1 2	Feedback mechanisms between precipitation and dissolution reactions across randomly heterogeneous conductivity fields
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9	Highlights
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10 11 12 13 14 15	 Regions of prevailing precipitation and dissolution are related to preferential flow patterns Large changes in non-Fickian transport parameters are observed while velocity variance display modest variations Initial heterogeneity facilitates attaining asymptotic average solute velocity value
16	Abstract
 17 18 19 20 21 22 23 24 25 26 27 28 	Our study investigates interplays between dissolution, precipitation, and transport processes taking place across randomly heterogeneous conductivity domains and the ensuing spatial distribution of preferential pathways. We do so by relying on a collection of computational analyses of reactive transport performed in two-dimensional systems where the (natural) logarithm of conductivity is characterized by various degrees of spatial heterogeneity. Our results document that precipitation and dissolution jointly take place in the system, the latter mainly occurring along preferential flowpaths associated with the conductivity field, the former being observed at locations close to and clearly separated from these. High conductivity values associated with the preferential flowpaths tend to further increase in time, giving rise to a self-sustained feedback between transport and reaction processes. The clear separation between regions where dissolution or precipitation takes place is imprinted onto the sample distributions of conductivity which tend to become visibly left skewed with time (with the appearance of a bimodal behavior at some times). The link between
29 30	conductivity changes and reaction-driven processes promotes the emergence of non-Fickian effective transport features. The latter can be captured through a continuous time random walk model where
29 30 31 32	conductivity changes and reaction-driven processes promotes the emergence of non-Fickian effective transport features. The latter can be captured through a continuous time random walk model where solute travel times are approximated with a truncated power law probability distribution. The parameters of such a model shift towards values associated with increasingly high non-Fickian
33	effective transport behavior as time progresses.

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1.Introduction

35 Diagnosis and characterization of the feedback between geochemical precipitation/dissolution reactions and solute transport processes in heterogeneous subsurface systems is key to a variety of 36 environmental and Earth science scenarios (Rege & Fogler, 1989; Berkowitz et al., 2016). A critical 37 challenge is the emergence of complex dependencies between physical and chemical processes taking 38 39 place across aquifer bodies (Saripalli et al., 2001). Heterogeneity of these systems promotes diverse 40 patterns of precipitation and/or dissolution that may imprint a variety of dynamic system responses, 41 including, e.g., wormholing and oscillatory behaviors of system attributes such as porosity and permeability (Edery et al., 2011; Garing et al., 2015; Golfier et al., 2002). Examples of practical
applications in this context include geologic CO₂ storage (e.g., Pawar et al., 2015; Noiriel & Daval,
2017; Cabeza et al., 2020 and references therein), acid injection in production wells (e.g., Liu et al.,
2017 and references therein), and reactive transport of contaminants (e.g., Ceriotti et al., 2018; Dalla
Libera et al., 2020 and references therein).

47 Computational studies can assist the analysis of patterns of chemical transport across heterogeneous subsurface systems in the presence of precipitation/dissolution phenomena. While 48 requiring an explicit description of the spatial heterogeneity of the system properties (Atchley et al., 49 2014), routine application of numerical simulations in practical settings is hampered by (i) our limited 50 51 knowledge of the system attributes, resulting in uncertainty affecting the parameterization of the 52 underlying physical and chemical processes and their variability, and (ii) the computational costs 53 required to properly quantify such uncertainties and their propagation onto environmental quantities of interest. In this context, we rely on an effective approach to characterize the evolution of key 54 55 features of solute transport in the presence of rock-fluid interactions across a porous medium whose 56 spatially heterogeneous conductivity field is interpreted according to a commonly employed 57 stochastic framework.

58 A critical element we tackle is related to the analysis of the dynamic feedback between reactive 59 transport and spatially heterogenous distributions of porous media attributes such as hydraulic conductivity. Following prior studies, we start by recognizing that, even under geochemical 60 61 equilibrium conditions, the spatial heterogeneity of system attributes typically imprints an uneven 62 spatial distribution of regions where chemical reactions take place, local fluctuations of conductivity being key to this element (Edery, Porta, et al., 2016). Further to this, our conceptualization of the 63 setting is grounded on the observation that rendering of transport features in geological formations 64 through effective formulations typically requires embedding non-Fickian features. To this end, we 65 rely on an upscaled description of transport where solute travel/waiting times are approximated with 66 a truncated power law probability density function (PDF), hereafter termed TPL (Berkowitz et al., 67 68 2006). This effective description is particularly relevant because the emergence of non-Fickian 69 transport features in heterogeneous formations has been observed at diverse scales of observation, 70 including pore-, laboratory- and field-scale scenarios (e.g., Edery et al., 2011; Muljadi et al., 2018; 71 Menke et al., 2018 and references therein).

