

1 **Bedrock depth influences spatial patterns of summer baseflow, temperature, and flow**
2 **disconnection for mountainous headwater streams**

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26 **Abstract**

27 In mountain headwater streams the quality and resilience of cold-water habitat is regulated by
28 surface streamflow connectivity and groundwater exchange. These critical hydrologic processes
29 are thought to be influenced by the stream corridor bedrock contact depth (sediment thickness),
30 which is often inferred from sparse hillslope borehole information, piezometer refusal, and
31 remotely sensed data. To investigate how local bedrock depth might control summer stream
32 temperature and channel disconnection (dewatering) patterns, we measured stream corridor
33 bedrock depth by collecting and interpreting 191 passive seismic datasets along eight headwater
34 streams in Shenandoah National Park (Virginia USA). In addition, we used multi-year stream
35 temperature and streamflow records to calculate summer baseflow metrics along and among the
36 study streams. Finally, comprehensive visual surveys of stream channel dewatering were
37 conducted in 2016, 2019, and 2021 during summer baseflow conditions (124 total km of stream
38 length). We found that measured bedrock depths were not well-characterized by soils maps or an
39 existing global-scale geologic dataset, where the latter overpredicted measured depths by 12.2 m
40 (mean), or approximately four times the average bedrock depth of 2.9 m. Half of the eight study
41 stream corridors had an average bedrock depth of less than 2 m. Of the eight study streams,
42 Staunton River had the deepest average bedrock depth (3.4 m), the coldest summer temperature
43 profiles, and substantially higher summer baseflow indices compared to the other study streams.
44 Staunton River also exhibited paired air and water annual temperature signals suggesting deeper
45 groundwater influence, and the stream channel did not dewater in lower sections during any
46 baseflow survey. In contrast, Paine Run and Piney River did show pronounced, patchy channel
47 dewatering, with Paine Run having dozens of discrete dry channel sections ranging 1 to greater
48 than 300 m in length. Stream dewatering patterns were apparently influenced by a combination

49 of discrete deep bedrock (20+ m) features and more subtle sediment thickness variation (1-4 m),
50 depending on local stream valley hydrogeology. In combination these unique datasets show the
51 first large-scale empirical support for existing conceptual models of headwater stream
52 disconnection based on underflow capacity and shallow groundwater supply.

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54 **1. Introduction**

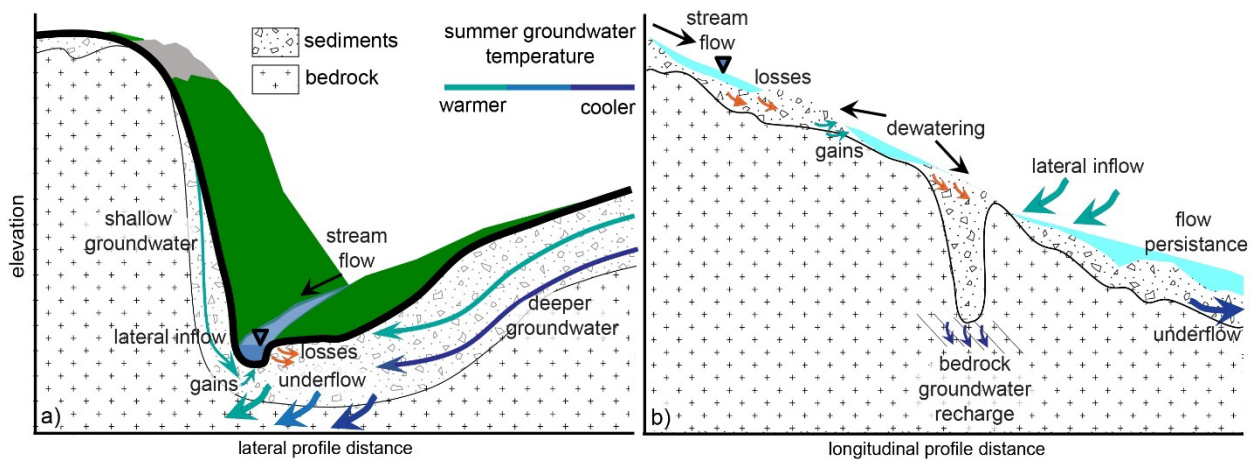
55 Mountain headwater stream habitat is influenced by hydrologic connectivity along the
56 surface channel, and connectivity between the channel and multiscale groundwater flowpaths
57 (Covino, 2017; Fausch et al., 2002; Wohl, 2017). Discharge from shallow groundwater within
58 the critical zone is a primary component of stream baseflow, attenuating maximum summer
59 temperatures and creating cold water habitat (Singha and Navarre-Sitchler, 2021; Sullivan et al.,
60 2021) and shaping catchment topography (Litwin et al., 2022). In headwater stream valleys
61 characterized by irregular bedrock topography and thin, permeable sediments, nested physical
62 processes interact to control the connectivity of groundwater/surface water exchange (Tonina
63 and Buffington, 2009). Between stormflow and snowmelt events, headwater streamflow
64 (baseflow) is primarily generated by groundwater discharge due to a relative lack of soil water
65 storage and release (Winter et al., 1998). Unlike in lower valley settings, mountain headwaters
66 accumulate less fine soil, facilitating efficient routing of quickflow to streams through
67 macropores and other preferential flowpaths within regolith and saprolite (Sidle et al., 2000).
68 Recharge that does percolate vertically contributes to shallow groundwater along steep hillslopes
69 and valley floors, where groundwater flowpath depths are constrained by bedrock topography
70 (Buttle et al., 2004). Although deeper groundwater may also represent an important contribution
71 to summer streamflow in systems with relatively permeable bedrock (Burns et al., 1998;
72 O’Sullivan et al., 2020), shallow, low permeability bedrock generally restricts stream-
73 groundwater connectivity to the thin layers of unconsolidated sediments (Briggs et al., 2018b).

74 In addition to baseflow drainage along headwater stream networks, down-valley shallow
75 groundwater ‘underflow’ can be substantial when high gradient streams lack sinuosity and flow
76 over permeable sediment (Figure 1a, Figure A1). In fact, headwater stream channels may only be
77 expected to show surface flow when the transmission ability of the underlying alluvium and

78 colluvium is exceeded, and bedrock depth is thought to be a primary control of this underflow
79 capacity (Ward et al., 2020). In some hydrogeologic settings, underflow can dominate
80 groundwater export from mountain catchments compared to groundwater drainage via the
81 surficial stream channel (Larkin and Sharp, 1992; Tiwari et al., 2017). Moreover, in addition to
82 longitudinal transport down-valley, underflow also acts as a reservoir of exchange for hyporheic
83 flowpaths that may mix with shallow groundwater before returning to channel flow (Payn et al.,
84 2009), transporting buffered temperature signals back to channel waters (Wu et al., 2020). Local
85 underflow is recharged from upgradient flowpaths and adjacent hillslopes, creating complex
86 seasonal and interannual patterns in groundwater connectivity and discharge to surface water
87 (Jencso et al., 2010; Johnson et al., 2017). A major challenge to understanding groundwater
88 exchange in headwaters is that attributes of the streambed subsurface, such as the depth to the
89 underlying bedrock contact, are often only available from limited direct measurements, coarse
90 spatial interpolations, or inferred remotely based on landscape forms. Therefore, methods that
91 allow efficient, local measurements of the streambed subsurface are critically needed.

92 Seasonal thermal regimes of mountain headwater streams can be profoundly impacted by
93 groundwater inflow from multiple depths (Briggs et al 2018a). In lower valley settings, the
94 temperature of groundwater discharge along stream networks is often assumed to be constant
95 throughout the year and approximate the average annual land surface temperature (Stonstrom
96 and Constantz, 2003). Conversely, shallow groundwater temperature (within several m from land
97 surface) can show pronounced seasonality (Bundschuh, 1993; Lapham, 1989) and high spatial
98 variability, even over small spatial extents (Snyder et al. 2015). The warming of shallow
99 groundwater during the summer and fall seasons can limit the ability of gaining mountain
100 streams to support cold-water fish populations during the low flow season, even if baseflow

101 (assumed to be dominated by groundwater discharge) fractions are large (Johnson et al., 2020).
 102 In systems with low permeability bedrock, thicker hillslope sediments may generate deeper,
 103 colder lateral groundwater flow to streams in summer (Figure 1a), increasing cold water habitat
 104 resiliency (Briggs et al., 2018b). For example, a recent meta-analysis of stream and air
 105 temperature records across the contiguous United States found that a substantial fraction of
 106 shallow groundwater dominated streams displayed summer warming trends in recent decades,
 107 while deeper groundwater dominated streams were more stable (Hare et al., 2021). Steep
 108 mountain stream systems such as those found in the Blue Ridge and Cascade mountains of the
 109 USA have been found to show annual thermal regimes dominated by the annual thermal signals
 110 indicative of shallow groundwater (Johnson et al., 2020), indicating such streams may also be at
 111 risk for warming over time, contrary to assumptions based on elevation alone.



112
 113 *Figure 1. A conceptual mountain stream valley cross section (panel a) and longitudinal profile*
 114 *(panel b) indicating the expected control of low permeability bedrock topography on*
 115 *groundwater temperature, stream-groundwater exchange, patchy stream dewatering, and the*
 116 *underflow reservoir.*

117 Beyond warm summer stream temperatures, the dewatering and disconnection of the
 118 active stream channel during summer low flows can adversely impact fish habitat by impeding
 119 fish movement (Edge et al., 2017; Labbe & Fausch, 2000; Rolls et al., 2012; Snyder et al.,
 120 2013), locally degrading water quality (Hopper et al., 2020), and increasing predation risks in

121 isolated pools (Magoulick and Kobza 2003). However, the physical controls on localized stream
122 channel dewatering are not well characterized and likely involve a spectrum of nested gaining
123 and losing flowpaths. For mountain headwater streams, previous research has documented major
124 contractions of drainage networks during seasonal drydown (Ilja Van Meerveld et al., 2019) and
125 general seasonal shifts in hydraulic gradients from gaining to losing, with closely coupled
126 streamflow and precipitation events, indicating a dominance of shallow routing rather than
127 deeper groundwater connectivity in maintaining streamflow (Zimmer and McGlynn, 2017).
128 Warix et al., (2021) found that although deeper/older groundwater was found to contribute to
129 their study streams during dry down, those sources were insufficient in preventing dry channel
130 sections from occurring, also indicating the importance of shallow groundwater inflows and local
131 geologic controls. Locally-losing sections of headwater stream channels can be associated with
132 coarse, permeable colluvial deposits from hillslope mass wasting processes (Costigan et al.,
133 2016; Weekes et al., 2015), as local enhancement of the total pore space under mountain streams
134 can drive downwelling of streamwater (Figure 1b, Tonina & Buffington, 2009). Main channel
135 dewatering occurs when the bed sediments have a storage and transport capacity that exceeds
136 stream discharge (Rolls et al., 2012; Ward et al., 2018), though stream water losses can also be
137 driven by local changes in bed morphology and slope (Costigan et al., 2016) and bedrock
138 permeability. The shallowing of the underlying bedrock contact may drive lateral underflow
139 toward the surface causing the channel to gain water (Herzog et al., 2019, Figure 1b), though
140 such hypothesized dynamics are not well documented in existing literature due to a relative lack
141 of bedrock topography data along headwater streams.

