The importance of vegetation to understand terrestrial water storage variations

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10 **Abstract.** So far, various studies aimed at decomposing the integrated terrestrial water storage variations observed by satellite gravimetry (GRACE, GRACE-FO) with the help of large-scale hydrological models. While the results of the storage decomposition depend on model structure, little attention has been given to the impact of the way how vegetation is represented in these models. Although vegetation structure and activity represent the crucial link between water, carbon and energy cycles, their representation in large-scale hydrological models remains a major source of uncertainty. At the same

15 time, the increasing availability and quality of Earth observation-based vegetation data provide valuable information with good prospects for improving model simulations and gaining better insights into the role of vegetation within the global water cycle.

In this study, we use observation-based vegetation information such as vegetation indices and rooting depths for spatializing the parameters of a simple global hydrological model to define infiltration, root water uptake and transpiration processes.

20 The parameters are further constrained by considering observations of terrestrial water storage anomalies (TWS), soil moisture, evapotranspiration (ET) and gridded runoff (Q) estimates in a multi-criteria calibration approach. We assess the implications of including vegetation on the simulation results, with a particular focus on the partitioning between water storage components. To isolate the effect of vegetation, we compare a model experiment with vegetation parameters varying in space and time to a baseline experiment in which all parameters are calibrated as static, globally uniform values.

25 Both experiments show good overall performance, but including vegetation data led to even better performance and more physically plausible parameter values. Largest improvements regarding TWS and ET were seen in supply-limited (semi-arid) regions and in the tropics, whereas Q simulations improve mainly in northern latitudes. While the total fluxes and storages are similar, accounting for vegetation substantially changes the contributions of snow and different soil water storage components to the TWS variations, with the dominance of an intermediate water pool that interacts with the fast plant

30 accessible soil moisture and the delayed water storage. The findings indicate the important role of deeper moisture storages as well as groundwater-soil moisture-vegetation interactions as a key to understanding TWS variations. We highlight the

need for further observations to identify the adequate model structure rather than only model parameters for a reasonable representation and interpretation of vegetation-water interactions.

1 Introduction

35 Since 2002 the Gravity Recovery and Climate Experiment (GRACE) mission facilitates global monitoring of terrestrial water storage (TWS) variations from space – a milestone of global hydrology (Rodell, 2004;Famiglietti and Rodell, 2013). Observed TWS variations from GRACE have since become a cornerstone for diagnosing trends in water resources due to climate change or anthropogenic activities (Rodell et al., 2018;Reager et al., 2015;Scanlon et al., 2018;Syed et al., 2009;Tapley et al., 2019), as well as for benchmarking and improving global hydrological models (GHMs) (Scanlon et al.,
40 2016;Döll et al., 2014;Werth et al., 2009;Zhang et al., 2017;Kumar et al., 2016;Eicker et al., 2014). Significant co-variations between GRACE-TWS and the global land carbon sink (Humphrey et al., 2018) and surface temperatures (Humphrey et al., 2021) highlight the importance of the water-cycle as nexus in the earth system.

However, GRACE TWS estimates represent a vertically integrated signal of all water in snow, ice, soil, surface and groundwater. Thus, understanding processes and mechanisms of TWS variations requires attribution of TWS variations to
45 individual storage components. Despite advancements in remote sensing, large-scale quantification of these components based on observations remains challenging. For example, remote sensing-based estimates of soil moisture only capture depths up to 5 cm and do not necessarily reflect the moisture availability in the deeper soil column (Dorigo et al., 2015). Therefore, GHMs are necessary to interpret TWS variations in terms of contributions by snow, soil moisture, ground or surface water. However, several studies suggested that current state-of-the-art GHMs cannot reproduce key patterns of
50 observed TWS variations and show partly diverging TWS partitioning (Scanlon et al., 2018;Schellekens et al., 2017;Zhang et al., 2017). This uncertainty of the available tools to interpret TWS variations is clearly a major obstacle for diagnosing and understanding global changes of the water cycle, which is increased by differing model structures and grown complexity of existing GHMs.

Among previous studies, Trautmann et al. (2018) showed that a simple large-scale hydrological model that is calibrated in a
55 multi-criteria fashion against multiple global hydrological data streams simultaneously yields very good model performance compared to state-of-the art GHMs. This study contributed important insights in the drivers of TWS variations across spatial and temporal scales in northern high latitudes, in particular with respect to contributions by snow vs liquid water storages. In this study, we follow a similar framework of using multiple observational data streams to constrain a simple hydrological model to understand the partitioning of TWS components, yet expand to global scale and with further partitioning of liquid
60 water storages.

Among liquid water storages, especially the differentiation between soil moisture and ground water poses a challenge. Reflecting on the determinants of rather shallow soil moisture vs. deeper groundwater storage variations, it is apparent that under most conditions the soil moisture state itself is the first order control valve. In particular, it determines the amount of

water that is available for soil water uptake for evapotranspiration but also for percolation into deeper soil layers and
65   consequently recharge into the groundwater storage. The two key processes that shape soil moisture dynamics, infiltration
and evapotranspiration (ET), are strongly mediated by the presence and properties of vegetation (Wang et al., 2018). For
example, vegetation promotes infiltration over surface runoff due to larger surface roughness, dampened precipitation
intensities, more soil macro pores due to rooting and biological activity. In fact, such roles of vegetation in a global climate
model were already envisioned and evaluated almost 4 decades ago (Rind, 1984). Besides, rooting depth and soil properties
70   like soil texture and organic matter content control the size of the soil moisture reservoir available for ET, and how ET
responds to drought stress conditions (Baldocchi et al., 2021;Yang et al., 2020). The significance of interactions between
vegetation and soil moisture are at the heart of ecohydrology (Rodriguez-Iturbe et al., 2001) and have become evident in
many theoretical and experimental studies. While many studies analysed effects of water availability on vegetation
functioning (Porporato et al., 2004;Reyer et al., 2013;Wang et al., 2001;Yang et al., 2014), the inverse pathway of how
75   vegetation properties influence dynamics of water pools and the partitioning of TWS in large scale models has received
surprisingly little attention.

Therefore, the objective of this study is to investigate the effect of vegetation-dependent parameterizations of key
hydrological processes on TWS partitioning in a multi-criteria model data fusion approach. The model, an expanded version
80   of Trautmann et al. (2018), is a simple conceptual 4-pool water balance model. Model parameters are calibrated against
TWS variations from GRACE (Wiese, 2015), ET from FLUXCOM (Jung et al., 2019), runoff from GRUN (Ghiggi et al.,
2019) and ESA CCI soil moisture (Dorigo et al., 2017). We contrast two experiments which differ only with respect to how
vegetation-related parameters are defined: 1) a baseline experiment with global uniform parameters, 2) a vegetation
experiment where vegetation parameters vary in space and partly in time. In contrast to the traditional approach of
85   spatializing vegetation parameters by plant functional types or land cover classes and keeping this a-priori parameterization
fixed during model application, we take advantage of continuous information on few key properties that link vegetation and
hydrological processes: 1) spatially distributed and time-varying active vegetation cover that influences transpiration demand
and interception storage, 2) spatial pattern of soil water supply for transpiration via roots, and 3) spatially distributed and
time-varying the influence of vegetation cover on infiltration and runoff generation. Specifically, we are addressing the
90   following questions:

1) Where, when, and by how much are hydrological simulations improved by spatially distributed vegetation
    parameters?
2) To what extent does the attribution and interpretation of TWS variations for individual storage components change
    when introducing spatial and temporal variation of vegetation parameters?

1.1 Methods

In the first section we give a general overview on the design of this study and it's spatial and temporal coverage. Subsequently, the used model and data streams as well as the calibration and evaluation approach are explained in detail.

1.2 Overview

100   To assess the potential effect of including continuous information on vegetation, we compare two model variants that are based on the same conceptual structure: 1) a base model with static, globally uniform parameter values (**B**), and 2) a model variant that includes spatially (and temporally) varying vegetation characteristics by defining vegetation parameters as function of global data products (**VEG**).

Forced with global climate-data, the parameters of each variant are calibrated for a spatial subset against multiple Earth-
105   observation based data. In the **B** experiment, the parameters themselves are calibrated and globally constant parameter values are obtained. In the **VEG** experiment, we describe vegetation related parameters as the linear product of a calibrated scalar and spatio-temporal varying vegetation variables. By calibrating the scalar, we include the continuous pattern from the data, but weight it to best fit with observational constraints.

Once the parameters are calibrated, the simulations for the whole domain (global) are used to evaluate the model
110   performance at different spatial and temporal scale. To finally delineate the effect of including vegetation data on the composition of simulated total water storage across temporal (mean seasonal, inter-annual) and spatial (local grid scale, spatially aggregated) scales, we use the Impact Index as defined by Getirana et al. (2017).

The model is run on daily time steps at a 1°x1° latitude/longitude resolution, focusing on vegetated regions under primarily
115   natural condition. To avoid biases of the calibrated model parameters due to processes that are not represented in the model structure, we exclude grid cells with > 10% permanent snow and ice cover, > 50% water fraction, > 20% bare land surface and > 10% artificial landcover fraction. These grid cells are masked out using the Globland20 fractional landcover v2 (Chen et al., 2014). Additionally, we exclude regions with a large human influence, mainly related to groundwater extraction, on the trend in GRACE TWS variations (Rodell et al., 2018). The final study area comprises 74% of global land area. All other
120   data sets used in this study were resampled to the 1°x1° grid and subset to the same grid cells.

Due to the temporal coverage of forcing data and observational constraints, we calibrate the model for the period 01/2002-12/2014, while the global-scale model runs and analyses are performed for the period 03/2000-12/2014. Prior to each model run, all states are initialized by a 8-year spin-up period. The forcing for the spin-up period is assembled by randomly rearranging complete years of the forcing data.

