



1 A Fast-response automated gas equilibrator (FaRAGE) for continuous in situ 2 measurement of methane dissolved in water Shangbin Xiao¹, Liu Liu²*, Wei Wang¹, Andreas Lorke^{1,3}, Jason Woodhouse² and Hans-Peter 3 Grossart²* 4 5 ¹ College of Hydraulic & Environmental Engineering, China Three Gorges University, 443002 Yichang, China 6 ² Department of Experimental Limnology, Leibniz Institute of Freshwater Ecology and Inland Fisheries, 16775 7 Stechlin, Germany 8 ³ Institute for Environmental Sciences, University of Koblenz-Landau, 76829 Landau, Germany 9 *Corresponding authors Emails: liu.liu@igb-berlin.de; grossart@igb-berlin.de 10 11 Abstract 12 Biogenic methane (CH₄) emissions from inland waters contribute substantially to global warming. In aquatic systems, CH₄ dissolved in freshwater lakes and reservoirs is 13 14 highly heterogeneous both in space and time. To better understand the biological and physical processes that affect sources and sinks of CH₄ in lakes and reservoirs, dissolved CH₄ needs to 15 be measured with a highest temporal resolution. To achieve this goal, we developed the Fast-16 Response Automated Gas Equilibrator (FaRAGE) for real-time in situ measurement of 17 18 dissolved CH₄ concentration at the water surface and in the water column. FaRAGE can achieve an exceptionally short response time ($t_{95\%} = 12$ s when including the response time of 19 20 the gas analyzer) while retaining an equilibration ratio of 63% and a measurement accuracy of 0.5%. An equilibration ratio as high as 91.8% can be reached at the cost of a slightly 21 22 increased response time (16 s). The FaRAGE is capable of continuously measuring dissolved

23 CH₄ concentrations in the nM-to-mM ($10^{-9} - 10^{-3}$ mol L⁻¹) range with a detection limit of sub-





nM (10^{-10} mol L⁻¹), when coupled with a cavity ring-down greenhouse gas analyzer (Picarro GasScouter). It enables the possibility of mapping dissolved CH₄ concentration in a "quasi" three-dimensional manner in lakes. The FaRAGE is simple to operate, inexpensive, and suitable for continuous monitoring with a strong tolerance to suspended particles. The easy adaptability to other gas analyzers such as Ultra-portable Los Gatos and stable isotopic gas analyzer (Picarro G2132-i) also provides the potential for many further applications, e.g. measuring dissolved ¹³ &C-CH₄ and CO₂.

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32 1 Introduction

Despite the well-established perception of inland waters as a substantial source of 33 34 atmospheric methane (CH₄) (Bastviken et al., 2011; Cole et al., 2007; Tranvik et al., 2009), 35 large uncertainties remain owing to poorly constrained sources and sinks (Saunois et al., 36 2019). Most freshwater lakes and reservoirs are often oversaturated with CH_4 (relative to atmosphere) and its distribution is characterized by high spatio-temporal heterogeneities 37 (Hofmann, 2013). Point-based and short-term measurements can result in biases in estimating 38 39 diffusive CH₄ flux (Paranaíba et al., 2018). Thus, resolving the spatio-temporal dynamics of dissolved CH₄ concentration in lake water is a prerequisite for better budgeting sources and 40 sinks in freshwater lakes. 41

Methane within lakes is often characterized by pronounced vertical and horizontal 42 43 concentration gradients, which can occur either below or above thermocline. In many deep stratified lakes, a sharp vertical gradient below the thermocline can develop in the anoxic 44 hypolimnion (mM range) (Encinas Fernández et al., 2014; Liu et al., 1996). In contrast, in 45 some stratified lakes with a fully oxygenated hypolimnion CH₄ can accumulate above the 46 thermocline (~µM range) (Grossart et al., 2011; Donis et al., 2017; Günthel et al., 2019). The 47 concentration of dissolved CH_4 is also regulated by loss due to oxidation and emission to the 48 atmosphere (Bastviken et al., 2004; Juutinen et al., 2009). Both rates can be highly variable, 49 particularly for the flux term which is strongly affected by wind and convective mixing (Read 50 51 et al., 2012; Vachon and Prairie, 2013). In addition to the uneven vertical CH₄ distribution, 52 apparent horizontal gradients have been observed in lakes where littoral sediments are identified as a CH₄ source (Murase et al., 2003). This horizontal CH₄ gradient can also 53 54 contribute to the epilimnetic CH₄ peak in pelagic waters via lateral transport (Hofmann et al., 55 2010; Fernández et al., 2016; Murase et al., 2005; Peeters et al., 2019). Nevertheless, dissolved CH₄ in lake water is not only featured with variable spatial patterns, it also changes 56





57 at different time scales as most processes that contribute to the spatial heterogeneity are not

58 always synchronized.

59 The rise and fall of lake CH₄ concentration often show strong seasonality that are 60 driven primarily by thermal stratification (Encinas Fernández et al., 2014) and phytoplankton 61 dynamics (Günthel et al., 2019). While the build-up of hypolimnetic CH_4 storage is a slow 62 process that is closely related to the development of lake hypoxia, the epilimnetic CH_4 maximum can be highly variable even at a daily basis as it is strongly affected by 63 64 phytoplankton dynamics (Günthel et al., 2019; Hartmann et al., 2020; Bižić et al., 2020). In addition, storms can act as another driver for short-term CH4 dynamics in the lake because it 65 often leads to higher evasion rates caused by strong vertical turbulent mixing (Zimmermann et 66 al., 2019) and enhanced horizontal transport (Fernández et al., 2016). While the seasonal 67 patterns of dissolved CH₄ concentration in lake water seem recurrent and can be simulated 68 (Bartosiewicz et al., 2019), the unpredictable effects of short-term phytoplankton dynamics 69 and storm events can present a challenge in modeling lake CH₄ dynamics. 70