142 At large scales, contiguous bedrock depth layers are interpolated from a combination of
143 relatively sparse borehole data and surface topography (Kauffman et al., 2018; Pelletier et al.,

144 2016; Shangguan et al., 2017). However, in steep headwater systems with little borehole data,
145 bedrock topography is difficult to predict accurately from land surface topography alone. The
146 development of improved tools for predicting bedrock depth is an active area of research which
147 has recently demonstrated promise when bedrock outcrop data are included (e.g. Furze et al.,
148 2021; Odom et al., 2021). The limitations of using landform data to predict bedrock depth are
149 compounded by inherent challenges in collecting physical data via soil pits and monitoring wells
150 in rugged, rocky terrain, and so direct measurement data are often limited to highly studied
151 experimental watersheds where bedrock depth is still only *inferred* from piezometer installation
152 refusal (e.g. Jencso et al., 2010; Ward et al., 2018). In more typical headwater systems, existing
153 wells may be preferentially installed to maximize the production of water and not broadly sample
154 the true range of bedrock depths.

155 Application of near surface geophysical methods to stream corridor research has
156 increased appreciably in recent years (McLachlan et al., 2017), and several methods are sensitive
157 to shallow subsurface flow and geologic attributes including bedrock depth. Active seismic
158 refraction measurements can provide high resolution (10s of cm) bedrock depth information
159 along transect-based cross-sections (e.g. Flinchum et al., 2018), but are less suited for
160 exploration throughout rugged mountain stream valleys at the many km-scale due to logistical
161 challenges in using active seismic methods to obtain a sufficient amount of data to effectively
162 characterize important variation in bedrock depth at relatively small, ecologically-relevant spatial
163 scales.

164 Point-based, efficient passive seismic measurements represent a unique combination of
165 high mobility and relative precision for measuring bedrock depth along mountain valleys. The
166 horizontal-to-vertical spectral ratio (HVSr) method is a passive seismic technique that evaluates

167 ambient seismic noise recorded using handheld instruments placed on the ground surface to
168 identify seismic resonance that develops due to strong vertical changes in subsurface acoustic
169 impedance (Yanamaka et al., 1994). While typically insensitive to variations in unconsolidated
170 sediment permeability (i.e., clay lenses), the HVSR method is effective at identifying the depth
171 to distinct unconsolidated sediment/bedrock interfaces at essentially the ‘point’ spatial scale.
172 HVSR measurements are often not successful in settings with highly weathered bedrock surfaces
173 such as those with pronounced epikarst and saprolite.

174 The control of stream to groundwater exchange (i.e., ‘transmission losses’) on streamflow
175 permanence has been highlighted as an important research need by the comprehensive review of
176 intermittent stream systems by Costigan et al., (2016). Following the conceptual model of Ward
177 et al., (2018) [for mountain stream corridors](#), a central hypothesis of our research was that
178 bedrock depth along the stream corridor will act as a first-order control on stream dewatering
179 patterns when shallow bedrock is of low permeability. Based on the concepts presented by
180 Tonina & Buffington, (2009), we postulated that relatively thick, permeable surficial sediment
181 zones could locally accommodate the entirety of low streamflow volumes, dewatering main
182 channel sections at varied scales when not balanced by groundwater inflow (Figure 1b). We
183 further hypothesized that summer stream channel thermal regimes would also be influenced by
184 bedrock depth, as the temperature of groundwater flowpaths that generate baseflow is depth
185 dependent (Briggs et al., 2018b), indicated conceptually in Figure 1a. To test our hypotheses, we
186 extended the existing mountain headwater bedrock depth surveys from Shenandoah National
187 Park (SNP), Virginia, USA (citation) to seven additional subwatersheds and compared results to
188 physical mapping of stream dewatering, multi-year stream temperature data and derived

189 groundwater influence metrics, and baseflow separation analysis to address the following
190 research questions:

- 191 1. Does stream corridor bedrock depth exhibit longitudinal spatial structure in mountainous
192 streams at ecologically relevant spatial scales? Can measured bedrock depth dynamics be
193 accurately extracted from existing large-scale datasets or inferred from high resolution soils
194 maps?
- 195 2. Does underflow generally represent a net source or sink of summer flow for headwater
196 streams based on observed dewatering patterns and groundwater influence metrics?
- 197 3. Does bedrock depth explain spatial variation in stream temperature and summer baseflow
198 indices within headwater streams?

199 **2. Study Area**

200 SNP is an 800 km² area of preserved headwater forest perched along a major ridgeline of
201 the Blue Ridge Mountains in northern VA, USA (Figure 2). The bedrock of the park is
202 predominantly low permeability basaltic and granitic material in the central and northern
203 sections, and siliciclastic along the southern section (Southworth et al., 2009), though many
204 subwatersheds also transition in dominant bedrock type from high to lower elevation. Stream
205 valleys of SNP are typically steep and feature a perennial channel with mainly non-perennial
206 tributaries (Johnson et al., 2017, Figure A1) and stream baseflow consists of less than 3-yr old
207 groundwater on average (Plummer et al., 2001). In contrast, water collected from SNP hillslope
208 wells completed in shallow fractured rock generally have higher ages of 10-20 yr (Plummer et
209 al., 2001), indicating minimal contributions from bedrock groundwater to streamflow. Previous
210 ecohydrological research in SNP has noted that some mainstem stream channels show patchy

211 dewatering at summer low flows (Snyder et al., 2013), though the physical controls on these
212 patterns of stream drying were not clear.

213 In SNP, stream baseflow is thought to be predominantly generated by near-surface
214 drainage of coarse unconsolidated alluvium and colluvium (DeKay, 1972; Nelms and Moberg,
215 2010). The mountain ridgeline streamflow systems are expected to drain near-surface flowpaths
216 and accommodate substantial down valley underflow below perennial stream channels (Figure
217 A1). A portion of hillslope recharge is expected to percolate downward through connected
218 bedrock fractures into the deeper groundwater reservoir contributing to mountain block recharge
219 along the Shenandoah River Valley. Narrow alluvium deposits mapped along the stream
220 corridors of SNP are thought to generally range up to 6 m in thickness and be more clay rich
221 when sourced by basaltic bedrock (Southworth et al., 2009). Data at sparse wells drilled along
222 the SNP ridgeline indicate bedrock depth can range over 20 m on hillslopes and be highly
223 variable (DeKay, 1972; Goodling et al., 2020; Lynch, 1987).

224 Previous research has inferred summer and annual groundwater discharge patterns
225 throughout SNP subwatersheds based on paired, local air and stream water temperature
226 dynamics (Briggs et al., 2018a; Johnson et al., 2017; Snyder et al., 2015). Combined, these
227 analyses indicated groundwater exchange is highly variable in space along singular stream
228 valleys and between subwatersheds, and dependent upon local- to subwatershed-scale
229 characteristics. A combination of landform features that include stream slope and stream valley
230 confinement operate in conjunction with seasonal precipitation to drive groundwater influence
231 on summer stream temperatures (Johnson et al., 2017). Multi-week lags in time between
232 streamwater and local air annual temperature signals (i.e., water phase shifts toward later time)
233 were observed from dozens of the 120 total monitored stream sites indicating a dominance of

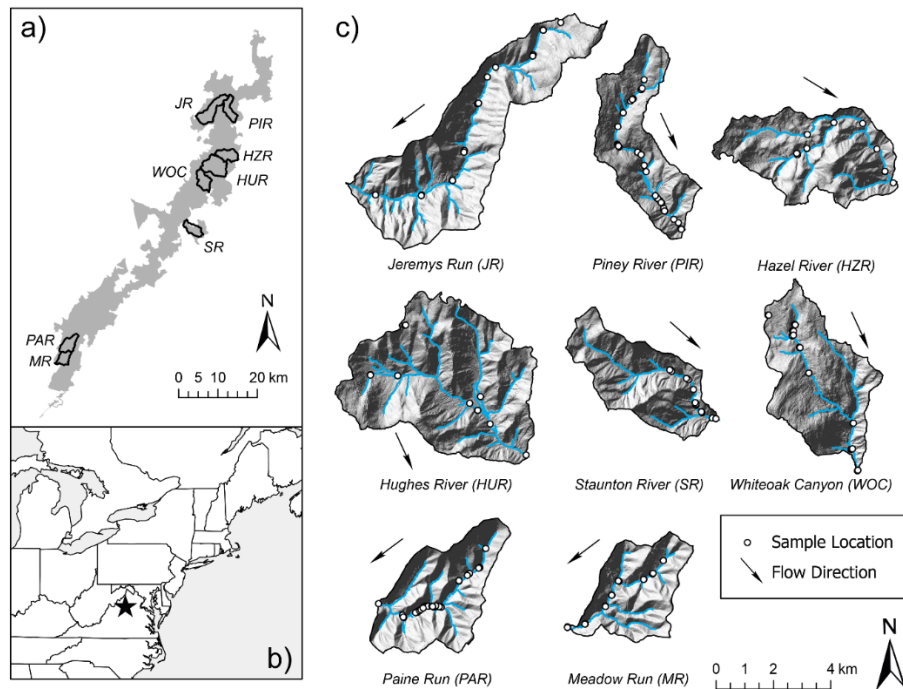
234 shallow groundwater discharge, originating generally within approximately 3 m of land surface
235 (Briggs et al., 2018a).

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241 *Figure 2. This study was based in Shenandoah National Park (panel a) located in the Blue Ridge*
242 *Mountains of northeast USA (panel b). LiDAR hillshade cutouts of each subwatershed illustrate*
243 *the rugged terrain and varied valley morphology (panel c). The mainstem stream channel and*
244 *tributaries are traced and passive seismic sample measurement locations noted.*

245 3.0 Methods

246 3.1 Passive Seismic Bedrock Depth Measurements

247 Periodically from the summer of 2016 to the spring of 2020, we acquired 323 HVSR

248 measurements across SNP. The geophysical data were collected along the perennial streams of

249 seven subwatersheds with extensive existing stream temperature and ecological datasets, and at
250 known ridgeline and hillslope borehole locations. This effort added to previously interpreted
251 HVSR data from 22 riparian sites collected along the Whiteoak Canyon subwatershed in late
252 2015 (Briggs et al., 2017), for a total of 8 mountain streams for analysis in this study (Figure 2).
253 In July 2016, HVSR data were collected in the following subwatersheds: Piney River, Paine
254 Run, Meadow Run, Jeremys Run, Hazel River, and Hughes River. Some stream sections were
255 inaccessible due to steep bedrock walls and waterfalls, resulting in poor data coverage in those
256 areas.



257
258 *Figure 3. Typical sections of a) Paine Run, b) Piney River, c) Staunton River, and d) a section of*
259 *Paine Run that was dewatered at baseflow leaving isolated pools. The passive seismic HVSR*
260 *instruments are shown deployed in panels a), b) and d) (Photographs by the U.S. Geological*
261 *Survey).*
262

263 Measurement locations mostly coincided with existing stream temperature monitoring
264 stations (described by Snyder et al., 2017), and were typically made at points immediately
265 adjacent to the stream or on larger rocks within the channel (Figure 3). In July 2019, HVSR data
266 were again collected along Paine Run and Piney River subwatersheds, and throughout the lower
267 Staunton River (Figure 3). The 2019 survey design differed in that transect measurements were
268 made at 4 locations along the stream channel waterline spaced approximately 25 m apart at
269 longitudinal locations that differed from the 2016 survey. This was done to assess potential
270 variation in bedrock depth along short subreaches of these three streams. Finally, clustered
271 HVSR data were collected in March 2020 in Paine Run and Piney River in zones previously
272 observed to show channel disconnection and streamflow re-emergence. Measurement locations
273 were chosen to test the hypothesis that the dewatering patterns were controlled by bedrock depth
274 as shown conceptually in Figure 1b.

275 HVSR data were collected using multi-component Tromino seismometers (MOHO,
276 S.R.L.) directly coupled to the land surface or placed on heavy metal plates where sediment was
277 loose. Collection times ranged 10-20 min at either 128 or 256 Hz sampling rates. HVSR data
278 collection locations were determined by a combination of internal Tromino GPS and external
279 GPS units. HVSR measurements were processed to derive a resonant frequency using a
280 commercially available program (GRILLA® v. 8.0 (2018)); further details regarding data
281 processing are given by Goodling et al., (2020).

