

1 **Assessing the influence of soil freeze-thaw cycles on catchment water storage – flux – age interactions** 2 **using a tracer-aided ecohydrological model**

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11

12 **Abstract**

13 Ecohydrological models are powerful tools to quantify the effects that independent fluxes may have on catchment storage
14 dynamics. Here, we adapted the tracer-aided ecohydrological model, EcH₂O-iso, for cold regions with the explicit
15 conceptualisation of dynamic soil freeze-thaw processes. We tested the model at the data-rich Krycklan site in northern Sweden
16 with multi-criteria calibration using discharge, stream isotopes and soil moisture in 3 nested catchments. We utilized the model's
17 incorporation of ecohydrological partitioning to evaluate the effect of soil frost on evaporation and transpiration water ages, and
18 thereby the age of source waters. The simulation of stream discharge, isotopes, and soil moisture variability captured the seasonal
19 dynamics at all three stream sites and both soil sites, with notable reductions in discharge and soil moisture during the winter
20 months due to the development of the frost front. Stream isotope simulations reproduced the response to the isotopically-depleted
21 pulse of spring snowmelt. The soil frost dynamics adequately captured the spatial differences in the freezing-front throughout the
22 winter period, despite no direct calibration of soil frost to measured soil temperature. The simulated soil frost indicated a maximum
23 freeze-depth of 0.25 m below forest vegetation. Water ages of evaporation and transpiration reflect the influence of snowmelt-
24 inputs, with a high proclivity of old water (pre-winter storage) at the beginning of the growing season and a mix of snowmelt and
25 precipitation (young water) toward the end of the summer. Soil frost had an early season influence of the transpiration water ages,
26 with water pre-dating the snowpack mainly sustaining vegetation at the start of the growing season. Given the long-term expected
27 change in the energy-balance of northern climates, the approach presented provides a framework for quantifying the interactions
28 of ecohydrological fluxes and waters stored in the soil and understanding how these may be impacted in future.

29 **1. Introduction**

30 Northern watersheds are sensitive hydrologic sites where a significant proportion of the annual water balance is controlled
31 by the spring melt period (Kundzewicz et al., 2007) and can thus be key sentinels for detecting climate change impacts (Woo,
32 2013). Recent data and long-term climate projections indicate a significant increase in warming for extensive areas of boreal forests
33 currently experiencing low-energy, low-precipitation hydroclimatic regimes (Pearson et al., 2013). The limited number of long-
34 term monitoring sites with high-quality data is a concern because it may prove difficult to document the anticipated hydrological
35 change in these catchments (Laudon et al., 2018; Tetzlaff et al., 2015). Within changing northern catchments, with high water loss
36 due to transpiration ($\sim 48 \pm 13\%$) (Schlesinger and Jasechko, 2014), and significant influence of evapotranspiration (*ET*) fluxes on
37 streamflow (Karlsen et al., 2016a), the long-term ecohydrological implications of vegetation adaptation, plant water use, and the
38 water sources that sustain growth are crucial to understand and quantify. Vegetation in boreal regions also exerts strong influences

39 on the energy balance of such catchments, with low leaf area index (LAI) conifer forests and shrubs affecting the surface albedo,
40 snow interception and affecting the timing and duration of the largest input fluxes of water during snowmelt (Gray and Male,
41 1981). However, the interactions between soil water storage and “green water” fluxes of transpiration and evaporation are poorly
42 constrained in northern regions, and the way in which sources of water from inputs of snowmelt and summer rainfall mix and
43 sustain plant growth is only just beginning to be understood (Sprenger et al., 2018a). Assessment of these interactions in northern
44 catchments is further complicated by large temperature variations, and the resulting stagnation of hydrological processes induced
45 by frequent frozen ground conditions. With increasing temperatures and potential changes to the winter soil freeze-thaw dynamics
46 (e.g. Venäläinen et al., 2001), it is important to establish how these affect current vegetation-soil water interactions to project the
47 implications of future change.

48 The intricate complexities of changes in the land surface energy balance, temporal changes in sub-surface storage due to frost
49 conditions, and vegetation and soil water usage (transpiration and soil evaporation, respectively), are notoriously challenging to
50 continuously monitor (Maxwell et al., 2019), particularly in northern environments, where site access is typically remote and
51 extreme cold can limit in-situ monitoring devices. In these circumstances, the fusion of sparsely available data with hydrological
52 models is an effective method to quantify water fluxes and storage dynamics at different temporal and spatial scales. While the
53 calibration of such models requires significant hydrometric data inputs, recent work has shown that incorporation of stable isotopes
54 can be an effective tool for constraining the model estimations of storage – flux interactions in the absence of direct in-situ
55 measurements. Such models include (but are not limited to); the STARR (Spatially distributed Tracer-Aided Rainfall-Runoff)
56 model (van Huijgevoort et al., 2016), which was developed for tracer-aided simulations and calibration, and adapted for additional
57 cold-regions processes (Ala Aho et al., 2017a and b; Piovano et al., 2018), CRHM (Cold Regions Hydrologic Model) specific for
58 cold regions (Pomeroy et al., 2007), but not currently using tracers, the isoWATFLOOD model (Stadnyk et al., 2013), which has
59 been used to isolate water fluxes with tracer-aided modelling in large-scale applications in northern regions of Canada, and the
60 EcH₂O-iso model (Maneta and Silverman, 2013; Kuppel et al., 2018 a and b), which was developed as a process-based, coupled
61 atmosphere-vegetation-soil energy balance ecohydrologic model, and modified to incorporate isotopic tracers (stable isotopes
62 deuterium and oxygen-18, $\delta^2\text{H}$ and $\delta^{18}\text{O}$, respectively). However, apart from EcH₂O-iso, which explicitly conceptualises short-
63 term (diurnal and seasonal) and long-term (growth-related) vegetation dynamics and biomass productivity, most of these existing
64 models were mainly developed with a focus on runoff generation (“blue water” fluxes). Consequently, they have very simplistic
65 representation of vegetation – soil – water interactions, estimating *ET* by approximating the physical transpiration controls of
66 vegetation (e.g. Penman-Monteith and Priestley-Taylor methods) and partitioning fluxes after estimation of actual *ET* (Fatichi et
67 al., 2016).

68 Currently, EcH₂O-iso, already incorporates some cold region processes, namely snowpack development, a snowmelt routine,
69 and the influence of temperature effects on vegetation productivity. While the depth of the snowpack is not directly estimated (only
70 snow water equivalent is tracked), the surface energy balance incorporates snowpack heat storage to estimate the warming phase
71 with effective snowmelt timing (Maneta and Silverman, 2013). The model additionally estimates soil temperature, however,
72 freezing temperature and soil frost development are adaptations that are needed for use in catchments with extensive freezing
73 conditions. Soil freeze-thaw has the potential to significantly influence soil moisture conditions, tracer dynamics, and the
74 magnitude and ages of all water fluxes. Soil freeze-thaw cycles have been estimated with a variety of methods, ranging from the
75 Stefan equation (i.e. cumulative freezing temperatures) to the more physically based Richards-Fourier calculations (e.g. Zhang et
76 al., 2007; Zhang et al., 2011). The simplicity of the Stefan equation is useful in many circumstances, including where computational
77 efficiency is important (as in with EcH₂O-iso) which restricts the use of small time and space steps of physically based models.
78 Additionally, the Stefan equation works well in most environments when soil latent heat is much larger than the sensible heat, and

79 there are linear gradients of soil temperature (Jumikis, 1977). The incorporation of tracer dynamics to EcH₂O-iso open
80 opportunities to strengthen the evaluation of the model processes (Kuppel et al., 2018b) and permits the use of tracers in calibration
81 (Douinot et al., 2019). Here, our overall aim was to provide a framework for assessing vegetation influences on the hydrology of
82 cold-regions by adapting the EcH₂O-iso model and testing it in the intensively monitored Krycklan catchment in northern Sweden.
83 The specific objectives of the study are three-fold; 1) to assess the capability of a spatially distributed, physically-based
84 ecohydrological model to capture the influence of snow and soil freeze-thaw processes on water storage dynamics, and the resulting
85 flux magnitudes under different vegetation communities (forest vs mire), 2) To examine the influence of soil frost on the dynamics
86 and age of water fluxes within the catchment, and 3) provide a generic modelling approach for application to other frost affected
87 catchments. In the adaptation of EcH₂O-iso to cold regions and the assessment of the simulated vegetation-soil water interactions
88 with frost conditions, we aim to improve the understanding and projecting the future role of vegetation in cold regions hydrology.

89 **2. Model description and extensions for this paper**

90 **2.1 EcH₂O-iso model**

91 Recent advances in hydrological modelling have included more explicit process-based conceptualisation of ecohydrological
92 interactions (Fatichi et al., 2016) and the integration of tracer-based data (Birkel and Soulsby, 2015). The EcH₂O model (Maneta
93 and Silverman, 2013) was developed as an ecohydrological model coupling land-surface energy balance models with a physically-
94 based hydrologic model. This explicitly includes the dynamics of vegetation growth and vertical and lateral ecohydrological
95 exchanges.

