



# Projection of irrigation water demand based on the simulation of synthetic crop coefficients and climate change

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**Abstract.** In the context of major changes (climate, demography, economy, etc.), the Southern Mediterranean area faces serious challenges with intrinsically low, irregular, and continuously decreasing water resources. In some regions, the proper dynamic of irrigated areas is very important so that it is needed to include it in future scenarios. A method for estimating the future evolution of irrigation water requirements is proposed and tested in the Tensift watershed, Morocco. Monthly synthetic crop coefficients ( $K_c$ ) of the different irrigated areas were obtained from a time series of remote sensing observations. An empirical model using the synthetic  $K_c$  and rainfall was developed and fitted to the actual data. The model appears to be reliable with an average  $r^2$  of 0.69 for the observation period (2000-2016). The sub-sampling tests suggested that a loss of performance ( $r^2=0.45$ ) is to be expected for two time periods after the observations (2050). This flexible system of equations has been used to reinterpret a local water management plan and to incorporate two downscaled climate change scenarios (RCP4.5 and RCP8.5). The examination of irrigation water requirements until 2050 revealed that the difference between the two climate scenarios was very small (<2%), while the two agricultural scenarios were strongly contrasted both spatially and in terms of their impact on water resources. The approach is generic and can be refined by incorporating irrigation efficiencies.

## 25 1 Introduction

Water resources are scarce in semi-arid areas and a major part is allocated to agriculture. In the south-Mediterranean region, irrigation allocation to agriculture represents 80% of total water abstraction. It varies from 46% in eastern countries up to 88% in Morocco in 2010 (FAO, 2016). This percentage has been decreasing in most south-Mediterranean countries during the last decades in particular due to the limitation of available resources and the increase of the urban water demand. In parallel, the pressure on water resources led to what Margat and Vallée (2000) called a "post-dam era" or what Molle et al. (2019) called a "groundwater rush" where subterranean water is overused to satisfy the growing water demand. In recent



years, overexploitation of groundwater has been facilitated by technological inventions, affordable cost of exploitation and weak monitoring by authorities (MED-EUWI working group on groundwater, 2007). In regions where the dynamics of irrigated areas is strong, it is necessary to make projections of the irrigation water demand with the actual trend in order to build alternative scenarios. In their review, March et al. (2012) defined scenario analysis as *internally consistent stories about ways that a specific system might evolve in the future. Scenarios are plausible accounts of the future rather than forecasts*. Narrative scenarios also help identify the drivers of change and the implications of current trajectories as well as the options for action (Raskin et al., 1998). Scenarios are therefore halfway between facts and speculations in terms of complexity and uncertainty (van Dijk, 2012). They are commonly used as a management tool for strategic planning and for helping managers strengthen decision making.

Various studies (Arshad et al., 2019; Lee and Huang, 2014; Maeda et al., 2011; Schmidt and Zinkernagel, 2017; Tanasijevic et al., 2014; Wang et al., 2016) have addressed the estimation of irrigation water requirements (IWR) of a region with an approach close to equation 1. A simplified balance between crop evapotranspiration ( $ET_c$ ) and effective precipitation ( $P_e$ ) is computed for each crop  $i$  and multiplied by the corresponding irrigated area.  $ET_c$  can be inferred from the crop coefficient method ( $K_c$ , Allen et al., 1998). Most of the effects of the various weather conditions are incorporated into  $ET_0$  which accounts for the water demand of a reference crop, while the  $K_c$  mainly accounts for crop characteristics.  $K_c$  varies according to four main characteristics: the crop height, the albedo of the crop and soil, the canopy resistance of the crop to vapor transfer (leaf area, leaf age and condition, and the degree of stomatal control) and the evaporation from soil, especially the soil exposed to solar radiations. Furthermore, as the crop develops, those different characteristics change during the various crop phenological stages.  $K_c$  values and stage lengths, for typical climate conditions, are considered to be well known for many crops and have been compiled in the FAO look-up tables.

$$IWR = \sum_{i=1}^n (ET_{c_i} - P_e) * Area_i = \sum_{i=1}^n [(K_{c_i} * ET_0) - P_e] * Area_i \quad (1)$$

This approach is also a very convenient way to account for both climate and crop changes. On the one hand,  $P_e$  and  $ET_0$  are obtained from meteorology or climatology. On the other hand,  $K_c$  values and stages lengths are taken from the tables. Most of the work consists of evaluating the future irrigated area of each crop and assessing the impact of more efficient irrigation techniques. Moreover, there is a significant amount of literature about the estimation of land use and land cover changes (Mallampalli et al., 2016; Noszczyk, 2018), with various techniques to estimate or predict them. Many land cover change approaches are based on transition probability which was introduced by (Bell, 1974) and have been eventually connected to Cellular Automata to account for geographical interrelationships (Houet et al., 2016; Marshall and Randhir, 2008). A very interesting technique has been to combine the top-down (demand-driven) and bottom-up (local conversion) processes of land cover change by proceeding to a simplification of local processes (van Asselen and Verburg, 2013; Verburg and Overmars, 2009). However, land cover change does not account for the intensification of cropping patterns.



In the present work, the hypothesis is that the dynamics of land-use change and intensity of the cropping patterns (hence irrigation water demand) can be reduced to a synthetic monthly  $K_c$  time series for every irrigated area of the studied region. Each synthetic time series would then account for the spatial variability of cropping patterns inside each irrigated area.

65 Unless no sudden change occurs, a curve fitting taking into account the trajectory and the actual rainfall should give an accurate fit that would allow extrapolating into the next few decades. The synthesis would also decrease substantially the amount of information compared to a classic approach of land cover, so that some information, like the amount of tree crops, should be retrieved back from the time series. This approach would prevent the need for tedious land cover classifications and reduce the difficulty of working with discrete values for developing future scenarios.

70 The objective of this study is to produce two scenarios of irrigation water demand for the different irrigated areas of the Tensift watershed in Morocco. One is the trend, the other one is an alternative scenario derived from a narrative scenario. To do so, a method for simulating and extrapolating  $K_c$  is proposed and is tested for the 2000-2016 period. The future time series of  $K_c$  is used with climate change scenarios to obtain an estimate of future irrigation demands. As the strategy of irrigation for tree crops is treated very differently from annual crops in the Tensift watershed, the yearly percentage of tree

75 crops is also retrieved from the  $K_c$  time series.

The content of the article is separated into four parts and a conclusion. The first part describes the study area and the dataset. The methodology is then detailed in three parts: 1)  $K_c$  time series are obtained from remote sensing observations for the observation period, 2) a model is constructed to project the IWR based on simulating the evolution of  $K_c$  and integrating climate change 3) an alternative scenario is built from the trend to reflect a narrative scenario built by local water managers.

80 In the results section, the performance of the model and its ability to make projections are analyzed. The alternative scenario is compared to the simulated trend. In the discussion, other paths are discussed, in particular those which are related to irrigation management.