282 Resonant frequency measurements that passed a series of quality criteria were then
283 converted to a bedrock depth estimate following Briggs et al., (2017). This conversion
284 necessitates a shear wave velocity estimate for the unconsolidated sediments over bedrock.
285 HVSR data collected at six spatially distributed boreholes with documented depth to varied-type

286 bedrock along the SNP ridgeline indicated a mean shear wave velocity of 358.7 +/- 56 m/s
287 (Goodling et al., 2020). A similar shear wave velocity of 346 m/s was measured at two locations
288 along the Whiteoak Canyon riparian zone spaced several km apart using active seismic methods
289 (Briggs et al., 2018b). This agreement indicates a common shear wave velocity can be assumed
290 for the unconsolidated material of SNP subwatersheds. For this study we used the average of
291 these spatially distributed active and passive seismic methods at 352 m/s. The average shear
292 wave velocity calculated in this study is comparable to the mean shear wave velocity ranges in
293 firm soils (180 - 360 m/s) and very dense soil and soft rock (360-760 m/s), according to National
294 Earthquake Hazards Reduction Program (NEHRP) guidelines (Building Seismic Safety Council,
295 1994). As an example of measurement sensitivity to the shear wave velocity parameter for
296 shallow bedrock contacts, a velocity change in either direction by 25 m/s would generally shift
297 the bedrock depth estimate by <0.2 m.

298 *3.2 Observations of spatial dewatering patterns*

299 Longitudinal (upstream to downstream) patterns of dewatering were determined in the
300 summers of 2016, 2019, and 2021 during baseflow conditions over 124 total km of stream length
301 for all surveys combined. In July-August of 2016 all eight subwatersheds (Figure 2) were
302 surveyed. In September of 2019 and August 2021, dewatering surveys were repeated in three
303 subwatersheds (Paine Run, Piney River, and Staunton River) to evaluate annual variation in
304 dewatering patterns. Data were collected by a team of investigators walking each stream from an
305 upstream location defined by the point along the stream draining 75-hectares (assumed capture
306 area required to generate perennial streamflow, determined using watershed tools in ArcGIS) to
307 the bottom of each watershed near the park boundary, and mapping transition points between
308 three hydrologic categories: Wet, dry, or isolated pools based upon investigator observation.
309 “Wet” segments were defined as reaches where entire channel was wet with flow between pools;

310 “Dry” segments were defined as reaches containing no water, or isolated pools of insufficient
311 depth to sustain 1+ year old brook trout; and “Isolated Pools” were defined as reaches containing
312 pools of sufficient depth to support brook trout but were hydrologically disconnected from other
313 parts of the channel. An example of isolated pools is photographically depicted in Figure 3d.
314 Spatial coordinates of transition points were mapped using a Trimble R2 GNSS receiver for <1-
315 meter accuracy. Surveys for each subwatershed were completed within a single day to minimize
316 effects of temporal variation in precipitation.

317 In addition to local variability in bedrock depth, spatial patterns of dewatering and stream
318 temperature are likely to be influenced by seasonal precipitation and air temperature proximate
319 to the period of measurement (i.e., summer conditions, 2016 and 2019). We used historical
320 weather records (1942 – 2020) collected from the nearby Luray Weather Station located within
321 SNP (Station No. GHCND:USC00445096) to compare weather conditions during these two
322 study years with historical norms. Finally, 3D surface area of each subwatershed was determined
323 from existing LiDAR data using the Add Surface Information tool in ArcGIS and mean valley
324 bottom width was evaluated from LiDAR data using 100-m transects measured approximately 2
325 m above the valley floor.

326 *3.3 Stream channel temperature data and baseflow separation*

327 Multi-year SNP stream temperature data were collected at hourly time intervals as
328 described by Snyder et al., (2017) using HOBO Pro V2 thermographs (+/- 0.2 °C expected
329 accuracy). From this larger dataset, 64 main channel locations within the 8 study subwatersheds
330 were extracted and processed for summary statistics such as the maximum and minimum of the
331 7-day running mean using Matlab R2019b software (Mathworks, Inc.). Only complete 7-day
332 periods were included in the running average. Warm season data (July, August, September) were

333 isolated and analyzed to coincide with the stream dewatering surveys and a larger body of
334 research regarding summer cold-water brook trout habitat in SNP. We utilized stream
335 temperature data processed to extract annual temperature signals by Briggs et al., (2018a) where
336 dry sensor periods were identified and removed, impacting a handful of the upper stream sites.
337 Data were visualized and downstream trends explored using Sigmaplot 14.0 software (Systat
338 Software Inc.). Baseflow separation was conducted for the three continuously gaged streams of
339 this study (Paine Run, Piney River, Staunton River) over summer months for the period of record
340 (1993-2020). Following the approach of Hare et al., (2021), the daily Baseflow Index (BFI) was
341 calculated using the USGS-R 'DVstats' package (version 0.3.4) following methods described by
342 Barlow et al., (2014), and dividing the calculated baseflow discharge by the corresponding
343 stream discharge, where a value of one would indicate stream discharge was entirely composed
344 of baseflow. BFI was then averaged (mean) across each summer season, along with the mean
345 and standard deviation of summer stream discharge.

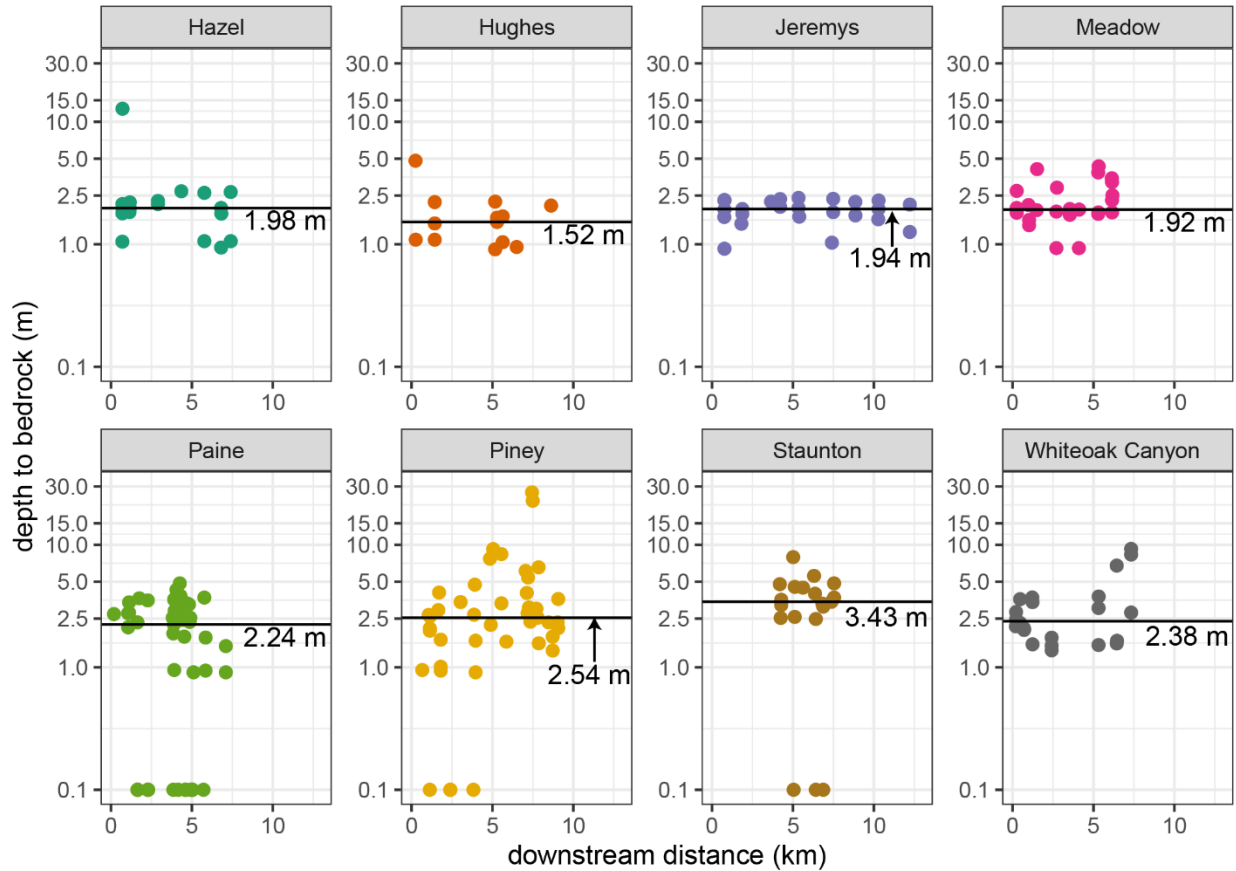
346 **4. Results**

347 *4.1 Stream Corridor Bedrock depth*

348 Approximately 60% of individual HVSR measurements (191 of the 323) were of high
349 enough quality to be interpreted for bedrock depth using objective data quality metrics reported
350 by the GRILLA software. This ratio of interpretable to total HVSR measurements was similar to
351 the previous 2015 Whiteoak Canyon Run study using the same instrument type (Briggs et al.,
352 2017). For the 132 datasets that could not be interpreted, the primary reason was no identifiably
353 resonant frequency 'peak' in the multicomponent seismic data, as described in more detail in the
354 data release of Goodling et al., (2020). The loosely consolidated, rocky surficial soils of many
355 SNP subwatershed riparian zones likely contributed to poor instrument coupling to the land

356 surface, and therefore reduced measurement sensitivity/success compared to firmer soils.
357 However, due to spatial redundancy in the measurements, the 191 locations where bedrock depth
358 was evaluated generally covered all the intended longitudinal stream measurement locations
359 throughout the subwatersheds.

360



361

362 *Figure 4. Measured depth to rock along the stream channel and riparian zones of the eight study*
363 *subwatersheds. Exposed bedrock (i.e., zero depth) observed at the intended measurement*
364 *location is noted here by a value of '0.1' on the log scale. The median value is shown as a*
365 *labelled horizontal line.*

366

367 The median bedrock depth was smallest for Hughes River (1.52 m), and similar for
368 Meadow Run, Jeremys Run, Hazel River (1.92, 1.94, 1.98, respectively, Table 1, Table B1,
369 Figure 4). Paine Run had a median of 2.24 m, Whiteoak Canyon of 2.38 m, and Piney River of

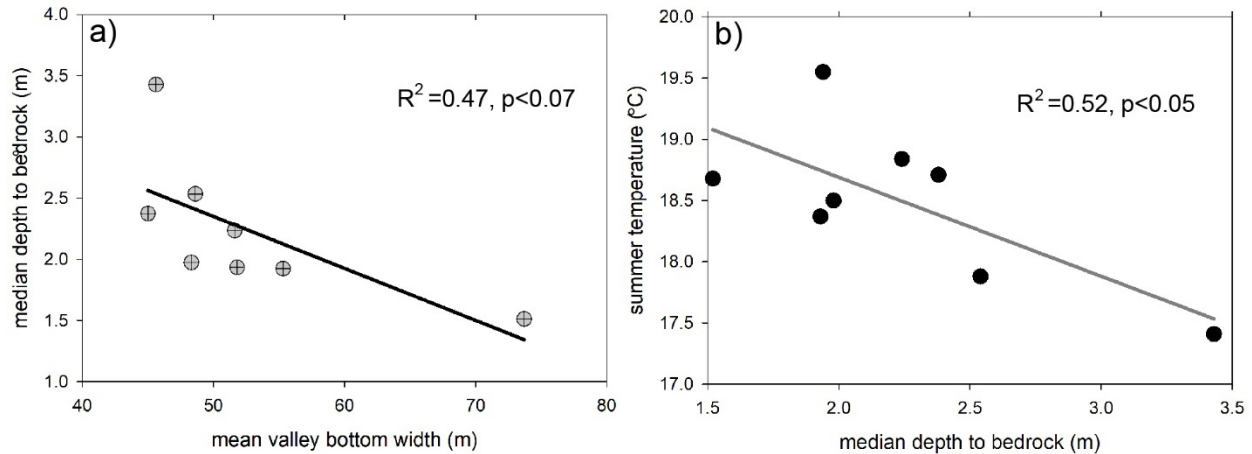
370 2.54 m. Lower Staunton River had the largest median depth to rock of 3.43 m (Table 1). Piney
 371 River had the largest variation in bedrock depth, including a discrete zone greater than 20 m
 372 deep, along with several zones of exposed bedrock along the channel. Visual observations of
 373 exposed channel bedrock were not incorporated into the bedrock depth averages presented in
 374 Table 1. Simple bivariate relations were explored between the physical valley parameters, and a
 375 negative relation was found between bedrock depth and mean valley bottom width while other
 376 relations were not significant.

