




1 The importance of non-stationary multiannual periodicities in the NAO index for forecasting

2 water resource extremes 

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14 **Abstract**

15 Drought forecasting and early warning systems for water resource extremes are increasingly
16 important tools in water resource management, particularly in Europe where increased
17 population density and climate change are expected to place greater pressures on water
18 supply. In this context, the North Atlantic Oscillation (NAO) is often used to indicate future
19 water resource behaviours (including droughts) over Europe, given its dominant control on
20 winter rainfall totals in the North Atlantic region. Recent hydroclimate research has focused
21 on the role of multiannual periodicities in the NAO in driving low frequency behaviours in
22 some water resources, suggesting that notable improvements to lead-times in forecasting
23 may be possible by incorporating these multiannual relationships. However, the importance
24 of multiannual NAO periodicities for driving water resource behaviour, and the feasibility of
25 this relationship for indicating future droughts, has yet to be assessed in the context of
26 known non-stationarities that are internal to the NAO and its influence on European
27 meteorological processes. Here we quantify the time-frequency relationship between the
28 NAO and a large dataset of water resources records to identify key non-stationarities that
29 have dominated multiannual behaviour of water resource extremes over recent decades.
30 The most dominant of these is a 7.5-year periodicity in water resource extremes since
31 approximately 1970 but which has been diminishing since 2005. Furthermore, we show that



32 the non-stationary relationship between the NAO and European rainfall is clearly expressed
33 at multiannual periodicities in the water resource records assessed. These multiannual
34 behaviours are found to have modulated historical water resource anomalies to an extent
35 that is comparable to the projected effects of a worst-case climate change scenario.
36 Furthermore, there is limited systematic understanding in existing atmospheric research for
37 non-stationaries in these periodic behaviours which poses considerable implications to
38 existing water resource forecasting and projection systems, as well as the use of these
39 periodic behaviours as an indicator of future water resource drought.

40

41 1. Introduction

42 Oscillatory ocean-atmosphere systems (such as El Nino Southern Oscillation (ENSO), North
43 Atlantic Oscillation (NAO) and Pacific Decadal Oscillation (PDO)) are known to modulate
44 hydrometeorological processes over a large domain, often driving multiannual periodicities in
45 hydrological records (Kuss and Gurdak, 2014; Labat, 2010; Trigo et al., 2002). As such,
46 indices of these systems can be useful when explaining decadal-scale variations in water
47 resource behaviour in Europe (Svensson et al, 2015; Kingston et al, 2006), North America
48 (Coleman and Budikova, 2013) and Asia (Gao et al, 2021). In the North Atlantic region, the
49 NAO represents the principal mode of atmospheric variability and is a leading control on
50 European winter rainfall totals (Hurrell, 1995; Hurrell and Deser, 2010). As such, many
51 studies have found strong and significant relationships between the winter NAO Index
52 (NAOI) and hydrological variables across Europe (Wrzesinski and Paluszkiwicz, 2011;
53 Brady et al, 2019; Burt and Howden, 2013), leading to the development of seasonal and
54 long-lead forecasting systems of hydrological behaviour (Svensson et al, 2015, Bonaccorso
55 et al, 2015).

56 A growing number of studies have identified stronger relationships between the NAOI and
57 certain water resource variables at multiannual periodicities (Holman et al, 2011; Neves et



58 al, 2019; Uvo et al, 2021), than at an annual scale. This is particularly apparent where longer
59 hydrological response times predominate (Rust et al 2021a). For instance, Neves et al
60 (2019) identified significant relationships between the NAOI and groundwater level in
61 Portuguese aquifers and at approximately 6- and 10-year periodicities, with associations to
62 episodes of recorded groundwater drought. Furthermore, Liesch and Wunsch (2019) found
63 significant coherence between NAOI and groundwater level at approximately 6- to 16-year
64 periodicities across the UK, Germany, Netherlands and Denmark. Rust et al (2019; 2021a)
65 identified a similar significant 6- to 9-year cycle across a large dataset of groundwater level
66 (59 boreholes) and streamflow (705 gauges) in the UK, which was associated with the
67 principal periodicity of the NAO (of a similar length (Hurrell et al., 2003; Zhang et al., 2011)).
68 In the instance of groundwater level, this periodicity was found to represent a notable portion
69 of overall behaviour (40% the standard deviation), and minima in the cycle were shown to
70 align with recorded instances of wide-spread groundwater drought (Rust et al, 2019). Given
71 their association with recorded droughts across Europe, these studies highlight the potential
72 benefit of an *a priori* knowledge of multiannual NAO periodicities in water resources for
73 improving preparedness for water resource extremes in Europe. Here we use extremes to
74 describe water resource deficit (i.e., drought) and periods of anomalously high water
75 resource stores. This is distinct from hydrological extremes, which infers the drought – flood
76 continuum.

77 However, the value of a multiannual relationship between the NAO and European water
78 resources has yet to be assessed in the context of reported non-stationarities in
79 hydroclimate systems. For instance, the NAO is an intrinsic mode of atmospheric variability
80 (Deser et al, 2017), but can also be influenced by multiple other teleconnection systems
81 such as the Madden-Julien Oscillation, Quasi-Biennial Oscillation (Feng et al 2021) or El-
82 Nino Southern Oscillation (Zhang et al, 2019). As such it is currently unclear whether
83 periodicities in the NAOI are emergent behaviours or the result of external forcing. This has
84 been compounded by a relatively weak signal-to-noise ratio for NAO periodicities, making



85 confident multiannual signal detection difficult (O'Reilly et al, 2018; Hurrell et al, 1997). While
86 stronger NAO-like multiannual periodicities have been detected in water resource variables,
87 due to the high-band filtering function of hydrological processes (van Loon, 2013), the
88 degree to which these behaviours are sufficiently stable to enable development of predictive
89 utilities is currently unclear. Furthermore, existing research has shown that the sign of the
90 relationship between NAOI and European rainfall is non-stationary at decadal timescales
91 (Rust et al, 2021b); Vicente-Serrano and López-Moreno (2008)). This is expected to add a
92 degree of uncertainty to the detection of lead times between multiannual periodic
93 components in the NAO and water resource response, which is necessary in the
94 development of early warning systems for water resource extremes. While some studies
95 have ascribed lags to this multiannual relationship for European water resources (Neves et
96 al, 2019; Holman et al, 2011), **the extent to which this non-stationarity is present at**
97 **multiannual periodicities has yet to be assessed.**

98 Finally, a critical application of early warning systems for water resource extremes is in the
99 design of drought management regimes for existing and projected climate change (Sutanto
100 et al, 2020). While some studies have quantified the degree of modulation that multiannual
101 ocean-atmosphere systems can have on water resources (Kuss and Gurdak, 2014; Neves et
102 al., 2019; Velasco et al., 2015), few have compared these to the expected modulations from
103 projected climate change scenarios. As such the benefit of incorporating multiannual NAO
104 periodicities into early warning systems for improving preparedness for water resource
105 extremes in climate change scenarios has not been assessed.

106 The aim of this paper is to assess the utility of multiannual relationships between the NAO
107 and water resources for improving preparedness for future water resource extremes. This
108 aim will be met by addressing the following research objectives:

- 109 1. Quantify significant covariances between multiannual periodicities in the NAOI and
110 water **resource extremes**, and assess the extent to which these periodicities are
111 stable over time



- 112 2. Assess multiannual periodicity phase differences between the NAOI and water
113 resources over time, to understand the extent to which annual-scale non-
114 stationarities between the NAO and **European rainfall** are expressed at multiannual
115 scales
- 116 3. Quantify the modulations of water resource variables caused by key multiannual
117 periodicities in **the NAO, during the dry season**, and compare this with projected
118 modulations of water resources due to climate change.

119 These objectives will be implemented on UK water resource records, given the considerable
120 coverage of recorded water resource data in time and across the space (Marsh and
121 Hannaford, 2008); however, the methodologies developed can be applied to any regions.

122

123 2. Data

124 2.1. Water resource data

125 The National Groundwater Level Archive (NGLA) and National River Flow Archive (NRFA)
126 provide high-resolution spatiotemporal coverage of groundwater level records and
127 streamflow across the UK.

128 2.1.1. Groundwater data

129 Monthly NGLA groundwater level data from 136 boreholes covering all of the major UK
130 aquifers, with record lengths of more than **20 years and** data gaps no longer than 24 months,
131 have been used (Figure 1). While some meta-analysis was conducted on monthly data, the
132 primary analysis was undertaken on seasonally averaged data, meaning a data gap of no
133 more than two points. They cover a range of unconfined and confined consolidated aquifer
134 types and have been categorised into generalised aquifer groups of Chalk (78 sites),
135 Limestone (12 sites), Oolite (12 sites), Sandstone (34) and variably cemented mixed clays
136 and sands (Lower Greensand Group, Allen et al., 1997) (3 sites). Given the spatially
137 heterogenous response of the Chalk aquifer to droughts (Marchant and Bloomfield, 2018),



138 Chalk sites have been subdivided into four groups based on aquifer region: Lincolnshire basin
139 (8 sites), East Anglian basin (17 sites), Thames and Chiltern basin (29 sites) and Southern
140 basin (21 sites) (Allen et al., 1997; Marchant and Bloomfield, 2018).

