four basins respectively. SURFEX model
 316 produced positive scores in three basins while ORCHIDEE, HTESSSEL, Watergap and the
 317 ensemble mean produced positive scores in two basins each. PCR-GLOBW produced negative
 318 scores in all the basins (Figure 3a).



319 **Figure 3:** Statistical evaluation of the models using (a) NSE, (b) KGE, and (c) PBIAS. Red and
 320 orange colours represent poor model performance in Figures 3a, 3b & 3c, however, the acceptable
 321 PBIAS range in Figure 3c is ± 2 %. Ensemble refers to the mean of WRR from the
 322 earthH2Observe product.
 323

324 Results of the KGE show that NOAH also produced positive scores (0.11–0.44) in all basins,
 325 followed by AWRAL, Lisflood and Terra models with positive scores in six, five and four basins
 326 respectively (Figure 3b). SURFEX and Watergap produced positive scores in three basins
 327 ORCHIDEE while HTESSSEL produced positive scores (0.31–0.76) in two basins, the ensemble

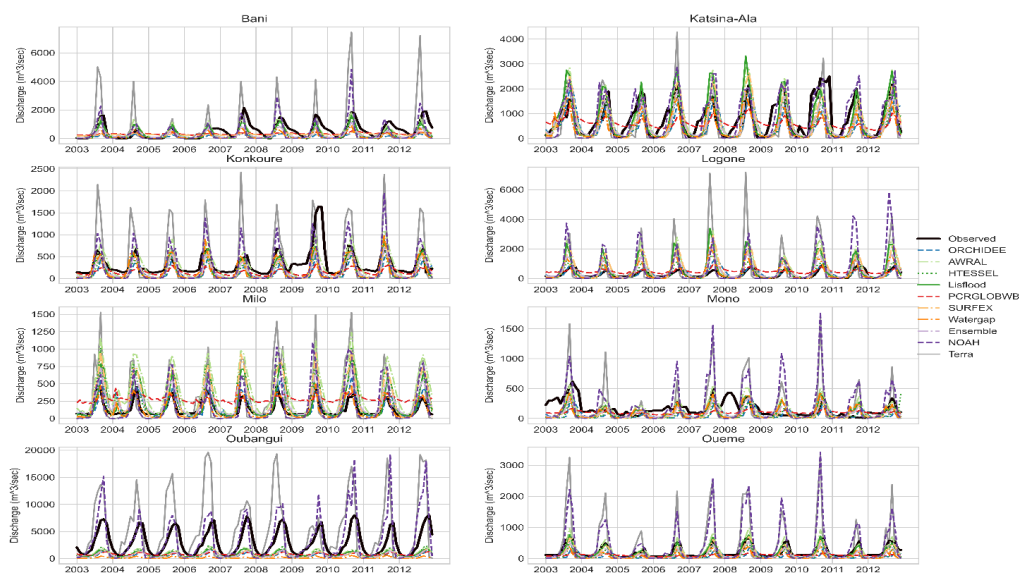


328 mean produced positive scores (0.09 – 0.42) in three basins. PCRGLOBW produced the worse
329 KGE scores (Figure 3b).

330 Positive and negative PBIAS values were obtained in the different basins. Negative values
331 indicate that the model overestimated discharge volumes compared to observed discharge while
332 positive values indicate the opposite. NOAH, Terra and AWRAL produced acceptable PBIAS
333 scores ($\pm 25\%$) in three basins, ORCHIDEE and Watergap produced similar scores in two basins
334 and HTESSEL in one basin (Figure 3c). The rest of the models including the ensemble mean
335 either grossly overestimated or underestimated discharge volumes in all the basins.

336 3.1.2. Temporal evaluation

337 The ability of the models to capture discharge variability was analysed by comparing the
338 simulated vs observed discharge in all the basins. Results show that most models were able to
339 capture the seasonal discharge variability including peak and low flows (Figure 4). However,
340 PCR-GLOBW systematically overestimated low flows and underestimated high flows across all
341 basins. In the Oubangui basin, all models were able to capture the seasonal variability but
342 consistently underestimated peak flows except NOAH and Terra models which both
343 overestimated peak flows (Figure 4). For example, peak discharge in the river exceeds 5000
344 m^3/sec , but all models except NOAH and Terra simulated this peak discharge to be less than 2000
345 m^3/sec (Figure 4).