96 *2.1.1 EcH₂O energy balance*

97 The energy balance is computed for two-layers, the canopy, and surface. The solution of the energy balance is used to calculate
98 the available energy reapportionment for transpiration, interception evaporation, soil evaporation, snowmelt, ground heat storage,
99 and canopy and soil temperature. The canopy energy balance is iteratively solved at each time step until canopy temperature
100 converges to the estimated value that balances radiative (incoming and outgoing short and long wave radiation), and turbulent
101 energy fluxes (sensible and latent heat) (Maneta and Silverman, 2013; Kuppel et al., 2018 a and b). Long- and shortwave radiation
102 transmitted through the canopy to the soil and longwave radiation emitted by the canopy toward the ground drive the surface
103 energy balance. The surface energy balance components include radiative exchanges (incoming and outgoing short and long-wave
104 radiation), sensible, latent, and ground heat fluxes, as well as heat storage in the soil and in the snowpack. Soil evaporation is
105 estimated with the latent heat, using the atmospheric conditions (air density and heat capacity), soil resistance to evaporation, and
106 the aerodynamic resistance (surface and canopy) of the evaporative surface (Maneta and Silverman, 2013). The transpiration is
107 estimated at the leaf and is dependent on the vapour pressure gradient from the leaf to the atmosphere, the canopy resistance to
108 vapour transport, vegetation properties, and the current soil saturation conditions (Maneta and Silverman, 2013). While the energy
109 balance apportions energy to each storage (i.e. soil and snowpack), when the snowpack is present, estimated surface temperatures
110 refer to the snowpack surface, and the surface temperature of the ground is assumed to be at the temperature of the snowpack,
111 which means that conductive heat transfer between soil and snowpack is 0 (no thermal gradient). Also, when the snowpack is
112 present latent heat for surface evaporation is set to 0. When no snowpack is present, the ground heat flux (and soil temperature) is
113 estimated with the one-dimensional (1D) diffusion equation with two thermal layers, where the bottom of the top thermal layer is
114 estimated with the thermal conductivity and heat capacity of the thermal layer (Maneta and Silverman, 2013). The 1D diffusion
115 equation is only used during the snow-free conditions since soil frost causes discontinuities in the estimation of the thermal layers.
116 While soil temperature is estimated within EcH₂O (at the interface of the first and second soil thermal layer), there is currently not
117 a freezing-routine for soil water below 0°C.

119 EcH₂O has previously been adapted to incorporate the tracking of hydrological tracers including stable isotopes (Kuppel et
 120 al., 2018b) and chloride (Douinot et al., submitted), and adapted to compute estimations of water age in water storage and fluxes.
 121 Isotopic fractionation is simulated in soil water using the Craig-Gordon model (Craig and Gordon, 1965), and tracer mixing is
 122 simulated using an implicit first-order finite difference scheme. Full details of the implementation of the isotopic module are in
 123 Kuppel et al., (2018a). These adaptations do not consider fractionation of snowmelt or open water evaporation. Water ages are
 124 estimated assuming complete mixing in each water storage compartment. Similar to other snowmelt tracer models (eg. Ala-aho et
 125 al., 2017a), the snowmelt ages are defined as the time the snow enters the catchment, rather than the time of melt. This results in
 126 older water estimations during the freshet period and a more complete estimate of the time that water has resided in the catchment.

127 2.2 Soil water freeze-thaw adaptation

128 Hydrology in cold regions can be greatly affected by the freeze-thaw cycles of soil water during the winter, resulting in
 129 reduced liquid water storage capacity during the spring melt and a restricted capability for infiltration due to the expansion of ice
 130 in pore spaces (Jansson, 1998). The depth of the soil frost can have a large influence on the timing of snowmelt runoff and provide
 131 an estimation of the liquid water available within a soil layer (Carey and Woo, 2005). The Stefan equation is a simple energy
 132 balance approach to estimate the progression of soil water freezing (Jumikis, 1977):

$$\Delta z_f = \left[\frac{2k_f(T_s - T_f)}{\lambda\theta} \right]^{1/2} \quad (1)$$

133 where Δz_f (m) is the change in depth of the frost and is a function of the thermal conductivity of the frozen soil layers between the
 134 frost depth and the soil surface (k_f , W·m⁻¹·C⁻¹), the soil surface temperature below the snowpack (T_s , °C), the temperature of freezing
 135 (T_f , °C), the latent heat of freezing (λ , J·m⁻³), and the liquid soil moisture (θ , m³·m⁻³). As with previous approaches (Jumikis, 1977;
 136 Carey and Woo, 2005), the progression of the soil frost is estimated by discretizing the total soil depth into smaller layers. Within
 137 EcH₂O-iso, the sub-surface soil regime is discretized into three soil layers, layer 1 (near the surface), layer 2, and layer 3
 138 (groundwater to bedrock), to resolve the water balance and estimate soil moisture. Here, the depths of layer 1, 2, and 3 were used
 139 as the layers since they intrinsically incorporate the soil moisture estimations without additional parameterisation. The thermal
 140 conductivity of frost affected layers is dependent on the moisture content of the soil:

$$k_f(i) = (k_{sat} - k_{dry}) \cdot \left(\frac{\theta(i)}{\phi(i)} \right) + k_{dry} \quad (2)$$

141 where $k_f(i)$ (W·m⁻¹·C⁻¹) is the thermal conductivity of frozen soil in layer i , k_{sat} (W·m⁻¹·C⁻¹) is the thermal conductivity of saturated
 142 soil, k_{dry} (W·m⁻¹·C⁻¹) is the thermal conductivity of dry soil, $\theta(i)$ (m³·m⁻³) is the soil moisture in layer i , and $\phi(i)$ (m³·m⁻³) is the soil
 143 porosity in layer i . The saturated thermal conductivity was estimated from the proportions of soil comprised of ice, liquid water,
 144 air, organic material, and mineral soil (Carey and Woo, 2005):

$$k_{sat} = \prod_{n=1}^5 k(j)^{f(j)} \quad (3)$$

145 where j is the thermal conductivity of each volume proportion, f is the fraction of total soil volume, and k is the thermal
 146 conductivity of volume j . Without proportions of soil organic and mineral material, the bulk soil thermal conductivity (k_{dry}) is
 147 considered the weighted average of organic and mineral thermal conductivity (only 4 total volumes in Eqn 3). Implementation of
 148 Eqns 1-3 for freezing layers below the ground surface are ideal for EcH₂O as the model estimates the parameters (T_s , and θ) or
 149 includes parameterisation of physical properties (λ , k_{dry} , ϕ , k_{water} , k_{air}), and only requires the addition of the thermal conductivity
 150 of ice (2.1 W·m⁻¹·C⁻¹, Waite et al., 2006). Within EcH₂O, the estimation of surface temperature (above the snowpack) is assumed
 151 to be isothermal with the snowpack and conduction through the snowpack is not considered. However, conduction through

152 snowpack is important for the Stefan equation (Eqn 1) as the surface temperature used is below the snowpack which is generally
 153 thermally insulated by the snowpack. To address the conduction through the snowpack without snow depth or density, the
 154 estimated surface temperature above the snowpack (T_{Est}) was damped with a single unitless parameter (D) such that $T_s = T_{Est} \cdot D$.

155 To account for the reduction of the infiltration rate due to ice, models have previously adjusted the soil hydraulic conductivity
 156 (e.g. Jansson, 1998). Here, the reduction in hydraulic conductivity is estimated using an exponential function:

$$K_{wf} = 10^{fc \cdot F} K_{sat} \quad (4)$$

157 where K_{wf} ($\text{m} \cdot \text{s}^{-1}$) is the hydraulic conductivity of the soil influenced by ice, K_{sat} ($\text{m} \cdot \text{s}^{-1}$) is the saturated hydraulic conductivity of
 158 ice-free soils, fc is a unitless ice-impedance parameter, and F is the fraction of frost depth to total soil depth. Equation (4) has two
 159 key assumptions: no ice lenses or frost heaving, and no soil volume expansion due to lower ice density (assumed 920 kg/m^3 at ice
 160 temperature $0\text{-}5^\circ\text{C}$).