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2.2 Model Description

The conceptual hydrological model is forced by daily precipitation, air temperature and net radiation (Table 1). It includes a snow component (see Trautmann et al. (2018)), a 2-layer soil water storage (*wSoil*), a deep soil water storage (*wDeep*) and a delayed, slow water storage (*wSlow*). The schematic structure of the model is shown in Fig. 1 and calibration parameters are

130   explained in Table 2.

Depending on air temperature (*Tair*), precipitation (*Precip*) is partitioned into snow fall (*Snow*), that accumulates in the snow storage (*wSnow*), and rain fall (*Rain*), that partly is retained in an interception storage. Interception throughfall together with snow melt are distributed among soil through infiltration and infiltration excess depending on the ratio of actual soil moisture and maximum soil water capacity following Bergström 1995:

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$$I_{exc} = I_{in} \cdot \left[\frac{\sum_{l=1}^{2} wSoil(l)}{\sum_{l=1}^{2} wSoil_{\max(l)}} \right]^{p_{berg}} \tag{1}$$

where, I_{exc} is the infiltration excess, I_{In} is the incoming water from throughfall and snow melt, *wSoil* is the soil moisture and *wSoil_max* the maximum soil water capacity of each soil layer *l*, and p_{berg} is a global calibration parameter.

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Part of the infiltration excess then replenishes a delayed water storage (*wSlow*), that acts as a linear reservoir and generates slow runoff (Q_{slow}). The remaining infiltration excess represents fast direct runoff (Q_{fast}). Q_{fast} and Q_{slow} together represent total runoff *Q*, that flows out of the system, i.e., grid cell.

Infiltrated water is distributed among 2 soil layers following a top-to-bottom approach, where the maximum capacity of the

145   first soil layer is prescribed as 4 mm, in order to match the tentative depth of satellite soil moisture observations, while the storage capacity of the 2nd soil layer is a calibration parameter ($wSoil_{max(2)}$). The 2nd soil layer is connected with a deeper water storage (*wDeep*). The size of *wDeep* is defined as a multiple of $wSoil_{max(2)}$ by the calibrated scaling parameter s_{deep} (Eq. 2). Depending on the moisture gradient between the two storages, water either percolates from the 2nd soil layer to the deeper soil, or it rises from the deeper storage into the 2nd soil layer, limited to a maximum flux rate.

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$$f_{pot} = f_{max} \cdot w_{grad} \tag{2}$$

Where, f_{pot} is the potential flux between both layers, f_{max} is the maximum flux rate (calibration parameter) and w_{grad} is the gradient of moisture calculated as

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$$w_{grad} = \frac{wDeep}{s_{deep} \cdot wSoil_{\max(2)}} - \frac{wSoil(2)}{wSoil_{\max(2)}} \tag{3}$$

where s_{deep} is the scaling parameter to derive the maximum capacity of *wDeep* as multiple of *wSoil$_{max(2)}$* (calibration parameter).

160   The deeper storage therefore acts as a storage buffer, that linearly discharges further to the delayed water storage (*wSlow*). The *wSlow*, which also receives part of the infiltration excess, is thus representative of all delayed storage components.

Evapotranspiration (*ET*) is represented by a demand-supply approach that is driven by a potential ET demand following Priestley-Taylor, and is limited by the available soil moisture supply. The ET is partitioned into interception evaporation (E_{Int}), bare soil evaporation from the first soil layer (E_{Soil}) and plant transpiration from the two soil layers (E_{Transp}).

165   Interception and plant transpiration are only calculated for the vegetated fraction of each grid cell, while bare soil evaporation is limited to the non-vegetated fraction of each grid.

$$E_{Soil} = min(E_{Soil-sup}, E_{Soil-dem}) \tag{4}$$

$$E_{Soil-dem} = PET.(1 - p_{veg}) \tag{5}$$

170   $$E_{Soil-sup} = wSoil(1).k_{Soil} \tag{6}$$

where, *PET* is the potential ET, p_{veg} is the vegetation fraction of each grid cell (calibration parameter), and k_{Soil} is the proportion of the first soil layer available for evaporation (calibration parameter).

Similarly, transpiration is calculated as

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$$E_{Transp} = min(E_{Transp-sup}, E_{Transp-dem}) \tag{7}$$

$$E_{Transp-dem} = PET.\alpha_{veg}.p_{veg} \tag{8}$$

$$E_{Transp-sup} = \sum_{l=1}^{2} wSoil(l).k_{Transp} \tag{9}$$

180   where, α_{veg} is the alpha coefficient of the Priestley-Taylor formula (calibration parameter) and k_{Transp} is the proportion of soil water available for transpiration (calibration parameter). In supply limited conditions, k_{Transp} effectively acts as a decay parameter that defines the depletion of the soil water storage via transpiration.

While water in *wSoil* is directly available for ET, *wDeep* is only indirectly accessible by capillary rise, and the water stored
185   in *wSlow* is not plant-accessible. Total water storage is the sum of all water storages, including *wSnow*, *wSoil*, *wDeep* and *wSlow*. Although groundwater and surface water storages are not implemented explicitly, they are effectively included in *wDeep* and *wSlow*, especially after calibration of associated storage parameters against GRACE TWS.
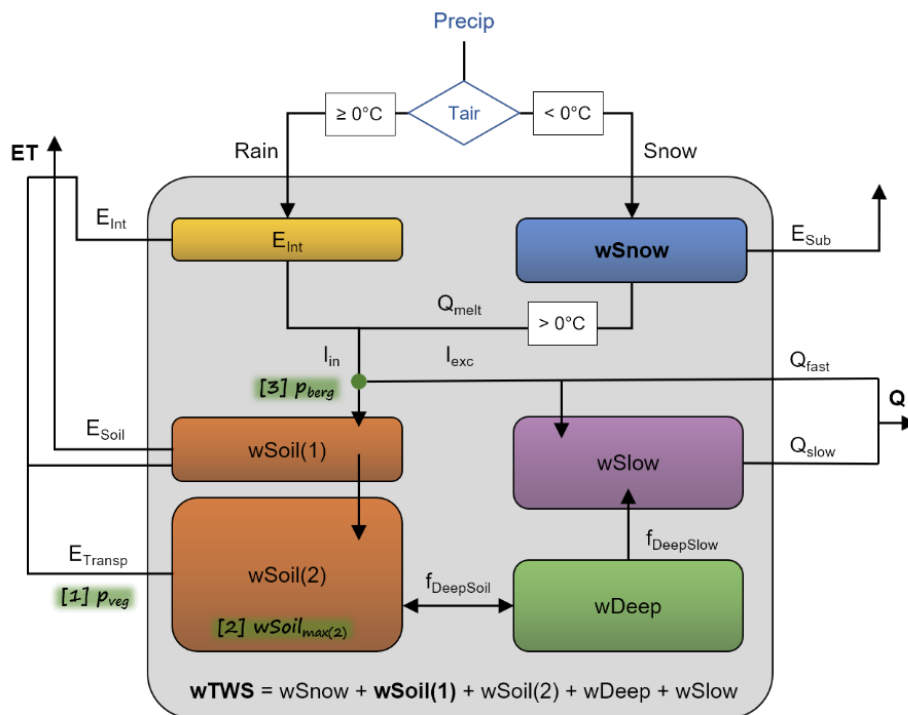
Figure 1 Schematic of the underlying model structure, with blue font denoting forcing data: *Precip* = precipitation, *Tair* = air temperature. **Boxes represent states:** E_{int} = interception storage, *wSnow* = snow water storage, *wSoil(1)* = upper soil layer, *wSoil(2)* = second soil layer, *wDeep* = deep water storage, and *wSlow* = slowly varying water storage. **Arrows denote fluxes:** *Rain* = rain fall, *Snow* = snow fall, E_{Sub} = sublimation, Q_{melt} = snow melt, I_{in} = incoming water from throughfall and snow melt, I_{exc} = infiltration excess, Q_{fast} = fast direct runoff, Q_{slow} = slow runoff, Q = total runoff, E_{Int} = evaporation from interception storage, E_{Soil} = soil evaporation, E_{Transp} = plant transpiration, ET = total evpotranspiration, $f_{DeepSoil}$ = flux between *wSoil* and *wDeep* (percolation resp. capillary rise), $f_{DeepSlow}$ = flux from *wDeep* to *wSlow*. **Bold print highlights model variables that are constraint in the calibration. Green highlights show where vegetation influence is included:** [1] the parameter p_{veg} to define each grid cell's vegetation fraction, [2] the parameter $wSoil_{max(2)}$ that defines the maximum plant available soil water, and [3] the parameter p_{berg} to define the infiltration and runoff generation partitioning.

Table 1 Data used for model forcing, description of vegetation characteristics and for model calibration.

	Product	Space	Time	Data Uncertainty	Reference
Forcing					
Precip	GPCP 1dd v1.2	global	daily		Huffmann et al. 2000
Tair	CRUNCEP v6	global	daily		Vivoy et al. 2015
Rn	CERES Ed4A	global	daily		Wielicki et al. 1996
Vegetation Characteristics					
EVI	based on MCD43C1 v6 (MODIS daily BDRF), calculated via MODIS standard EVI formula		daily climatology		Schaaf & Wang 2015
RD1	maximum rooting depth		static		Fan et al. 2017
RD2	effective rooting depth		static		Yang et al. 2016
RD3	maximum soil water storage capacity		static		Wang-Erlandson et al.