71 While there is urgent need for resolving the spatio-temporal variabilities of CH₄ in large water bodies (e.g. deep, stratified lakes), we recognize limitations in the available 72 methodology. Like most gases in dissolved phase, CH₄ cannot be measured directly in water. 73 74 Instead, a carrier gas (synthetic air or at air concentration) is added to achieve (full/partial) gas-water equilibration. The headspace gas sample is then measured with a gas spectrometer 75 and the concentration of targeted gas can be calculated according to Henry's law (Magen et 76 al., 2014). To save sampling effort, continuous gas equilibration devices have been developed, 77 78 which generally can be classified to four categories: 1) Membrane type (Schlüter and Gentz, 2008; Boulart et al., 2010; Gonzalez-Valencia et al., 2014; Hartmann et al., 2018) - gases are 79 80 extracted from water by using a gas-permeable membrane; 2) Marble type (Frankignoulle et 81 al., 2001; Santos et al., 2012) - gas exchange is enhanced by pumping water through marbles





82 that increases the gas-water contact area; 3) Bubble type (Schneider et al., 1992; Körtzinger et al., 1996; Gülzow et al., 2011) - dissolved gases are stripped out by bubbling the water sample; 83 4) Showerhead type (Weiss-type) (Johnson, 1999; Rhee et al., 2009; Li et al., 2015) - water is 84 pumped from top and then mixed with a circulated headspace carrier gas. A full evaluation on 85 the performance of these devices was provided in a recent review (Webb et al., 2016), where, 86 the most important paprameter, response time, was found to vary between 2-34 min for 87 88 dissolved CH₄. While it is already encouraging, improvements are expected to further shorten the response time. 89

Driven by the need to resolve temporal and spatial variabilities of dissolved CH₄ in 90 lakes/reservoirs with sufficient precision, we developed a novel, low-cost equilibrator to 91 92 achieve fast gas-water equilibration. The Fast-Response Automated Gas Equilibrator 93 (FaRAGE) can be coupled with a portable gas analyzer, which makes it perfect for field use. Here, the performance of the FaRAGE is evaluated by investigating its response time, 94 95 detection limit and equilibration ratio. Applications are provided exemplarily to demonstrate 96 the potential of the FaRAGE for improving our understanding on the spatial distribution and temporal dynamics of dissolved CH₄ in inland waters. 97

98 2 Materials and Methods

99 2.1 Device description

The design of the FaRAGE is modified from two types of equilibrators: Bubble-type (Schneider et al., 1992) and Weiss-type (Johnson, 1999). In contrast to the traditional bubbletype and Weiss-type equilibrators that create a large-volume headspace and circulates air back to the headspace, the FaRAGE is a flow-through system that adds gas flow into a constant water flow to produce a minimal headspace for continuous concentration measurement of CH₄ dissolved in water.





106 The operation principle of the FaRAGE is depicted in Fig. 1 and photos of the main 107 parts of the prototype are provided in Fig. S1. A list of information on suppliers and cost of each part can be found in Table S1. A mass flow controller (SIERRA C50L, Netherlands) is 108 used to generate a constant carrier gas (normal air/synthetic air) flow (1 L min⁻¹) from a 109 compressed air tank coupled with a pressure regulator. Water samples are taken continuously 110 using a peristaltic pump (500 mL min⁻¹), and the flow is monitored using a flow meter 111 112 (Brooks Instrument, Germany). The two flows mix in a gas-water mixing unit that is composed of a gas bubble generating unit and a coiled hose for further gas-water turbulent 113 mixing. In the bubble unit (modified from a 10 mL plastic syringe), a jet flow is created by 114 adapting narrowed tubing (2 mm inner diameter) to the water pumping hose (3.2 mm inner 115 diameter). Degassing occurs when the jet flow enters the chamber with a sudden enlarged 116 117 diameter (14 mm). Degassing is further enhanced by micro-bubbles that are generated by a bubble diffusor attached to the carrier gas hose (inside the bubble unit). The gas-water 118 mixture flows through the 2-m long Tygon tube (3.2 mm inner diameter) where additional 119 equilibration occurs. The flow is finally introduced to a gas-water separation unit (a 30 mL 120 plastic syringe) where the headspace gas is separated from the water. In this chamber, water 121 falls down freely to the bottom while the headspace gas is taken directly to a greenhouse gas 122 analyzer (1 L min⁻¹ gas pumping rate; GasScouter G4301, Picarro, USA). A 2-m long Tygon 123 124 tube (3.2 mm inner diameter) is attached to the top of the chamber for venting excess gas flow 125 while stabilizing gas pressure in the headspace. The bottom water is discharged back to the lake using another peristaltic pump (500 mL min⁻¹). To protect the gas analyzer from 126 damaging high water vapor content, a Teflon membrane filter (pore size 0.2 µm) is placed 127 before the gas intake (resulting in a ~210 mL min⁻¹ reduction in flow rate of gas sample, 128 which is vented from the bypass at the top of the gas separation unit). A desiccant (a 20 mL 129 plastic syringe filled with dried silicone beads) is used to reduce moisture concentration to < 130 131 0.1% when attaching to a Picarro G2132-i isotope analyzer (Picarro, USA), in which < 1%





moisture level is required for ${}^{13}\partial C$ -CH₄ measurement. The temperature of the water sample at the point of equilibration with the headspace gas is monitored using a fast thermometer (precision 0.001 °C, 1 Hz, TR-1050, RBR, Canada) attached to the end of water discharging hose.

As concerns might arise from the availability of gas analyzer coupled to the FaRAGE, in addition to Gas Scouter from Picarro, two additional widely used models of greenhouse gas analyzers were tested. They are the Ultraportable Los Gatos (Los Gatos Research, USA) and stable isotopic CH₄ analyzer (G2132-i, Picarro, USA). The main technical details of all three tested gas analyzers are listed in Table S2.



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Fig. 1 Schematic design of the FaRAGE. The components include: Air tank containingcompressed carrier gas (air or synthetic air) with a pressure regulator, a mass flow controller





144 (MFC) for generating constant carrier gas flow, two peristaltic pumps for taking and 145 discharging water, respectively, a flow meter for monitoring water sample flow, a gas-water 146 mixing unit, a gas-water separation unit, a gas analyzer, and a thermometer for measuring water temperature at phase equilibration. A Teflon membrane filter is placed after the MFC 147 and another is added before the gas analyzer to protect from being flooded. A desiccant is 148 used to dry the gas flow to the gas analyzer (if Picarro isotopic analyzer is used). The red 149 150 color marks the flow of carrier gas, dark blue line indicates the water sample, purple line shows the flow of gas-water mixture, the light brown line shows the flow of gas sample (after 151 partial equilibration) and the light blue line depicts the water discharged back to lake. The 152 thickness of the lines scales with the gas/water flow rates. The arrows show the flow 153 154 directions.

