

1                   Effects of aquifer geometry on seawater intrusion in annulus  
2   segment island aquifers

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21                   Resubmitted to *Hydrology and Earth System Sciences* on 27 August 2021

## 22 **Abstract**

23        Seawater intrusion in island aquifers was considered analytically, specifically for annulus  
24 segment aquifers (ASAs), i.e., aquifers that (in plan) have the shape of an annulus segment.  
25        Based on the Ghijben-Herzberg and hillslope-storage Boussinesq equations, analytical  
26 solutions were derived for steady-state seawater intrusion in ASAs, with a focus on the  
27 freshwater-seawater interface and its corresponding watertable elevation. Predictions of the  
28 analytical solutions compared well with experimental data, and so they were employed to  
29 investigate the effects of aquifer geometry on seawater intrusion in island aquifers. Three  
30 different ASA geometries were compared: convergent (smaller side facing the lagoon),  
31 rectangular and divergent (larger side facing the sea). Depending on the aquifer geometry,  
32 seawater intrusion was found to vary greatly, such that the assumption of a rectangular aquifer  
33 to model an ASA can lead to poor estimates of seawater intrusion. Other factors being equal,  
34 compared with rectangular aquifers, seawater intrusion is more extensive and watertable  
35 elevation is lower in divergent aquifers, with the opposite tendency in convergent aquifers.  
36        Sensitivity analysis further indicated that the effects of aquifer geometry on seawater intrusion  
37 and watertable elevation vary with aquifer width and distance from the circle center to the  
38 inner arc (the lagoon boundary for convergent aquifers while the internal no-flow boundary  
39 for divergent aquifers). A larger aquifer width and distance from the circle center to the inner  
40 arc weaken the effects of aquifer geometry and hence differences in predictions for the three  
41 geometries become less pronounced.

42 **Keywords:** sharp-interface; steady-state analytical solution; atoll aquifer; annulus segment  
43 aquifer, seawater intrusion

44 **Key Points**

- 45 ➤ Analytical solutions of steady-state seawater intrusion were derived for annulus segment  
46 aquifers
- 47 ➤ Among three different aquifer geometries, divergent aquifers have the lowest watertable  
48 and hence the most extensive seawater intrusion
- 49 ➤ Aquifer geometry effects on seawater intrusion depend on the aquifer width and distance  
50 from the circle center to the inner arc

## 51 **1. Introduction**

52 Islands are extensively distributed throughout the world's oceans. Unfortunately, their  
53 groundwater resources are impacted by sea-level rise and increased demands. According to a  
54 recent estimate, there are approximately 65 million people living in oceanic islands where  
55 groundwater may be the only source of freshwater (Thomas et al., 2020). Fresh groundwater  
56 stored on oceanic islands is mainly from precipitation (usually in the form of a freshwater  
57 lens) and its availability varies due to different factors, e.g., island topography, rainfall  
58 patterns, tides, episodic storms and human activities (White & Falkland, 2010; Storlazzi et al.,  
59 2018). Seawater intrusion is thus an important issue due to its deleterious effect on oceanic  
60 island freshwater storage (e.g., Werner et al., 2017; Lu et al., 2019; Memari et al., 2020).

61 In contrast to coastal aquifers where seawater intrudes into freshwater from one direction  
62 only, seawater intrusion occurs from two directions for narrow strip islands and from all  
63 directions for circular islands. Over the past few decades, seawater intrusion in oceanic  
64 islands has been extensively investigated in field observations (e.g., Röper et al., 2013; Post et  
65 al., 2019), laboratory experiments (e.g., Stoeckl et al., 2015; Bedekar et al., 2019; Memari et  
66 al., 2020), numerical simulations (e.g., Lam, 1974; Gingerich et al., 2017; Liu & Tokunaga,  
67 2019) and analytical solutions (e.g., Fetter, 1972; Ketabchi et al., 2014; Lu et al., 2019).

68 Among these, analytical solutions are effective tools to assess the extent of seawater intrusion  
69 (i.e., the location of the freshwater-seawater interface), although they cannot incorporate  
70 complex factors (e.g., dispersive mixing and transient oceanic dynamics) (Werner et al.,  
71 2013). The advantages of analytical solutions are that they are computationally efficient, can

72 be used as test cases for numerical models, and can reveal the explicit relationships between  
73 parameters that influence seawater intrusion (e.g., Fetter, 1972; Ketabchi et al., 2014; Liu et  
74 al., 2014; Lu et al., 2019;).

75 Based on the Dupuit-Forchheimer approximation (i.e., ignoring vertical flow) and the  
76 Ghijben-Herzberg equation (Drabbe & Badon Ghijben, 1889, English translation given by  
77 Post (2018); Herzberg, 1901), Fetter (1972) presented analytical solutions describing the  
78 freshwater-seawater interface location and watertable elevation in a circular island. Bailey et  
79 al. (2010) further compared these single-layered analytical solutions with field measurements,  
80 indicating that the analytical solutions perform well in estimating the freshwater-seawater  
81 interface location and watertable elevation. Fetter's solutions formed the foundation for many  
82 subsequent analytical studies on seawater intrusion in island aquifers. Again, for a single  
83 layer, Chesnaux and Allen (2008) and Greskowiak et al. (2013) developed analytical solutions  
84 to predict the steady-state groundwater age distribution in freshwater lenses. In addition, using  
85 single-layered analytical solutions, Morgan and Werner (2014) proposed vulnerability  
86 indicators of freshwater lenses under sea-level rise and recharge change.

87 Since aquifers are usually heterogeneous, the single-layer analytical solutions were  
88 subsequently extended to two-layered island aquifers. Vacher (1988) derived solutions for the  
89 freshwater-seawater interface location and watertable elevation for infinite-strip islands  
90 composed of different layers. Dose et al. (2014) conducted laboratory experiments to validate  
91 and confirm the reliability of analytical solutions proposed by Fetter (1972) and Vacher  
92 (1988). Ketabchi et al. (2014) extended Fetter's analytical solutions to calculate the

93 freshwater-seawater interface location and watertable elevation in two-layered circular islands  
94 subject to sea-level rise. Their results indicated that land-surface inundation caused by sea-  
95 level rise has a considerable impact on fresh groundwater lenses. Recently, Lu et al. (2019)  
96 derived analytical solutions for the freshwater-seawater interface location and watertable  
97 elevation for both strip and circular islands with two adjacent layers, i.e., a less permeable  
98 slice along the shoreline of an island, and a more permeable zone inland.

99 All the abovementioned analytical solutions apply to either strip or circular islands.  
100 According to the classification of sand dunes developed by Stuyfzand (1993; 2017), there are  
101 different island layouts that should be considered, e.g., where the shape of the island is an  
102 annulus segment, instead of a strip or circular disk (Figure 1). Annulus segment-shaped  
103 islands are found in various atolls (i.e., circular chains of islands surrounding a central  
104 lagoon) as found in the Pacific and Indian Oceans (Werner et al., 2017; Duvat, 2019).  
105 Nevertheless, analytical solutions of seawater intrusion are not yet available for annulus  
106 segment aquifers (ASAs). In general, ASAs are conceptually treated as a 2D cross section,  
107 similar to strip islands (e.g., Ayers & Vacher, 1986; Underwood et al., 1992; Bailey et al.,  
108 2009; Werner et al., 2017). Evidently, topography plays an important role in groundwater flow  
109 and hence seawater intrusion (e.g., Zhang et al., 2016; Liu & Tokunaga, 2019). It remains  
110 unclear whether analytical solutions of seawater intrusion for strip islands are appropriate for  
111 ASAs. It is moreover additionally unclear how island geometry affects the freshwater-  
112 seawater interface location and watertable elevation of ASAs.

113 In this study, analytical solutions are derived for steady-state seawater intrusion for ASAs,

114 with a focus on the freshwater-seawater interface location and its corresponding watertable  
115 elevation. After comparing their predictions with experimental data (Memari et al., 2020), the  
116 analytical solutions are employed to investigate the effects of aquifer geometry on the  
117 freshwater-seawater interface location and watertable elevation in ASAs.

## 118 **2. Conceptual Model**

119 Figure 2 shows the conceptual model of an ASA (a slice of an atoll island). The plan  
120 view of the model domain is represented as a sector ( $EFGH$ ) with an angle  $\theta$  (Figure 2a).  
121 Radial flow only is considered. The sea ( $EF$ ) and lagoon ( $HG$ ) boundaries are located at  $L +$   
122  $L_0$  [L] and  $L_0$  [L] from the circle center, respectively. Since the longitudinal length is usually  
123 much longer than the lateral length for an atoll island (Werner et al., 2017), seawater intrusion  
124 from the lateral sides ( $EH$  and  $FG$ , Figure 2a) is negligible in comparison to the longitudinal  
125 side, especially for the middle portion of an ASA. Therefore,  $EH$  and  $FG$  are treated as lateral  
126 no-flow boundaries. Note that treating the lateral sides as no-flow boundaries is often used in  
127 studies of freshwater lenses on atoll islands (e.g., Ayers & Vacher, 1986; Underwood et al.,  
128 1992; Bailey et al., 2009; Werner et al., 2017). The side view of the model domain is  
129 conceptualized as a rectangle ( $ABCD$ ) along the radial direction with dimensions of  $L$  [L]  
130 (width)  $\times d$  [L] (height) (Figure 2b, c).  $AD$  is the impermeable base while  $BC$  is the land  
131 surface through which aquifer recharge flows.

132 Both the sea and lagoon water levels are set to  $H_s$  [L], which results in an internal no-  
133 flow boundary (water divide, where the slope of the watertable is zero) between the sea and  
134 lagoon (location of the  $z$ -axis in Figure 2b,c). The segment between the sea and the internal

135 no-flow boundary is referred to as Unit 1, whereas the segment between the internal no-flow  
136 and lagoon boundaries is referred to as Unit 2 (Figure 2). The widths of Units 1 and 2 are  $l_1$   
137 [L] and  $l_2$  [L], respectively. In addition, the flow is asymmetrical in Units 1 and 2, with  
138 divergent flow (the aquifer width  $w$  [L] increases along the flow direction) in Unit 1 and  
139 convergent flow ( $w$  decreases along the flow direction) in Unit 2.