72 In line with our objective, we rest on the framework of analysis developed in (Edery et al., 73 2014; Edery, Porta, et al., 2016), where an effective depiction of transport processes is parametrized 74 as a function of the statistics of solute residence times in randomly heterogeneous conductivity fields. 75 A main element of this framework is that it yields a link between the TPL weighting times and the 76 occurrence of preferential pathways that can be obtained from computational studies of transport in 77 such conductivity fields (Edery et al., 2014). As a result, the methodology is conducive to an effective 78 (or upscaled) representation of local features to identify signatures of non-Fickian transport (see also, 79 e.g., Dentz et al., 2011; Edery et al., 2014). To illustrate the main features associated with the scenario 80 of interest, we consider a Darcy-scale formulation of a reactive transport setup, where precipitation 81 and/or dissolution of minerals are driven by the injection of an acid compound establishing local equilibrium with the resident fluid and a solid matrix of the host porous medium which is considered 82 83 to be composed of calcite mineral. While the geochemical processes we consider are somehow 84 streamlined with respect to a field-scale scenario (see, e.g., Lichtner, 1988; Dreybrodt et al., 1996), they embed the main elements characterizing the interplay between solute transport and rock-fluid 85 interactions in Darcy-scale systems (e.g., Edery et al., 2011). Within this conceptual picture, our study 86 aims at investigating (i) the interplay between the reactive process and the ensuing spatial distribution 87

of preferential pathways associated with spatially heterogeneous conductivities and (*ii*) the link
 between locally occurring reaction-driven phenomena and emerging non-Fickian effective transport
 features, as captured by the TPL formulation of the PDF of particle travel/waiting times.

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2.Methodology

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2.1 Chemical model

We simulate a reactive transport scenario where calcite $(CaCO_{3}(s))$, subscript (s) denoting solid mineral) can dissolve or precipitate locally in the presence of chemical equilibrium between dissolved carbonic acid (H_2CO_3) and pH. The amount of dissolved H_2CO_3 as a function of pH (see Figure S1 in the supplementary material) is then governed by equilibrium conditions, which is tantamount to assuming a locally instantaneous reaction (i.e., as the reactive time scale approaches zero at equilibrium, this scenario corresponds to a local Damköhler number which tends to infinity). The formulation describing the chemical reactions can then be streamlined as:

$$100 \quad CaCO_{3(s)} \leftrightarrow Ca^{2+} + CO_3^{2-} \tag{1a}$$

101
$$H_2CO_3 \leftrightarrow 2H^+ + CO_3^{2-}$$
 (1b)

according to which two protons H^+ in (1b) react with CO_3^{2-} to produce H_2CO_3 that in turn drives 102 dissolution of the host calcium carbonate solid matrix. While simplified (see, e.g., Lichtner, 1988; 103 Dreybrodt et al., 1996), the chemical set-up follows previous work by Edery et al. (2011), where there 104 105 is an extensive assessment of the employed formulation (see also the Supplementary Material). In this context, and consistent with typical experimental practice, we consider the injected fluid and the 106 porous medium to be associated with a source of H^+ and an abundance of Ca^{2+} , respectively. Thus, 107 Ca^{2+} is not rate limiting and the spatial distribution of H^+ , as driven by transport and reaction, 108 109 governs pH. The rate limiting reaction is then (1b), that is controlled by the available H^+ (or pH), 110 similar to observations associated with other studies (Singurindy & Berkowitz, 2004; Edery et al., 2011). The chemical reaction system (1a) and (1b) is here simplified (see, e.g., Krauskopf & Bird, 111 1967) through: 112

$$113 \quad co \leftrightarrow 2h + c \tag{2}$$

114 where *co* denotes H_2CO_3 , *h* and *c* representing H^+ and $CaCO_{3(s)}$, respectively.

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2.2 Flow and transport modeling

116 Our computational setting is intended to mimic a laboratory scale scenario where a 60×24 cm² two-dimensional flow cell is filled with a porous system formed by a $CaCO_{3(s)}$ solid matrix. The 117 system is initially fully saturated with water and an injection of low pH water takes place across the 118 119 upstream side of the cell. To investigate the influence of the dissolution/precipitation reaction on 120 solute transport, we consider a uniform in the mean groundwater flow, taking place within a two-121 dimensional domain where the (natural) logarithm of conductivity, $y = \ln(k)$, is considered as a 122 zero-mean, second-order stationary random field. The latter is further characterized by an isotropic, 123 simple exponential, covariance function, with (normalized) correlation length l/L, L being the length 124 of the domain along the main flow direction. Various degrees of heterogeneity of the system are analyzed upon considering values of log-conductivity variance $\sigma_0^2 = [1, 3, 5]$, subscript 0 denoting 125 that these values refer to the initially generated conductivity distributions (i.e., prior to the occurrence 126 127 of reactions). The domain is discretized through 300×120 elements of uniform size $\Delta = 0.2$ cm,

yielding a field size of 60×24 cm². Each field is synthetically generated through the widely tested 128 sequential Gaussian simulator GCOSIM3D (Gómez-Hernández & Journel, 1993) and is characterized 129 by l/L = 0.016. This yields a value of $\Delta/l = 0.2$, which is deemed adequate to capture the local features 130 of the covariance of y and their impact on the main statistics of the velocity field and travel times 131 132 (Ababou et al., 1989; Riva et al., 2009). We note that, while our study is representative of a laboratory scale analysis, the dimensions of the domain have no particular implication and they are only selected 133 to ensure a meaningful description of the correlation structure that is included in the initially generated 134 135 conductivity fields.