## 2 Study Site and Data

### 2.1 Study Area: the Haouz Plain in Morocco

85 The study area includes the Haouz Plain, underlain by a large Plio-Quaternary alluvial aquifer. It is located within the Tensift watershed in southern Morocco around the city of Marrakech (Figure 3). The plain is bordered at the south by the High Atlas range. The High-Atlasian watersheds that peak at 4167 masl receive about 600 mm/year of precipitation of which a major portion falls as snow; 25% of the streamflow is generated by snowmelt (Boudhar et al., 2009). The Plain, which is located at altitudes ranging between 480 and 600 masl, has a semi-arid climate (rainfall of 250 mm/year and potential

90 evapotranspiration of 1600 mm/year) with mild, wet winters and very hot and dry summers. Irrigation is essential for good crop development and yield. The so-called “traditional” areas (purple) which originally received water from the Atlasian wadis are more and more dependent on groundwater. The areas irrigated from lake reservoirs (green) receive water from the Lalla Takerkoust reservoir, the Moulay Youssef reservoir and also from the Hassan 1er reservoirs through a 140km canal



called the Rocade Canal. The rest of the irrigation water comes from the groundwater. The increase in the irrigated area and  
95 the intensification of irrigation during recent decades have caused a long-lasting fall in the groundwater table (Boukhari et  
al., 2015). A multi-model analysis of the area (Fakir et al., 2015) has shown that the groundwater table falls from 1 to 3  
m/year and that the mean annual groundwater deficit (about 100 hm<sup>3</sup> since 2000) is equivalent to 50% of the reserves lost  
during the previous 40 years. Among the main causes of this depletion, we find a reduction and higher irregularity of  
precipitation (Marchane et al., 2017) for crop growing and groundwater recharge, a reduction of snow water storage  
100 (Marchane et al., 2015), an increase and intensification of irrigated areas, a progressive conversion to arboriculture due to  
national plans. Since irrigation relies increasingly upon groundwater abstraction, questions are inevitably raised concerning  
the future of local agriculture and of the groundwater.

## 2.2 Data

In order to compute the crop-water demand, Regional Climate Model (RCM) simulations over the Mediterranean region  
105 from the Med-CORDEX initiative (<https://www.medcordex.eu>, Ruti et al., 2016) were utilized to provide future values of  
near-surface atmospheric variables, namely the 2m air temperature and total precipitation. RCM results from two contrasting  
greenhouse-gas concentration RCP (Representative Concentration Pathway) trajectory scenarios were selected: the RCP4.5  
scenario, which represents the optimistic scenario for CO<sub>2</sub> emissions (a stabilization of emissions) and the RCP8.5  
(increasing emissions). Simulations from Centro Euro-Mediterraneo sui Cambiamenti Climatici (CMCC), Météo-France  
110 (CNRM) and IPSL-Laboratoire de Météorologie Dynamique (LMD) were used in the current study. The three RCMs give a  
good representation of the spread typically found in such simulations. The so-called "delta-change" or perturbation method  
(Anandhi et al., 2011) was used to downscale the RCM data using the Global Soil Wetness Project Phase 3 forcings  
(GSWP3: <http://hydro.iis.u-tokyo.ac.jp/GSWP3>) as the reference historical data. The GSWP3 data were evaluated using  
station observations of air temperature, relative humidity, and precipitation from the surface network ( $T^\circ$  daily  $r = 0.88$ ,  
115  $\text{bias}(C^\circ) = 0.73$ , Precipitation daily  $r = 0.09$ ,  $\text{bias}(\%) = -23$ ). The 0.5° data were resampled to a resolution of 1 km, applying  
a correction of altitude between the GSWP3 geopotential and a digital elevation model (GTOPO30 available at  
<https://www.usgs.gov>) for air and dew point temperatures. The meteorological dataset extends from 2000 to 2050.  
For generating vegetation time series, we used the temporal 16-day composite series of MODIS NDVI (MOD13Q1,  
Collection 6, K. Didan, 2015), at a spatial resolution of 250 m. This product is computed from atmosphere-corrected (Justice  
120 et al., 2002), daily bidirectional surface reflectance observations, using a compositing technique based on product quality  
(Wan et al., 2015). We used the 16-day composites to reduce cloud coverage. The data has been compiled for the period  
2000-2016.



### 3. Methodology

#### 3.1 $K_c$ assessment from Remote Sensing Time Series

125 Neale et al. (1990) proposed to estimate the  $K_c$  coefficient from vegetation indices obtained from remote-sensing imagery. In this work, we have been using empirical linear equations where the slope (a) and intercept (b) have been previously calibrated for common crop type in local field experiments (Duchemin et al., 2006; Er-Raki et al., 2007, 2008):

$$K_{c_{sat}} = a * NDVI + b \quad (2)$$

Equation (2) accounts for combined Evaporation from the soil and Transpiration from the crop in a very simple way, but as no water balance is performed, several assumptions must be kept in mind: 1) the extra evaporation due to wetting events (rainfall, irrigation) is not computed, but is assumed to be represented in the calibrated NDVI/ $K_{c_{sat}}$  relationship, which is related to the precipitation of the year used for calibration. The inability of this method to increase the evaporation part of evapotranspiration when frequent small rainfalls events occur should be assessed and corrected for, either at the level of the  $ET_c$  computation or in the  $P_e$  computation; 2) the lower evaporation due to a reduction of the wetted fraction characteristic of micro-irrigation is not taken into account; 3) The crop water stress is not taken into account in this calculation, as the time step of one month causes this to be quite complex to represent. However, significant stress may impact the evolution of the NDVI signal. The approach of estimating  $K_{c_{sat}}$  is displayed in the light-gray area of Figure 2.

#### 3.2 Trend projection of $K_c$

The simulation of  $K_c$  was carried out as follows: A linear adjustment from the original  $K_{c_{sat}}$  is obtained from remote sensing, secondly, this first guess is corrected according to the rainfall of the year, then finally a correction is made according to the percentage of tree crops. Figure 3 illustrates the quantities used to perform this calculation. The four steps detailed below were reproduced for each of the 29 irrigation areas of the study.

(1) The first step was to perform a linear adjustment (eq. 3). For each month  $M$ , of the year  $y$ , a linear least-squares fit was made to the  $K_{c_{sat}}$  values for all years in the series, thus we obtained a set of 12 monthly regression equations for  $K_{c_{sat}}$  as a function of the year. These twelve linear regressions thus formed the time series  $K_{c_{lin}}$ . An example is plotted in blue in Fig. 3. :

$$K_{c_{linM}}(y) = a1_M \cdot y + b1_M \quad (a1 \text{ and } b1 \text{ are fitted to } K_{c_{satM}}(y) \text{ for } M \in [1,12]) \quad (3)$$

(2) The second step consisted of correcting the first approximation of  $K_{c_{lin}}$  for the impact of rainfall. In this semi-arid region, sowing and the development of the vegetation depend very much on the time distribution and accumulation of rainfall during the agricultural season. We chose March as the month most representative of these interannual differences (Le Page and Zribi, 2019). On the one hand, we calculated the difference in  $K_c$  ( $\Delta K_c$ ) between the observation ( $K_{c_{sat}}$ ) and the linear regression ( $K_{c_{lin}}$ ) in the month of March as in Eq. (4). On the other hand, the rainfall accumulation  $P_y$  was summed between