					2016
RD4	maximum plant available water capacity		static		Tian et al. 2019
Calibration					
wTWS	GRACE mascon RL06	global	monthly	with product	Wiese et al. 2018
wSoil	ESA CCI SM v4.04 (combined product)	~global	daily	with product	Dorigo et al. 2017
ET	FLUXCOM RS ensemble	global	daily	with product	Jung et al. 2018
Q	GRUN v1	global	monthly	~ 50 %	Ghiggi et al. 2019

2.2.1 Including Vegetation Characteristics

We include three aspects of vegetation influence on hydrological processes: 1) the specific transpiration demand by vegetation, 2) the soil water supply for transpiration via roots, and 3) the influence of vegetation on infiltration and runoff

205 generation. These three aspects are controlled by three corresponding model parameters, namely the grid cell's (active) vegetation fraction (p_{veg}), the plant available soil water ($wSoil_{max}2$), and the runoff generation/infiltration coefficient (p_{berg}). In the **VEG** experiment, scalar parameters are used as linear multipliers of observation-based spatio-temporal patterns. This step is considered necessary to handle the differences among different datasets, while still harvesting the information of spatial and temporal patterns from the observations. Alternatively, the scalars can be interpreted as scaling factors or weights

210 of different observational data streams. By calibrating the scalars, the weight given to observational pattern is constrained by the data streams used for model calibration.

2.2.1.1 Vegetation Fraction

The parameter p_{veg} reflects the active vegetation cover of each grid cell that influences the grid's interception storage,

215 transpiration demand, and partitioning of evapotranspiration components. To describe its spatial and seasonal variations, we include the mean seasonal cycle (MSC) of the Enhanced Vegetation Index (EVI). Therefore, p_{veg} at time step is defined as linear function of EVI, where s_{EVI} is the calibrated scaling parameter:

$$p_{veg} = s_{EVI} \cdot EVI \qquad (10)$$

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with $0 \leq p_{veg} \leq 1$.

EVI data is calculated via the MODIS standard formula (Didan & Barreto-Munoz) using the daily BRDF, nadir BRDF adjusted reflectance values MCD43C1 v6 (Schaaf &Wang 2015) for the period 01.2001 – 12.2014:

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$$EVI = 2.5 \; \frac{NIR - Red}{NIR + 6 \cdot Red - 7.5 \cdot Blue + 1} \qquad (11)$$

Since the daily EVI time series are not continuous due to noise and missing values during cloudy conditions, snow and darkness, the data was preprocessed to be used in the model. For each grid cell, we calculate the median seasonal cycle, fill
230 long gaps during winter time with a low value, interpolate missing values, and smooth the time series. Therefor, winter is defined as days with negative net radiation and gaps are considered long when 10 consecutive days of EVI data is missing. The winter time gaps are filled with the 5th percentile of available winter time data. The remaining missing values are linearly interpolated and finally the resulting seasonal cycle is smoothed by a local regression with weighted linear least squares and a 1st order polynomial model.

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2.2.1.2 Plant available Soil Water

In order to determine the soil water supply for transpiration as a function of vegetation, we define the maximum soil water capacity of the 2nd soil layer $wSoil_{max(2)}$ based on either rooting depth or soil water capacity data. We include the maximum rooting depth by Fan et al. (2017) (RD1), effective rooting depth by Yang et al. (2016) (RD2), maximum soil water capacity
240 by Wang-Erlandsson et al. (2016) (RD3) and maximum plant accessible water capacity by Tian et al. (2019) (RD4). Due to our definition of $wSoil_{max(2)}$ as maximum plant accessible water, all four data are, theoretically, suitable when focusing on spatial patterns. Practically, though, they vary in their definition, underlying approaches, spatial coverage and derived spatial pattern. The RD1 and RD2 are based on principles of vegetation optimality and plant adaption, and RD3 and RD4 are based on a water-balance perspective but using Earth-observations and/or data assimilation techniques. Therefore, we employ an
245 approach in which the weight of each data is calibrated, and their calibrated values are necessarily constrained by either the ET or TWS data. The maximum soil water capacity is therefore calculated as:

$$wSoil_{\max(2)} = \sum_{d=1}^{4} s_{RD(d)} \cdot RD(d) \qquad (12)$$

250 where $RD(d)$ is the data from each data stream d and $s_{RD(d)}$ are the corresponding scaling factors that are calibrated specifically for each RD data. As RD4 from Tian et al. (2019) is only available for arid to moderately humid vegetated land area and excludes tropical forests (Tian et al., 2019), resulting gaps in the study area are filled by the calibration parameter $wSoil_{max(RD4)}$ prior to scaling RD4.

255 ### 2.2.1.3 Runoff/Infiltration Coefficient

Finally, vegetation structure also affects the infiltration and runoff generation process as it alters the surface and sub-surface characteristics. To reflect this influence, we describe the infiltration/runoff parameter p_{berg} as linear function of vegetation fraction p_{veg}:

$$p_{berg} = s_{berg} \cdot p_{veg} \tag{13}$$

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where s_{berg} is the calibrated scaling parameter.

2.4 Model Calibration

In order to keep computational costs low and avoid overfitting of parameters, we perform model calibration for a spatial
265 subset of selected grid cells that include only 8% of the total study area. The calibration grid cells are chosen by a stratified
random sampling method that maintains the overall proportion of different climate and hydrological regimes defined by
Köppen-Geiger climate regions (Kottek et al., 2006).

Since this study focuses on the impact of vegetation and in order to keep the number of calibration parameters low, we do
not optimize snow related parameters and focus on the vegetation related parameters. Instead, the optimized snow
270 parameters from Trautmann et al. 2018 are used. This results in a total of 11 calibration parameters for the **B** model and a
total of 16 parameters for the **VEG** model (Table 2).

In order to constrain different aspects of the water cycle, we use a multi-criteria calibration approach similar to Trautmann et
al. 2018. The parameters of each model variant are simultaneously optimized against multiple observational constraints,
including monthly TWS anomalies from GRACE (Wiese et al. 2018), ESA CCI Soil Moisture (Dorigo et al., 2017),
275 evapotranspiration estimates from FLUXCOM-RS ensemble (Jung et al., 2019) and gridded runoff from GRUN (Ghiggi et
al., 2019) (Table 1).

For each of the observational constraints, we calculate a cost term that considers the data's specific strengths and
uncertainties, as well as each grid cell's area. Thereby, we only use grid cells and time steps with available observations,
which vary for the different data streams. To retrieve one cost term per observational constraint, we concatenate the
280 timeseries of all grid cells into a single vector for which costs are calculated. The individual cost terms are considered to
have the full weight of 1, resulting in a total cost value ($cost_{total}$) as a sum of individual cost. The total cost is then
minimized during the optimization process using a global search algorithm, the Covariance Matrix Evolutionary Strategy
(CMAES) algorithm (Hansen and Kern, 2004).

$$cost_{total} = \sum_{ds=1}^{nds} cost(ds) \tag{14}$$

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where, *cost(ds)* is the cost for each data stream *ds*. For wTWS, ET and Q, the cost terms are based on the weighted Nash
Sutcliff Efficiency (Nash and Sutcliffe, 1970), which explicitly considers the observational uncertainty σ:

$$cost = \frac{\sum_{i=1}^{n} \frac{(x_{obs,i} - x_{mod,i})^2}{\sigma i}}{\sum_{i=1}^{n} \frac{(x_{obs,i} - \bar{x}_{obs})^2}{\sigma_i}} \tag{15}$$

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where $x_{mod,i}$ is the modelled variable, $x_{obs,i}$ is the observed variable, $\overline{x}_{obs}$ is the average of $x_{obs,}$ and σ_i is the uncertainty of x_{obs} of each data point i. The cost criterion reflects the overall fit in terms of variances and biases, with an optimal value of 0 and a range from 0-∞.

295 Owing to the larger uncertainties of Q_{obs} on inter-annual scales (Ghiggi et al., 2019), we only use the monthly mean seasonal cycle, while for the other variables, full monthly time series were used.

To define σ of ET_{obs}, we utilize the median absolute deviation of the FLUXCOM-RS ensemble. For Q_{obs}, we assume an average uncertainty of 50% based on values reported in Ghiggi et al. (2019). For $wTWS_{obs}$, the spatially and temporally varying uncertainty information provided with the GRACE data is used. Besides, the largest monthly values of $wTWS_{obs}$ (< -

300 500 mm and > 500 mm) were masked out to avoid the effect of outliers on optimization results. Note that these outliers represent less than 0.5%, and are mainly located in coastal arctic regions, and are, thus, potentially related to land and sea-ice and/or leakage from neighboring grid cells over ocean. Before calculating $cost_{wTWS}$, the monthly means of observed and modelled wTWS are respectively removed to calculate anomalies over a common time period 01.01.2002–31.12.2012.

305 Since remote sensing-based soil moisture only captures the top few centimeters of soil depth, usually 5 cm, $cost_{wSoil}$ is calculated based on the modelled soil moisture in the first soil layer. As the combined ESA CCI soil moisture imposes absolute values and ranges from GLDAS-Noah (Dorigo et al., 2015), we use Pearson's correlation coefficient as $cost_{wSoil,}$ and focus on soil moisture dynamics that is most reflective of the original remote sensing observation. Only estimates from 01.01.2007 onwards are considered, as data before that period is sparse. Further, $cost_{wSoil}$ is calculated from the monthly

310 average values to avoid large variations and potentially large noise in the daily data. Thereby, only months with observations available for at least 10 days are considered. Due to snow cover, the temporal coverage of the product decreases with increasing latitude. Therefore, to prevent a bias towards northern summer months, we also exclude grid cells that lack more than 40% of monthly estimates. After these filtering for missing data, monthly time series for 56% of the total study area and 51% of the calibration grids are available.