For each value of σ_0^2 , 20 random realizations of y are generated, each being then subject to a deterministic pressure drop ($\Delta H = 100$ cm) between the inlet (left) and the outlet (right) sides. The local distribution of fluid velocity is computed through

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$$\nabla \cdot \boldsymbol{q}(\boldsymbol{x}) = 0; \, \boldsymbol{q}(\boldsymbol{x}) = -k(\boldsymbol{x}) \cdot \nabla h(\boldsymbol{x}) \tag{3}$$

140 where q(x) is the local Darcy flux, vector x corresponding to spatial location. The local fluid velocity 141 field is then obtained as $v = q/\theta$, a constant initial porosity $\theta = 0.4$ being here considered for the 142 porous medium.

143 Solute transport is then simulated across each conductivity field by a particle tracking approach (Le Borgne et al., 2008). A number of $10^5 h$ particles (see (2)), which is selected to represent a full 144 pore volume (whose magnitude is evaluated through the initial condition, i.e., before porosity and 145 permeability are altered by the reactive processes) at constant pH = 3.5, is divided by the domain 146 length and multiplied by the mean velocity ($\hat{\mathbf{v}}$, as evaluated from (4) across the whole domain). A 147 total amount of particles evaluated as $Integer(10^5/L \cdot \Delta t \cdot \hat{\mathbf{v}})$ is then injected into the system at 148 regular time intervals ($\Delta t = 0.1$ min). Particles are injected at the very beginning of each time step Δt 149 and are flux-weighted according to the conductivity distribution at the inlet. The particles representing 150 a full pore volume correspond to $M_{H^+} = 10.79$ moles of H^+ , the same amount being injected across 151 the simulation course to obtain a constant pH = 3.5 in the injected fluid, while absence of h particles 152 is taken to correspond to pH = 8. We then evaluate the pH value (or H^+ molar mass) associated with 153 154 each h particle by dividing the total number of H^+ moles required to obtain a pH = 3.5 (i.e., 10.79 155 mol of H^+) by the pore volume (as represented by $10^5 h$ particles).

156 The upper and lower boundaries of the domain are reflective while the outlet boundary is 157 absorbing. Particle migration is simulated through

158
$$\boldsymbol{d} = \boldsymbol{v}[\boldsymbol{x}(t_k)]\boldsymbol{\delta}t + \boldsymbol{d}_D \tag{4}$$

159 where d is particle displacement, $x(t_k)$ is the vector identifying spatial coordinates of particle location 160 at time t_k , v is fluid velocity at $x(t_k)$, $\delta t = \delta s/v$ is the temporal displacement magnitude (v is the norm of v), and d_D is the diffusive displacement. The latter is evaluated as $d_D = N[0,1]\sqrt{2D_m\delta t}$, where 161 N[0,1] represents a two-dimensional vector of random variables, whose entries are mutually 162 163 independent and sampled from a Gaussian distribution with zero mean and unit variance, $D_m = 10^{-5}$ cm²/min representing diffusion. The value of δs is selected to be an order of magnitude less than Δ , 164 165 to accurately sample the velocity variability within a conductivity block. It is noted that, while taking into account pore-scale processes within a continuum-scale model through a local scale dispersion 166 167 could be a modeling option, this would add an additional level of complexity to our numerical 168 simulation without modifying the key elements of our work, which is focused on the interaction 169 between flow patterns and reactive processes. In this context, our modeling choice is to represent 170 macro-dispersive effects through averaging (in a multi-realization context) the effect of fluctuations of velocity arising between diverse realizations of the conductivity field. Consistent with this, we rely on a constant and isotropic diffusion coefficient that can be associated with an advection-dominated transport regime (as quantified in terms of a Péclet number, as seen in the following). This choice is also consistent with previous works (e.g., Aquino and Bolster, 2017; Wright et al., 2021).

175 Coupling between particle evolution and the geochemical setup illustrated in Section 2.1 is 176 achieved in two steps. First, we advance all particles according to the displacement mechanism described above. Second, we satisfy the equilibrium condition (2) by equilibrating both co and h177 within each cell, leading to precipitation or dissolution of a calcite mineral. The calcite volume to 178 mole ratio is taken as $M_{CaCO_3} = 37 \frac{cm^3}{mal}$ (Morse & Mackenzie, 1993) and the equilibrium between h 179 and *co* particles (according to (2)) leads to a local precipitation (or dissolution) of the solid. We update 180 181 in time the spatial distribution of porosity assuming that it is characterized by a uniform change within 182 each individual domain cell. We finally update conductivity through the Kozeny-Carman (KC) 183 formulation

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$$k(ar)_{ij} = k(br)_{ij} \cdot \frac{\theta(ar)_{ij}^{3}}{(1-\theta(ar)_{ij})^{2}} \cdot \frac{(1-\theta(br)_{ij})^{2}}{\theta(br)_{ij}^{3}}; \qquad \theta(ar)_{ij} = \theta(br)_{ij}(1 \mp M_{CaCO_{3}} \cdot M_{H^{+}})$$
(5)

where $k(ar)_{ij}$ and $\theta(ar)_{ij}$ are conductivity and porosity, respectively, after the reaction (ar) has 185 taken place, while $k(br)_{ij}$ and $\theta(br)_{ij}$ are their counterparts before the reaction is observed, 186 187 subscripts *i* and *j* being identifiers of a given cell. The process is repeated for each particle in each of the cells until an equilibrium between *co* and *h* is reached. We set an upper and a lower bound of 0.1 188 and 0.9, respectively, for porosity, to avoid the occurrence of unphysical porosity values. These 189 190 constraints are set for consistency with our assumptions, i.e., to consider Darcy flow across the porous 191 domain. A complete clogging (or opening) of a void space would require a different mathematical 192 and conceptual treatment, which is beyond the scope of our study. Precipitation is treated numerically 193 in a corresponding way.

