

1 Effects of aquifer geometry on seawater intrusion in annulus
2 segment island aquifers

3
4 Zhaoyang Luo^{1,2}, Jun Kong^{1,3,#}, Chengji Shen¹, Pei Xin¹, Chunhui Lu¹, Ling Li⁴,
5 David Andrew Barry²

6
7 ¹State Key Laboratory of Hydrology-Water Resources and Hydraulic Engineering, Hohai
8 University, Nanjing, China

9
10 ²Ecological Engineering Laboratory (ECOL), Environmental Engineering Institute (IIE),
11 Faculty of Architecture, Civil and Environmental Engineering (ENAC), École Polytechnique
12 Fédérale de Lausanne (EPFL), Lausanne, Switzerland

13
14 ³Jiangsu Key Laboratory of Coast Ocean Resources Development and Environment Security,
15 Hohai University, Nanjing, China

16
17 ⁴School of Engineering, Westlake University, Hangzhou, China

18
19 [#]Corresponding author: Jun Kong (kongjun999@126.com)

20
21 Resubmitted to *Hydrology and Earth System Sciences* on 28 November 2021

22 **Abstract**

23 Seawater intrusion in island aquifers was considered analytically, specifically for annulus
24 segment aquifers (ASAs), i.e., aquifers that (in plan) have the shape of an annulus segment.
25 Based on the Ghijben-Herzberg and hillslope-storage Boussinesq equations, analytical
26 solutions were derived for steady-state seawater intrusion in ASAs, with a focus on the
27 freshwater-seawater interface and its corresponding watertable elevation. Predictions of the
28 analytical solutions compared well with experimental data, and so they were employed to
29 investigate the effects of aquifer geometry on seawater intrusion in island aquifers. Three
30 different ASA geometries were compared: convergent (smaller side facing the lagoon, larger
31 side is the internal no-flow boundary, flow converges towards the lagoon), rectangular and
32 divergent (smaller side is the internal no-flow boundary, larger side facing the sea, flow
33 diverges towards the sea). Depending on the aquifer geometry, seawater intrusion was found
34 to vary greatly, such that the assumption of a rectangular aquifer to model an ASA can lead to
35 poor estimates of seawater intrusion. Other factors being equal, compared with rectangular
36 aquifers, seawater intrusion is more extensive and watertable elevation is lower in divergent
37 aquifers, with the opposite tendency in convergent aquifers. Sensitivity analysis further
38 indicated that the effects of aquifer geometry on seawater intrusion and watertable elevation
39 vary with aquifer width and distance from the circle center to the inner arc (the lagoon
40 boundary for convergent aquifers or the internal no-flow boundary for divergent aquifers). A
41 larger aquifer width and distance from the circle center to the inner arc weaken the effects of
42 aquifer geometry and hence differences in predictions for the three geometries become less

43 pronounced.

44 **Keywords:** sharp-interface; steady-state analytical solution; atoll aquifer; annulus segment
45 aquifer, seawater intrusion

46 **Key Points**

47 ➤ Analytical solutions of steady-state seawater intrusion were derived for annulus segment
48 aquifers

49 ➤ Among three different aquifer geometries, divergent aquifers have the lowest watertable
50 and hence the most extensive seawater intrusion

51 ➤ Aquifer geometry effects on seawater intrusion depend on the aquifer width and distance
52 from the circle center to the inner arc

53 1. Introduction

54 Islands are extensively distributed throughout the world's oceans. Unfortunately, their
55 groundwater resources are impacted by sea-level rise and increased demands. According to a
56 recent estimate, there are approximately 65 million people living in oceanic islands where
57 groundwater may be the only source of freshwater (Thomas et al., 2020). Fresh groundwater
58 stored on oceanic islands is mainly from precipitation (usually in the form of a freshwater
59 lens) and its availability varies due to different factors, e.g., island topography, rainfall
60 patterns, tides, episodic storms and human activities (White and Falkland, 2010; Storlazzi et
61 al., 2018). Seawater intrusion is thus an important issue due to its deleterious effect on
62 oceanic island freshwater storage (e.g., Werner et al., 2017; Lu et al., 2019; Memari et al.,
63 2020).

64 Over the past few decades, seawater intrusion in oceanic islands has been extensively
65 investigated in field observations (e.g., Röper et al., 2013; Post et al., 2019), laboratory
66 experiments (e.g., Stoeckl et al., 2015; Bedekar et al., 2019; Memari et al., 2020), numerical
67 simulations (e.g., Lam, 1974; Gingerich et al., 2017; Liu and Tokunaga, 2019) and analytical
68 solutions (e.g., Fetter, 1972; Ketabchi et al., 2014; Lu et al., 2019). Among these, analytical
69 solutions are effective tools to assess the extent of seawater intrusion (i.e., the location of the
70 freshwater-seawater interface), although they cannot incorporate complex factors (e.g.,
71 dispersive mixing and transient oceanic dynamics) (Werner et al., 2013). The advantages of
72 analytical solutions are that they are computationally efficient, can be used as test cases for
73 numerical models, and can reveal the explicit relationships between parameters that influence

74 seawater intrusion (e.g., Fetter, 1972; Ketabchi et al., 2014; Liu et al., 2014; Lu et al., 2019).

75 Based on the Dupuit-Forchheimer approximation (i.e., ignoring vertical flow) and the
76 Ghijben-Herzberg equation (Drabbe and Badon Ghijben, 1889, English translation given by
77 Post (2018); Herzberg, 1901), Fetter (1972) presented analytical solutions describing the
78 freshwater-seawater interface location and watertable elevation in a circular island. Bailey et
79 al. (2010) further compared these single-layered analytical solutions with field measurements,
80 indicating that the analytical solutions perform well in estimating the freshwater-seawater
81 interface location and watertable elevation. Fetter's solutions formed the foundation for many
82 subsequent analytical studies on seawater intrusion in island aquifers. Again, for a single
83 layer, Chesnaux and Allen (2008) and Greskowiak et al. (2013) developed analytical solutions
84 to predict the steady-state groundwater age distribution in freshwater lenses. In addition, using
85 single-layered analytical solutions, Morgan and Werner (2014) proposed vulnerability
86 indicators of freshwater lenses under sea-level rise and recharge change.

87 Since aquifers are usually heterogeneous, the single-layer analytical solutions were
88 subsequently extended to two-layered island aquifers. Vacher (1988) derived solutions for the
89 freshwater-seawater interface location and watertable elevation for infinite-strip islands
90 composed of different layers. Dose et al. (2014) conducted laboratory experiments to validate
91 and confirm the reliability of analytical solutions proposed by Fetter (1972) and Vacher
92 (1988). Ketabchi et al. (2014) extended Fetter's analytical solutions to calculate the
93 freshwater-seawater interface location and watertable elevation in two-layered circular islands
94 subject to sea-level rise. Their results indicated that land-surface inundation caused by sea-

95 level rise has a considerable impact on fresh groundwater lenses. Recently, Lu et al. (2019)
96 derived analytical solutions for the freshwater-seawater interface location and watertable
97 elevation for both strip and circular islands with two adjacent layers, i.e., a less permeable
98 slice along the shoreline of an island, and a more permeable zone inland.

99 All the abovementioned analytical solutions apply to either strip or circular islands.
100 According to the classification of sand dunes developed by Stuyfzand (1993; 2017), there are
101 different island layouts that should be considered, e.g., where the shape of the island is an
102 annulus segment, instead of a strip or circular disk (Figure 1). Annulus segment-shaped
103 islands are found in various atolls (i.e., circular chains of islands surrounding a central
104 lagoon) as found in the Pacific and Indian Oceans (Werner et al., 2017; Duvat, 2019).
105 Nevertheless, analytical solutions of seawater intrusion are not yet available for annulus
106 segment aquifers (ASAs). In general, ASAs are conceptually treated as a 2D cross section,
107 similar to strip islands (e.g., Ayers and Vacher, 1986; Underwood et al., 1992; Bailey et al.,
108 2009; Werner et al., 2017). Evidently, topography plays an important role in groundwater flow
109 and hence seawater intrusion (e.g., Zhang et al., 2016; Liu and Tokunaga, 2019). It remains
110 unclear whether analytical solutions of seawater intrusion for strip islands are appropriate for
111 ASAs. It is also unclear how island geometry affects the freshwater-seawater interface
112 location and watertable elevation of ASAs.

113 In this study, analytical solutions are derived for steady-state seawater intrusion for ASAs,
114 with a focus on the freshwater-seawater interface location and its corresponding watertable
115 elevation. After comparing their predictions with experimental data (Memari et al., 2020), the

116 analytical solutions are employed to investigate the effects of aquifer geometry on the
117 freshwater-seawater interface location and watertable elevation in ASAs.

118 **2. Conceptual Model**

119 Figure 2 shows the conceptual model of an ASA (a slice of an atoll island). The plan
120 view of the model domain is represented as a sector ($EF GH$) with an angle θ (Figure 2a).
121 The sea (EF) and lagoon (HG) boundaries are located at $L + L_0$ [L] and L_0 [L] from the circle
122 center, respectively. Since the longitudinal length is usually much longer than the lateral
123 length for an atoll island (Werner et al., 2017), seawater intrusion from the lateral sides (EH
124 and FG , Figure 2a) is negligible in comparison to the longitudinal side, especially for the
125 middle portion of an ASA. Therefore, EH and FG are treated as lateral no-flow boundaries.
126 Note that treating the lateral sides as no-flow boundaries is often used in studies of freshwater
127 lenses on atoll islands (e.g., Ayers and Vacher, 1986; Underwood et al., 1992; Bailey et al.,
128 2009; Werner et al., 2017). The lateral vertical cross section of the model domain is
129 conceptualized as a rectangle ($ABCD$) along the radial direction with dimensions of L [L]
130 (width) \times d [L] (height) (Figure 2b, c). AD is the impermeable base while BC is the land
131 surface through which aquifer recharge flows.

