1	Rapid reduction in ecosystem productivity caused by flash drought based on
2	decade-long FLUXNET observations
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Abstract. Flash drought is characterized by its rapid onset and arouses wide concerns 15 due to its devastating impacts on the environment and society without sufficient early 16 17 warnings. The increasing frequency of soil moisture flash drought in a warming climate highlights the importance of understanding its impact on terrestrial 18 19 ecosystems. Previous studies investigated the vegetation dynamics during several extreme cases of flash drought, but there is no quantitative assessment on how fast the 20 carbon fluxes respond to flash drought based on decade-long records with different 21 22 climates and vegetation conditions. Here we identify soil moisture flash drought 23 events by considering decline rate of soil moisture and the drought persistency, and detect the response of ecosystem carbon and water fluxes to soil moisture flash 24 drought during its onset and recovery stages based on observations at 29 FLUXNET 25 26 stations from croplands to forests. Corresponding to the sharp decline in soil moisture and higher VPD, gross primary productivity (GPP) drops below its normal conditions 27 in the first 16 days and reduces to its minimum within 24 days for more than 50% of 28 29 the 151 identified flash drought events, and savannas show highest sensitivity to flash drought. Water use efficiency increases for forests but decreases for cropland and 30 31 savanna during the recovery stage of flash droughts. These results demonstrate the rapid responses of vegetation productivity and resistance of forest ecosystems to flash 32 33 drought.

34 Keywords: Flash drought; GPP; Soil moisture; Water use efficiency; FLUXNET

1. Introduction

Terrestrial ecosystems play a key role in the global carbon cycle and absorb 36 37 about 30% of anthropogenic carbon dioxide emissions during the past five decades (Le Ou ét ét al., 2018). With more climate extremes (e.g. droughts, heat waves) in a 38 39 warming climate, the rate of future land carbon uptake is highly uncertain regardless of the fertilization effect of rising atmospheric carbon dioxide (Green et al., 2019; 40 Reichstein et al., 2013; Xu et al., 2019). Terrestrial ecosystems can even turn to 41 42 carbon source during extreme drought events (Ciais et al., 2005). Record-breaking 43 drought events have caused enormous reduction of the ecosystem gross primary productivity (GPP), such as the European 2003 drought (Ciais et al., 2005; Reichstein 44 45 et al., 2007), USA 2012 drought (Wolf et al., 2016), China 2013 drought (Xie et al., 46 2016; Yuan et al., 2016), Southern Africa 2015/16 drought (Yuan et al., 2017) and Australia Millennium drought (Banerjee et al., 2013). The 2012 summertime drought 47 in USA was classified as flash drought with rapid intensification and insufficient early 48 49 warning, which caused 26% reduction in crop yield (Hoerling et al., 2014; Otkin et al., 2016). Flash drought may only need several weeks to develop into its maximum 50 51 intensity, and the rapid onset distinguishes it from traditional drought which is assumed to be a slowly evolving climate phenomenon taking several months or even 52 years to develop (Otkin et al., 2018). Several extreme flash droughts would ultimately 53 propagate into long-term droughts due to persistent precipitation deficits, e.g., 2012 54 55 flash drought over the USA Midwest Plain (Basara et al., 2019). Flash drought has aroused wide concerns for its unusually rapid development and detrimental effects 56

(Basara et al., 2019; Christian et al., 2019; Ford & Labosier, 2017; Nguyen et al.,
2019; Otkin et al., 2018a; Otkin et al., 2018b; Wang and Yuan, 2018; Yuan et al., 2015;
Yuan et al., 2017; Yuan et al., 2019b). Despite the increasing occurrence and clear
ecological impacts of flash droughts, our understanding of their impacts on carbon
uptake in terrestrial ecosystems remains incomplete.

Previous studies mainly focused on the response of vegetation to long-term 62 droughts, and found that the response time ranged from several months to years 63 through correlation analysis (Vicente-Serrano et al., 2013; Xu et al., 2018). The 64 65 response time of vegetation to flash droughts might be different, which requires further investigation for quantification. Recent studies assessed the impact of flash 66 drought on vegetation including the 2012 central USA flash drought and the 2016 and 67 68 2017 northern USA flash drought. For instance, Otkin et al. (2016) used the evaporative stress index (ESI) to detect the onset of the 2012 central USA flash 69 drought, and found the decline in ESI preceded the drought according to the United 70 71 States Drought Monitor (Svoboda et al., 2002). He et al. (2019) assessed the impacts 72 of the 2017 northern USA flash drought (which also impacted parts of southern 73 Canada) on vegetation productivity based on GOME-2 solar-induced fluorescence (SIF) and satellite-based evapotranspiration in the US Northern plains. Otkin et al. 74 (2019) examined the evolution of vegetation conditions using LAI from MODIS 75 during the 2015 flash drought over the South-Central United States and found that the 76 LAI decreased after the decline of soil moisture. Besides, the 2016 flash drought over 77 U.S. northern plains also decreased agricultural production (Otkin et al., 2018b). 78

However, previous impact studies only focused on a few extreme flash drought cases without explicit definition of flash drought events. As the baseline climate is changing (Yuan et al., 2019b), it is necessary to systematically investigate the response of terrestrial carbon and water fluxes to flash drought events based on long-term records rather than one or two extreme cases.

In fact, there are numerous studies on the influence of drought on ecosystem 84 productivity (Ciais et al., 2005; Stocker et al., 2018; Stocker et al., 2019). It is found 85 that understanding the coupling of water-carbon fluxes during drought is the key to 86 87 revealing the adaptation and response mechanisms of vegetation to water stress (Boese et al., 2019; Nelson et al., 2018). Water use efficiency (WUE) is the metric for 88 89 understanding the trade-off between carbon assimilation and water loss through 90 transpiration (Beer et al., 2009; Cowan and Farquhar, 1977; Zhou et al., 2014, 2015), and it is influenced by environmental factors including atmospheric dryness and soil 91 moisture limitations (Boese et al., 2019). Although WUE has been widely studied for 92 93 seasonal to decadal droughts, few studies have investigated WUE during flash 94 droughts that usually occur at sub-seasonal time scale (Xie et al., 2016; Zhang et al., 95 2019).