140 The  $x$ - $z$  coordinate origin is placed at the intersection of the internal no-flow boundary  
141 and impermeable base, with the  $x$ -axis pointing to the circle center and the  $z$ -axis pointing  
142 vertically upward. Further,  $\phi$  [L] is the watertable height,  $h$  [L] is the vertical distance  
143 between the watertable and the interface,  $h_s$  [L] is the vertical distance between the sea level  
144 and the interface, and  $h_c = H_s - h_s$  [L] is the vertical distance from the impermeable base to  
145 the interface for given  $x$  (Figure 2b,c). Constant recharge into the saturated zone,  $N$  [ $LT^{-1}$ ], is  
146 assumed. There are two possibilities for the interface tip (i.e., the location where the  
147 freshwater-seawater interface connects to the  $z$ -axis or the bottom boundary): above the  
148 aquifer bed (Figure 2b) or on the aquifer bed (Figure 2c). The  $x$ -coordinates of the interface  
149 tip in Units 1 and 2 are denoted as  $x_{t1}$  [L] and  $x_{t2}$  [L], respectively (Figure 2c). Note that  $x_{t1} =$   
150  $x_{t2} = 0$  when the interface tip is above the aquifer bed, as in Figure 2b.

151 Consistent with previous studies (e.g., Ketabchi et al., 2014; Lu et al., 2016; 2019), the  
152 following assumptions are made: (1) steady-state flow, (2) sharp freshwater-seawater  
153 interface, (3) homogeneous and isotropic aquifer, (4) negligible unsaturated flow, (5) recharge  
154 is less than the saturated hydraulic conductivity (else overland flow which will appear



155 following ponding occurs), and (6) vertical flow in the saturated zone is negligible (Dupuit-  
 156 Forchheimer approximation).

### 157 3. Analytical Solutions

158 Groundwater flow in an ASA (Figure 2) can be described as (Paniconi et al., 2003; Troch  
 159 et al., 2003),

$$160 \quad -\frac{\partial}{\partial x}(wq) + Nw = \frac{\partial S}{\partial t} \quad (1)$$

161 where  $q$  [ $L^2T^{-1}$ ] is the Darcy flux per unit length along the aquifer;  $S$  [ $L^2$ ] is the total  
 162 water storage per unit distance along the aquifer, and  $t$  [T] is time. Equation (1) is derived  
 163 from the hillslope-storage Boussinesq equation reformulated in terms of soil water storage  
 164 rather than watertable elevation, as widely used previously (e.g., Stagnitti et al., 1986; Troch  
 165 et al., 2003; Hilberts et al., 2005; Kong et al., 2016; Luo et al., 2018). At steady state, equation  
 166 (1) reduces to,

$$167 \quad -\frac{\partial}{\partial x}(wq) + Nw = 0 \quad (2)$$

168 According to Darcy's law and the Dupuit-Forchheimer approximation, the freshwater  
 169 flux in the aquifer segment between the seaward boundary and interface tip can be calculated  
 170 as ( $\phi$  is independent of  $z$ ),

$$171 \quad q = -\int_{h_c}^{\phi} K_s \frac{d\phi}{dz} dz = -K_s (\phi - h_c) \frac{d\phi}{dx} \quad (3)$$

172 where  $K_s$  [ $LT^{-1}$ ] is the saturated hydraulic conductivity.

#### 173 3.1. Interface Tip above the Aquifer Bed

174 We first consider the situation where the interface tip is above the aquifer bed (Figure  
 175 2b). In Unit 1 where  $w = \theta(L_0 + l_2 - x)$ , substituting equation (3) into equation (2) and then

176 integrating gives (based on the Dupuit-Forchheimer approximation),

$$177 \quad -\frac{1}{2} \left[ (L_0 + l_2 - x)^2 - (L_0 + l_2)^2 \right] N = -(L_0 + l_2 - x) K_s (\phi - h_c) \frac{d\phi}{dx} \quad (4)$$

178 According to the Ghijben-Herzberg equation, the vertical thickness of the freshwater zone ( $h$ )

179 in the interface zone is given by,

$$180 \quad h = \phi - h_c = (1 + \alpha)(\phi - H_s) \quad (5)$$

181 where  $\alpha = \rho_f / (\rho_s - \rho_f)$  is the dimensionless density difference, and  $\rho_f$  [ $\text{ML}^{-3}$ ] and  $\rho_s$

182 [ $\text{ML}^{-3}$ ] are the freshwater and seawater densities, respectively. Substitution of equation (5)

183 into equation (4) yields,

$$184 \quad -\frac{1}{2} \left[ (L_0 + l_2 - x)^2 - (L_0 + l_2)^2 \right] N = -K_s (L_0 + l_2 - x)(1 + \alpha)(\phi - H_s) \frac{d\phi}{dx} \quad (6)$$

185 Rearranging equation (6) produces,

$$186 \quad -\frac{(L_0 + l_2 - x)N}{2} + \frac{N(L_0 + l_2)^2}{2(L_0 + l_2 - x)} = -K_s (1 + \alpha)(\phi - H_s) \frac{d\phi}{dx} \quad (7)$$

187 Integrating equation (7) leads to,

$$188 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - x) - \frac{1}{2}(L_0 + l_2)Nx + \frac{1}{4}Nx^2 + C_1 = -K_s (1 + \alpha) \frac{(\phi - H_s)^2}{2} \quad (8)$$

189 where  $C_1$  is the integration constant that is determined by the sea boundary condition (i.e.,

190  $x = -l_1, \phi = H_s$ ),

$$191 \quad C_1 = \frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 + l_1) - \frac{1}{2}(L_0 + l_2)l_1N - \frac{1}{4}l_1^2N \quad (9)$$

192 The relation between  $h_s$  and  $\phi$  is given by,

$$193 \quad h_s = \alpha(\phi - H_s) \quad (10)$$

194 Combining equation (8) with equation (10) and eliminating  $\phi$  yields,

$$-\frac{(L_0+l_2)^2 N}{2} \ln(L_0+l_2-x) - \frac{1}{2}(L_0+l_2)Nx + \frac{1}{4}Nx^2 + C_1 = -K_s(1+\alpha)\frac{h_s^2}{2\alpha^2} \quad (11)$$

Equation (11) gives the freshwater-seawater interface location in Unit 1 once  $l_1$  and  $l_2$  are determined.

Equation (8) applies to Unit 2 by replacing  $C_1$  with  $C_2$ ,

$$-\frac{(L_0+l_2)^2 N}{2} \ln(L_0+l_2-x) - \frac{1}{2}(L_0+l_2)Nx + \frac{1}{4}Nx^2 + C_2 = -K_s(1+\alpha)\frac{(\phi-H_s)^2}{2} \quad (12)$$

where  $C_2$  is chosen to satisfy the lagoon boundary condition ( $x=l_2$ ,  $\phi=H_s$ ),

$$C_2 = \frac{(L_0+l_2)^2 N}{2} \ln(L_0) + \frac{1}{2}(L_0+l_2)l_2N - \frac{1}{4}l_2^2N \quad (13)$$

Combining equations (10) and (12) and eliminating  $\phi$  generates,

$$-\frac{(L_0+l_2)^2 N}{2} \ln(L_0+l_2-x) - \frac{1}{2}(L_0+l_2)Nx + \frac{1}{4}Nx^2 + C_2 = -K_s(1+\alpha)\frac{h_s^2}{2\alpha^2} \quad (14)$$

Equation (14) gives the freshwater-seawater interface location in Unit 2 once  $l_2$  is determined. Since the sea level and lagoon water level are the same, an internal no-flow boundary exists between the sea and lagoon, i.e.,

$$x=0, \quad (h_s)_{unit1} = (h_s)_{unit2} \quad (15)$$

where  $(h_s)_{unit1}$  and  $(h_s)_{unit2}$  represent  $h_s$  in Units 1 and 2, respectively.

Combining equations (11), (14) and (15) leads to expressions for  $l_1$  and  $l_2$ ,

$$l_1 = L + L_0 - \sqrt{\frac{2LL_0 + L^2}{2\ln(L+L_0) - 2\ln(L_0)}} \quad (16)$$

$$l_2 = \sqrt{\frac{2LL_0 + L^2}{2\ln(L+L_0) - 2\ln(L_0)}} - L_0 \quad (17)$$

As indicated by equations (16) and (17), the internal no-flow boundary between the sea and

213 lagoon only depends on  $L$  and  $L_0$ . For known  $l_1$  and  $l_2$ , equations (11) and (14) can be  
 214 employed to predict the freshwater-seawater interface location in Units 1 and 2, respectively.