194 We numerically calculate the updated local head and fluid velocity distributions from (4) at 195 time intervals of 10 Δt , to reduce constraints associated with computational costs. The computational cost of each realization is between 1~3 days (depending on the value of σ_0^2), upon relying on a 16 196 Xeon 2.6 Ghz processor with 64 GB RAM. With reference to the sensitivity of the results to the 197 198 numerical parameters, we note that: (a) when considering single realizations, the results showed only 199 minute sensitivity to increasing the number of particles by a factor of 10 (less than 3% difference in 200 the results was observed); (b) results display a slightly larger sensitivity to the time step, otherwise 201 resulting in a difference of less than about 5% in the amount of reaction when decreasing the time 202 interval by an order of magnitude; and (c) preliminary analyses aimed at assessing possible influences 203 of the grid size on the key results of the study imbued us with confidence about the quality of the 204 results obtained with the grid size employed in the study (details not shown), which we select as good 205 compromise between computational accuracy and execution time constraints, in light of our 206 objectives.

The updated conductivity field is extracted and stored at the above mentioned regular intervals of $10 \Delta t$. Transport of a non-reactive solute pulse is then simulated across each of these updated fields to capture the temporal evolution of the key parameters driving effective transport (see Section 2.3). While noting that natural porous media can exhibit complex relationships between permeability and porosity (Luquot & Gouze, 2009), which may not always be interpreted through the KC model (6), we employ the latter formulation because it is considered as a reference model in the literature and can serve as a proxy for alternative improved parameterizations (Erol et al., 2017). We also recall that we consider local equilibrium, this choice being consistent with the specific objective of this work which is related to the characterization of transport under dynamic evolution of continuum-scale quantities such as conductivity, rather than being focused on a detailed characterization of the effects of the reactive process on the pore-scale structure (as reflected, e.g., by local changes of specific surface area).

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2.3 Quantities of interest

220 The workflow described in Section 2.2 enables one to extract computationally-based quantities employed to characterize the analyzed reactive transport setup. As stated in Section 2.2, we simulate 221 222 a tracer test within the original fields as well as within those modified by precipitation/dissolution. 223 Particles are displaced through the action of advection and diffusion following a pulse (flux-weighted) 224 injection at the inlet. These non-reactive transport simulations are performed to assess base values of 225 parameters characterizing solute transport (a) prior to starting the reactive transport simulation as well 226 as (b) at specific times after reaction changed the field. The empirical PDF of particle waiting times 227 is assessed from the corresponding histogram starting by evaluating particle waiting times within a given domain cell through the inverse of the particle velocity computed at each time step multiplied 228 229 by the cell length and weighted by the number of particles visiting the cell. This PDF is then used to 230 estimate the parameters of the TPL model

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$$\psi(t_w) = \frac{n}{t_1} \frac{\exp(-t/t_2)}{\left(1 + \frac{t}{t_1}\right)^{1+\beta}}$$
 (6)

where t_w is the waiting time of a particle within a given domain cell, t_1 , t_2 , and β are model calibration parameters, which are estimated through a standard least square technique. Note that previous results have shown that the parameters obtained from (6) can be readily used to interpret breakthrough curves associated with non-reactive solutes (Edery et al., 2014).

236 The velocity fields are examined upon computing the evolution of the velocity and conductivity 237 fields statistics, as described in the following. Let us consider a discrete field of a generic quantity z_{ii} evaluated in a given cell ij. In the particle tracking numerical simulations we quantify $n_{ii}(t)$ as the 238 number of particles that have visited cell ij along the simulation up to a given time t. Thus, we 239 240 evaluate two relative frequency (or empirical probability) distributions, i.e., $f(z_{ii})$ and $f(nz_{ii})$, hereafter termed as unweighted and weighted distribution of the variable z_{ij} , respectively. We define 241 the weighted variable $nz_{ij}(t) = n_{ij}(t)z_{ij}/\bar{n}(t)$, where \bar{n} is the average value of n_{ij} . Note that the 242 243 adopted weighting scheme corresponds to weighting z_{ij} by the solute mass distribution. Average 244 values of the weighted and unweighted distributions (hereafter denoted as \overline{z} and \overline{nz} , respectively) can 245 then be evaluated. In the following we perform particle weighting in the non-reactive as well as in 246 the reactive transport scenarios. Distribution weighting by reactive particles is indicated by n_R , meaning that weighting is performed based on the reactive transport simulations (i.e., considering h 247 and co particles as explained above). The plain symbol n indicates weighting by non-reactive 248 particles, employed to simulate conservative tracer tests as detailed above. The variable z_{ii} is taken 249 to correspond to either the cell log-conductivity y_{ij} or fluid velocity v_{ij} in the results illustrated in 250 251 Section 3.