161 **2.3 Soil frost volume, depth, and water age**

162 As soil frost progresses through the layers, the proportion of liquid water is assumed to decrease at the same rate as the
 163 proportion of unfrozen soil. Similar to other approaches estimating the moisture content of frost-affected soils (Jansson, 1998), a
 164 minimum liquid soil moisture was retained in all frozen soils. This minimum was assumed to be the residual soil moisture (θ_r),
 165 the minimum moisture content required for evaporation and root-uptake. Following the estimation of the soil water infiltration
 166 and redistribution of soil water within EcH_2O , the change in soil moisture due to freezing in each layer is estimated:

$$\Delta\theta = (\theta(i) - \theta_r) \cdot \frac{\Delta z_f}{d(i) - d_f(i)} \quad (5)$$

167 where $\Delta\theta$ ($\text{m}^3 \cdot \text{m}^{-3}$) is the change in liquid water and ice content, $\theta(i)$ ($\text{m}^3 \cdot \text{m}^{-3}$) is the initial liquid content in layer i , θ_r ($\text{m}^3 \cdot \text{m}^{-3}$) is
 168 the residual moisture content, $d(i)$ (m) is the total depth of layer i , and $d_f(i)$ (m) is the depth of frost in layer i . Step-wise
 169 estimation of freeze and thaw for each layer is provided in more detail in Appendix A. The water age of the ice is estimated in a
 170 similar way to the liquid water ages of the soil layers (Kuppel et al., 2018b):

$$V_{res}^{t+\Delta t} A_{res}^{t+\Delta t} - V_{res}^t A_{res}^t = q_{in} A_{in}^{t+\Delta t} - q_{out} A_{res}^{t+\Delta t} \quad (6)$$

171 where t is time (sec), Δt is the time-step (sec), V_{res} (m^3) is the volume of ice in storage, q_{in} (m^3) is the volume of water from the
 172 change in soil moisture during freeze-up (using $\Delta\theta$ in Eqn 5), q_{out} (m^3) is the volume of water from the change in soil moisture
 173 during thaw (from Eqn 5), and A (sec) is the water age (subscripts *res* and *in* are the water ages in storage and inflow, respectively).
 174 Similar to the isotope and vegetation modules in EcH_2O , the frost dynamics (i.e. frost depths and water ages) were implemented
 175 as an option within EcH_2O .

176 **2.4 Isotope snowmelt fractionation**

177 Isotopic fractionation of snowmelt can have a significant influence on the composition of streams (Ala-aho et al., 2018a).
 178 Previous successful applications of a simple approach equation to estimate the isotopic fractionation of snowmelt at multiple
 179 locations has shown that low-parameterised fractionation models can be used to spatially approximate snowmelt fractionation. One
 180 of the noted limitations of the simple snowmelt fractionation approach used in Ala-aho et al., (2018a), is the dependence of the
 181 snowmelt fractionation on the past snowmelt volumes (total days of melt, d_{melt}) rather than current snowmelt rate ($\delta^2 H_{melt} =$
 182 $\delta^2 H_{pack} - M_{frac}/d_{melt}$, M_{frac} is a fractionation parameter, in Ala-aho et al., 2018a). The approach was modified to include the
 183 snowmelt rate with one additional parameter using an exponential function:

$$\delta^2 H_{melt} = \delta^2 H_{pack} - \left(S \cdot \exp \left(-S \cdot \left(1 - \frac{SWE - M}{SWE_{max}} \right) \right) \right) \cdot C \quad (7)$$

184 where δ^2H_{melt} (‰) is the isotopic composition of the snowmelt, δ^2H_{pack} (‰) is the composition of the snowpack at the beginning
185 of the time-step, SWE (m) is the snow water equivalent at the current time, SWE_{max} (m) is the maximum snow water equivalent
186 before melt, M (m) is the total depth of snowmelt in the current time-step, S is a unitless slope parameter describing the shape of
187 the exponential change of the snowmelt fractionation, and C (‰) is an amplification factor. Eq (7) serves as the mass-balance for
188 the snowpack isotopes throughout the winter and spring melt period. In comparison to Ala-aho et al (2018a), the exponential form
189 works to temporally change the shape of the fractionation as a relationship to the amount of melt (i.e. replacing $1/d_{melt}$). Higher
190 values of S (10-20) result in larger early melt fractionation and limited late melt fractionation, while low values of S result in a
191 lower, but more consistent fractionation throughout the melt period. The amplification factor behaves as a simplification of
192 atmospheric effects on the snowmelt fractionation. The isotopic composition of the snowpack is updated at the end of each time-
193 step.

194 **3. Data and study site**

195 **3.1 Study site**

196 Svartberget (C7, 0.49 km²) is a small subcatchment situated in the headwaters of the Krycklan catchment (64°14'N, 19°46'E)
197 in northern Sweden. Svartberget is a well-studied site with long-term data collection including: streamflow (1991-present), stream
198 chemistry (2000-present), and hillslope transect measurements (soil moisture and water chemistry). Svartberget has two
199 subcatchments, Västrabäcken (C2, 0.12 km²) and Mire (C4, 0.18 km²) (Fig. 1). The topographic relief of C7 is 71 m (235 – 306 m
200 a.s.l.), with 57 m of relief in C2 (247 – 304 m a.s.l.) and only 26 m of relief in C4 (280 – 306 m a.s.l.) (Fig 1). The climate is
201 subarctic (in the Köppen classification index), with annual precipitation of 614 mm, evapotranspiration (ET) of 303 mm, mean
202 relative humidity of 82 %, and a 30 year mean annual temperature of 1.8 °C (Laudon et al., 2013). The relatively low topography
203 results in no observable influence of elevation on precipitation (Karlsen et al., 2016b). The catchment experiences continuous
204 snowpack development throughout the winter, accounting for approximately a third of the annual precipitation and lasting on
205 average 167 days (Laudon and Löfvenius, 2016). The large quantity of snowfall results in a dominant snowmelt-driven freshet
206 period (Karlsen et al., 2016a). Till (10 – 15 m thick) covers the majority of the downstream catchment area (C7, 92% downstream
207 of C4) with intermittent shallow soils in the headwaters of C2 (Fig. 1a). The catchment is predominantly forest covered (82% total,
208 98% downstream of C4), with Scots Pine (*Pinus sylvestris*), Norway Spruce (*Picea abies*), and Birch (*Betula spp.*). The Mire (Fig
209 1b) is dominated by *Sphagnum* mosses.

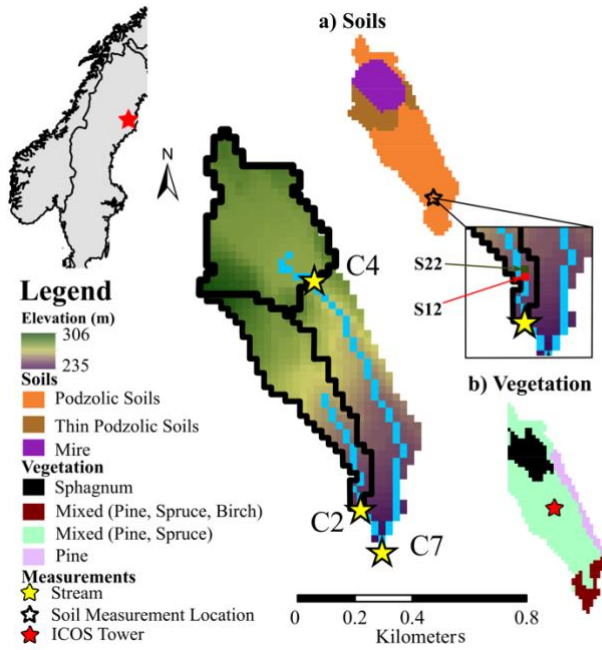


Figure 1: Location of the Svartberget within Sweden and its elevation profile with the channels and stream measurement locations (yellow). Inset figures show (a) catchment soils with the locations of S12 and S22, and (b) catchment vegetation.

3.2 Input and calibration datasets

3.2.1 Stream discharge and isotope datasets

The discharge at the three streamflow locations has been measured with hourly stream stage measurements using pressure transducers. V-notch weirs improve measurement accuracy, aided by monthly salt dilution gauging to validate results. Average discharge in the catchment varies from $9 \times 10^{-4} \text{ m}^3/\text{s}$ (C2) to $4 \times 10^{-3} \text{ m}^3/\text{s}$ at the outlet (C7), with maximum discharge events up to $0.1 \text{ m}^3/\text{s}$ (C7) during spring freshets ($0.02 \text{ m}^3/\text{s}$ and $0.03 \text{ m}^3/\text{s}$ at C2 and C4, respectively). The evaluation of the stable isotopes $\delta^2\text{H}$ and $\delta^{18}\text{O}$ of stream water were carried out for the stream water samples collected every two weeks at each site. Long-term average $\delta^2\text{H}$ is similar between streams (-95.5 , -94.5 and -95.6 ‰ for C7, C2, and C4 respectively), with the highest isotopic variability at site C4 (standard deviation (SD) of 7.9 ‰) and lowest at C2 (SD of 4.5 ‰) with C7 intermediate.

3.2.2 Meteorological datasets

Precipitation (rain and snowfall), temperature, wind speed, and relative humidity were measured daily at the Svartberg meteorological station, 150 m southwest of the catchment. Radiation data, incoming longwave and shortwave radiation, were obtained at 3-hourly time-steps and $0.75 \times 0.75 \text{ km}$ grid resolution from ERA-Interim climate reanalysis (Dee et al., 2011). During the study, a 150 m observation tower (Integrated Carbon observation system, ICOS Tower) was installed within the catchment. Data from the ICOS tower were available from 2014 to 2015. The ICOS tower measures energy fluxes, latent and sensible heat, and net radiation, among other atmospheric parameters. The isotopic composition of precipitation was determined on daily bulk samples following each major rain and snow event. The average precipitation $\delta^2\text{H}$ (weighted mean -95.1 ‰) is similar to the stream isotopic composition, though the isotopic variability is between 4.4 – 7.8 times larger.