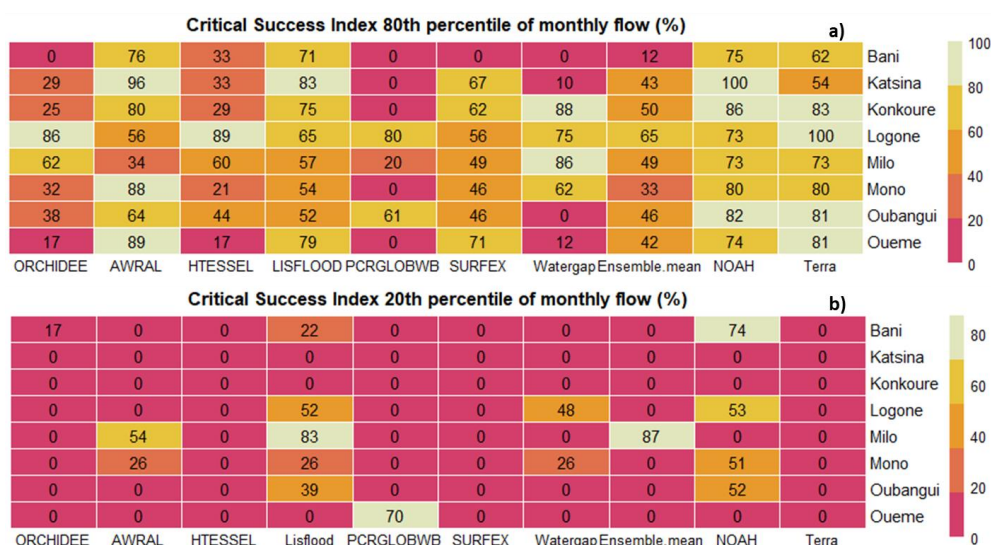
346

347 **Figure 4:** Evaluation of temporal flow variability simulated by the different model



348 **3.1.3. Critical Success Index**

349 Figure 5 shows the performance of the models in simulating the 80th and 20th percentiles
 350 monthly discharge. For the 80th percentile flows, results show that NOAH and Terra produced
 351 CSI scores above 50 % in all basins followed by Lisflood and AWRAL in seven and six basins
 352 respectively while Surfex and Watergap produced similar scores in four basins each (Figure
 353 5a). For the 20th percentile flows, only NOAH produced CSI scores above 50 % in four basins
 354 while Lisflood produced similar scores in two basins. The performance of the other models in
 355 simulating the 80th percentile flow shows a large spread while most models including the
 356 ensemble mean failed to simulate the 20th percentile flow across all the basins. Taking together,
 357 results suggest that the models simulated high flows better than the low flows with only the
 358 NOAH model capable of capturing both flow regimes in most basins (Figure 5b).



359 **Figure 5:** Critical Success Index for 80th and 20th percentile of monthly flow across all basins
 360

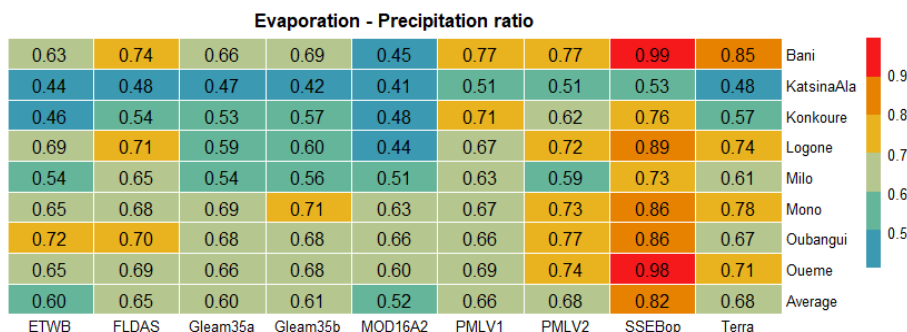
361 **3.2. Evapotranspiration products**