In this paper, we address the ecological impact of soil moisture flash droughts through analyzing FLUXNET decade-long observations of CO_2 and water fluxes. Here we consider not only the rapid onset stage of soil moisture flash droughts but also the recovery stage to assess the ecological impacts. The ecological responses to water stress vary under different ecosystems and drought characteristics, and the focus 101 on the soil moisture flash droughts would detect the breakdown of ecosystem functioning of photosynthesis. The specific goals are to (1) examine the response of 102 103 carbon and water fluxes to soil moisture flash droughts from the onset to the recovery stages, and (2) investigate how WUE changes during soil moisture flash drought for 104 105 different ecosystems. The methodology proposed by Yuan et al. (2019b) enables the analysis of the flash drought with characteristics of duration, frequency, and intensity 106 in the historical observations. All the flash drought events occurred at the FLUXNET 107 108 stations are selected to investigate the response of carbon fluxes and WUE. More than 109 10-year records of soil moisture, carbon and water fluxes are available (Baldocchi et al., 2002), which makes it possible to assess the response of vegetation to flash 110 droughts by considering different climates and ecosystem conditions. 111

112 **2. Data and Methods**

113 **2.1 Data**

FLUXNET2015 provides daily hydrometeorological variables including precipitation, 114 115 temperature, saturation vapor pressure deficit (VPD), soil moisture (sm), shortwave radiation (SW), evapotranspiration (ET) inferred from latent heat, and carbon fluxes 116 117 including GPP and net ecosystem productivity (NEP). We use GPP data based on night-time partitioning method (GPP NT VUT REF). Considering most sites only 118 measure the surface soil moisture, here we use daily soil moisture measurements 119 120 mainly at the depth of 5-10 cm averaged from half-hourly data. Soil moisture observations are usually averaged over multiple sensors including time domain 121 reflectometer (TDR), frequency domain reflectometer (FDR), and water content 122

reflectometer etc. However, the older devices may be replaced with newer devices at 123 certain sites, which may decrease the stability of long-term soil moisture observations 124 and the average observation error of soil moisture is $\pm 2\%$. All daily 125 hydrometeorological variables and carbon fluxes are summed to 8-day time scale to 126 study the flash drought impact. There are 34 sites from FLUXNET 2015 dataset 127 (Table 1) consisting of 8 vegetation types, where the periods of observations are no 128 less than 10 years ranging from 1996 to 2014, and the rates of missing data are lower 129 130 than 5%. Here we only select the FLUXNET observations including 12 evergreen 131 needleleaf forest sites (ENF), 5 deciduous broadleaf forests (DBF), 6 crop sites (CROP; 5 rain-fed sites and 1 irrigated site), 3 mixed forests (MF), and 3 savannas 132 (SAV). The sites for grasslands, evergreen broadleaf forests, and shrublands are 133 134 excluded because there are less than 10 soil moisture flash drought events. The vegetation classification is according to International Geosphere-Biosphere Program 135 (IGBP; Belward et al., 1999), where MF is dominated by neither deciduous nor 136 137 evergreen tree type with tree cover larger than 60% and the land tree cover is 10-30% 138 for SAV. The detailed information is listed in Table 1.

139 **2.2 Methods**

140 **2.2.1 Definition of soil moisture flash drought events**

The definition of soil moisture flash drought should account for both its rapid intensification and the drought conditions (Otkin et al., 2018a; Yuan et al., 2019b). Here we used soil moisture percentile to identify soil moisture flash drought according to Yuan et al. (2019b) and Ford et al. (2017). Figure 1 shows the procedure

for soil moisture flash drought identification, including five criteria to identify the 145 rapid onset and recovery stages of soil moisture flash drought. 1) Soil moisture flash 146 drought starts at the middle day of the 8-day period when the 8-day mean soil 147 moisture is less than the 40th percentile, and the 8-day mean soil moisture prior to the 148 starting time should be higher than 40th percentile to ensure the transition from a 149 150 non-drought condition. 2) The mean decreasing rate of 8-day mean soil moisture percentile should be no less than 5% per 8 days to address the rapid drought 151 intensification. 3) The 8-day mean soil moisture after the rapid decline should be less 152 153 than 20% in percentile, and the period from the beginning to the end of the rapid decline is regarded as the onset stage of soil moisture flash drought (those within red 154 dashed line in Figure 1). 4) If the mean decreasing rate is less than 5% in percentile or 155 156 the soil moisture percentile starts to increase, the soil moisture flash drought enters into the "recovery" stage, and the soil moisture flash drought event (as well as the 157 recovery stage) ends when soil moisture recovers to above 20th percentile (those 158 159 within blue dashed line in Figure 1). The recovery stage is also crucial to assess the impact of soil moisture flash drought (Yuan et al., 2019b). 5) The minimum duration 160 of a flash drought event is 24 days to exclude those dry spells that last for a too short 161 period to cause any impacts. 162

At least decade-long observations of 8-day mean soil moisture are used to calculate soil moisture percentile with a moving window of 8-day before and 8-day after the target 8-day, resulting in at least 30 samples for deriving the cumulative distribution function of soil moisture before calculating percentiles. Besides, the target 8-day soil moisture percentiles are only based on the target 8-day soil moisture in the context of the expanded samples. For example, the soil moisture percentile of June 22nd in 1998 is calculated by firstly ranking June 14th, June 22nd, and June 30th soil moisture in all historical years (N samples) from lowest to highest, identifying the rank of soil moisture of June 22nd, 1998 (e.g., M), and obtaining the percentile as M/N*100. We focus on growing seasons during April-September for sites in the North Hemisphere and October-March for sites in the South Hemisphere.