377 **Table 1.** The median bedrock depth along with the elevation, mean, and 7-d maximum summer
 378 temperatures over the period of record collected at most downstream site location in each
 379 subwatershed.

<i>site</i>	<i>3D subwatershed surface area</i> (km ²)	<i>mean valley bottom width</i> (m)	<i>median bedrock depth</i> (m)	<i>mean stream slope</i> (°)	<i>most downstream stream temperature site</i>		
					<i>elevation</i> (m)	<i>mean</i> (°C)	<i>7-d max</i> (°C)
Hughes River	42.2	73.7	1.52	22.7	307	18.7	21.2
Meadow Run	15.0	55.3	1.93	14.2	450	18.4	20.4
Jeremys Run	37.5	51.8	1.94	16.3	286	19.6	23.6
Hazel River	22.5	48.3	1.98	13.0	328	18.5	21.7
Paine Run	21.7	51.6	2.24	15.6	426	18.8	20.9
Whiteoak Cyn.	22.4	45.0	2.38	17.2	348	18.7	21.2
Piney River	20.6	48.6	2.54	14.9	371	17.9	20.6
Staunton River	18.0	45.6	3.43	20.7	309	17.4	19.9

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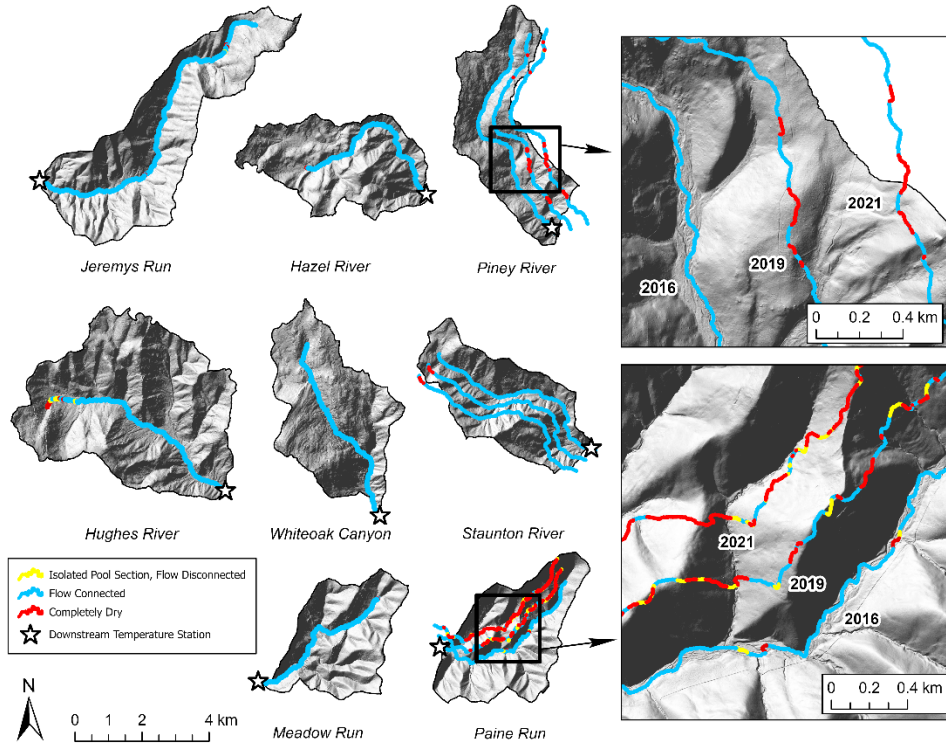
383 *Figure 5. Median study stream corridor bedrock depth showed a negative relation to valley*
 384 *bottom width (panel a), and mean summer stream temperature at the lower study stream*
 385 *boundaries was negatively related to median bedrock depth (panel b).*

386

387 4.2 Spatial Dewatering Patterns and Climate Data

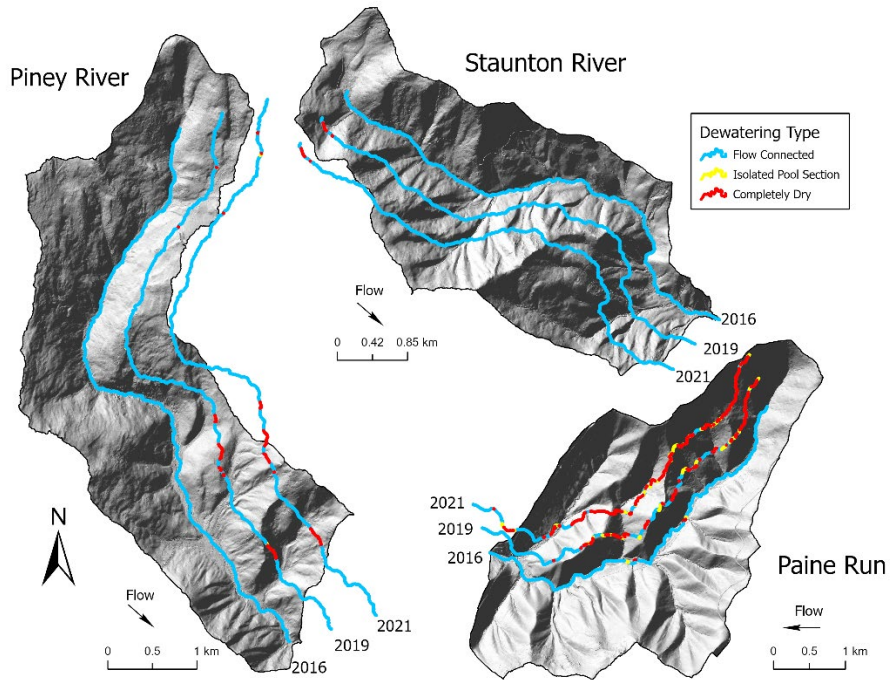
388 Cumulative monthly precipitation during baseflow summer (July-September) was higher
 389 than normal in 2016 and near average or lower than average (period of record 1942-2020),
 390 depending on the month, in 2019 (Figure A2). Mean monthly air temperatures were higher than
 391 average for both study years during baseflow summer reflecting the long-term trend of
 392 increasing air temperatures in the park (Luray weather station GHCND:USC00445096). Patches
 393 of stream dewatering were observed along five of the eight study subwatersheds between 19-27
 394 July, 2016, when over 98 km of total stream length were mapped (Figure 6). However, for
 395 Meadow Run, Hazel River, and Hughes River stream dewatering only occurred near the upper
 396 stream origination point. In contrast, Paine Run and Jeremys Run had several discrete dewatering
 397 sections further from their origination points (examples shown in Figure 3d, Figure A3). During
 398 the drier period 17-19 September 2019, no dewatering was found along lower Staunton River,
 399 though Piney River had seven discrete dry patches where none were mapped in 2016, and similar
 400 patterns were observed for those two streams in 2021 (Figure 7). Paine Run had 29 points of

401 dewatering in 2019, distributed mainly along the central and upper sections of the stream
402 corridor, and showed extensive dewatering in 2021 (Figures 6, 7, 8). The two Paine locations
403 that were dry in 2016 were also dry in 2019 and 2021.



404

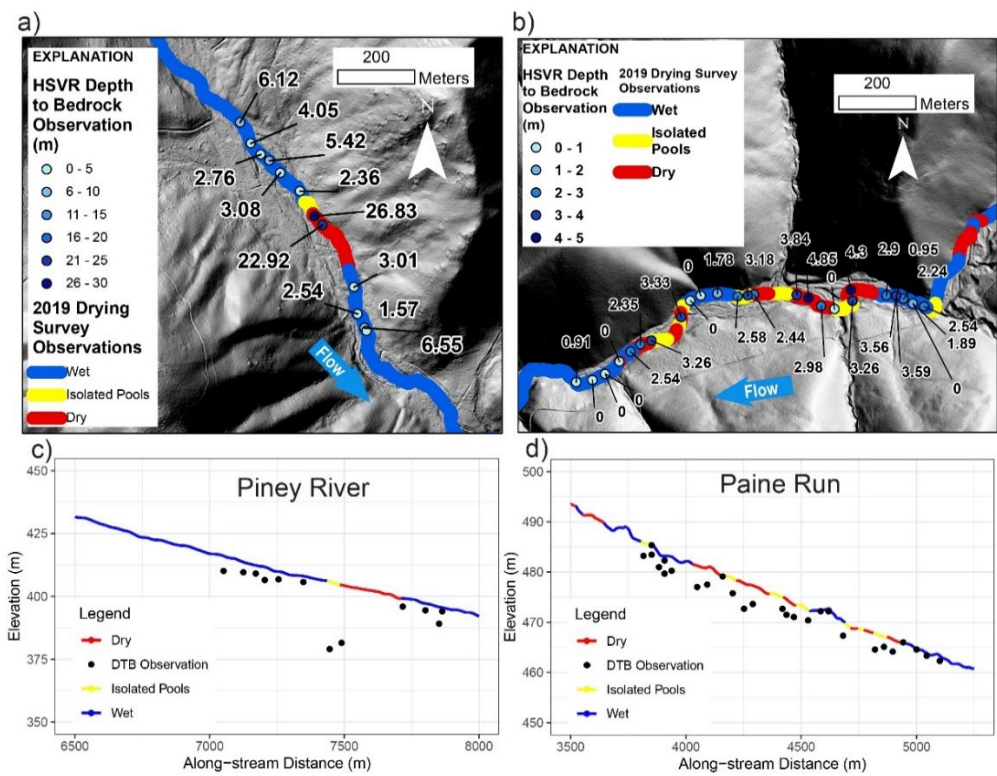
405 *Figure 6. Results from 2016, 2019, and 2021 longitudinal channel dewatering surveys conducted*
406 *by physical observation, where the 2019 and 2021 data are shown offset laterally from the*
407 *stream channel where those surveys occurred.*



408

409 *Figure 7. Zoom views for the three study subwatersheds where stream dewatering observations*
 410 *were also collected over three summer seasons (2016 ,2019, 2021).*