315 **2.5 Model Evaluation and Analysis**

For model evaluation, we contrast the optimized parameter values and their uncertainties. The relative uncertainty in the optimized parameter vector is estimated by quantifying each parameter's standard error according to Omlin and Reichert (1999) and Draper and Smith (1981), similar to Trautmann et al. (2018).

320 For each experiment, the optimized parameter sets are used to produce model simulations for the entire study area. Their performances are then evaluated using Pearson's correlation coefficient and uncertainty weighted Nash-Sutcliff efficiency for wTWS, ET and Q observations (Eq. 16). The performances are evaluated on local (for each grid cell individually), regional and on global scales.

325    $wNSE = 1 - \dfrac{\sum_{i=1}^{n} \frac{(x_{obs,i}-x_{mod,i})^2}{\sigma i}}{\sum_{i=1}^{n} \frac{(x_{obs,i}-\overline{x}_{obs})^2}{\sigma_i}}$ (16)

For the regional analysis, we derive 5 hydroclimatic regions by performing a cluster analysis using the spatiotemporal characteristics of wTWS, ET and Q observations, as well as each grid cell's latitude. By that, each zone is characterized by similar seasonal dynamics and amplitudes of the constraints, allowing a better comparison when evaluating regional

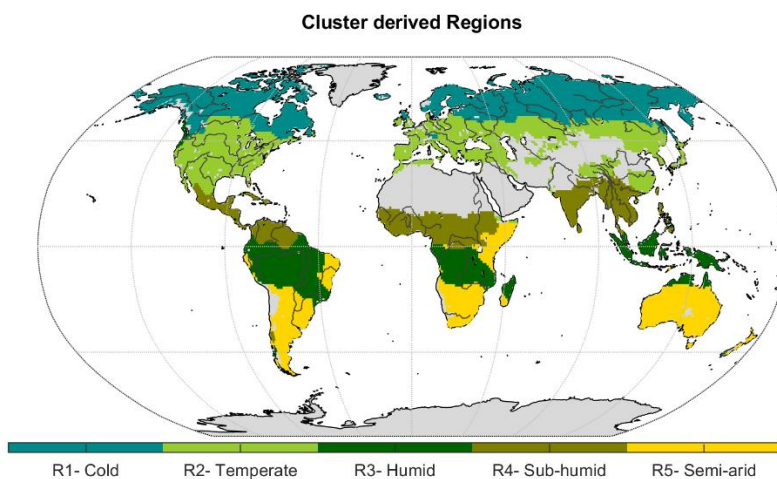
330   averages. The resulting regions are shown in Fig. 2. Region 1 comprises the snow dominated northern latitudes (*Cold*), while region 2 includes the moderate mid latitudes (*Temperate*). Very humid and mostly tropical regions are combined in region 3 (*Humid*). Region 4 is characterized by a distinct rain season (*Sub-humid*), while region 5 includes semi-arid areas in low latitudes (*Semi-arid*).

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Figure 2 Regions used for model evaluation and analysis.

Finally, we assess the contributions of the four water storage components, *wSnow*, *wSoil*, *wDeep* and *wSlow*, to seasonal and inter-annual variations of the total water storage across spatial scales, i.e. the local grid cell, the regional and the global

340   average. To do so, we apply the Impact Index *I* following Getirana et al. (2017). The metric describes the contribution *C* of each water storage *s* as the sum of its absolute monthly anomaly:

$C_s = \sum_{t=1}^{nt} |s_t - \overline{s}|$ (17)

345   Where, $\overline{s}$ is the average storage of the timesteps *t-nt*; with *nt* = 12 for mean seasonal and *nt* = 178 for inter-annual dynamics.

The Impact Index I_s is then defined as the ratio of each water storage component contribution C_s to the total contributions from all storage components:

$$I_s = \frac{C_s}{\sum_{s=1}^{n} C_s} \tag{18}$$

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The value of I_s range from 0-1, with 0 indicating no impact and 1 indicating full control of all variations.

3 Results

In the following section we first evaluate the performance of both calibrated model variants by comparing the calibrated model parameters and validate modelled wTWS, ET and Q against observations at global, regional and local scale.

355 Subsequently, we show the contribution of individual storage components to TWS variability for **B** and **VEG** on different spatial and temporal scales.

3.1 Model Evaluation

3.1.1 Calibrated Parameters

Table 2 summarizes the calibrated parameters and their uncertainties for the **B** and **VEG** model experiments. Overall,

360 including vegetation leads to more plausible parameter values after calibration, while in **B** several parameters hit their expected bounds. Furthermore, very high parameter uncertainties, indicating unconstrained values, could be drastically reduced in **VEG** and the interactions among different parameters are reduced with fewer parameters showing correlation (S3).

365 For **B**, p_{veg} suggests that on average only 37% of each grid cell are covered with vegetation globally. Despite having a good performance of modelled ET, the vegetation fraction is relatively low, but transpiration is increased by a high α_{veg} value (2.25), which is much higher than commonly used alpha coefficient of the Priestley-Taylor equation of 1.2 or 1.26 (Lu et al., 2005). At the same time, a very low fraction of the first layer soil layer is available for evaporation, as k_{Soil} hits the lower bound of 10%. Besides, the parameters controlling the drainage from deep and slow water storage (d_{Deep}, d_{Slow}) are high

370 resulting in a fast drainage, and effectively discarding the hydrological influences of these water pools.

For **VEG**, the median vegetation fraction is 73 %, leading to a more realistic fraction of soil moisture being available for evaporation ($k_{Soil} = 0.4$), which is similar to the mode value of 0.33 reported by McColl et al. (2017), and a more realistic α_{veg} value of 0.92, that effectively leads to the median Priestley-Taylor alpha coefficient of 0.81 (S2). In comparison to **B**, the resulting $wSoil_{max(2)}$ of **VEG** is with a median value of 52 mm considerably lower. Its spatial pattern mainly originates from

375 RD3 and RD4 data, while RD1 contributes only little and RD2 data is negligible. The resulting spatial patterns of the

maximum soil water capacity from the combination of all datasets are yet consistent with those from other estimates and patterns of rooting depth (e.g., Schenk and Jackson (2005)). We note here that the soil water capacity data are favoured over the rooting depth data. This agrees with Küçük et al. (2020), who suggest that estimating plant storage capacity based on Earth-Observation data may be more suitable than those using optimality principles. Besides, $wSoil_{max(2)}$ is defined as

380 maximum plant accessible soil storage, but also used to describe the maximum soil water storage. Due to their retrieval, the rooting depth data RD1 and RD2 suggest deep roots in warm/dry regions, while shallow roots would be sufficient in wet conditions. Therefore, RD1 and RD2's spatial pattern with deep roots in dry/warm regions and shallow roots in wet conditions behaves anti-proportional to the actual (maximum) soil water storage aimed for in this study. Related to the limited size of wSoil, calibration enforces a deeper and a slow water storage with reasonable depletion parameters (d_{Deep},

385 d_{Slow}). Overall, in **VEG,** only the k_{Transp} parameter, which describes the fraction of the 2nd soil layer that is available for transpiration, is higher than reported values of ET decay between 0.02- 0.08 (Teuling et al., 2006).

Table 2 Calibrated model parameters, their description, range and calibrated values for experiments B and VEG. Red fonts highlight calibrated values at the predefined parameter bounds.

parameter	description	units	default value	range	calibrated values ± uncertainty (%) B		VEG	
vegetation fraction								
p_{veg}	active vegetation fraction of the grid cell		0.5	0.3 - 1	0.37	± 0.05		
s_{EVI}	scaling parameter to derive active vegetation fraction from EVI data		1	0 - 5			3.89	± 0.05
evapotranspiration								
p_{Int}	interception storage	mm	1	0 - 10	1.0	± 0.08	0.6	± 0.02
k_{Soil}	fraction of 1st soil layer available for evaporation		0.5	0.1 - 0.95	0.1	± 0.01	0.4	± 0.08
α_{veg}	alpha parameter of the Priestley-Taylor equation		1	0.2 - 3	2.25	± 0.15	0.92	± 0.00
k_{Transp}	fraction of soil water available for transpiration		0.02	0 - 1	0.12	± 0.32	0.48	± 1.76
infiltration/runoff								
p_{berg}	runoff-infiltration coefficient		1.1	0.1 - 5	1.32	± 0.02		
s_{berg}	scaling parameter to derive the runoff-infiltration coefficient from p_{veg}		3	0.1 - 10			3.08	± 0.02
soil moisture								
$wSoil_{max(2)}$	maximum (available) water capacity of the 2nd soil layer	mm	300	10 - 1000	752	± 0.02		
$s_{RD(1)}$	weight to include maximum rooting depth by Fan et al. 2017		0.05	0 - 5			0.01	± 0.00
$s_{RD(2)}$	weight to include effective rooting depth by Yang et al. 2016		0.05	0 - 5			0.00	± 0.00
$s_{RD(3)}$	weight to include maximum soil water storage capacity by Wang-Erlandson et		0.05	0 - 5			0.15	± 0.06

	al. 2016							
$s_{RD(4)}$	weight to include plant available water capacity by Tian et al. 2019		0.05	0 - 5			0.15	± 0.07
$wSoil_{\max(RD4)}$	maximum (available) water capacity of the 2nd soil layer for grids with missing estimates in Tian et al. 2019	mm	50	0 - 1000			145	± 0.08
deep soil								
s_{deep}	scaling parameter to derive the maximum deep soil storage from $wSoil_{\max(2)}$		0.5	0 - 50	9.1	± 461317	5.6	± 0.21
f_{max}	maximum flux rate between deep soil and the 2nd soil layer	mm d^{-1}	10	0 - 20	1.5	± 0.00	5.1	± 0.01
d_{Deep}	depletion coefficient from deep soil to delayed water storage		0.5	0 - 1	1.0	± 5.61	0.01	± 0.00
delayed water storage								
rf_{Slow}	recharge fraction of infiltration excess into delayed water storage		0.5	0 - 1	0.78	± 1.72	0.68	± 0.01
d_{Slow}	depletion coefficient from delayed water storage to slow runoff		0.01	0 - 1	1.0	± 2329	0.02	± 0.03

390

3.1.2 Model Performance

Table 3 contrasts the overall model performance metrics for wTWS, ET and Q for the two experiments for the calibration subset of 8% grid cells (*opti*) and the entire study domain (*global*). The metrics are calculated in the same way as during optimization, i.e., by concatenation of the timeseries of all grids into a single vector for which statistics are calculated. In
395 general, the differences between *opti* and *global*, as well as between **B** and **VEG** are marginal. For **VEG**, results mainly improve for wTWS, and slightly for ET. Although the models were only calibrated for the spatial subset in *opti*, equally good or even better performances are obtained when the calibrated parameters are applied over the entire study domain. This suggests that the calibration subset was representative of the entire study domain and the calibration did not overfit the model parameters.