132 Both the sea and lagoon water levels are set to H_s [L], which results in an internal no-
133 flow boundary (water divide, where the slope of the watertable is zero) between the sea and
134 lagoon (location of the z -axis in Figure 2b,c). The segment between the sea and the internal
135 no-flow boundary is referred to as Unit 1, whereas the segment between the internal no-flow
136 and lagoon boundaries is referred to as Unit 2 (Figure 2). The widths of Units 1 and 2 are l_1

137 [L] and l_2 [L], respectively. In addition, the flow is asymmetrical in Units 1 and 2, with
138 divergent flow (the aquifer length w [L] increases along the flow direction) in Unit 1 and
139 convergent flow (w decreases along the flow direction) in Unit 2.

140 The r - z coordinate origin is placed at the intersection of the internal no-flow boundary
141 and impermeable base, with the r -axis pointing to the circle center (radial direction) and the z -
142 axis pointing vertically upward. Further, ϕ [L] is the watertable height, h [L] is the
143 vertical distance between the watertable and the interface, h_s [L] is the vertical distance
144 between the sea level and the interface, and $h_c = H_s - h_s$ [L] is the vertical distance from the
145 impermeable base to the interface for given r (Figure 2b,c). Constant recharge into the
146 saturated zone, N [LT⁻¹], is assumed. There are two possibilities for the interface tip (i.e., the
147 location where the freshwater-seawater interface connects to the z -axis or the bottom
148 boundary): above the aquifer bed (Figure 2b) or on the aquifer bed (Figure 2c). The r -
149 coordinates of the interface tip in Units 1 and 2 are denoted as r_{t1} [L] and r_{t2} [L], respectively
150 (Figure 2c). Note that $r_{t1} = r_{t2} = 0$ when the interface tip is above the aquifer bed, as in Figure
151 2b.

152 Consistent with previous studies (e.g., Ketabchi et al., 2014; Lu et al., 2016; 2019), the
153 following assumptions are made: (1) steady-state flow, (2) sharp freshwater-seawater
154 interface, (3) homogeneous and isotropic aquifer with a horizontal bottom, (4) aquifer
155 recharge, due to rainfall infiltration, is assumed to be stationery and uniform throughout the
156 domain, (5) vertical flow in the saturated zone is negligible (the Dupuit-Forchheimer
157 approximation), and (6) the same velocity is assumed on the arc (w) for a given radial

158 distance r , leading to radial flow only. Based on this last assumption, the 3D flow problem can
159 be simplified to 1D, making it possible to consider geometry effects analytically (Fan and
160 Bras, 1998; Paniconi et al., 2003; Troch et al., 2003).

161 **3. Analytical Solutions**

162 Under the abovementioned assumptions, groundwater flow in an ASA (Figure 2) can be
163 described as (Fan and Bras, 1998; Paniconi et al., 2003; Troch et al., 2003),

$$164 \quad -\frac{d}{dr}(wq) + Nw = 0 \quad (1)$$

165 where q [L^2T^{-1}] is the radial flux per unit length along the radial direction r [L]. Equation
166 (1) is a special case of the hillslope-storage Boussinesq equation proposed by Troch et al.
167 (2003). Paniconi et al. (2003) validated the hillslope-storage Boussinesq equation by
168 comparing it with a 3D Richards' equation model and found that predictions of the hillslope-
169 storage Boussinesq equation matched well those of the 3D model for seven different
170 geometries. For conciseness, readers are referred to Paniconi et al. (2003) for more details
171 about the validation. Subsequently, the hillslope-storage Boussinesq equation was used to for
172 different analyses (Hilberts et al., 2005, 2007; Hazenberg et al., 2015, 2016; Kong et al.,
173 2016; Luo et al., 2018), all of which focus on hillslope aquifers where the aquifer bottom is
174 usually sloping. The hillslope-storage Boussinesq equation assumes that groundwater flow is
175 parallel to the aquifer bottom (the Dupuit-Forchheimer approximation). Therefore, it can be
176 applied to coastal unconfined aquifers where the aquifer bottom slope is usually mild (Lu et
177 al., 2016).

178 According to Darcy's law and the Dupuit-Forchheimer approximation, the freshwater

179 flux in the aquifer segment between the seaward boundary and interface tip can be calculated
 180 as (ϕ is independent of z),

$$181 \quad q = -\int_{h_c}^{\phi} K_s \frac{d\phi}{dr} dz = -K_s (\phi - h_c) \frac{d\phi}{dr} \quad (2)$$

182 where K_s [LT^{-1}] is the saturated hydraulic conductivity.

183 3.1. Interface Tip above the Aquifer Bed

184 We first consider the situation where the interface tip is above the aquifer bed (Figure
 185 2b). In Unit 1 where $w = \theta(L_0 + l_2 - r)$, substituting equation (2) into equation (1) and then
 186 integrating gives,

$$187 \quad -\frac{1}{2} \left[(L_0 + l_2 - r)^2 - (L_0 + l_2)^2 \right] N = -(L_0 + l_2 - r) K_s (\phi - h_c) \frac{d\phi}{dr} \quad (3)$$

188 According to the Ghijben-Herzberg equation, the vertical thickness of the freshwater zone (h)
 189 in the interface zone is given by,

$$190 \quad h = \phi - h_c = (1 + \alpha)(\phi - H_s) \quad (4)$$

191 where $\alpha = \rho_f / (\rho_s - \rho_f)$ is the dimensionless density difference, and ρ_f [ML^{-3}] and ρ_s
 192 [ML^{-3}] are the freshwater and seawater densities, respectively. Substitution of equation (4)
 193 into equation (3) yields,

$$194 \quad -\frac{1}{2} \left[(L_0 + l_2 - r)^2 - (L_0 + l_2)^2 \right] N = -K_s (L_0 + l_2 - r) (1 + \alpha) (\phi - H_s) \frac{d\phi}{dr} \quad (5)$$

195 Rearranging equation (5) produces,

$$196 \quad -\frac{(L_0 + l_2 - r)N}{2} + \frac{N(L_0 + l_2)^2}{2(L_0 + l_2 - r)} = -K_s (1 + \alpha) (\phi - H_s) \frac{d\phi}{dr} \quad (6)$$

197 Integrating equation (6) leads to,

$$198 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r) - \frac{1}{2} (L_0 + l_2) Nr + \frac{1}{4} Nr^2 + C_1 = -K_s (1 + \alpha) \frac{(\phi - H_s)^2}{2} \quad (7)$$

199 where C_1 is the integration constant that is determined by the sea boundary condition (i.e.,

200 $r = -l_1, \phi = H_s$),

$$201 \quad C_1 = \frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 + l_1) - \frac{1}{2}(L_0 + l_2)l_1 N - \frac{1}{4}l_1^2 N \quad (8)$$

202 The relation between h_s and ϕ is given by,

$$203 \quad h_s = \alpha(\phi - H_s) \quad (9)$$

204 Combining equation (7) with equation (9) and eliminating ϕ yields,

$$205 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r) - \frac{1}{2}(L_0 + l_2)Nr + \frac{1}{4}Nr^2 + C_1 = -K_s(1 + \alpha)\frac{h_s^2}{2\alpha^2} \quad (10)$$

206 Equation (10) gives the freshwater-seawater interface location in Unit 1 once l_1 and l_2 are

207 determined.

208 Equation (7) applies to Unit 2 by replacing C_1 with C_2 ,

$$209 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r) - \frac{1}{2}(L_0 + l_2)Nr + \frac{1}{4}Nr^2 + C_2 = -K_s(1 + \alpha)\frac{(\phi - H_s)^2}{2} \quad (11)$$

210 where C_2 is chosen to satisfy the lagoon boundary condition ($r = l_2, \phi = H_s$),

$$211 \quad C_2 = \frac{(L_0 + l_2)^2 N}{2} \ln(L_0) + \frac{1}{2}(L_0 + l_2)l_2 N - \frac{1}{4}l_2^2 N \quad (12)$$

212 Combining equations (9) and (11) and eliminating ϕ leads to,

$$213 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r) - \frac{1}{2}(L_0 + l_2)Nr + \frac{1}{4}Nr^2 + C_2 = -K_s(1 + \alpha)\frac{h_s^2}{2\alpha^2} \quad (13)$$

214 Equation (13) gives the freshwater-seawater interface location in Unit 2 once l_2 is

215 determined. Since the sea level and lagoon water level are the same, an internal no-flow

216 boundary exists between the sea and lagoon, i.e.,

$$217 \quad r = 0, (h_s)_{unit1} = (h_s)_{unit2} \quad (14)$$

218 where $(h_s)_{unit1}$ and $(h_s)_{unit2}$ represent h_s in Units 1 and 2, respectively.

219 Combining equations (10), (13) and (14) leads to expressions for l_1 and l_2 ,

$$220 \quad l_1 = L + L_0 - \sqrt{\frac{2LL_0 + L^2}{2\ln(L + L_0) - 2\ln(L_0)}} \quad (15)$$

$$221 \quad l_2 = \sqrt{\frac{2LL_0 + L^2}{2\ln(L + L_0) - 2\ln(L_0)}} - L_0 \quad (16)$$

222 As indicated by equations (15) and (16), the internal no-flow boundary between the sea and
 223 lagoon only depends on L and L_0 . For known l_1 and l_2 , equations (10) and (13) can be
 224 employed to predict the freshwater-seawater interface location in Units 1 and 2, respectively.