September and March according to Eq. (5). This has been done for each year from 2000 to 2016. We found that the relationship between  $\Delta K_c$  and  $P_y$  was best modeled using a second-degree polynomial according to Eq. (6). An example of this fit is given on the right side of Fig.3.

$$\Delta Kc(y) = Kc_{lin}(y) - Kc_{sat}(y) \quad (4)$$

$$P_y = \sum_{m=sep}^{mar} P_m \quad (5)$$

$$Kc_{cor}(P_y) = a2.P_y^2 + b2.P_y + c2 \quad (a2, b2, c2 \text{ are fitted to } \Delta Kc_M \text{ for } P_y) \quad (6)$$

(3) In the third stage, the tree crops and their preservation were taken into account. In fact, in semi-arid Mediterranean regions, evergreen tree orchards (olive, citrus) are the major crops responsible for green vegetation in the dry season of summer, although the cultivation of summer vegetables is also possible. In the Tensift region, orchards come first in order of preference for water allocation in agriculture, especially in dry years. The aim is to preserve the tree crops from drought. To ensure that the tree crop area would not be reduced from one year to the next, the minimum  $Kc_{sim}$  was forced to not run in the opposite direction to the trend slope (it won't be reduced when the trend is growing and inversely). In other words, the yearly rate of minimum  $K_c$  is growing, a new value is ensured to be greater or equal to the minimum value of the previous year. Finally, to ensure that the resulting synthetic  $K_c$  would be in line with the proportion of trees in the irrigated area, the percentage of tree crop has been obtained during the dry season when it is assumed that there is no other crop than trees:

$$\%Trees(y) = \frac{minKc(y)}{Kc_{Trees}} \quad (7)$$

where  $minKc$  is the minimum  $Kc_{sim}$  for the hydrological season (see the blue line on Fig. 3) and  $Kc_{Trees}$  is the maximum  $K_c$  for the tree crop considered (see the orange line on Fig. 3). For the study area, in which olive trees are the dominant tree crop,  $Kc_{Trees}$  has been set to 0.55 (Allen et al., 1998). The potential maximum  $K_c$  of each year ( $maxKc$ ) was computed with the weighted sum as follows:

$$maxKc(y) = Kc_{max}.(1 - \%Trees(y)) + Kc_{Trees}.\%Trees(y) \quad (8)$$

where the non-tree crop area has a maximum  $K_c$  of  $Kc_{max}$ , which was set according to winter wheat as in Allen et al. (1998). Winter wheat is a dominant crop of the study area, and its maximum  $K_c$  (1.15) is among the highest known.

(4) The final estimate of monthly  $K_c$  ( $Kc_{sim}$ ) integrated the linear trend, the yearly variation due to rainfall and the limitation due to the coverage of tree crops (eq. 9).

$$Kc_{sim_M}(y, P) = \max[\min(Kc_{lin_M}(y) + Kc_{cor}(P), maxKc(y)), Kc_{min}] \quad (9)$$

where the variability of  $K_c$  due to yearly rainfall ( $Kc_{cor}$ ) was added to the first guess  $Kc_{lin}$ . It was ensured that this sum did not exceed the potential maximum  $K_c$  ( $maxKc$ , Eq. 8). It was also ensured that it did not go below the  $K_c$  of bare soil ( $Kc_{min} = 0$ ).



The set of equations 3 to 9 together model the future trend of  $K_c$ . The water demand was then computed from Equation 1, where  $ET_0$  incorporated the weather conditions, and in particular the modifications of temperature predicted by the GCMs.

180 The performance of the model has been assessed using different sampling techniques to obtain calibration and validation data sets. First, calibration years are selected to be equally spaced along the time axis: one out of every two (8 years), one out of every three (6 years) and one out of every four (4 years) of calibration. For the ‘every two years’ sampling, there were two sets of calibration years, for the ‘every three years’ sampling, there were three calibrations sets, and so on. The three different versions of the model were run for the different combinations of calibration and validation data sets. The versions

185 tested were the linear fit ( $K_{c_{lin}}$ ), the linear fit with rain correction ( $K_{c_{cor}}$ ) and then with tree correction ( $K_{c_{sim}}$ ). The performance indicators were averaged for each group (1/2, 1/3, 1/4). The years have also been separated into two groups: years with less precipitation are called the ‘dry years’ (2001, 02, 05, 07, 08, 11, 12, 14, 16) and the rest being called ‘wet’ years. Three statistical metrics have been computed: the determination coefficient  $r^2$  (Eq. 10), the root mean squared error RMSE (Eq. 11), and the standard deviation of error  $stderr$  (Eq. 12):

$$r^2 = \sqrt{\frac{Cov(X, Y)}{\sigma_X \sigma_Y}} \quad (10)$$

$$RMSE = \sqrt{\frac{\sum_1^n (X - Y)^2}{n}} \quad (11)$$

$$stderr = \sqrt{\frac{\sum_1^n (X - Y)}{n}} \quad (12)$$

### 190 3.3 Alternative evolution of $K_c$

The set of equations 3 to 9 were used as a basis for simulating the long-term trend of  $K_c$ . The coefficients  $a_1$  and  $b_1$  of Eq. 3 were fitted with the data from the entire time series. The parameters  $a_2$ ,  $b_2$  and  $c_2$  of Eq. 6 were fitted using the rainfall data taken from the downscaled climate scenario.

The system of equations allows the consideration of various possibilities for “bending the curve” (Raskin et al., 1998), which

195 here means modifying the trend of the cropping scenario: (1) the parameters  $a_1$  and  $b_1$  in Eq. 3 can be modified to account for an overall change in the trend; (2) the coefficients  $a_2$ ,  $b_2$ , and  $c_2$  in Eq. 6 can be modified to account for the influence of cumulative rainfall. (3) The hypothetical law that determines that the rate of change in tree coverage always has the same sign could be relaxed so that a decrease of tree coverage could be possible.

As the status of the Haouz aquifer is becoming critical, alternative trend scenarios have been proposed by the Tensift Basin

200 Agency under the AGIRE Convention developed with the support of GIZ (*Gesellschaft für Internationale Zusammenarbeit*) (ABHT GROUP AG - RESING, 2017). It lists several measures designed to curb the increasing trend in groundwater abstraction. Here, we proposed to only modify the main trend of the  $K_c$  curve in order to represent the idea of a pause in crop-cover expansion. The set of equations to simulate the bending of the curve are:





$$\begin{aligned}a1' &= a1 \cdot b_c \\b1' &= (a1 \cdot b_c + b1) - y_c \cdot a1' \\Kc'_{sim} &= a1' \cdot Kc_{sim} + b1'\end{aligned}\tag{13}$$

205 A new slope ( $a1'$ ) and intercept ( $b1'$ ) were computed according to a bending coefficient ( $b_c$ ) and applied for the desired years were  $y_c$  is the beginning year of the bending. The bending coefficient was expressed as a percentage.