400 Among the variables, the best model performances in terms of *wNSE* and *corr* is obtained for ET. While the correlation between observed and simulated wTWS is high, the overall *wNSE* is relatively low, which mainly results from higher uncertainties in TWSobs and a larger variance error, likely originating from grid cells with low observed TWS variance.

Table 3 Overall model performance metrics in terms of weighted Nash-Sutcliff efficiency (wNSE) and Pearson's correlation
405 **coefficient (corr) of total water storage (wTWS), evapotranspiration (ET) and runoff (Q) in B and VEG experiments for the calibration subset (opti) and the entire study domain (global).**

	wTWS				ET				Q			
	wNSE		*corr*		*wNSE*		*corr*		*wNSE (MSC)*		*corr (MSC)*	
	opti	global	opti	global	opti	global	opti	global	opti	global	opti	global
B	0.33	0.33	0.69	0.69	0.97	0.97	0.90	0.90	0.63	0.63	0.86	0.86
VEG	0.38	0.41	0.71	0.72	0.98	0.98	0.90	0.91	0.60	0.57	0.85	0.85

Similar to the global metrics, the average mean seasonal cycle of different regions shows an equally good or slightly better
410   performance of **VEG** compared to **B** regarding all variables (Fig. 2). At regional scale (Fig. 4), the general pattern of grid-
wise Pearson correlation is similar for both experiments. However, the difference between the correlation coefficients
highlights an improvement using **VEG** for a large proportion of grid cells, and regarding all variables (indicated by brown
color). In the next section, regional and local performances are explained for wTWS, ET and finally Q.

415   For wTWS, the amplitude at the global scale is well-captured, yet with a phase difference of ~1 month in both model
variants, which both lead the timing of peak storage (Fig. 3). The phase shift is also apparent in the *Temperate* and *Cold*
regions, while the seasonal dynamics in *Sub-humid* and *Humid* region is captured well, yet with an underestimation of the
amplitude. Though differences are small, **VEG** obtains higher correlation and a smaller bias except for the *Semi-arid* region.
At local scale, correlation with GRACE TWS is lowest in rather semi-arid grid cells (Fig. 4), where wTWS variation is low
420   and it has lower impact on the global *wNSE*, which determines the cost in model calibration. However, including spatial
pattern of vegetation improves wTWS mainly in these (semi-)arid regions.

Regarding ET, both experiments reproduce seasonal dynamics in all regions quite well, yet tend to underestimate ET in the
Semi-arid, *Sub-humid* and *Humid* regions, especially in months with low ET (Fig. 3). At grid-scale (Fig. 4), correlation of
ET is very high, except for tropical regions. Compared to **B**, **VEG** improves correlation here, as well as in some (semi-)arid
425   regions such as the Sahel zone and the Western US.

In contrast to ET, performance for Q is generally the best in regions with poorer model performance in terms of ET (*Semi-
arid*, *Sub-humid* and *Humid* regions) (Fig. 3), suggesting a trade-off between the two different observation data streams.
Nonetheless, including vegetation improves peak runoff in all regions and reduces the underestimation of Q especially in the
Cold region. While the improvement of Q simulations in Northern latitudes gets even more obvious at grid-scale, **B** shows
430   higher correlation with observations in Africa and the Mediterranean (Fig. 4).
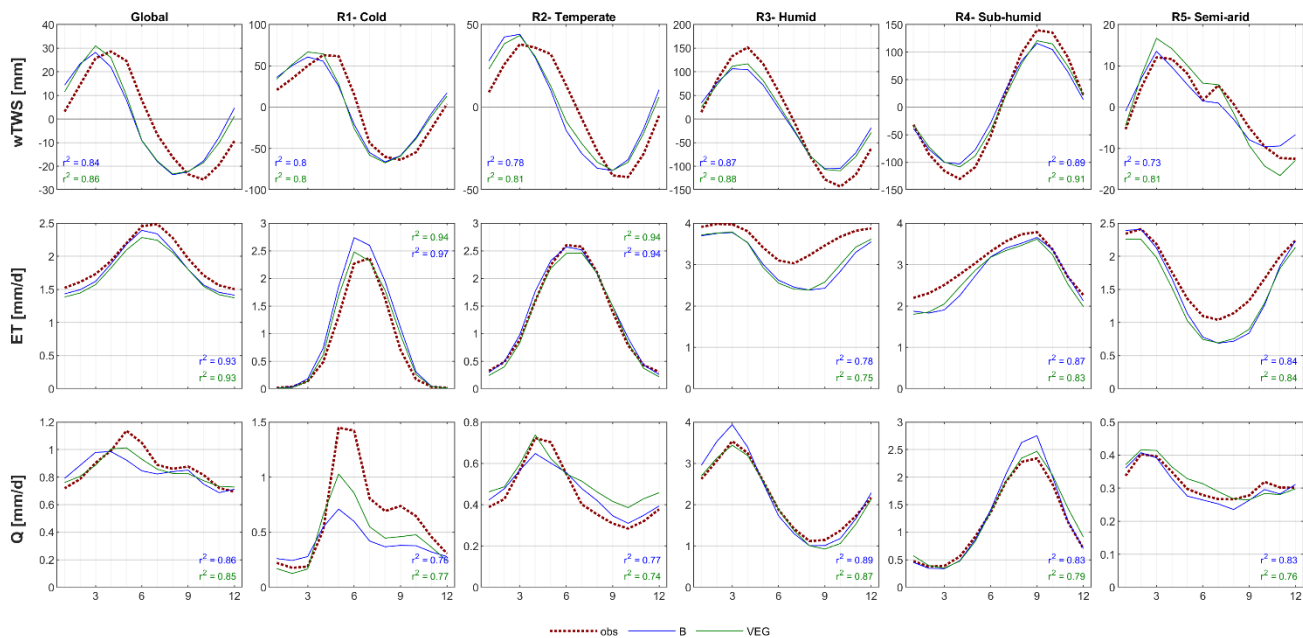
Figure 3 Global and regional mean seasonal cycles of total water storage (wTWS), evapotranspiration (ET) and runoff (Q) for the B and VEG experiments compared to the observational constraints by GRACE (wTWS), FLUXCOM (ET) and GRUN (Q).

435
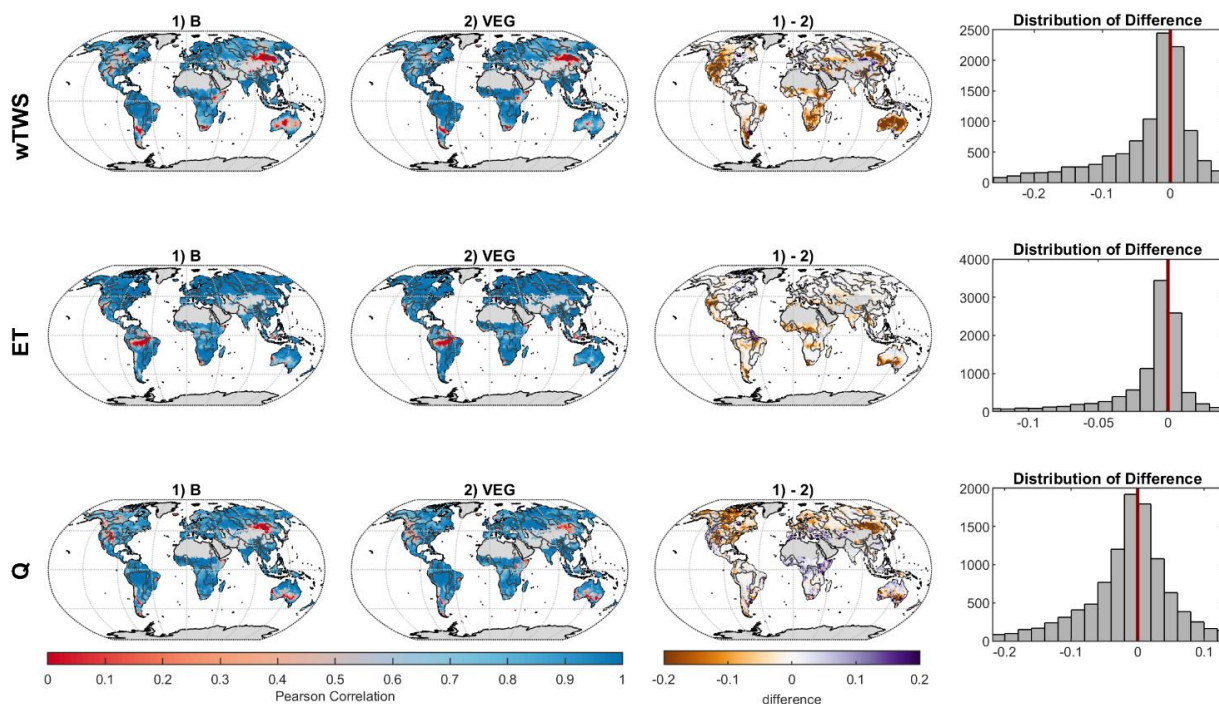
Figure 4 Grid-wise Pearson's correlation coefficient for total water storage (wTWS), evapotranspiration (ET) and runoff (Q) between 1) observations and B, and 2) observations and VEG, as well as differences between 1) and 2) (brown color, i.e., negative values indicate higher correlations for VEG, while purple color, i.e., positive values indicate better correlation values for B).