225 Once the interface location is determined, h and ϕ are given by,

$$226 \quad h = \frac{1+\alpha}{\alpha} h_s \quad (17)$$

$$227 \quad \phi = \frac{h_s}{\alpha} + H_s \quad (18)$$

228 3.2. Interface Tip on the Aquifer Bed

229 When the interface tip is on the aquifer bed, the location of the internal no-flow
 230 boundary remains the same as for the interface tip above the aquifer bed. The freshwater-
 231 seawater interface for Units 1 and 2 can be determined by equations (10) and (13),
 232 respectively. Then, from equation (17), h at the aquifer segment between the sea boundary and
 233 the interface tip is determined. To calculate h for the aquifer segment between the interface tip
 234 and the internal no-flow boundary, the r -coordinate of the interface tip is found. At the
 235 interface tip of Unit 1 ($r = r_{t1}$),

$$236 \quad h_s = H_s \quad (19)$$

$$237 \quad \phi = \frac{1+\alpha}{\alpha} H_s \quad (20)$$

238 With equations (10) and (20), r_{t1} is given by,

$$239 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r_{t1}) - \frac{1}{2}(L_0 + l_2)Nr_{t1} + \frac{1}{4}Nr_{t1}^2 = -C_1 - K_s(1 + \alpha)\frac{H_s^2}{2\alpha^2} \quad (21)$$

240 Let,

$$241 \quad a = \frac{1}{4}N \quad (22a)$$

$$242 \quad b = -\frac{1}{2}(L_0 + l_2)N \quad (22b)$$

$$243 \quad c = -\frac{(L_0 + l_2)^2 N}{2} \quad (22c)$$

244 and

$$245 \quad m = -C_1 - K_s(1 + \alpha)\frac{H_s^2}{2\alpha^2} \quad (22d)$$

246 then equation (21) becomes,

$$247 \quad ar_{t1}^2 + br_{t1} + c \ln(L_0 + l_2 - r_{t1}) = m \quad (23)$$

248 which is solved by a root-finding method.

249 The freshwater discharge for the aquifer segment between the interface tip and the
250 internal no-flow boundary is calculated as,

$$251 \quad -\frac{1}{2}[(L_0 + l_2 - r)^2 - (L_0 + l_2)^2]N = -(L_0 + l_2 - r)K_s\phi\frac{d\phi}{dr} \quad (24)$$

252 Repeating the steps from equations (3) to (7) gives,

$$253 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - r) - \frac{1}{2}(L_0 + l_2)Nr + \frac{1}{4}Nr^2 + C_3 = -\frac{K_s}{2}\phi^2 \quad (25)$$

254 where C_3 is determined by substituting equation (20) into equation (25). Then, equation (25)
255 can be adopted to calculate h for the segment between the interface tip and the internal no-
256 flow boundary where $h = \phi$.

257 Similarly, the r -coordinate of the interface tip in Unit 2 (r_{i2}) is obtained by substituting
258 equation (19) into equation (13). Then, the watertable (h) of the aquifer segment between the
259 interface tip and the internal no-flow boundary for Unit 2 is computed by repeating the steps
260 from equations (21) to (25).

261 **4. Results and Discussion**

262 **4.1. Validation of the Analytical Solutions**

263 The analytical solutions were validated by comparing their predictions with experimental
264 data compiled from Memari et al. (2020), who reported experiments carried out using a 15°
265 radial tank. The tank contained three distinct chambers: internal no-flow boundary condition,
266 porous medium and constant-head boundary condition (i.e., sea or lagoon). The internal no-
267 flow and seaward boundaries were respectively located at 10 and 55.5 cm from the circle
268 center, i.e., 45.5 cm from the internal no-flow boundary to the constant-head boundary along
269 the radial direction. Note that the experimental tank corresponds to Unit 1 of the radial aquifer
270 with $l_1 = 45.5$ cm and $l_2 = 0$, so the analytical results were calculated using equations (10)
271 and (23). The thicknesses of the porous medium and sea level were 28 and 25 cm,
272 respectively, with $K_s = 1.23 \times 10^{-2}$ m s⁻¹. The measured saltwater and freshwater densities
273 were respectively 1.015 and 0.999 g ml⁻¹, leading to $\alpha = 62$. Two different recharge events
274 with constant N , 2.46×10^{-4} and 1.08×10^{-4} m s⁻¹, were considered in the experiments.

275 Figure 3 shows the comparison between analytical and experimental results of the
276 freshwater-seawater interface for different recharge events. In general, the analytical solution
277 predicts the freshwater-seawater interface well for both recharge events, despite there being

278 some differences between the analytical results and the measurements, particularly in the zone
279 near the constant-head boundary ($r = -45$ cm). These deviations are likely due to assumptions
280 made in the analytical solution, i.e., (i) a sharp freshwater-seawater interface, (ii) ignoring the
281 effect of freshwater discharge, and (iii) neglecting the vertical flow (the Dupuit-Forchheimer
282 approximation).

283 **4.2. Effects of Aquifer Geometry on Seawater Intrusion**

284 Previous studies showed that boundary conditions play a critical role in estimates of
285 seawater intrusion (Werner and Simmons, 2009; Lu et al., 2016). Therefore, the internal no-
286 flow boundary between the sea and lagoon was examined for various ASAs. As indicated by
287 equations (15) and (16), this internal no-flow boundary depends only on L and L_0 . The values
288 of l_1 and l_2 calculated respectively from equations (15) and (16) are shown in Figure 4 for
289 three typical values of L (500, 1000 and 2000 m) with L_0 varying from 10^2 to 10^6 m. In
290 general, the internal no-flow boundary deviates from the middle of the ASA. When L_0 is less
291 than 10^5 m, l_1 is larger than l_2 for the three different values of L , indicating an internal no-
292 flow boundary closer to the lagoon boundary. For example, taking $L = 2000$ m and $L_0 = 100$ m
293 leads to $l_1 = 1240$ m and $l_2 = 760$ m, with a deviation of 240 m (12% of 2000 m) from the
294 middle of the ASA. When L_0 exceeds 10^5 m, however, the location of the internal no-flow
295 boundary can be approximated as being at the middle of the ASA for all considered values of
296 L . This is in contrast to strip and circular aquifers where the internal no-flow boundary is
297 always in the middle of aquifer due to symmetry.

298 Since the internal no-flow boundary location between the sea and lagoon deviates from

299 the middle of the ASA, we expect aquifer geometry to play a significant role in controlling
300 seawater intrusion. As mentioned previously, ASAs can be convergent (Unit 1) or divergent
301 aquifers (Unit 2) where the extent of seawater intrusion may be different. However, for strip
302 aquifers, both Units 1 and 2 are rectangular with the same extent of seawater intrusion.
303 Therefore, three geometries were compared in this study: convergent, rectangular and
304 divergent (Figure 5). These geometries have been widely examined in hillslope hydrology
305 regrading to the effects of aquifer geometry on runoff generation (Troch et al., 2003; Kong et
306 al., 2016; Luo et al., 2018). To present the results more conveniently, we placed the r - z
307 coordinate origin at the intersection of the constant-head boundary (sea or lagoon) and the
308 impermeable base, with the r -axis pointing horizontally to the internal no-flow boundary and
309 the z -axis vertically upward (Figure 5). In addition, the distance between the constant-head
310 boundary and the internal no-flow boundary (aquifer width) is denoted as L^* (Figure 5) while
311 the other parameters remain the same.

312 Following previous studies (e.g., Lu et al., 2016; 2019), different cases were selected to
313 show the effects of aquifer geometry on seawater intrusion (Cases 1 and 2 in Table 1).
314 According to Werner et al. (2017), the width of atoll islands generally varies from 100 to 1500
315 m along the radial direction. In order to focus on the effects of aquifer geometry on seawater
316 intrusion, the same L^* and L_0 were assumed for the three aquifers, with L^* and L_0 equal to
317 1000 and 200 m, respectively. Note that L_0 is the distance from the circle center to the lagoon
318 boundary for convergent aquifers, whereas it represents the distance from the circle center to
319 internal no-flow boundary for divergent aquifers hereafter. The sand characteristics were the

320 same as in the experiments of Memari et al. (2020). Two recharge events were considered
321 (Cases 1 and 2, Table 1). The freshwater-seawater interface was calculated using the
322 analytical solutions for the three different aquifers. Note that the Appendix presents analytical
323 solutions for seawater intrusion in strip aquifers deduced from Lu et al. (2019).

324 Figure 6 shows the freshwater-seawater interface calculated for Cases 1 and 2. As can be
325 seen, the extent of seawater intrusion is noticeably different for the three aquifer geometries.
326 For high recharge ($1 \times 10^{-6} \text{ m s}^{-1}$), the interface tip is located at around 500 m for the
327 divergent aquifer, which is about twice the value of the rectangular aquifer and six times the
328 value for the convergent aquifer (Figure 6a). When the recharge decreases to $3 \times 10^{-7} \text{ m s}^{-1}$,
329 the interface tip moves further landward for the three aquifers as expected, but the difference
330 between results is still great (Figure 6b). The interface tip is displaced above the aquifer bed
331 for both the rectangular and divergent aquifers, while it remains on the aquifer bed for the
332 convergent aquifer. Regardless of the recharge rate, the most landward freshwater-seawater
333 interface occurs in the divergent aquifer and vice versa for the convergent aquifer. This
334 underlines that aquifer geometry plays a major role in controlling seawater intrusion and
335 hence it is necessary to account for aquifer geometry in analyses of seawater intrusion.