141 Broad aquifer groups can be described as follows: Chalk, a limestone aquifer comprising of a
142 dual porosity system with localized areas where it exhibits confined characteristics;
143 characterised by fast-responding fracture porosity (Bloomfield, 1996); Oolite characterised by
144 a highly fractured lithology with low intergranular permeability; Sandstone, comprised of sands
145 silts and muds with principle inter-granular flow but fracture flow where fractures persist; and
146 Lower Greensand, characterised by intergranular flow with lateral fracture flow depending on
147 depth and formation (Allen et al, 1997).

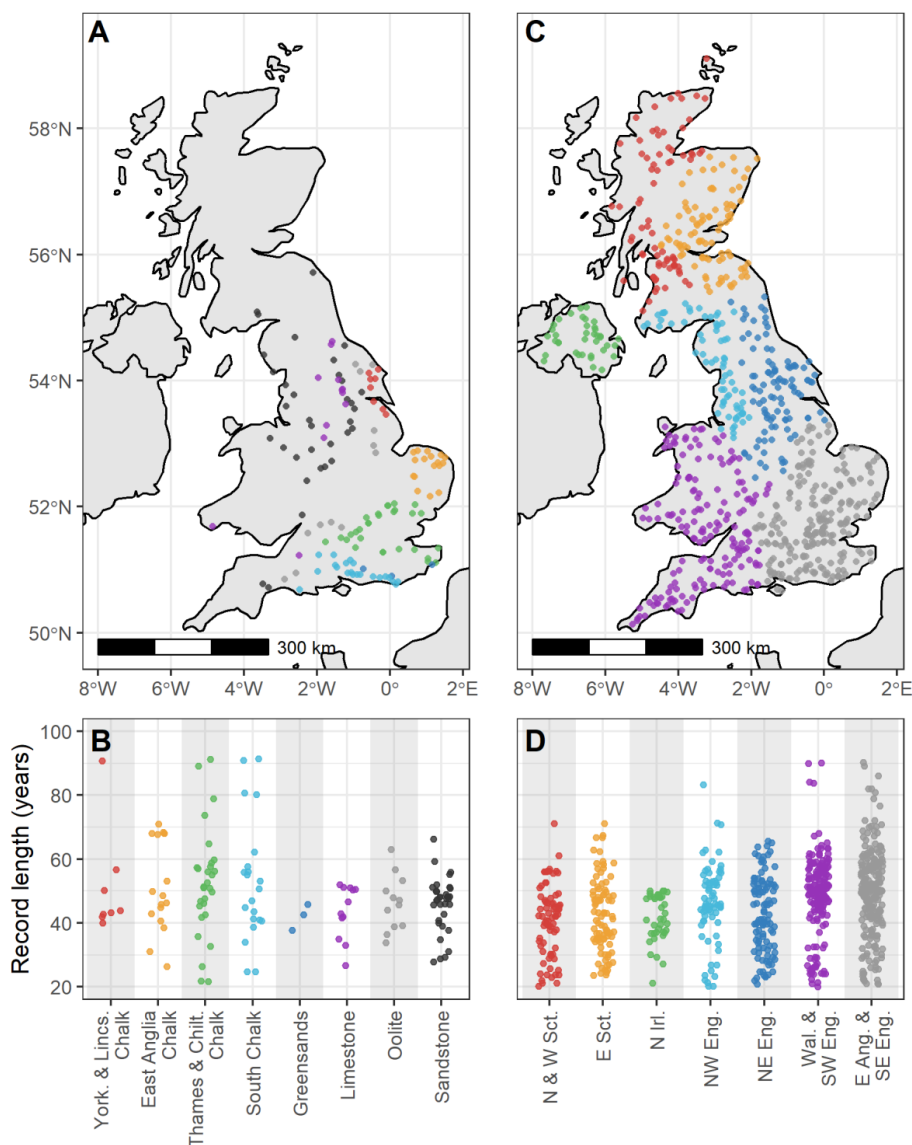
148 2.1.2. Streamflow data

149 Monthly streamflow data from the UK National River Flow Archive (NRFA; Dixon et al., 2013:
150 <http://nrfa.ceh.ac.uk/>) has been used. Gauging stations with more than 20 years of continuous
151 streamflow data and no data gaps greater than 24 months were initially selected. Sites serving
152 the largest catchment were selected where there are multiple sites within a single river
153 catchment. This produced a final list of 767 streamflow gauging stations for use. To
154 understand broad spatial relationships across the streamflow dataset, records have been
155 divided into groups based on the NRFA river drainage basin (RDB). These are grouped by
156 seven generalised regions of the UK; North and West Scotland (75 records), East Scotland
157 (89 records), Northern Ireland (38 records), North-west England (70 records), North-east
158 England (102 records), Wales & South-west England (170 records), East Anglia & South-east
159 England (223 records). Streamflow with minimal influence from human factors is often used
160 in hydroclimate studies to avoid confounding mechanisms, however no such large-scale
161 dataset exists for the UK. Furthermore, over the period of analysis and the broad scale of this
162 assessment, inconsistencies in the way water resource management practices are
163 implemented is expected to result in noise to the observations rather than some systematic
164 signal or bias that would affect the results of this paper.



165 2.2. North Atlantic Oscillation data

166 Monthly North Atlantic Oscillation Index (NAOI) data calculated by the National Centre for
167 Atmospheric Research (NCAR) using the principal component (PC) method for the period
168 1989 – 2021 has been used. The PC NAOI is a time series of the leading empirical orthogonal
169 functions (EOFS) of sea level pressure grids across the north Atlantic region (20°-80°N, 90°W-
170 40°E).



171

172 Figure 1 – Spatial and temporal distributions of water resource records; a) location of
173 groundwater boreholes coloured by associated aquifer group, b) jitter plot of groundwater
174 record lengths within each aquifer group, c) location of streamflow gauges coloured by
175 associated regional group, d) jitter plot of streamflow record lengths within each regional group

176

177



178 **3. Methods**

179 **3.1. Data Pre-processing**

180 In this study we use the continuous and cross-wavelet transform to understand behaviours
181 and relationships across different periodicities within the different water resource variable time
182 series.

183 For all datasets, gaps less than two years were infilled to a monthly time step using a cubic
184 spline to produce a complete time series for the wavelet transform. For time series with gaps
185 greater than two years, the shortest time period before or after the data gap was removed.

186 The records were not trimmed to obtain a common period of data coverage. **Instead, all data**
187 **was trimmed to start at a minimum of 1930.** This was to allow the analysis of the fewer records
188 that cover a longer time period while still capturing a time periods with adequate record
189 coverage. All of the time series were standardised by dividing by their standard deviation and
190 subtracting their mean.

191 **3.2. Quantifying wide-spread water resource extremes**

192 In order to meet objective 1, we produced a time series which describes the behaviour of
193 wide-spread water resource extremes across each resource variable (i.e., groundwater or
194 streamflow). In this study we have assessed water resource extremes using a drought
195 threshold methodology proposed in Peters (2003). While other measures of drought are
196 available (e.g., Standardised Precipitation Index (SPI) and Standardised Groundwater Index
197 (SGI)) (Bloomfield and Marchant, 2013), a threshold approach has been adopted as its can
198 be easily applied to both streamflow and groundwater variables.

199 To calculate a drought series from monthly groundwater level and streamflow series, we first
200 used the threshold methodology given by equation 4.3 in Peters (2003):

201



$$\int_0^M (x_t(c) - x(t))_+ dt = c \int_0^M (\bar{x} - x(t))_+ dt \quad (\text{Eq. 5})$$

202 Where:

$$203 \quad x_+ = \begin{cases} x & \text{if } x \geq 0 \\ 0 & \text{if } x < 0 \end{cases}$$

204 and M is the full length of the data series. Here we use a threshold level of $c = 0.3$ for
205 groundwater level and $c = 0.01$ for streamflow. Peters et al (2003) found that a value of 0.3
206 for groundwater level was comparable to other commonly used thresholds. A value of 0.01
207 for streamflow was chosen as it produced a similar distribution of drought events as the
208 groundwater drought series. The chosen value of c for either variable is not expected to
209 affect the outcomes of the study as the focus is on the frequency structure of water resource
210 extremes, rather than magnitude.

211 For each measurement site, the monthly time series of drought status (whether in drought
212 according to the threshold criteria or not) was converted into a yearly series describing
213 whether that site experienced a drought in the calendar year. Then, for each year, the
214 number of sites that experienced drought were summed and divided by the number of sites
215 with coverage of that year. This produced a time series of the proportion of sites
216 experiencing drought each year, for groundwater level and streamflow variables. This is
217 referred to as the drought coverage time series.

218 3.3. Frequency Transformations

219 3.3.1. Continuous Wavelet Transform (CWT)

220 The Continuous Wavelet Transform (CWT) was performed on the drought coverage time
221 series for groundwater and streamflow to understand the frequency behaviour of wide-
222 spread water resource extremes over time. The CWT is often used in geoscience to
223 understand non-stationarities of a variable over time and frequency space (Sang, 2013).