174 2.2.2 Response time of GPP to soil moisture flash drought

175 Drought has a large influence on ecosystem productivity through altering the plant photosynthesis and ecosystem respiration (Beer et al., 2010; Green et al., 2019; 176 Heimann & Reichstein, 2008; Stocker et al., 2018). GPP dominates the global 177 178 terrestrial carbon sink and it would decrease due to stomatal closure and non-stomatal limitations like reduced carboxylation rate and reduced active leaf area index (de la 179 Motte et al., 2019) under water stress. The negative anomalies of GPP during soil 180 181 moisture flash drought are considered as the onset of ecological response. Here, we use two response time indices to investigate the relationship between soil moisture 182 flash drought and ecological drought (Crausbay et al., 2017; Niu et al., 2018; Song et 183 al., 2018; Vicente-Serrano et al., 2013): 1) the response time of the first occurrence 184 (RT) of negative standardized GPP anomaly (SGPPA= $\frac{GPP-\mu_{GPP}}{\sigma_{GPP}}$, where μ_{GPP} and 185 σ_{GPP} are mean and standard deviation of the time series of GPP at the same dates as 186 the target 8-days for all years, which can remove the influence of seasonality. For 187 instance, all Apr 1-8 during 1996-2014 would have a μ_{GPP} and a σ_{GPP} based on a 188

climatology same as soil moisture percentile calculation which consists of March 189 24-31, Apr 1-8, and Apr 9-16 in all years, and Apr 9-16 would have another μ_{GPP} 190 191 and another σ_{GPP} , and so on), which is the lag time between the start of flash drought and the time when SGPPA becomes negative during flash drought period; and 2) the 192 193 response time of occurrence of minimum SGPPA (RTmin), which is the lag time between the start of flash drought and the time when SGPPA decreases to its 194 minimum values during the flash drought period. If the response time is 8 days for the 195 first occurrence of negative SGPPA, it means that the response of GPP starts at the 196 197 beginning of flash drought (the first time step of flash drought). Considering flash drought is identified through surface soil moisture due to the availability of 198 FLUXNET data, vegetation with deeper roots may obtain water in deep soil and 199 200 remain healthy during flash drought. The roots vary among different vegetation types and forests are assumed to have deeper roots than grasslands, which may influence the 201 response to soil moisture flash droughts. 202

203 2.2.3 Water use efficiency

Carbon assimilation and transpiration are coupled by stomates, and plants face a tradeoff between carbon uptake through photosynthesis and water loss through transpiration under the influence of water and energy availability (Boese et al., 2019; Gentine et al., 2019; Huang et al., 2016; Nelson et al., 2018). WUE can be used to quantify the trade-off between carbon and water cycles, which is defined as the assimilated amount of carbon per unit of water loss (Peters et al., 2018). At the ecosystem scale, WUE is the ratio of GPP over ET (Cowan and Farquhar, 1977).

Drought would cause stomatal closure and non-stomatal adjustments in biochemical 211 functions thus altering the coupling between GPP and ET. Underlying WUE (uWUE) 212 is calculated as $GPP \times \sqrt{VPD}/ET$ considering the nonlinear relationship between 213 GPP, VPD and ET (Zhou et al., 2014). uWUE is supposed to reflect the relationship of 214 215 photosynthesis-transpiration via stomatal conductance at the ecosystem level by considering the effect of VPD on WUE (Beer et al., 2009; Boese et al., 2019; Zhou et 216 al., 2014, 2015). WUE varies under the influence of VPD on canopy conductance 217 (Beer et al., 2009; Tang et al., 2006), whereas uWUE is considered to remove this 218 219 effect and be more directly linked with the relationship between environmental conditions (e.g., soil moisture) and plant conditions (e.g., carboxylation rate; Lu et al., 220 221 2018). The standardized anomalies of WUE and uWUE are calculated the same as 222 SGPPA, where different sites have different mean values and standard deviations for different target 8-days to remove the spatial and temporal inhomogeneity. 223

224 2.2.4 The relations between meteorological conditions and GPP

225 Considering the compound impacts of temperature, radiation, VPD and soil 226 moisture on vegetation photosynthesis, the partial correlation is used to investigate the 227 relationship between GPP and each climate factor, with the other 3 climate factors as 228 control variables as follows:

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$$r_{ij(m_1,m_2...m_n)} = \frac{r_{ij(m_1,...m_{n-1})} - r_{im_n(m_1,...m_{n-1})} r_{jm_n(m_1,...m_{n-1})}}{\sqrt{(1 - r_{in(m_1,...m_{n-1})})(1 - r_{jn(m_1,...m_{n-1})}^2)}}$$
(1)

where *i* represents GPP, *j* represents the target meteorological variables and $m_1, m_{2...} and m_n$ represent the control meteorological variables. $r_{ij(m_1,m_{2...}m_n)}$ is the partial correlation coefficient between *i* and *j*, and $r_{ij(m_1,...m_{n-1})}$, $r_{im_n(m_1,...m_{n-1})}$ and 233 $r_{jm_n(m_1,\dots,m_{n-1})}$ are partial correlation coefficients between *i* and *j*, *i* and *m_n*, *j* and 234 m_n respectively under control of m_1, m_2 and m_{n-1} .

235 **3. Results**

236 **3.1 Identification of flash drought events at FLUXNET stations**

Based on FLUXNET data, we have identified 151 soil moisture flash drought 237 events with durations longer than or equal to 24 days using soil moisture observations 238 of 371 site years. Figure 2a shows the distribution of the 29 sites with different 239 240 vegetation types, which are mainly distributed over North America and Europe. The 241 number of soil moisture flash drought ranges from 13 to 70 events among different vegetation types. There are 12 ENF sites in this study, and the number of soil moisture 242 243 flash droughts for ENF (70) is the most among all the vegetation types. The duration 244 for flash drought events ranges from 24 days to several months. In some extreme cases, the flash droughts would develop into long-term droughts without enough 245 rainfall to alleviate drought conditions. Mean durations of soil moisture flash droughts 246 247 for different vegetation types range from around 30 days to 50 days (Figure 2c).