336 4.3. Sensitivity Analysis

337 A sensitivity analysis was conducted to investigate to what extent aquifer geometry
338 affects seawater intrusion. Since we focus on the effects of aquifer geometry on the locations
339 of the freshwater-seawater interface and watertable, values of L_0 and L^* were varied, with
340 other parameters kept constant. When conducting the sensitivity analysis of L_0 , L^* was fixed

341 at 1000 m, which is a typical value for ASAs (Werner et al., 2017). Figure 7 shows the
342 sensitivity of the locations of the freshwater-seawater interface and watertable to changes in
343 L_0 (Case 3, Table 1). The freshwater-seawater interface and watertable elevation are
344 independent of L_0 for rectangular aquifers (Appendix). However, the freshwater-seawater
345 interface and watertable elevation differ greatly when varying L_0 for both convergent and
346 divergent aquifers, highlighting that L_0 plays an important role in affecting seawater intrusion.
347 Specifically, as L_0 increases, the freshwater-seawater interface moves more landward (larger
348 r/L^* , Figure 7a) and its corresponding watertable elevation decreases (Figure 7c) for
349 convergent aquifers. In contrast, for divergent aquifers increasing L_0 moves the freshwater-
350 seawater interface more seaward (smaller r/L^* , Figure 7b) and its corresponding watertable
351 elevation increases (Figure 7d). For a given L_0 , divergent aquifers have the largest extent of
352 seawater intrusion and the lowest watertable elevation, and conversely for convergent aquifers
353 (Figure 7).

354 Regardless of the freshwater-seawater interface and watertable elevation, the deviation
355 between rectangular aquifers and divergent or convergent aquifers is significant when L_0 is
356 less than 2000 m (Figure 7). For example, the r -coordinate of the interface tip ($z = 0$) is 262 m
357 for the rectangular aquifer at $L_0 = 200$ m, whereas it is 78 (31% of that in the rectangular
358 aquifer) and 500 m (191% of that in the rectangular aquifer) for the convergent and divergent
359 aquifers, respectively. As L_0 increases, the deviation between the three aquifers decreases.
360 When $L_0 = 2000$ m, the r -coordinate of the interface tip is 262, 209 (80% of that in the
361 rectangular aquifer) and 318 m (121% of that in the rectangular aquifer) for the rectangular,

362 convergent and divergent aquifers, respectively. As L_0 increases to 6000 m, the freshwater-
363 seawater interface and watertable elevation of both convergent and divergent aquifers tend to
364 those of rectangular aquifers, i.e., geometry effects decrease with increasing L_0 . These results
365 highlight the critical role played by the shape of aquifers. As a result, ignoring the aquifer
366 geometry may lead to an inappropriate management strategy for groundwater resources in
367 atoll islands.

368 The sensitivity of the freshwater-seawater interface and watertable elevation to L^* was
369 investigated by varying L^* from 600 to 1600 m while fixing L_0 to 200 m (Case 4, Table 1). As
370 shown in Figure 8, contrary to the results for varying L_0 , in this case the freshwater-seawater
371 interface and watertable elevation in all three topographies are related to L^* . Again, the extent
372 of seawater intrusion is greatest in divergent aquifers and least in convergent aquifers for
373 given L^* . When L^* increases, the freshwater-seawater interface moves seaward and the
374 watertable elevation increases, regardless of aquifer geometry, i.e., the seawater intrusion
375 decreases (Figures 8a-c). This is because the total freshwater flux increases with increasing
376 L^* , leading to a higher hydraulic gradient and hence less seawater intrusion (Figures 8d-f).
377 Moreover, an increase in L^* reduces the differences in the seawater intrusion distance among
378 the three geometries, i.e., the effects of aquifer geometry on seawater intrusion are more
379 significant at small L^* . However, even at the maximum L^* considered (1600 m), the deviation
380 between three aquifers remains significant: The r -coordinate of the interface tip is about 148
381 m for the rectangular aquifer, whereas it is about 32 (22% of that in the rectangular aquifer)
382 and 278 m (188% of that in the rectangular aquifer) for the convergent and divergent aquifers,

383 respectively. Both L_0 and L^* can greatly impact seawater intrusion estimates for divergent and
384 convergent aquifers, highlighting the necessity to include geometry effects in analytical
385 solutions of seawater intrusion.

386 **5. Conclusions**

387 Based on the Ghijben-Herzberg and hillslope-storage Boussinesq equations, we derived
388 analytical solutions of steady-state seawater intrusion for ASAs, with a focus on the
389 freshwater-seawater interface and its corresponding watertable elevation as affected by
390 recharge. After comparing with experimental data of Memari et al. (2020), the analytical
391 solutions were employed to examine the effects of aquifer geometry on seawater intrusion in
392 island aquifers. Three different shapes of island aquifer were compared: convergent,
393 rectangular and divergent. The results lead to the following conclusions:

- 394 • The presented analytical solutions perform well in predicting the experimental freshwater-
395 seawater interface, suggesting that these analytical solutions can predict seawater intrusion
396 reasonably in different aquifer geometries.
- 397 • Island geometry plays a significant role in affecting the freshwater-seawater interface and
398 watertable elevation. Other factors being equal, the extent of seawater intrusion is greatest
399 in divergent aquifers, and conversely least in convergent aquifers. In contrast, the
400 watertable elevation is lowest in divergent aquifers and highest in convergent aquifers.
- 401 • The effects of aquifer geometry on seawater intrusion are dependent on the aquifer width
402 and distance from the circle center to the internal no-flow boundary (Figures 7 and 8). A
403 larger aquifer width and distance from the circle center to the inner arc (the lagoon

404 boundary for convergent aquifers or the internal no-flow boundary for divergent aquifers)
405 weaken the role played by aquifer geometry and hence lead to a smaller deviation of the
406 extent of seawater intrusion between the three topographies.

407 Real island aquifers are expected to exhibit more complexity than considered here, e.g.,
408 they will have more complex shapes and are subjected to transient flow conditions caused by
409 tides, waves and groundwater pumping (Mantoglou et al. 2003; Pool and Carrera., 2011;
410 Werner et al., 2013). In addition, since the experimental scale of Memari et al. (2020) is
411 necessarily small, future experiments and field data are needed to further validate and
412 facilitate the analytical solutions. Despite this, the new analytical solutions, validated against
413 experiments, can be used as a tool for rapid estimation of seawater intrusion in ASAs once
414 known island geometry and corresponding soil properties are given.

415 **Appendix: Analytical Solutions for Rectangular Aquifers**

416 For rectangular aquifers, the seawater intrusion in Unit 1 is identical to that in Unit 2
 417 because of symmetry. With the interface tip on the aquifer bed, analytical solutions for the
 418 freshwater-seawater interface (h_s), watertable elevation (h), and r -coordinate of the interface
 419 tip in Unit 2 (r_{t2}) can be respectively written as (Lu et al., 2019),

$$420 \quad h_s = \alpha \sqrt{\frac{N}{(1+\alpha)K_s} \left(\frac{L^2}{4} - r^2 \right)} \quad (\text{A1})$$

$$421 \quad h = \begin{cases} \sqrt{\frac{N}{K_s} (r_{t2}^2 - r^2) + \left(\frac{H_s}{\alpha} + H_s \right)} & 0 \leq r \leq r_{t2} \\ \sqrt{\frac{N}{(1+\alpha)K_s} \left(\frac{L^2}{4} - r^2 \right) + H_s} & r_{t2} < r \leq \frac{L}{2} \end{cases} \quad (\text{A2})$$

$$422 \quad r_{t2} = \sqrt{\frac{L^2}{4} - \frac{(1+\alpha)K_s}{N} \left(\frac{H_s^2}{\alpha^2} \right)} \quad (\text{A3})$$

423 When the interface tip is above the aquifer bed, the analytical solution for the freshwater-
 424 seawater interface location and watertable elevation in Unit 2 are the same as equations (A1)
 425 and (A2), respectively.

426 **Code/Data availability**

427 Experimental data used in this study were compiled from Memari et al. (2020).

428 **Author contributions**

429 All authors contributed to the design of the research. ZL carried out data collation,
430 developed the analytical solutions and prepared the manuscript with contributions from all
431 co-authors. All authors contributed to the interpretation of the results and provided feedback.

432 **Competing interests**

433 The authors declare that they have no conflicts of interest.

434 **Acknowledgments**

435 This research was supported by the National Key R&D Program of China
436 (2019YFC0409004) and the National Natural Science Foundation of China (51979095 and
437 41807178). ZL acknowledges EPFL for financial support and JK acknowledges the Qing Lan
438 Project of Jiangsu Province (2020). We appreciate the constructive comments from the
439 handling Editor Mauro Giudici and three anonymous reviewers, which led to significant
440 improvement of the paper.

