

Supplements of the manuscript: modeling inorganic carbon dynamics in the Seine River continuum in France

Audrey Marescaux, Vincent Thieu, Nathalie Gypens, Marie Silvestre, Josette Garnier

S1: List of state variables for the RIVE model

GROUP *	VARIABLE	DESCRIPTION
	Q	mean discharge during the 10 day period in $\text{m}^3 \text{s}^{-1}$
PHY	DIA	diatom biomass in mgC l^{-1}
	GRA	green algae (chlorophyceae) biomass in mgC l^{-1}
	CYA	cyanobacteria biomass in mgC l^{-1}
	MES	mineral suspended solid in mg l^{-1}
NUTS	NO3	nitrate in $\mu\text{mol l}^{-1}$
	NH4	ammonium in $\mu\text{mol l}^{-1}$
	PO4	phosphates in $\mu\text{mol l}^{-1}$
	PIT	total inorganic phosphorus in $\mu\text{molP l}^{-1}$ [**]
	SIO	dissolved silica in $\mu\text{mol l}^{-1}$
	OXY	dissolved oxygen in $\mu\text{mol l}^{-1}$
ZOO	ZOR	rotiferan-like zooplankton in mgC l^{-1}
	ZOC	cladoceran-like zooplankton in mgC l^{-1}
BACT	BAG	heterotrophic bacteria $> 1\mu$ in mgC l^{-1}
	BAP	heterotrophic bacteria $< 1\mu$ in mgC l^{-1}
	NIT	ammonium oxidizing nitrifying bacteria in mgC l^{-1}
	NAT	nitrite oxidizing nitrifying bacteria in mgC l^{-1}
OC	DOC1	rapidly biodegradable dissolved organic matter in mgC l^{-1}
	DOC2	slowly biodegradable dissolved organic matter in mgC l^{-1}
	DOC3	refractory dissolved organic matter in mgC l^{-1}
	POC1	rapidly biodegradable particulate organic matter in mgC l^{-1}
	POC 2	slowly biodegradable particulate organic matter in mgC l^{-1}
	POC 3	refractory particulate organic matter in mgC l^{-1}
	DSS	low molecular weight directly assimilable organic substrates in mgC l^{-1}
Benthic	SIB	biogenic non-living particulate (amorphous) silica in $\mu\text{mol l}^{-1}$
	BOC1	benthic (deposited) rapidly biodegradable organic matter in gC m^{-2}
	BOC2	benthic (deposited) slowly biodegradable organic matter in gC m^{-2}
	BOC3	benthic (deposited) refractory organic matter in gC m^{-2}
	BPI	benthic (deposited) inorganic phosphorus in mmolP m^{-2}
	BBS	benthic (deposited) biogenic silica in mmol m^{-2}
	BFE	benthic (deposited) faecal bacteria in $1000 \text{ m}^{-2} \text{l}^{-1}$
	SED	deposited inorganic material in g m^{-2}
	FEL	free living faecal bacteria in nb l^{-1}
	FEA	attached faecal bacteria in nb l^{-1}
	N2O	nitrous oxide in $\mu\text{mol l}^{-1}$
	NO2	nitrite in $\mu\text{mol l}^{-1}$
	CH4	methane in $\mu\text{mol l}^{-1}$
IC	CO2 (***)	Carbon dioxide in mgC l^{-1}
	TA (***)	Total alkalinity in $\mu\text{mol l}^{-1}$
	DIC (***)	Dissolved inorganic carbon in mgC l^{-1}

(*) “Group” refer to generic group name provided in **Erreur ! Source du renvoi introuvable.** in this paper

(**) An instantaneous equilibrium is considered for adsorption of ortho-phosphate on MES, so that PIT is the only primary state variable to be considered.