Figure 3 shows the meteorological conditions during different stages of soil moisture flash drought including the standardized anomalies of temperature, precipitation, VPD, and shortwave radiation and soil moisture percentiles. Here the onset and recovery stages of flash droughts refer to certain periods characterized by the soil moisture decline rates. The standardized anomalies of temperature, precipitation, VPD, and shortwave and soil moisture percentiles are composited to show the meteorological conditions during different stages of flash droughts. The

onset stage of soil moisture flash droughts mainly refers to the rapid intensification, 255 and the flash droughts may or may not develop into long-term droughts depending on 256 257 the deficits in precipitation. There is a slight reduction in precipitation during 8 days prior to soil moisture flash drought (Figure 3b). During the onset of soil moisture 258 259 flash drought, soil moisture percentiles decline rapidly from nearly 50% during 8 days before flash drought to 18% during onset stages (Figure 3e). The rapid drying of soil 260 moisture is always associated with a large precipitation deficits, anomalously high 261 temperature and shortwave radiation and large VPD indicate increased atmospheric 262 263 dryness (Ford et al., 2017; Koster et al., 2019; Wang et al., 2016), which persist until the recovery stage except for shortwave radiation. The soil moisture percentiles are 264 averaged during the onset and recovery stages and the soil moisture percentiles during 265 266 recovery stages are slightly lower than those during onset stages (Figure 3e) considering the soil moisture is not quite dry during the early period of onset stages. 267 Sufficient precipitation occurs during the 8 days after soil moisture flash droughts to 268 269 relieve the drought condition and soil moisture percentiles increase from 12% during recovery stages to 36% during 8 days after flash droughts. 270

3.2 Climatological statistics of the response time of GPP to flash drought

By analyzing all the 151 soil moisture flash drought events across 29 FLUXNET sites, we find that negative GPP anomalies occur during 81% of the soil moisture flash drought events. Figure 4 shows the probability distributions of the response time of GPP to soil moisture flash drought as determined by soil moisture reductions for the first occurrence of negative SGPPA, the minimum negative value of SGPPA and

the minimum soil moisture percentiles for different vegetation types, respectively. To 277 reduce the uncertainty due to small sample sizes, only the results for vegetation types 278 (SAV, CROP, MF, DBF, ENF) with more than 10 flash drought events are shown. For 279 soil moisture flash droughts from all vegetation types, the first occurrences of 280 281 negative SGPPA are concentrated during the first 24 days, and GPP starts to respond to soil moisture flash drought within 16 days for 57% flash droughts (Figures 4a-e). 282 The occurrences of minimum value of SGPPA rise sharply at the beginning of soil 283 moisture flash drought, and reach the peak during 17-24 days, and then slow down 284 285 (Figures 4f-j), which is similar to the decline in soil moisture. Although the first occurrences of negative SGPPA mainly occur in the onset stage, GPP would continue 286 to decrease in the recovery stages for 60% of soil moisture flash drought events. 287 288 Different types of vegetation including herbaceous plants and woody plants all react to soil moisture flash drought in the early stage (Figures 4a-e). Among them, SAV 289 shows the fastest reaction to water stress (Figures 4a and 4f), and the RT is within 8 290 days for 63% events, suggesting that SAV responds concurrently with soil moisture 291 flash drought onset. Ultimately, 88% events for SAV show reduced vegetation 292 photosynthesis. The result is consistent with previous studies regarding the strong 293 response of semi-arid ecosystems to water availability (Gerken et al., 2019; 294 Vicente-Serrano et al., 2013; Zeng et al., 2018), and the decline in GPP for SAV is 295 related to isohydric behaviors during soil moisture drought and higher VPD, through 296 closing stomata to decrease water loss as transpiration and carbon assimilation 297 (Novick et al., 2016; Roman et al., 2015). For ENF, only 27% of soil moisture flash 298

droughts cause the negative SGPPA during the first 8 days. When RT is within 40 299 days, the cumulative frequencies range from 74% to 88% among different vegetation 300 types. The response frequency of RTmin and the response time of minimum soil 301 moisture percentiles are quite similar, although there are discrepancies among the 302 303 patterns of the response frequency for different vegetation types. The response frequency of RTmin for SAV increases sharply during 17-24 days of soil moisture 304 flash droughts (Figure 4f). GPP is derived from direct eddy covariance observations 305 of NEP and nighttime terrestrial ecosystem respiration, and temperature-fitted 306 307 terrestrial ecosystem respiration during daytime. The response of NEP to flash droughts shows the compound effects of vegetation photosynthesis and ecosystem 308 respiration. In terms of RT, the response of NEP is slower than GPP for SAV, but is 309 310 quicker for DBF and ENF (Figure 5). The discrepancies between NEP and SM in terms of RTmin are more obvious than those between GPP and SM, and the RTmin of 311 NEP is much shorter than the RTmin of soil moisture especially for DBF and ENF, 312 313 which may be related to the increase of ecosystem respiration (Figures 5 i and j).

Figure 6 shows the temporal changes of SGPPA and soil moisture percentiles during 8 days before soil moisture flash droughts and during the first 24 days of the droughts. During 8 days before flash droughts, there is nearly no obvious decline for SGPPA, while SAV, DBF and ENF shows small increase in GPP. The decline in SGPPA is more significant during the first 9-24 days of soil moisture flash droughts for different vegetation types, and SGPPA for SAV and CROP show quicker decline even during the first 8 days of soil moisture flash droughts. The decline rates in soil moisture are mainly concentrated within the first 16 days of flash droughts. There are various lag times for the response of GPP to the decline in soil moisture among different vegetation.

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3.3 The coupling between carbon and water fluxes under soil moisture stress

Figure 7 shows the standardized anomalies of WUE and uWUE and their 325 components for different ecosystems during 8 days before and after soil moisture 326 flash droughts and the onset and recovery stages. Here, we select 81% of soil moisture 327 flash drought events with GPP declining down to its normal conditions to analyze the 328 329 interactions between carbon and water fluxes, while GPP during the remaining 19% of soil moisture flash drought events may stay stable and is less influenced by drought 330 conditions. During 8 days before soil moisture flash drought, WUE and uWUE are 331 332 generally close to the climatology (Figure 7a) and there are no significant changes in GPP, ET, and ET/\sqrt{VPD} (Figures 7e and 7i). However, the median value of SGPPA 333 for SAV is positive (Figure 7e). WUE is stable during the onset stage, whereas uWUE 334 335 increases for all ecosystems except for CROP (Figure 7b). For CROP, both GPP and 336 ET decrease, and the decline in WUE is related with a greater reduction in GPP relative to ET (Figure 7f and 7j). The positive anomalies of uWUE are correlated with 337 decrease in ET/\sqrt{VPD} mainly induced by the high VPD. Increasing VPD and 338 339 deficits in soil moisture would decrease canopy conductance (Grossiord et al., 2020) but not GPP for MF and ENF. During the onset stage, GPP and ET reduce only for 340 341 SAV, and CROP, and DBF, and the magnitudes of GPP and ET reduction are highest for SAV. ET is close to normal conditions for MF, DBF, and ENF, thus enhancing the 342