441 **References**

- 442 Ayers, J. F., and Vacher, H. L.: Hydrogeology of an atoll island: A conceptual model from
443 detailed study of a Micronesian example, *Groundwater*, 24, 185-198,
444 <https://doi.org/10.1111/j.1745-6584.1986.tb00994.x>, 1986.
- 445 Bailey, R. T., Jenson, J. W., and Olsen, A. E.: Estimating the ground water resources of atoll
446 islands, *Water*, 2, 1-27, <https://doi.org/10.3390/w2010001>, 2010.
- 447 Bailey, R. T., Jenson, J. W., and Olsen, A. E.: Numerical modeling of atoll island
448 hydrogeology, *Groundwater*, 47, 184-196, [https://doi.org/10.1111/j.1745-](https://doi.org/10.1111/j.1745-6584.2008.00520.x)
449 [6584.2008.00520.x](https://doi.org/10.1111/j.1745-6584.2008.00520.x), 2009.
- 450 Bedekar, V. S., Memari, S. S., and Clement, T. P.: Investigation of transient freshwater storage
451 in island aquifers, *J. Contam. Hydrol.*, 221, 98-107,
452 <https://doi.org/10.1016/j.jconhyd.2019.02.004>, 2019.
- 453 Chesnaux, R., and Allen, D. M.: Groundwater travel times for unconfined island aquifers
454 bounded by freshwater or seawater, *Hydrogeol. J.*, 16, 437-445,
455 <https://doi.org/10.1007/s10040-007-0241-6>, 2008.
- 456 Dose, E. J., Stoeckl, L., Houben, G. J., Vacher, H. L., Vassolo, S., Dietrich, J., and
457 Himmelsbach, T.: Experiments and modeling of freshwater lenses in layered aquifers:
458 Steady state interface geometry, *J. Hydrol.*, 509, 621-630,
459 <https://doi.org/10.1016/j.jhydrol.2013.10.010>, 2014.
- 460 Drabbe J. and Badon Ghijben, W. (1889): Nota in verband met de voorgenomen put boring
461 nabij Amsterdam, *Tijdschrift van het Koninklijk Instituut van Ingenieurs*. pp. 8-22,

462 Gravenhage, Netherlands, 1889.

463 Duvat, V. K. E.: A global assessment of atoll island planform changes over the past decades,
464 Wiley Interdiscip. Rev. Clim. Change, 10, e557, <https://doi.org/10.1002/wcc.557>,
465 2019.

466 Fan, Y., and Bras, R. L.: Analytical solutions to hillslope subsurface storm flow and saturation
467 overland flow, Water Resour. Res., 34, 921-927, <https://doi.org/10.1029/97WR03516>,
468 1998.

469 Fetter, C. W.: Position of the saline water interface beneath oceanic islands, Water Resour.
470 Res., 8, 1307-1315, <https://doi.org/10.1029/WR008i005p01307>, 1972.

471 Gingerich, S. B., Voss, C. I., and Johnson, A. G.: Seawater-flooding events and impact on
472 freshwater lenses of low-lying islands: Controlling factors, basic management and
473 mitigation, J. Hydrol., 551, 676-688, <https://doi.org/10.1016/j.jhydrol.2017.03.001>,
474 2017.

475 Greskowiak, J., Röper, T., and Post, V. E. (2013).: Closed-form approximations for two-
476 dimensional groundwater age patterns in a fresh water lens, Groundwater, 51, 629-
477 634. <https://doi.org/10.1111/j.1745-6584.2012.00996.x>, 2013.

478 Hazenberg, P., Fang, Y., Broxton, P., Gochis, D., Niu, G. Y., Pelletier, J. D., Troch., P. A., and
479 Zeng, X.: A hybrid-3D hillslope hydrological model for use in Earth system models.
480 Water Resour. Res., 51, 8218-8239, <https://doi.org/10.1002/2014WR016842>, 2015.

481 Hazenberg, P., Broxton, P., Gochis, D., Niu, G. Y., Pangle, L. A., Pelletier, J. D., Troch., P. A.,
482 and Zeng, X. (2016).: Testing the hybrid-3-D hillslope hydrological model in a

483 controlled environment, *Water Resour. Res.*, 52, 1089-1107,
484 <https://doi.org/10.1002/2015WR018106>, 2016.

485 Herzberg, A.: Die wasserversorgung einiger Nordseebäder. *Journal für Gasbeleuchtung und*
486 *Wasserversorgung*, 44, 815-819, 45, 842-844, 1901.

487 Hilberts, A. G. J., Troch, P. A., and Paniconi, C.: Storage-dependent drainable porosity for
488 complex hillslopes, *Water Resour. Res.*, 41, W06001,
489 <https://doi.org/10.1029/2004WR003725>, 2005.

490 Hilberts, A. G., Troch, P. A., Paniconi, C., and Boll, J.: Low-dimensional modeling of
491 hillslope subsurface flow: Relationship between rainfall, recharge, and unsaturated
492 storage dynamics, *Water Resour. Res.*, 43, W03445,
493 <https://doi.org/10.1029/2006WR004964>, 2007.

494 Ketabchi, H., Mahmoodzadeh, D., Ataie-Ashtiani, B., Werner, A. D., and Simmons, C. T.
495 (2014). Sea-level rise impact on fresh groundwater lenses in two-layer small islands,
496 *Hydrol. Process.*, 28, 5938-5953, <https://doi.org/10.1002/hyp.10059>, 2014.

497 Kong, J., Shen, C., Luo, Z., Hua, G., and Zhao, H.: Improvement of the hillslope-storage
498 Boussinesq model by considering lateral flow in the unsaturated zone, *Water Resour.*
499 *Res.*, 52, 2965-2984, <https://doi.org/10.1002/2015WR018054>, 2016.

500 Lam, R. K.: Atoll permeability calculated from tidal diffusion, *J. Geophys. Res.*, 79, 3073-
501 3081, <https://doi.org/10.1029/JC079i021p03073>, 1974.

502 Liu, J., and Tokunaga, T.: Future risks of tsunami-induced seawater intrusion into unconfined
503 coastal aquifers: Insights from numerical simulations at Niijima Island, Japan, *Water*

504 Resour. Res., 55, 10082-10104, <https://doi.org/10.1029/2019WR025386>, 2019.

505 Liu, Y., X. Mao, J. Chen, and D. A. Barry.: Influence of a coarse interlayer on seawater intrusion
506 and contaminant migration in coastal aquifers, *Hydrol. Process.*, 28, 5162-5175.
507 <https://dx.doi.org/10.1002/hyp.10002>, 2014.

508 Lu, C., Cao, H., Ma, J., Shi, W., Rathore, S. S., Wu, J., and Luo, J.: A proof-of-concept study
509 of using a less permeable slice along the shoreline to increase fresh groundwater
510 storage of oceanic islands: Analytical and experimental validation, *Water Resour. Res.*,
511 55, 6450-6463, <https://doi.org/10.1029/2018WR024529>, 2019.

512 Lu, C., Xin, P., Kong, J., Li, L., and Luo, J.: Analytical solutions of seawater intrusion in
513 sloping confined and unconfined coastal aquifers, *Water Resour. Res.*, 52, 6989-7004,
514 <https://doi.org/10.1002/2016WR019101>, 2016.

515 Luo, Z., Shen, C., Kong, J., Hua, G., Gao, X., Zhao, Z., Zhao, H., and Li, L.: Effects of
516 unsaturated flow on hillslope recession characteristics, *Water Resour. Res.*, 54, 2037-
517 2056, <https://doi.org/10.1002/2017WR022257>, 2018.

518 Mantoglou, A.: Pumping management of coastal aquifers using analytical models of saltwater
519 intrusion, *Water Resour. Res.*, 39, 1335, <https://doi.org/10.1029/2002WR001891>,
520 2003.

521 Memari, S. S., Bedekar, V. S., and Clement, T. P.: Laboratory and numerical investigation of
522 saltwater intrusion processes in a circular island aquifer, *Water Resources Research*,
523 56(2), e2019WR025325, <https://doi.org/10.1029/2019WR025325>, 2020.

524 Morgan, L. K., and Werner, A. D.: Seawater intrusion vulnerability indicators for freshwater

525 lenses in strip islands, *J. Hydrol.*, 508, 322-327,
526 <https://doi.org/10.1016/j.jhydrol.2013.11.002>, 2014.

527 Paniconi, C., Troch, P. A., Van Loon, E. E., and Hilberts, A. G.: Hillslope-storage Boussinesq
528 model for subsurface flow and variable source areas along complex hillslopes: 2.
529 Intercomparison with a three-dimensional Richards equation model, *Water Resour.*
530 *Res.*, 39, 1317, <https://doi.org/10.1029/2002WR001730>, 2003.

531 Pool, M., and Carrera, J.: A correction factor to account for mixing in Ghyben-Herzberg and
532 critical pumping rate approximations of seawater intrusion in coastal aquifers, *Water*
533 *Resour. Res.*, 47, W05506, <https://doi.org/10.1029/2010WR010256>, 2011.

534 Post, V. E.: Annotated translation of “Nota in verband met de voorgenomen putboring nabij
535 Amsterdam [Note concerning the intended well drilling near Amsterdam]” by J.
536 Drabbe and W. Badon Ghijben (1889), *Hydrogeol. J.*, 26, 1771-1788,
537 <https://doi.org/10.1007/s10040-018-1797-z>, 2018.

538 Post, V. E. A., Houben, G. J., Stoeckl, L., and Sültenfuß, J.: Behaviour of tritium and
539 tritiogenic helium in freshwater lens groundwater systems: Insights from Langeoog
540 Island, Germany, *Geofluids*, Volume 2019, Article ID 1494326,
541 <https://doi.org/10.1155/2019/1494326>, 2019.

542 Röper, T., Greskowiak, J., Freund, H., and Massmann, G.: Freshwater lens formation below
543 juvenile dunes on a barrier island (Spiekeroog, Northwest Germany), *Estuar. Coast.*
544 *Shelf Sci.*, 121-122, 40-50, <https://doi.org/10.1016/j.ecss.2013.02.004>, 2013.