3.2 Contribution of Vegetation to TWS Variability

440 In this section, we present the influences of vegetation on TWS partitioning into snow (*wSnow*), plant-accessible soil moisture (*wSoil*), not directly plant-accessible deep soil water (*wDeep*) and non-plant-accessible slow water storages (*wSlow*) at different spatial and temporal scales. We first focus on mean seasonal dynamics and continue with the contribution of each component to inter-annual TWS variability at local grid-cell and regional scales, respectively, before presenting the analysis at the global scale.

3.2.1 Local & Regional Scale

Figure 5 shows the contribution of individual water storages to mean seasonal TWS variations at local grid-scale. For both **B** and **VEG**, *wSnow* has the highest impact in Northern latitudes and high altitudes where snow fall occurs regularly. Locally, the contribution of liquid water increases gradually with decreasing latitude and, finally, causes all TWS variations south of ~45° N. Within the liquid water storages, **B** attributes nearly all variations to directly plant accessible soil moisture *wSoil*,

450 with an average of 76% over all grid cells. While showing a similar pattern of increasing contribution towards lower latitudes, the **VEG** experiment only has an average of 17% contribution from *wSoil*. Instead, most variations (40%) are due to variability in the deeper soil storage, *wDeep*. Besides, the average impact of slow water storages *wSlow* (20%) is comparable to that of *wSnow* (22%) in **VEG**, though it is spatially much more limited to tropical regions, such as the Amazon basin.

455 Mean seasonal dynamics averaged globally and for different regions are shown in Fig. 6. As indicated by the grid-scale results, *wSnow* dominates TWS variations in the northern *Cold* region (73% in **B**, resp. 69% in **VEG**), and plays a considerable role in the *Temperate* region (28% resp. 26%). For the other regions, **B** attributes nearly all remaining variability to *wSoil*, while in **VEG** *wDeep* has the highest Impact Index (59% in *Semi-arid*, 50% in *Sub-humid* and 43% in *Humid*).

460
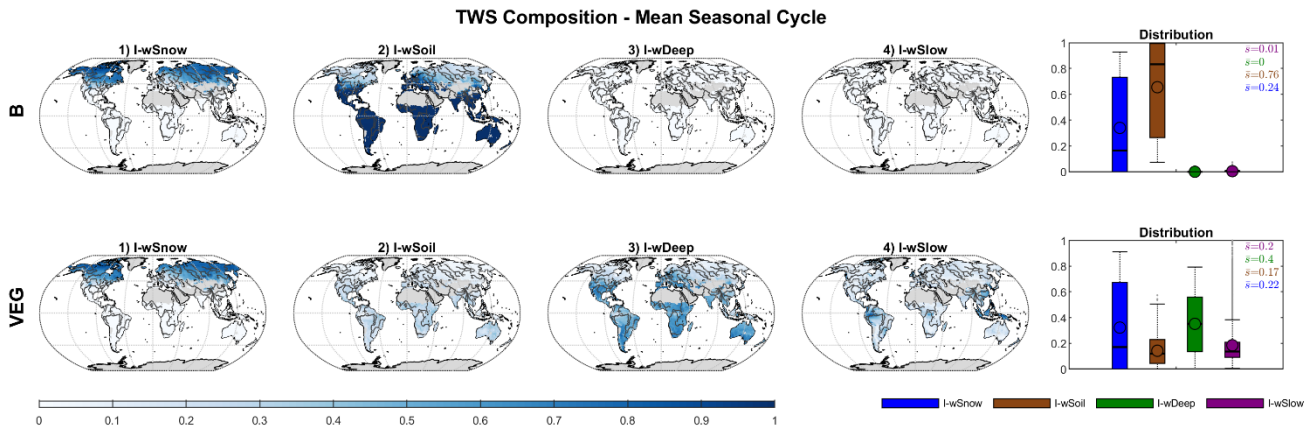
Figure 5 Global distribution of the Impact Index, I, for the contribution of simulated snow (*wSnow*), soil (*wSoil*), deep water storage (*wDeep*) and delayed water storage (*wSlow*) to the mean seasonal cycle of total water storage, for B and VEG.
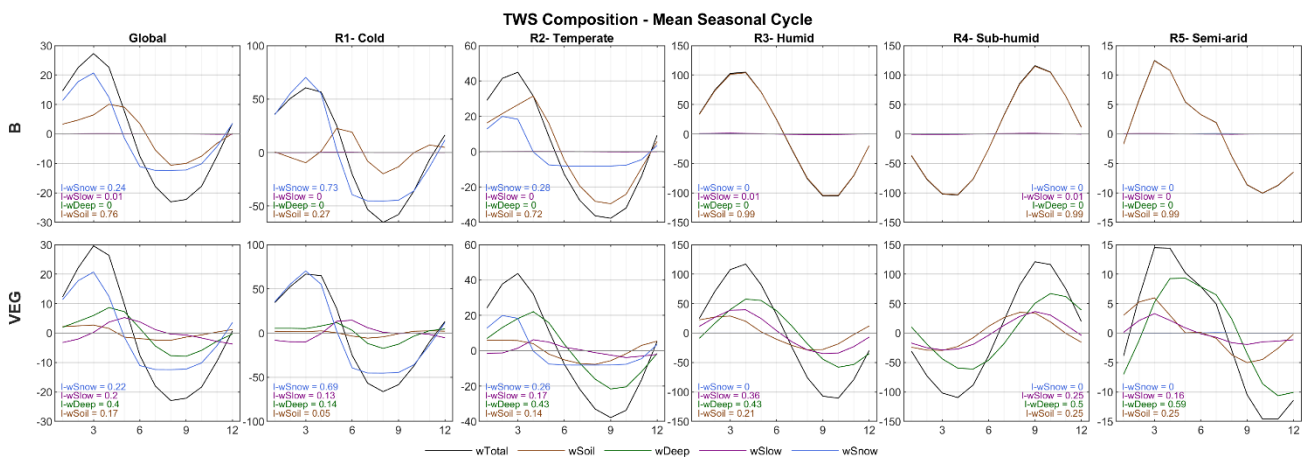
Figure 6 Global and regional average mean seasonal cycles of simulated total water storage and its components for B and VEG, including the regional Impact Index *I* for each storage.

At the inter-annual scales, the impact of *wSnow* decreases to 10% (**B**) respectively 12% (**VEG**) locally (Fig. 7). For most of the grid cells, all inter-annual TWS variations are caused by *wSoil* in **B**. In **VEG** however, the deeper soil layer *wDeep* is again the most important storage, with an average Impact Index of 53% for all grid cells. The contribution of *wSoil* and *wSlow* remain more or less the same as those for seasonal TWS variations.

Average contributions for different regions and globally (S4) show again that, in **B,** nearly all inter-annual wTWS variability is caused by *wSoil* (87-99%). Only in the *Cold* region, the impact of *wSoil* decreases to 69% in the favor of *wSnow* (31%). Similar to the local scale, in **VEG,** *wDeep* explains > 50% of wTWS variability in most regions, only in the *Cold* region, the contribution of *wDeep* is similar to *wSnow* (39% vs. 38%). The contribution of *wSoil* ranges from 9% (*Cold*) to 19% (*Semi-*

arid), while the impact of *wSlow* is between 16-18% in most regions and increases in *Sub-humid* (24%) and *Humid* (34%) regions.
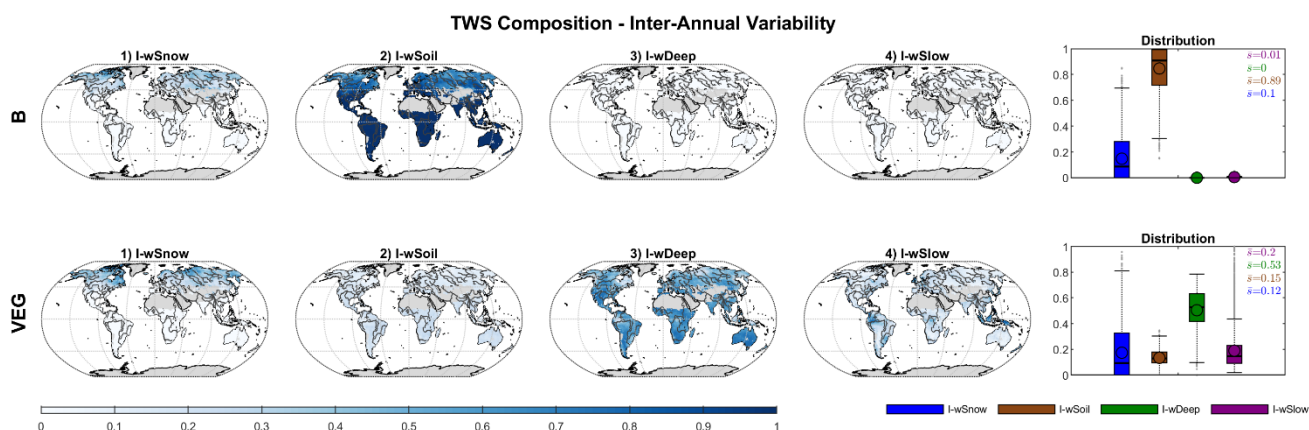
Figure 7 Global distribution of the Impact Index *I* for the contribution of simulated snow (*wSnow*), soil (*wSoil*), deep water storage (*wDeep*) and delayed water storage (*wSlow*) to the inter-annual variability of total water storage, for B and VEG.