155 2.2 Lab validation

The FaRAGE prototype was first tested intensively in the lab to determine both the 156 equilibration ratio and response time. The equilibration ratio is defined as the percentage of 157 the gaseous CH₄ concentrations at the outlet of the gas equilibrator in comparison to the 158 equilibrium concentration (full gas-water equilibration). The equilibration ratio was 159 established by measuring a range of CH₄ stock solutions (nano-to-micro molar dissolved gas 160 161 concentrations). These standard solutions were prepared by adding different amounts of CH₄ into a 200 mL headspace of a 2 L Schott bottle filled with Milli-Q water. The exact dissolved 162 CH₄ concentrations in these solutions were tested with the traditional manual headspace 163 method: a 400 mL headspace was created in a 500 mL plastic syringe with nitrogen gas. The 164 165 CH₄ concentration of the headspace gas was then measured using GasScouter G4301 (Picarro, USA). At the same time, CH_4 concentrations of these standard solutions were measured with 166 167 the FaRAGE for at least 2 min and an average was calculated from more than 60 individual





- data points. For simplicity, we directly compared dissolved CH_4 concentrations measured using the two different methods, i.e., our equilibrator and manual headspace method.
- The response time of the device was investigated by switching the water sample inlet between two water samples with different CH_4 concentrations. Triplicated measurements were performed. An exponential fit was applied to the concentration change curve and the response time was determined as time needed to reach 95% of the final concentration.

The effect of water-to-gas mixing ratio on equilibration ratio and response time of the 174 device was investigated. By fixing the carrier gas flow rate to 1 L min⁻¹, the water-to-gas 175 mixing ratio was varied from 0.04, 0.08, 0.12, 0.15, 0.24, 0.29, 0.36, 0.43 and 0.5 by adjusting 176 177 the water sample flow rate. The effect of tube length on performance of the device was also examined by adapting 1, 2, 4.4 and 8.4 m Tygon tube onto the gas-water mixing unit. For all 178 179 these tests, triplicated measurements of the equilibration ratio and response time were performed corresponding to different mixing ratios and the mean values were used for 180 analysis. 181

Tests were performed to investigate the performance of the device when adapting to two other types of gas analyzers. As the equilibration ratio is unaffected by the model of gas analyzers, only response time was determined. This was done by fixing carrier gas and water sample flow rates to 1 and 0.5 L min⁻¹, respectively. The surplus gas was vented to the air as Ultraportable Los Gatos and Picarro G2132-i have a gas intake flow rate of only 500 and 25 mL min⁻¹, respectively. The effect of desiccant on response time of Picarro G2132-i was checked by measuring gas samples with and without a desiccant installed.

189 2.3 Field tests

190 Two lakes in Germany were chosen for field test. Lake Stechlin is a deep meso-191 oligotrophic lake with a maximum depth of 68 m and Lake Arend is a eutrophic lake with a





maximum depth of 48 m. Pronounced CH_4 peaks in the epilimnion of Lake Stechlin have been previously reported that were measured with various methods (Grossart et al., 2011; Hartmann et al., 2018; Tang et al., 2014). This makes it ideal for our testing purpose. While CH_4 profiles at Lake Arend have never been reported, the metalimnetic oxygen minimum in the lake observed during summer (Kreling et al., 2017) renders it interesting for CH_4 profiling throughout the entire water column.

198 Due to the high potential of the FaRAGE for real-time in situ measurement of 199 dissolved CH₄ concentrations, we explored potential field applications. These field tests included depth profiling of dissolved CH4 concentrations in Lake Arend and Lake Stechlin 200 and investigations of the horizontal distribution of surface dissolved CH₄ concentration across 201 202 the entire Lake Stechlin. For the first application, a fast-response CTD (conductivity, 203 temperature and depth) profiler (XR-620 CTD+, RBR, Canada) was mounted onto a winch with a 30 m long water hose (4 mm inner diameter) attached. The CTD profiler with hose was 204 lowered down continuously at a constant speed (1 m min⁻¹). The exact depth and temperature 205 206 of sampled water can be extracted from the CTD profiler by correcting for the travel time of 207 water sample flow in hose. For the spatial mapping, a GPS antenna (Taoglas, AA.162, USA) was attached to the Picarro gas analyzer. The water intake was submerged 0.5 m below the 208 209 water surface together with the CTD profiler and fixed to one side of the boat. The boat was driven at a constant speed of 5 km h^{-1} . 210

211 **2.4** Theoretical background and data processing

The FaRAGE shares a similar working principle to the Weiss-type gas equilibrator described by Johnson (1999). The theoretical background and equations are provided in S3.

A simplified calculation is described by referring to the manual headspace method. In principle the gas-water mixture is analogous to the static headspace method with the final CH₄





216 concentration in the gas phase assumed to reach a full equilibrium with that dissolved in the 217 aqueous phase. Therefore, by specifying the mixing ratio of air and water, the total mass of 218 CH_4 can be calculated by summing up the CH_4 in the headspace with the dissolved CH_4 (at equilibrium according to Henry's law, which is temperature and pressure dependent) in the 219 aqueous phase and subtracting the mass of background CH₄ (from carrying gas with known 220 concentration). The dissolved CH₄ concentration is then expressed as the volumetric 221 222 concentration of total net mass of CH₄ in the dissolved phase in the given sample volume. A separated exemplary calculation sheet (excel file S5) is provided, which allows for correction 223 for temperature and pressure change. 224

As the equilibration is only partially reached (< 92%), a correction coefficient is needed. This can be obtained by measuring the water samples with known concentrations across a large gradient. By referring to the results measured with the manual headspace method assuming full equilibration (Magen et al., 2014), an equation for precise correction of the measured CH₄ concentrations can be obtained.