362 **3.2.1. Evapotranspiration–precipitation ratio**

363 Figure 6 shows the annual ET–precipitation ratio for all basins. It can be observed that average
 364 annual ET–precipitation ratio ranges between 0.52–0.82 over a period of 10 years (2003–2012)
 365 across all basins. SSEBop produced the highest ET–precipitation ratios (0.53–0.99) while
 366 MOD16A2 produced the lowest ratio (0.41–0.66) (Figure 5). Results show that the evaporation
 367 ratios from most of the ET products are in the same order of magnitude with the ratio from



368 ET_{WB} across all the basins with the only exception being SSEBop and MOD16A2 which
 369 respectively overestimated and underestimated this value.

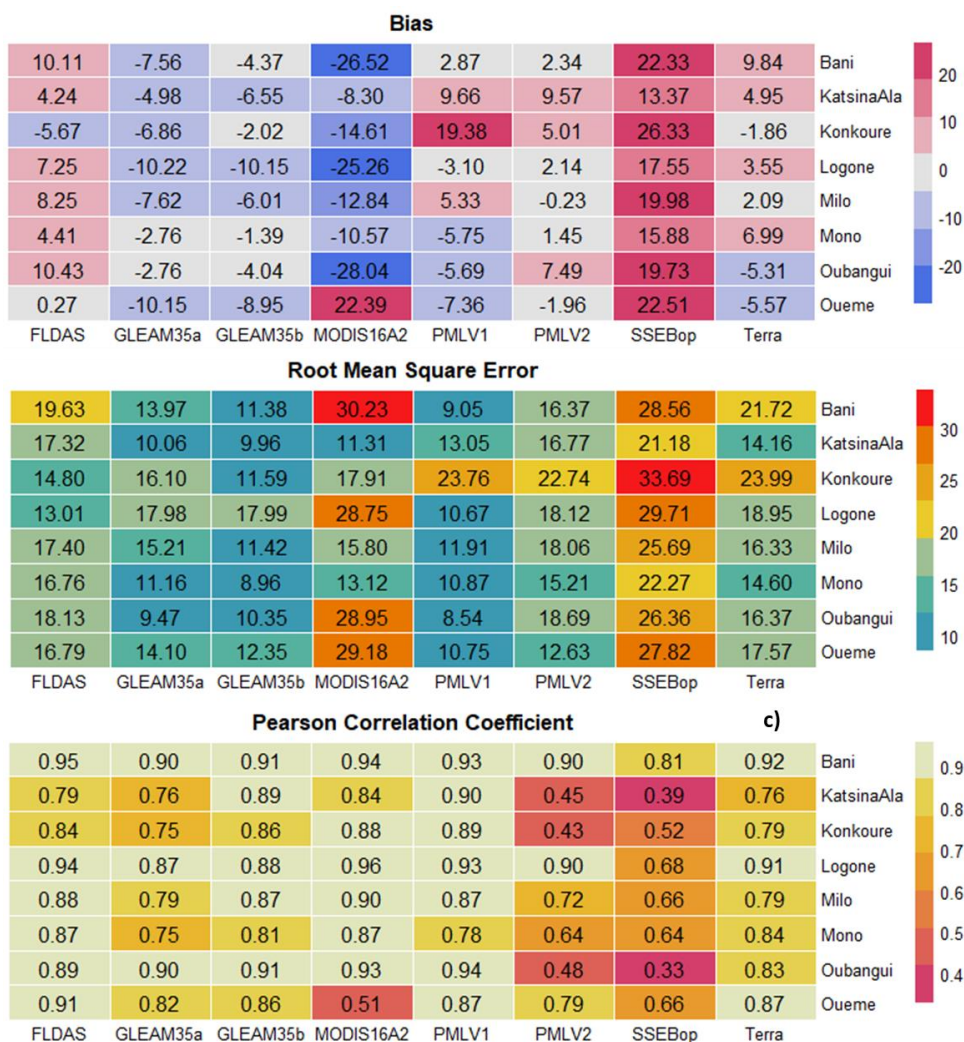


370
 371

Figure 6: Annual evapotranspiration – precipitation ratio 2003 – 2012

372 3.2.2. Basin-wide water balance estimates

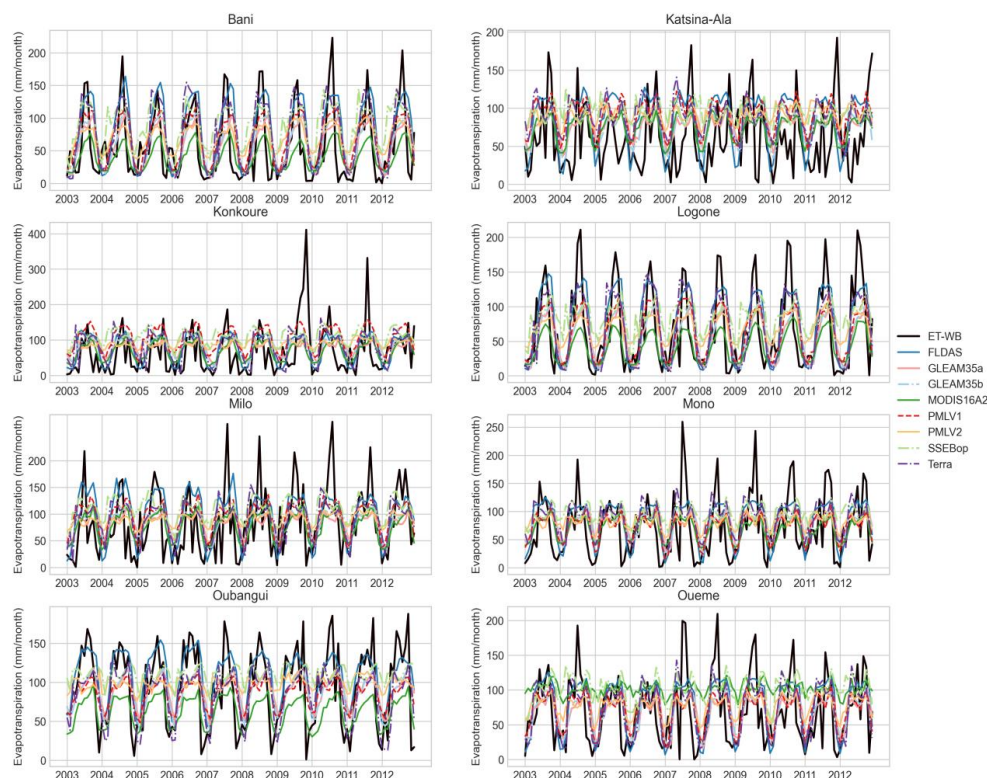
373 Figure 7 shows the results of the statistical metrics used in evaluating the ET products using
 374 monthly ET_{WB} as a reference. Considering bias as a performance metric, several products e.g.,
 375 FLDAS, PMLV2, Terra, and GLEAM3.5a & 3.5b produced low bias scores ranging from -6 to
 376 11 mm/month. However, GLEAM products systematically underestimated monthly ET with
 377 respect to ET_{WB} in all the basins while FLDAS, Terra and PMLV2 produced mixed results
 378 (7a). While SSEBop systematically overestimated monthly ET in all the basins, MODIS16A2
 379 underestimated this variable in all but one basin with respect to monthly ET_{WB} (Figure 7a). The
 380 lowest bias values ranging from -8.30 to 13.37 mm/month were obtained in the Katsina-Ala
 381 basin while the highest bias values ranging from -14.61 to 26.33 mm/month were recorded in
 382 the Konkoure basin.



383

384 **Figure 7:** Bias, RMSE, and Pearson correlation coefficient between monthly ET_{WB} and
 385 different ET products.