It was determined that the introduction of two bending points to the trend scenario (Eq. 13) gave a good representation of the vision of the narrative scenario (-50% on the trend in 2020 and another -50% by 2040). Note that the narrative scenario didn't give a detailed view of the changes so that those coefficients have been applied uniformly to the 29 irrigated areas.

## 4 Results

### 210 4.1 Performance of the simulated $Kc_{sim}$

Figure 4 shows the calibrated and original time series of  $K_c$  for some selected irrigated areas. In these graphs, the calibration has been carried out over the entire time series. The Chichaoua Private area gave the worst result. As can be seen, the low  $r^2$  was mainly due to the poor distribution of the data at very low  $K_c$ . It is very likely that most of the area was not irrigated continuously during the whole study period. The N'Fis-RD is located on the outskirts of the city of Marrakech, under high  
215 urban pressure. The negative trend was well predicted and the correlation score was satisfactory ( $r^2=0.77$ ) despite the probably erroneous pause in the original  $K_c$  data in 2006 and 2007. The Ourika traditional area is located at the outlet of the Ourika River. The scores were average ( $r^2 = 0.646$ , Slope = 0.747). The general growing trend was reproduced, especially for the base of the curve. The maximum and minimum  $K_c$  simulated with the amount of annual rainfall was generally accurate, except for the two years 2002 and 2003. Those are extremely dry years where rainfed cereals couldn't develop and  
220 there were strong irrigation restrictions. The Bouidda district is an area irrigated by the Moulay Youssef reservoir in the eastern part of the study area. It obtained one of the best scores ( $r^2=0.863$ ). The area is dominated by cereals, which explains the typical seasonal  $K_c$  curve. The increase in permanent tree crops was well simulated at the baseline. There was also an intensification of cereal crops that increased the peaks at the end of the study period. The lower  $K_c$  years 2006 and 2007 were well simulated. The very dry year 2001 was simulated but did not reach the minima observed in the original curve.

225 The performance of the algorithm may be summarized as follows. The simple linear relationship provided an average approximation. The rainfall correction improved the average performance if the calibration years were representative of the validation years, elsewhere it worsened the prediction. There was no visible effect of the tree correction during the period of time where observations are available. The bounding constraint was basically not reached. Finally, the model results were much better if the calibration years were representative of the rainfall variability (wet/dry years). Given that the remote  
230 sensing dataset was processed for the 2000-2016 period (16 years), the performance of the prediction at the 2050 horizon should be comparable to the 1/3 test:  $r^2 = 0.5$ , RMSE = [0.1-0.14], stderr = [0.02-0.03].





#### 4.2 Evaluation of the trend and alternative evolution under climate change

Figure 6 shows examples of the extrapolation of  $K_c$  for four contrasted irrigated areas, with the trend and alternative  $K_c$  scenarios and RCP4.5 and RCP8.5 climatic scenarios. The region of private irrigation in Mejjat was all but covered with drylands until the early 90s. A boom of groundwater exploitation has provoked the development of irrigated areas.  $K_c$  plummets in the rare wet years, but most of the time, the amplitude of the signal is very small, showing that the area is mostly composed of bare soils and tree crops. The trend projection predicts that the  $K_c$  will tend toward an average of 0.2 by the year 2050, which means the impact of irrigation will be much greater than at the beginning of the XX century. The alternative  $K_c$  greatly reduced the trend and stabilized around 0.15. Ourika is an irrigated area located at the piedmont of the Atlas mountains. This area has traditionally relied on spate irrigation from the Atlas wadis feed by snowmelt. This area is also located over the Haouz aquifer where it benefits from recharge from below-river flows. The trend toward the growth of the irrigated area and full coverage with tree crops around 2040 was simulated on the trend  $K_c$ . On the alternative trajectory,  $K_c$  is less saturated, meaning that there is still a place for annual crops. As stated earlier, the N'Fis right bank irrigated area, which is located near Marrakech already had a trend toward the disappearance of irrigated crops. The two projections maintained the  $K_c$  level near zero except for wet years. The Bouidda irrigated area in the Tessaout region follows a trend of reconversion toward tree crops without a trend of cropped area development, which is almost fulfilled by 2050. In the alternative scenario, this trend was controlled and the area maintained a mix of annual and tree crops.

In those four examples, the climatic scenarios of rainfall only had a minimum impact on the  $K_c$  time series.

Figure 7 shows the relative evolution of water demand compared to the year 2000, and the expected impact of the "bending" effect for the four different planning areas in the watershed (see Fig. 1). The expected temperature increase between RCP4.5 and RCP8.5 is about 1°C by 2050, leading to a 2% difference in yearly  $ET_0$ . The impact on irrigation water demand is therefore minimal by the end of the simulation and has not been plotted in figure 7. However, we must keep in mind that if climate change only slightly impacts the water demand, it will probably reduce the water offer, because the run-off decrease will reduce the surface water available for irrigation (snowmelt, runoff, reservoirs). The Area Equipped for Irrigation (AEI) also remains unchanged while the actual irrigated area may change inside those AEIs. Those simulations showed that each area had its dynamic (bearing in mind that sub-areas also had their own dynamics). The Mejjat area, which was mostly unexploited in the year 2000, has been experiencing significant growth since. We can also see that by 2040, the demands tend to stabilize in the trend lines owing to the saturation effect of the irrigated areas. In the alternative scenario, the reduction in the agricultural development rate stabilizes at a lower level. The N'Fis area showed the lowest increase, and the modeled trend towards a flat line showed that the area will be fully dominated by tree crops by 2050. The Haouz and Tessaout areas showed the most significant increases in irrigation water demand, as it almost tripled during the 50 years of simulation. However, the alternative scenario mitigated these increases.



## 5. Discussion

The proposed method allowed projecting irrigation water demand by including both the anthropogenic and climatic vector of changes. The output is given for the next thirty years for each demand site at the monthly time step which are scales commonly used by water planners. Various improvements to the method have been identified: 1)  $K_c$  was determined by a pre-calibrated linear relation to the NDVI time series. If an actual evapotranspiration product (see for example Xu et al., 2019) would be available for the whole period of study and at a resolution compatible with the size of the irrigated areas,  $K_c$  could be determined by computing the ratio between actual evapotranspiration products and reference evapotranspiration. 2) The gridded data were synthesized at the monthly time step for each irrigated area, by computing arithmetic averages, which resulted in a loss of information. If needed, other statistical indicators, such as the dominant land cover of the irrigated area could be kept for further analysis. 3) The main temporal trend was fitted with monthly linear regressions (Eq. 3). The degree of the regression could be easily adapted if the model does not fit adequately with the observations.