411
 412
 413

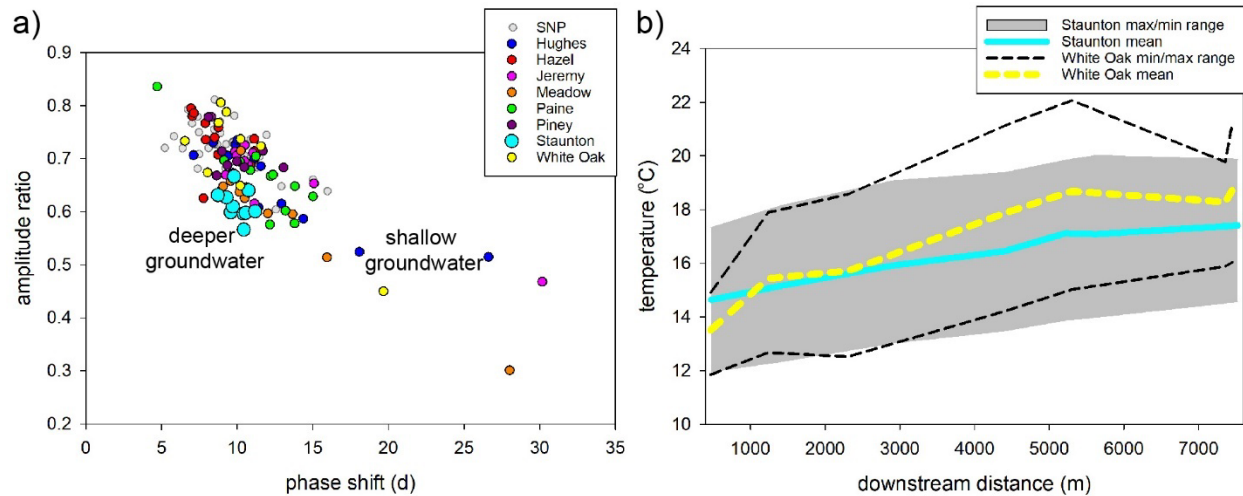


414

415 *Figure 8. The results of the 2019 stream drying survey and 2020 high spatial resolution HVSR*
 416 *measurements are shown over the LiDAR hillshade in plan view (panels a, b) and along a*
 417 *LiDAR-derived stream elevation profile cross-section view (panels c, d) for Piney River (panels*
 418 *a, c) and Paine Run (panels b, d).*

419 *4.3 Stream Temperature Patterns*

420 Paired air and water annual temperature signals exhibited a spectrum of shallow
 421 groundwater influences as indicated by extracting fundamental sinusoids from each multi-year
 422 temperature dataset per methods described by Briggs et al. (2018). Observed phase shifts
 423 between stream and local air temperature signals ranged from approximately 5 to 30 d with a
 424 mean of 11 d. Reduced annual temperature signal amplitude ratio generally corresponded with
 425 increased phase shift when all SNP stream monitoring sites are plotted in aggregate (Figure 9a).
 426 Staunton River stream sites cluster together and show less signal phase shift (mean of 10 d) for
 427 similarly low amplitude ratio values (mean of 0.6) observed in other subwatersheds.



428

429 *Figure 9. Panel a) shows the annual temperature signal metrics for the study subwatersheds*
 430 *highlighted within the larger SNP dataset with conceptual groundwater end member signature*
 431 *trajectories. Panel b) displays the downstream mean summer temperature profiles and 7-d*
 432 *maximum and minimum temperature ranges for Staunton River and Whiteoak Canyon.*

433

434

Although originating in a similar place (Table 1), the downstream mean, 7-d maximum,
 435 and 7-d minimum stream temperature profiles differed between Staunton River and Whiteoak
 436 Canyon, where the latter had greater temperature variation and warming with downstream
 437 distance (Figure 9b). The mean summer stream temperature had an approximate 2 °C total range
 438 over the period of record. The warmest average (19.6 °C) and 7-d maximum (23.6 °C) was
 439 observed for the lower Jeremys Run site, which was also at the lowest elevation. However, only
 440 23 m higher in elevation, the downstream Staunton River site had the coldest average (17.4 °C)
 441 and 7-d maximum (19.9 °C) summer temperature. Piney River, which has the second largest
 442 median bedrock depth (2.54 m), had the second lowest average temperature (17.4 °C) at the
 443 lower site. No significant relation was observed between elevation and mean summer
 444 temperature at the lower stream monitoring site, but a significant negative linear relation
 445 ($R^2=0.52$; $p<0.05$) was determined between median stream corridor bedrock depth and mean
 446 summer stream temperature (Figure 5b), with strong leverage on the linear fit imparted by the
 447 Staunton River datapoint.

448

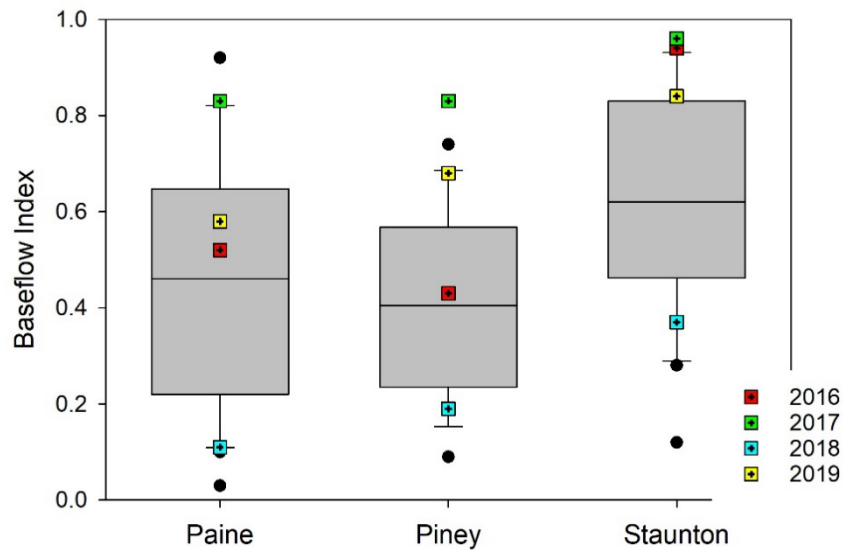
449 *4.4 Baseflow Separation (Index)*

450 The summer season BFI determined for Paine Run, Piney River, and Staunton River over
451 the period of flow record show substantial variability, but the median summer BFI for Staunton
452 River (0.62) is approximately 50% greater than Paine Run and Piney River (0.46 and 0.41,
453 respectively, Table 2). For the primary study years of 2016-2019, Staunton River BFI is always
454 largest, and all sites are above their respective interquartile range in 2017 but below their
455 interquartile range in 2018 (Figure 10). The anomalously low 2018 BFI values can be explained
456 by extremely high summer precipitation that year (Figure S2), resulting in total streamflow being
457 dominated by runoff and quickflow as determined by with baseflow separation. Mean summer
458 streamflow over the period of record was highest for Piney River and lowest for Pain Run, and
459 overall summer streamflow was most stable for Staunton River (lowest coefficient of variation).

460 **Table 2.** The median summer Baseflow Index (BFI), mean summer streamflow, and mean summer
461 standard deviation (SD) streamflow for three gaged streams from 1993-2020.

<i>site</i>	<i>median BFI</i>	<i>mean streamflow</i> (L/s)	<i>mean coefficient of</i> <i>streamflow variation</i>
Paine Run	0.46	93.0	1.6
Piney River	0.41	164.4	1.7
Staunton River	0.62	157.3	0.7

462



463

464 *Figure 10. Summer Baseflow Index metrics summarized from 1993-2020 for three streams, with*
 465 *specific values from the primary study years identified.*

466 **5.0 Discussion**

467 *5.1 Longitudinal Spatial Structure in Observed Bedrock Depth*

468 Seminal groundwater/surface water exchange research has indicated that bedrock
 469 topography along headwater streams may be a first-order control on the arrangement of nested
 470 gaining and losing flowpaths (e.g. Tonina & Buffington, 2009), and increased depth to low
 471 permeability bedrock contacts is recognized as a primary driver of stream disconnection during
 472 dry periods that could be exacerbated by climate change (Ward et al 2020). However, despite the
 473 apparent importance to a range of headwater stream physical processes and cold-water habitat,
 474 local bedrock depth data are almost universally lacking, even in heavily studied experimental
 475 watersheds. Our study provides new inferences regarding the effects of bedrock depth on
 476 groundwater exchange and consequent effects on stream dewatering and temperature patterns at
 477 ecologically relevant spatial scales in mountain streams. The combined datasets indicate stream

478 channel bedrock depth assessments may be necessary to support stream habitat assessments and
479 predictions of stream connectivity under drought and climate change when existing large-scale
480 geologic datasets are not of sufficient spatial resolution to support natural resource management
481 applications.

482 Bedrock depth varied substantially within and among several of the eight study SNP
483 subwatersheds but was predominantly shallow. For half of the subwatersheds (Hughes River,
484 Meadow Run, Jeremys Run, and Hazel River), median bedrock depth along the stream channel
485 and lateral riparian zone was less than 2 m and did not show notable variability with distance,
486 outside of one 12.8 m depth to rock location at upper Hazel River (Figure 4). This anomalous
487 measurement at Hazel was collected lateral to the stream on a valley terrace of colluvium, in the
488 vicinity of the only cold (approximately 10 °C at land surface) riparian spring that was observed
489 during all HVSR surveys. Bedrock depths of greater than 8 m were found along the upper
490 Whiteoak Canyon riparian zone as well (Briggs et al 2018a), also associated with surficial
491 seepage. Two anomalous bedrock depth measurements of 22.9 and 26.8 m were collected along
492 the Piney River channel, but instead of being associated with groundwater springs, they
493 coincided with a discrete sections of channel dewatering at baseflow during 2019 and 2021.
494 Therefore, it appears that discrete zones of thick surficial material are the exception along SNP
495 streams, though they can be important to localized processes such as focused riparian discharge
496 and streamflow disconnection (latter discussed in Section 5.2).

497 There are several existing sources of bedrock depth data that could potentially be used to
498 inform headwater stream modeling and habitat assessment, but the accuracy of such datasets
499 along headwater streams (typically away from existing boreholes) has generally not been
500 evaluated. We conducted a point-scale comparison of our relatively high-resolution bedrock

501 depth measurements to the global bedrock depth map of Shangguan et al., (2017) and found that
502 bedrock depths were almost universally overpredicted at the SNP by large margins (Figure A4).
503 Specifically, predictions from the global-scale dataset exceeded HVSR measured depths by
504 +12.2 m (mean), or approximately four times the average bedrock depth (2.9 m). This result may
505 not be surprising, as Shangguan et al., (2017) recognizes that large scale bedrock depth
506 interpolations are likely to overpredict shallow bedrock contacts, especially in mountainous
507 terrain with minimal available borehole data constraints. However, given that baseflow
508 generation is expected to be dominated by shallow groundwater sourced from unconsolidated
509 sediment in headwater systems with low permeability bedrock, our study highlights that use of
510 large-scale bedrock depth layers may propagate substantial uncertainty into process-based
511 groundwater flow model predictions when used to inform model structure in the absence of local
512 measurements.

513 Publicly available maps of surficial geologic materials are another potential source of
514 bedrock depth information. High-resolution digital soils maps are now widely available,
515 including for the catchments of SNP, and these maps do capture some of the general depth to
516 rock transitions between subwatersheds observed in this study. For example, NRCS (2020)
517 (<https://websoilsurvey.sc.egov.usda.gov/App/WebSoilSurvey.aspx>, accessed 12/10/2020)
518 indicate that the Whiteoak Canyon stream corridor is comprised of silts, loams, and stony soils
519 with a general bedrock depth of approximately 1.2 m., which is in a similar range as most HVSR
520 measurements made along the upper stream section (Figure 4). However, the generalized soil
521 units may not offer needed detail regarding site-specific valley sediment thickness for
522 hydrogeological and ecological studies where information regarding within-watershed variation
523 is critical. Along lower Piney River, where HVSR data had depths to rock ranging 1.4 to 3.6 m,

524 the NRCS soils map universally indicates silt and stony material > 2 m. Along Paine Run, where
525 the stream is often scoured to bedrock, the soils map shows consistent highly permeable sandy
526 material with > 2 m thickness. This discrepancy is understandable given most of the test pits
527 were likely substantially further downstream in better terrain for agriculture. In conclusion,
528 analysis of large-scale patterns from existing soils maps and interpolated/predicted bedrock
529 depth layers indicates that more precise geophysical mapping of bedrock depth may be needed to
530 inform stream research and management, particularly in shallow, low-permeability bedrock
531 terrain.