224 The cross-wavelet transform, W , consists of the convolution of the data sequence (x_t) with
225 scaled and shifted versions of a mother wavelet (daughter wavelets):

$$W(\tau, s) = \sum_t x_t \frac{1}{\sqrt{s}} \psi * \left(\frac{t - \tau}{s} \right) \quad (\text{Eq. 1})$$

226 where the asterisk represents the complex conjugate, τ is the localized time index, s is the
227 daughter wavelet scale and dt is increment of time shifting of the daughter wavelet. The
228 choice of the set of scales s determines the wavelet coverage of the series in its frequency
229 domain. The Morlet wavelet was favoured over other candidates due to its good definition in
230 the frequency domain and its similarity with the signal pattern of the environmental time
231 series used (Tremblay et al. 2011; Holman et al. 2011).

232 The modulus of the transform can be interpreted as the continuous wavelet power (CWP):

$$P(\tau, s) = |W(\tau, s)| \quad (\text{Eq. 2})$$

233 We use the package “WaveletComp” produced by Rosch & Schmidbauer (2018) for all
234 wavelet transformations in this paper.

235 **3.3.2. Cross-Wavelet Transform (XWT)**

236 The bivariate XWT was applied between the NAOI and each of the water resources records
237 (groundwater level (GWL) and streamflow (SF)). This produces a cross-wavelet power which
238 is analogous to the covariance between the two variables over a time and frequency
239 spectrum. This has been selected over the cross-wavelet coherence (analogous to
240 correlation) as this metric requires a high degree of spectral smoothing, making the resultant
241 coherence spectra sensitive to the choice of smoothing approach (Rosch & Schmidbauer
242 (2018)). Here we use the covariance spectrum to compare against the drought series
243 frequency spectrum to understand where strong coherences are reflective of dominant
244 behaviours in water resource extremes.



245 In order to calculate cross-wavelet power (XWP) for the bivariate case, it is first necessary to
246 calculate the continuous wavelet transform (CWT) for each of the variables separately. The
247 XWT between variables x and y is given by:

$$W.xy(\tau, s) = \frac{1}{s} \cdot W.x(\tau, s) \cdot W.y^*(\tau, s) \quad (\text{Eq. 3})$$

248 The modulus of the transform can be interpreted as the cross-wavelet power (XWP):

$$P.xy(\tau, s) = |W.xy(\tau, s)| \quad (\text{Eq. 4})$$

249

250 3.3.3. Wavelet Significance

251 Lag-1 autocorrelations (AR1) in environmental datasets can produce emergent low frequency
252 behaviours, making the detection of externally-forced behaviours more difficult (Allen and
253 Smith, 1996; Meinke et al., 2005; Velasco et al., 2015). In this study, a significance test was
254 undertaken to test the red-noise null hypothesis that wavelet powers calculated are the result
255 of the recorded variables' AR1 properties. This was based on 1000 synthetic Monte Carlo
256 series with the original AR1 values. In this paper we test significance to the 95% CI.

257 The significance spectra for the XWT for each variable pair (e.g., GWL and NAOI) form the
258 primary results for the XWT method in this paper, since the cross-wavelet power is heavily
259 dependent on the individual series and its frequency composition. The overall relationship
260 between the NAOI and water resources as a whole are investigated by showing the proportion
261 of sites over time and frequency that exhibit a significant relationship with the NAOI (95% CI).
262 This average significance spectrum is produced by summing the significance matrices across
263 each resource (groundwater level or streamflow) and dividing by the number of records used
264 in year each.

265

266 3.3.4. Phase Difference



267 In the bivariate case, the instantaneous phase difference for the XWP spectrum (between
268 wavelets pairs from the CWT spectrum for each variable) can also be calculated as:

$$\text{Angle}(\tau, s) = \text{Arg}(W_{xy}(\tau, s)) \quad (\text{Eq. 5})$$

269

270 This is the difference of the individual phases from both variables at an instantaneous time
271 and frequency (period), converted to an angle between $-\pi$, and π . Values close to 0 indicate
272 the two series move in-phase, with absolute values close to π indicating an out-of-phase
273 relationship. Values between 0 and π indicate degrees of phase difference or phase shift.
274 Phase differences between 0 and π can indicate the degree to which variable x is leading
275 variable y, however a phase difference between 0 and $-\pi$ can either indicate that variable y is
276 leading variable x, or that variable x is leading by more than half the phase rotation (period
277 length). The degree to which a certain variable is leading is analogous to a lag between the
278 two variables.

279

280 3.4. Modulation measurement

281 In order to understand the degree of modulation that the NAO teleconnection has on water
282 resources, an absolute and relative modulation value has been calculated for each series.
283 Here, we use modulation to describe the degree to which the NAO (or other process) has
284 increased or decreased a water resource measure from its mean. This has been derived by
285 reconstructing a specific principal periodicity range from the cross-wavelet powers using the
286 following equation:

$$(x_t) = \frac{dj \cdot dt^{1/2}}{0.776 \cdot \psi(0)} \sum_s \frac{\text{Re}(W(., s))}{s^{1/2}} \quad (\text{Eq. 6})$$

287 Where dj is the frequency step and dt is the time step.



288 This produces a periodic reconstruction of a component of the original dataset that conforms
289 to the set of periodicities (scale steps) selected. The mean and maximum amplitude of this
290 periodic reconstruction was calculated from the absolute values of minima and maxima.
291 Since the data were standardised by dividing by the standard deviation prior to the wavelet
292 transform, this calculated mean and maximum amplitude are also relative to the sd of the
293 original data. Multiplying the calculated amplitude by the original sd converts this back into a
294 real-valued measurement. This was only done for groundwater, since streamflow is highly
295 dependent on catchment size. In the case of streamflow, amplitudes are reported as relative
296 to the standard deviation of the streamflow record. All calculated modulations were produced
297 using reconstructed wavelets from after 1970 where the majority of records are present in
298 both groundwater and streamflow variables. This was done to mitigate the effect of differing
299 record lengths.

300

301 4. Results

302 4.1. Multiannual water resource extremes covariance with NAOI

303 Figure 2 shows the NAOI covariance significance spectrum (fig 2a and 2b) and drought
304 frequency spectrum (fig 2c and 2d) for the groundwater level records. These have been
305 plotted together to allow for easier interpretation and comparison of the results. Black lines in
306 the spectral plots show the 95% CI. The calculated drought series (fig 2e) and record
307 coverage (fig 2f) have also been plotted alongside for comparison.

308 Figure 2a shows the results from the XWT significance testing between the NAOI and the
309 136 groundwater level records. Results are displayed as contours showing the percentages
310 of sites that exhibited a significant (0.05 a) XWP within the time-frequency spectrum. There
311 are five localised regions within the NAOI x GWL XWP spectrum that denote a wide-spread
312 significance between the GWL records and the NAOI. The greatest significance contours of
313 these regions (referred to here as focal points (FPs)) are labelled on figure 2a as: FP 1: 1934



314 at the 4.2 years periodicity (80% of records); FP 2: 1974 at the 8.5 years periodicity (40% of
315 records); FP 3: 1995 at 5.4 years (80% of records); FP 4: 2005 at 7 years (90% of records)
316 and; FP 5: 2012 at 2.9 years (60% of records).

317 These focal points are grouped into three larger regions within the 10% contour; between
318 1933 – 1940 spanning the 3- to 5-year periodicity; 1964 – 2020 spanning the 4- to 12-year
319 periodicity and; 2007 – 2017 spanning the 2- to 4-year periodicity. There is a single peak in
320 the time-averaged percentage plots (figure 2b) at the 7.5-year periodicity (average of 26% of
321 records)

322 Figure 2c shows the results from the CWT of the groundwater drought series (shown in Fig
323 2e). There are five regions of significant wavelet power in the groundwater drought
324 frequency spectrum that are labelled in figure 2c as follows; region 1: 1930 - 1950 in the 4-
325 to 8-year periodicity range (greatest power at 4.8 years); region 2: 1930 – 1945 in the 10- to
326 13-year periodicity range (greatest power at 11.7 years); region 3: 1960 – 1965 in the 2.5- to
327 3.5-year periodicity range (greatest power at 2.8 years); region 4: 1960 – 1990 centred at the
328 12- to 17-year periodicity range (greatest power at 15.4 years); and region 5: 1980 to 2020
329 at the 6- to 8-year periodicity range (greatest power at 7 years). There is a sixth significant
330 region starting in 2019 and covering periods between 2 and 5 years, however this is very
331 close to the end of the record and may be subject to edge effects. As such this region has
332 not been taken forward for discussion.

333 There are also two notable non-significant regions of medium strength wavelet power (\geq
334 0.4); 1930 - 2000 at the 14- to 23-year periodicity range (centred at 16 years), and between
335 1960 and 1970 at the 8- to 16-year periodicity range (centred at 9 years). There are two
336 notable peaks in time-averaged wavelet power for the GWL drought series (figure 2d); the
337 greatest at the 7-year periodicity (average wavelet power of 0.38), and the second at the 14-
338 year periodicity (average wavelet power of 0.24).



