



Partitioning the forest water balance within a boreal catchment using sapflux, eddy covariance and process-based model

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Abstract

20 In the hydrological cycle, water is lost from terrestrial ecosystems either laterally through stream runoff or vertically as evapotranspiration (ET) back to the atmosphere. Although it is well known that ET losses represents an important water loss pathway at local to global scales, the magnitude and relative importance of ET and its individual flux components varies considerable among different ecosystems. In this study, we combined empirical sapflux and

25 eddy covariance measurements with estimates from a process-based model to partition the water balance in a boreal forested catchment. This study was conducted within the Krycklan Catchment, which has state-of-the-art infrastructure for hydrological measurements, thereby providing us the unique opportunity to compare the absolute and relative magnitude of ET and its flux components to other water loss pathways (i.e., stream runoff). During the

30 growing season, ET was the major water loss pathway, representing *ca.* 85 % of the incoming precipitation and being roughly 7 times greater than stream runoff. Both the empirical results and model estimates suggested that tree transpiration (T) and canopy interception (IL) represented 45 % and 35 % of total ET loss; respectively, and thus together were responsible for 70 % of the water loss during the growing season. Understory evapotranspiration (ET_u)

35 was less important than T and IL during most of the study period, except for late autumn when ET_u was the largest ET flux component. Overall, our study highlights the importance of trees in regulating the water cycle of boreal catchments, implying that forest management



impacts on stand structure as well as climate change effects on tree growth are likely to have large cascading effects on the way water moves through these forested landscapes.

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1 Introduction

In the hydrological cycle, water enters terrestrial ecosystems mainly through precipitation.

This water is lost either through surface runoff or evapotranspiration (ET) back to the

atmosphere. At a global scale, ET losses can account for *ca.* 60 % of the annual terrestrial

45 precipitation (Oki and Kanae, 2006). Yet the magnitude, as well as the relative importance, of

ET varies considerably among different ecosystems (Peel et al., 2010). Understanding this

variation in ET loss is crucial, as the difference between incoming precipitation and ET

represents the available water to a hydrological system, which in turn has cascading effects

on streamflow (Karlsen et al., 2016;Koster and Milly, 1997), groundwater recharge (Githui et

50 al., 2012) and the ecosystem carbon cycle (Wang et al., 2002;Öquist et al., 2014).

For a long time, most hydrological studies have treated ET as a single water flux.

However, ET represents two fundamentally different pathways of water loss from terrestrial

ecosystems: (1) transpiration through stomata of plants and (2) evaporation from wet

surfaces. These two different pathways are controlled in contrasting ways and to varying

55 degrees by environmental factors and thus are likely to respond differently to climate change

and land management practices. Specifically, transpiration occurs mainly during the



vegetation season and is thus governed by plant physiological processes, whereas evaporation occurs throughout the year and is strongly controlled by vapor pressure deficit, surface wetness, and aerodynamic conductance (Stoy et al., 2006). Thus, quantifying the magnitude and drivers of transpiration and evaporation are crucial to better understanding the spatiotemporal variation of water fluxes in terrestrial ecosystems.

During the past decade, there has been increasing number of studies partitioning ET into transpiration and evaporation (see Schlesinger and Jasechko (2014)). Based on the isotopic composition of water ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) collected from large lakes, Jasechko et al. (2013) estimated that at a global scale transpiration represents 80 to 90 % of terrestrial ET. The high estimate of T/ET reported by Jasechko et al. (2013) has been strongly contested (Coenders-Gerrits et al., 2014), with a more conservative estimate of transpiration representing *ca.* 60 % of ET being more generally accepted (Schlesinger and Jasechko, 2014; Wei et al., 2017). Despite these recent advances in partitioning ET, most ecohydrological studies typically partition ET at the stand or plot scale without considering the broader hydrological cycle (e.g. Cienicala et al., 1997; Grelle et al., 1997; Schlesinger and Jasechko, 2014; Wang et al., 2017; Ohta et al., 2001; Iida et al., 2009; Hamada et al., 2004; Maximov et al., 2008; Warren et al., 2018). We are aware of only a few investigations that have provided a detailed water balance at the catchment scale (Sarkkola et al., 2013b; Telmer and Veizer, 2000), and thus we have little empirical data about how the



amount of water that is lost via transpiration compares to other water fluxes (i.e., streamflow) in the terrestrial hydrological cycle.