(***) New state variable added in the RIVE model

All biomasses (DIA, GRA, CYA, BAG, BAP, NIT, NAT, ZOR, ZOC) as well as organic matter pools (DOC 1-2-3, POC 1-2-3, BOC 1-2-3) are considered to have constant massic C:N ratios, namely $C/N=7$

S2: Modeling objects of the Seine basin

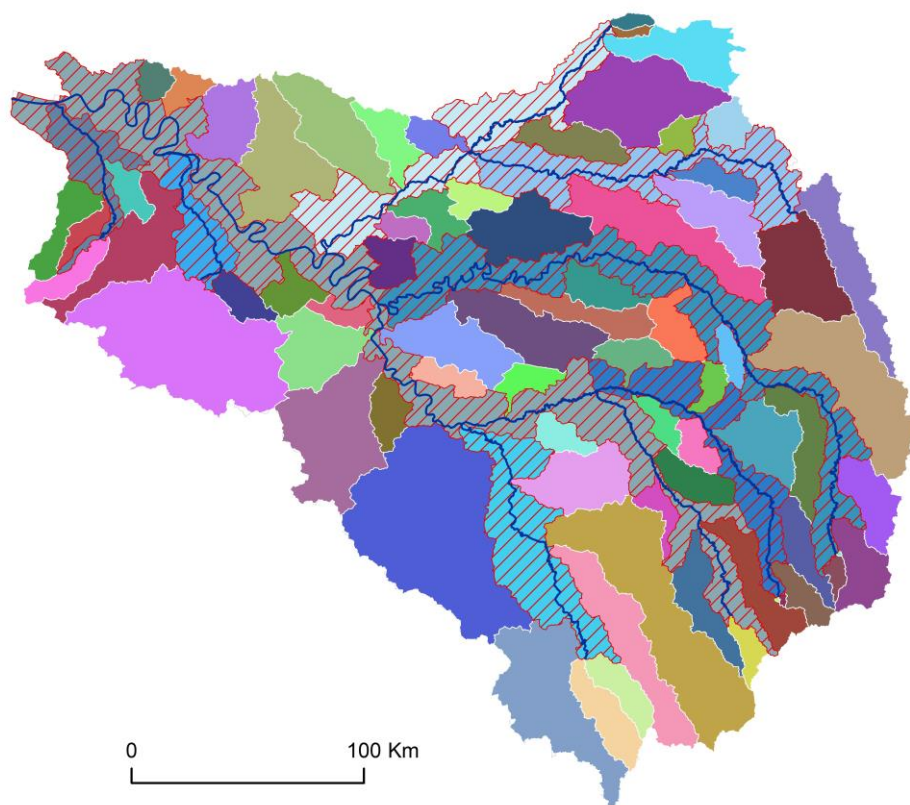
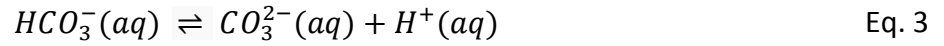
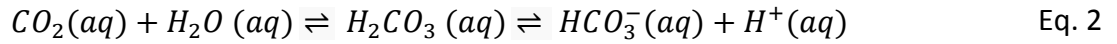


Fig. S1. PyNuts-Riverstrahler modeling objects. Eight “Axis” objects (main river branches, mapped with red hatching) and 72 “basin” objects (upstream tributaries connected to “axis” objects).

S3: Description of the inorganic carbon module implemented in pyNuts-Riverstrahler

3.1. The carbonate systems

The major dissolved forms of the carbonate system are CO_2 (aq) aqueous carbon dioxide, H_2CO_3 (aq) carbonic acid, HCO_3^- (aq) bicarbonate ion, CO_3^{2-} (aq) carbonate ion.



The acid dissolution constants from the second and the third equations are, respectively:

$$K_1 = \frac{[H^+][HCO_3^-]}{[CO_2]} \quad \text{Eq. 4}$$

$$K_2 = \frac{[H^+][CO_3^{2-}]}{[HCO_3^-]} \quad \text{Eq. 5}$$

$$\text{with } pK = -\log K \quad \text{Eq. 6}$$

3.2. Carbonate system as a function of DIC

Carbonate dissociation can be determined by total inorganic carbon.