3.Results

253 We start our analyses by simulating transport of a non-reactive tracer across the generated heterogeneous conductivity domains. As log-conductivity variance increases, the range of 254 conductivity values naturally increases, this being reflected in the distribution $f(y_{ii})$ (see, e.g., Figure 255 1 a-c (blue circles)). The shape of weighted conductivity distributions, $f(ny_{ij})$, differs from the one 256 of $f(y_{ii})$, consistent with the observation that particles are chiefly channelled towards preferential 257 flow pathways. The latter distributions tend to be shifted towards high conductivity values and are 258 characterized by an enhanced mean conductivity value as compared against their generated 259 (unweighted) counterparts (see conductivity mean and weighted mean values in Table 1, and the 260 261 results corresponding to the blue and red circles depicted in Figure 1 a-c). This shift is imprinted onto 262 the probability density function (PDF) of the waiting times and onto its associated TPL parameters (see Figure 2 a-c), consistent with prior studies (Edery et al., 2014; Edery, Geiger, et al., 2016; Edery, 263 2020). We then simulate reactive transport across the collection of generated fields, allowing for 264 precipitation (and/or dissolution) of calcite and assessing the evolution of the conductivity field 265 according to the Kozeny-Carman formulation introduced in Section 2. Conductivity, head, and 266 velocity fields, as well as particle visitations, $n_{ii}(t)$, associated with species h and co are sampled 267 268 across time.

269 After 200 Δt have elapsed (corresponding to a total simulation time of 20 min, i.e., a full pore 270 volume) a set of h particles connecting the inlet to the outlet of the system is clearly visible (see Figure 3 a and b), these particles being non-uniformly distributed in space. Figure 3a and b depict a 271 272 heat map of the *h* particles distribution at time $t = 20 \min$ (i.e., corresponding to the first pore volume), 273 clearly evidencing the emergence of regions of preferential flow (PF). We also note that the number 274 of h particles density (corresponding to concentration) tends to decrease with increasing distance 275 from the inlet, these being replaced by *co* particles, consistent with the observation that they are 276 consumed during the course of the reactive process which induces dissolution of the host solid matrix. The *h* and *co* particles attain equilibrium within cells away from the inlet. As such, reaction can only 277 278 take place if a particle leaves (or enters) a cell under the action of advection and/or diffusion leading 279 to a new equilibrium state. When examining the alteration of conductivities due to the dissolution/precipitation reaction, we note that dissolution (corresponding to an increase of 280 281 permeability values) is primarily tied to the preferential flow pathways. Otherwise, precipitation is 282 seen to take place in regions close (on average) to these pathways. The highest strength of 283 precipitation is observed in the proximity of the preferential pathways, to then decrease with distance 284 from these.

285 Figures 3c and e depict the regions where conductivity has increased (due to dissolution) or 286 decreased (due to precipitation), respectively. The h particles invading the domain closely follow the PFs displaying a fingering pattern, leading to a corresponding dissolution pattern associated with 287 locally increased conductivities. Since conductivity values along the PFs are typically higher (on 288 289 average), dissolution is increasing these conductivities even further, giving rise to a self-sustained 290 enhancing mechanism. The concentration of h particles reaches a local (i.e., within a given cell) 291 equilibrium with the produced co particles. Hence, dissolution will take place where transport induces 292 shifts in concentration that need to be compensated by the dissolution/precipitation process to 293 maintain local equilibrium. Such scenarios can be attained (i) by co particles exiting the preferential 294 flow pathways due to the action of diffusion (i.e., they leave locations where concentration of hparticles is large upon diffusing towards higher pH regions where they precipitate) or (ii) by h295 296 particles traveling through the fast preferential paths and advancing through these. Figure 3g and h 297 display regions with dominating dissolution or precipitation for cells outside and within the PFs, 298 respectively. Here cells associated with PFs are identified upon relying on particle visitations following Edery et al. (2014). Dissolution dominates within the PFs (as indicated by the red cells in 299 300 Figure 3h), because h particles are injected through a flux-weighted boundary condition. On the other hand, the produced co particles do not precipitate at locations corresponding to the high h301 302 concentration residing in the PFs. These may precipitate away from these regions, where they experience low concentrations of h particles. Thus, we observe a reduction of conductivity taking 303 304 place in regions adjacent to the PFs (Figure 3b and g). In summary, our computational results 305 document an increase of conductivity along the preferential pathways jointly with a conductivity 306 reduction within regions close to these and along directions approximately normal to them.