Transpiration can be further partitioned into canopy tree transpiration and understory vegetation transpiration. Few studies have measured understory transpiration, yet the ones that have suggest that understory transpiration represents a small fraction of total transpiration (Kulmala et al., 2011; Palmroth et al., 2014) and is strongly dependent on canopy tree structure (Constantin et al., 1999; Baldocchi et al., 1997; Domec et al., 2012). Similarly, total evaporation can be partitioned into canopy interception losses (IL) and evaporation from the forest floor. At a global scale, roughly 20 % of incoming precipitation is lost via IL (Wang et al., 2007) and in many forested ecosystems IL represents a substantial portion of total evaporation (Barbier et al., 2009; Gu et al., 2018). By separating transpiration and evaporation into their different flux components, it is possible to directly assess the relative importance of trees (i.e., canopy transpiration + IL) to other water loss pathways and in turn determine the important role trees play in the terrestrial hydrological cycle.

There are numerous different approaches and methodology to partition ET into its different flux components (Kool et al., 2014). One approach is to use a combination of eddy-covariance methods to estimate ET and sap flow sensors to separately estimate transpiration in the dominant canopy trees (Mitchell et al., 2009; Cavanaugh et al., 2011), although this approach often ignores sub-canopy vegetation. Combining this approach with the placement



95 of rain gauges under the forest canopy and in open sky it is also possible to estimate IL and
its relative importance to total ET losses. With recent advances in techniques and
instrumentation it is now also possible to use the stable hydrogen and oxygen isotopes ($\delta^2\text{H}$
and $\delta^{18}\text{O}$) of liquid water to partition ET (Evaristo et al., 2015; Good et al., 2014; Sutanto et
al., 2014). In addition to these empirical approaches, there are a number of different process-
100 based models that can be used to partitioning ET into its different flux components (Sutanto
et al., 2014; Stoy et al., 2019). Each of these different approaches have their advantages and
disadvantages and moreover it has been shown that the relative contribution of ET flux
components differs depending on the approach used (Schlesinger and Jasechko, 2014). It has
therefore been highlighted that the use of multiple methods is desirable to more accurately
105 partition ET into its individual flux components (Stoy et al., 2019).

In this study, we use a combination of empirical data derived from eddy-covariance
and sapflux measurements as well as rain gauges collecting open sky and throughfall
precipitation to partition ET into its individual flux components during the growing season in
a boreal catchment. Additionally, we used a multi-layer, multi-species soil-vegetation-
110 atmosphere transfer model (APES model based on Launiainen et al., 2015) as another
independent approach to partition ET. In doing so, the main objective of this study was to *i*)
constrain the absolute and relative magnitude of ET flux components by using both empirical
data and model simulations and *ii*) to explore how they vary during the course of the growing



season. This study was conducted within the Krycklan Catchment, which has state-of-the-art
115 infrastructure for hydrological measurements, thereby providing us the unique opportunity to
compare different ET fluxes to other water loss pathways (i.e., streamflow) and to directly
assess the important role trees play in the boreal hydrological cycle.

2. Material and Methods

120 2.1 Study site

The study was conducted in the 14 ha subcatchment C2 (64.26° N, 19.77° E) within the 64
km² Krycklan Catchment Study area (Laudon et al., 2013) in northern Sweden (Fig. 1). The
Krycklan Catchment Study area is unique as it is one of the oldest long-term catchment
monitoring sites in northern latitudes with continuous hydrological and meteorological
125 measurements dating back to the early 1980s. The 30-year mean annual temperature in
Krycklan (1986-2015) was 2.1° C; with highest mean monthly temperature in July and lowest
temperature in January (14.6°C and -8.6°C; respectively). The mean annual precipitation was
619 mm, with the majority (ca. 60%) falling in the form of rain. Soils within the C2 are
dominated by glacial till (84%), predominately of stony, sandy texture on gneiss and granite.
130 There is considerable variation in the thickness of the humus layer, yet the average is 8 cm
(Odin, 1992). The average slope is 6% and the outlet of the C2 subcatchment is located at
243 m a.s.l.