The very low difference in impact between the two RCP scenarios is interesting. It can be explained by the short time range (2050) which resulted in a difference of only 1°C and almost no difference in precipitation over the study region. It should be noted that the predicted impact of climate change on precipitation is more variable in space and less certain than that for temperature. It also may be due to the absence of the impact of including increasing atmospheric CO<sub>2</sub> concentration prescribed by the RCP scenario in our approach. Different strategies are possible. For example, Fares et al. (2015) proposed a modification of the ratio bulk resistance over stomatal resistance  $r_s/r_a$  in the ET<sub>0</sub> equation based upon the CO<sub>2</sub> emission scenario. The low impact of RCP scenarios could also be explained by the fact that the potential decrease in the annual crop growth cycle duration associated with temperature rise is not represented. According to (Bouras et al., 2019), the reduction of the cycle can reach 30% at mid-century under RCP8.5 scenario for wheat in the study region. Similar results were obtained with the ISBA A-gs model (Calvet et al., 1998) when studying the impact of climate change on Mediterranean crops (Garrigues et al., 2015). It is also interesting to note that if the water demand for seasonal crops is reduced with the crop cycle duration, the yield may also be affected by extreme temperatures (Hatfield and Prueger, 2015) so that the trend might be affected in order to reach the production objective.

Finally, the assessment of irrigation water demand should also take into account different losses at the system level (storage, transportation and operating losses) and the plot level (deep percolation and runoff). Equation 1 can be rewritten to express the Gross Irrigation Water Requirement ( $IWR_G$  expressed in m<sup>3</sup>):

$$IWR_G = 1/\sigma \frac{AEI}{10} \frac{\varphi ET_0 K_c - \alpha P}{\beta} \quad (14)$$

where AEI (hectares) is the area equipped for irrigation. The system efficiency  $\sigma$  is the ratio of water delivered upstream to the water that is distributed to the plots (Blinda, 2012). The coefficient  $\alpha$  produces the efficient rainfall similar to that proposed by (USDA-SCS, 1970). The coefficient  $\varphi$  is used to account in a simple manner for a reduction of evapotranspiration due to micro-irrigation.  $\beta$  is the effective irrigation coefficient; it is low for surface irrigation (50-70%) and close to 100% for drip irrigation. The behavior of equation 14 is displayed in Figure 8 for two sample years (September



295 2011 to August 2013) for the R3 irrigated sector where wheat is the dominant crop. According to those different efficiency  
configurations, IWR can vary from single to more than quadruple between the ideal best and the worst situation. However,  
note that those configurations are theoretical. The best case with drip is virtually unachievable in real conditions.

## 6. Conclusions

Owing to their inherent complexity, scenarios of agricultural evolution based on interviews are very difficult to translate into  
300 numbers (when and how much), and to represent spatially (where). The simple and flexible statistical approach proposed  
here is a possible solution for quantifying spatially distributed scenarios of agriculture evolution in the context of climate  
change and irrigated areas that are rapidly changing owing to socio-economical influences. This is the case for many  
Mediterranean areas like the Tensift watershed which was the case used for the current study. The performance of the model  
was acceptable in most irrigated areas, giving monthly correlation coefficients of  $K_c$  of up to 0.92 but with significant  
305 differences between the irrigated areas. It was shown that the prediction performance of the model was a function of the  
length of the calibration period. With 16 years of data (2000-2016), the prediction could only be done for two other periods  
of 16 years (until 2050). In order to demonstrate the flexibility of the model for scenario building, a local scenario of water  
resource management was reinterpreted to build an alternative scenario upon the trend scenario. Two climate scenarios  
(RCP4.5 and RCP8.5) were introduced into the trend and alternative scenarios. The local anthropogenic scenarios of  
310 irrigation water demand had rapidly changing dynamics and were spatially contrasted. On the contrary, the difference of  
impact between the two climate change scenarios appeared to be very small over the next 30 years, although the impacts of  
atmospheric carbon concentration on stomatal controls of the plant were neglected. Finally, the discussion showed that the  
approach could easily be combined with irrigation efficiency scenarios. The conjunction of the present approach with  
irrigation efficiencies would probably be a suitable combination for the modeling of water resources management.

## 315 Data Availability

Some or all data, models, or code generated or used during the study are available in a repository or online in accordance  
with funder data retention policies. (K. Didan 2015, <https://www.medcordex.eu>). Some or all data, models, or code used  
during the study were provided by a third party (GSWP3). Some or all data, models, or code generated or used during the  
study are proprietary or confidential in nature and may only be provided with restrictions (ABHT GROUP AG - RESING,  
320 2017).

## Code availability

Models or code generated or used during the study are available from the corresponding author by request.



### Author contribution

MLP analyzed and processed the satellite and meteorological data and proposed the model. LJ and AB supervised the work  
325 and reviewed the paper and took part in critical discussions. AB disaggregated the climate change forcings. YF participated  
in the formulation of the research question and the writing of the paper. BB coordinated the work with the Tensift Basin  
Agency. LJ, SK and MZ participated in the research of funding and reviewed the paper. All authors read and agreed with the  
manuscript.

### Competing interests

330 The authors declare that they have no conflict of interest.

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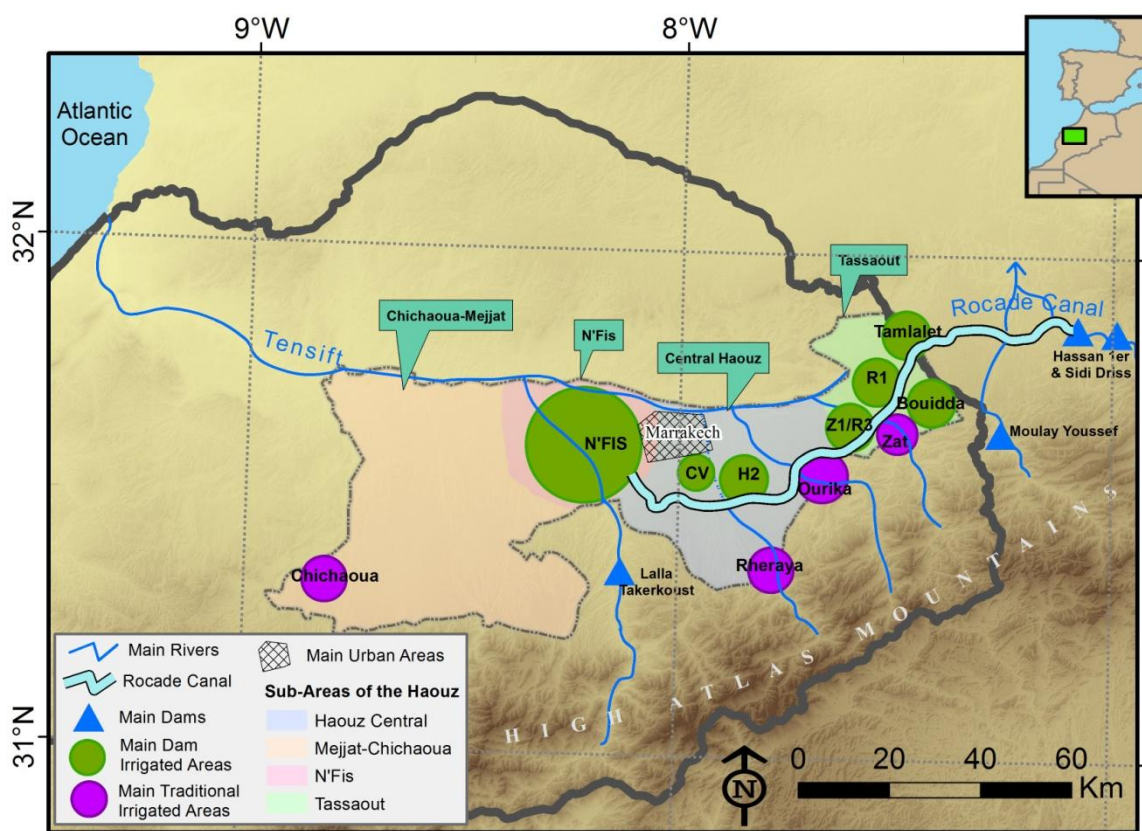