339 Figure 3 shows the same as Figure 2 but for the streamflow (SF) case. There are six
340 localised regions within the NAOI x SF XWP spectrum that denote a wide-spread
341 significance between the SF records and the NAOI. FPs of these regions are labelled on
342 figure 2a; FP 1: 1940 at the 6.7-year periodicity (30% of records); FP 2: 1962 at the 5.2-year
343 periodicity (50% of records); FP 3: 1975 at the 8.5-year periodicity (40% of records); FP 4:
344 1994 at the 5.2-year periodicity (80% of records); FP 5: 2007 at the 7-year periodicity (90%
345 of records) and; FP 6: 2011 to 2015 at the 3.2-year periodicity (60% of records). These
346 centres are grouped into larger regions within the 10% contour; these are between 1933 –
347 1947 spanning the 5.5- to 8-year periodicity; 1960 – 1970 spanning the 4- to 8-year
348 periodicity; 1965 – 1990 spanning the 7- to 11-year periodicity; 1988 – 2000 spanning the 4-
349 to 5.5-year periodicity; 1995 – 2020 spanning the 4.5- to 11-year periodicity and 2007 –
350 2017 spanning the 2.5- to 4.5-year periodicity. There is a single peak in the time-averaged
351 percentage plots (figure 3b) at the 7.5-year periodicity (average of 29% of records)

352 Figure 3c shows the results from the CWT of the streamflow drought series (shown in Fig
353 3e). There are three regions of significant wavelet power in the groundwater drought
354 frequency spectrum that are labelled on Figure 3c; region 1: 1930 – 1935 in the 21 year
355 periodicity (this region appears clipped by the record start date, so the strongest wavelet
356 power for this region may not be captured); region 2: 1930 - 1937 in the 2.5- to 6.5-year
357 periodicity range (strongest power at 4.3 years) and; region 3: 1930 – 1960 in the 11- to 15-
358 year periodicity range (strongest power at 13 years);

359 There are four non-significant regions of medium strength wavelet power (≥ 0.4); 1935 –
360 1945 at the 2- to 3-year periodicity; 1955 – 1965 at the 2- to 4-year periodicity; 1960 – 2015
361 at the 5.5- to 8-year periodicity; and 2000 – 2005 at the 2- to 5-year periodicity. The time-
362 averaged wavelet power for the SF drought series (figure 3d) contains multiple peaks
363 suggesting no dominant periodicity. The greatest peak is at the 7-year periodicity with an
364 average wavelet power of 0.21.

365 **4.2. Cross-wavelet phase difference**



366 The cross-wavelet phase difference (ϕ) between water resource variables and the NAOI at
367 the 7.5-year periodicity has been displayed in figure 4 for the GWL records and figure 5 for
368 the streamflow records. The phase difference is a circular measurement where 0 indicates
369 an in-phase relationship (analogous to zero lag) and $\pm\pi$ indicates an out-of-phase
370 relationship between the selected periodicity within the two variables (analogous to half a
371 periodicity lag (3.75-years)). The purpose of these plots of phase differences are to visualise
372 and understand the difference in phase between the NAO and water resources. **Records**
373 have been split by their aquifer group in Figure 4, and by catchment region in figure 5, to
374 understand if there are any general differences between regions.

375 The majority of groundwater level records cover the period 1970 to present, meaning
376 general trends are more clearly presented for this time period. The phase difference of most
377 GWL records can be defined by a sudden shift at approximately 1990 (figure 4). Values of ϕ
378 generally range from between $-1/4\pi$ and $-3/4\pi$ (-0.76 to -2.36 rads; generally anti-phase) for
379 the period 1975 to 1990 to between $+1/4\pi$ and $+3/4\pi$ (0.76 to 2.36 rads; generally in-phase)
380 for the period 1990 to 2019 across all sites. This is with the exception of 17 sites across the
381 South Chalk and Thames & Chiltern Chalk which have shorter ~anti-phase periods (between
382 approximately 1985 and 1990). Average ϕ values for the period 1970 – 1990 (1990 – 2020)
383 for each aquifer region are: -1.26 (1.41) in East Anglian Chalk; -2.25 (1.21) in Lincolnshire
384 Chalk, 0.52 (0.83) in South Chalk, -1.37 (0.83) in Thames & Chiltern Chalk, 1.51 (1.21) in
385 Greensands, -0.78 (0.66) in Limestone, -1.36 (1.09) in Oolite, -0.70 (1.35) in Sandstone. As
386 such most aquifer regions experience an average reversal of polarity at 1990. Greensand
387 GWL show no reversal when assessing average ϕ values, however 1 of the 3 sites in this
388 aquifer group does show this reversal.

389 Similar to the GWL records, most SF records exhibit a shift in phase difference at
390 approximately 1990, with catchment groups in the north of the UK showing minimal shifts
391 (i.e., NW Scotland, E Scotland, NI, and NW England) (figure 5). In the southern catchment
392 groups, values of ϕ generally range from between $-1/2\pi$ and $\pm\pi$ (generally anti-phase) for



393 the period 1970-1990 (approximately prior to the shift) to between 0 and $+3/4\pi$ (generally in-
394 phase) for the period 1990 to 2020 (approximately after the shift). Furthermore, catchment
395 groups in the east of the UK (i.e., E Scotland, NE England, East Anglia & SE England)
396 during the in-phase period (1990-2020) exhibit a notable transition to increased phase
397 difference (to approximately $+3/4\pi$) between 2000 and 2010 before decreasing to
398 approximately $+1/4\pi$ in 2020. Average ϕ values for the period 1970 – 1990 (1990 – 2020)
399 for each catchment region are: -0.21 (0.14) in North and West Scotland, 0.49 (0.86) in East
400 Scotland, -0.43 (0.46) in Northern Ireland, -0.44 (0.47) in NW England, 2.32 (1.08) in NE
401 England, 0.77 (0.64) in Wales and SE England, and 2.53 (0.99) in East Anglia and SE
402 England.

403 **4.3. Modulation of dry season water resources**

404 Figure 6 shows two boxplots for each aquifer group, representing the distribution of mean (in
405 blue) and maximum (in red) dry-season GWL deviation as a result of the 7.5-year periodicity
406 (over the length of each of the record). Median values from each of these mean and
407 maximum boxplots are described below, and are referred to as med.mean and med.max
408 respectively.

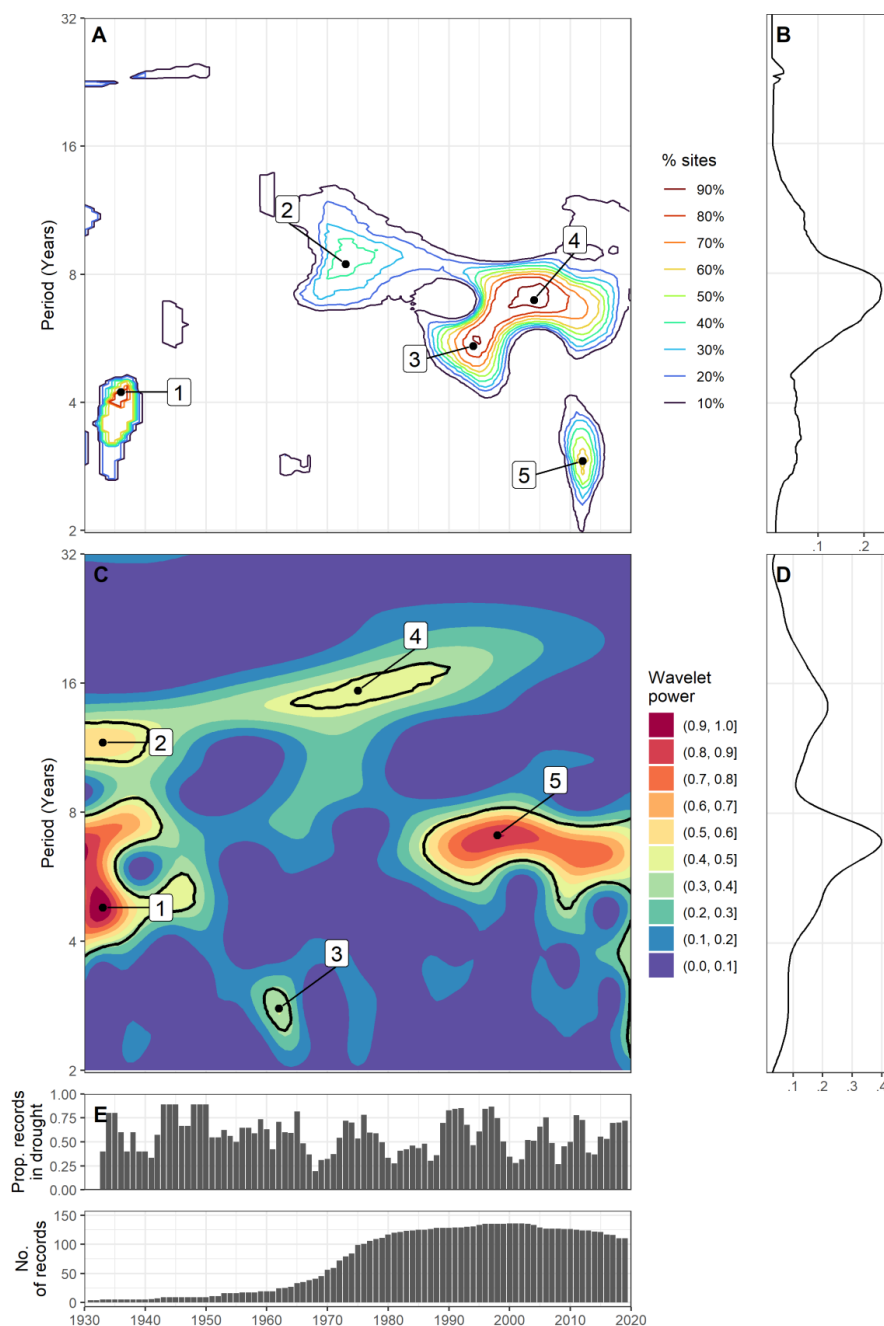
409 The 7.5 year periodicity accounts for the greatest deviation of-dry season GWL in the Chalk
410 aquifer regions, with the Thames & Chiltern basin GWL showing the greatest modulation of
411 all groups showing med.mean of 0.94m and a med.max of 1.38m. Two other Chalk groups
412 showed similarly strong modulations; the South Chalk basin GWL (med.mean: 0.7m,
413 med.max: 1.07m); and the Lincolnshire Chalk GWL (med.mean: .56m, med.max: 0.77m).
414 The East Anglia GWL show lowest modulation of the Chalk (med.mean: 0.16m, med.max:
415 0.34m), similar to GWL in the Limestone (med.mean: 0.35m, med.max: 0.51m) and the
416 Oolite (med.mean: 0.21m, med.max: 0.33m). Lowest overall modulations are found in the
417 Sandstone (med.mean: 0.15m, med.max: 0.25m) and Greensands aquifers (med.mean:
418 0.12m, med.max: 0.17m).