$$DIC = [CO_2] + [HCO_3^-] + [CO_3^{2-}] \quad \text{Eq. 7}$$

With the insertion of [4] and [5] in [7], DIC becomes:

$$DIC = \frac{[H^+][HCO_3^-]}{K_1} + [HCO_3^-] + \frac{[HCO_3^-]K_2}{[H^+]} \quad \text{Eq. 8}$$

$$\leftrightarrow [HCO_3^-] = \frac{DIC}{\frac{[H^+]}{K_1} + \frac{K_2}{[H^+]} + 1} \quad \text{Eq. 9}$$

and with $[HCO_3^-]$, we can calculate $[CO_3^{2-}]$ and $[CO_2]$ with [8] and [9].

$$[CO_2] = \frac{[H^+]}{K_1} \frac{DIC}{\frac{[H^+]}{K_1} + \frac{K_2}{[H^+]} + 1} \quad \text{Eq. 10}$$

$$[CO_3^{2-}] = \frac{K_2 C}{[H^+]} \frac{DIC}{\frac{[H^+]}{K_1} + \frac{K_2}{[H^+]} + 1} \quad \text{Eq. 11}$$

3.3. The dissociation constants of carbonic acid

According to Millero et al. (2006), Harned and Scholes (1941), Harned et al. (1943) the dissociation constants of carbonic acid (molar concentration unit) are calculated as:

$$pK_1 = -126.34048 + \frac{6320.813}{T_K} + 19.568224 * \ln(T_K) \quad \text{Eq. 12}$$

$$pK_2 = -90.18333 + \frac{5143.692}{T_K} + 14.613358 * \ln(T_K) \quad \text{Eq. 13}$$

where T_k is the water temperature in Kelvin (°K).

and:

$$K_1 = 10^{-pK_1} \quad \text{Eq. 14}$$

$$K_2 = 10^{-pK_2} \quad \text{Eq. 15}$$

3.4. pH calculation

In the pyNuts-Riverstrahler model, pH had not yet been calculated. In the inorganic carbon module, we added the calculation of pH derived from Culberson (1980) to calculate the speciation of carbonate (Eqs. 9–11).

$$DIC_{conv} = \frac{DIC}{12 \rho} 10^6 \quad \text{Eq. 16}$$

DIC_{conv} (conversion from mgC L^{-1} to $\mu\text{mol kg}^{-1}$), where ρ (kg m^{-3}) is the water density from Millero and Poisson, 1981)

$$Xdiss = \left(1 - \frac{Bor0}{CA}\right) \cdot K_b + \left(1 - \frac{DIC_{conv}}{CA}\right) \cdot K_1 \quad \text{Eq. 17}$$

$$Ydiss = \left(1 - \frac{Bor0 + DIC_{conv}}{CA}\right) \cdot K_1 \cdot K_b + \left(1 - 2 \cdot \frac{DIC_{conv}}{CA}\right) \cdot K_1 \cdot K_2 \quad \text{Eq. 18}$$

$$Zdiss = \left(1 - \frac{Bor0 + 2 \cdot DIC_{conv}}{CA}\right) \cdot K_1 \cdot K_2 \cdot K_b \quad \text{Eq. 19}$$

$$aCulb = \frac{(Xdiss^2 - 3 \cdot Ydiss)}{9} \quad \text{Eq. 20}$$

$$bCulb = \frac{-(2Xdiss^3 - 9 \cdot Xdiss \cdot Ydiss + 27 \cdot Zdiss)}{54} \quad \text{Eq. 21}$$

$$phyCulb = \cos\left(\frac{bCulb}{(aCulb^3)^{0.5}}\right) \quad \text{Eq. 22}$$

$$[H_3O^+] = 2 \cdot aCulb^{0.5} \cdot \cos\left(\frac{phyCulb}{3}\right) - \frac{Xdiss}{3} \quad \text{Eq. 23}$$

$$pH = -\log_{10}[H_3O^+] \quad \text{Eq. 24}$$

where Bor0 is the total dissolved boron concentration that can generally be ignored in freshwaters (Emiroglu et al., 2010). “Xdiss”, “Ydiss”, “Zdiss” are the coefficient of the general cubic equation. “aCulb” and “bCulb” are the zero of the general cubic equation. H_3O^+ is the positive roots of the general cubic equation that can be calculated from phyCulb (where $\arccos(\text{phyCulb})$ is the inverse trigonometric function) (see Culberson et al. 1980).