Table 1: Datasets used for forcing, calibration and validation within the Svartberg catchment

Input Meteorological Forcing Data			
	Location	Resolution	Time-Period
Air Temperature (minimum, maximum, and mean) ($^{\circ}\text{C}$)	Svartberg	Daily	2005-2016
Precipitation ($\text{m}\cdot\text{s}^{-1}$)	Svartberg	Daily	2005-2016

Wind speed ($\text{m}\cdot\text{s}^{-1}$)	Svartberg	Daily	2005-2016
Relative Humidity (%)	Svartberg	Daily	2005-2016
Longwave Radiation ($\text{W}\cdot\text{m}^{-2}$)	ERA-interim	Daily	2005-2016
Shortwave Radiation ($\text{W}\cdot\text{m}^{-2}$)	ERA-interim	Daily	2005-2016
$\delta^2\text{H}$ Precipitation (‰)	Svartberg	Event-based	2005-2016
Calibration and Validation Datasets			
	Location	Resolution	Time-Period
Discharge ($\text{m}^3\cdot\text{s}^{-1}$)	C7	Hourly	2005 – 2016
	C2	Hourly	2005 – 2016
	C4	Hourly	2005 – 2016
Stream isotopes (‰)	C7	Biweekly	2005 – 2016
	C2	Biweekly	2005 – 2016
	C4	Biweekly	2005 – 2016
Soil Moisture ($\text{m}^3\cdot\text{m}^{-3}$)	S12	Hourly at depth of 5, 10, 20, 30, 40, 60 cm	2013 – 2016
	S22	Hourly at depth of 6, 12, 20, 50, 60, 90 cm	2013 – 2016
Validation Only Datasets			
	Location	Resolution	Time-Period
Soil Isotopes (Lysimeter, ‰)	S12	9 samples: 10, 20, 30, 40, 60, and 70 cm	2012
	S22	9 samples: 10, 20, 35, 50, 75, and 90 cm	2012
Soil Isotopes (Bulk Water, ‰)	S12		2015 – 2016
	S22	7 samples: 10, 20, 30, 40, 60, and 70 cm	2015 – 2016
Soil Temperature ($^{\circ}\text{C}$)	ICOS Tower	30 min @ 4 locations at depths 5, 10, 15, 30, and 50 cm	2014 – 2015
Net Radiation ($\text{W}\cdot\text{m}^{-2}$)		30 min	
Latent Heat ($\text{W}\cdot\text{m}^{-2}$)		30 min	
Sensible Heat ($\text{W}\cdot\text{m}^{-2}$)		30 min	

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3.2.3 Soil moisture and isotope datasets

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Soil moisture sensors were installed in 1997 and replaced at the beginning for 2013. The soil moisture sensors were installed at the hillslope transect location at 4, 12, 22, and 28 m locations from the C2 stream. The depths of the soil moisture measurements slightly differ between sites (Table 1); however, the depths encompass shallow and deep soil waters. Soil sensors have also been installed in the area surrounding the ICOS tower, measuring soil temperature at 4 locations and 6 depths (10, 20, 30, 40, 60, and 70 cm, Appendix B) (Table 1) which can provide a proxy for the depth of the frost. Soil isotopes ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) were measured at multiple depths (2.5 cm increments) measured via lysimeters (2012) and bulk water samples (2015 – 2016).

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3.3 Model set-up and calibration

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All simulations were conducted on a daily time-step between January 2005 and September 2016. The period from January 2005 to December 2009 was used as a spin-up period with measured hydrologic data, to stabilize $\delta^2\text{H}$, $\delta^{18}\text{O}$ composition, and water ages in each of the model storage units. Initial analysis of the measured discharge from 2000-2016 revealed the highest and lowest annual discharge years were between 2010 and 2014. Consequently, calibration was carried out for the period between January 2010 and December 2014. The validation set used was the remaining period from January 2015 – September 2016. The C7 catchment was defined with a grid resolution of $25 \times 25 \text{ m}^2$ to balance adequate differentiation of multiple locations on the soil water transect while maintaining computational efficiency. The 25 m grid includes adjacent soil pixels for S12 and S22, with sites S04 and S28 within the same grids as S12 and S22, respectively. Within the biomass module, the vegetation dynamics for leaf growth and carbon allocation were held at steady state to minimize the parameterisation and focus on the soil freeze-thaw cycles. As temperature effects and water stress are less sensitive for conifer trees, a relatively constant leaf area index and needle growth/decay rate were maintained (Liu et al., 2018). Evaporative soil water fractionation was activated using similar parameterisation to Kuppel et al. (2018b), as this has previously been identified as an influential summer process in the catchment

253 (Ala-aho et al., 2017a). Soil relative humidity was estimated using Lee and Pielke’s (1992) approach, and values of kinematic
254 diffusion were estimated as presented by Vogt (1976) (0.9877 and 0.9859 for H^2/H^1 and O^{18}/O^{16} ratios, respectively).
255 Parameterisation of the model was conducted for each soil type (3 soil types, Fig 1a) and vegetation type (4 types, Fig 1b).

256 A sensitivity analysis established the most sensitive parameters to be used in calibration using the Morris sensitivity analysis
257 (Soheir et al., 2014). Parameters were assessed using 10 trajectories using a radial step for evaluating the parameter space. The
258 parameter sensitivity was evaluated using the mean absolute error. Results of this are shown in Appendix C. Sensitive parameters
259 were calibrated using Latin Hypercube sampling (McKay et al., 1979) with 150,000 parameter sets and a Monte Carlo simulation
260 approach to optimize the testing of the model parameter space.

261 **3.4 Model evaluation**

262 The model output was constrained using measurements of stream discharge (3 sites, Fig. 1), stream δ^2H (3 sites, Fig. 1), and
263 soil moisture (2 sites, Fig. 1a). The 8 measurement datasets were combined into a multi-criteria calibration objective function using
264 the mean absolute error (MAE) with the cumulative distribution functions (CDFs) of the model goodness-of-fit (GOF) (Ala-aho
265 et al., 2017a; Kuppel et al., 2018 a and b). The MAE moderated over-calibration of peak flow events, typical for functions like the
266 root mean square error, and Nash-Sutcliffe efficiency, as well as being consistent with previous studies in the region (Ala-aho et
267 al., 2017a). To focus on the dynamics of soil moisture, given the coarse model grid, measured and simulated values were
268 standardized by their respective mean values prior to analysis. From the CDF method, the 30 “best” simulations were selected for
269 evaluation and are presented using 95% spread of predictive uncertainty (Kuppel et al., 2018b). The parameters achieved through
270 calibration are shown in Appendix D. Model results were verified against the remaining years of discharge, soil moisture, and
271 stream flow δ^2H , as well as independent time series of soil isotopes (bulk and lysimeter), net radiation, sensible heat, latent heat,
272 and frost depth (estimated from depth-dependent soil temperatures).

273 The evaluation of changes to water ages due to soil frost was conducted by comparing the ages within the catchment for
274 simulations of the 30 “best” parameter sets with and without frost. These were conducted without frost by turning frost dynamics
275 off within the model. Freeze-thaw effects on evaporation and transpiration ages were evaluated as the difference between frost and
276 non-frost simulations. Positive values indicate older water with the frost while negative values indicate older water with frost-free
277 simulations. The age differences were only considered on days when both frost and non-frost simulations simulate a flux greater
278 than 0 mm/day.