532 *5.2 Summer Stream Dewatering Related to Bedrock depth*

533 Aligned with the conceptual model of Ward et al., (2018), our central hypothesis was
534 bedrock depth along the stream corridor acts as a primary control on longitudinal stream
535 dewatering and flow disconnection during summer low flows (visual example shown in Figure
536 A3). We postulated that permeable streambed thickness may undulate along mountain stream
537 channels, and relatively thick sub-stream sediment zones could accommodate the entirety of low
538 streamflow volumes, locally disconnecting channels during seasonal flow recession. We found
539 mixed support for this simple hypothesis. Hazel River and Hughes River were two of the three
540 subwatersheds that had dry channel zones just downstream of their respective stream origination
541 points in 2016, and these two riparian corridors also had their deepest riparian bedrock depths in
542 those high-elevation areas. However, as discussed above, Whiteoak Canyon had relatively thick,
543 porous sediment zone near the subwatershed outlet but did not show any zones of dewatering,
544 nor did lower Staunton River in 2016, 2019, or 2021, despite having the deepest median bedrock
545 contact. Jeremys Run had three mapped dry zones in 2016 (not surveyed in 2019), yet depth to
546 rock in those areas was only approximately 2 m, though the HVSR data collection points were
547 not perfectly aligned with the dry patches. To address this spatial mismatch in stream dewatering

548 and HVSR data, we used the stream dewatering maps to guide two new high-resolution HVSR
549 surveys in March 2020 along sections of Paine Run and Piney River with dynamic patterns of
550 channel drying, as described below.

551 When bedrock depth data were collected at high-resolution, even more variability in
552 bedrock topography/sediment thickness was revealed than in the original larger-scale surveys,
553 and that finer scale of information was relevant to understanding stream dewatering patterns.
554 For example, during summer 2019, a 291 m length section of lower Piney River was observed to
555 be dry, and immediately preceded by 62 m of isolated stream channel pools, and a nearly
556 identical dewatering pattern was observed there in 2021 (Figures 7,8a,c). The upper portion of
557 this major feature of stream disconnection corresponded directly with a transition in bedrock
558 depth along the channel from approximately 3 m to adjacent measurements of 27 and 23 m. This
559 ‘trough’ in the bedrock surface can likely act as a streamwater sink (shown conceptually in
560 Figure 1b), routing surface water downward to the point of draining the channel locally in the
561 summers of 2019 and 2021, but not in 2016 when precipitation (groundwater supply) was higher
562 than normal. Further downstream, the bedrock depth returned to approximately 3 m near the
563 furthest downstream measurement point, and flowing channel water was again noted during the
564 drying surveys. Such a section of stream dewatering in the lower watershed would serve to
565 impede fish passage along Piney River during the lowest flows, likely corresponding to times of
566 maximum thermal stress when fish mobility is critical to seeking thermal refuge (Magoulick and
567 Kobza, 2003).

568 Not all variability in bedrock depth below streams associated with stream drying was as
569 dramatic as the Piney River example but can be important in disconnecting channel habitat in
570 summer. Paine Run is a more strongly confined stream valley that had 29 discrete zones of

571 stream channel dewatering during September of 2019 and extensive dewatering in 2021 (Figures
572 6,7,8b,d), when numerous dead brook trout were also noted. Paine also had the greatest total
573 exposed bedrock out of any of the SNP subwatersheds in this study, indicating a highly
574 constrained valley underflow reservoir. High resolution bedrock depth data was collected over a
575 Paine Run subreach with seven discrete dry patches ranging from 17 m to 185 m in channel
576 length, with many bordered by zones of isolated pools (Figure 8b). A comparison of these
577 patterns with bedrock depth along the channel shows the flowing sections of stream were
578 dominated by exposed bedrock surfaces or thin sediment. However, a notable exception is
579 toward the upstream end of this focus reach, where depth to rock was consistently > 2 m over the
580 run up to a large zone of disconnected channel with some isolated pools (Figure 8b,d). This
581 result suggests the losses of stream water accumulated over this approximately 80 m channel
582 distance. In the following downstream contiguous sections of dry channel and/or isolated pools,
583 bedrock depth averaged a larger 3.3 m, indicating the entirety of streamflow was accommodated
584 by the subsurface, congruent with our original hypothesis. However, knowledge of bedrock
585 depth in isolation is clearly not sufficient to predict stream channel gaining, losing, and
586 disconnection patterns as the stream with the largest average bedrock depth, lower Staunton
587 River (median depth to rock 3.4 m, Figure 4), was not observed to dewater during any of the
588 three physical surveys (Figures 6,7).

589 *5.3 Summer Stream Temperature and Groundwater Exchange Dynamics*

590 Although headwater stream heat budgets are complex, our data indicates groundwater
591 connectivity plays an important role when stream temperatures are already close to aquatic
592 species thermal tolerances. The apparent dominance of shallow (<3 m depth) groundwater
593 discharge along Whiteoak Canyon contributed to the Briggs et al. (2018b) prediction that the
594 lower reaches would not provide suitable brook trout habitat by the end of the century given

595 anticipated atmospheric warming. Jeremys Run, a long (13.4 km) stream consistently underlain
596 by a shallow bedrock contact (median depth < 2 m), already shows a 7-d maximum summer
597 temperature that exceeds expected brook trout tolerances (i.e., >23.3 °C mean weekly average
598 temperature, Wehrly et al., (2007)) along the lowest reach.

599 The underflow reservoir of headwater stream valleys integrates upgradient and lateral
600 hillslope groundwater flowpaths, which accumulate with distance when bounded by low
601 permeability bedrock. The two subwatersheds with largest median bedrock depth along their
602 respective upstream corridors had the coldest mean summer temperatures, with Staunton River
603 standing out as distinctly colder, and having the only 7-d max temperature below 20 °C (Table
604 1). There was a significant relation between median bedrock depth and mean summer stream
605 temperature at the lower stream sites but not with elevation (Figure 5b), indicating exchange
606 with groundwater had disrupted the expected elevation control on lower reach cold water habitat.
607 Surficial hillslope contributing area is often assumed a primary control on potential groundwater
608 discharge at the stream subreach scale. However, Staunton River also had the second smallest
609 drainage surface area of all study subwatersheds, and it is often assumed that lateral groundwater
610 inflow to headwater streams is related to presumed upslope contributing area. Further, Staunton
611 River did not have an average valley bottom width that was greater than other streams that were
612 observed to dewater. This apparent conundrum indicates the importance of bedrock depth
613 (suprabedrock aquifer thickness) in facilitating spatially persistent baseflow generation during
614 dry times, and we also found that the more narrow headwater stream valleys of this study tended
615 to have deeper bedrock depth (Figure 5a).

616 Our research indicates that the vertical shallow aquifer dimension, as represented by
617 bedrock depth, is likely an important control of groundwater storage and connectivity to the

618 stream corridor. This conclusion is supported by the paired air/water annual temperature signal
619 metrics, indicating Staunton River sites cluster toward stronger, deeper groundwater influence
620 compared to most observations along the other SNP streams (Figure 9a). Therefore, it seems
621 there are important tradeoffs between bedrock depth along the stream channel as a driver of
622 stream dewatering and sediment thickness along the valley floor and hillslopes as a potential
623 source of stream baseflow.

624 For a more in-depth analysis of the paired bedrock depth and groundwater inflow
625 controls on headwater summer stream dynamics, Staunton River can be contrasted with Paine
626 Run. The latter had a similar total drainage surface area to Staunton River with a >5 m (average)
627 wider stream valley bottom, but a 1.2 m shallower bedrock depth on average. Paine Run had
628 dozens of dewatered stream channel sections in 2019 and 2021, and had a downstream boundary
629 summer stream temperature that was 1.4 °C warmer than Staunton River. In addition to a
630 reduced average bedrock depth, Paine Run had numerous sections of exposed bedrock adjacent
631 to localized pockets of stream channel alluvium and colluvium (Figure 4, 7), while extensive
632 colluvial deposits along the Staunton River channel limited exposed bedrock to a few m-scale
633 sections associated with pool steps (Figure 4). Lower Staunton experienced major debris flows in
634 June, 1995 (Morgan and Wiczorek, 1996), events that likely created an enhanced local
635 groundwater reservoir within coarse hillslope material compared to other SNP subwatersheds.

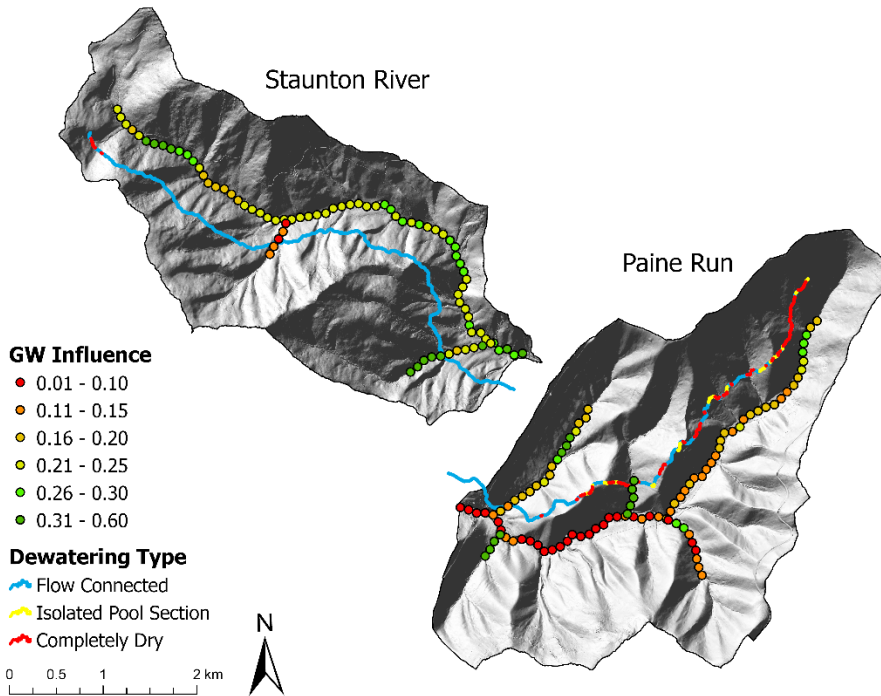
636 Based on the integrated datasets from these two SNP streams, we conclude that
637 groundwater exchange is a critical factor determining whether headwater streams will warm and
638 dewater in summer, which in turn is controlled in part by the thickness of supra-bedrock
639 unconsolidated aquifer. As noted above, annual temperature metrics indicated a consistently
640 deeper groundwater discharge influence along Staunton River, while Paine Run had annual

641 signal metrics that mainly indicated reduced and/or more shallow groundwater influence (Figure
642 9a). Long-term streamflow and baseflow analysis from these streams showed Staunton River had
643 higher, but more stable summer discharge (Table 1), and substantially higher median summer
644 BFI (0.62 vs 0.46), indicating greater dominance of groundwater as a generator of streamflow
645 compared to runoff and quickflow. Previous research in SNP used paired air/stream water
646 temperature records, precipitation, and landscape characteristics to statistically model
647 ‘groundwater influence’ by year on a scale of 0-1 at the 100-m scale along the streams of this
648 study, where details are described by Johnson et al., (2017). Although this previous work only
649 extended to 2015, that year had analogous BFI scores to 2019 for Staunton River (0.88 vs 0.84)
650 and Paine Run (0.60 vs 0.58). Comparing the 2019 drying survey observations to the 2015 high
651 spatial resolution modeling of groundwater influence we found that Paine Run was predicted to
652 have groundwater influenced tributaries, but along the mainstem, where extensive dewatering
653 was observed, there was substantially reduced modeled groundwater influence compared to the
654 mainstem of Staunton River (Figure 11). Johnson et al., (2017) also found a negative relation
655 between valley bottom width and their metrics of groundwater influence on SNP streams. In the
656 context of our finding that bedrock depth is negatively related to valley bottom width, we find
657 further support for the hypothesis that thicker headwater stream valley sediments are influential
658 to baseflow generation in low permeability bedrock settings.