307 Changes of conductivity values ensuing precipitation/dissolution are clearly visible by the broadening of the unweighted log-conductivity distribution $f(y_{ij})$, see Figure 1d-f, and Figure 1g-i 308 309 (blue circles), evaluated at times t = 10 and t = 20 min, respectively. The reaction dynamics leads to 310 a conductivity field characterized by a slightly increased average value, given that our computational 311 analyses entail the injection of an acid fluid into the system (see Table 1 for details). Detailed 312 inspection of Figure 1d and f reveals that precipitation takes place across a slightly larger area than 313 dissolution, i.e. values of the frequency distribution $f(y_{ij})$ associated with low conductivities tend to increase at a larger rate rather than those corresponding to high conductivities (the left tail of the 314 distributions becomes heavier than the right tail with the progress of reaction). The 315 weighted $(f(ny_{ij}))$ and unweighted $(f(y_{ij}))$ distributions (red and blue circles in Figure 1d-f, and 316 Figure 1g-i at time t = 10 and t = 20 min, respectively) are visibly broadening, being associated with 317 an average conductivity which is higher than the one of the originally generated conductivity domains 318 319 (see Table 1). As stated above, dissolution is focused along the preferential pathways, which comprise 320 an area of limited extent with respect to the whole field.

321 The above documented mechanism and its signature on the weighted and unweighted conductivity frequency distributions are sensitive to the initial log-conductivity variance, σ_0^2 . When 322 considering both distributions $f(y_{ij})$ and $f(ny_{ij})$, associated with the case $\sigma_0^2 = 1$, the distributions 323 variance are seen to increase in time, as compared to the values attained at the beginning of the 324 simulation (i.e., prior to reaction; see Figure 1a to d and g, and Table 1). Otherwise, as the initial 325 heterogeneity increases (see, e.g., $\sigma_0^2 = 3,5$) mean and variance associated with the weighted and 326 unweighted conductivity distributions display only minor changes (approximately 10%) across the 327 328 temporal window considered. The conductivity fields characterized by the lowest σ_0^2 value are associated with preferential pathways that are not starkly recognizable when analyzed under non-329 330 reactive transport conditions. These channels become more clearly distinguishable as reactions induce 331 an increase of the conductivities along the PFs. At the same time, precipitation causes a decrease of the conductivity outside the PF. This leads to an increased importance of the left tail of $f(y_{ii})$, 332 333 corresponding to an increase of low conductivity values (see Figure 1, left middle and bottom rows).

334 With reference to the highest conductivity variance analyzed, the reaction patterns for the 335 precipitation and dissolution lead to a smaller relative change between conductivity frequency 336 distributions evaluated prior and after the reaction. Relative changes between the unweighted and 337 weighted conductivity frequency distributions (including the ensuing mean and variance listed in 338 Table 1) evaluated before and after the reaction are less pronounced as the variance of the generated 339 conductivity field increases (see Figure 1 middle and right columns and Table 1). This is related to 340 the observation that, as log-conductivity variance increases, preferential pathways in the originally 341 generated field become markedly more distinct. Thus, relative differences between unweighted and 342 weighted conductivity histograms are seen to diminish in time because the flow field is already organized according to well-identified pathways and tends to preserve its initial pattern (Figure 1d-f, and Figure 1g-i at time t = 10 and t = 20 min, respectively). Note that low order statistics (i.e., mean and variance) of velocity and conductivity display only a minute evolution with the progress of reaction, in spite of the relevant changes exhibited by the tails of the frequency distributions (see Figure 1) for all considered values of σ_0^2 , the latter feature being relevant when addressing non-Fickian transport, as further discussed below.

349 As stated in Section 2.2, the conductivity fields altered through precipitation/dissolution and 350 extracted at regular time intervals of 10 Δt are subject to non-reactive transport analyses and the 351 ensuing evolution of the parameters of the TPL model (6) is analyzed. Key results of these analyses 352 are listed in Table 1 with reference to the original (unaltered) conductivity fields and at the final simulation time (i.e., at time $t = 200 \Delta t$). Analysis of the results associated with transport across the 353 log-conductivity field characterized by the smallest original variance (i.e., $\sigma_0^2 = 1$) and listed in Table 354 1 indicates that the changes of the sample log-conductivity PDF induced by the progress of the 355 356 reaction are reflected by the parameters of the TPL model (6). These transition from estimated values corresponding to an effective Fickian transport regime (corresponding to $\beta = 2$, see also Figure 2a) 357 to values denoting a highly non-Fickian effective transport setting, manifested by the widening of the 358 support of the waiting time PDF $\psi(t_w)$ (see also Figure 2d and 2g, for results obtained at t = 10 and 359 20 min, respectively). Effective transport in the domain with the highest variance (i.e., $\sigma_0^2 = 5$; see 360 Table 1) is characterized by estimated TPL parameters corresponding to a non-Fickian signature also 361 362 prior to the occurrence of precipitation/dissolution (Figure 2c). Such a signature is then further enhanced after reaction has altered the conductivity field, yet displaying a less marked evolution of 363 the shape of the $\psi(t_w)$ as compared to the case $\sigma_0^2 = 1$ (see also Figure 2f and i for t = 10 and 20 min, 364 365 respectively).