3.5. CO₂ flux calculation

The carbonate speciation function makes it possible to calculate the concentration in aqueous carbon dioxide (CO₂) (eq. 10).

The flux ($\text{gC m}^{-2} \text{h}^{-1}$) of CO_2 at the interface of the river and the atmosphere is calculated as:

$$F_{\text{CO}_2} = \frac{k}{24} (\text{CO}_2 - \text{CO}_{2\text{atm}}) \cdot \text{rho} \cdot 10^{-3} \quad \text{Eq. 25}$$

Monthly atmospheric CO_2 ($\text{CO}_{2\text{atm}}$, mgC L^{-1}) was measured at Mauna Loa Observatory (Hawaii, USA) and provided by the NOAA/ESRL:

(<http://www.esrl.noaa.gov/gmd/ccgg/trends/>, last accessed 2018/11/05), Scripps Institution of Oceanography (scrippsco2.ucsd.edu/, last accessed 2018/11/05). Concentrations in ppm from Mauna Loa Observatory were converted to mgC L^{-1} using solubility according to Weiss, (1974). The CO_2 concentration in water (CO_2) (mgC L^{-1}) was calculated according to section 3.2. Carbonate system as a function of DIC". rho (kg m^{-3}) is the water density calculated by the International One Atmosphere Equation (Millero and Poisson, 1981).

The flux of CO_2 depends on the gas transfer velocity (k) that can be determined from the temperature-normalized gas transfer velocity (k_{600}). k_{600} is the gas transfer velocity at a water temperature of 20°C . Parametrization related to the gas exchange used k_{600} to compare systems excluding temperature's physical effect. According to Wilke and Chang (1955) and Wanninkhof (1992), the gas transfer velocity $k(\text{m d}^{-1})$ can be calculated as:

$$k = k_{600} \cdot \sqrt{\frac{600}{Sc_{\text{CO}_2}(T)}} \quad \text{Eq. 26}$$

k_{600} is the gas transfer velocity for a Schmidt number of 600 (m d^{-1}), and $Sc_{\text{CO}_2}(T)$ is the Schmidt number (dimensionless) calculated with the water temperature (T) in Celsius ($^\circ\text{C}$) calculated as:

$$Sc_{\text{CO}_2}(T) = 1911.1 - 118.11T + 3.4527T^2 - 0.04132T^3 \quad \text{Eq. 27}$$

For rivers less than 100 m wide (SO 1–5), the k_{600} equation used was retrieved from Alin et al. (2011):

$$k_{600} = \frac{13.82 + 0.35v * 100}{100} \quad \text{Eq. 28}$$

where k_{600} (m d^{-1}) depends on v , the water velocity (m s^{-1}).

In general, k_{600} equations for rivers of widths greater than 100 m require wind velocity (e.g., Chu and Jirka, 2003; Alin et al., 2011; Raymond et al., 2012a). At this stage the pyNuts Riverstrahler model does not consider the wind as an input, which would have required a much higher spatiotemporal resolution to reflect its heterogeneity in the Seine basin with the diurnal cycle affected by phenomena such as breeze (Quintana-Seguí et al., 2008). Therefore, for river widths greater than 100 m, the k_{600} equation from O'Connor and Dobbins, (1958) and Ho et al. (2016), neglecting the term related to the wind, was selected (see **Erreur ! Source du renvoi introuvable.**):

$$k_{600} = \frac{a \sqrt{\frac{v}{depth}}}{100} \quad \text{Eq. 29}$$

where k_{600} (m d^{-1}) depends on v , the water velocity (m s^{-1}), and the river depth (m) and the coefficient “a”. The coefficient was tested from 0 to 2 for SO6 and SO7. We found that SO6 showed a behavior near the original coefficient of 1.539 from O'Connor and Dobbins (1958) with 1.55, and SO7 a behavior near the 0.77 proposed by and Ho et al. (2016) with 0.55.