279 **4. Results**

280 **4.1 Simulation results**

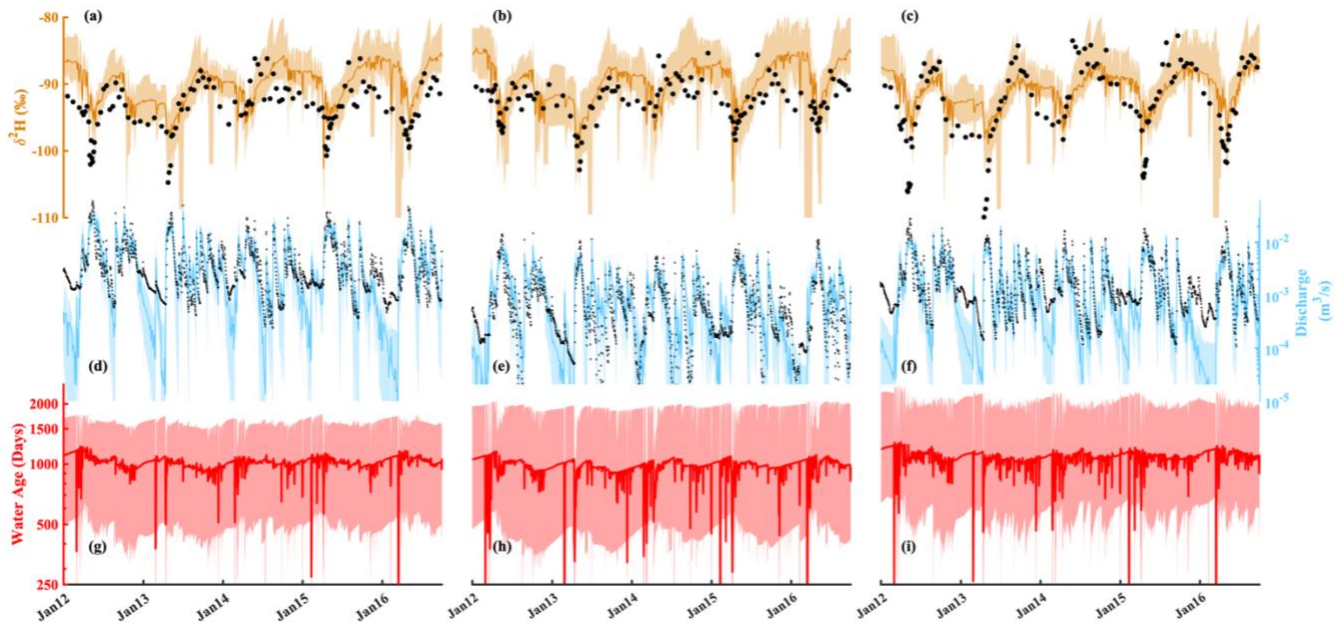
281 Calibration captured dynamics of both high and low flow discharge periods through both the calibration period (2010 – 2014)
282 and validation period (2015 – 2016), with a maximum mean stream flow MAE of $2 \times 10^{-3} \text{ m}^3/\text{s}$ for C7, and a maximum mean stream
283 δ^2H MAE of 5.8 ‰ at C4 (Table 2). Due to extreme high and low flow periods in the calibration period, it was unsurprising that
284 the resulting MAE was higher than in the validation. The MAE of the soil moisture calibration was also reasonable, with average
285 MAE of 0.05 and 0.09 for sites S12 and S22, respectively. With the standardization of the soil moisture, the low MAE indicates
286 that the dynamics in the model correspond well to those measured. The optimization of the GOF for 3 measures (discharge, stream
287 δ^2H , and soil moisture) at 8 locations resulted in a compromise for all streams. Simulations yielded better (lower) MAE for
288 discharge and isotopes of individual streams.

289 **Table 2: Calibration and validation efficiency criteria, shown as mean efficiency for all multi-calibration criteria**

		Calibration (2010 – 2014)	Validation (2015 – 2016)
	Site	MAE	MAE
Discharge	C7	$2 \times 10^{-3} \text{ m}^3/\text{s}$	$6 \times 10^{-4} \text{ m}^3/\text{s}$

	C2	$1 \times 10^{-3} \text{ m}^3/\text{s}$	$1 \times 10^{-4} \text{ m}^3/\text{s}$
	C4	$1 \times 10^{-3} \text{ m}^3/\text{s}$	$3 \times 10^{-4} \text{ m}^3/\text{s}$
$\delta^2\text{H}$	C7	4.8 ‰	4.0 ‰
	C2	4.6 ‰	3.8 ‰
	C4	5.8 ‰	3.9 ‰
Soil Moisture	S12	0.05	0.09
	S22	0.09	0.09
Latent Heat	ICOS Tower	N/A	13.1 W/m ²
Sensible Heat	ICOS Tower	N/A	29.5 W/m ²
Net Radiation	ICOS Tower	N/A	31.0 W/m ²
Soil Frost Depth	ICOS Tower	N/A	0.03 m

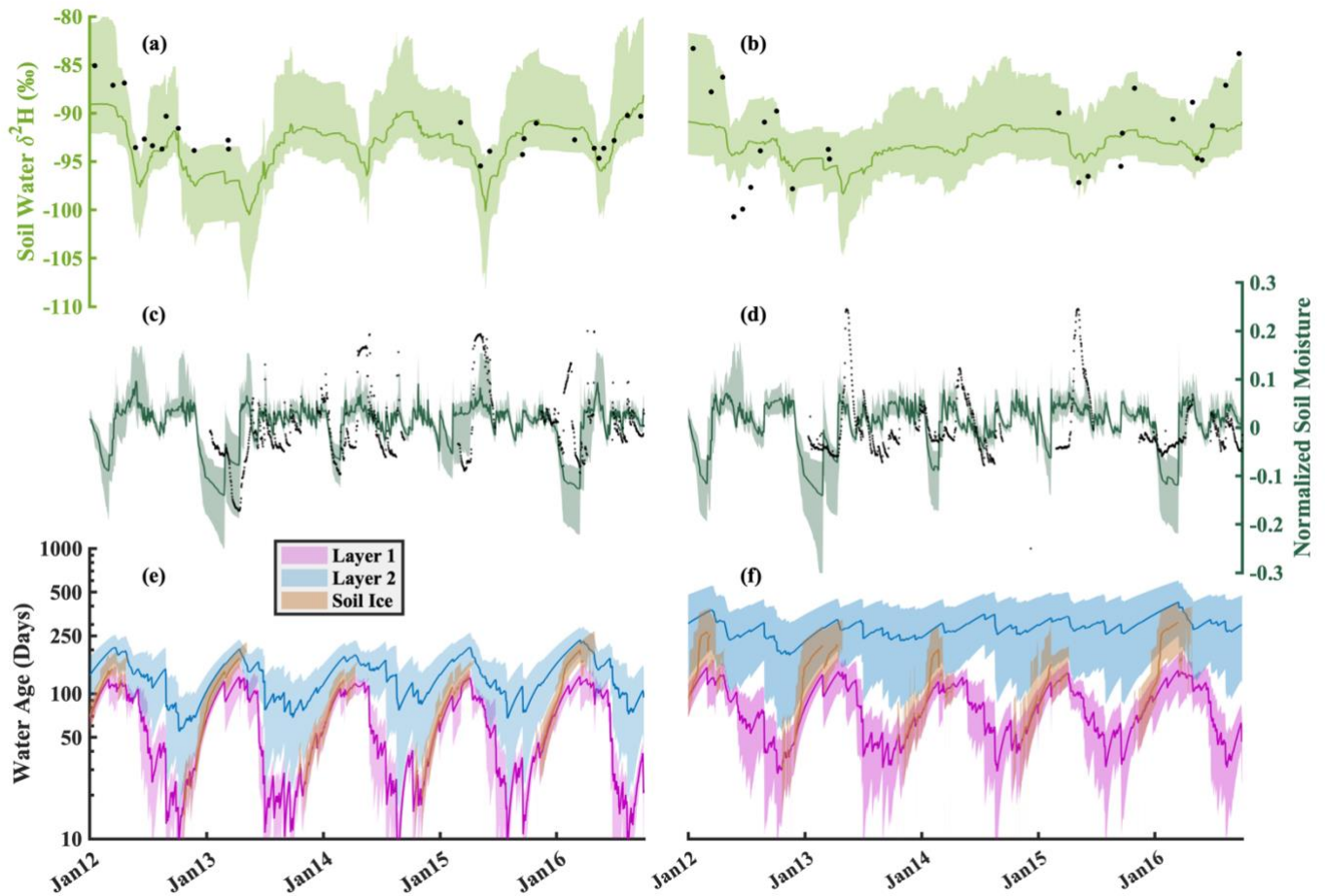
290 Temporal variability of $\delta^2\text{H}$ in each of the streams was captured quite well throughout the calibration and validation periods
 291 (Fig 2 a – c). The largest offsets in modelled isotopic composition occurred during the winter low flow conditions. The simulated
 292 stream isotopes tended to retain a slight “memory” effect from the contributions of more enriched inflow in late summer. This was
 293 likely due to the underestimation of discharge during winter (Fig 2 d – f) which slowed the flushing of the more enriched water.
 294 Overall though, discharge was adequately simulated for each site, notably during the spring melt and summer months. While flows
 295 were underestimated during the winter, the difference between simulations and measurements were typically $< 1 \times 10^{-3} \text{ m}^3/\text{s}$. The
 296 weight-median water ages of each of the three streams were broadly similar, 2.8, 2.6, and 3.1 years for C7, C2, and C4, respectively
 297 (Fig 2 g – i). These stream ages were generally older than previous estimates, with deeper soil layers and complete mixing in each
 298 compartment tending to increase the average age. The depth of the soil layers in the peat and podzolic areas are the primary drivers
 299 for water age, with a ~1:1 relationship (Appendix E). Water age decreased during the annual freshet, driven by the younger
 300 snowmelt and frozen soil water ages (typically 150 – 200 days old). The rapid runoff during the freshet limited the long-term
 301 influence of the younger water ages on the stream water at each of the sites as older groundwater dominated low flows. Rain on
 302 snow events resulted in some rapid, yet un-sustained, influences on the soil water ages, as observed with the sudden decreases in
 303 stream water age during winter months (Fig 2 g-i; log-scale with lower bound of 250 days).