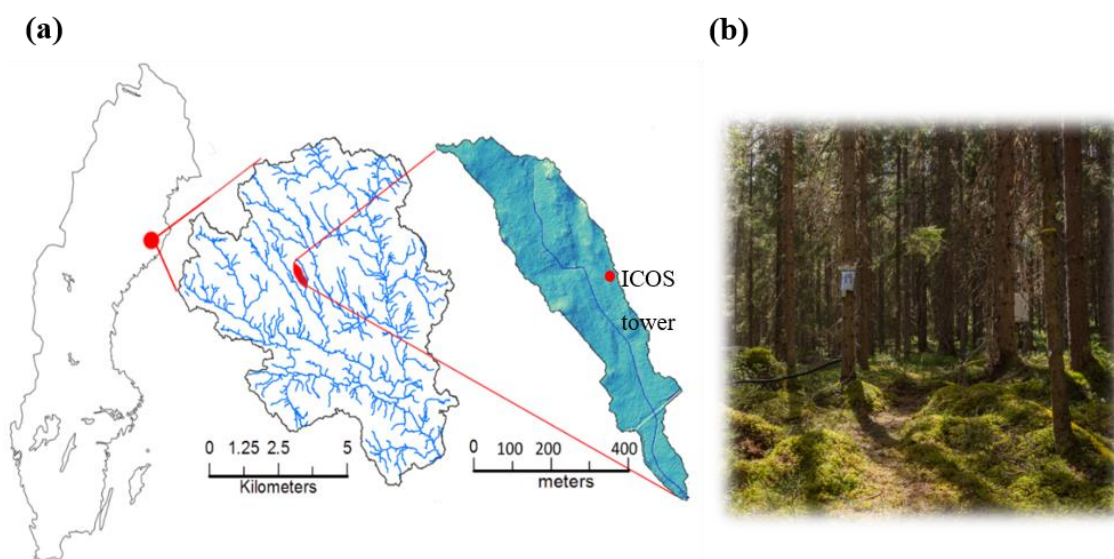


Figure 1. Location of the studied C2 subcatchment within the Krycklan Catchment in

135 northern Sweden (a) and a picture of forest stand and rich understory vegetation that is
 characteristic of the C2 subcatchment (b).

The catchment is characterized by an old (>100 yr.) mixed forest stand of *Picea abies*
 (61 %), *Pinus sylvestris* (34 %), and *Betula* (5 %) (Laudon et al. 2013). The understory
 140 consists of a rich layer of bilberry (*Vaccinium myrtillus*), lingonberry (*Vaccinium vitis-idaea*),
 and mosses (*Pleurozium schreberi* and *Hylocomium splendens*). Similar forest stands extend
 to the east and west of the C2 subcatchment boundaries by several hundred meters. Within
 the C2 subcatchment, there is also the Integrated Carbon Observation System (ICOS)

Svartberget ecosystem-atmosphere station which provides data on greenhouse gas, water and

145 energy fluxes as well as meteorological, vegetation and soil environmental variables



(www.icos-sweden.se/station_svaltberget.html). Our study period was from July to October 2016 which corresponded to the snow free growing season, spanning from after the spring flood until leaf senescence for deciduous species. The 2016 year was a typical year in terms of precipitation and stream runoff (Fig. S1).

150

2.2 Empirical measurements of the water balance components

To quantify water balance components at the catchment scale, we combined multi-scale measurements of vertical and lateral water fluxes made within the C2 subcatchment. The underlying assumption in our analysis is that forest stand characteristics are uniform across
155 and beyond the C2 subcatchment, which allows us to directly compare water fluxes per unit area regardless of the spatial scale the measurements represents: stand vs. catchment.