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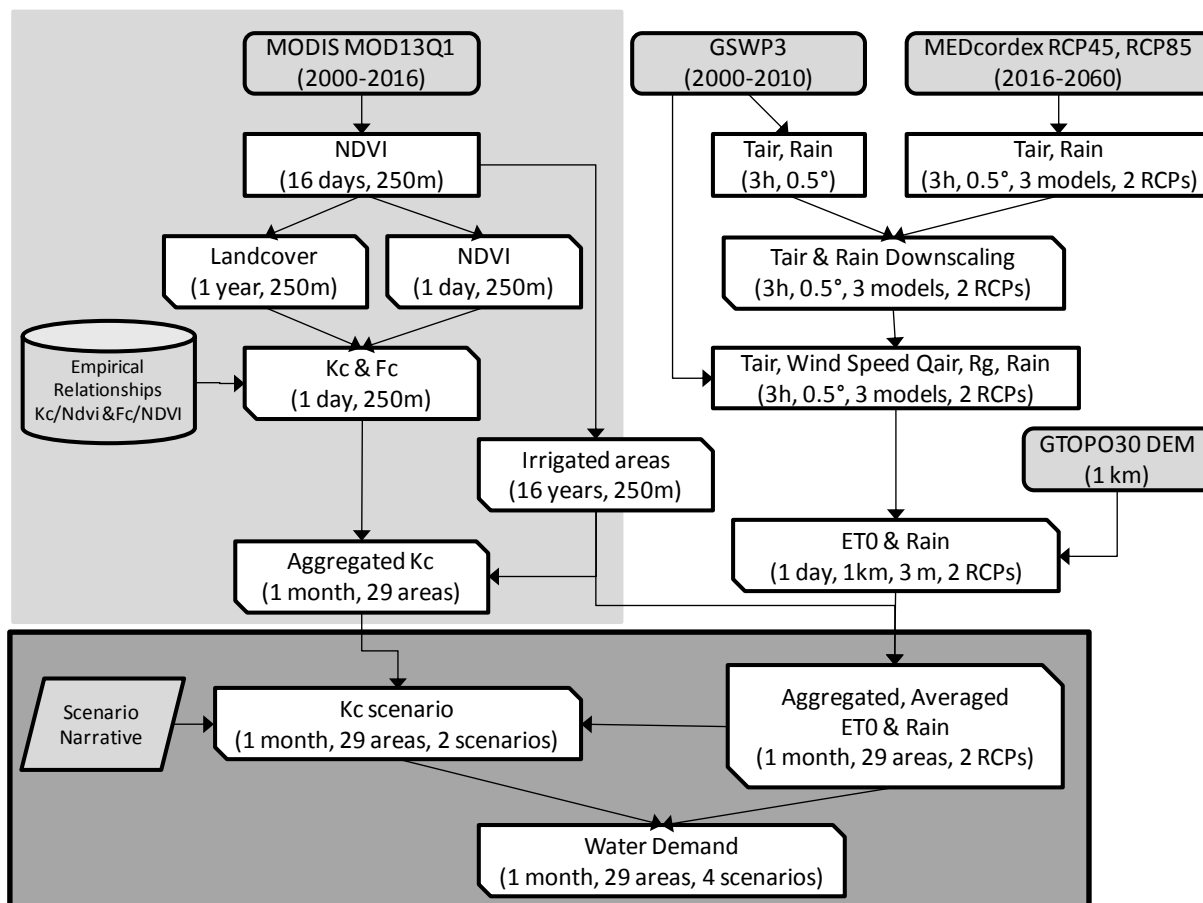


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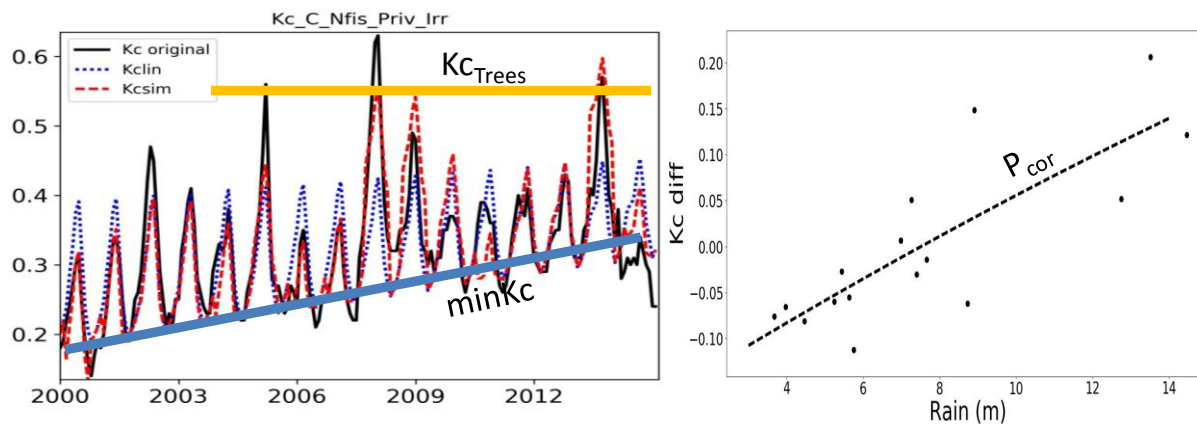
**Figure 1:** The Haouz-Mejjat aquifer is located within the Tensift watershed (grey) and separated into four subareas. The most important traditional and dam irrigated areas are shown on the map. The N'Fis area is a mix of traditional and modern irrigation. All of these areas also use groundwater, but outside them, the irrigated areas (private irrigation) rely exclusively on groundwater.

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**Figure 2: Processing Flowchart.** The light gray area of the Flowchart was presented in (Le Page et al., 2012). The calculations on  $K_c$  are performed on 29 irrigated areas. The input data are indicated in gray boxes. Variable extraction is indicated by a square box. Processings (and their results) are indicated by boxes with rounded corners. The indications between parenthesis show the temporal, spatial and scenario resolutions. The scenario generation described in this article is indicated in dark gray.