419 Figure 7 shows the same as **figure 6** but for the streamflow case. Streamflow modulations
420 are measured as relative to the standard deviation of each record. Modulation of streamflow
421 for each catchment group are (in descending order of med.mean); Wales & south-west
422 England (med.mean: 0.32, med.max: 0.50); East Anglia & south-east England (med.mean:
423 0.31, med.max: 0.53); Northern Ireland (med.mean: 0.29, med.max: 0.50); West Scotland
424 (med.mean: 0.27, med.max: 0.46); north-east England (med.mean: 0.27, med.max: 0.47),
425 north-west England (med.mean: 0.26, med.max: 0.46), east Scotland (med.mean: 0.21,
426 med.max: 0.39).

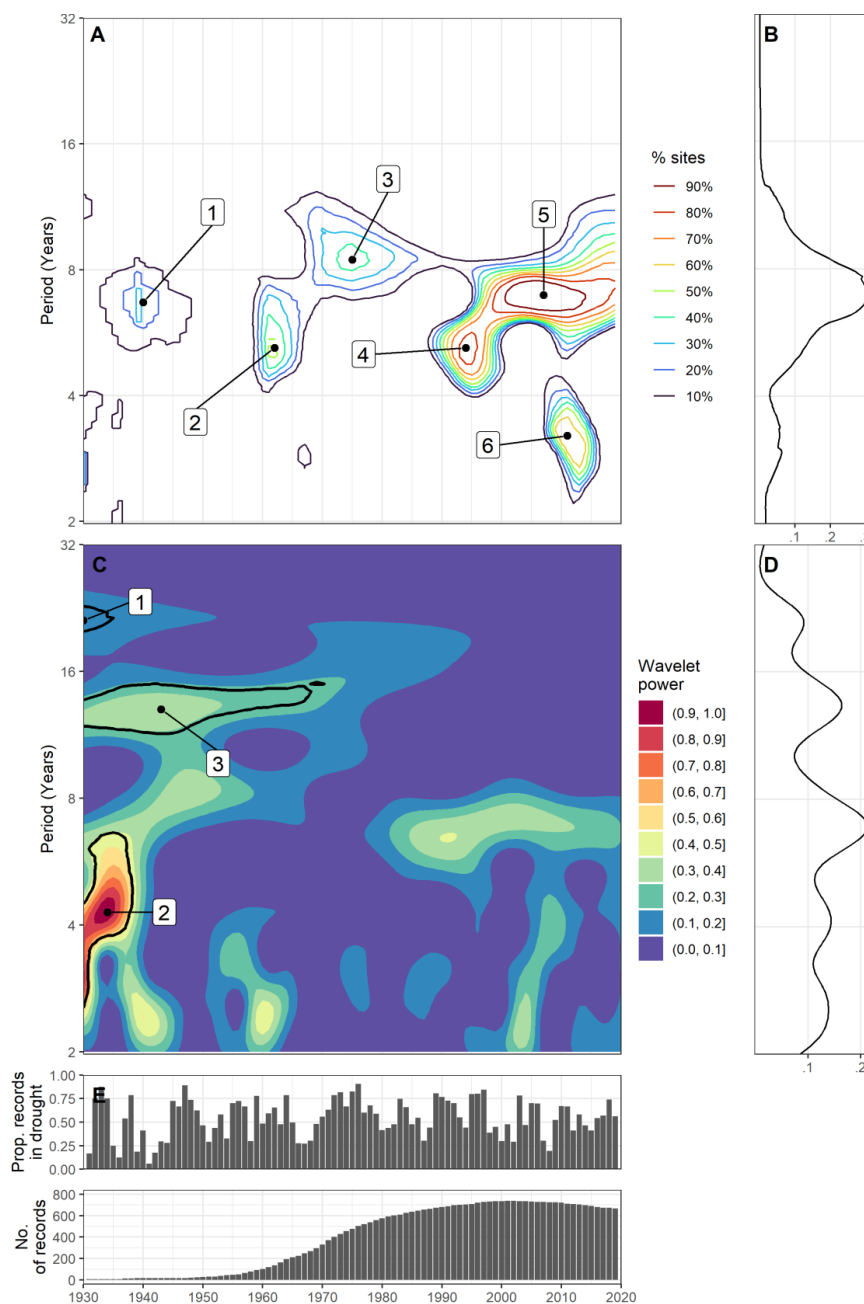
427

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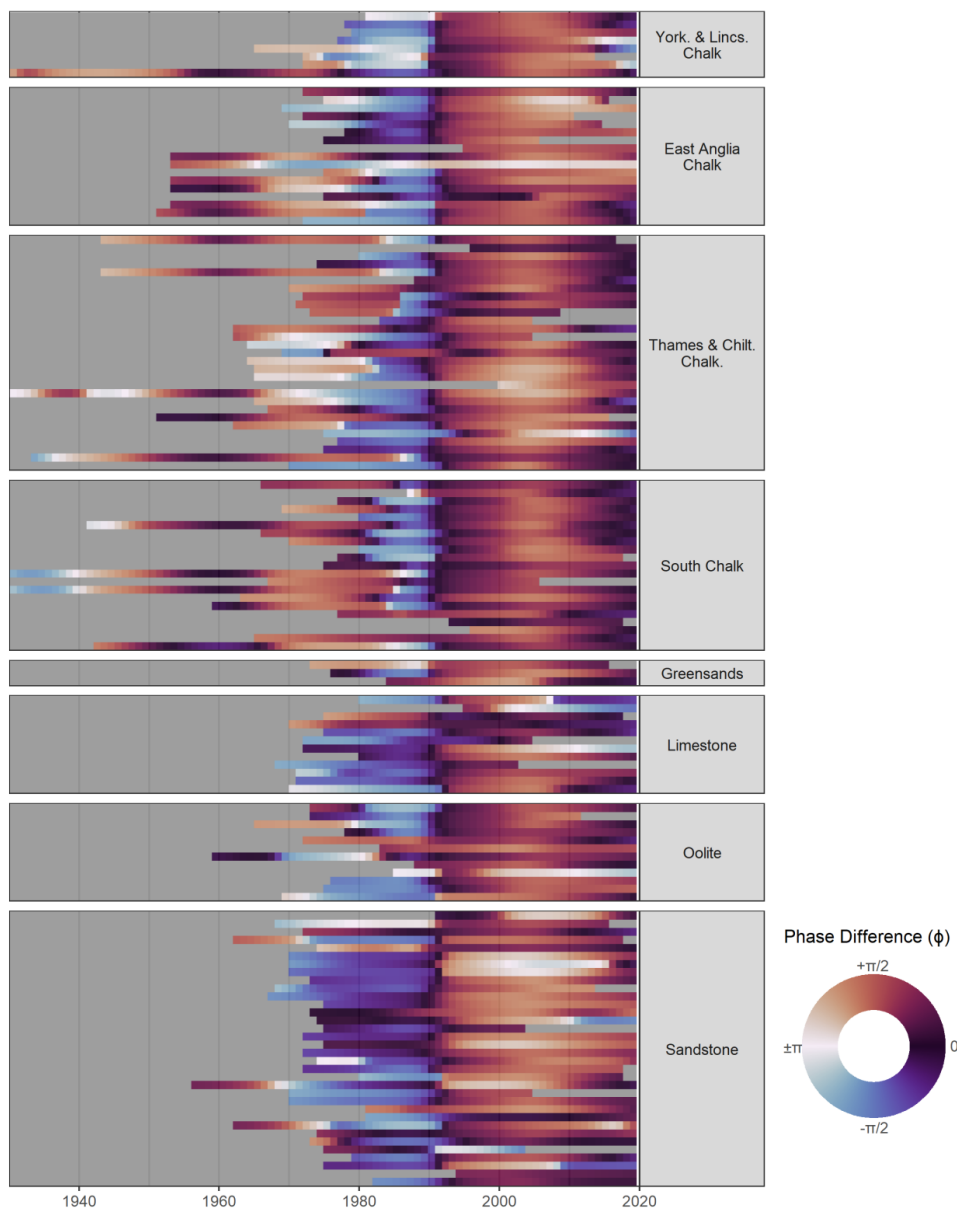
429

430 Figure 2 – a) Significance (95% CI) contours between GWL and NAOI, b) time-averaged
 431 proportion of gwl records with a significant XWP with the NAOI (measured as a decimal
 432 fraction), c) wavelet (spectral) power of GWL drought series, d) time-averaged wavelet
 433 (spectral) power of GWL drought series, e) GWL drought coverage time series, f) temporal
 434 coverage of records.