215 Once the interface location is determined,  $h$  and  $\phi$  are given by,

$$216 \quad h = \frac{1+\alpha}{\alpha} h_s \quad (18)$$

$$217 \quad \phi = \frac{h_s}{\alpha} + H_s \quad (19)$$

### 218 **3.2. Interface Tip on the Aquifer Bed**

219 When the interface tip is on the aquifer bed, the location of the internal no-flow  
 220 boundary remains the same as for the interface tip above the aquifer bed. The freshwater-  
 221 seawater interface for Units 1 and 2 can be determined by equations (11) and (14),  
 222 respectively. Then, from equation (18),  $h$  at the aquifer segment between the sea boundary and  
 223 the interface tip is determined. To calculate  $h$  for the aquifer segment between the interface tip  
 224 and the internal no-flow boundary, the  $x$ -coordinate of the interface tip is found. At the  
 225 interface tip of Unit 1 ( $x = x_{t1}$ ),

$$226 \quad h_s = H_s \quad (20)$$

$$227 \quad \phi = \frac{1+\alpha}{\alpha} H_s \quad (21)$$

228 With equations (11) and (21),  $x_{t1}$  is given by,

$$229 \quad -\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - x_{t1}) - \frac{1}{2}(L_0 + l_2) N x_{t1} + \frac{1}{4} N x_{t1}^2 = -C_1 - K_s (1 + \alpha) \frac{H_s^2}{2\alpha^2} \quad (22)$$

230 Let,

$$231 \quad a = \frac{1}{4} N \quad (23a)$$

$$232 \quad b = -\frac{1}{2}(L_0 + l_2) N \quad (23b)$$

233 
$$c = -\frac{(L_0 + l_2)^2 N}{2} \quad (23c)$$

234 and

235 
$$m = -C_1 - K_s (1 + \alpha) \frac{H_s^2}{2\alpha^2} \quad (23d)$$

236 then equation (22) becomes,

237 
$$ax_{i1}^2 + bx_{i1} + c \ln(L_0 + l_2 - x_{i1}) = m \quad (24)$$

238 which is solved by a root-finding method.

239 The freshwater discharge for the aquifer segment between the interface tip and the  
240 internal no-flow boundary is calculated as,

241 
$$-\frac{1}{2} \left[ (L_0 + l_2 - x)^2 - (L_0 + l_2)^2 \right] N = -(L_0 + l_2 - x) K_s \phi \frac{d\phi}{dx} \quad (25)$$

242 Repeating the steps from equations (4) to (8) gives,

243 
$$-\frac{(L_0 + l_2)^2 N}{2} \ln(L_0 + l_2 - x) - \frac{1}{2} (L_0 + l_2) Nx + \frac{1}{4} Nx^2 + C_3 = -\frac{K_s}{2} \phi^2 \quad (26)$$

244 where  $C_3$  is determined by substituting equation (21) into equation (26). Then, equation (26)  
245 can be adopted to calculate  $h$  for the segment between the interface tip and the internal no-  
246 flow boundary where  $h = \phi$ .

247 Similarly, the  $x$ -coordinate of the interface tip in Unit 2 ( $x_{i2}$ ) is obtained by substituting  
248 equation (20) into equation (14). Then, the watertable ( $h$ ) of the aquifer segment between the  
249 interface tip and the internal no-flow boundary for Unit 2 is computed by repeating the steps  
250 from equations (22) to (26).

## 4. Results and Discussion

### 4.1. Validation of the Analytical Solutions

The analytical solutions were validated by comparing their predictions with experimental data compiled from Memari et al. (2020), who reported experiments carried out using a 15° radial tank. The tank contained three distinct chambers: internal no-flow boundary condition, porous medium and constant-head boundary condition (i.e., sea or lagoon). The internal no-flow and seaward boundaries were respectively located at 10 cm and 55.5 cm from the circle center, i.e., 45.5 cm from the internal no-flow boundary to the constant-head boundary along the radial direction. Note that the experimental tank only corresponds to Unit 1 of the radial aquifer with  $l_1 = 45.5$  cm and  $l_2 = 0$ , so the analytical results were calculated using equations (11) and (24). The thicknesses of the porous medium and sea level were 28 and 25 cm, respectively. The sand used in experiments had a saturated hydraulic conductivity of  $1.23 \times 10^{-2}$  m s<sup>-1</sup> and an effective porosity of 0.40. The measured saltwater and freshwater densities were respectively 1.015 and 0.999 g ml<sup>-1</sup>, leading to  $\alpha = 62$ . Two different recharge events with constant  $N$ ,  $2.46 \times 10^{-4}$  and  $1.08 \times 10^{-4}$  m s<sup>-1</sup>, were considered in the experiments.

Figure 3 shows the comparison between analytical and experimental results of the freshwater-seawater interface for different recharge events. In general, the analytical solution predicts the freshwater-seawater interface well for both recharge events, despite there being some differences between the analytical results and the measurements, particularly in the zone near the constant-head boundary ( $x = -45$  cm). These deviations are likely due to assumptions made in the analytical solution, i.e., (i) a sharp freshwater-seawater interface, (ii) ignoring the

272 effect of freshwater discharge, and (iii) neglecting the vertical flow (the Dupuit-Forchheimer  
273 approximation).

#### 274 **4.2. Effects of Aquifer Geometry on Seawater Intrusion**

275 Previous studies showed that boundary conditions play a critical role in estimates of  
276 seawater intrusion (Werner & Simmons, 2009; Lu et al., 2016). Therefore, the internal no-  
277 flow boundary between the sea and lagoon was examined for various ASAs. As indicated by  
278 equations (16) and (17), this internal no-flow boundary depends only on  $L$  and  $L_0$ . The values  
279 of  $l_1$  and  $l_2$  calculated respectively from equations (16) and (17) are shown in Figure 4 for  
280 three typical values of  $L$  (500, 1000 and 2000 m) with  $L_0$  varying from  $10^2$  to  $10^6$  m. In  
281 general, the internal no-flow boundary deviates from the middle of the ASA. When  $L_0$  is less  
282 than  $10^5$  m,  $l_1$  is larger than  $l_2$  for the three different values of  $L$ , indicating an internal no-  
283 flow boundary closer to the lagoon boundary. For example, taking  $L = 2000$  m and  $L_0 = 100$  m  
284 leads to  $l_1 = 1240$  m and  $l_2 = 760$  m, with a deviation of 240 m (12% of 2000 m) from the  
285 middle of the ASA. When  $L_0$  exceeds  $10^5$  m, however, the location of the internal no-flow  
286 boundary can be approximated as being at the middle of the ASA for all considered values of  
287  $L$ . This is in contrast to strip and circular aquifers where the internal no-flow boundary is  
288 always in the middle of aquifers due to symmetry.

289 Since the internal no-flow boundary location between the sea and lagoon deviates from  
290 the middle of the ASA, we expect aquifer geometry to play a significant role in controlling  
291 seawater intrusion. As mentioned previously, ASAs can be convergent (Unit 1) or divergent  
292 aquifers (Unit 2) where the extent of seawater intrusion may be different. However, for strip

293 aquifers, both Units 1 and 2 are rectangular with the same extent of seawater intrusion.  
294 Therefore, three geometries were compared in this study: convergent, rectangular and  
295 divergent (Figure 5). These geometries have been widely examined in hillslope hydrology  
296 regrading to the effects of aquifer geometry on runoff generation (Troch et al., 2003; Kong et  
297 al., 2016; Luo et al., 2018). To present the results more conveniently, we replaced the  $x$ - $z$   
298 coordinate origin at the intersection of the constant-head boundary (sea or lagoon) and the  
299 impermeable base, with the  $x$ -axis pointing horizontally to the internal no-flow boundary and  
300 the  $z$ -axis vertically upward (Figure 5). In addition, the distance between the constant-head  
301 boundary and the internal no-flow boundary (aquifer width) is denoted as  $L^*$  (Figure 5) while  
302 the other parameters remain the same.

303 Following previous studies (e.g., Lu et al., 2016; 2019), different cases were selected to  
304 show the effects of aquifer geometry on seawater intrusion (Cases 1 and 2 in Table 1).  
305 According to Werner et al. (2017), the width of atoll islands generally varies from 100 to 1500  
306 m along the radial direction. In order to focus on the effects of aquifer geometry on seawater  
307 intrusion, the same  $L^*$  and  $L_0$  were assumed for the three aquifers, with  $L^*$  and  $L_0$  equal to  
308 1000 and 200 m, respectively. Note that  $L_0$  is the distance from the circle center to the lagoon  
309 boundary for convergent aquifers, whereas it represents the distance from the circle center to  
310 internal no-flow boundary for divergent aquifers hereafter. The sand characteristics were the  
311 same as in the experiments of Memari et al. (2020). Two recharge events were considered  
312 (Cases 1 and 2, Table 1). The freshwater-seawater interface was calculated using the  
313 analytical solutions for the three different aquifers. Note that the Appendix presents analytical



314 solutions for seawater intrusion in strip aquifers deduced from Lu et al. (2019).

315 Figure 6 shows the freshwater-seawater interface calculated for Cases 1 and 2. As can be  
316 seen, the extent of seawater intrusion is noticeably different for the three aquifer geometries.  
317 For high recharge ( $1 \times 10^{-6} \text{ m s}^{-1}$ ), the interface tip is located at around 500 m for the  
318 divergent aquifer, which is about twice the value of the rectangular aquifer and six times the  
319 value for the convergent aquifer (Figure 6a). When the recharge decreases to  $3 \times 10^{-7} \text{ m s}^{-1}$ ,  
320 the interface tip moves further landward for the three aquifers as expected, but the difference  
321 between results is still great (Figure 6b). The interface tip is displaced above the aquifer bed  
322 for both the rectangular and divergent aquifers, while it remains on the aquifer bed for the  
323 convergent aquifer. Regardless of the recharge rate, the most landward freshwater-seawater  
324 interface occurs in the divergent aquifer and vice versa for convergent aquifer. This underlines  
325 that aquifer geometry plays a major role in controlling seawater intrusion and hence it is  
326 necessary to account for aquifer geometry in analyses of seawater intrusion.