230 3 Results and Discussion

231 **3.1 Detection limit, equilibration ratio and response time**

The FaRAGE is capable of achieving a high gas equilibration ratio. We observed a 232 high correlation ($R^2 = 0.999$, p < 0.01) between the concentrations obtained using the 233 traditional headspace method and those measured using the FaRAGE (Fig. 2a) across a wide 234 range of dissolved CH₄ concentrations. The measurement accuracy is 0.5% (standard 235 236 deviation in relation to final concentration) once a stable plateau was reached (Fig. 2b). The FaRAGE reaches a high equilibration ratio (63%) and ensures a rapid response. The 237 238 determined response time $t_{95\%}$ is only 12 s when switching from low-to-high (nano-to-micro 239 molar) dissolved CH₄ concentrations while the $t_{95\%}$ is a little longer (15 ± 2 s) when switching





from high-to-low concentration (Fig. 2b). For the current design specifications that allow for a high equilibration ratio, the detection is theoretically limited by the sensitivity of the coupled gas analyzer. In the lab tests, a clear response was observed at least for CH_4 concentration at air saturation (16.9 nM inside the lab building). The measureable CH_4 concentrations should be at least sub-nM (10^{-10} mol L⁻¹) given the high performance of cavity-ring-down gas analyzers. This is more than sufficient for applications in inland waters where dissolved CH_4 concentrations are often above air saturation.

The response time for the FaRAGE results from two components: 1) the response of the gas analyzer to changes in gas concentration and 2) the physical gas-water exchange process. The response time for the gas analyzer is 5 s when the CH_4 concentration increases (Fig. S2). The FaRAGE itself needs < 10 s to reach 95% of the final steady-state concentration.

Equilibration ratio and response time of the FaRAGE is not sensitive to water-to-gas 252 mixing ratio (Fig. 2c) but rather to the length of the tube in the gas-water mixing unit (Fig. 2d). 253 Little effect was observed on the equilibration ratio in response to the increase of water-to-gas 254 mixing ratio. Also, the increase of water-to-gas mixing ratio did not significantly change the 255 response time of the device (on average 9 s for low-to-high and 13 s for high-to-low, 256 257 respectively). This is in contrast to other types of equilibrators in which an increase of waterto-gas mixing ratio was found to result in a faster response (Webb et al., 2016). However, a 258 sharp enhancement of equilibration ratio was observed due to the extended length of the tube 259 for the gas-water mixing unit. A 91.8% equilibration ratio can be achieved by extending the 260 261 tube length to 13 m while extended response times (low-to-high 17 s and high-to-low 47.5 s, respectively). Further enhancement of the equilibration ratio was not possible when a longer 262 tube (e.g. 18 m) was used. The gas flow rate cannot be stabilized at 1 L min⁻¹ due to the 263 264 increased resistance in response to the further extension of tube length.





265	As shown in Table S2 and Fig. S2, the fast response of the FaRAGE is partly due to
266	the extremely fast response of the Picarro Gas Scouter. This makes it unfair to compare with
267	other equilibrators in which different gas analyzers were used. Tests were performed by
268	adapting the FaRAGE to two other greenhouse gas analyzers (Ultraportable Los Gatos and
269	Picarro G2132-i) and the response times are listed in Table S3. Comparisons were made in
270	Webb et al. (2016) and Hartmann et al. (2018) where both CH ₄ and ¹³ & -CH ₄ were measured
271	using a Picarro G2201-i (Picarro, USA). Here we used a similar Picarro stable isotopic gas
272	analyzer (Picarro G2132-i) and unified all previous reported response time τ to $t_{95\%}$ by
273	applying the equation $t_{95\%} = 3\tau$. The comparison between up-to-date previous studies and this
274	study (Table S4) demonstrated the extraordinary fast response relative to all existing gas
275	equilibration devices. A 53 s response time was achieved when the FaRAGE was adapted to
276	the Picarro G2132-i, which is significantly faster than others (171-6744 s).







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Fig. 2 Performance of the Fast-Response Automated Gas Equilibrator (FaRAGE). (a) 279 280 Exemplary correlation between measurements with the FaRAGE (with a 2-m tube in the gaswater mixing unit) and expected concentrations measured using the manual headspace method. 281 (b) Exemplary response time of FaRAGE for low-to-high and high-to-low concentration 282 changes (with a 2-m tube in the gas-water mixing unit; water-to-gas mixing ratio 0.5). 283 Triplicated tests were performed and averaged response time was taken at the time point when 284 95% of the final concentration was reached. (c) Equilibration ratio and response time in 285 response to changing water/gas mixing ratio (with a 2-m tube in the gas-water mixing unit). 286 Black cross symbols are equilibration ratios, and low-to-high and high-to-low response times 287 are represented by red open and solid squares, respectively. (d) Equilibration ratio and 288





response time in response to changing tube length of gas-water mixing unit (with a fixed water-to-gas mixing ratio of 0.5). Black cross symbols are equilibration ratios, and low-tohigh and high-to-low response times are represented by red open and solid squares, respectively.

3.2 Depth profiles of dissolved CH₄ from multiple lakes

294 Good agreement was observed between depths profiles of dissolved CH₄ concentration measured using two different methods (Fig. 3). The observed occurrence of a maximum in the 295 296 vertical profile of dissolved CH_4 concentration in the upper layer of Lake Stechlin (Fig. 3b) is consistent with previous observations (Grossart et al., 2011; Tang et al., 2014; Hartmann et al., 297 298 2018). In Lake Arend we also observed a CH_4 peak (Fig. 3a), although the overall concentration was lower. In contrast, with the headspace method the FaRAGE allowed for the 299 300 localized CH₄ concentration maximum to be described at a high vertical resolution, similar to that obtained with more sophisticated membrane filter equilibrators (Hartmann et al., 2018; 301 302 Gonzalez-Valencia et al., 2014). The FaRAGE was capable of resolving differences in 303 dissolved CH₄ concentration in lake water at decimeter scales with ease. Whilst care should be taken to ensure the sampling hose moves smoothly and slowly through the water column, 304 continuous profiling of a 20 m deep lake can be completed in 30 min. This is a big advantage 305 306 since in situ CH₄ concentrations can vary at very short time scales (hours to days) subject to 307 internal production, oxidation, weather conditions and etc. (cf. Hartmann et al. (2020)).