545 Stoeckl, L., Houben, G. J., and Dose, E. J.: Experiments and modeling of flow processes in

546 freshwater lenses in layered island aquifers: Analysis of age stratification, travel times
547 and interface propagation, *J. Hydrol.*, 529, 159-168,
548 <https://doi.org/10.1016/j.jhydrol.2015.07.019>, 2015.

549 Storlazzi, C. D., Gingerich, S. B., van Dongeren, A., Cheriton, O. M., Swarzenski, P. W.,
550 Quataert, E., Voss, C. I., Field, D. W., Annamalai, H., Piniak, G. A., and McCall, R.:
551 Most atolls will be uninhabitable by the mid-21st century because of sea-level rise
552 exacerbating wave-driven flooding, *Sci. Adv.*, 4, eaap9741,
553 <https://doi.org/10.1126/sciadv.aap9741>, 2018.

554 Strack, O. D. L.: A single-potential solution for regional interface problems in coastal
555 aquifers, *Water Resour. Res.*, 12, 1165-1174,
556 <https://doi.org/10.1029/WR012i006p01165>, 1976.

557 Stuyfzand, P. J.: Observations and analytical modeling of freshwater and rainwater lenses in
558 coastal dune systems, *J. Coast. Conserv.*, 21, 577-593, [https://doi.org/10.1007/s11852-](https://doi.org/10.1007/s11852-016-0456-6)
559 [016-0456-6](https://doi.org/10.1007/s11852-016-0456-6), 2017.

560 Stuyfzand, P. J.: Hydrochemistry and hydrology of the coastal dune area of the Western
561 Netherlands, Ph.D. Thesis, Vrije University, Amsterdam, KIWA, ISBN 90-74741-01-0,
562 <http://dare.uvu.vu.nl/handle/1871/12716>, 1993.

563 Thomas, A., Baptiste, A., Martyr-Koller, R., Pringle, P., and Rhiney, K.: Climate change and
564 small island developing states, *Annu. Rev. Environ. Resour.*, 45, 1-27,
565 <https://doi.org/10.1146/annurev-environ-012320-083355>, 2020.

566 Troch, P. A., Paniconi, C., and Emiel van Loon, E.: Hillslope-storage Boussinesq model for

567 subsurface flow and variable source areas along complex hillslopes: 1. Formulation
568 and characteristic response, *Water Resour. Res.*, 39, 1316,
569 <https://doi.org/10.1029/2002WR001728>, 2003.

570 Underwood, M. R., Peterson, F. L., and Voss, C. I.: Groundwater lens dynamics of atoll
571 islands, *Water Resour. Res.*, 28, 2889-2902, <https://doi.org/10.1029/92WR01723>,
572 1992.

573 Vacher, H. L.: Dupuit-Ghyben-Herzberg analysis of strip-island lenses. *Geol. Soc. Am. Bull.*,
574 100, 580-591, [https://doi.org/10.1130/0016-
575 7606\(1988\)100<0580:DGHAOS>2.3.CO;2](https://doi.org/10.1130/0016-7606(1988)100<0580:DGHAOS>2.3.CO;2), 1988.

576 Werner, A. D., Sharp, H. K., Galvis, S. C., Post, V. E., and Sinclair, P.: Hydrogeology and
577 management of freshwater lenses on atoll islands: Review of current knowledge and
578 research needs, *J. Hydrol.*, 551, 819-844,
579 <https://doi.org/10.1016/j.jhydrol.2017.02.047>, 2017.

580 Werner, A. D., Bakker, M., Post, V. E., Vandenbohede, A., Lu, C., Ataie-Ashtiani, B.,
581 Simmons, C. T., and Barry, D. A.: Seawater intrusion processes, investigation and
582 management: Recent advances and future challenges, *Adv. Water Resour.*, 51, 3-26,
583 <https://doi.org/10.1016/j.advwatres.2012.03.004>, 2013.

584 Werner, A. D., and Simmons, C. T.: Impact of sea-level rise on sea water intrusion in coastal
585 aquifers, *Groundwater*, 47, 197-204, [https://doi.org/10.1111/j.1745-
586 6584.2008.00535.x](https://doi.org/10.1111/j.1745-6584.2008.00535.x), 2009.

587 White, I., and Falkland, T.: Management of freshwater lenses on small Pacific islands,

588 Hydrogeol. J., 18, 227-246, <https://doi.org/10.1007/s10040-009-0525-0>, 2010.

589 Zhang, Y., Li, L., Erler, D. V., Santos, I., and Lockington, D.: Effects of alongshore
590 morphology on groundwater flow and solute transport in a nearshore aquifer, *Water*
591 *Resour. Res.*, 52, 990-1008, <https://doi.org/10.1002/2015WR017420>, 2016.

592 **Table 1.** List of parameters use in different simulations.

	No.	L^* (m)	L_0 (m)	H_s (m)	d (m)	α (-)	K_s (m s ⁻¹)	N (m s ⁻¹)
Cases	1	1000	200	38	45	40	1.23×10^{-2}	1×10^{-6}
	2	1000	200	38	45	40	1.23×10^{-2}	3×10^{-7}
	3	1000	†	38	45	40	1.23×10^{-2}	1×10^{-6}
	4	†	200	38	45	40	1.23×10^{-2}	1×10^{-6}

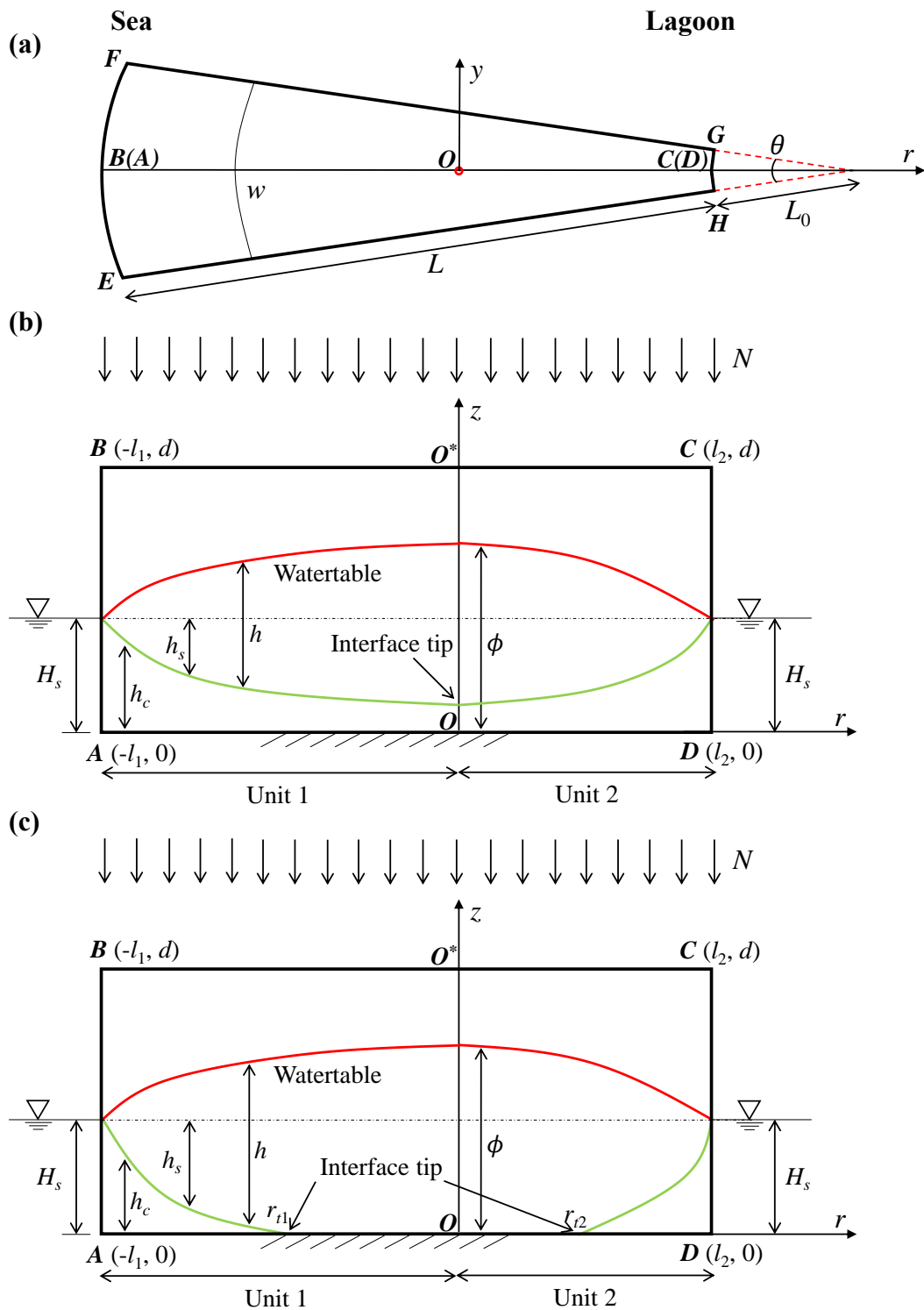
593 †The parameter is varied: The range of L_0 is from 200 to 6000 m, whereas the range of L^* is
 594 from 600 to 1600 m.