3.2.2 Global Scale

Finally, Fig. 8 contrasts the impact of water storage components to the total storage, in **B** and **VEG**, at the global scale. As with the local and regional scales, including vegetation differentiates the composition of global TWS variations drastically. In both experiments, *wSnow* clearly dominates the spatially aggregated mean seasonal cycle with an Impact Index of 71% (**B**) and 69% (**VEG**). These contributions are considerably higher than the average local Impact Index over all grid cells (**B** 24%, **VEG** 22%; Fig. 5). As already seen at local scale, liquid water storages dominate the inter-annual TWS variability, whereby **B** and **VEG** differ in the attribution to different components of the liquid water storage. In **B,** all variations other than *wSnow* originates from *wSoil*, but *wDeep* dominates in **VEG**. Especially at inter-annual scales, *wDeep* accounts for half of all TWS variations. In contrast to **B**, in **VEG**, *wSoil* only has a minor impact of 7% at seasonal and 13% at inter-annual scale. Instead, *wSlow* has a moderate contribution of 11% (mean seasonal) and. 17% (inter-annual). In contrast to the mean seasonal dynamics in which the dominating storage are different at local and global scales, the inter-annual dynamics are consistent across scales with the same storage component dominating at both local and global scale (Fig. 5,7,8).
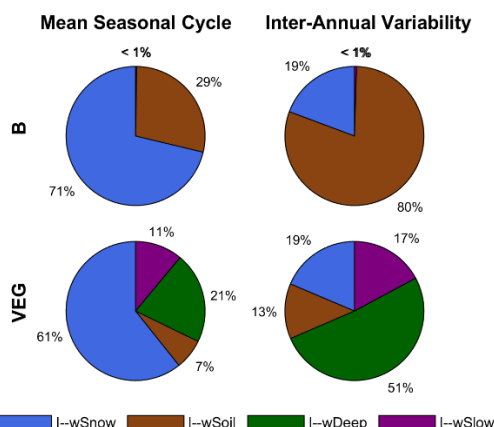
Figure 8 Impact Index *I* for the contribution of simulated snow (*wSnow*), soil (*wSoil*), deep water storage (*wDeep*) and delayed water storage (*wSlow*) to the global average mean seasonal cycle and inter-annual variability of total water storage, for B and VEG.

4 Discussion

In order to address the two main research questions of this study, the following section discusses the above shown differences between **B** and **VEG,** first regarding model performance and finally regarding the modelled partitioning of wTWS.

4.1 Model Performance

Both experiments have good performance against the observational constraints, and the differences between **B** and **VEG** are marginal over the global scale. However, there are not only systematic spatial differences at the regional and local scale, but also calibrated parameter values are more realistic and with smaller interactions among each other (less equifinal) in the **VEG** experiment. The latter suggests a more realistic representation of fluxes and states in **VEG** even though they are not directly constrained during the calibration process. Remaining discrepancies compared to observations can be associated with shortcomings and uncertainties in the observational data, as well as to the processes that are not represented in the rather simple model structure.

The differences in the seasonal phase of global wTWS in both model experiments mainly originate from the *Temperate* and *Cold* regions, and such differences have been reported previously (Döll et al., 2014;Schellekens et al., 2017;Trautmann et al., 2018). One of the potential reasons is the intermediate storage of melt water during spring in rivers and other surface water bodies, which occurs coherently over large areas in mid-to-high latitudes (Döll et al., 2014;Schellekens et al., 2017;Schmidt et al., 2008;Kim et al., 2009), and thus potentially delays the storage decay. In this context, also lateral water transport may additionally affect the wTWS variations in downstream grid cells. Yet, such processes and conditions are neither represented in **B** nor **VEG**.

Weaker performance of wTWS in (semi-) arid regions, on the one hand, originates from a low influence of these regions on global *wNSE*, because of low observed TWS variations and a high signal-to-noise ratio (Scanlon et al. 2016). Therefore, less weight is given to those grid cells in the cost component during calibration. On the other hand, alteration by human activities like groundwater withdrawal, dams and irrigation to overcome the natural water shortage in such regions as North-East

525 China and the American (Mid-)West can be regionally large in relative terms. While we aimed to exclude grid cells with large human impact a priori, we cannot completely exclude the influence of the aforementioned anthropogenic processes, that are not explicitly represented in our model experiments. It should, however, be noted the observational EVI data used in the **VEG** experiment do have an imprint (of the effects) of irrigated agriculture in terms of increased vegetation activity. This may be associated with an improved simulation of wTWS variations in such regions in the **VEG** experiment.

530 While overall ET performance is good, tropical regions peak out with low correlation. These areas are associated with higher uncertainties in the FLUXCOM ET estimates due to issues with the energy closure gap. Nonetheless, including vegetation data improves simulated ET here, suggesting a better representation of the characteristic highly active vegetation compared to other regions and to global averages. Besides, **VEG** improves ET mainly in supply-limited regions for the reasons already presented above for improved TWS performance in (semi-) arid regions.

535 The trade-off between the performances, more specifically bias, of Q and ET simulations suggest possible inconsistencies between the ET and Q constraints from independent sources. Additionally, the bias regarding either ET or Q, may relate to shortcomings in the precipitation forcing that doesn't provide sufficient input to support both outgoing water fluxes. Lastly, some remaining deterioration of performance of Q may originate from deficiencies in the GRUN product itself which was generated with climatic drivers only, disregarding information on spatio-temporal variations in vegetation (Ghiggi et al.,

540 2019).

The improvement of Q in Northern latitudes is associated with the activation of the slow and delayed storage in the **VEG** experiment with spatial varying parameterization of soil water storage capacity. The slow storage represents better the runoff delay in surface water and rivers in these regions that results in improvements of low flow during winter as well as the increase of runoff during spring (Fig. 3). Such delayed runoff also improves the simulation of peak runoff in the *Sub-humid*

545 and *Humid* regions.

The remaining deficiencies in model performance, especially in the *Cold* region, indicate missing processes in the simple model structure. Such processes include freeze/thaw dynamics, permafrost and ice jam in river channels that would increase surface water storage and allow high spring flood. Besides, snow parameters have been calibrated against remote sensing-based GlobSnow Snow Water Equivalent that is known to saturate for deep snow conditions (Luojus et al., 2014) (see

550 Trautmann et al. (2018)). Although the calibration process considered this shortcoming, an underestimation of modelled snow accumulation is possible – leading to an underestimation of peak snow pack in winter that would result in an underestimation of runoff due to lower snowmelt in spring.

While the **VEG** experiment presented here considers all 3 aspects of vegetation influences on hydrological processes (see
section 2.2.1), we also run experiments with including these aspects separately into model calibration (not shown). These
analyses found that the largest improvement was obtained when including soil water storage capacity as a function of rooting
depth and storage capacity data, and a rather low impact when considering the runoff/infiltration partitioning as a function of
vegetation fraction. This highlights the central role of soil water storages and the importance of adequately describing soil
moisture pattern and dynamics in hydrological models.

4.2 Contribution to TWS Variability

Both model experiments agree with previous studies (Trautmann et al., 2018) in showing a dominating role of snow
accumulation and depletion on global seasonal TWS variability, and the same of liquid water storage on inter-annual scales.
At the same time, the contribution of individual storages to TWS variations differ at the local grid-scale compared to when
they are averaged over a region or globally. In **B,** all variations other than *wSnow* originates from directly plant accessible
soil moisture, whereas, in **VEG**, the deeper soil storage *wDeep* becomes the most important. Therefore, including
observation-based information on vegetation changes the attribution of TWS variations drastically, while the variations of
total TWS themselves do not change significantly.

In the following, we briefly focus on the varying TWS composition among spatial scales, and finally discuss the systematic
differences between **B** and **VEG**.

Differences among scales

Albeit their global coverage, the presented results agree with previous regional studies focusing on Northern mid-to-high
latitudes (Trautmann et al., 2018). Snow dominates seasonal TWS variations locally and regionally in higher northern
latitudes and altitudes (Güntner et al., 2007). Its stronger contribution on spatially aggregated signals can be explained by the
spatial coherence of snow accumulation over larger areas. Liquid water storages, on the other hand, are more spatially
heterogenous, with increasing and decreasing dynamics across regions that cancel out and compensate each other when
spatially aggregated (Trautmann et al., 2018;Jung et al., 2017).

In contrast to the mean seasonal dynamics, the inter-annual Impact Indices of the storage components at the global scale are
similar to the average local Impact Indices (Fig. 7 and Fig. 8). This suggests that at inter-annual time scales, there is no
spatially coherent pattern of one single storage component that leads to higher accumulated Impact Indices than the local
averages.