3.6. Parameters, constants and equations used in the inorganic carbon module.

Table S1 Parameters used in the inorganic carbon module for freshwaters

Parameters	Acronym	Units	Sources
Dissociation constants of carbonic acid	K_1, K_2	$\mu\text{mol l}^{-1}$	Harned and Scholes (1941), Harned et al. (1943), and, Millero et al. (2006)
Solubility	k_0	$\text{mol kg}^{-1} \text{atm}^{-1}$	Weiss (1974)
gas transfer velocity	<i>k-value</i>	m h^{-1}	Wanninkhof (1992), Wilke and Chang (1955) Alin et al. (2011, eq. for river width inferior to 100m) and O'Connor and Dobbins (1958) updated in Ho et al. (2016), for river width superior to 100m.
water density	ρ	kg m^{-3}	Millero and Poisson (1981)
pH	pH	-	Culberson (1980)
Apparent ionization constant of boric acid	K_b	hPa	Schubert (2011)

S4 Representation of superficial MESO water bodies

Groundwaters are under strong anthropogenic pressures due to water withdrawn for drinking water production and pollution by agriculture. The European directive (Water Framework Directive 2000/60/CE) aims to reach the good environmental status of surface and groundwaters and has introduced the notion of groundwater bodies (French acronym: Masse d'Eau SOuerraine - MESO) (Fig. S2). These MESO units are defined by hydrogeological criteria (e.g., extent and characteristics of the geological layers, feeding area, hydraulic connection between geological layers, interaction with surface waters and associated terrestrial ecosystems, etc.) and nonhydrogeological criteria (capture or the possibility of capture; impact of pressures, potential pollution, administrative boundaries, etc.). The status and upward trends in the concentrations of any pollutant in groundwaters are monitored at the MESO scale. The Seine basin includes 48 unconfined (fully or partially) MESO units (see map), which were regrouped according to the lithology and geological ages (colors) to simplify the representation.

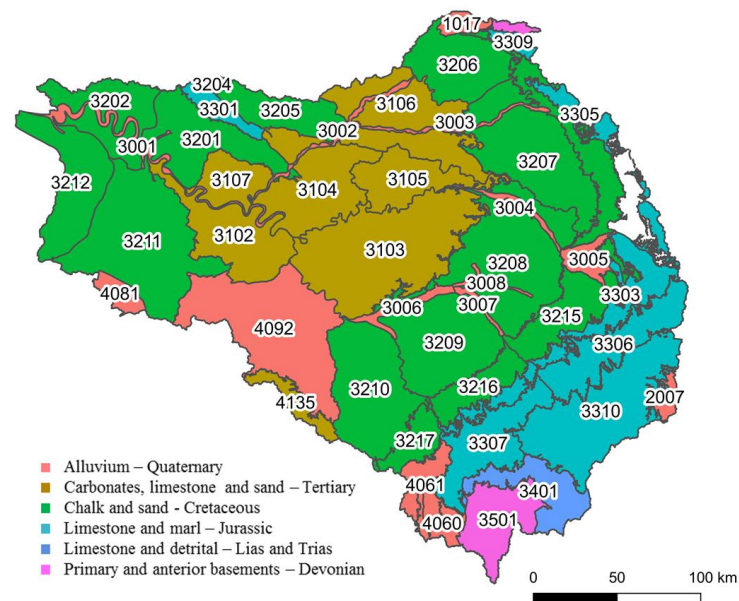


Fig. S2. Superficial MESO (Masse d'Eau SOuerraine) water bodies of the Seine basin grouped according to the lithology and geological ages

Table S2 Mean concentrations and standard deviations of total alkalinity (TA, $\mu\text{mole L}^{-1}$) and dissolved inorganic carbon (DIC, mgC L^{-1}) grouped by MESO units in the Seine Basin for the 2010-2013 period.