435

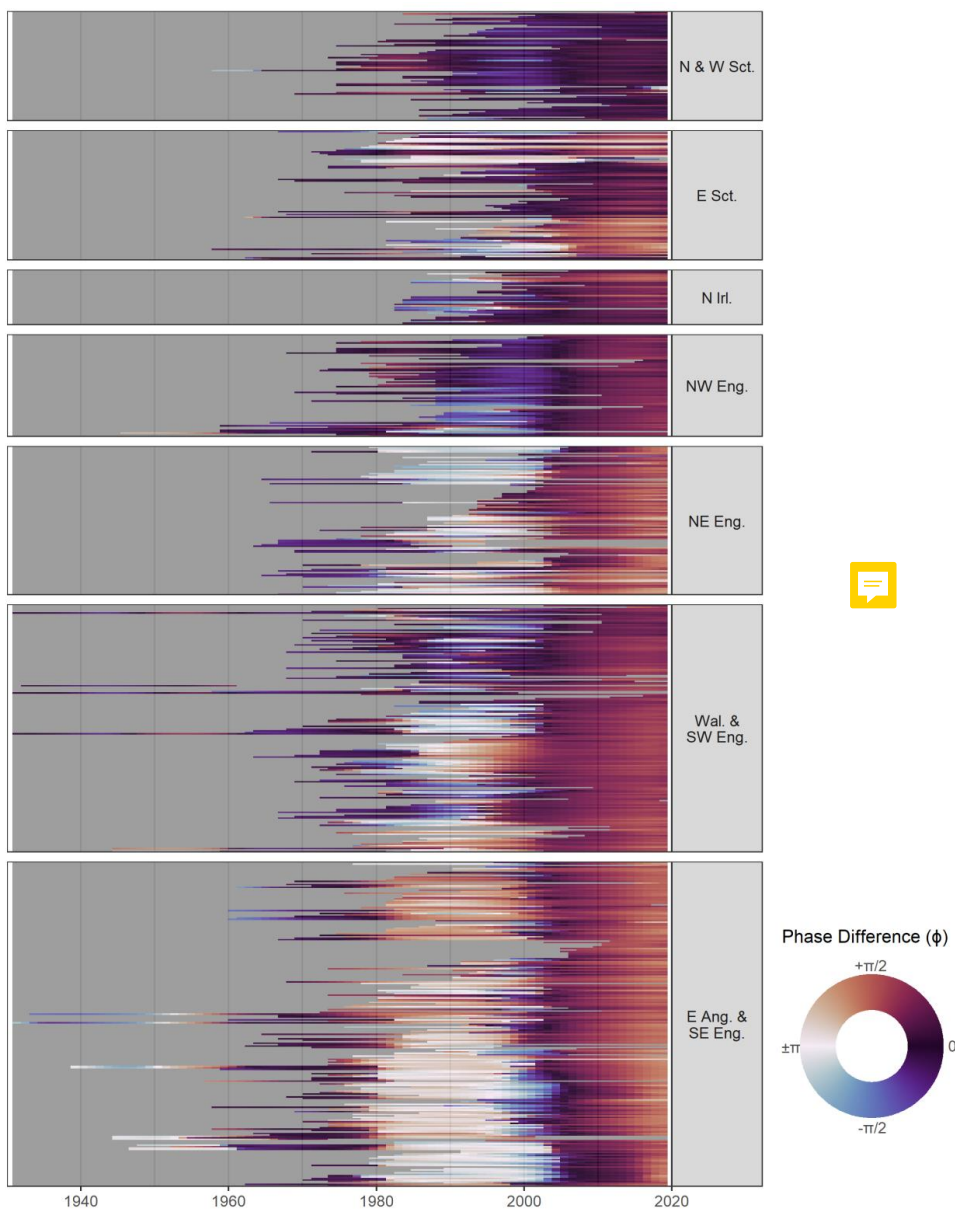
436 Figure 3 – a) Significance (95% CI) contours between SF and NAOI, b) time-averaged
 437 proportion of SF records with a significant XWP with the NAOI (measured as a decimal
 438 fraction), c) wavelet (spectral) power of SF drought series, d) time-averaged wavelet
 439 (spectral) power of SF drought series, e) SF drought series showing proportion of records in
 440 drought each year, f) temporal coverage of records.



441

442 Figure 4 – Phase difference between the NAOI and each GWL record for the GWL record
443 period. Results are grouped by aquifer regions. $\phi = 0$ is equivalent to an in-phase
444 relationship and $\phi = \pm\pi$ is equivalent to an antiphase relationship.

445



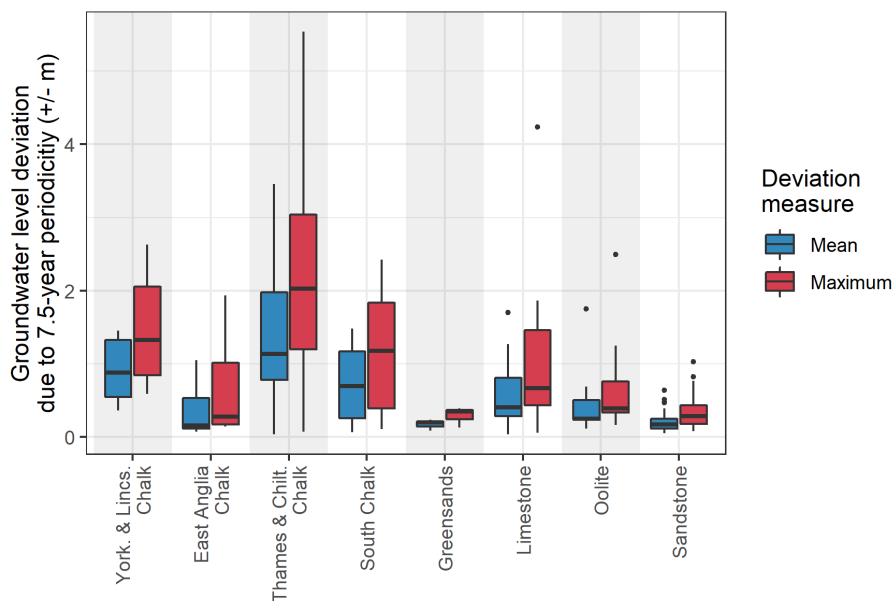
446

447 Figure 5 – Phase difference between the NAOI and each streamflow record for the
448 streamflow record period. Results are grouped by regions. $\phi = 0$ is equivalent to an in-phase
449 relationship and $\phi = \pm\pi$ is equivalent to an antiphase relationship.

450

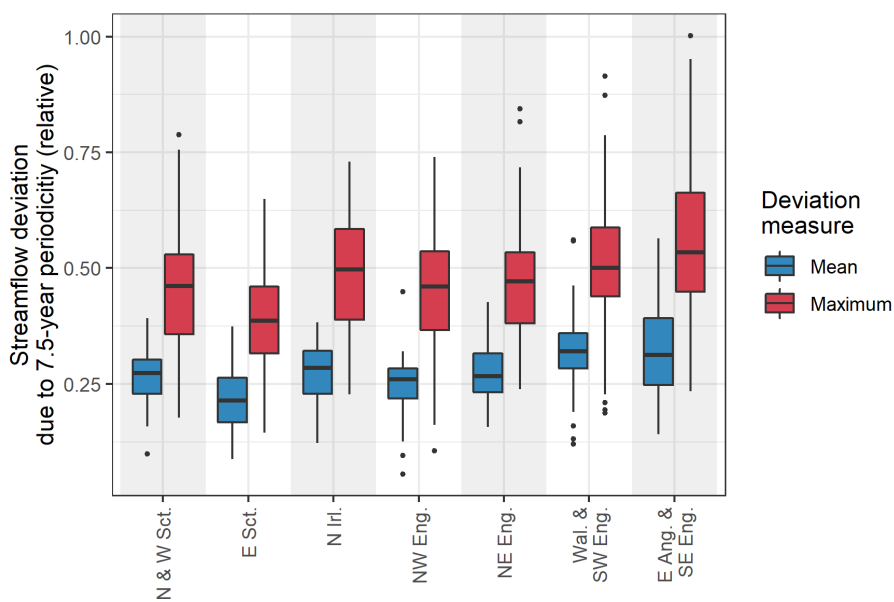
451

452



453

454 Figure 6 – Distribution of absolute mean and maximum modulation of summer groundwater
455 level as a result of the principal cross-wavelet periodicity between the NAOI and winter
456 Groundwater level by aquifer region



457



458 Figure 7 – Modulation of summer streamflow (relative to record standard deviation) as a
459 result of the principal cross-wavelet periodicity between the NAOI and winter streamflow.