595

596 **Figure 1.** Island with an annulus segment in the Namu Atoll, Marshall Islands (© Google

597 Earth 2021).



598

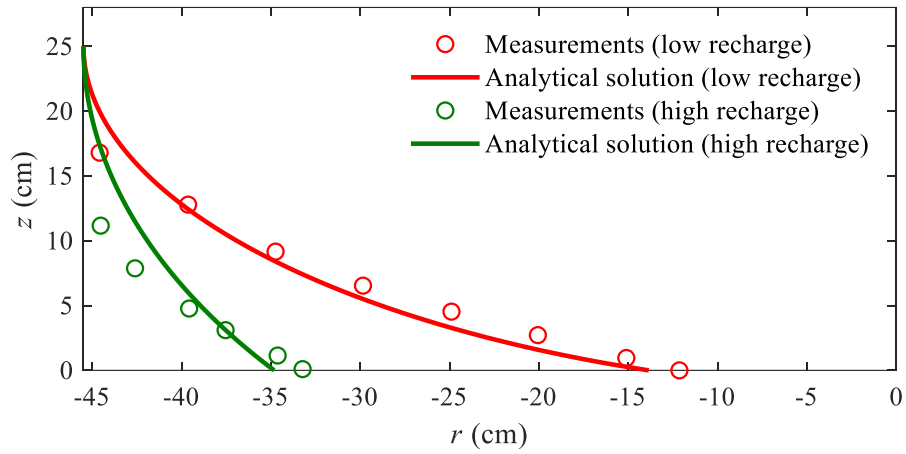
599 **Figure 2.** Conceptual model of an annulus segment aquifer (a slice of an atoll island). (a) Plan

600 view and (b, c) lateral vertical cross section with the saltwater interface tip (b) above the

601 aquifer bed (single location) and (c) on the aquifer bed (two locations). In (a), the sea

602 boundary is on EF and the atoll lagoon boundary is on HG ; In (b) and (c), AD is the

603 impermeable base and OO^* is the internal no-flow boundary.

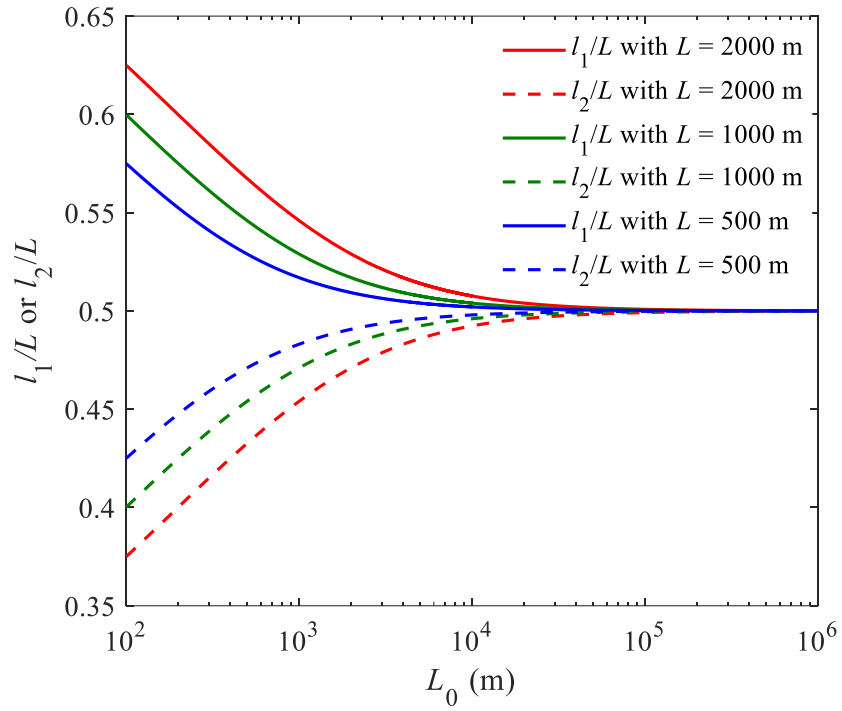


604

605 **Figure 3.** Comparison between analytical and experimental (data compiled from Memari et

606 al., 2020) results for the freshwater-seawater interface location for different recharge events.

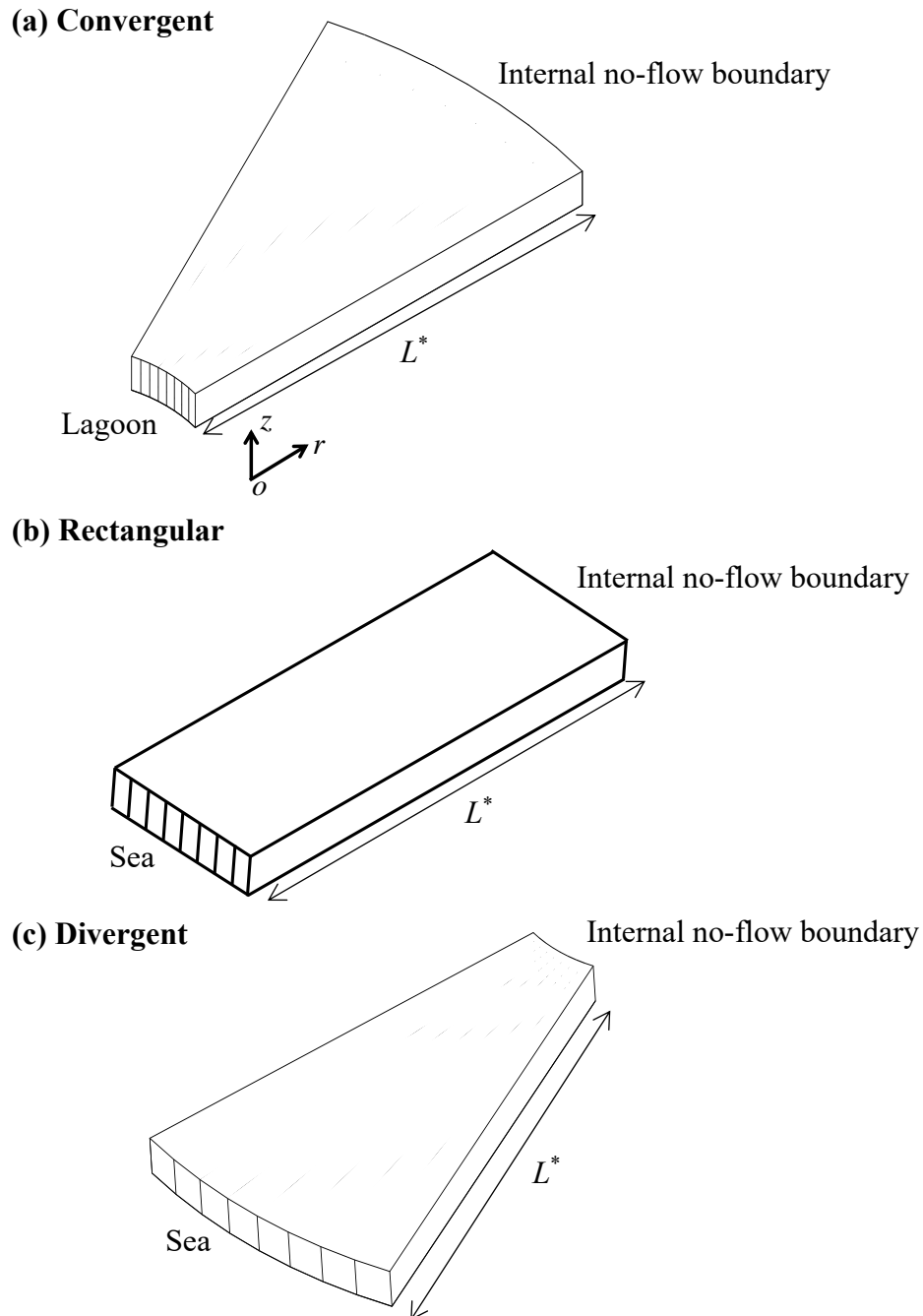
607 Note that the left and right sides are the sea and internal no-flow boundaries, respectively.



608

609

Figure 4. Widths of Unit 1 and Unit 2 versus L_0 for aquifers with different total width L .