386 GLEAM3.5a & b produced the lowest RMSE (9.47–18 mm/month), followed by FLDAS (13–
 387 20 mm/month) and PMLV1 (8.50–12 mm/month) with this score exceeding 20 mm/month in
 388 only one basin. The rest of the ET products produced substantially higher RMSE scores with
 389 SSEBop and MODIS16A2 producing the highest RMSE scores (Figure 7b). Most ET products
 390 produced high Pearson correlation scores (≥ 0.75) with respect to ET_{WB} in all basins except
 391 PMLV2 and SSEBop which both produced low scores (< 0.50) in three and two basins
 392 respectively (Figure 7c).



393

394 **Figure 8:** Seasonal cycle of ET products and basin-wide water balance evapotranspiration.
395 ET_{WB} represents monthly evapotranspiration estimated by the water balance method, while the
396 rest are model-derived ET products.

397

3.2.3. Monthly ET variability

398 Figure 8 shows the seasonal cycle of ET_{WB} against the ET products for all basins. It can be
399 observed that most products were able to replicate the seasonal ET cycle across the basins.
400 However, most ET products underestimated monthly ET compared to ET_{WB} during the rainy
401 season with MOD16A2 producing the poorest results. Furthermore, most products were not
402 able to replicate the high peaks produced by ET_{WB} during the rainy season,.

403

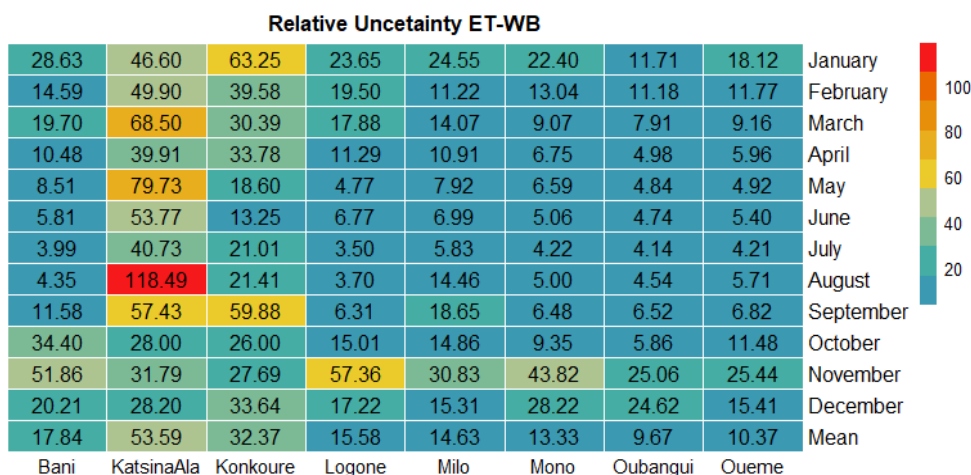
3.2.4. Estimating relative uncertainty in ET_{WB}

404 An assessment of absolute uncertainties in monthly ET_{WB} indicated that the dominant sources
405 of uncertainty vary from one basin to another and by each month. For example, in the Katsina-
406 Ala, Konkoure, and Milo basins, the dominant source of uncertainty in monthly ET_{WB} was
407 from river discharge (**Appendix A**). Although the absolute uncertainty in precipitation and
408 TWS also appears to be substantial in the three river basins, the uncertainty in river discharge



409 takes precedence over the other sources of uncertainty due to its higher magnitude. On the
 410 contrary, the dominant source of uncertainty in ET_{WB} in the Bani, Logone, and Oubangui basins
 411 was from TWSC. It can also be observed across the basins that there was no significant
 412 variation in monthly TWSC uncertainty which is consistent with the results of a similar study
 413 in the Amazon basin (Baker et al., 2021). Results also revealed that the magnitude of TWSC
 414 uncertainty were similar across the basins irrespective of the basin size (**Supplementary**
 415 **material**).

416 Figure 9 shows the relative uncertainty in ET_{WB} across all the basins. It can be observed
 417 that relative uncertainty values are generally <30 % but vary from month to month. However,
 418 the values were exceptionally high in the Katsina-Ala and Konkoure basins. The relative
 419 uncertainty in ET_{WB} also appears to be exceptionally high in the months of September–
 420 November which corresponds to high flow season. Taking together, the average monthly
 421 relative uncertainty in ET_{WB} for all basins ranges from 10–18% except in the Katsina-Ala and
 422 Konkoure basins where this range is grossly exceeded.