304
 305 **Figure 2: Calibration 95% maximum and 5% minimum bounds, median simulation (solid line), and measured data (black circles) of**
 306 **$\delta^2\text{H}$ for (a) C7, (b) C2, and (c) C4; discharge for (d) C7, (e) C2, and (f) C4; and stream water age for (g) C7, (h) C2, and (i) C4.**

307 **4.2 Soil moisture, isotope, and water ages**

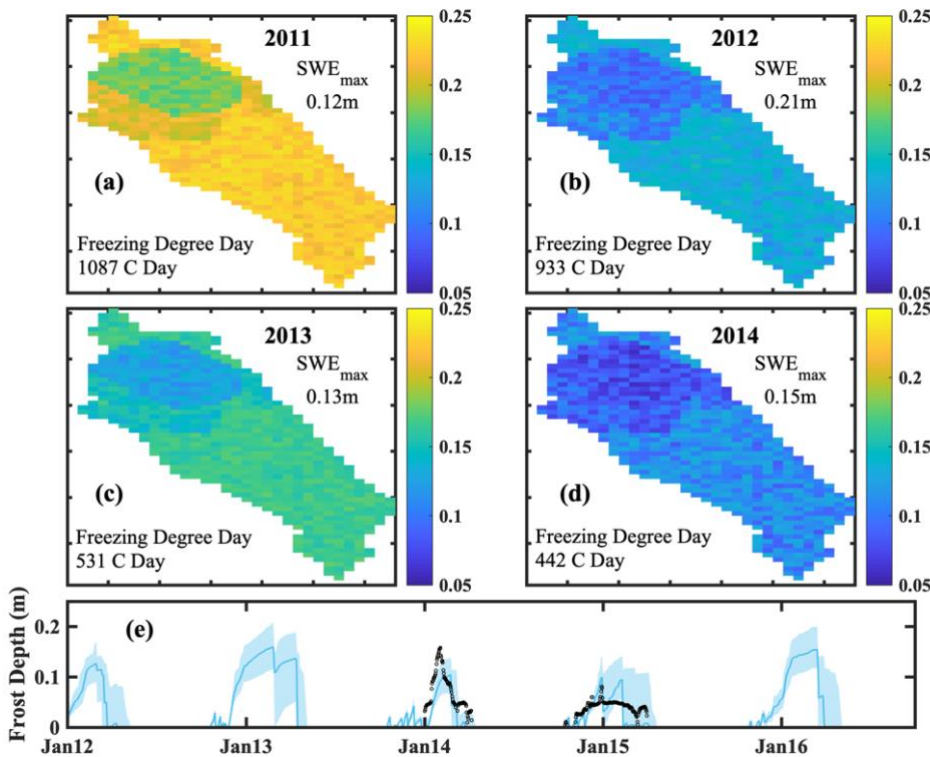
308 Simulated soil water isotopes (note that the model did not use isotopes during calibration) mostly captured those measured in
 309 both bulk water (2015 – 2016) and lysimeter water (2012) within the 90% simulation bounds at the S12 and S22 sites (Fig 3 a &
 310 b). Isotope dynamics were best captured at site S12, with early season depletion due to snowmelt and enrichment of the previous
 311 summer. While most variability was captured within the 90% bounds, the magnitude of the intra-annual contrasts at site S22 was
 312 not fully reproduced. Similar to the soil isotopes, dynamics of simulated soil moisture (calibrated) were captured at both S12 and
 313 S22, with better simulation performance at S12 (Fig 3 c & d). The model struggled to simultaneously reproduce the more damped
 314 soil moisture at S12 with the relatively dynamic soil moisture post-melt at S22 in the adjacent cell under the same soil
 315 parameterisation. Rather, the same parameterisation resulted in balancing the conditions observed at S12 and S22. The large
 316 declines in measured soil moisture during the winter months were captured with the soil frost module (Fig 3 c & d). The modelled
 317 decline in the soil moisture resulted from the transition of soil water from liquid to ice. Water ages in layers 1 and 2 at each site
 318 showed noticeable intra-annual variability, and gradually declined during the growing season (May – September) and increased
 319 during the winter due to negligible water inflow (Fig 3 e & f). The gradual decrease in the soil water age during the summer was
 320 the result of younger rainfall flushing the older snowmelt and pre-winter soil water as the growing season progressed. The
 321 variability of the soil water ages in layers 1 and 2 was similar, though the ages in layer 2 were significantly older. While S12 is
 322 closer to the stream, water ages in S22 were generally older in both layers 1 and 2.



323
 324 **Figure 3: Simulation 90% bounds and mean simulation (solid line) for the average of layer 1 and 2 $\delta^2\text{H}$ for (a) site S12, and (b) site S22;**
 325 **the average of layer 1 and 2 soil moisture standardized by the mean for (a) site S12 and (b) site S22; and water ages of soil layers 1, 2 and**
 326 **soil frost for (e) site S12, and (f) site S22.**

327 **4.3 Soil freeze-thaw simulations**

328 Simulations of frost depth revealed large inter-annual variability throughout the catchment (Fig 4 a-d), depending on winter
 329 temperatures, snowpack depth, and the soil moisture conditions. Wetter conditions in the mire generally show shallower frost
 330 depths than the podzolic soils elsewhere in the catchment. Similar soil conditions for the podzolic and thin podzolic soils (Fig 1a)
 331 resulted in negligible differences for estimated frost depth. Overall, estimated frost depth was generally limited by the total number
 332 of freezing days. Colder winters (larger numbers of freezing degree days) resulted in deeper frost depths for an equivalent snowpack
 333 depth (e.g. Fig 4a vs Fig 4c). Conversely, a deeper snowpack (higher maximum SWE) resulted in a shallower simulated frost depth
 334 for years with similar temperatures (e.g. Fig 4 a vs c) as the deeper snowpack was a larger storage for incoming radiation. Using
 335 0°C in the soil temperature probes at the ICOS tower as a proxy for the depth of the soil frost, a direct comparison of simulated
 336 frost depth and the measured catchment frost depth was completed without calibration. Simulated frost depth showed good
 337 agreement with observed 0°C soil temperature depth, imitating the rapid increase in frost depth in 2014 and a more gradual increase
 338 in 2015 (Fig 4e). Late winter soil frost depth was estimated to be shallower and varied more rapidly than the observed 0°C soil
 339 temperature depth (Fig 4e). The median estimated soil depth against the measured 0°C soil temperature depth showed that estimate
 340 soil thaw was too rapid, and thaw completed too early.

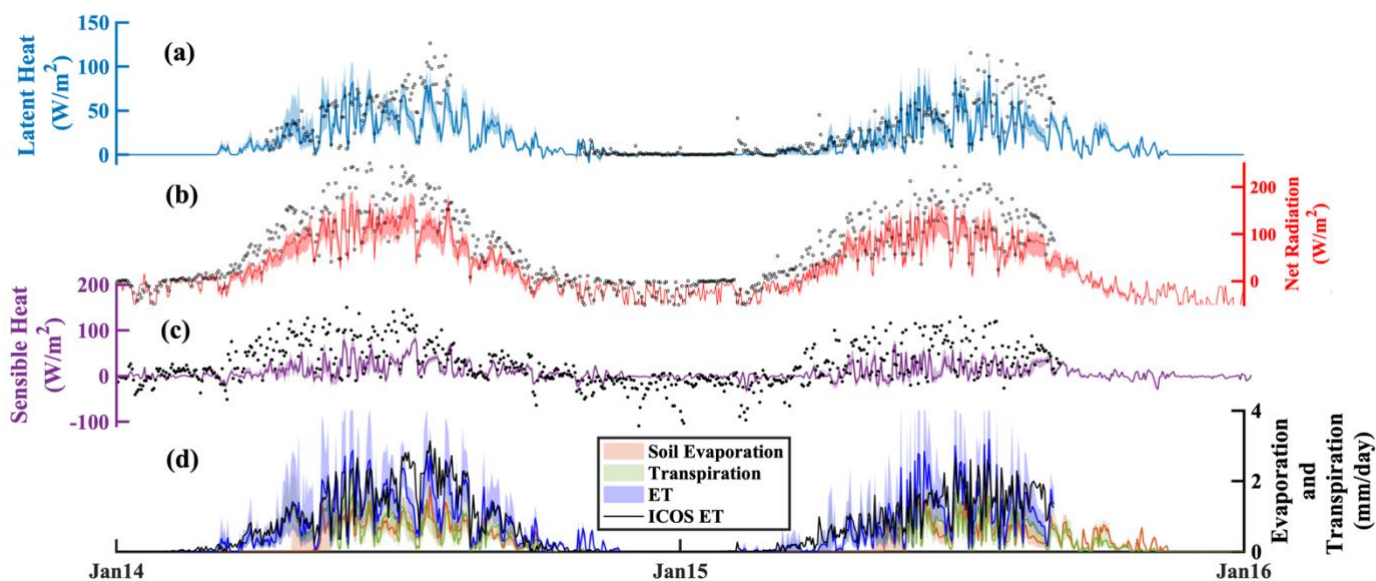


341
 342 **Figure 4: Mean simulated soil frost depth during the peak soil frost depth in winter (a) 2010-2011, (b) 2011-2012, (c) 2012-2013, and (d)**
 343 **2013-2014. 90% uncertainty bounds of the simulated frost depth at the ICOS Tower with the depth of the 0°C soil temperature measured**
 344 **at the ICOS Tower (black circles)**

345 **4.4 Evaporation and transpiration**

346 While the evaporation, transpiration, and energy balance datasets were not included in the calibration, modelled energy balance
 347 components (sensible heat, latent heat, and net radiation) showed reasonable agreement to observed values in 2014-2016 at the
 348 ICOS Tower. There was an underestimation of net radiation and sensible heat throughout the growing season (Fig 5 b & c), and
 349 an underestimation of latent heat late in the year (Fig 5a). While the MAE of the latent heat was relatively small (13.1 W/m²)
 350 considering that they were not used for calibration, net radiation and sensible heat had a notable maximum bias (~30 W/m²) during

351 summer. Simulations of total daily evaporation (soil and interception) and transpiration had a similar pattern, with transpiration
 352 accounting, on average, for 54% of total evapotranspiration. Throughout the year, the simulated proportion of transpiration to total
 353 evapotranspiration ranged from 31 – 72% except for the spring periods (Fig 5d). The late onset of evaporation resulted from the
 354 assumption that soil evaporation was negligible while the snowpack remains, which potentially lead to an underestimation of
 355 evaporation during the melt.