The observed temporal changes in conductivity and the ensuing local dynamics of transport 366 367 pattern yield global variations in the reaction rate. Consistent with prior studies and the imposed boundary conditions, the mean velocity associated with the originally generated conductivity domains 368 increases with σ_0^2 . As the reaction progresses and the conductivity fields change, the increased area 369 370 subject to dissolution leads to a slight increase of the mean velocity for all of the σ_0^2 analyzed. To analyze the influence of the preferential flow on the velocity that is affecting particle transport, we 371 consider the average value $\overline{n_R v}$, evaluated upon considering weighting by the number of reactive 372 particles, n_R , visiting each cell (where the term reactive particles denotes both h and co particles 373 374 employed in the context of the reactive transport simulations). The weighted average velocity displays an initial increase over time due to the increase of conductivity within the preferential pathways. 375 376 When considering the relative change across the whole simulation time, values of the temporal increase of $\overline{n_R v}$ are similar across the three heterogeneity levels examined, i.e., they are seemingly 377 independent of σ_0^2 . However, results in Figure 4a also reveal that the average velocity $\overline{n_R v}$ displays 378 distinct temporal histories depending on σ_0^2 . In particular, the value of $\overline{n_R v}$ tends to attain an 379 asymptotic value at time $t \approx 7$ min for $\sigma_0^2 = 5$, while showing a sustained increasing trend for $\sigma_0^2 = 5$ 380 1. This result suggests that the feedback between reaction and flow patterns reaches an asymptotic 381 382 condition faster in systems characterized by higher heterogeneity.

The temporal evolution of the velocity fields due to the precipitation/dissolution reaction and the resulting conductivity changes lead to time-dependent reaction pattern and reaction rates. The Damköhler number is infinite on a local scale in our computational analyses, because the reaction is instantaneous. When considering the entire system, transport processes induce a net overall reaction rate that can be quantified as the sum of the total conductivity changes across time

 $(Sum(|\Delta k_{ij}|)[\frac{cm}{sec}])$. The latter incorporates both positive and negative changes of hydraulic 388 conductivity and therefore quantifies the overall intensity of precipitation and dissolution processes 389 in the domain. The quantity $Sum(|\Delta k_{ij}|)$ is evaluated across all realizations for each of the σ_0^2 values 390 considered and is depicted in Figure 4b as a function of time. These results indicate that the overall 391 392 reaction rate increases in time with a similar rate for all considered values of σ_0^2 (figure 4b) at early times. The observed increase is consistent with the initially advancing of the reaction front across the 393 domain. We observe that the reactive processes magnitude is proportional to σ_0^2 . For low initial levels 394 of heterogeneity, conductivity values along the preferential pathways are closer to the average field 395 396 conductivity than what can be observed for the highly heterogeneous domains. As such, the portion of the domain where precipitation or dissolution can take place increases at a rate proportional to σ_0^2 . 397 As the reaction front reaches the domain outlet, the dissolving front found in the PF leaves the domain. 398 399 Hence, the global variation in conductivity (which is proportional to the magnitude of reactive 400 processes) tends towards an asymptotic value, corresponding to the diffusion-controlled solute 401 exchange along a direction transverse to the preferential pathways (see Figure 3). In agreement with 402 results shown in Figure 4a, this transition towards an asymptotic regime takes place earlier for larger values of σ_0^2 , while a smaller initial heterogeneity implies a longer transient period. 403

4.Conclusions

Our computational study tackles the quantitative characterization of the feedbacks between
 precipitation and dissolution reaction dynamics taking place in randomly heterogenous conductivity
 fields associated with various degrees of spatial heterogeneity. Our work leads to the following key
 conclusions.

- Joint occurrence of precipitation and dissolution is tightly coupled with the existence of preferential flow pathways. Conductivity increase due to the dissolution reaction along such paths leads to enhance particle migration along these. The dominance of preexisting preferential flow regions on the (reactive) transport pattern across the field is therefore further reenforced and self-sustained across time. At the same time, diffusion promotes displacement of particles, leading to precipitation (and hence a progressive reduction over time of local conductivities) at locations in the proximity of these.
- Reactive processes yield an increase over time of the range of conductivity values across the domain, eventually leading to a widening of the support of solute waiting times and conductivity distributions. The clear separation between regions where dissolution or precipitation takes place is reflected in sample distributions of conductivity which tend to become visibly left skewed with time, a feature which is associated with precipitation taking place in low conductivity cells located in the proximity of existing preferential flow pathways.
- 423 Solute mass weighted conductivity and velocity distributions are at the basis of our • characterization of the parameters of a TPL model which is at the core of the characterization 424 of the probability density function of particle travel/waiting times and enables us to capture 425 non-Fickian 426 effective (upscaled) transport behaviors. With the progress of precipitation/dissolution reactions, transport shifts towards an increasingly acute non-Fickian 427 effective behavior (see Figure 2 and ensuing parameter values listed in Table 1). The latter is 428 429 then seen as a direct outcome of the documented feedbacks between transport and reactions 430 taking place in heterogeneous porous media. The evolution of TPL model parameters towards a pronounced non-Fickian behavior is associated with only minor changes of the mean and 431 432 variance of log-conductivity values. This result is consistent with the conceptual picture that

the tails of flux and hydraulic conductivity distributions carry critical information to
characterize transport while displaying only a minor effect on low order statistics associated
with these quantities. Our results suggest that this feature must be acknowledged to properly
characterize transport in the presence of precipitation/dissolution. In this context, we recall
that our study is not aimed at exploring the skill of any specific transport model to interpret
non-reactive transport across the conductivity fields prior and/or after reaction takes place,
this particular analysis being deferred to a future study.