423 **Figure 9:** Average (2003 – 2012) monthly relative uncertainty in monthly ET_{WB} (%)
 424

425 4. Discussion

426 The overarching goal of this paper was to assess the performance of gridded water resources
 427 reanalysis and evapotranspiration products and to estimate the relative uncertainty in monthly
 428 basin-wide evapotranspiration (ET_{WB}) estimates. Below we provide a discussion and
 429 implications of our results in water security assessment in ungauged basins.

430



431 **4.1. Water resources reanalysis**

432 The performance of WRR products was assessed through commonly used model evaluation
433 metrics, discharge variability, and verification skill scores (critical success index) using
434 observed river discharge data. Our results show strong differences in the performance of the
435 different models in simulating river discharge across the basins. NOAH model produced
436 positive NSE and KGE values in all basins and PBIAS values within the acceptable range
437 ($\pm 25\%$) in three basins. Temporal evaluation of the WRR products showed that NOAH, Terra,
438 AWRAL and Lisflood were able to capture the seasonal variability in discharge as
439 demonstrated by high KGE scores. Indeed, high KGE values suggest that some models were
440 able to capture the temporal dynamics (strong correlation), and low bias scores indicate that
441 the variability errors between the observed discharge and simulation was also low (Gupta et
442 al., 2009). Nevertheless, Terra consistently overestimated peak flows in all the basins.

443 Apart from NOAH model which is a LSM used in FLDAS, most GHMs used in
444 earthH2Observe tier 1 product performed better than the LSMs, which is consistent with results
445 from other studies (Lakew et al., 2020). The strong performance of GHMs compared to LSMs
446 can be attributed to the differences in the model structure and parametrisation schemes between
447 LSMs and GHMs (Gründemann et al., 2018; Koukoulou et al., 2020). For example, some GHMs
448 such as Watergap are able to simulate lakes and reservoirs and water withdrawal while LSMs
449 can only simulate natural processes. Such differences in model structure can significantly
450 influence discharge volumes simulated by both types of models (Gründemann et al., 2018).
451 Although PCRGLOBW is a GHM, it produced substantially low performance compared to the
452 LSMs which is consistent with results from other studies in the region (Gründemann et al.,
453 2018; Lakew et al., 2020). This suggest that PCRGLOBW model may not be suitable for
454 assessing water security in the region.

455 The ability of the models to simulate flow thresholds was evaluated using the CSI.
456 Results show that NOAH, Terra, AWRAL and Lisflood were able to capture more than 50%
457 of 80th percentile monthly flow in most basins. We also noted that apart from NOAH model,
458 the rest of the GHMs performed better than the LSMs from earthH2Observe in their ability to
459 capture the 80th percentile monthly flows across the basins while only NOAH was able to
460 capture 20th percentile flows in three basins. The better performance of NOAH model
461 compared to other models evaluated in this study can be attributed to the fact that FLDAS was
462 specially designed and optimized to produce physically meaningful quantitative data for
463 monitoring food and water security in data-scarce regions in Africa (Mcnally et al., 2017). The
464 slight better performance of NOAH can also be attributed to its higher spatial resolution (0.1°)



465 compared to other models with coarser spatial resolution (0.5°). Terra with a spatial resolution
466 of 0.041° also performed slightly better than the other models with coarser spatial resolution.
467 In fact, a previous study (Gründemann et al., 2018), has shown that WRR products with higher
468 spatial resolution perform better than products with coarser resolution in their ability to
469 simulate discharge. The better performance of NOAH can also be attributed to the fact the
470 FLDAS is driven by a combination of different precipitation products which reduces the
471 uncertainties in the input data while earth2observe tier 1 product are driven by only one data
472 source (WFDEI) with uncertainties in the input data which is propagated to the model outputs.
473 Our results also showed that Lisflood performed better than most other earth2observe models
474 which can also be attributed to the fact that Lisflood has been extensively used in research and
475 operational settings in Africa (Thiemig et al., 2015; Smith et al., 2020). As such, the model
476 parameters may have been better constrained in the region than other models from
477 earth2Observe. Taking together, results from this study highlight the importance of evaluating
478 outputs from WRR products in representative basins before applying them in studies that may
479 have wider policy and financial implications. However, our results suggest a need to enhance
480 the spatial resolution of WRR products and for these products to be driven by data from
481 multiple sources to reduce the uncertainties input data.