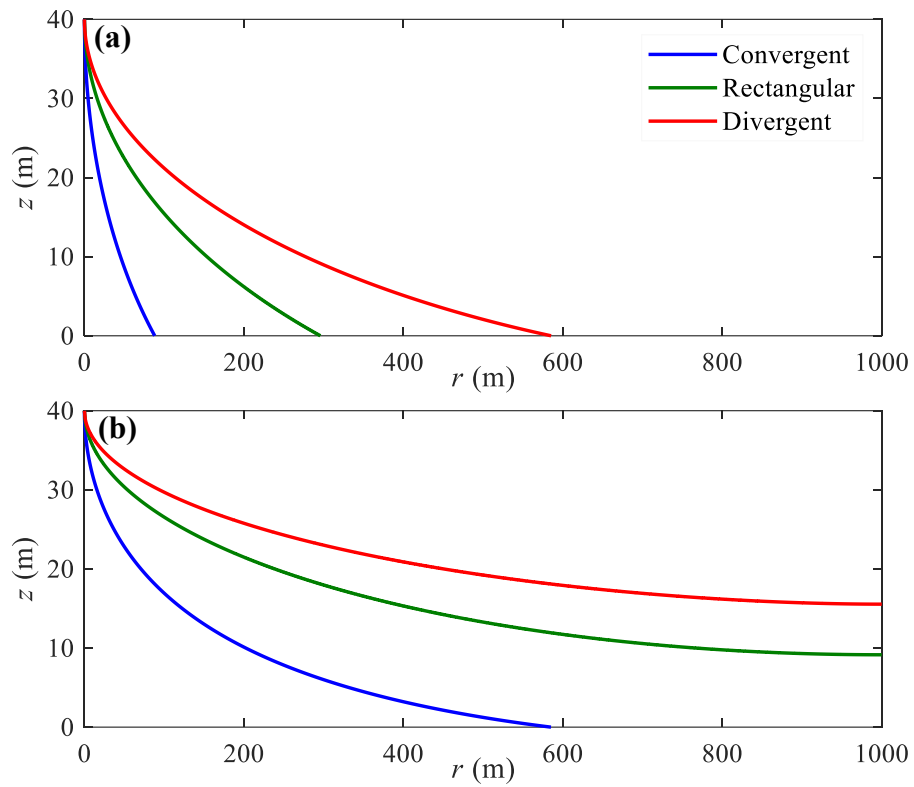


610

611

612

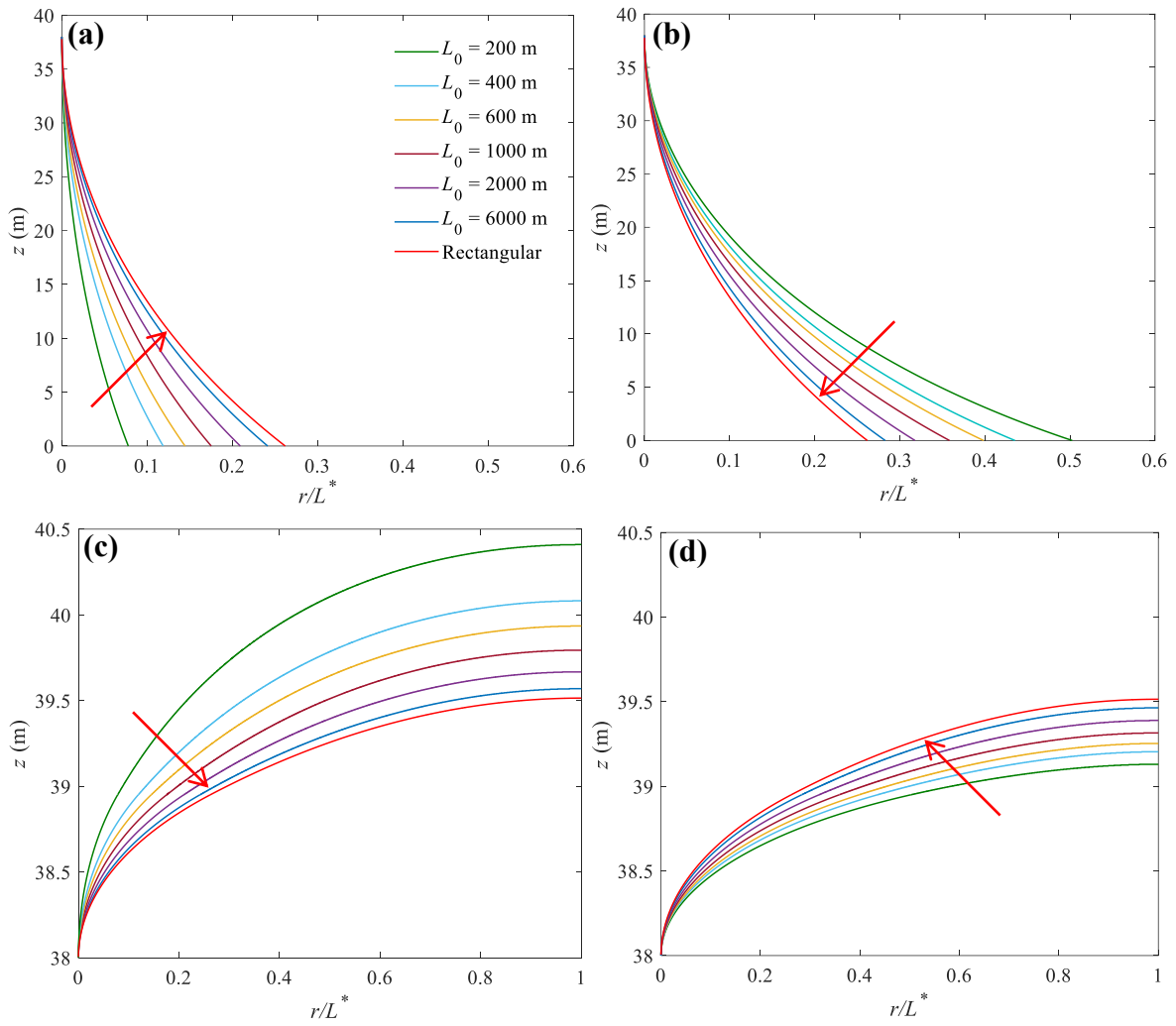
613 **Figure 5.** Three-dimensional view of (a) convergent (smaller side facing the lagoon), (b)
 614 rectangular and (c) divergent aquifers (larger side facing the sea) compared in this study. L^*
 615 represents the distance from the sea/lagoon to the internal no-flow boundary, i.e., l_1 or l_2 in
 616 Figure 2. The internal no-flow boundary corresponds to the z -axis in Figure 2.



617

618

619 **Figure 6.** Freshwater-seawater interface predicted by analytical solutions for three different
 620 aquifers with (a) high and (b) low recharge (Cases 1 and 2 in Table 1). Note that $r = 1000$ m is
 621 the internal no-flow boundary in Figure 5.



622

623

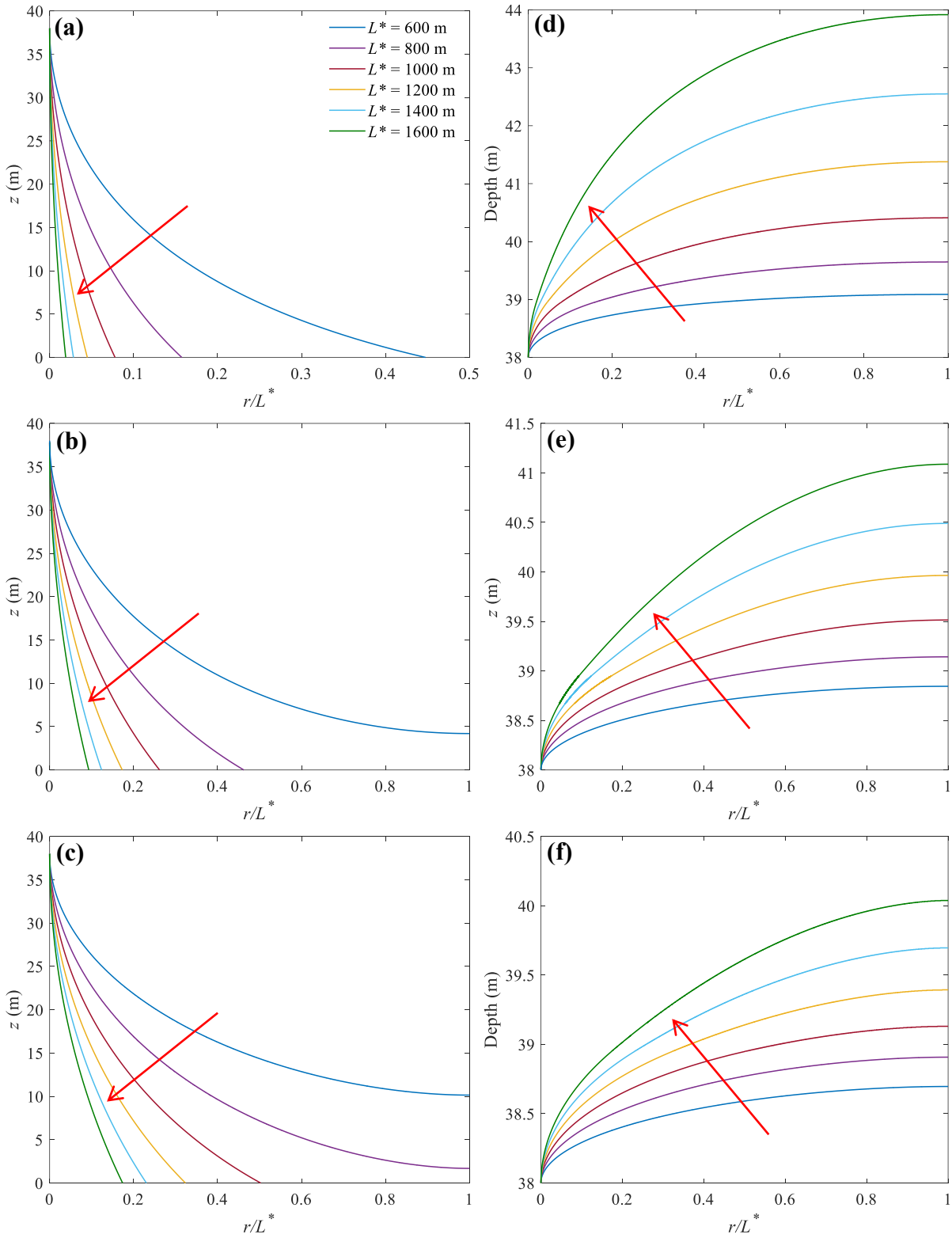
624

625

626

627

Figure 7. Sensitivity of (a, b) the locations of the freshwater-seawater interface and (c, d) watertable to L_0 for convergent (left panel) and divergent (right panel) aquifers. The arrow in each plot shows the direction of increasing L_0 (values given in (a), used to produce the different curves). Note that predictions for rectangular aquifers are independent of L_0 .



628

629

630

631 **Figure 8.** Sensitivity of (a-c) the locations of the freshwater-seawater interface and (d-f)

632 watertable to L^* for convergent (a, d), rectangular (b, e) and divergent (c, f) aquifers. The

633 arrow in each plot points to the increase of L^* values used to construct each curve (values

634 indicated in (a)).