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663 *Figure 11. The 2019 stream dewatering survey data (lines; this study), plotted offset of the*
 664 *mainstem, and 100-m groundwater influence predictions (points from Johnson et al., 2017),*
 665 *plotted along the mainstem and tributaries, of Staunton River and Paine Run.*

666 This observation and model comparison represents another line of evidence that groundwater
 667 connectivity at the sub-reach scale is key in determining whether local increases in depth to
 668 bedrock drive channel dewatering at low flow. The impact of reduced underflow groundwater
 669 supply on stream disconnection is likely exasperated by the extensive zones of exposed bedrock
 670 along Paine Run (Figure 4, 7d), which locally reduce groundwater mounding in stream valley
 671 sediments as shown conceptually in Figure 1b, such that abrupt increases in bedrock depth cause
 672 stream dewatering. Among the eight streams investigated here, Staunton River likely represents
 673 the most resilient summer cold water habitat, which could not be predicted using bedrock depth
 674 data alone but necessitated paired assessment of groundwater discharge dynamics.

675 **6 Conclusions**

676 In steep mountain valley stream systems underlain by low-permeability bedrock, the
677 longitudinal underflow reservoir serves as a complex mechanism of streamflow generation,
678 streamflow losses, and stream temperature control (Figure 1, Supplementary Figure S1). Our
679 study utilized complimentary geophysical, temperature, and hydrologic data at the scale of eight
680 subwatersheds to highlight apparent tradeoffs in bedrock depth, shallow groundwater supply, and
681 the quality of cold-water habitat. Certain mountain stream corridor parameters may be
682 reasonable to assume or infer from high-resolution topographic data, such as surficial sediment
683 permeability (based on land surface roughness) and stream valley width, which are primary
684 controls on whether underflow serves as a net source or sink of stream water (Flinchum et al.,
685 2018; Ward et al., 2018). However, as shown here, advances in predicting hydrologic
686 connectivity and thermal variation along mountain stream networks may also require local
687 evaluation of bedrock depth and stream-groundwater exchange.

688 When local increases in bedrock depth are not balanced by groundwater inflow, streams
689 may be expected to dewater and disconnect under low flow conditions, and streams with reduced
690 deep groundwater influence or shallower-sourced groundwater show warmer summer
691 temperatures. Contrary to what might be expected, we found that mean summer stream
692 temperature was not significantly related to elevation at all lower study boundaries, but instead
693 was (negatively) related to average stream bedrock depth. Staunton River had the coldest
694 summer stream temperatures and most pronounced deeper groundwater signatures. However,
695 that subwatershed was of relatively small total surface area and average valley bottom width.
696 The defining physical feature of Staunton River was that it had the largest average bedrock depth
697 of all the eight SNP study streams at 3.4 m, allowing greater overall storage of recharge and

698 baseflow generation. The other two gaged streams had substantially reduced baseflow indices,
699 indicating streamflow generation was dominated by runoff and quickflow.

700 Overall, SNP streams tended to have consistently shallow bedrock depth, though a subset
701 were more variable or had spatial trends and discrete features. Observed channel dewatering
702 patterns during late summer baseflow periods were related to local scale variation in bedrock
703 depth, such as a discrete feature of greater than 20 m depth observed along Piney River that
704 caused repeated streamflow disconnection. However, in other streams more subtle bedrock depth
705 variation also caused channel dewatering, indicating the importance of local hydrogeological
706 context in determining the importance of bedrock depth on streamflow connectivity. For
707 example, patchy 2-4 m deposits of sediment adjacent to exposed bedrock along Paine Run
708 caused extensive summer dewatering in 2019 and 2021, and during the latter survey many dead
709 brook trout were noted in the disconnected sections. Paine and Piney also showed enhanced
710 dewatering during the summers of 2019 and 2021 compared to the wetter 2016 summer,
711 demonstrating the additional control of recent precipitation on stream disconnection in headwater
712 systems that do not efficiently store water.

713 Lateral groundwater inflow through high permeability, unconsolidated sediments is a
714 critical component of headwater stream baseflow (Tran et al., 2020). Shallow, low permeability
715 bedrock can constrain lateral flowpaths and underflow to the near surface critical zone, where it
716 is highly sensitive to enhanced evapotranspiration, temperature increase, and drought under
717 climate change (Condon et al., 2020; Hare et al., 2021). As it becomes increasingly important to
718 understand and predict the resilience of mountain cold-water stream habitat at a fine spatial
719 grain, continued coupled advances in geophysical characterization, stream temperature
720 monitoring, and groundwater exchange analysis are needed.

721 **Data Availability**

722 The data described in this manuscript are available at: doi.org/10.5066/F7B56H72,

723 doi.org/10.5066/F7JW8C04, and doi.org/10.5066/P9IJMGIB

724 **Author contribution**

725 *Conceptualization*: M.A. Briggs, Z.C. Johnson, C.D. Snyder, N.P. Hitt; *Investigation*: M.A.
726 Briggs, P. Goodling, Z.C. Johnson, C.D. Snyder, K.M. Rogers, N.P. Hitt; *Visualization*: M.A.
727 Briggs, K.M. Rogers, P. Goodling, J.B. Fair, C.D. Snyder. All authors contributed to the formal
728 analysis and varied stages of writing.

729 **Competing interests**

730 The authors declare that they have no conflict of interest.

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737 product names is for descriptive purposes only and does not imply endorsement by the U.S.
738 Government.

739

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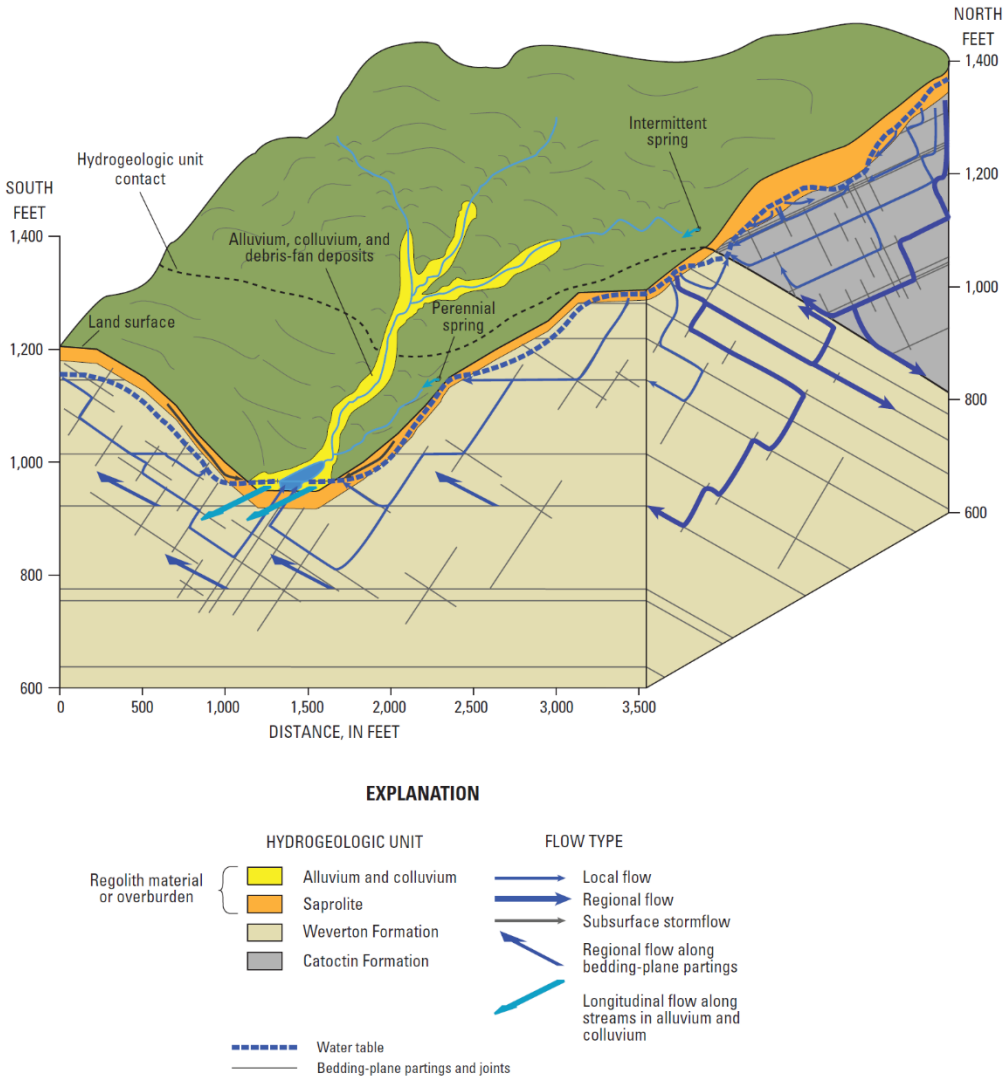
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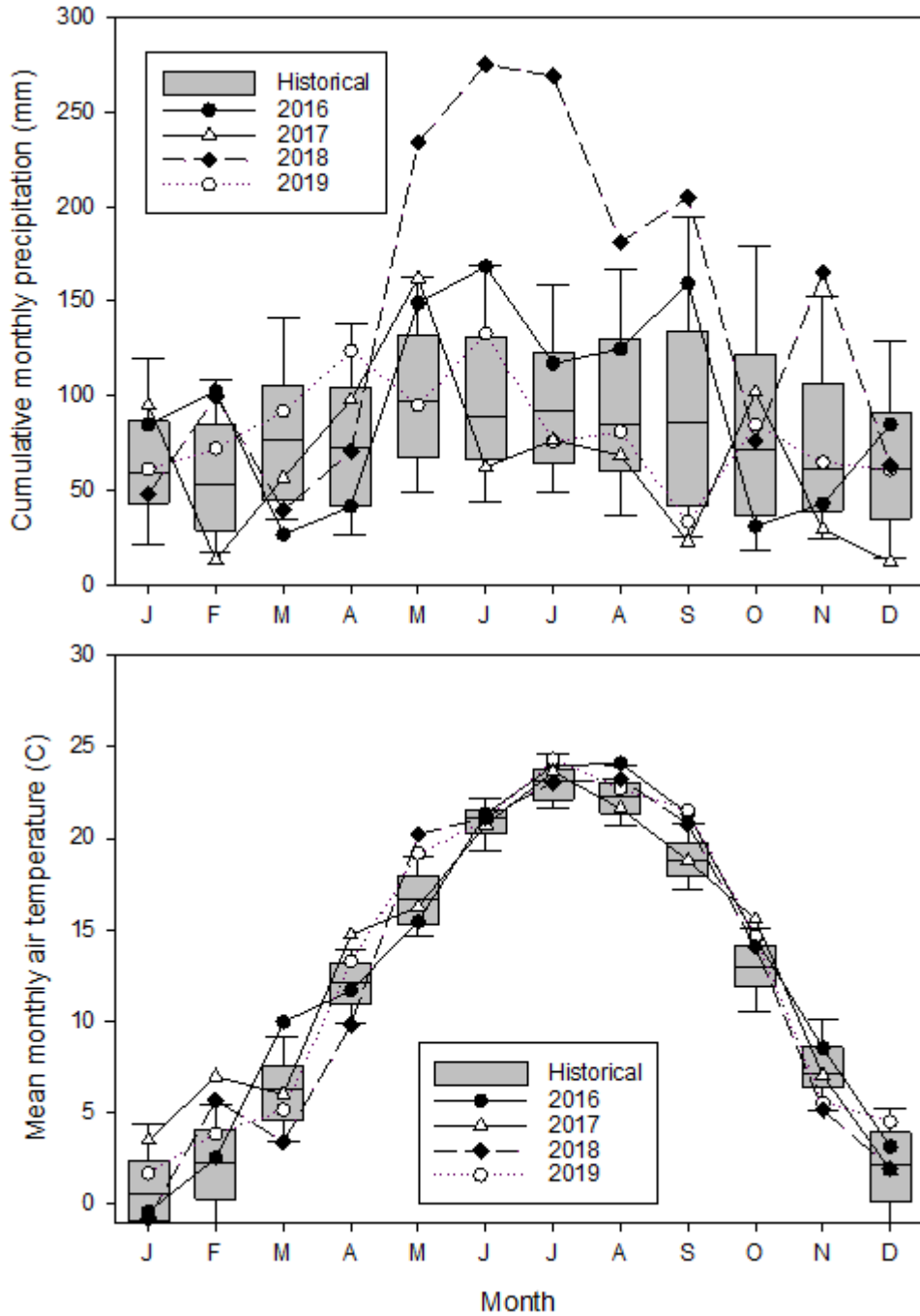
927 **Appendix A**