482 **4.2. Evapotranspiration products**

483 The annual ET – precipitation ratio produced by the ET products in this study is in the same
484 order of magnitude with that produced by ET_{WB} except for SSEBop and MOD16A2 which are
485 within the range estimated for the global land regions (Rodell et al., 2015). This indicates that
486 most ET products performed well in this aspect of the ET evaluation. The annual ET –
487 precipitation ratios obtained in this study suggests that annual ET does not exceed annual
488 precipitation in any of the basins during the period under evaluation which is an indication of
489 available water resources in each basin.

490 Taking together all the ET evaluation criteria, FLDAS, GLEAM3.5a & 3.5b, Terra and
491 PMLV2 appear to outperform the other products even though GLEAM products systematically
492 underestimated ET in all the basins. Conversely, SSEBop and MOD16A2 produced poor did
493 not perform well in all the basins and may not be suitable for water security assessments in the
494 region. Our results are generally consistent with those from other studies indicating that
495 GLEAM and MODIS16A2 underestimate evapotranspiration, while SSEBop overestimates
496 this variable in most parts of Africa (Weerasinghe et al., 2020; Adeyeri and Ishola, 2021;
497 Mcnamara et al., 2021). Given that FLDAS ET estimate is derived from a LSM (NOAH) with



498 other water balance components (runoff, soil moisture and baseflow), it may be more preferable
499 for assessing water security in ungaged basins because of water balance closure. Our results
500 also revealed that the performance of the ET products was not influenced by spatial resolution
501 which is consistent with results from previous studies (Weerasinghe et al., 2020; Jiang and Liu,
502 2021). For example, Gleam products with a spatial resolution of 0.25° outperformed products
503 such as MODIS16A2 and SSEBop with higher spatial resolutions. Weerasinghe et al. (2020)
504 reported that re-gridding ET products to the same spatial resolution did not have any significant
505 impact on the performance of the product.

506 Although all the products were able to capture the temporal ET cycle in the basins, there
507 were substantial differences in the magnitude of monthly ET from each model. This finding is
508 consistent with results from other studies showing strong differences in ET estimates produced
509 by different models over Africa (Weerasinghe et al., 2020; Adeyeri and Ishola, 2021). The
510 discrepancies in monthly ET estimates from the models may be attributed to differences in the
511 equations underpinning each ET model, model parameters, and uncertainties in the input data
512 used in driving the models. This is also in-line with findings from another study in West Africa
513 highlighting the impact of model parameters and precipitation input uncertainty on ET
514 estimates (Jung et al., 2019). Considering the aforementioned factors, it may be difficult to
515 expect the products to produce similar results. ET_{WB} estimates across all the basins produced
516 very high peaks during the rainy season which is also similar to the results of a related study in
517 West Africa (Andam-Akorful et al., 2015). The high peaks observed in ET_{WB} may be attributed
518 to errors inherent in monthly precipitation, river discharge, and TWSC estimates used in
519 estimating monthly ET_{WB} .