We observe the emergence of an asymptotic regime in highly heterogeneous systems, where 440 • 441 the (averaged) solute velocity attains a constant value even in the presence of reaction. This 442 suggests the occurrence of an equilibrium state between reactive processes and transport 443 under the flow conditions analyzed. This regime is attained because the effects of locally 444 occurring precipitation and dissolution balance each other at the overall scale of the system, 445 so that the ensuing (ensemble-averaged) solute velocity remains unaffected. The time required to attain such an asymptotic state increases with decreasing initial heterogeneity of 446 447 the conductivity field, thus suggesting that pre-asymptotic behaviors may be more relevant 448 in initially (i.e., prior to reaction taking place) homogeneous systems.

449 Our results are based on numerical simulations and may be used to inform upscaling 450 approaches to capture the pre-asymptotic and asymptotic dynamics of reactive transport in 451 heterogeneous systems through simplified models. Future computational studies might also include 452 an assessment of the importance of initial/boundary conditions and/or of solute injection mode on the 453 emergence of non-Fickian transport features, with special emphasis on locations close to the inlet 454 boundary which might then have an impact on the overall solute residence times and related statistics.

455

	$\sigma_0^2 = 1$			$\sigma_0^2 = 3$			$\sigma_{0}^{2} = 5$		
<i>t</i> [min]	0	10	20	0	10	20	0	10	20
\overline{y}	0.15	0.16	0.17	0.45	0.46	0.46	0.66	0.67	0.69
\overline{ny}	0.32	0.33	0.35	0.83	0.84	0.86	1.20	1.19	1.21
$\sigma^2(y_{ij})$	0.97	1.20	1.37	2.80	3.10	3.26	4.60	4.96	5.11
$\sigma^2(ny_{ij})$	0.99	1.21	1.41	2.80	3.07	3.21	4.60	4.85	5.00
$\overline{v_T}$	0.36	0.37	0.38	0.41	0.42	0.43	0.45	0.46	0.47
$\overline{nv_T}$	0.45	0.47	0.50	0.67	0.70	0.72	0.86	0.88	0.90
β	2.00	1.40	1.05	1.70	1.20	0.95	1.40	0.80	0.60
t_1	0.10	0.09	0.06	0.10	0.07	0.03	0.08	0.03	0.01
$\text{Log}_{10}(t_2)$	1.0	2.5	2.8	1.5	2.2	2.9	1.6	2.5	3.0

456

457 Table 1. Values of mean and variance of unweighted and weighted log-conductivity distributions and estimated parameters of the 458 effected TPL model obtained through calibration of (6) against the computed distributions of particle waiting times in the domain

458 effected 1PL model obtained inrolign calibration of (6) against the computed distributions of particle waiting times in the domain 459 cells. Results are listed for the three values of the initial log-conductivity variance (σ_0^2) and are obtained from non-reactive transport 460 simulations performed across conductivity fields resulting from reactive transport simulations at selected times.



463 Figure 1. Relative frequency distributions $f(y_{ij})$ (blue circles) and $f(ny_{ij})$ (red circles) for a tracer test performed on the 464 conductivity field prior to reaction and those associated with reactive simulation times of 10 and 20 minutes. Results correspond to

 $\sigma_0^2 = 1,3,5$ (left, middle and right columns, respectively) and to t = (a-c) 0, (d-f) 10, and (g-i) 20 min. Mean and variance of these 466 distributions are listed in Table 1.



conductively field prior to reaction those associated with reactive simulation times of 10 and 20 minutes. Results correspond to $\sigma_0^2 = 1,3,5$ (left, middle and right columns, respectively) and t = (a-c) 0, (d-f) 10 min, and (g-i) 20 min. Values of TPL parameters

estimated by calibrating model (6) on the sample distributions are listed in Table 1.



474 × × × 475 Figure 3. Heat map representing $(a,b) \log_{10}(n_{Rij})$, i.e., the number of h particles visiting each cell for $\sigma_0^2 = 1$ and 5, respectively, 476 and (c-f) relative change in hydraulic conductivity at time t = 20 min (corresponding to the first pore volume) with respect to the 477 initially generated values for $\sigma_0^2 = 1$ (c and e) and $\sigma_0^2 = 5$ (d and f). Panels c and d display positive changes in conductivity with 478 respect to the initial field, while panels e and f display negative changes in conductivity, both positive and negative changes being 479 represented in log-scale. Results correspond to a selected realizaton of the the log-conductivity fields. The highlighted box 480 illustrates the separation between regions where precipitation or dissolution take place. Panels g and h display cells associated with a net decrease (green) and increase (red) of conductivity for cells outside (g) or within the PF (h), for $\sigma_0^2 = 1$.



483

484 Figure 4. Temporal evolution of (a) the weighted mean velocity $\overline{n_R v}$ (open symbols) and (b) the sum of all conductivity changes over 485 1 minut. Results correspond to $\sigma_0^2 = 1$ (circles), 3 (squares), and 5 (diamond).

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