MESO code	Dominant lithology and geological age	TA ($\mu\text{mole L}^{-1}$)		DIC (mgC L^{-1})	
		<i>mean</i>	<i>sd</i>	<i>mean</i>	<i>sd</i>
1017	Alluvium - Quaternary	5113	388	79	3
2007		4454	671	75	7
3001		4030	1018	71	12
3002		5058	677	85	6
3003		5288	1025	90	11
3004		4532	992	72	10
3005		4624	487	79	8
3006		3960	493	69	7
3007		4254	382	67	7
3008		4274	832	70	10
4060		4701	291	74	2
4061		4701	291	74	2
4081		1568	1462	35	16
4092		4245	785	72	10
3102	carbonate, limestone and sand – Tertiary	4719	878	81	15
3103		5453	988	76	11
3104		5580	673	92	10
3105		5387	768	91	7
3106		5321	360	89	10
3107		5264	488	86	5
4135		3774	1011	64	11
3201	chalk and sand - Cretaceous	5264	488	83	4
3202		4574	546	76	6
3204		4831	389	81	1
3205		5192	470	83	5
3206		5134	576	83	7
3207		4268	941	70	11
3208		3911	866	64	10
3209		4169	680	69	8
3210		4669	590	75	6
3211		4173	801	67	10
3212		4713	401	77	4
3214		5172	941	87	12
3215		4394	1202	73	13
3216		4732	1011	77	11
3217		5002	481	77	4
3218		4590	995	75	12
3301	Limestone and marl - Jurassic	1660	346	41	5
3303		5120	467	82	6
3304		4939	265	78	4
3305		5236	971	88	12
3306		4629	754	77	8
3307		4907	585	80	6
3309		4938	365	81	4
3310		4454	671	75	7
3401	Limestone and detrital – Lias and Trias	5053	699	82	8
3501	Primary and anterior basements - Devonian	765	228	25	4
3508		663	683	27	9

S5: Sampling strategy and protocol for measuring emissions from wastewater treatment plants in the Seine basin

Table S3. Location of the sampled WWTPs. Their nominal capacity and type of treatment are retrieved from <http://assainissement.developpement-durable.gouv.fr/>, last accessed 2019/03/07. Measured values for TA, DIC, pH and pCO₂.

WWTPs	Location	Nominal capacity (Inhab. Eq.)	Water treatment	TA (μmole L ⁻¹)	DIC (mgC L ⁻¹)	pH	pCO ₂ (ppmv)
Butry	49°04'60N, 2°12'00E	6700	Low load activated sludge	5885	76.27	7.3	31723
Auvers	49°04'00N, 2°10'00E	34,300	Secondary treatment, denitrification, dephosphatation	3700	53.39	7.26	28723
Dammarie	48°30'57"N, 2°36'43"E	80,000	Secondary treatment, denitrification, dephosphatation	4800	68.55	7.45	17268
Rosny	49°0'7"N, 1°39'11"E	135,417	Secondary treatment, denitrification, dephosphatation	4500	55.50	7.4	18651
Troyes	48°20'5"N, 4°2'37"E	260,000	Secondary treatment, denitrification, dephosphatation	#N/A	86.31	7.4	29930
Seine Centre	48°55'58"N, 2°14'43"E	800,000	Secondary treatment, denitrification, dephosphatation	3590	47.44	8.17	#N/A
St-Thibault-des-Vignes	48°52'20"N, 2°40'27"E	350,000	Secondary treatment, denitrification, dephosphatation	5660	73.55	7.40	15732
SAV	48°58'25" N, 2°09'56"E	7,500,000	Secondary treatment, denitrification, dephosphatation	2915	#N/A	7.91	#N/A
Inputs of the model (weighted average per capacity):				3993	69.75		

Water samples were filtrated on combusted filters (4 h at 500°C: GF/F 0.7 μm, 25 mm), and filtrates enabled measurement of DIC and TA concentrations. Dissolved inorganic carbon was analyzed with a TOC analyzer (Aurora 1030). Nongaseous DIC analyses required acidification of the filtrated sample by adding sodium persulfate reagents (100 g L⁻¹) to dissociate the carbonate in the CO₂ that was detected with nondispersive infrared gas analysis (IRGA). TA (μmol kg⁻¹) was analyzed using an automatic titrator (TitroLine® 5000) on three 20-mL replicates of filtered water (GF/F: 0.7 μm), with hydrochloric acid (0.1 M).