928

929 Figure A1. The headwater streams of Shenandoah National Park, Virginia, USA are expected to
930 flow over coarse alluvium and colluvium and have connectivity to shallow hillslope groundwater
931 and underflow, but reduced connectivity to deeper bedrock groundwater (Modified Figure 26 in
932 (Nelms and Moberg, 2010) *U.S. Geol. Surv. Investigations Rep. 2010–5190*.

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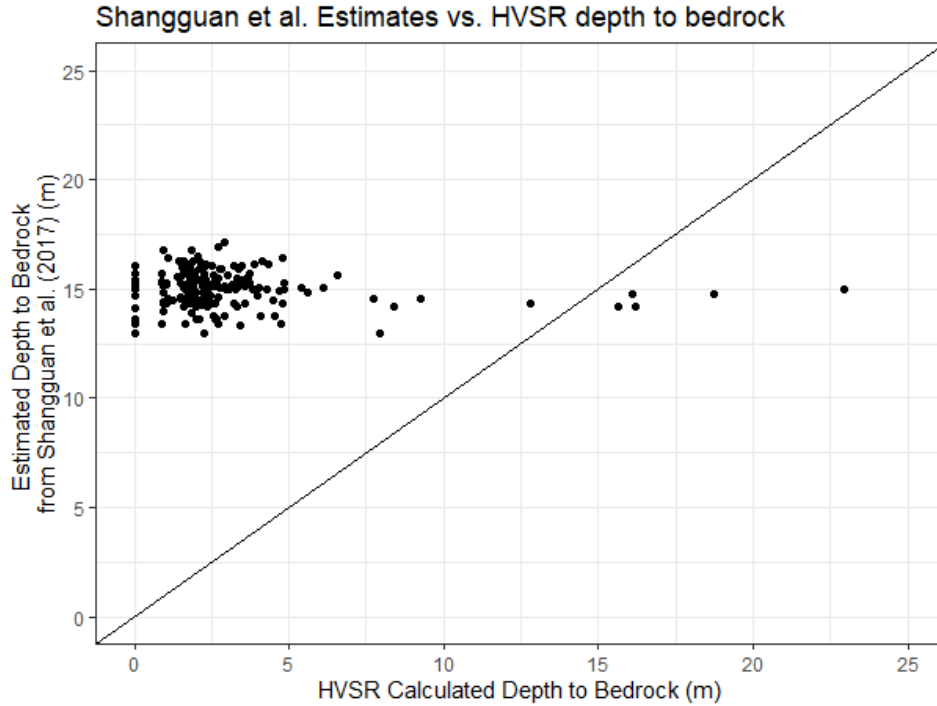
935 Figure A2. Monthly precipitation and air temperature data derived from the Luray weather
 936 station (GHCND:USC00445096) located within Shenandoah National Park. Box plots show the
 937 distribution of values for the period of record (1942-2020) with the limits of the box containing
 938 50% of the values, whiskers containing 90% of the values, and solid line in boxes depicting the
 939 median value. The lines represent values for the four primary study years.

940



941

942 Figure A3. Images from the same vantage point along Paine Run during a) high and b) low flow
943 times, the latter showing channel dewatering associated with a deposit of coarse alluvium across
944 the channel.



945

946 Figure A4: Comparison between bedrock depth modeled for the globe by Shangguan et al.,
 947 (2017) at a 250m resolution and the HVSr-calculated depths to bedrock in this study.

948

949 **Appendix B**

950 Table B1. Summer stream temperature metrics for each study subwatershed determined from the
 951 data set of Snyder et al. (2017), doi.org/10.5066/F7B56H72.

Subwatershed	SiteID	Easting	Northing	Downstream Distance (m)	summer mean (°C)	7-d min (°C)	7-d max (°C)	Stdev (°C)
Hughes	HUR1MP	730038	4276000	242.50	13.43	11.50	16.07	0.98
Hughes	HUR3LCP	731058	4275970	1634.44	15.73	13.23	18.21	1.24
Hughes	HUR5LCP	732278	4275850	3308.93	16.21	13.48	18.62	1.24
Hughes	HUR6MP	733348	4275060	5163.46	16.39	13.86	17.95	1.13
Hughes	HUR12MP	733698	4274880	5620.73	16.79	14.09	18.34	1.24
Hughes	HUR8LCP	733988	4274619	6219.50	18.80	15.19	21.49	1.57
Hughes	HUR9LCP	733968	4274529	6284.35	17.59	14.49	20.09	1.34
Hughes	HUR10MP	734928	4273520	8187.04	18.05	15.01	20.05	1.40
Hughes	HUR13MP	735258	4273330	8667.16	18.68	15.15	21.18	1.62
Hazel	HZR1MP	735158	4278560	707.45	16.80	13.26	19.75	1.46
Hazel	HZR3LCP	735498	4278760	1190.18	16.74	13.08	19.41	1.50
Hazel	HZR11MP	736378	4279640	2951.89	18.16	15.34	20.32	1.52

Hazel	HZR5MP	736638	4279790	3331.66	17.59	13.59	20.62	1.67
Hazel	HZR6MP	737498	4279059	5095.01	18.16	14.03	21.40	1.74
Hazel	HZR7MP	738048	4277990	6820.13	18.48	14.50	21.77	1.72
Hazel	HZR9MP	738368	4277620	7478.33	18.50	14.74	21.72	1.63
Jeremys	JR1MP	734618	4293430	102.97	15.48	12.42	18.74	1.41
Jeremys	JR2MP	733908	4293130	1268.08	16.49	13.38	19.42	1.38
Jeremys	JR4MP	732498	4292250	3699.53	16.84	14.23	18.22	1.11
Jeremys	JR5MP	731778	4290670	5961.22	17.54	14.15	20.24	1.43
Jeremys	JR13MP	731498	4289490	7506.87	18.16	14.57	21.28	1.67
Jeremys	JR7MP	730068	4288080	10327.83	18.76	16.79	21.07	1.10
Jeremys	JR9LCP	729888	4288080	10539.49	17.78	15.33	20.01	1.04
Jeremys	JR12MP	728758	4288080	12030.09	18.61	14.46	22.08	1.73
Jeremys	JR10MP	727758	4288440	13376.47	19.55	14.93	23.55	1.98
Meadow	MR0MP	695318	4228150	0.00	14.16	12.01	15.57	1.07
Meadow	MR1MP	695038	4227980	217.46	16.82	13.82	18.39	1.28
Meadow	MR2MP	694678	4227520	979.43	17.11	13.71	18.78	1.48
Meadow	MR9MP	693488	4227270	2757.87	18.10	15.34	19.71	1.31
Meadow	MR4LCP	693428	4227240	2854.69	17.01	13.92	19.02	1.44
Meadow	MR8MP	693078	4226450	4036.50	17.53	14.32	19.48	1.35
Meadow	MR6LCP	692918	4226170	4446.20	17.08	14.34	19.32	1.29
Meadow	MR7MP	691738	4225700	6209.68	18.37	15.33	20.44	1.44
Paine	PAR1MP	696938	4232031	249.71	16.86	13.96	18.72	1.36
Paine	PARB1	696718	4231390	1115.08	17.20	14.81	18.70	1.15
Paine	PAR2MP	696468	4231210	1542.16	17.15	15.22	18.61	1.03
Paine	PAR3MP	695685	4230400	3169.18	17.48	14.93	19.28	1.15
Paine	PAR5LCP	695369	4230040	3861.10	17.87	15.01	19.53	1.32
Paine	PAR9MP	694568	4229850	5016.00	18.04	15.43	19.60	1.12
Paine	PAR6MP	694218	4229700	5563.29	18.39	14.86	20.32	1.52
Paine	PAR10MP	694068	4229730	5829.23	18.62	14.71	20.50	1.65
Paine	PARB2	693248	4230140	7055.48	18.60	14.54	20.57	1.67
Paine	PAR8MP	693137	4230180	7122.47	18.84	14.50	20.91	1.97
Piney	PIR1MP	736308	4292604	402.61	15.67	12.24	19.16	1.65
Piney	PIR3LCP	736218	4291980	1199.93	16.43	12.76	19.78	1.65
Piney	PIR4MP	735598	4291160	2480.47	16.55	13.24	19.65	1.51
Piney	PIR5MP	735458	4290050	3955.00	16.82	13.62	19.97	1.48
Piney	PIR6MP	736408	4289180	5862.79	17.74	15.87	20.49	1.15
Piney	PIR7MP	736748	4288300	7115.97	17.40	15.07	20.22	1.17
Piney	PIR8MP	737538	4287390	8756.79	17.88	14.63	20.55	1.39
Staunton	SR1MP	725248	4260810	477.07	14.64	11.96	17.33	1.32
Staunton	SR2MP	725908	4260450	1412.13	15.16	12.34	18.15	1.44
Staunton	SR5MP	726948	4259890	2907.57	15.92	13.03	19.08	1.51

Staunton	SR6MP	728018	4259921	4398.87	16.45	13.48	19.38	1.48
Staunton	SR10MP	728598	4259660	5220.08	17.12	13.88	19.84	1.48
Staunton	SR7MP	728718	4259390	5627.21	17.09	13.99	20.02	1.52
Staunton	SR9MP	729448	4258420	7519.72	17.41	14.57	19.88	1.33
White Oak	WOC1MP	728788	4273701	469.03	13.51	11.85	14.91	0.74
White Oak	WOC3MP	728998	4273160	1237.37	15.43	12.67	17.90	1.29
White Oak	WOC4MP	729268	4272400	2307.96	15.71	12.52	18.58	1.48
White Oak	WOC5MP	730288	4271180	4428.05	17.90	14.23	21.16	1.77
White Oak	WOC7LCP	730758	4270690	5302.94	18.69	15.02	22.07	1.79
White Oak	WOC8MP	730948	4269150	7356.87	18.29	15.88	19.78	1.07
White Oak	WOCB	731018	4269110	7448.09	18.71	16.04	21.18	1.28

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