### 327 **4.3. Sensitivity Analysis**

328 A sensitivity analysis was conducted to investigate to what extent aquifer geometry  
329 affects seawater intrusion. Since we focus on the effects of aquifer geometry on the locations  
330 of the freshwater-seawater interface and watertable, values of  $L_0$  and  $L^*$  were varied, with  
331 other parameters kept constant. When conducting the sensitivity analysis of  $L_0$ ,  $L^*$  was fixed  
332 at 1000 m, which is a typical value for ASAs (Werner et al., 2017). Figure 7 shows the  
333 sensitivity of the locations of the freshwater-seawater interface and watertable to changes in  
334  $L_0$  (Case 3, Table 1). The freshwater-seawater interface and watertable elevation are

335 independent of  $L_0$  for rectangular aquifers (Appendix). However, the freshwater-seawater  
336 interface and watertable elevation differ greatly when varying  $L_0$  for both convergent and  
337 divergent aquifers, highlighting that  $L_0$  plays an important role in affecting seawater intrusion.  
338 Specifically, as  $L_0$  increases, the freshwater-seawater interface moves more landward (larger  
339  $x/L^*$ , Figure 7a) and its corresponding watertable elevation decreases (Figure 7c) for  
340 convergent aquifers. In contrast, for divergent aquifers increasing  $L_0$  moves the freshwater-  
341 seawater interface more seaward (smaller  $x/L^*$ , Figure 7b) and its corresponding watertable  
342 elevation increases (Figure 7d). For a given  $L_0$ , divergent aquifers have the largest extent of  
343 seawater intrusion and the lowest watertable elevation, and conversely for convergent aquifers  
344 (Figure 7).

345       Regardless of the freshwater-seawater interface and watertable elevation, the deviation  
346 between rectangular aquifers and divergent or convergent aquifers is significant when  $L_0$  is  
347 less than 2000 m (Figure 7). For example, the  $x$ -coordinate of the interface tip ( $z = 0$ ) is 262 m  
348 for the rectangular aquifer at  $L_0 = 200$  m, whereas it is 78 (31% of that in the rectangular  
349 aquifer) and 500 m (191% of that in the rectangular aquifer) for the convergent and divergent  
350 aquifers, respectively. As  $L_0$  increases, the deviation between the three aquifers decreases.  
351 When  $L_0 = 2000$  m, the  $x$ -coordinate of the interface tip is 262, 209 (80% of that in the  
352 rectangular aquifer) and 318 m (121% of that in the rectangular aquifer) for the rectangular,  
353 convergent and divergent aquifers, respectively. As  $L_0$  increases to 6000 m, the freshwater-  
354 seawater interface and watertable elevation of both convergent and divergent aquifers tend to  
355 those of rectangular aquifers, i.e., geometry effects decrease with increasing  $L_0$ . These results

356 highlight the critical role played by the shape of aquifers. As a result, ignoring geometry  
357 effects may lead to an inappropriate management strategy for groundwater resources in atoll  
358 islands.

359 The sensitivity of the freshwater-seawater interface and watertable elevation to  $L^*$  was  
360 investigated by varying  $L^*$  from 600 to 1600 m while fixing  $L_0$  to 200 m (Case 4, Table 1). As  
361 shown in Figure 8, contrary to the results for varying  $L_0$ , in this case the freshwater-seawater  
362 interface and watertable elevation in all three topographies are related to  $L^*$ . Again, the extent  
363 of seawater intrusion is greatest in divergent aquifers and least in convergent aquifers for  
364 given  $L^*$ . When  $L^*$  increases, the freshwater-seawater interface moves seaward and the  
365 watertable elevation increases, regardless of aquifer geometry, i.e., the seawater intrusion  
366 decreases (Figures 8a-c). This is because the total freshwater flux increases with increasing  
367  $L^*$ , leading to a higher hydraulic gradient and hence less seawater intrusion (Figures 8d-f).  
368 Moreover, an increase in  $L^*$  reduces the differences in the seawater intrusion distance among  
369 the three geometries, i.e., geometry effects on seawater intrusion are more significant at small  
370  $L^*$ . However, even at the maximum  $L^*$  considered (1600 m), the deviation between three  
371 aquifers remains significant: The  $x$ -coordinate of the interface tip is about 148 m for the  
372 rectangular aquifer, whereas it is about 32 (22% of that in the rectangular aquifer) and 278 m  
373 (188% of that in the rectangular aquifer) for the convergent and divergent aquifers,  
374 respectively. Both  $L_0$  and  $L^*$  can greatly impact seawater intrusion estimates for divergent and  
375 convergent aquifers, highlighting the necessity to include geometry effects in analytical  
376 solutions of seawater intrusion.

## 377 **5. Conclusions**

378 Based on the Ghijben-Herzberg and hillslope-storage Boussinesq equations, we derived  
379 analytical solutions of steady-state seawater intrusion for ASAs, with a focus on the  
380 freshwater-seawater interface and its corresponding watertable elevation as affected by  
381 recharge. After comparing with experimental data of Memari et al. (2020), the analytical  
382 solutions were employed to examine the effects of aquifer geometry on seawater intrusion in  
383 island aquifers. Three different shapes of island aquifer were compared: convergent,  
384 rectangular and divergent. The results lead to the following conclusions:

385 (1) The presented analytical solutions perform well in predicting the experimental  
386 freshwater-seawater interface, suggesting that these analytical solutions can predict seawater  
387 intrusion reasonably in different aquifer geometries.

388 (2) Island geometry plays a significant role in affecting the freshwater-seawater interface  
389 and watertable elevation. Other factors being equal, the extent of seawater intrusion is greatest  
390 in divergent aquifers, and conversely least in convergent aquifers. In contrast, the watertable  
391 elevation is lowest in divergent aquifers and highest in convergent aquifers.

392 (3) The effects of aquifer geometry on seawater intrusion are dependent on the aquifer  
393 width and distance from the circle center to the internal no-flow boundary (Figures 7 and 8).  
394 A larger aquifer width and distance from the circle center to the inner arc (the lagoon  
395 boundary for convergent aquifers while the internal no-flow boundary for divergent aquifers)  
396 weakens the role played by aquifer geometry and hence lead to a smaller deviation of the  
397 extent of seawater intrusion between the three topographies.

398 Real island aquifers are expected to exhibit more complexity than considered here, e.g.,  
399 that will have more complex shapes and are subjected to transient flow conditions caused by  
400 tides, waves and pumping (Mantoglou et al. 2003; Pool & Carrera., 2011; Werner et al.,  
401 2013). In addition, since the experimental scale of Memari et al. (2020) is necessarily small,  
402 future experiments and field data are needed to further validate and facilitate the analytical  
403 solutions. Despite this, the new analytical solutions, validated against experiments, can be  
404 used as a tool to rapidly estimate seawater intrusion in ASAs once known island geometry and  
405 corresponding soil properties are given.

406 **Appendix: Analytical Solutions for Rectangular Aquifers**

407 For rectangular aquifers, the seawater intrusion in Unit 1 is identical to that in Unit 2  
 408 because of symmetry. With the interface tip on the aquifer bed, analytical solutions for the  
 409 freshwater-seawater interface ( $h_s$ ), watertable elevation ( $h$ ), and  $x$ -coordinate of the interface  
 410 tip in Unit 2 ( $x_{t2}$ ) can be respectively written as (Lu et al., 2019),

$$411 \quad h_s = \alpha \sqrt{\frac{N}{(1+\alpha)K_s} \left( \frac{L^2}{4} - x^2 \right)} \quad (\text{A1})$$

$$412 \quad h = \begin{cases} \sqrt{\frac{N}{K_s} (x_{t2}^2 - x^2) + \left( \frac{H_s}{\alpha} + H_s \right)} & 0 \leq x \leq x_{t2} \\ \sqrt{\frac{N}{(1+\alpha)K_s} \left( \frac{L^2}{4} - x^2 \right) + H_s} & x_{t2} < x \leq \frac{L}{2} \end{cases} \quad (\text{A2})$$

$$413 \quad x_{t2} = \sqrt{\frac{L^2}{4} - \frac{(1+\alpha)K_s}{N} \left( \frac{H_s^2}{\alpha^2} \right)} \quad (\text{A3})$$

414 When the interface tip is above the aquifer bed, the analytical solution for the freshwater-  
 415 seawater interface location and watertable elevation in Unit 2 are the same as equations (A1)  
 416 and (A2), respectively.

417 **Code/Data availability**

418 Experimental data used in this study are compiled from Memari et al. (2020).

419 **Author contribution**

420 All authors contributed to the design of the research. ZL carried out data collation,  
421 developed the analytical solutions and prepared the manuscript with contributions from all  
422 co-authors. All authors contributed to the interpretation of the results and provided feedback.



423 **Competing interests**

424 The authors declare that they have no conflicts of interest.

425 **Acknowledgments**

426       This research was supported by the National Key R&D Program of China  
427 (2019YFC0409004) and the National Natural Science Foundation of China (51979095,  
428 41807178). ZL acknowledges EPFL for financial support and JK acknowledges the Qing Lan  
429 Project of Jiangsu Province (2020).

430 **References**

431 Ayers, J. F., & Vacher, H. L. (1986). Hydrogeology of an atoll island: A conceptual model  
432 from detailed study of a Micronesian example. *Groundwater*, 24(2), 185-198.

433 <https://doi.org/10.1111/j.1745-6584.1986.tb00994.x>

434 Bailey, R. T., Jenson, J. W., & Olsen, A. E. (2010). Estimating the ground water resources of  
435 atoll islands. *Water*, 2(1), 1-27. <https://doi.org/10.3390/w2010001>

436 Bailey, R. T., Jenson, J. W., & Olsen, A. E. (2009). Numerical modeling of atoll island  
437 hydrogeology. *Groundwater*, 47(2), 184-196. <https://doi.org/10.1111/j.1745->

438 [6584.2008.00520.x](https://doi.org/10.1111/j.1745-6584.2008.00520.x)

439 Bedekar, V. S., Memari, S. S., & Clement, T. P. (2019). Investigation of transient freshwater  
440 storage in island aquifers. *Journal of Contaminant Hydrology*, 221, 98-107.

441 <https://doi.org/10.1016/j.jconhyd.2019.02.004>

442 Chesnaux, R., & Allen, D. M. (2008). Groundwater travel times for unconfined island  
443 aquifers bounded by freshwater or seawater. *Hydrogeology Journal*, 16(3), 437-445.