Fig. 3 Depth profiles of dissolved CH₄ concentration from two lakes in Germany: (a)
eutrophic Lake Arend on June 17, 2019 and (b) meso-oligotrophic Lake Stechlin on July 23,
2019. Results from the headspace method are designated as red open circles and
measurements using the FaRAGE are shown as solid black dots.

314 **3.3 Resolving spatial variabilities of dissolved CH₄ concentrations**

We confirmed the capability of the FaRAGE to operate continuously over a 7-h period 315 without notable decreases in performance (Fig. 4a). Benefitting from its fast response rate, 316 surface water CH₄ concentrations across the 4.52 km² Lake Stechlin was mapped with great 317 detail within one day. During the cruise, 10 reference measurements were made at different 318 times, which were consistent with nonstop online *in situ* measurements. The cruising survey 319 demonstrated the capability of this device for resolving not just vertical dynamics of CH₄ in 320 lake water, but also the potential for studying horizontal distributions of CH₄ across large 321 distances, for instance large lakes and rivers. With a driving speed of 5 km h⁻¹ and a response 322





- time of 12 s, a spatial resolution of 17 m can be achieved, which is sufficient for such a
- 324 medium-sized lake. The relative higher dissolved CH₄ concentrations in the shallow littoral
- 325 zone of Lake Stechlin (Fig. 4b) reflect higher CH₄ release from the local sediment.



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Fig. 4 Map of surface dissolved CH₄ concentration at Lake Stechlin. (a) Time series of 7-h
continuous surface water CH₄ measurement on March 28, 2019. The reference headspace
measurements are shown as red circles. (b) Spatial distribution of surface water CH₄
concentration is given on top of the lake's bathymetry. Colored symbols show CH₄





- 331 concentration according to the color bar. Black lines show the outline of the lake with depth
- 332 contours.
- 333 4 Comments and Recommendations
- 334 4.1 Adaptability to different gas analyzers

335 The reasons for the significantly shortened response time of the FaRAGE compared to 336 other types of gas equilibrators are two-fold. While the working principle of the FaRAGE is based on the bubble-type (Schneider et al., 1992) and Weiss-type equilibrators (Johnson, 337 338 1999), a reduced headspace volume is adopted, which enhances the physical gas-water 339 exchange. Another reason is the use of extremely fast-response gas analyzer (Picarro Gas 340 Scouter 4301). It is a highly recommended combination for concentration measurement when the best time-wise performance is preferred due to its great mobility (Table S2). However, 341 coupling to other Cavity-Ring-Down gas analyzers is also possible (Table S3). This feature 342 enables a possibility to investigate stable isotopic nature of dissolved CH₄, which is important 343 344 when sources of CH₄ need to be identified.

When a portable gas analyzer (Picarro Gas Scouter or Ultraportable Los Gatos) is used 345 for measuring CH₄ concentration only, the gas equilibrator can be optimized for different 346 application environments. The length of coiled tube for gas-water mixing can be adjusted to 347 change the response time (Fig. 2d). For smaller lakes a higher spatial resolution can be 348 obtained by shortening the equilibration tubing, which shortens the response time, and hence 349 increases the spatial resolution, whilst maintaining an acceptable equilibration ratio (51% 350 351 when tube length is 1 m). In environments with extremely low dissolved CH_4 concentrations, e.g. ocean waters, a longer gas-water mixing tube should be used to ensure a high gas 352 353 equilibration ratio.





354 To measure stable isotopic CH₄ in water, the sensitivity of the FaRAGE can be modified to better adapt to the choice of gas analyzer (e.g., when Picarro G2201-i or G2132-i 355 is used). For example, high dissolved CH₄ concentrations (e.g. µM-to-mM range) can be 356 357 measured with greater accuracy by increasing the flow rate of the carrier gas relative to the sample water flow, therefore diluting the CH_4 concentrations to the range of the gas analyzer. 358 359 This can be particularly useful, for instance, when an instrument has an optimal precision at a low concentration range (1.8-12 ppm for Picarro isotopic gas analyzer) for ¹³&-CH₄ 360 361 measurements. By using pure N₂ gas or carrier gases (e.g. Helium and Argon) and corresponding gas analyzers, it would be possible to measure other dissolved gas 362 concentrations, e.g. CO_2 can be measured simultaneously (CO_2 was tested in this study, but 363 not shown for simplicity). In addition, benefited from the high equilibration ratio of this 364 365 device (max. 91.8%), it would be possible to measure dissolved CH₄ (and other gases) close to equilibrium concentrations. 366

367 4.2 Uncertainties due to suspended solids, temperature and pressure change

The FaRAGE is proven to be resistant to suspended solids in freshwater lakes without 368 having to use additional accessories. As shown in Fig. S3, apparent phytoplankton blooms 369 were observed in the two studied lakes each with a high biomass (Chl-a > 30 μ g L⁻¹) in the 370 epilimnetic water. The measurements were unaffected, without any interruptions during 371 measurements. As algal particles are a large component in freshwater systems, it is safe to 372 373 claim the resistance of this device to suspended solids in such a system. However, care must be taken to avoid the water intake hose hitting the bottom sediment, which could cause 374 375 blockage of the water hose.

The temperature and hydrostatic pressure could both change when water is pumped out through a water hose. To consider the temperature effect, a fast temperature logger is used





(Fig. 1) which allows for corrections in calculation. Instead of using *in situ* lake temperature, the temperature measured at the gas equilibrator should be used where gas equilibration occurs. Our measurements found a minor effect when measuring surface waters but an apparent warming for hypolimnetic water in deep lakes. While a calibration can be done directly by taking water samples from multiple depths of the lake (e.g., Fig. 3) to consider this effect, one could make the calculation without taking many samples by applying temperature correction.