drying rate of soil moisture with less precipitation supply during the onset stage. But 343 during recovery stage of soil moisture flash drought, GPP and ET show significant 344 345 reductions except for MF (Figures 7g and 7k), and the responses of WUE and uWUE are different between herbaceous plants (SAV and CROP) and forests (MF, DBF, and 346 ENF), where WUE and uWUE decrease significantly for SAV and CROP but increase 347 slightly for forests (Figure 7c). The decrease in uWUE for SAV and CROP during 348 recovery stages indicates that SAV and CROP are likely brown due to carbon 349 350 starvation caused by the significant decrease in stomatal conductance (McDowell et 351 al., 2008). The decrease in GPP during recovery stage is not only related to the reduction in canopy conductance, but also the decrease in uWUE under drought for 352 SAV and CROP which is possibly influenced by suppressed state of enzyme and 353 354 reduced mesophyll conductance (Flexas et al., 2012). However, the positive anomalies of uWUE for DBF and ENF during the recover stage imply that the decline 355 in GPP mainly results from the stomata closure. ET starts to decrease during the 356 357 recovery stage due to the limitation of water availability, and the decreasing ET also reflects the enhanced water stress for vegetation during the recovery stage. The 358 average soil moisture conditions are 12% in percentile for recovery stage but 18% for 359 onset stage. So, drier soil moisture in the recovery stage exacerbates ecological 360 response. Figure 7c also shows the higher WUE and uWUE for forests, which 361 indicates their higher resistance to flash drought than herbaceous plants during 362 363 recovery stage. During 8 days after flash drought, the standardized anomalies of uWUE are still positive for forests, whereas SGPPA and ET are both lower than the 364

365 climatology for all ecosystems. The ecological negative effect would persist after the366 soil moisture flash drought.

3.4 The impact of climate factors on GPP during soil moisture flash drought

Figure 8 shows the partial correlation coefficients between standardized 368 anomalies of GPP and meteorological variables and soil moisture percentiles during 369 different stages of soil moisture flash droughts. The correlation between climate 370 factors and GPP is not statistically significant during 8 days before soil moisture flash 371 droughts. During onset stages of soil moisture flash droughts, the partial correlation 372 373 coefficients between SGPPA and soil moisture percentiles are 0.44, 0.49 and 0.29, respectively for SAV, CROP, and ENF (p<0.05). Besides, shortwave radiation is 374 positively correlated with SGPPA for MF, DBF, and EBF (Figure 8b) during onset 375 376 stages and the positive anomalies of shortwave radiation could partially offset the loss of vegetation photosynthesis due to the deficits in soil moisture. SGPP is also 377 positively correlated with temperature during onset stages for SAV and DBF. The 378 partial correlation coefficients between SGPPA and VPD are -0.53 and -0.22 379 respectively for DBF and ENF, and the higher VPD would further decrease GPP 380 381 during onset stages. The influence of VPD on GPP is much more significant during recovery stages and 8 days after. SGPPA is positively correlated with soil moisture 382 and negatively with VPD for SAV both during recovery stages and 8 days after. 383

384 4. Discussion

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385 Previous studies detected the vegetation response for a few extreme drought cases386 without a specific definition of flash drought from a climatological perspective (Otkin

et al., 2016; He et al., 2019). Moreover, less attention has been paid to the coupling 387 between carbon and water fluxes during soil moisture flash drought events. This study 388 389 investigates the response of carbon and water fluxes to soil moisture flash drought based on decade-long FLUXNET observations during different stages of flash 390 391 droughts. The responses vary across different phases of flash drought, and different ecosystems have different responses, which provide implications for eco-hydrological 392 modeling and prediction. Besides, the influence of different climate factors including 393 VPD and soil moisture also differs during different stages of soil moisture flash 394 395 droughts.

396 **4.1 The responses of carbon and water fluxes to flash droughts**

Based on 151 soil moisture flash drought events identified using soil moisture 397 398 from decade-long FLUXNET observations, the response of GPP to flash drought is found to be quite rapid. For more than half of the 151 soil moisture flash drought 399 events, the GPP drops below its normal conditions during the first 16 days and 400 reaches its maximum reduction within 24 days. Due to the influence of ecosystem 401 respiration, the responses of NEP for DBF and ENF to flash droughts are much 402 quicker than GPP, implying that the sensitivity of ecosystem respiration is less than 403 that of vegetation photosynthesis (Granier et al., 2007). Eventually, 81% of soil 404 moisture flash drought events cause declines in GPP. During the drought period, 405 plants would close their stomata to minimize water loss through decreasing canopy 406 407 conductance, which in turn leads to a reduction in carbon uptake. The soil moisture flash droughts are always accompanied by high temperature and VPD. Through 408

partial correlation analysis, positive anomalies in VPD and the deficits in soil 409 moisture would both decrease the rate of photosynthesis. High VPD further reduces 410 411 canopy conductance to minimize water loss at the cost of reducing photosynthesis during soil moisture flash drought (Grossiord et al., 2020). The suppression of GPP 412 413 and ET is more obvious for flash drought recovery stage determined by soil moisture 414 than the onset stage. The discrepancy of GPP responses between different phases of 415 soil moisture flash drought may result from 1) soil moisture conditions which are drier during the recovery stage, and 2) the damaged physiological functioning for 416 417 specific vegetation types. The anomalies of uWUE for ecosystems are always positive or unchanged during soil moisture flash drought except for croplands and savannas 418 during recovery stage. The decrease in canopy conductance would limit 419 420 photosynthetic rate, however, the increase of uWUE may indicate adaptative regulations of ecosystem physiology which is consistent with Beer et al. (2009). 421 uWUE is higher than WUE during onset stage of soil moisture flash drought, which is 422 423 due to the decreased conductance under increased VPD. However, there is no obvious difference between WUE and uWUE during recovery stage, which indicates that 424 photosynthesis is less sensitive to stomatal conductance and may be more correlated 425 with limitations of biochemical capacity (Flexas et al., 2012; Grossiord et al., 2020). 426 During 8 days after the soil moisture flash drought, the anomalies of GPP and ET are 427 still negative, indicating that the vegetation does not recover immediately after the 428 soil moisture flash drought. The legacy effects of flash droughts may be related to the 429 vegetation and climate conditions (Barnes et al., 2016; Kannenberg et al., 2020). 430