444 <https://doi.org/10.1007/s10040-007-0241-6>

445 Dose, E. J., Stoeckl, L., Houben, G. J., Vacher, H. L., Vassolo, S., Dietrich, J., &

446 Himmelsbach, T. (2014). Experiments and modeling of freshwater lenses in layered  
447 aquifers: Steady state interface geometry. *Journal of Hydrology*, 509, 621-630.

448 <https://doi.org/10.1016/j.jhydrol.2013.10.010>

449 Drabbe J. & Badon Ghijben, W. (1889). *Nota in verband met de voorgenomen put boring*  
450 *nabij Amsterdam*. Tijdschrift van het Koninklijk Instituut van Ingenieurs. pp. 8-22,

451 Gravenhage, Netherlands.

452 Duvat, V. K. E. (2019). A global assessment of atoll island planform changes over the past  
453 decades. *Wiley Interdisciplinary Reviews: Climate Change*, 10(1), e557.  
454 <https://doi.org/10.1002/wcc.557>

455 Fetter, C. W. (1972). Position of the saline water interface beneath oceanic islands. *Water  
456 Resources Research*, 8(5), 1307-1315. <https://doi.org/10.1029/WR008i005p01307>

457 Gingerich, S. B., Voss, C. I., & Johnson, A. G. (2017). Seawater-flooding events and impact  
458 on freshwater lenses of low-lying islands: Controlling factors, basic management and  
459 mitigation. *Journal of Hydrology*, 551, 676-688.  
460 <https://doi.org/10.1016/j.jhydrol.2017.03.001>

461 Greskowiak, J., Röper, T., & Post, V. E. (2013). Closed-form approximations for two-  
462 dimensional groundwater age patterns in a fresh water lens. *Groundwater*, 51(4), 629-  
463 634. <https://doi.org/10.1111/j.1745-6584.2012.00996.x>

464 Herzberg, A. (1901). Die wasserversorgung einiger Nordseebäder. *Journal für  
465 Gasbeleuchtung und Wasserversorgung*, 44, 815-819, 45, 842-844.

466 Hilberts, A. G. J., Troch, P. A., & Paniconi, C. (2005). Storage-dependent drainable porosity  
467 for complex hillslopes. *Water Resources Research*, 41(6), W06001.  
468 <https://doi.org/10.1029/2004WR003725>

469 Ketabchi, H., Mahmoodzadeh, D., Ataie-Ashtiani, B., Werner, A. D., & Simmons, C. T.  
470 (2014). Sea-level rise impact on fresh groundwater lenses in two-layer small islands.  
471 *Hydrological Processes*, 28(24), 5938-5953. <https://doi.org/10.1002/hyp.10059>

472 Kong, J., Shen, C., Luo, Z., Hua, G., & Zhao, H. (2016). Improvement of the hillslope-storage  
473 Boussinesq model by considering lateral flow in the unsaturated zone. *Water*  
474 *Resources Research*, 52(4), 2965-2984. <https://doi.org/10.1002/2015WR018054>

475 Lam, R. K. (1974). Atoll permeability calculated from tidal diffusion. *Journal of Geophysical*  
476 *Research*, 79(21), 3073-3081. <https://doi.org/10.1029/JC079i021p03073>

477 Liu, J., & Tokunaga, T. (2019). Future risks of tsunami-induced seawater intrusion into  
478 unconfined coastal aquifers: Insights from numerical simulations at Niijima Island,  
479 Japan. *Water Resources Research*, 55(12), 10082-10104.  
480 <https://doi.org/10.1029/2019WR025386>

481 Liu, Y., X. Mao, J. Chen, and D. A. Barry. 2014. Influence of a coarse interlayer on seawater  
482 intrusion and contaminant migration in coastal aquifers. *Hydrological Processes*, 28(20),  
483 5162-5175. <https://dx.doi.org/10.1002/hyp.10002>

484 Lu, C., Cao, H., Ma, J., Shi, W., Rathore, S. S., Wu, J., & Luo, J. (2019). A proof-of-concept  
485 study of using a less permeable slice along the shoreline to increase fresh groundwater  
486 storage of oceanic islands: Analytical and experimental validation. *Water Resources*  
487 *Research*, 55(8), 6450-6463. <https://doi.org/10.1029/2018WR024529>

488 Lu, C., Xin, P., Kong, J., Li, L., & Luo, J. (2016). Analytical solutions of seawater intrusion in  
489 sloping confined and unconfined coastal aquifers. *Water Resources Research*, 52(9),  
490 6989-7004. <https://doi.org/10.1002/2016WR019101>

491 Luo, Z., Shen, C., Kong, J., Hua, G., Gao, X., Zhao, Z., Zhao, H., & Li, L. (2018). Effects of  
492 unsaturated flow on hillslope recession characteristics. *Water Resources Research*,

493 54(3), 2037-2056. <https://doi.org/10.1002/2017WR022257>

494 Mantoglou, A. (2003). Pumping management of coastal aquifers using analytical models of  
495 saltwater intrusion. *Water Resources Research*, 39(12), 1335.  
496 <https://doi.org/10.1029/2002WR001891>

497 Memari, S. S., Bedekar, V. S., & Clement, T. P. (2020). Laboratory and numerical  
498 investigation of saltwater intrusion processes in a circular island aquifer. *Water*  
499 *Resources Research*, 56(2), e2019WR025325. <https://doi.org/10.1029/2019WR025325>

500 Morgan, L. K., & Werner, A. D. (2014). Seawater intrusion vulnerability indicators for  
501 freshwater lenses in strip islands. *Journal of Hydrology*, 508, 322-327.  
502 <https://doi.org/10.1016/j.jhydrol.2013.11.002>

503 Paniconi, C., Troch, P. A., Van Loon, E. E., & Hilberts, A. G. (2003). Hillslope-storage  
504 Boussinesq model for subsurface flow and variable source areas along complex  
505 hillslopes: 2. Intercomparison with a three-dimensional Richards equation model.  
506 *Water Resources Research*, 39(11), 1317. <https://doi.org/10.1029/2002WR001730>

507 Pool, M., & Carrera, J. (2011). A correction factor to account for mixing in Ghyben-Herzberg  
508 and critical pumping rate approximations of seawater intrusion in coastal aquifers.  
509 *Water Resources Research*, 47(5), W05506. <https://doi.org/10.1029/2010WR010256>

510 Post, V. E. (2018). Annotated translation of “Nota in verband met de voorgenomen putboring  
511 nabij Amsterdam [Note concerning the intended well drilling near Amsterdam]” by J.  
512 Drabbe and W. Badon Ghijben (1889). *Hydrogeology Journal*, 26(6), 1771-1788.  
513 <https://doi.org/10.1007/s10040-018-1797-z>

514 Post, V. E. A., Houben, G. J., Stoeckl, L., & Sültenfuß, J. (2019). Behaviour of tritium and  
515 tritiogenic helium in freshwater lens groundwater systems: Insights from Langeoog  
516 Island, Germany. *Geofluids*, Volume 2019, Article ID 1494326.  
517 <https://doi.org/10.1155/2019/1494326>

518 Röper, T., Greskowiak, J., Freund, H., & Massmann, G. (2013). Freshwater lens formation  
519 below juvenile dunes on a barrier island (Spiekeroog, Northwest Germany). *Estuarine,  
520 Coastal and Shelf Science*, 121-122, 40-50. <https://doi.org/10.1016/j.ecss.2013.02.004>

521 Stagnitti, F., Parlange, M. B., Steenhuis, T. S., & Parlange, J.-Y. (1986). Drainage from a  
522 uniform soil layer on a hillslope. *Water Resources Research*, 22(5), 631-634.  
523 <https://doi.org/10.1029/WR022i005p00631>

524 Stoeckl, L., Houben, G. J., & Dose, E. J. (2015). Experiments and modeling of flow processes  
525 in freshwater lenses in layered island aquifers: Analysis of age stratification, travel  
526 times and interface propagation. *Journal of Hydrology*, 529, 159-168.  
527 <https://doi.org/10.1016/j.jhydrol.2015.07.019>

528 Storlazzi, C. D., Gingerich, S. B., van Dongeren, A., Cheriton, O. M., Swarzenski, P. W.,  
529 Quataert, E., Voss, C. I., Field, D. W., Annamalai, H., Piniak, G. A., & McCall, R.  
530 (2018). Most atolls will be uninhabitable by the mid-21st century because of sea-level  
531 rise exacerbating wave-driven flooding. *Science Advances*, 4(4), eaap9741.  
532 <https://doi.org/10.1126/sciadv.aap9741>

533 Strack, O. D. L. (1976). A single-potential solution for regional interface problems in coastal  
534 aquifers. *Water Resources Research*, 12(6), 1165-1174.

535 <https://doi.org/10.1029/WR012i006p01165>

536 Stuyfzand, P. J. (2017). Observations and analytical modeling of freshwater and rainwater  
537 lenses in coastal dune systems. *Journal of Coastal Conservation*, 21(5), 577-593.  
538 <https://doi.org/10.1007/s11852-016-0456-6>

539 Stuyfzand, P. J. (1993). *Hydrochemistry and hydrology of the coastal dune area of the Western*  
540 *Netherlands*. Ph.D. Thesis. Vrije University, Amsterdam, KIWA, ISBN 90-74741-01-  
541 0. <http://dare.ubvu.vu.nl/handle/1871/12716>

542 Thomas, A., Baptiste, A., Martyr-Koller, R., Pringle, P., & Rhiney, K. (2020). Climate change  
543 and small island developing states. *Annual Review of Environment and Resources*,  
544 45(1), 1-27. <https://doi.org/10.1146/annurev-environ-012320-083355>

545 Troch, P. A., Paniconi, C., & Emiel van Loon, E. (2003). Hillslope-storage Boussinesq model  
546 for subsurface flow and variable source areas along complex hillslopes: 1.  
547 Formulation and characteristic response. *Water Resources Research*, 39(11), 1316.  
548 <https://doi.org/10.1029/2002WR001728>

549 Underwood, M. R., Peterson, F. L., & Voss, C. I. (1992). Groundwater lens dynamics of atoll  
550 islands. *Water Resources Research*, 28(11), 2889-2902.  
551 <https://doi.org/10.1029/92WR01723>

552 Vacher, H. L. 1988. Dupuit-Ghyben-Herzberg analysis of strip-island lenses. *Geological*  
553 *Society of America Bulletin*, 100, 580-591. [https://doi.org/10.1130/0016-7606\(1988\)100<0580:DGHAOS>2.3.CO;2](https://doi.org/10.1130/0016-7606(1988)100<0580:DGHAOS>2.3.CO;2)

554 [7606\(1988\)100<0580:DGHAOS>2.3.CO;2](https://doi.org/10.1130/0016-7606(1988)100<0580:DGHAOS>2.3.CO;2)

555 Werner, A. D., Sharp, H. K., Galvis, S. C., Post, V. E., & Sinclair, P. (2017). Hydrogeology



556 and management of freshwater lenses on atoll islands: Review of current knowledge  
557 and research needs. *Journal of Hydrology*, 551, 819-844.