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480 **Figure 3: Quantities used to perform the simulation.** On the left, the three curves are the  $K_c$  time series obtained from remote sensing processing (solid line), the linear estimation ( $K_{clin}$ , Eq. 3, dotted blue line) and the estimation corrected with rainfall ( $K_{sim}$ , Eq. 6, dashed red line).  $K_{cTrees}$  has been set to 0.55 in Eq. 8, and  $minK_c$  is the yearly minimum of  $K_c$  on  $K_{sim}$ . On the right, the fitted correction curve of  $K_c$  according to the cumulated rainfalls ( $P_{corr}$ ) is almost linear in this example.

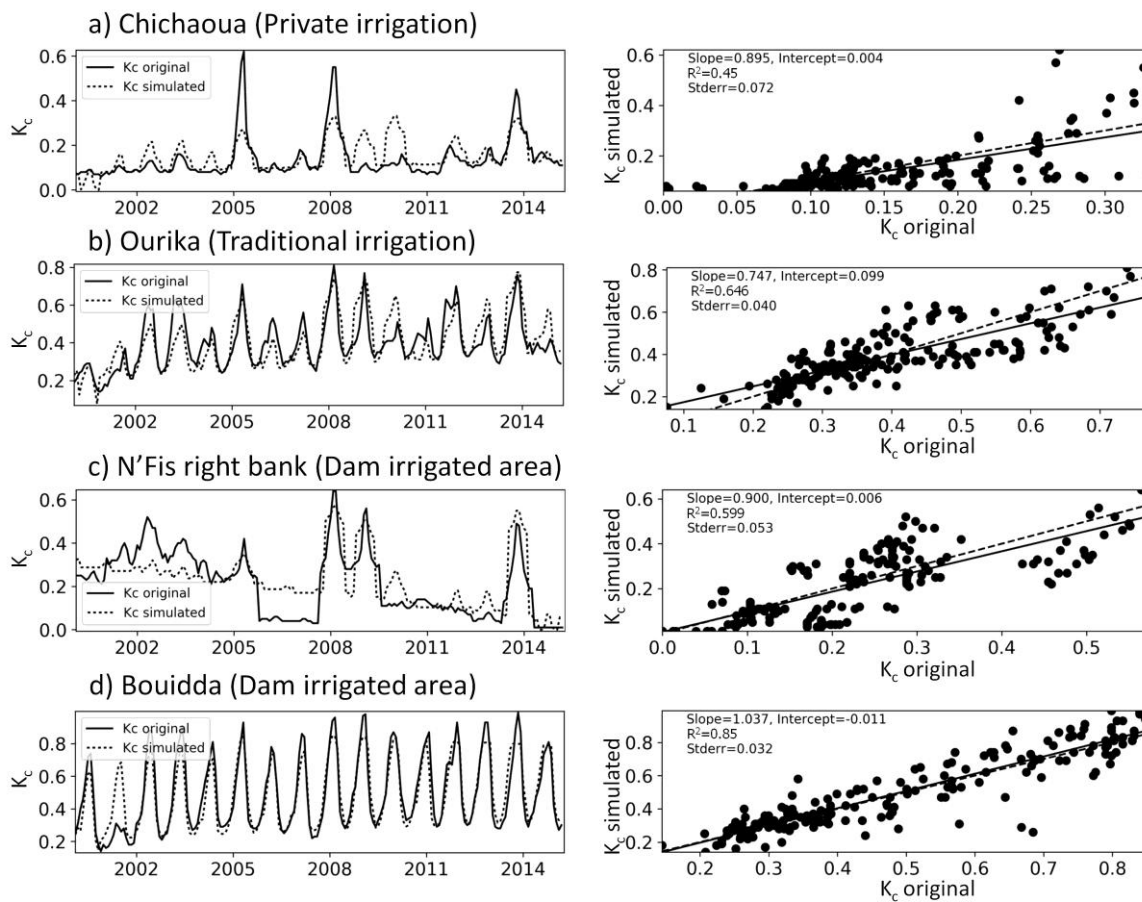
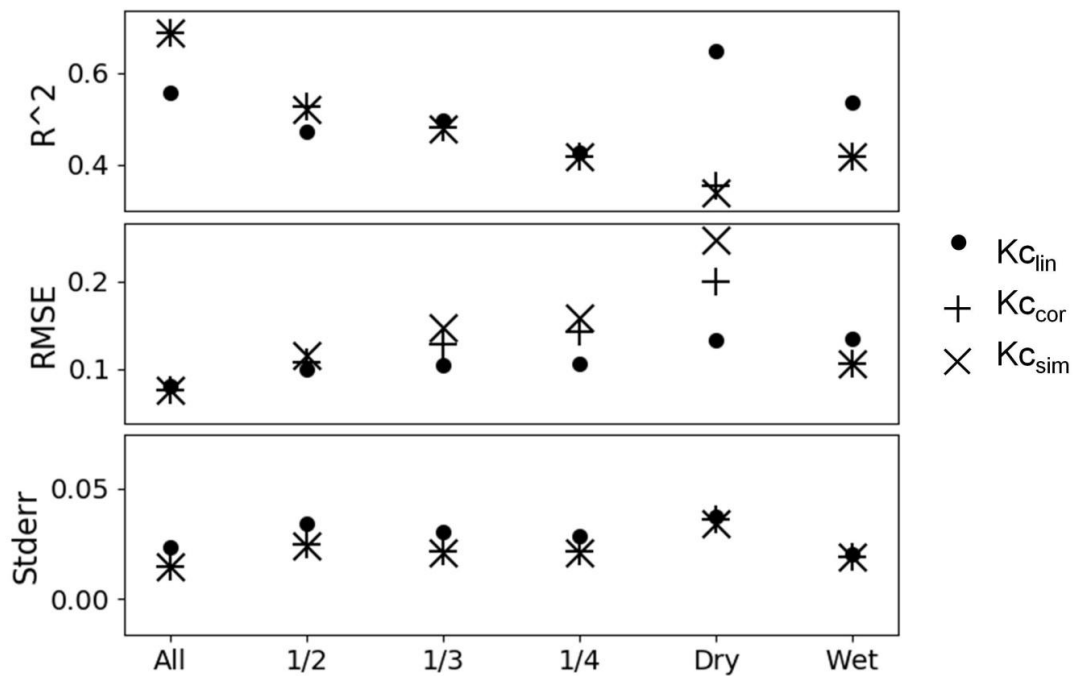
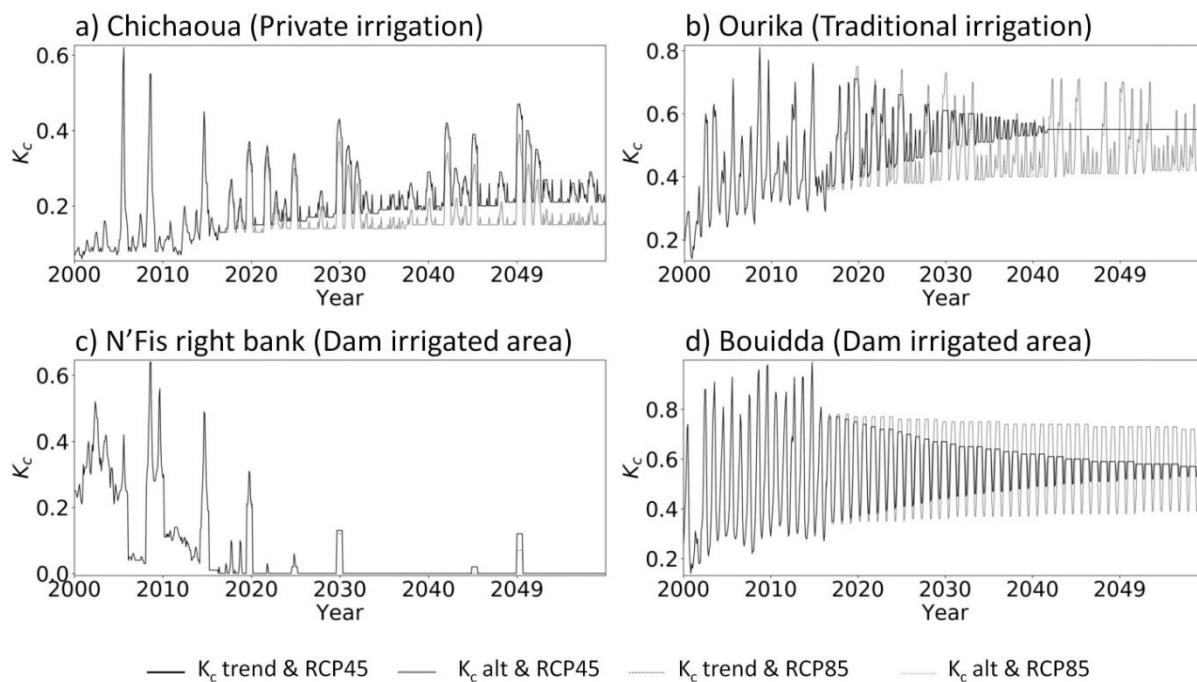