385 The temperature correction can be made by referring to the manual headspace method. The constant gas and water flow can be used as headspace and water volume, respectively. By 386 considering the temperature and pressure effects on gas solubility, the dissolved CH₄ 387 concentrations can be calculated (an example calculation sheet is provided in Table S5). The 388 calibration curve can be established using the manual headspace measurements as standards. 389 The final concentrations can be corrected for partial equilibration by applying the equation 390 391 from the calibration curve (e.g., Fig. 2a). The response time should be deduced when calculating CH₄ depth profiles and spatial distributions, in addition to the time lag caused by 392 pumping water samples by using an extended water hose. 393

394 **4.3 Calibration, maintenance and mobility**

395 The FaRAGE can be readily adopted for measuring other trace gases when coupled 396 with other portable gas analyzers. Due to differences in gas solubility (Duan and Sun, 2003; 397 Wiesenburg and Guinasso Jr, 1979), for each new gas, it would be necessary to establish the relative equilibration efficiency and response time, following the approach we outlined here 398 399 for CH₄. Once set, a new calibration is only required when the tubing diameter or length is changed (when the old one is filthy due to biofilm growth). This can be done by referring to a 400 number of known concentrations that covers a wide range (at least 5), e.g., taking water 401 402 samples from different water depth of the lake or a gradient from littoral to pelagic zones.





Once this full calibration is made, the calibration curve can be used for calculating the 403 404 subsequent measurements. A one-point reference measurement should be performed between depth profiles or transects to check for apparent drifting. This can usually be done by taking 405 one surface water sample from a lake for manual headspace measurement. Care should be 406 taken when measuring in lakes with an anoxic hypolimnion where hydrogen sulfide is likely 407 to accumulate. The performance of Cavity-Ring-Down gas analyzers can be potential affected 408 by organics, ammonia, ethane, ethylene, or sulfur containing compounds (Kohl et al., 2019). 409 At these sites, it is always recommended to take additional samples and measure them with 410 traditional methods (e.g., with a Gas Chromatograph Analyzer). 411

The gas equilibrator should be carefully maintained. Replacement of parts is 412 recommended at a monthly basis provided the device is heavily in use. They include bubble 413 diffusor and the coiled gas-water mixing tube. In addition, to ensure the performance and 414 prevent biofilm formation the gas-water mixing and separation units should be cleaned after 415 use. Running with distilled or Milli-Q water would help to rinse the device and reduce the risk 416 of biofilm development in the inner tubes. The performance of peristaltic pumps should be 417 418 also regularly checked and the inner pump tubes need to be replaced to ensure a constant water flow. 419

The combination of FaRAGE with the Picarro Gas Scouter provides the most mobility. The system can be easily carried by one person and work in a small aluminum or inflatable boat with a maximum capacity of three people is possible. The device can also work in bad weather with additional measures based on protecting the gas analyzer from water damage by rain or flooding.

425 Code availability

426 Not applicable.

21





427 Data Availability

- An example calculation sheet (raw data of Fig. 2a) is provided as part of supporting information for device calibration and for temperature and pressure correction when calculating dissolved methane concentration. The full data sets associated with lab and field tests are available upon request.
- 432 Supplement link
- 433 From Copernicus.

434 Author contributions

SBX and WW proposed the idea and built the first prototype. LL improved the prototype and conducted lab and field tests. JW contributed to the field tests. AL contributed to the derivation of equations; HPG advised the development of the modified prototype. LL drafted the initial manuscript. All authors discussed the results and commented on the manuscript.

440 Competing interests

441 The authors declare that they have no conflict of interest.

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448 **References**





- 449 Bartosiewicz, M., Przytulska, A., Lapierre, J. F., Laurion, I., Lehmann, M. F., and Maranger,
- 450 R.: Hot tops, cold bottoms: Synergistic climate warming and shielding effects increase carbon
- 451 burial in lakes, Limnol. Oceanogr. Lett., 4, 132-144, https://doi.org/10.1002/lol2.10117, 2019.
- 452 Bastviken, D., Cole, J., Pace, M., and Tranvik, L.: Methane emissions from lakes:
- 453 Dependence of lake characteristics, two regional assessments, and a global estimate, Global
- 454 Biogeochem. Cy., 18, https://doi.org/10.1029/2004GB002238, 2004.
- 455 Bastviken, D., Tranvik, L. J., Downing, J. A., Crill, P. M., and Enrich-Prast, A.: Freshwater
- 456 methane emissions offset the continental carbon sink, Science, 331, 50-50,
- 457 https://doi.org/10.1126/science.1196808, 2011.
- 458 Bižić, M., Klintzsch, T., Ionescu, D., Hindiyeh, M. Y., Günthel, M., Muro-Pastor, A. M.,
- 459 Eckert, W., Urich, T., Keppler, F., and Grossart, H.-P.: Aquatic and terrestrial cyanobacteria
- 460 produce methane, Sci. Adv., 6, eaax5343, https://doi.org/10.1126/sciadv.aax5343, 2020.
- 461 Boulart, C., Connelly, D., and Mowlem, M.: Sensors and technologies for in situ dissolved
- 462 methane measurements and their evaluation using Technology Readiness Levels, Trends Anal.
- 463 Chem., 29, 186-195, https://doi.org/10.1016/j.trac.2009.12.001, 2010.
- 464 Cole, J. J., Prairie, Y. T., Caraco, N. F., McDowell, W. H., Tranvik, L. J., Striegl, R. G.,
- 465 Duarte, C. M., Kortelainen, P., Downing, J. A., and Middelburg, J. J.: Plumbing the global
- 466 carbon cycle: integrating inland waters into the terrestrial carbon budget, Ecosystems, 10,
- 467 172-185, https://doi.org/10.1007/s10021-006-9013-8, 2007.
- 468 Donis, D., Flury, S., Stöckli, A., Spangenberg, J. E., Vachon, D., and McGinnis, D. F.: Full-
- 469 scale evaluation of methane production under oxic conditions in a mesotrophic lake, Nat.
- 470 Commun., 8, 1661, https://doi.org/10.1038/s41467-017-01648-4, 2017.
- 471 Duan, Z., and Sun, R.: An improved model calculating CO₂ solubility in pure water and
- 472 aqueous NaCl solutions from 273 to 533 K and from 0 to 2000 bar, Chem. Geol., 193, 257-
- 473 271, https://doi.org/10.1016/S0009-2541(02)00263-2, 2003.