460 5. Discussion

461 5.1. Historical covariances between the NAOI and water resources at multiannual 462 periodicities

463 Our results show that the dominant mode of multiannual covariance between the NAOI and
464 UK water resources is at the ~7.5-year periodicity. This is apparent in the time-averaged
465 covariance significance plots for groundwater (figure 2b) and streamflow (figure 3b). The
466 same 7.5-year periodicity is also the strongest average mode of periodic behaviour in water
467 resource extremes. Periodicities of similar lengths have previously been detected in
468 European GWL records, such as those in the UK (Rust et al, 2018 Holman et al, 2011),
469 Hungary (Garamhegyi et al, 2016), Spain (Luque-Espinar et al, 2008), Italy (De Vita et al
470 2011), and Germany (Liesch and Wunsch, 2019); and European streamflow records, for
471 example in the UK (Rust et al 2021; Burt and Howden, 2013) and Sweden (Uvo et al, 2021).
472 Our results therefore are consistent with principal periodicities detected in wider European
473 water resources and highlight the NAO's wide-scale control on water resource extremes.

474 Despite the prominence of the average 7.5-year periodicity in water resource variables, the
475 wider time-frequency spectra show that the NAO's multiannual control on water resources is
476 subject to considerable transience and non-stationarity across time and frequency. For
477 instance, the percentage of water resource records with a significant covariance with the
478 NAOI at the 7.5-year periodicity remains below 10% until between 1960 and 1965, with
479 significance becoming abruptly widespread (> 30%) between 1980 and 1985. As such this
480 suggests that the NAO's control on water resources, at the 7.5-year periodicity, has only
481 been prominent over the past four to five decades. Furthermore, prior to this mode of
482 behaviour, an approximate 16-year periodicity predominated the water resource extremes
483 record that did not covary with NAOI. Previous studies have associated a minimum in this
484 16-year cycle in water resources with the wide-scale 1976 drought (Rust et al, 2019) that



485 affected most UK water resources, particularly in the south of the country (Rodda and
486 Marsh, 2011). These findings are also consistent with Barker et al (2019) who demonstrate
487 longer duration drought events in the UK for the period 1940 to 1980 (approximately), and
488 comparatively shorter drought durations for the period 1980 to present. This may be
489 explained by a more prominent low-frequency influence on water resources and extremes
490 during this former period (1940 – 1980), causing longer negative anomalies on drought
491 indices. Finally, Holman et al (2011) linked a 16-year periodic behaviour in groundwater
492 records with the **East Atlantic pattern**, the second-most dominant mode of atmospheric
493 variability in the North Atlantic region. Our results could be interpreted as suggesting an
494 abrupt shift towards increased frequency of water resource extremes around 1970 to 1980
495 as a result of a transition of periodic control from the EA to the NAO. This interpretation may
496 expand on findings from Neves et al (2019) who demonstrate that historical droughts in
497 southwest Europe are better explained with a combination of NAO and EA influence.

498

499 Multiple studies have noted a marked change in European hydrological drought trends since
500 the 1970s, often in the context of the ongoing effects of climate change on water resources
501 (Tanguy et al 2021; Rodda and Marsh, 2011; Bloomfield et al., 2019). These impacts vary
502 depending on the water resource and region but can include changing drought frequency
503 (Spinoni et al, 2015; Bloomfield et al., 2019; Chiang et al, 2021), severity (Hanel et al, 2018;
504 Bloomfield et al., 2019), and increasing divergence of drought characteristic across Europe
505 (Cammalleri et al, 2020). We show here that a dominant 7.5-year periodicity, driven by the
506 NAO, has occurred coincident to these reported changing trends, and proceeded a
507 secondary periodicity of approximately 16 years. As such our results suggest that some of
508 the change in drought frequency that has been noted to have occurred since the 1970s, may
509 be in-part driven by the NAO's increased periodic control on water resources. Hydroclimate
510 studies often highlight that the interaction between climate change, ocean-atmosphere
511 processes and land-surface processes may be complex, resulting in non-linear hydrological



512 responses to increasing global temperatures (Rial et al 2004, Wu et al, 2018). As such, the
513 abrupt emergence of a 7.5-year periodicity between the NAO and water resource extremes
514 between 1980 and 1985, and its weakening since 2005, may be evidence of this type of non-
515 linear response. While there have been many studies assessing the impact of climate
516 change projections on the NAO (e.g. Rind et al (2005); Woolings and Blackburn (2012)),
517 there have been few that have investigated potential interactions between climate change
518 and multiannual periodicities in the NAO. As such, the role of climate change in affecting the
519 non-stationary periodicities (detected in this study) is currently unknown.

520 Yuan et al (2017) highlight the importance of suitable calibration period selection for the
521 development of drought early warning systems, particularly in climate change scenarios.
522 Many of these systems in Europe (e.g. Hall and Hanna, 2018; Svensson et al., 2015) rely on
523 high-resolution hydrometeorological datasets for calibration of historical relationships, many
524 of which are only available for recent decades (Rust et al, 2021b, Sun et al 2018). We show
525 here that frequency statistics potentially used as calibration bases for water resource early
526 warning systems can exhibit both multidecadal periods of stability and abrupt sub-decadal
527 non-stationarities, driven by multiannual behaviours in the NAO. Furthermore, we show a
528 weakening of the dominant 7.5-year periodicity since 2005, suggesting a different frequency
529 structure may predominate water resource extremes from the 2020s. This further highlights
530 the need for continuous recalibration of critical forecasting utilities, and the potential benefit
531 of including the NAOI as a covariate when understanding multiannual periodic variability in
532 European water resources.

533



534 **5.2. Phase difference between NAO and water resource records at 7.5-year**
535 **periodicity**

536 The quantification of lead times between meteorological processes and water resource
537 response is critical in the development of early warning systems for water resource
538 management. As such, hydroclimate studies have sought to investigate temporal lags
539 between multiannual periodicities in the NAO and water resource variables across Europe
540 (Uvo et al, 2021, Neves et al 2019, Holman et al 2011). However, previous research has
541 highlighted that the relationship strength and sign between the NAO and European rainfall is
542 non-stationary at sub-decadal to decadal timescales (Rust et al 2021, Vicente-Serrano &
543 López-Moreno, 2008). The extent to which this non-stationarity is projected to multiannual
544 periodicities in water resources was previously unknown. Sign change is synonymous with a
545 phase difference shift of approximately π between periodic components of the NAO and
546 water resources, and as such has the potential to disrupt the projection of lead times into
547 future scenarios. Here we assess the phase difference between the NAO and water
548 **resources at a country scale to identify** the extent to which this non-stationary is present at
549 multiannual periodicities.

550 Most water resources records exhibit an abrupt shift in phase difference of approximately $-\pi$
551 around 1990. An earlier shift (of approximately $+\pi$) is also apparent between 1970 and 1980,
552 however this is less temporally aligned across the fewer records that cover this period. This
553 suggests that, for the period of approximately 1970 to 1990, the relationship sign between
554 the NAO and water resources was inverted. Furthermore, the timing of this period of
555 inversion generally aligns with reported periods of sign inversion in existing studies between
556 the NAO and UK rainfall (Rust et al 2021, Vicente-Serrano & López-Moreno, 2008). It is
557 interesting to note that this period of inversion is notably shorter for some groundwater level
558 records of the Chalk (e.g., those in South Chalk and Thames and Chiltern Chalk). Rust et al
559 (2021) showed the south and south east of the UK was subject to the increased non-
560 stationarity of the NAO-precipitation relationship when compared to other regions, which





561 may explain these relatively short periods of relationship inversion. A similar spatial pattern
562 is shown in the streamflow records, with minimal phase difference shifts in northwest
563 England, Scotland, and Northern Ireland where more stable signs have been found by Rust
564 et al (2021b).

565 Localisation of this non-stationarity between the NAO and water resources at multiannual
566 periodicities suggests it is possible to identify a discrete time period of sufficient stationarity
567 from which to calculate lead-in times for early warning systems (for instance, between 1990
568 and 2020). However, phase differences for this period also show a degree of non-

569 stationarity, varying by up to approximately $\pm\frac{1}{4}\pi$. Some of this variance may be due to

570 changing storage dynamics within a catchment over time (Rust et al, 2014; Beverly and

571 Hocking, 2012), but also the introduction of red noise from reconstructing from non-
572 significant wavelets. This also explains the increased variance seen in aquifer groups
573 characterised by higher autocorrelation (e.g., Sandstone) (Bloomfield and Marchant, 2013),
574 and the relatively low variance seen in streamflow records which often have lower
575 autocorrelation when compared to groundwater level (Hannaford et al, 2021). While this
576 can be minimised by calculating phase difference from significant wavelets only, we have
577 shown in the previous section that the significance between the NAO and water resources
578 and multiannual periodicities is also subject to notable non-stationarity.