This study is based on the sites that are mainly distributed over North America 431 and Europe. It is necessary to investigate the impact of flash drought on vegetation 432 433 over other regions with different climates and vegetation conditions. In addition, this study used in-situ surface soil moisture at FLUXNET stations to detect vegetation 434 response due to the lack of soil moisture observations at deep soil layers. There would 435 be more significant ecological responses to flash drought identified through using 436 root-zone soil moisture because of its close link with vegetation dynamics. Due to the 437 438 limitation of FLUXNET soil moisture measurements, here we used soil moisture 439 observations mainly at the depths of 5 to 10 cm. We also analyzed the response of GPP to flash drought identified by 0.25-degree ERA5 soil moisture reanalysis data at 440 441 the depths of 7cm and 1m. The response of GPP to flash droughts identified by 442 FLUXNET surface soil moisture are quite similar to those identified by ERA5 soil moisture at the depth of 1m (not shown). There are less GPP responses to flash 443 droughts identified by ERA5 surface soil moisture. Although we select the ERA5 grid 444 445 cell that is closest to the FLUXNET site and use the ERA5 soil moisture data over the 446 same period as the FLUXNET data, we should acknowledge that the gridded ERA5 data might not be able to represent the soil moisture conditions as well as flash 447 droughts at in-situ scale due to strong heterogeneity of land surface. Therefore, the 448 in-situ surface soil moisture from FLUXNET is useful to identify flash droughts 449 compared with reanalysis soil moisture, although the in-situ root-zone soil moisture 450 451 would be better.

452 **4.2 Variation in ecological responses across vegetation types**

The responses of GPP, ET and WUE to soil moisture flash drought vary among 453 different vegetation types. The decline in GPP and ET only occurs across croplands 454 455 and savannas during onset stage. For most forests, the deterioration of photosynthesis and ET appears during the recovery stage with higher WUE and uWUE. For CROP 456 457 and SAV, both WUE and uWUE decrease during the recovery stage and they may be brown due to reduced photosynthesis. The positive anomalies of WUE and uWUE for 458 forests suggest that their deeper roots can obtain more water than grasslands during 459 flash drought. Xie et al. (2016) pointed out that WUE and uWUE for a subtropical 460 461 forest increased during the 2013 summer drought in southern China. The increased WUE in forest sites and unchanged WUE in grasslands were also found in other 462 studies for spring drought (Wolf et al., 2013). In general, herbaceous plants are more 463 464 sensitive to flash drought than forests, especially for savannas. The correlation between soil moisture and GPP is more significant for SAV, CROP, and ENF during 465 onset stages of flash droughts, which is consistent with the strong response to water 466 467 availability of SAV and CROP (Gerken et al., 2019). SAV is more isohydric than forests and would reduce stomatal conductance immediately to prohibit water loss 468 that further exacerbates drought (Novick et al., 2016; Roman et al., 2015). However, 469 almost all vegetation types show high sensitivity to VPD during the recovery stage of 470 flash droughts. 471

472 **4.3 Potential implications for ecosystem modelling**

The study reveals the profound impact of soil moisture flash droughts on ecosystem through analyzing eddy covariance observations. It is found that the

responses of carbon and water exchanges are quite distinguishing for forests and 475 herbaceous plants. For the ecosystem modeling, the response of stomatal conductance 476 477 under soil moisture stress has been addressed in previous studies (Wilson et al., 2000), but there still exists deficiency to capture the impacts of water stress on carbon uptake 478 479 (Keenan et al., 2009), which is partly due to the different responses across species. Incorporating physiological adaptations to drought in ecosystem modeling especially 480 481 for forests would improve the simulation of the impact of drought on the terrestrial ecosystems. 482

483 **5. Conclusion**

This study presents how carbon and water fluxes respond to soil moisture flash 484 485 drought during 8 days before flash droughts, onset and recovery stages, and 8 days after flash droughts through analyzing decade-long observations from FLUXNET. 486 Ecosystems show high sensitivity of GPP to soil moisture flash drought especially for 487 savannas, and GPP starts to respond to soil moisture flash droughts within 16 days for 488 489 more than half of the flash drought events under the influence of the deficit in soil moisture and higher VPD. However, the responses of WUE and uWUE vary across 490 491 vegetation types. Positive WUE and uWUE anomalies for forests during the recovery stage indicate the resistance to soil moisture flash drought through non-stomatal 492 regulations, whereas WUE and uWUE decrease for croplands and savannas during the 493 recovery stage. For now, the main concern about the ecological impact of soil 494 495 moisture flash drought is concentrated on the period of flash drought and the legacy effects of flash drought are not involved. It still needs more efforts to study the 496

subsequent effects of soil moisture flash droughts which would contribute to assessing 497 the accumulated ecological impacts of flash drought. Nevertheless, this study 498 highlights the rapid response of vegetation productivity to soil moisture dynamics at 499 sub-seasonal timescale, and different responses of water use efficiency across 500 ecosystems during the recovery stage of soil moisture flash droughts, which 501 complements previous studies on the sensitivity of vegetation to extreme drought at 502 longer time scale. Understanding the response of carbon fluxes and the coupling 503 between carbon and water fluxes to drought, especially considering the effects of 504 505 climate change and human interventions (Yuan et al., 2020), might help assessing the resistance and resilience of vegetation to drought. 506

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515

516 Data availability statement

517 Carbon fluxes and hydrometeorological variables from FLUXNET2015 are available

518 through <u>https://fluxnet.fluxdata.org/data/fluxnet2015-dataset/</u>.