520 Given that there was no uncertainty information on the river discharge data used in this
521 study, we adopted a value of 20 % following a previous study in the region (Burnett et al.,
522 2020). In fact, we feel that this value is conservative considering that uncertainties in river
523 discharge in tropical regions have been shown to range from 41 to 200 % (Kiang et al., 2018).
524 The mean monthly relative uncertainty for ET_{WB} for most basins appears to be in the same
525 order of magnitude (16 %) with results obtained in the Amazon basin (Baker et al., 2021).
526 Results also showed that the relative uncertainty in ET_{WB} is not influenced by basin size as
527 both large and small basins produced similar (same order of magnitude) uncertainty estimates.
528 Relative uncertainty in monthly ET_{WB} was higher during the rainy season. This can be linked
529 to high rainfall input during the rainy season which translates to high river discharge and TWSC
530 thereby increasing the absolute uncertainties in the different water balance components terms
531 used in estimating monthly ET_{WB} . Another study has shown that rainfall input is a major source



532 of uncertainty in river discharge due to its sensitivity to rainfall changes (Berghuijs et al., 2017).
533 Results from this study suggest that the relative the uncertainty in monthly ET_{WB} may be
534 substantial which may influence the performance scores of the ET products when they are
535 evaluated using the ET_{WB} method. We therefore recommend that evaluating the performance
536 of ET products at this monthly timescale should be accompanied with the estimataion of
537 relative uncertainties in monthly ET_{WB} .

538 **5. Conclusions**

539 The objectives of this study were to assess the performance of water resources reanalysis and
540 evapotranspiration products and to estimate the relative uncertainties in monthly ET_{WB} across
541 eight basins in Africa. Results show varying strengths and weaknesses for the different models
542 used in the WRR products. Some models were able to capture the river discharge dynamics in
543 the basins while other models could not adequately capture this patter. Differences in the model
544 performance can be attributed to differences model structure, parameters, input data used in
545 driving the models and the spatial resolution of the WRR products. Apart from NOAH which
546 is a land surface model (LSM), global hydrological models (GHMs) evaluated in this study
547 performed better than LSMs while PCRGLOBW which is a GHM did not perform well.

548 Evaluation of gridded ET products also revealed varying strengths and weaknesses for
549 the different products. Based on the different evaluation criteria (bias, RMSE, Pearson
550 correlation coefficient, and temporal ET variability), FLDAS appears to outperform most of
551 other ET products and may therefore be recommended for water security assessment in the
552 region. More so, because of water balance closure and the availability of other water balance
553 components (runoff, soil moisture and baseflow). Our results also suggest that the performance
554 of the ET products is not influenced by spatial resolution, while differences in monthly ET
555 estimates may be attributed to differences in the equations underpinning each ET model and
556 the sources of input data used to drive the model. We also observed that while spatial resolution
557 may have an impact on the performance of WRR products, this was not the case with ET
558 products as their performance appears to not be dependent on the spatial resolution.

559 Our results also revealed that relative uncertainties in monthly ET_{WB} were substantially
560 higher during the rainy season which can be attributed to uncertainties emanating from higher
561 rainfall input leading to an increase in discharge magnitude and TWSC during this period.
562 Results also revealed that uncertainty in river discharge is the dominant source of uncertainty
563 in ET_{WB} . This underscores the need to prioritize the installation of new gauging stations while
564 upgrading existing stations because such large uncertainties could constrain our ability to



565 understand hydrologic variability and flow forecast and could seriously undermine the
566 evaluation results of WRR and ET products and the calibration of hydrological models.

567 Results from this study suggest that WRR and ET products may be used for water
568 security assessment in ungauged basins. However, it is imperative to evaluate the performance
569 of these products in representative gauged basins before applying them in ungauged basins.
570 This is because applying the products in ungauged basins without evaluating their performance
571 may lead to poor water management decisions with wider policy and financial implications.
572 However, there is also a need for WRR and ET products to be driven by input data from
573 multiple sources to reduce uncertainties in the input data while the spatial resolution of WRR
574 products also need to be enhanced. Results from this study may be used by the products
575 developers to improve on the quality of future generations of WRR and ET products.

576 **Author contributions:** EN and RGB designed the methodological framework and contributed
577 to the entire strategic and conceptual framework of the study. EN prepared the data, performed
578 the analyses, interpreted the results and wrote the original draft. JN and EIB provided discharge
579 data for the Mono and Oueme basins respectively. All authors read the paper and provided
580 feedback.

581 **Competing interests:** The authors declare that they have no conflict of interest.

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