558 <https://doi.org/10.1016/j.jhydrol.2017.02.047>

559 Werner, A. D., Bakker, M., Post, V. E., Vandenbohede, A., Lu, C., Ataie-Ashtiani, B.,  
560 Simmons, C. T., & Barry, D. A. (2013). Seawater intrusion processes, investigation  
561 and management: Recent advances and future challenges. *Advances in Water*  
562 *Resources*, 51, 3-26. <https://doi.org/10.1016/j.advwatres.2012.03.004>

563 Werner, A. D., & Simmons, C. T. (2009). Impact of sea-level rise on sea water intrusion in  
564 coastal aquifers. *Groundwater*, 47(2), 197-204. [https://doi.org/10.1111/j.1745-](https://doi.org/10.1111/j.1745-6584.2008.00535.x)  
565 [6584.2008.00535.x](https://doi.org/10.1111/j.1745-6584.2008.00535.x)

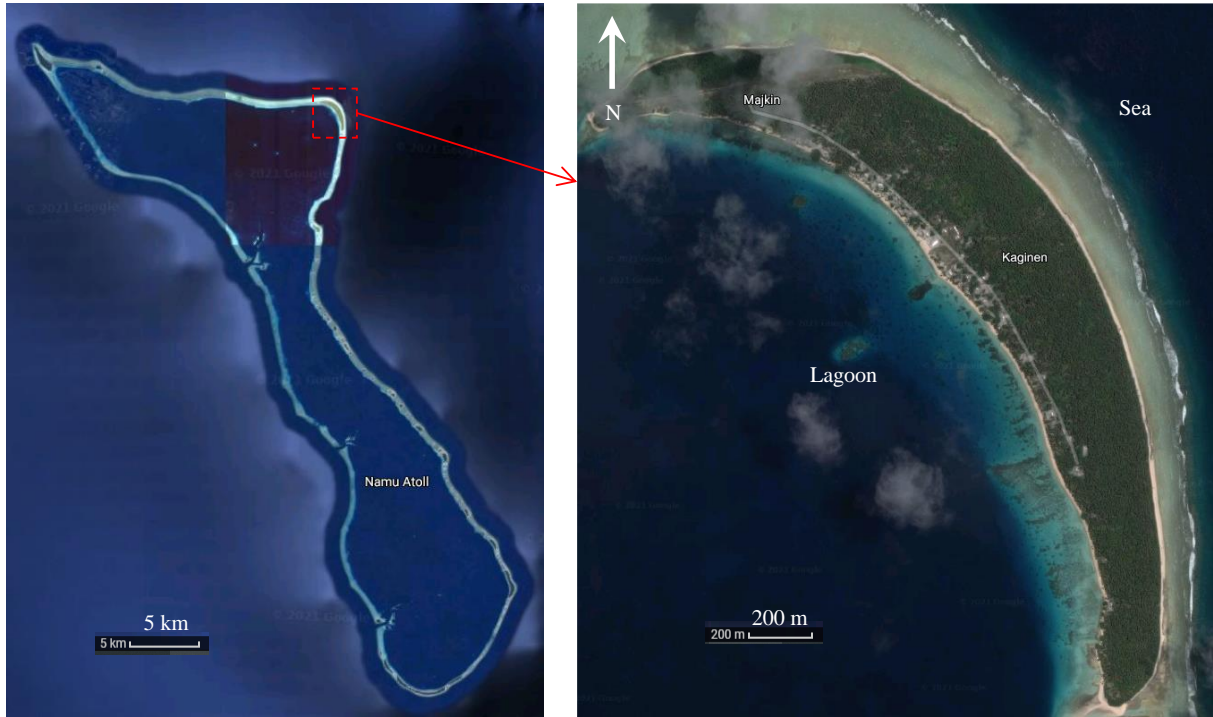
566 White, I., & Falkland, T. (2010). Management of freshwater lenses on small Pacific islands.  
567 *Hydrogeology Journal*, 18(1), 227-246. <https://doi.org/10.1007/s10040-009-0525-0>

568 Zhang, Y., Li, L., Erler, D. V., Santos, I., & Lockington, D. (2016). Effects of alongshore  
569 morphology on groundwater flow and solute transport in a nearshore aquifer. *Water*  
570 *Resources Research*, 52(2), 990-1008. <https://doi.org/10.1002/2015WR017420>

571 **Table 1.** List of parameters use in different simulations.

	No.	$L^*$ (m)	$L_0$ (m)	$H_s$ (m)	$d$ (m)	$\alpha$ (-)	$n_e$ (-)	$K_s$ (m s <sup>-1</sup> )	$N$ (m s <sup>-1</sup> )
	1	1000	200	38	45	40	0.4	$1.23 \times 10^{-2}$	$1 \times 10^{-6}$
Cases	2	1000	200	38	45	40	0.4	$1.23 \times 10^{-2}$	$3 \times 10^{-7}$
Simulated	3	1000	†	38	45	40	0.4	$1.23 \times 10^{-2}$	$1 \times 10^{-6}$
	4	†	200	38	45	40	0.4	$1.23 \times 10^{-2}$	$1 \times 10^{-6}$

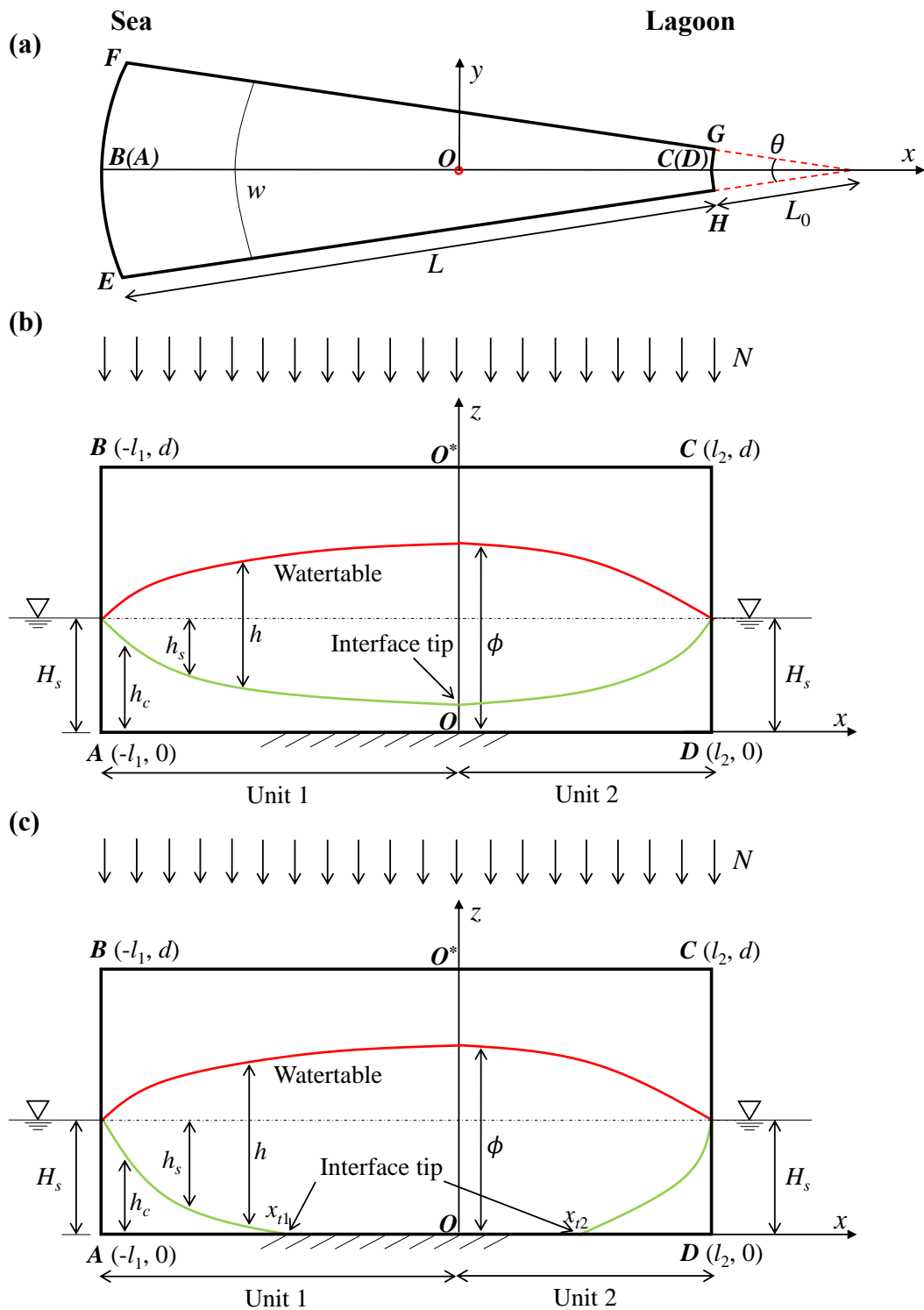
572 †The parameter is varied: The range of  $L_0$  is from 200 to 6000 m, whereas the range of  $L^*$  is from 600 to 1600 m.