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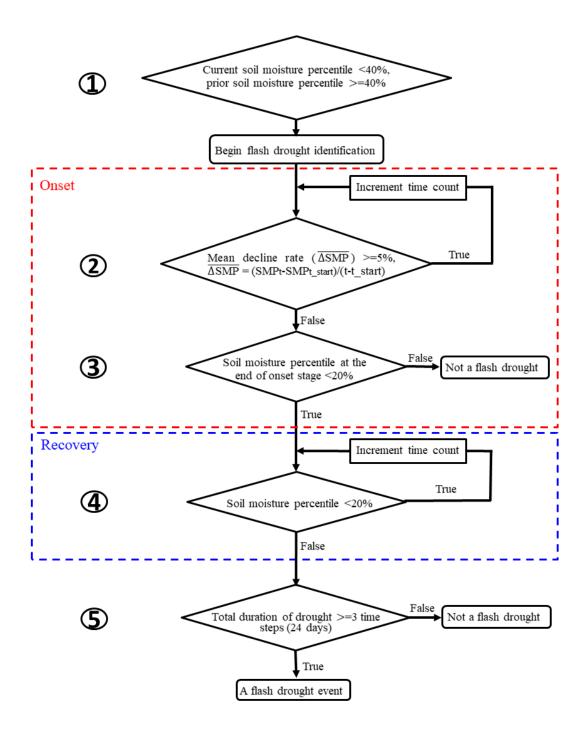


Figure 1. A flowchart of flash drought identification by considering soil moisture

818 decline rate and drought persistency.

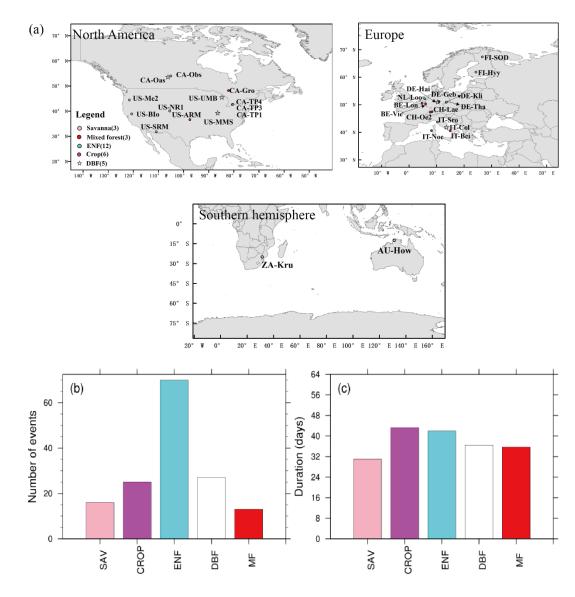


Figure 2. Global map of 29 FLUXNET sites used in this study (a) and flash drought
characteristics (b&c). (b) Total numbers (events) and (c) mean durations (days) of
flash drought events for each vegetation type during their corresponding periods (see
Table 1 for details). Different colors represent different vegetation types.

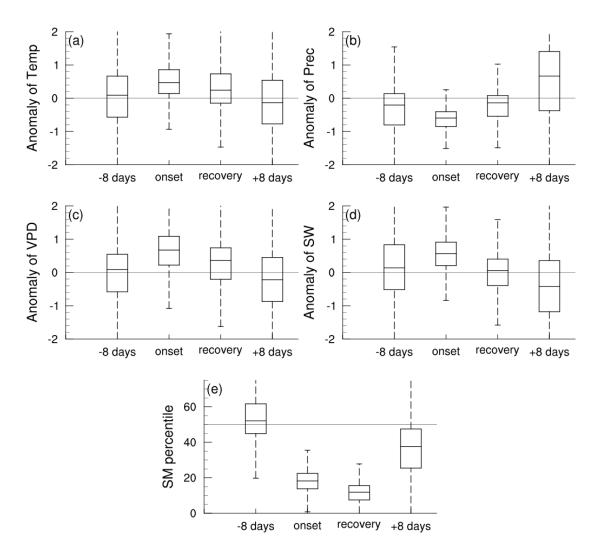


Figure 3. Standardized 8-day anomalies of (a) temperature, (b) precipitation, (c) VPD,
(d) short wave radiation (SW), and (e) soil moisture (SM) percentiles during 8 days
prior to flash drought onset, onset and recovery stages of flash drought, and 8 days
after flash drought.

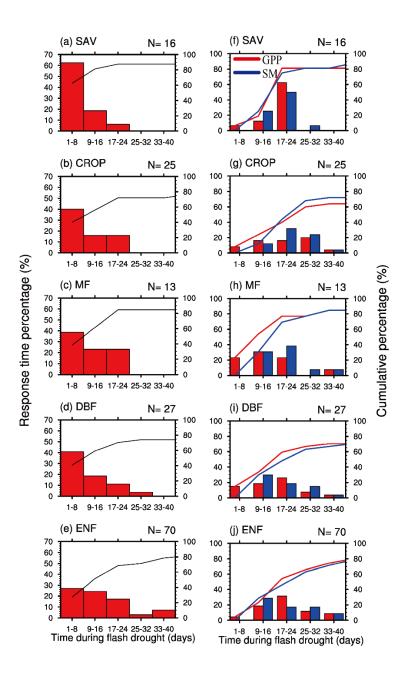


Figure 4. Percentage of the response time (days) of the first occurrence of negative
GPP anomaly (a-e), minimum GPP anomaly and minimum soil moisture percentile
(f-j) during soil moisture flash drought for different vegetation types. SAV: savanna,
CROP: rainfed cropland, MF: mixed forest, DBF: deciduous broadleaf forest and
ENF: evergreen needleleaf forest.

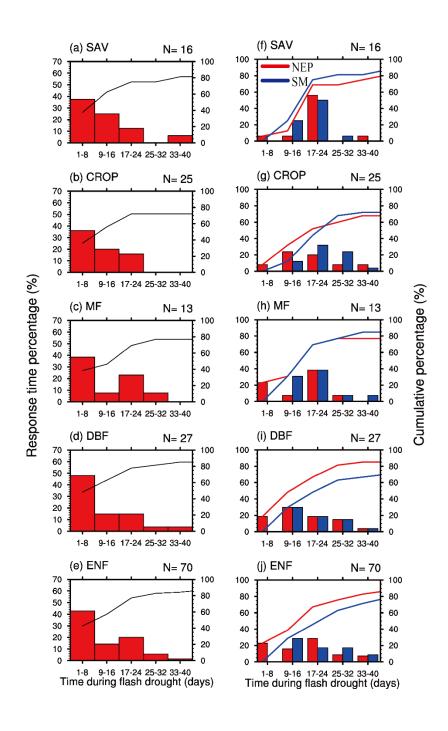




Figure 5. The same as Figure 4, but for net ecosystem productivity (NEP).