- 474 Encinas Fernández, J., Peeters, F., and Hofmann, H.: Importance of the autumn overturn and
- 475 anoxic conditions in the hypolimnion for the annual methane emissions from a temperate lake,
- 476 Environ. Sci. Technol., 48, 7297-7304, https://doi.org/10.1021/es4056164, 2014.
- 477 Fernández, J. E., Peeters, F., and Hofmann, H.: On the methane paradox: Transport from
- 478 shallow water zones rather than in situ methanogenesis is the major source of CH_4 in the open
- 479 surface water of lakes, J. Geophys. Res.: Biogeosciences, 121, 2717-2726,
- 480 https://doi.org/10.1002/2016JG003586, 2016.
- 481 Frankignoulle, M., Borges, A., and Biondo, R.: A new design of equilibrator to monitor
- 482 carbon dioxide in highly dynamic and turbid environments, Water Res., 35, 1344-1347,
- 483 https://doi.org/10.1016/S0043-1354(00)00369-9, 2001.
- 484 Gonzalez-Valencia, R., Magana-Rodriguez, F., Gerardo-Nieto, O., Sepulveda-Jauregui, A.,
- 485 Martinez-Cruz, K., Walter Anthony, K., Baer, D., and Thalasso, F.: In situ measurement of
- 486 dissolved methane and carbon dioxide in freshwater ecosystems by off-axis integrated cavity
- 487 output spectroscopy, Environ. Sci. Technol., 48, 11421-11428,
- 488 https://doi.org/10.1021/es500987j, 2014.
- 489 Grossart, H.-P., Frindte, K., Dziallas, C., Eckert, W., and Tang, K. W.: Microbial methane
- 490 production in oxygenated water column of an oligotrophic lake, Proc. Natl. Acad. Sci., 108,
- 491 19657-19661, https://doi.org/10.1073/pnas.1110716108, 2011.
- 492 Gülzow, W., Rehder, G., Schneider, B., Deimling, J. S. v., and Sadkowiak, B.: A new method
- 493 for continuous measurement of methane and carbon dioxide in surface waters using off-axis
- 494 integrated cavity output spectroscopy (ICOS): An example from the Baltic Sea, Limnol.
- 495 Oceanogr.: Methods, 9, 176-184, https://doi.org/10.4319/lom.2011.9.176, 2011.
- 496 Günthel, M., Donis, D., Kirillin, G., Ionescu, D., Bizic, M., McGinnis, D. F., Grossart, H.-P.,
- 497 and Tang, K. W.: Contribution of oxic methane production to surface methane emission in
- 498 lakes and its global importance, Nat. Commun., 10, 1-10, https://doi.org/10.1038/s41467-019-
- 499 13320-0, 2019.





- 500 Hartmann, J. F., Gentz, T., Schiller, A., Greule, M., Grossart, H. P., Ionescu, D., Keppler, F.,
- 501 Martinez-Cruz, K., Sepulveda-Jauregui, A., and Isenbeck-Schröter, M.: A fast and sensitive
- 502 method for the continuous in situ determination of dissolved methane and its δ^{13} C-isotope
- ratio in surface waters, Limnol. Oceanogr.: Methods, 16, 273-285,
- 504 https://doi.org/10.1002/lom3.10244, 2018.
- 505 Hartmann, J. F., Gunthel, M., Klintzsch, T., Kirillin, G., Grossart, H.-P., Keppler, F., and
- 506 Isenbeck-Schröter, M.: High Spatio-Temporal Dynamics of Methane Production and
- 507 Emission in Oxic Surface Water, Environ. Sci. Technol., 54, 1451-1463,
- 508 https://doi.org/10.1021/acs.est.9b03182, 2020.
- 509 Hofmann, H., Federwisch, L., and Peeters, F.: Wave-induced release of methane: Littoral
- zones as source of methane in lakes, Limnol. Oceanogr., 55, 1990-2000,
- 511 https://doi.org/10.4319/lo.2010.55.5.1990, 2010.
- 512 Hofmann, H.: Spatiotemporal distribution patterns of dissolved methane in lakes: How
- 513 accurate are the current estimations of the diffusive flux path?, Geophys. Res. Lett., 40, 2779-
- 514 2784, https://doi.org/10.1002/grl.50453, 2013.
- 515 Johnson, J. E.: Evaluation of a seawater equilibrator for shipboard analysis of dissolved
- 516 oceanic trace gases, Anal. Chim. Acta, 395, 119-132, https://doi.org/10.1016/S0003-
- 517 2670(99)00361-X, 1999.
- 518 Juutinen, S., Rantakari, M., Kortelainen, P., Huttunen, J. T., Larmola, T., Alm, J., Silvola, J.,
- 519 and Martikainen, P. J.: Methane dynamics in different boreal lake types, Biogeosciences, 6,
- 520 209–223, https://doi.org/10.5194/bg-6-209-2009, 2009.
- 521 Kohl, L., Koskinen, M., Rissanen, K., Haikarainen, I., Polvinen, T., Hellén, H., and Pihlatie,
- 522 M.: Interferences of volatile organic compounds (VOCs) on methane concentration
- 523 measurements, Biogeosciences, 16, 3319-3332, https://doi.org/10.5194/bg-16-3319-2019,
- 524 2019.