573

574 **Figure 1.** Island with an annulus segment in the Namu Atoll, Marshall Islands (© Google

575 Earth).



576

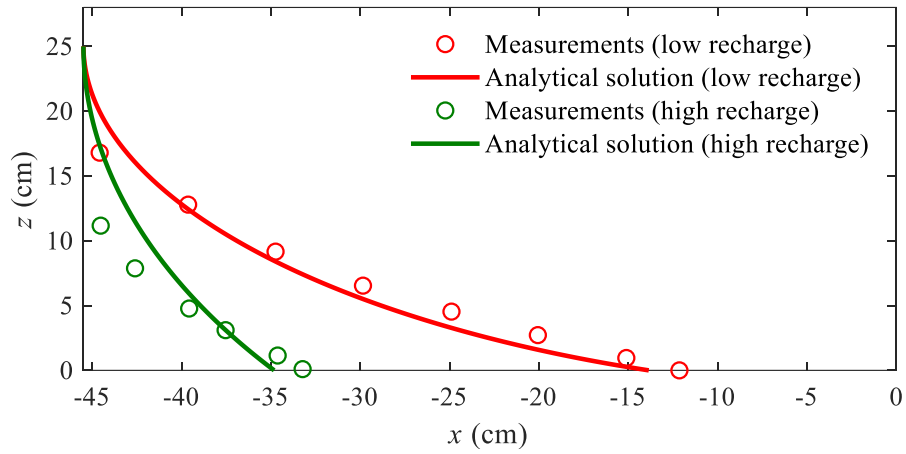
577 **Figure 2.** Conceptual model of an annulus segment aquifer (a slice of an atoll island). (a) Plan

578 view and (b, c) side view with the saltwater interface tip (b) above the aquifer bed (single

579 location) and (c) on the aquifer bed (two locations). In (a), the sea boundary is on  $EF$  and the

580 atoll lagoon boundary is on  $HG$ ; In (b) and (c),  $AD$  is the impermeable base and  $OO^*$  is the

581 internal no-flow boundary.

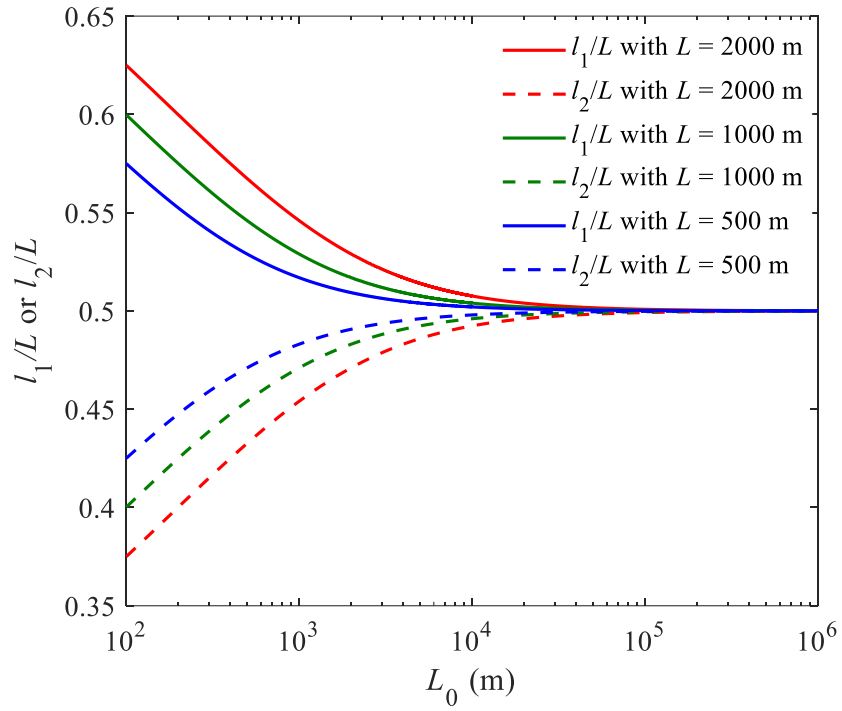


582

583 **Figure 3.** Comparison between analytical and experimental (data compiled from Memari et

584 al., 2020) results for the freshwater-seawater interface location for different recharge events.

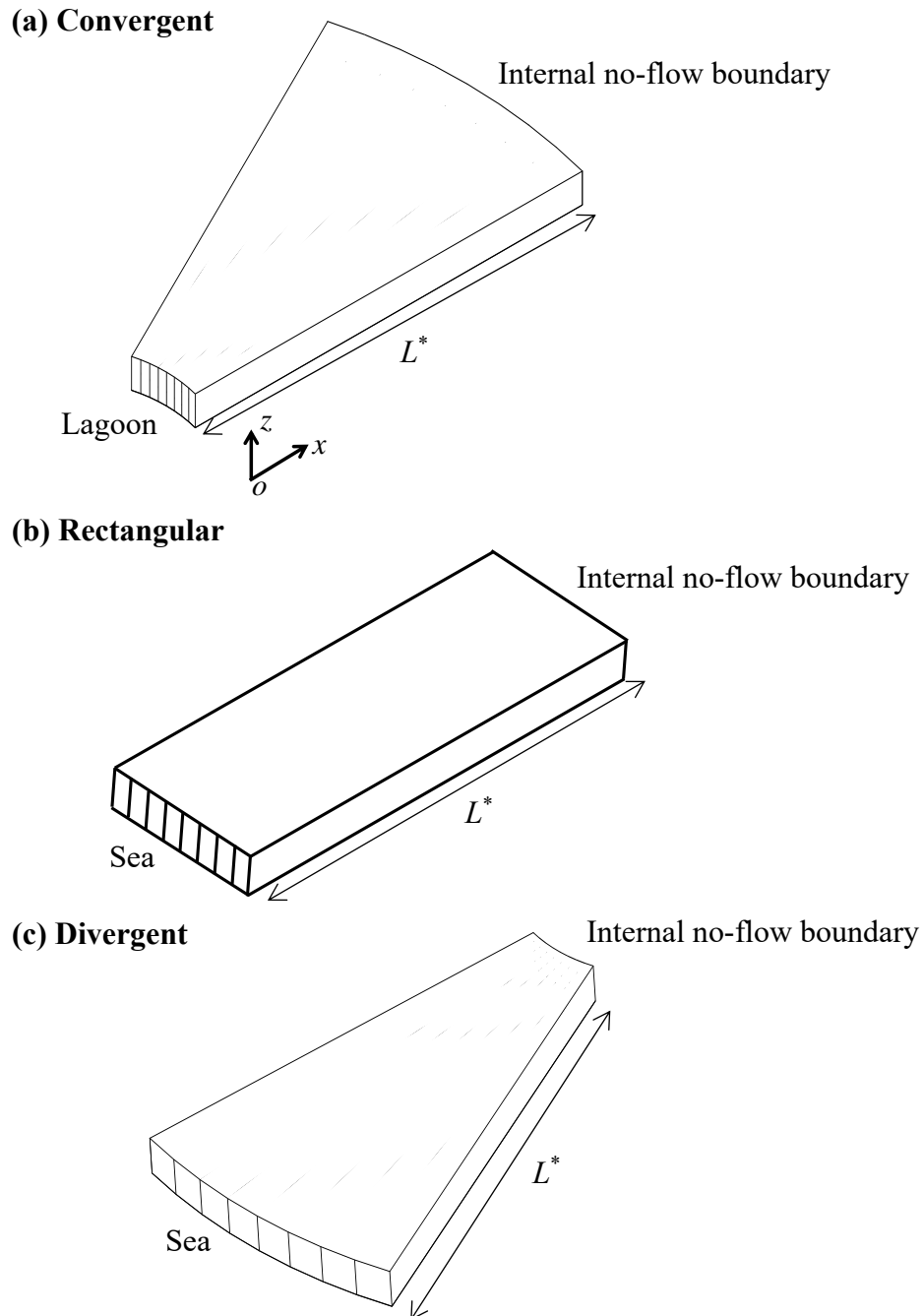
585 Note that the left and right sides are the sea and internal no-flow boundaries, respectively.



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**Figure 4.** Widths of Unit 1 and Unit 2 versus  $L_0$  for aquifers with different total width  $L$ .



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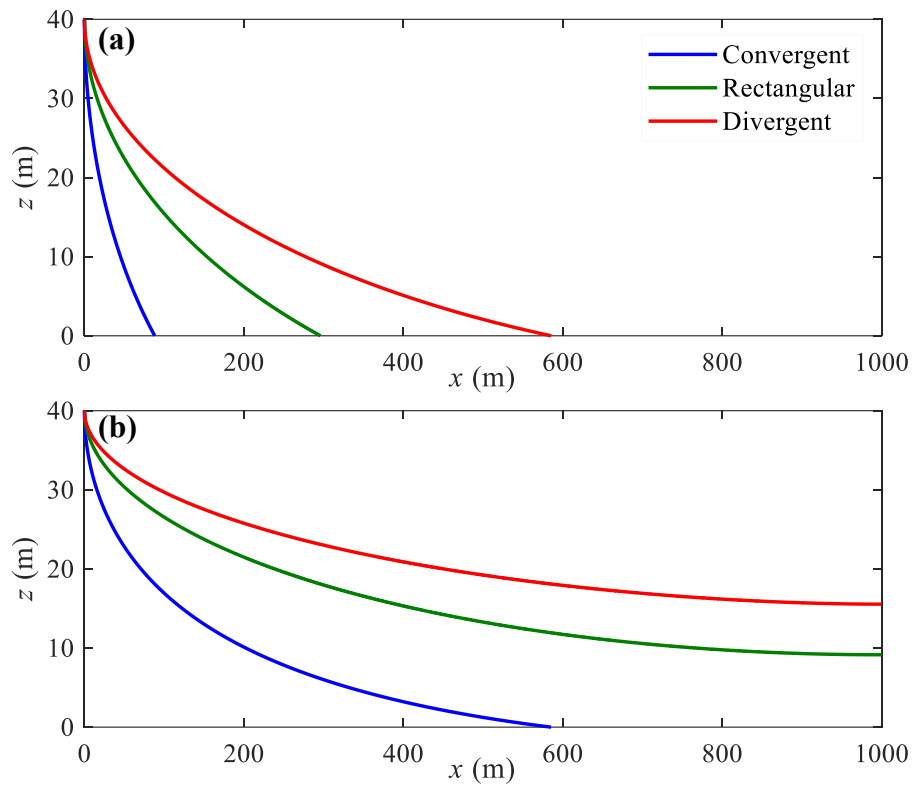
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**Figure 5.** Three-dimensional view of (a) convergent (smaller side facing the lagoon), (b) rectangular and (c) divergent aquifers (larger side facing the sea) compared in this study.  $L^*$  represents the distance from the sea/lagoon to the internal no-flow boundary, i.e.,  $l_1$  or  $l_2$  in Figure 2. The internal no-flow boundary corresponds to the  $z$ -axis in Figure 2.