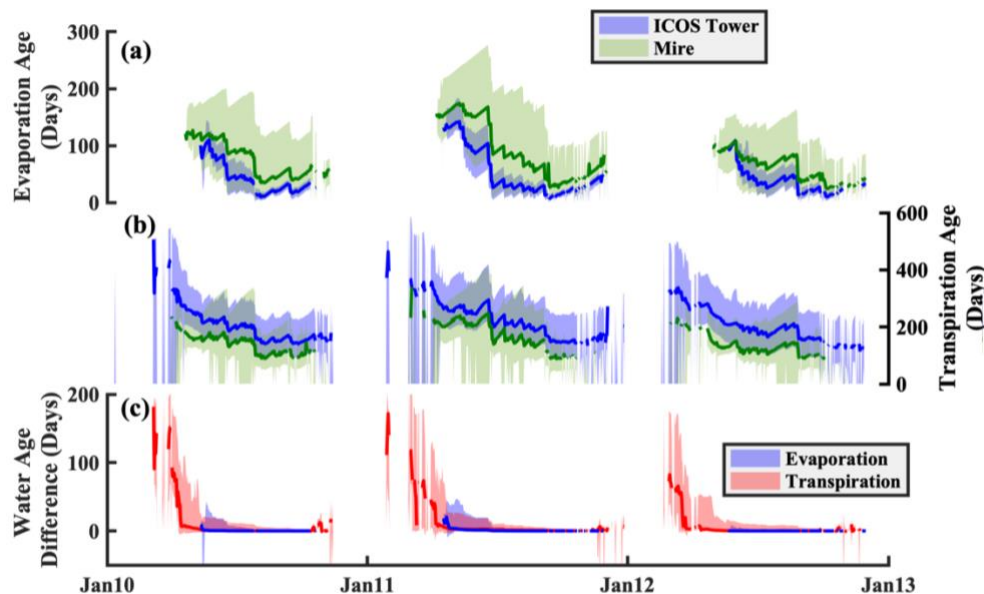
356
 357 **Figure 5: Energy balance component of (a) estimated latent heat (90% and mean values), (b) estimated net radiation (90% and median**
 358 **values), (c) estimated sensible heat (90% and median values) and (d) estimated soil evaporation, transpiration, and evapotranspiration**
 359 **(ET) (90% and mean values), at the ICOS Tower site with the estimated total evapotranspiration from energy fluxes at the ICOS Tower**
 360 **(black circles where data are available).**

361 Ages of soil evaporation and transpiration decreased throughout the year (Fig 6 a and b), tracing the decline in soil water ages
 362 estimated with the addition of precipitation (age of 0 days). Older water present in evaporation and transpiration water at the start
 363 of the year was a mixture of the snowmelt water age and frozen water ages (from the previous summer). Spatial differences in
 364 evaporation and transpiration ages were evident throughout the catchment; shown by the difference between the forested ICOS
 365 tower site (blue, Fig 6 a & b), and the average for shrubs in the mire (green, Fig 6 a & b). The annual flux-weighted median water
 366 age of transpiration was 200 ± 42 and 141 ± 40 days for the ICOS tower and mire, respectively, while evaporation ages were $48 \pm$
 367 11 and 85 ± 36 days for the ICOS tower and mire, respectively.

368 Differences between the evaporation and transpiration ages were determined by comparing water ages with the soil frost
 369 module activated, against those with the frost module deactivated. Generally, including frost in the simulations resulted in older
 370 water (water age difference > 0 Fig 6c) for both evaporation and transpiration. Differences in evaporation age were not as
 371 pronounced as transpiration ages due to the slight bias of the evaporation timing (always following the snowmelt). Due to the
 372 estimated completion of soil thaw prior to the snowmelt period, the difference between the water ages of evaporation with the
 373 influence of frozen ground was modest. Rapid flushing of the soil water due to large snowmelt inputs and spring precipitation
 374 resulted in a rapid decline in the differences of transpiration water ages. Within the first month of transpiration, the difference for
 375 the frost and non-frost simulations were more notable and approached 200 days when frost limited water movement. However, the
 376 relatively lower transpiration rates, which occurred during the spring within these simulations, resulted in a moderate effect on the
 377 overall annual transpiration water ages. The effects of soil frost on stream water ages showed little effect, with negligible
 378 differences given the relatively old water bias in the stream that only shows some flashes of younger water influence (Fig 2 g – i).

379 While the soil frost increased the stream water ages throughout the year, the effect is well within the relatively large uncertainty
380 bounds of the stream water ages.

381



382

383 **Figure 6: 90% bounds and median values of the (a) estimated soil evaporation water ages at the ICOS Tower (blue) and in the Mire (green), (b) estimated transpiration water age at the ICOS Tower (blue) and in the Mire (green), and (c) mean difference of evaporation and transpiration water ages when soil frost is not considered.**

385

386 5. Discussion

387 5.1 Modelling soil freeze-thaw processes in tracer-aided models

388 Hydrologic models are powerful tools for exploring the internal functioning of catchments, particularly when intensive and
389 long-term monitoring programs are in place to help calibration and testing (Maxwell et al., 2019). Here, the development of a
390 spatially distributed, process-based tracer-aided model for northern climates produced encouraging results reproducing soil frost
391 dynamics despite the model not being directly calibrated to match frost depths observations. The use of streamflows, stream
392 isotope ratios and soil moisture dynamics in calibration proved to be adequate for estimating the dynamics of soil frost depth and
393 timing of the frost onset (Fig 4) and revealed spatial differences in frost depth due to contrasting soil types and moisture
394 conditions. However, there are limitations with the current approach that results in some uncertainty of the effect of soil freeze-
395 thaw on catchment hydrology. To improve the computational efficiency of the model, the temperature of the snowpack was
396 assumed to be isothermal (Maneta and Silverman, 2013), and modified here to include only a single temperature damping
397 parameter. However, snowpacks may have a variable thermal gradient (e.g. Filippa et al., 2014), and is dependent on snow
398 density (e.g. Riche and Schneebeli, 2013), snow surface albedo, wind speed, and liquid water component, among others
399 (USACE, 1956; Meløysund et al., 2007; Sturm et al., 2010). While these additional components may contribute to an
400 improvement in the estimation of soil frost, it likely would not have a significant improvement compared to the simple
401 temperature damping used here with additional calibration to constrain snow water equivalent for more dynamic energy
402 exchange (e.g. Lindström et al, 2002). The simplistic consideration of negligible soil sensible heat storage effects on the soil
403 freeze/thaw processes, consistent with other process-based cold region models (e.g. CHRM, Krogh and Pomeroy, 2018), may
404 result in dampened rates of freezing and rapid melting during the spring (Kurylyk and Hayashi, 2016). More delayed melting of
405 the soil frost may have implications for snowmelt runoff, increasing the dynamics of the streamflow isotopic compositions

406 towards more depleted isotopic compositions (Fig 2 a-c). Finally, the simplification of a single soil frost front may have some
407 implications for the snowmelt infiltration to the soil. The single front does not allow for near-surface soil thaw to occur prior to
408 deeper soils and thereby has implications for shallow root-water uptake and evaporation.

409 Energy fluxes in northern catchments can be highly sensitive to the timing of snowmelt, yielding differences in the surface
410 and canopy net radiation due to changing albedo and to turbulent fluxes due to alterations in surface temperature. Here, the under-
411 estimated sensible heat flux during the spring and the growing season could be the result of many different processes. Some of
412 these processes include the aerodynamic resistance (r_a) to transpiration, an underestimated thermal gradient between the simulated
413 soil and air temperature, an overestimation of the incoming short- or long-wave radiation from the ERA-interim dataset,
414 sublimation and snowpack energy storage, or an over estimation of a heat sink of ground heat through the thermal layers of the
415 soil. An overestimated aerodynamic resistance can simultaneously reduce the transpiration (increases the surface temperature), as
416 well as decreases the sensible heat flux (Maneta and Silverman, 2013). An over estimated surface temperature can result in both a
417 decreased thermal gradient from the soil to the air used for the sensible heat flux estimation, as well as result in the shallower, and
418 earlier, simulated soil frost melt relative to the measured 0°C soil temperature depth. While the short- and long-wave radiation
419 from the dataset had no consistent deviation from the shorter measured time-series at the ICOS Tower (1:1 ratio for both short-
420 and long-wave radiation, not shown), intermittent periods of under-estimated ERA-interim radiation could contribute an
421 underestimation of the net and sensible heat flux. Lastly, the lack of snowpack heat storage and sublimation could have an influence
422 on the energy balance by limiting incoming winter radiation into the soils (i.e. decreasing soil temperatures).