Figure 4: Examples of  $K_c$  simulations compared to the original time series for four selected areas.



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Figure 5: Performance indicators ( $r^2$ , RMSE, stderr) for the three variations of the model ( $Kc_{lin}$ ,  $Kc_{cor}$ ,  $Kc_{sim}$ ) as a function of the splitting strategy for determining the calibration/validation data set.



490 **Figure 6: Examples of long-term simulation of  $K_c$  in four different irrigated areas with the trend and alternative scenarios of  $K_c$  and the RCP4.5 and RCP8.5 climatic scenarios.**



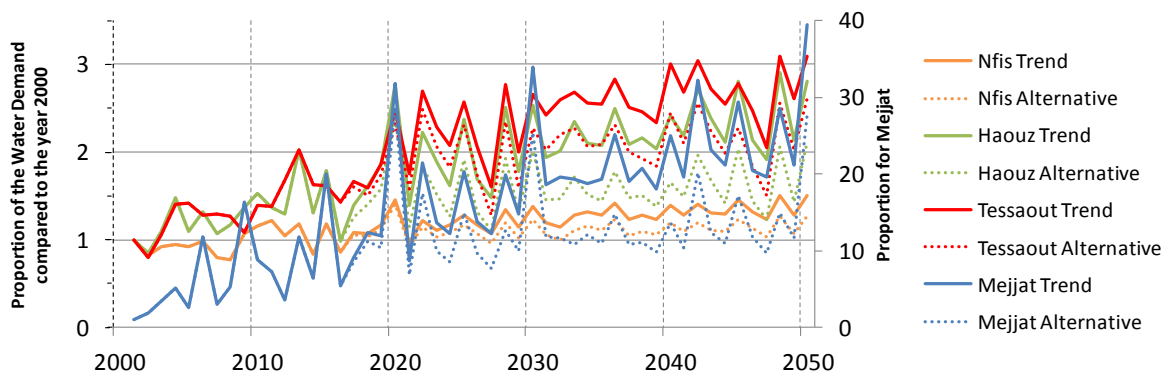


Figure 7: Trend and alternative scenarios for irrigation-water demand in the four planning areas of the Tensift with RCP8.5.

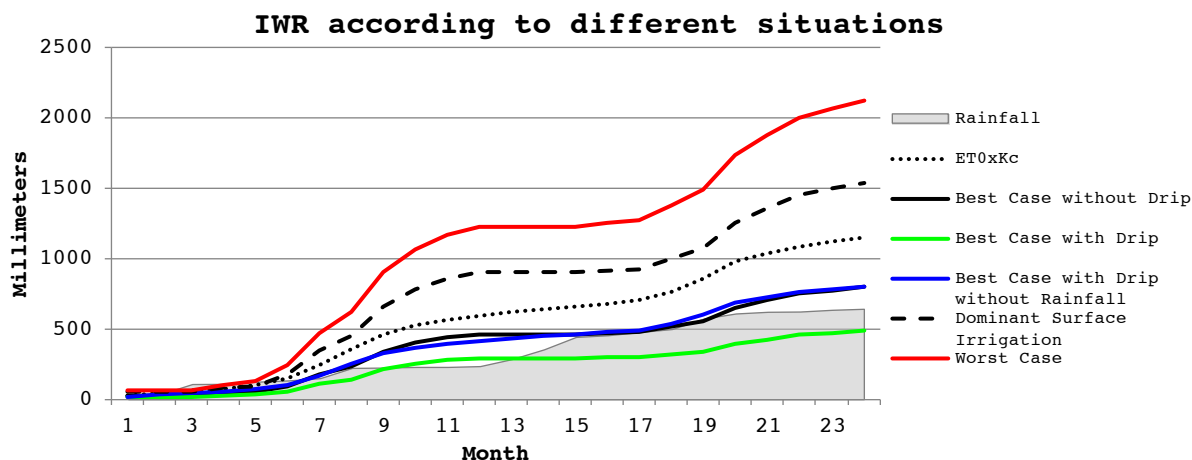


Figure 8:  $IWR_G$  according to different efficiencies for two years of example (September 2011 to August 2013) in the case of an irrigated wheat. In the best case without drip curve, all coefficients are set to one except the efficient rainfall which is set to a classical 0.75 ( $\sigma=1$ ,  $\varphi=1$ ,  $\alpha=0.75$ ,  $\beta=1$ ). In the best case with drip, The  $\sigma$  coefficient of localized irrigation is set to 0.7 ( $\sigma=1$ ,  $\varphi=0.7$ ,  $\alpha=0.75$ ,  $\beta=1$ ). The best case with drip but without taking into account rainfall simulates the irrigation practice done to prevent soil leaching ( $\sigma=1$ ,  $\varphi=0.7$ ,  $\alpha=0$ ,  $\beta=1$ ). The dominant surface irrigation simulates an irrigated sector where surface (gravitatory) irrigation dominates ( $\alpha=0.5$ ) and where the system performance is good ( $\sigma=0.9$ ) ( $\sigma=0.9$ ,  $\varphi=1$ ,  $\alpha=0.9$ ,  $\beta=0.5$ ). The worst-case depicts an irrigated sector with low irrigation efficiencies ( $\sigma=0.75$ ,  $\varphi=1$ ,  $\alpha=0.75$ ,  $\beta=0.5$ ).

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