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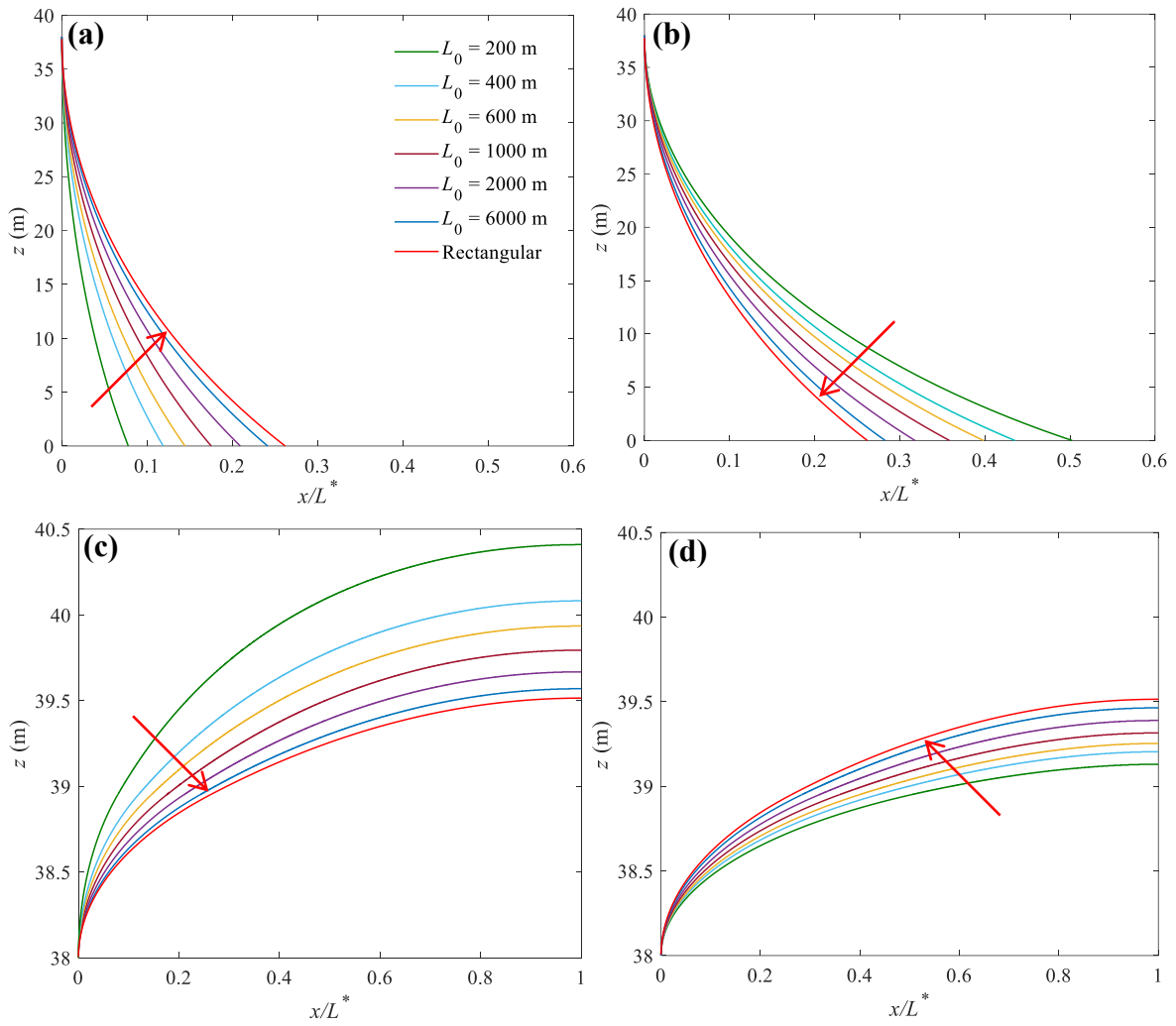
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**Figure 6.** Freshwater-seawater interface predicted by analytical solutions for three different aquifers with (a) high and (b) low recharge (Cases 1 and 2 in Table 1). Note that  $x = 1000$  m is the internal no-flow boundary in Figure 5.



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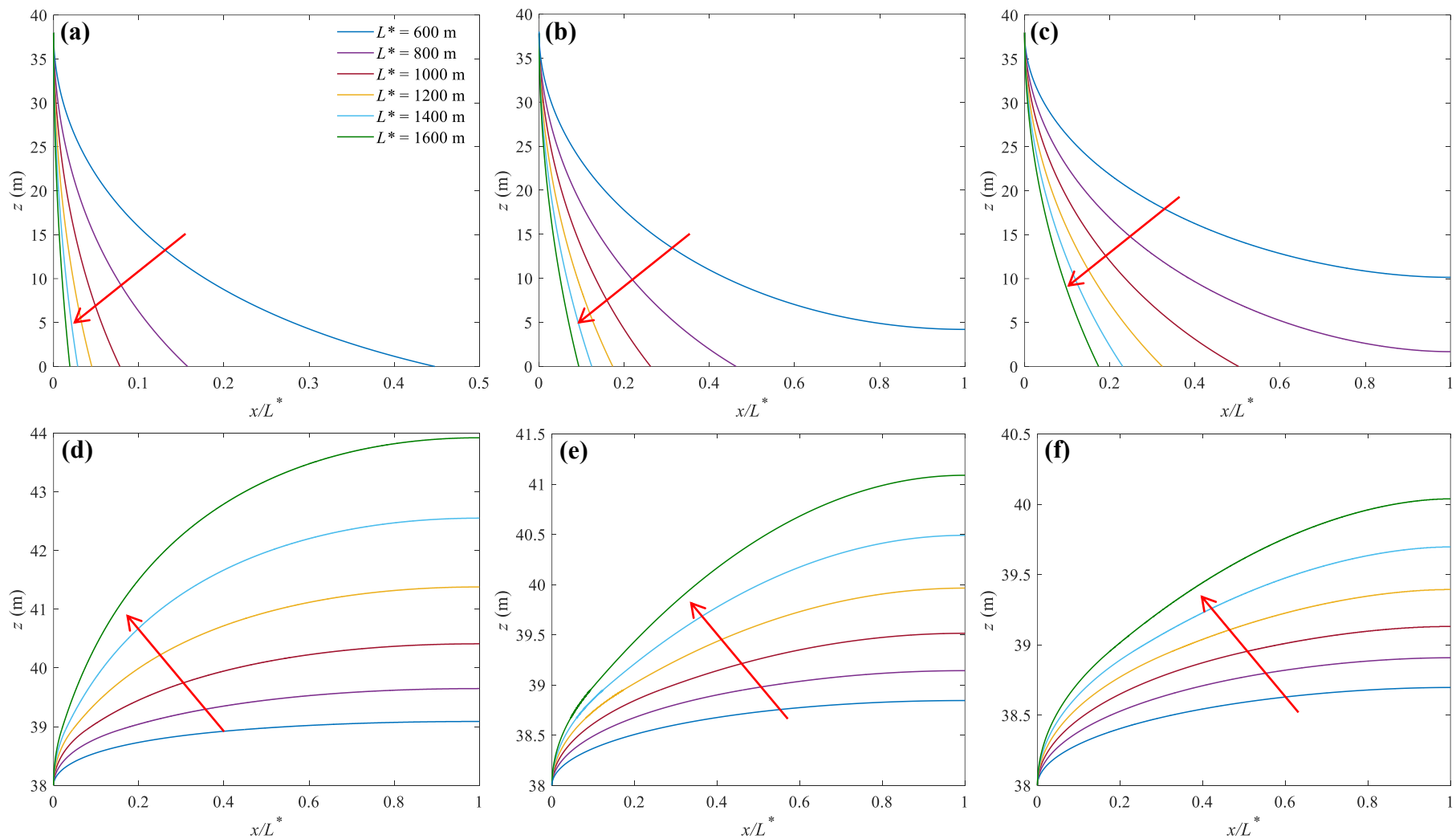
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**Figure 7.** Sensitivity of (a, b) the locations of the freshwater-seawater interface and (c, d) watertable to  $L_0$  for convergent (left panel) and divergent (right panel) aquifers. The arrow in each plot shows the direction of increasing  $L_0$  (values given in (a), used to produce the different curves). Note that predictions for rectangular aquifers are independent of  $L_0$ .



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**Figure 8.** Sensitivity of (a-c) the locations of the freshwater-seawater interface and (d-f) watertable to  $L^*$  for convergent (a, d), rectangular (b, e)

and divergent (c, f) aquifers. The arrow in each plot points to the increase of  $L^*$  values used to construct each curve (values indicated in (a)).