423 ***5.2 Effect of soil freeze-thaw on water ages and implications for northern catchments***

424 The adaptation here of a process-based, spatially distributed model to additionally incorporate fundamental cold regions
425 processes provides both the opportunity to improve the representation of key hydrologic functions of cold catchments, and assess
426 the effect that these processes have on transit times and ages of ecohydrological fluxes. While studies in northern catchments have
427 aimed to assess the partitioning and transit or residence times of ecohydrologic fluxes (e.g. Sprenger et al., 2018a), few studies
428 have included the influence of frozen conditions on the water movement. Frozen conditions may be significant for the effective
429 transit times during the spring freshet period (Tetzlaff et al., 2018) and flow path modelling (Laudon et al., 2007; Sterte et al.,
430 2018). Traditionally, isotopic tracers water at catchment outlets have been the primary metrics for assessing the water ages in
431 streams. Here, the relatively old age of stream water, and the underestimation of soil-thaw result in only slightly older water ages
432 when soil frost conditions are considered, potentially due to the smaller proportion of wetland areas (Sterte et al., 2018). The
433 limited effect of soil frost effects on stream water age was compounded from both the wide uncertainty of the stream water ages
434 (Fig 2g-i), and the late deviation of soil water frost ages from the soil water (Fig 3e & f). Generally, the uncertainty bounds of the
435 stream water ages were greater than the difference of the soil water and soil frost ages. The smaller difference of soil water and soil
436 frost ages thereby resulted in small effective changes in stream water age. The deeper frost depth in the forested regions likely did
437 not reduce the spring infiltration due to the low moisture content in the soil relative to the more saturated wetlands (Laudon et al.,
438 2007). Additionally, the relatively wide uncertainty bounds of stream water age estimates present difficulties in assessing the
439 relatively moderate effects of soil frost on the stream water age (Fig 2). The large dependence of the flows and stream water ages
440 at C7 on the outlet of the large mire at C4 indicates that the water age progressing through the mire will be a strong determinant of
441 long-term change. The flux-weighted median water age estimations for the streams here were estimated to be substantially older
442 than other tracer-aided hydrologic models for the catchment (Ala aho et al., 2017a), though were on the upper end of other stream
443 and hillslope transit times from transit time methods (Peralta-Tapia et al., 2016; Ameli et al., 2017). The reasons for this are largely
444 three-fold. Firstly, the model was calibrated with soil depths comparable to those observed in the catchment. The calibrated model
445 used soil depths ranging from 1.5 – 6 m, where the shallower soil depths yield stream water ages are comparable to previous

446 studies. Secondly, the complete mixing assumption within the model does not allow for rapid preferential movement of young
447 water that has been observed in numerous other recent studies (e.g. Botter et al., 2010; Harman 2015). Incomplete mixing within
448 the model framework would allow for deeper soil profiles to yield younger water fluxes, as estimated from isotopes alone, albeit
449 at the expense of additional parameterisation. Lastly, the previous transit time estimations (Peralta-Tapia et al., 2016; Ameli et al.,
450 2017) do not account for older water ages of the snowpack, or the immobility and aging of frozen soil water, which would increase
451 the estimated water ages.

452 Unlike stream or soil water ages, low uncertainty of transpiration and soil evaporation ages helps bring new understanding to
453 how soil frost affects the source contributions of these ecohydrological fluxes which were the focus of the study. Ages of both
454 transpiration and soil evaporation are consistent with soil profile modelling conducted in the region using the SWIS model
455 (Sprenger et al., 2018b). However, the dynamics of the age variation are notably different due to the differences in the input water,
456 where the snowmelt input to the SWIS model assumes a water age of 0 days and does not account for the “green” water fluxes
457 during the spring months. While the transpiration ages show notable differences when frost, and the corresponding discontinuity
458 of transit times, is included in the simulation, the evaporation water ages are not greatly affected. Transpiration fluxes are
459 influenced by the frost due to the reduced liquid soil water availability and increased soil resistance to transpiration. The higher
460 soil water deficit for transpiration does not fully restrict the transpiration flux, but reduces the fluxes until soil frost thaws and
461 mixes the older frost water in the upper soil layers. The differences are reduced for both fluxes due to a few potential reasons.
462 Firstly, the timing of the soil thaw has a significant influence on age estimation of soil water available for both evaporation and
463 root-uptake. While the general timing and magnitude of the soil frost depth development seems appropriately captured by the
464 model, even without calibration, soil thaw in late winter was simulated faster than observations (Fig 4e). There are notable
465 differences between the ages of soil water, soil frost, and the snowpack, where soil frost is representative of the previous fall soil
466 water, soil water is a younger water mix of the fall soil water and newer precipitation (e.g. from rain-on-snow and early spring
467 snowmelt), and snowpack is the amount weighted age of solid precipitation. Here, shallower soil frost and early melting of soil
468 frost in the spring results in step-wise mixing, firstly of soil frost (oldest water, e.g. 200 – 300 days in Fig 3) and soil water
469 (moderate age, e.g. 100 – 125 days in Fig 3), then of the soil water mixture and snowmelt (youngest water, e.g. 70 – 90 days).
470 Since evaporation, and its corresponding age, only begins following the end of snowmelt, the greater degree of older soil frost with
471 the younger soil water and snowmelt reduces the influence of the soil frost on the evaporation ages. Delaying the simulated timing
472 of soil thaw would result in a larger influence of the soil water ages on both the evaporation and root-uptake.

473 While the influence of soil frost on stream water ages was limited in this catchment, the results have potentially significant
474 implications for modelling other catchments with frozen soils. The effect on water ages will likely be the greatest in catchments
475 where winter precipitation is limited, allowing the soil frost depth to increase from the surface, delaying the soil thaw until after
476 the primary snowmelt. For evaporation and transpiration water ages, notable spatial differences highlight an essential consideration
477 for northern climates in the influence of vegetation-type on the source of water fluxes. In many northern areas, past glaciation
478 results in significant wetlands typically dominated with shrub and herbaceous vegetation. Reductions in soil frost will result in
479 greater water availability throughout the year, aiding in vegetation growth (Woo, 2013). With increased water availability
480 throughout the year, the water use of vegetation will likely increase, and thereby limit the amount of young water percolating
481 through the rooting zone. A reduction in the amount of young water percolating through the rooting-zone will likely increase the
482 age of soil water and catchment outflows. Finally, the timing of the evaporation and root-uptake needs to be strongly considered,
483 at both seasonal and diurnal time scales. Soil frost had a strong influence on the timing of evaporation and transpiration, where the
484 magnitude of both fluxes was greater in simulations without soil frost and timing of the root-uptake and soil evaporation was
485 delayed due to ice-restricted pore spaces. While such changes are anticipated, many studies have focused on plot scale studies and

486 with estimated long-term reductions of soil frost depth, larger scale estimations of these differences are essential to understanding
487 how catchment ecosystems will respond.

488 **6. Conclusion**

489 In northern environments, with a rapidly changing climate, quantitative evaluation of vegetation interactions with catchment
490 soil water is crucial for understanding and projecting catchment responses. The process-based evaluation here of a well-monitored,
491 long-term study catchment in the northern boreal forest region using a tracer-aided, surface-atmosphere energy-balance model has
492 provided significant insights into the importance of soil freeze-thaw processes. Tracers were used, not only as a calibration tool,
493 but as validation metrics, and highlighted the effectiveness multi-criteria calibration of a model at nested scales using discharge,
494 isotopes, and soil moisture to constrain additional, un-measured, features (e.g. soil frost depth). The progressively younger ages of
495 evaporation and transpiration throughout the growing season show the dependence of both “green water” fluxes on spring
496 snowmelt, which remains in soil water towards the end of the growing season. Adaptation of the EcH₂O-iso model provided an
497 opportunity to examine spatial patterns of frost depth throughout the catchment and its ecohydrological influence. Soil frost
498 responded to both lower winter temperatures (increasing frost depths) and greater snowpack depth (decreasing frost depth). While
499 there was little influence on the overall timing of water movement at the catchment scale as stream water ages, the greatest influence
500 was observed within the ecohydrological partitioning, notably with the transpiration ages. Soil frost delays the onset of vegetation
501 growth and soil evaporation, resulting in older soil water from the previous autumn to sustain early-season transpiration rather than
502 younger snowmelt. With the implications of reduced numbers of cold days (Guttorp and Xu, 2011), and the dependence of
503 vegetation growth on the summer temperatures (Schöne et al., 2004) in northern latitudes, this assessment of ecohydrological
504 partitioning is timely in understanding the effect of climatic change.

505

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