579 Finally, in order to calculate accurate lead-in times between periodicities in the NAO and
580 water resources in future scenarios, a sufficient systematic understanding of the NAO sign
581 non-stationarity is required. However, there is limited research that has investigated the
582 causes for these modes of multiannual non-stationarity. Vicente-Serrano & López-
583 Moreno (2008) suggest that an eastward shift of the NAO's southern centre of action may
584 account for a portion of this variability, but highlight that further work is required for this to be
585 a sufficient explanation of a changing correlation between the NAO and European rainfall.

586 As such, existing non-stationarities between the NAO and water resources at multiannual



587 periodicities remains a considerable barrier to its application in improving preparedness for
588 future water resource extremes.

589 **5.3. NAO multiannual modulations on water resources in future scenarios**

590 Water resource management systems are in place across Europe to improve planning and
591 preparedness for the projected effects of climate change. As such, in order for multiannual
592 NAO modulations of water resources to have sufficient utility for water management systems
593 in future scenarios, they need to exhibit a comparable influence on water resources to the
594 projected effects of climate change. Here, we present historical modulations of summer
595 water resource variables from the principal NAO periodicity alongside expected impacts on
596 water resources from climate change projections in order to discuss their comparative
597 influence.

598 Jackson et al (2015) estimated median groundwater level change due to climate change in
599 24 boreholes across Chalk, limestone, sandstone and greensand aquifer groups in the UK
600 for the 2050s under a high emission scenario for September (as a typical annual minima of
601 groundwater levels in the UK). Median level from each site in Jackson et al (2015) have
602 been regrouped and averaged across the broad aquifer groups used in this study to allow
603 comparison with historical deviations in water resource results as a result of the NAO's 7.5-
604 year periodicity. This comparison is provided in Table 1. A mapping table of this comparison
605 is available in the supplementary material.

606



Aquifer group	50 th %ile gwI change due to climate change (m)	Gwl deviation due to 7.5-year NAO periodicity (± m) (med.mean)	Gwl deviation due to 7.5-year NAO periodicity (± m) (med.max)
Chalk (East Anglia)	-0.21	0.16	0.31
Chalk (Lincolnshire)	-0.31	0.71	1.03
Chalk (South)	-0.64	0.73	1.08
Chalk (Thames / Chilterns)	-0.69	0.86	1.33
Limestone	-0.28	0.35	0.51
Oolite	-0.36	0.21	0.33
Sandstone	-0.07	0.15	0.25
Greensands	-0.10	0.12	0.17

607 Table 1 – synthesis of Table 3 from Jackson et al (2015). Median results from the absolute
 608 teleconnection modulation on groundwater level from Figure 3 of this paper are also
 609 presented for the mean and maximum modulation cases. NAO teleconnection modulations
 610 greater than the reported 50th percentile climate change modulation are shaded in grey.

611
 612
 613 Historical modulations in groundwater level due to multiannual periodicities in the NAO were
 614 greater than projected GWL modulation from a high emissions climate change scenario, in
 615 all but two aquifer groups for mean NAO modulation (East Anglia Chalk, Oolite), and all but
 616 one for maximum NAO modulation (Oolite). Similar degrees of GWL modulation from climate
 617 change scenarios have been shown for wider European aquifer systems (e.g., Dams et al,
 618 2011), and our results for NAO modulations of GWL are of a similar degree to those reported
 619 by Neves et al (2019) for aquifers in the Iberian Peninsula. While few studies have looked at
 620 multiannual NAO modulations of groundwater level across Europe, our results here suggest
 621 a similar response across Western Europe, where the NAO has a greater influence on
 622 precipitation (Trigo et al, 2002). However, existing studies notable uncertainties in the future
 623 trends of groundwater level change due to climate change. For instance, Yusoff et al. (2002)
 624 demonstrated that it was not possible to predict whether groundwater level would rise or fall
 625 between 2020s and 2050s, Bloomfield et al. (2003) showed that groundwater levels were
 626 expected to rise in the 2020s but fall in the 2050s, and, Jackson et al (2015) showed



627 reductions in annual and average summer levels but increases in average winter levels by
628 the 2050s. For streamflow, Kay et al (2020) give estimated modulations to low flows (Q95)
629 as a result of climate change (2050 horizon). While no Scottish catchments were used in the
630 study, percentage modulations for low flows were found to be mostly between 0 to -20%
631 change with some catchments showing up to -40% change for catchments in the West and
632 South West of the UK. Schnieder et al (2013) show similar low flow modulations across
633 Europe as a result of climate change, ranging from +20% for northwest Europe to -40% in
634 the Iberian Peninsula. As such, our results for streamflow (Figure 7) indicate that multiannual
635 NAO modulation of streamflow has been, on average, comparable to the expected change
636 due to climate change scenarios. NAO modulations in streamflow are notably less than
637 those found in groundwater level, as may be expected given the established sensitivity of
638 groundwater processes to long-term changes in meteorological fluxes (Forootan et al., 2018;
639 Van Loon, 2015; Folland et al., 2015).

640 Given the scale of multiannual NAO influence on water resource compared to the estimated
641 effects of climate change, the NAO may have the potential to impact the projected trend of
642 water resource variability in certain future scenarios more than was previously understood,
643 and therefore effect the required adaptive management response. However, existing
644 research has shown that that current GCMs do not fully replicate low frequency behaviours
645 in the NAO that have been historical recorded (Eade et al, 2021). Given the importance of
646 multiannual periodicities the NAO in defining water resource behaviour, demonstrated here
647 and in other research (e.g., Uvo et al, 2021; Neves et al, 2019), this raises notable
648 uncertainties in the use of GCMs outputs for projecting European water resource behaviour
649 into future scenarios. Findings reported here suggest that current projections from these
650 GCMs may contain error that is comparable to the current projected effect of climate change
651 on water resources. This therefore highlights the need for improved low frequency
652 representation in GCMs, and for an understanding of the non-stationary atmospheric
653 behaviours are can considerably influence wide-scale water resource behaviour.



654 Rust et al (2018) set out a conceptual model for how multiannual modulations of water
655 resources due to the NAO may provide a system for improving water resource forecasts and
656 management regimes. This model highlights the need for a systematic understanding of how
657 multiannual periodicities affect water resources over time, including temporal lags and
658 amplitude modulation between the NAO and water resources. We demonstrate that the
659 degree to which the NAO's 7.5-year periodicity has modulated historical water resources is
660 of a similar order of magnitude to the estimated impacts on water resource variables from
661 climate change projections. These results further show the importance of including the
662 influence of multiannual NAO periodicities on water resources in the understanding of future
663 extremes, as they have the potential to affect the required management regime for certain
664 resources in climate change scenarios. However, we also show that there are notable non-
665 stationarities in NAO periodicities over time and their relationship with water resource
666 response, for which there is limited systematic understanding in existing hydroclimate
667 literature.

668

669 **6. Conclusions**

670 This paper assesses the utility of the relationship between the NAO and water resources, at
671 multiannual periodicities, for improving preparedness of water resource extremes in Europe.
672 We review this relationship in the context of non-stationary dynamics within the NAO and its
673 control on UK meteorological variables, as well as its potential impact on water resources in
674 climate change scenarios. We provide new evidence for the time-frequency relationship
675 between the NAO and water resources in western Europe showing that a wide-spread 7.5-
676 year periodicity, which predominates the multiannual frequency structure of many European
677 water resources, is the result of a non-stationary control from the NAO between
678 approximately 1970 and 2020. Furthermore, we show that known non-stationarities of the
679 relationship sign between the NAOI and European rainfall at the annual scale are present in
680 water resources at multiannual scales. A current lack of systematic understanding of both



681 these forms of non-stationarity, in existing atmospheric or meteorological literature, is a
682 considerable barrier to the application of this multiannual relationship for improving
683 preparedness for future water resource extremes. However, we also show that the degree of
684 modulation from multiannual NAO periodicities on water resources can be comparable to
685 modulations from a worst-case climate change scenario. As such multiannual periodicities
686 offer a valuable explanatory variable for ongoing water resource behaviour that have the
687 potential to heavily impact the required management regimes for individual resources in
688 climate change scenarios. Therefore, we highlight knowledge gaps in atmospheric research
689 (e.g. the ability of climate models to simulate NAO non-stationarities) that need to be
690 addressed in order for multiannual NAO periodicities to be used in improving early warning
691 systems or improving preparedness for water resource extremes.

692 **Data availability.**

693 The groundwater level data used in the study are from the WellMaster Database in the
694 National Groundwater Level Archive of the British Geological Survey. The data are available
695 under license from the British Geological Survey at <https://www.bgs.ac.uk/products/hydrogeology/WellMaster.html> (last accessed: 24/10/2021).

697 The streamflow data as well as the metadata used in this study are freely available at the
698 NRFA website at <http://nrfa.ceh.ac.uk/> (last accessed: 25/10/2021).

699 The data that support the findings of this study are available in CORD at
700 10.17862/cranfield.rd.16866868. This study was a re-analysis of existing data that are
701 publicly available from NCAR at <https://climatedataguide.ucar.edu/climate-data>.

702

703 **Author contributions.**

704 WR designed the methodology and carried them out with supervision from all co-authors. WR
705 prepared the article with contributions from all co-authors.



706 **Competing interests.**

707 The authors declare that they have no conflict of interest.

708

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717

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722

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