Differences among model experiments

Differences in the composition of TWS variability between **B** and **VEG** are effectively reflected in the differences of
calibrated parameters. In **B**, the directly plant accessible soil water storage is larger, due to a higher effective $wSoil_{max(2)}$,

while delayed water storages are 'turned off' because of increased drainage (d_{Deep}, d_{Slow}), reducing the variations in *wDeep* and *wSlow*. Although **VEG** has been calibrated in the same way with the same observational constraints, calibrated model parameters differ as the included data on vegetation characteristics provides complementary information on spatial and temporal patterns. Therefore, the resulting calibrated parameters can be assumed to be more realistic. For example, they enable (delayed) longer-term water storage as well as capillary rise from the deeper soil water storage when the directly plant accessible storage dries out. Due to this process, TWS variations are mainly controlled by *wDeep* in **VEG**.

In detail, the increasing importance of the indirect plant-accessible storage *wDeep* in **VEG** can be related to the limited maximum soil water capacity $wSoil_{max(2)}$ that is constrained by rooting depth/soil water capacity data, and to a higher k_{Transp} parameter. The smaller *wSoil* storage forces percolation to *wDeep*, but the water is still available when needed due to the capillary rise from *wDeep* to *wSoil*.

Removing capillary flux from *wDeep* to *wSoil* in fact increases the contribution of *wSoil* to seasonal variability, while the impact of *wDeep* remains high on inter-annual scales (S7). Thus, the question is whether the derived contributions to TWS variability are robust or an artefact of the model formulation. While the contribution of capillary rise to total ET is < 20% for most grid cells, it becomes more important in arid-to-wet transition regions, e.g., sub-Saharan Sahel, Savannas, northern Australia and the Indian subcontinent (Fig. 9). These are regions with high precipitation seasonality, where vegetation grows deep roots to access deep unsaturated zone storage and groundwater during the dry season. The spatial patterns do not only agree with the findings of Koirala et al. (2014), who applied the physically-based model MATSIRO to investigate the effect of capillary flux to hydrological variables, but also with observations. For example, the spatial patterns are in line with the predicted probability of deep rooting by Schenk and Jackson (2005), and are supported by Tian et al. 2019 who found that vegetation remains active long into the dry season in Africa, suggesting that soil-deep soil/groundwater interaction plays a considerable role. Therefore, the spatial pattern of where *wDeep* interacts with *wSoil* is reasonable, indirectly validating the results from **VEG**.

While defined as 'fraction of soil water available for transpiration', k_{Transp} is also an effective decay parameter for the depletion of *wSoil* via transpiration processes. Plausible values are in the order of $10^{-3} - 10^{-2}$, similar in magnitude to delay coefficients for baseflow. By calibrating a model against GRACE TWS, it is difficult to decide whether water leaves the system slowly via ET or by Q. Additional including ET and Q data in model calibration should have ideally reduced equifinality. However, our results suggest that the model is still not constrained well enough: in **B**, k_{Transp} is more plausible, yet other slow depleting storages are 'turned off'. In contrast, **VEG** with additional vegetational data, simulates a reasonable slow storage that contributes to Q, but has a rather high calibrated k_{Transp}. Fixing k_{Transp} in **VEG** to 0.05 and optimizing all other parameters would result again in most TWS variations being caused by *wSoil*, but with less improvement in model performance compared to **B** (S8). Therefore, TWS decomposition is very sensitive to (certain) parameter values. However, **VEG** and **VEG** with fixed k_{Transp} qualitatively agree in the importance of the slow water storage in *Humid* regions, which was also shown by Getirana et al. (2017).

We showed, that even with the same calibration procedure and underlying model structure, TWS composition differs drastically between model experiments, and that is indicative of the large impact of the role of vegetation and of transpiration water supply within the model.

625   As with the presented model variants, TWS composition simulated with existing large-scale hydrological models differs widely (Scanlon et al., 2018;Schellekens et al., 2017;Zhang et al., 2017). For example, PCR-GLOBWB and W3RA attribute seasonal TWS variations in the tropics to groundwater, while other models suggest it is mainly caused by soil moisture. Those results are largely dependent on model structure and parametrization, which is potentially a challenge when models are used to decompose the integrated GRACE TWS signal, and when implications of different processes and interactions are drawn. For example, Humphrey et al. (2018) analysed how the CO_2 growth rate is affected by inter-annual fluctuations in

630   GRACE TWS, assuming that these represent fluctuations in plant accessible water that influence the carbon uptake of land ecosystems. In contrast, our study, along with previous reports, show that a significant proportion of the GRACE TWS signal in tropics is not directly plant accessible soil moisture, but deeper soil water and slow storage component. The latter comprises surface water storage, whose importance for TWS variations in tropical regions has been shown by several studies (e.g. Güntner et al. 2007, Getirana et al. (2017)).

635   Although **VEG** can be considered more reliable because of more realistic parameter values and better model performance, the current study still has some shortcomings. Despite using a multi-variate calibration, individual component fluxes and states may not necessarily be well constrained. To further improve and solidify conclusions, especially on TWS partitioning, more constraints, such as deep soil moisture estimates or high-quality observations of surface water are needed. Furthermore, spatial constraints for defining the depletion of water storages via ET and Q – either with spatial information on the delay

640   parameters (k_{Transp} for ET; d_{Slow} for Q), or for their sub fluxes (transpiration or evaporation; baseflow or direct runoff) would be beneficial. In this context, runoff characteristics as the baseflow index or the baseflow recession coefficient provided by Beck et al. (2015) are potentially useful to define spatial pattern of the slow runoff component. Besides, a GRACE product with daily resolution (Eicker et al., 2020) could enable better decomposition and differentiation of fast and slow storages whose short-term imprints are lumped in the monthly TWS signal.

645
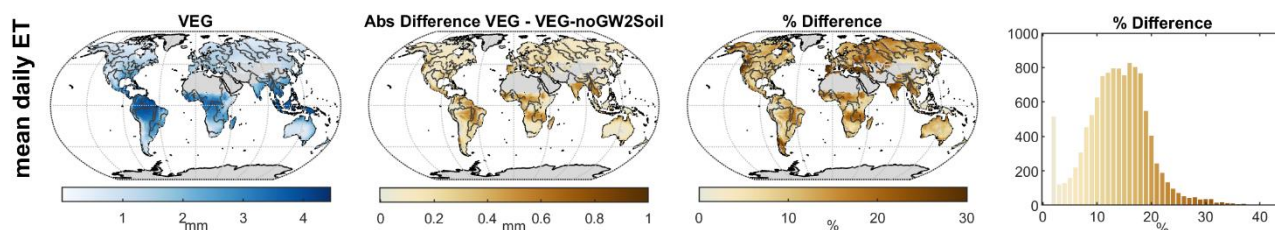
Figure 9 Total evapotranspiration of VEG with capillary flux from the deep soil water storage (left), and difference compared to a model version without capillary flux in mm (center) and as percentage difference (right).

5 Conclusion

In this study, we investigated the effect of vegetation on global hydrological simulations and in particular on the partitioning
650 of TWS variations among snow, plant accessible soil moisture, a deep soil and groundwater storage, and a slowly varying
water pool that represent surface and near-surface water storage. To do so, we included observation-based vegetation
information to parameterize the hydrological processes of evapotranspiration, soil water storage and runoff generation in a
large-scale hydrological model. With the parsimonious model that was constrained against multiple observations, we
highlight the value of observation-based datasets in constraining model parameters of global hydrological models, while
655 maintaining simple model formulations to evaluate the influences of vegetation in the global hydrological cycle.

First, we find that using a multi-criteria calibration approach allows for different model variants to perform relatively well
despite major differences in model parameterization among them. In fact, even without vegetation, the model performance
can be interpreted as reasonable, and more so at the global scale. However, including spatial pattern of vegetation further
improved the model performance. For example, large improvements were found in supply-limited regions, i.e., (semi-) arid
660 regions (TWS and ET) and in tropical regions (ET), and Q simulations both globally and regionally in the northern
hemisphere. Undoubtedly, spatio-temporal variations of vegetation characteristics are relevant for the regional and global
hydrological simulations.

Interestingly, we find that the calibrated parameter values are also more reasonable when the model is fed with the
vegetation information. In particular, parameter interactions and equifinality were reduced even though the same
665 observational constraints were used for calibration.

Lastly, we show how vegetation can modulate surface and subsurface hydrological process representation in the model,
changing the spatial-temporal dynamics of individual storage components while maintaining the same overall response of
total hydrological fluxes and storage variations. With or without vegetation, seasonal storage variations are dominated by
snow at the global scale. However, including vegetation drastically changes the attribution of TWS variations among soil
670 moisture, deep soil water and slow water storages. Without vegetation, the soil moisture effectively controls most of the
TWS variation, but with vegetation the role of deeper and delayed water storage becomes prominent.

In summary, this study highlights the value of including vegetation characteristics to further constrain model parameters and
to reliably represent hydrological processes despite a parsimonious model structure. The findings further suggest an
important role of groundwater-soil moisture-vegetation interactions for TWS variations. In particular, the representation of
675 vegetation related processes in global hydrological models seems to be a key factor controlling TWS partitioning of their
simulations and associated uncertainties. Besides, it emphasizes the need for further observations and observational-based
estimates, as well as multi-model experiments to provide and understand identifiability in model structure.

Besides, this study motivates further multi-model experiments to understand the need and potential of existing and novel
observational constraints to increase the identifiability not only regarding model parameters, but also of model structure.

Author contribution

TT designed the research in extensive collaboration with SK, NC and MJ. TT and SK programmed the model experiments. NC contributed to parameter estimation and uncertainty analysis. TT performed the analysis and prepared the first draft of the manuscript. All co-authors provided recommendations for the data analysis, participated in discussions about the results, and edited the manuscript.

Competing interests

The authors declare that they have no conflict of interest.

Code/data availability

All code and data are available from the authors upon request.

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