- 525 Körtzinger, A., Thomas, H., Schneider, B., Gronau, N., Mintrop, L., and Duinker, J. C.: At-
- sea intercomparison of two newly designed underway pCO2 systems encouraging results,
- 527 Mar. Chem., 52, 133-145, https://doi.org/10.1016/0304-4203(95)00083-6, 1996.
- 528 Kreling, J., Bravidor, J., Engelhardt, C., Hupfer, M., Koschorreck, M., and Lorke, A.: The
- 529 importance of physical transport and oxygen consumption for the development of a
- metalimnetic oxygen minimum in a lake, Limnol. Oceanogr., 62, 348-363,
- 531 https://doi.org/10.1002/lno.10430, 2017.
- 532 Li, Y., Zhan, L., Zhang, J., and Chen, L.: Equilibrator-based measurements of dissolved
- 533 methane in the surface ocean using an integrated cavity output laser absorption spectrometer,
- 534 Acta Oceanol. Sin., 34, 34-41, https://doi.org/10.1007/s13131-015-0685-9, 2015.
- 535 Liu, R., Hofmann, A., Gülaçar, F. O., Favarger, P.-Y., and Dominik, J.: Methane
- 536 concentration profiles in a lake with a permanently anoxic hypolimnion (Lake Lugano,
- 537 Switzerland-Italy), Chem. Geol., 133, 201-209, https://doi.org/10.1016/S0009-
- 538 2541(96)00090-3, 1996.
- 539 Magen, C., Lapham, L. L., Pohlman, J. W., Marshall, K., Bosman, S., Casso, M., and
- 540 Chanton, J. P.: A simple headspace equilibration method for measuring dissolved methane,
- 541 Limnol. Oceanogr.: Methods, 12, 637-650, https://doi.org/10.4319/lom.2014.12.637, 2014.
- 542 Murase, J., Sakai, Y., Sugimoto, A., Okubo, K., and Sakamoto, M.: Sources of dissolved
- 543 methane in Lake Biwa, Limnology, 4, 91-99, https://doi.org/10.1007/s10201-003-0095-0,
- 544 2003.
- 545 Murase, J., Sakai, Y., Kametani, A., and Sugimoto, A.: Dynamics of methane in mesotrophic
- 546 Lake Biwa, Japan, Ecol. Res. 20, 377-385, https://doi.org/10.1007/s11284-005-0053-x, 2005.
- 547 Paranaíba, J. R., Barros, N., Mendonça, R., Linkhorst, A., Isidorova, A., Roland, F. b.,
- 548 Almeida, R. M., and Sobek, S.: Spatially resolved measurements of CO₂ and CH₄
- 549 concentration and gas-exchange velocity highly influence carbon-emission estimates of
- 550 reservoirs, Environ. Sci. Technol., 52, 607-615, https://doi.org/10.1021/acs.est.7b05138, 2018.





- 551 Peeters, F., Fernandez, J. E., and Hofmann, H.: Sediment fluxes rather than oxic
- 552 methanogenesis explain diffusive CH₄ emissions from lakes and reservoirs, Sci. Rep., 9,
- 553 https://doi.org/10.1038/s41598-018-36530-w, 2019.
- 554 Read, J. S., Hamilton, D. P., Desai, A. R., Rose, K. C., MacIntyre, S., Lenters, J. D., Smyth, R.
- 555 L., Hanson, P. C., Cole, J. J., and Staehr, P. A.: Lake-size dependency of wind shear and
- convection as controls on gas exchange, Geophys. Res. Lett., 39,
- 557 https://doi.org/10.1029/2012GL051886, 2012.
- 558 Rhee, T., Kettle, A., and Andreae, M.: Methane and nitrous oxide emissions from the ocean:
- 559 A reassessment using basin-wide observations in the Atlantic, J. Geophys. Res.: Atmospheres,
- 560 114, https://doi.org/10.1029/2008JD011662, 2009.
- 561 Santos, I. R., Maher, D. T., and Eyre, B. D.: Coupling automated radon and carbon dioxide
- measurements in coastal waters, Environ. Sci. Technol., 46, 7685-7691,
- 563 https://doi.org/10.1021/es301961b, 2012.
- 564 Saunois, M., Stavert, A. R., Poulter, B., et al.: The Global Methane Budget 2000–2017, Earth
- 565 Syst. Sci. Data Discuss., https://doi.org/10.5194/essd-2019-128, in review, 2019.
- 566 Schlüter, M., and Gentz, T.: Application of membrane inlet mass spectrometry for online and
- 567 in situ analysis of methane in aquatic environments, J. Am. Soc. Mass Spectrom., 19, 1395-
- 568 1402, https://doi.org/10.1016/j.jasms.2008.07.021, 2008.
- 569 Schneider, B., Kremling, K., and Duinker, J. C.: CO₂ partial pressure in Northeast Atlantic
- 570 and adjacent shelf waters: Processes and seasonal variability, J. Marine Syst., 3, 453-463,
- 571 https://doi.org/10.1016/0924-7963(92)90016-2, 1992.
- 572 Tang, K. W., McGinnis, D. F., Frindte, K., Brüchert, V., and Grossart, H.-P.: Paradox
- 573 reconsidered: Methane oversaturation in well-oxygenated lake waters, Limnol. Oceanogr., 59,
- 574 275-284, https://doi.org/10.4319/lo.2014.59.1.0275, 2014.
- 575 Tranvik, L. J., Downing, J. A., Cotner, J. B., Loiselle, S. A., Striegl, R. G., Ballatore, T. J.,
- 576 Dillon, P., Finlay, K., Fortino, K., and Knoll, L. B.: Lakes and reservoirs as regulators of





- carbon cycling and climate, Limnol. Oceanogr., 54, 2298-2314,
- 578 https://doi.org/10.4319/lo.2009.54.6_part_2.2298, 2009.
- 579 Vachon, D., and Prairie, Y. T.: The ecosystem size and shape dependence of gas transfer
- velocity versus wind speed relationships in lakes, Can. J. Fish. Aquat. Sci., 70, 1757-1764,
- 581 https://doi.org/10.1139/cjfas-2013-0241, 2013.
- 582 Webb, J. R., Maher, D. T., and Santos, I. R.: Automated, in situ measurements of dissolved
- 583 CO₂, CH₄, and δ^{13} C values using cavity enhanced laser absorption spectrometry: Comparing
- response times of air-water equilibrators, Limnol. Oceanogr.: Methods, 14, 323-337,
- 585 https://doi.org/10.1002/lom3.10092, 2016.
- 586 Wiesenburg, D. A., and Guinasso Jr, N. L.: Equilibrium solubilities of methane, carbon
- 587 monoxide, and hydrogen in water and sea water, J. Chem. Eng. Data, 24, 356-360,
- 588 https://doi.org/10.1021/je60083a006, 1979.
- 589 M. Zimmermann, M. J. Mayr, D. Bouffard, W. Eugster, T. Steinsberger, B. Wehrli, A. Brand,
- and Bürgmann H.: Lake overturn as a key driver for methane oxidation, bioRxiv,
- 591 https://doi.org/10.1101/689182, 2019.