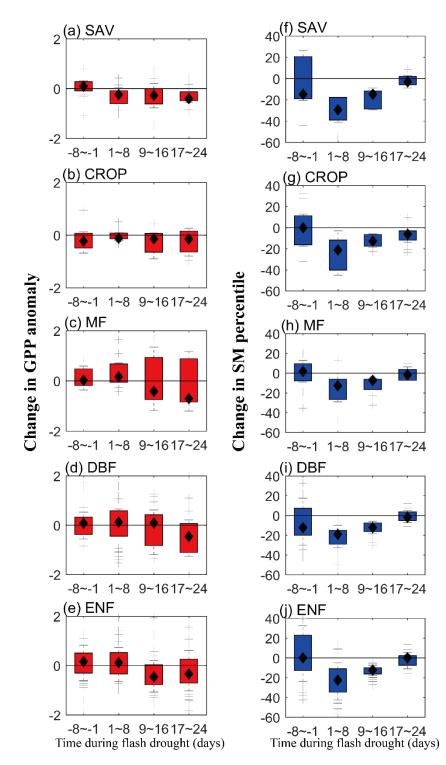




Figure 6. The temporal change rates of standardized GPP anomalies (a-e) and soil moisture percentiles (f-j) for different vegetation types. SAV: savanna, CROP: rainfed cropland, MF: mixed forest, DBF: deciduous broadleaf forest and ENF: evergreen needleleaf forest.

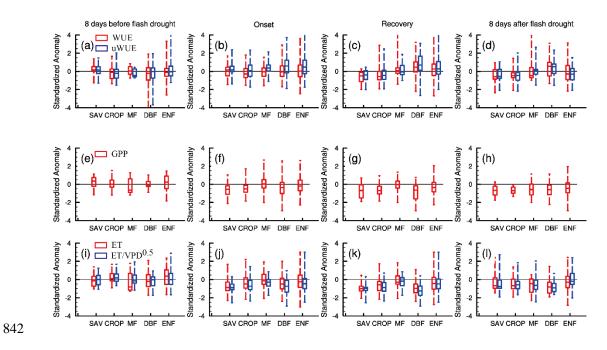


Figure 7. Standardized anomalies of water use efficiency (WUE), underlying WUE

844 (uWUE), GPP, ET and ET/\sqrt{VPD} during 8 days before flash drought onset, onset 845 and recovery stages of flash drought events, and 8 days after flash drought.

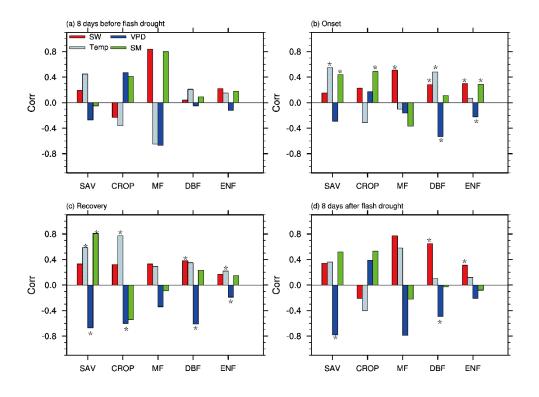


Figure 8. The partial correlation coefficients between GPP and soil moisture (SM), shortwave radiation (SW), temperature (Temp) and vapor pressure deficit (VPD) for different vegetation types including savannas (SAV), rain-fed croplands (CROP), mixed forests (MF), deciduous broadleaf forests (DBF), and evergreen needleleaf forests (ENF) during 8 days before soil moisture flash drought, onset and recovery stages and 8 days after soil moisture flash drought. * indicates the correlation is statistically significant at the 95% level.

Table 1. Locations, vegetation types and data periods of Flux Tower Sites used in this
study. WSA: woody savanna; CROP: cropland; EBF: evergreen broadleaf forests; MF:
mixed forest; DBF: deciduous broadleaf forest; ENF: evergreen needleleaf forest;

station	lat	lon	IGBP	period
AU-How	-12.49	131.15	WSA	2002-2014
BE-Lon	50.55	4.75	CROP-rainfed	2004-2014
BE-Vie	50.31	6.00	MF	1997-2014
CA-Gro	48.22	-82.16	MF	2004-2013
CA-Oas	53.63	-106.20	DBF	1996-2010
CA-Obs	53.99	-105.12	ENF	1999-2010
CA-TP1	42.66	-80.56	ENF	2002-2014
CA-TP3	42.71	-80.35	ENF	2002-2014
CA-TP4	42.71	-80.36	ENF	2002-2014
CH-Lae	47.48	8.37	MF	2005-2014
CH-Oe2	47.29	7.73	CROP-rainfed	2004-2014
DE-Geb	51.10	10.91	CROP-rainfed	2001-2014
DE-Hai	51.08	10.45	DBF	2000-2012
DE-Kli	50.89	13.52	CROP-rainfed	2005-2014
DE-Tha	50.96	13.57	ENF	1997-2014
FI-Hyy	61.85	24.29	ENF	1997-2014
FI-Sod	67.36	26.64	ENF	2001-2014
IT-Bci	40.52	14.96	CROP-irrigated	2005-2014
IT-Col	41.85	13.59	DBF	2005-2014
IT-Sro	43.73	10.28	ENF	2000-2012
NL-Loo	52.17	5.74	ENF	1999-2013
US-ARM	36.61	-97.49	CROP-rainfed	2003-2013
US-Blo	38.90	-120.63	ENF	1998-2007
US-Me2	44.45	-121.56	ENF	2002-2014
US-MMS	39.32	-86.41	DBF	1999-2014
US-NR1	40.03	-105.55	ENF	2002-2014
US-SRM	31.82	-110.87	WSA	2004-2014
US-UMB	45.56	-84.71	DBF	2002-2014
ZA-Kru	-25.02	31.50	SAV	2000-2010

857 GRA: grassland; SAV: savanna.