Precipitation (P) and other environmental data at 30 min temporal resolution were obtained from the ICOS portal, Svartberget station (<http://www.icos-sweden.se/data.html>). For this study we used data of daily mean air temperature, relative humidity, wind speed,
160 total precipitation, atmospheric pressure, incoming short and long-wave radiation, photosynthetic active radiation (PAR), as well as soil temperature and moisture measured at 0.05 m depth. Daily stream runoff (Q) was calculated as ratio of daily discharge, obtained from the Svartberget data portal (<https://franklin.vfp.slu.se/>), and catchment area. Changes in soil storage (ΔS) were not empirically measured in this study.



165 Evapotranspiration (ET) data were derived from the ICOS-Svartberget eddy
 covariance (EC) system installed at 32.5 m above the ground. The setup consists of a 3D
 ultrasonic anemometer (uSonic3 Class-A, METEK Meteorologische Messtechnik GmbH,
 Germany) for measuring wind components (u , v , w) and a closed-path infrared gas analyzer
 (LI-7200, LI-COR Biosciences, USA) for measuring CO_2 and H_2O concentrations. The EC
 170 raw data were recorded at 10 Hz frequency and 30-min fluxes were processed in the
 EddyPro[®] software (version 6.2.0, LI-COR Biosciences, USA) according to common settings
 described in details in Chi et al. (2019). Briefly, the ET data were computed as a covariance
 of vertical wind speed (w') and H_2O concentration (q') for each 30-min period and corrected
 for water vapor storage in the air volume between the measurement height and the forest
 175 ground:

$$ET = \overline{w'q'} + \int_0^h \overline{\rho_d} \frac{\partial \bar{q}}{\partial t} \delta z \quad (1)$$

where $h = 32.5$ m, $t = 30$ min, $\overline{\rho_d}$ is dry air density, q is profile measurements of water vapor
 concentrations at different height (z) from 4.2 to 30 m above the ground. ET data were then
 filtered for the non-ideal conditions for EC measurements, i.e. non-stationarity, wind
 180 distortion, power failure, as well as site maintenance activities (Chi et al., 2019). The
 potential advection effects on ET data were detected using the turbulence intensity
 parameters (Wharton et al., 2009) and data were discarded when advection was no longer
 negligible. Gaps in the half-hourly ET data were filled using the REdyProcWeb online tool



(Wutzler et al., 2018). Based on the Kljun footprint model (Kljun et al., 2015), the EC

185 footprint (90 %) covers a measurement area of $\sim 0.5 \text{ km}^2$ with a mean upwind fetch of $\sim 400 \text{ m}$ surrounding the tower. The uncertainty in the EC-based ET was estimated by the Monte Carlo simulation (Richardson and Hollinger, 2007).

Evaporation of intercepted water from tree canopy was estimated based on canopy interception losses (IL) and was determined by subtracting throughfall (TF) from open sky
190 precipitation (P): $IL = GP - TF$. Measurements of TF were made 1 km from the study subcatchment by installing 25 rain gauges to measure throughfall in a similar mature mixed coniferous forest stand. The design of rain gauges followed WMO (Bidartondo et al., 2001) requirements, which included a stable rim with sharp edge, orifice area of 200 cm^2 , hydrophobic plastic material and a narrow entrance to the receiving container to prevent
195 evaporation. To test custom made gauges, three of them were installed next to a standardized precipitation collector Geonor T200BM (Geonor Inc., New Jersey, USA) at the Svartberget field station for the entire period and the difference in captured rain was always less than 3%. Measurements of TF were made between the beginning of July and the end of October 2016, at an event-based frequency. The IL (mm) for a single rain event was estimated as weighted
200 average from the 25 throughfall gauges. Weighting was based on spatial forest canopy density data acquired by airborne laser scanning (ALS), using a canopy metric calculated from all canopy height measurements within a two-meter horizontal distance of the collector,



and frequency distribution of that canopy metric within the C2 subcatchment. The canopy metric used for this procedure was the median of the absolute deviation of ALS height

205 measurements from the overall median height (ElevMADmedium). This particular metric was chosen from the set of metrics calculated by the FUSION software (McGaughey, 2012) because it had the highest correlation with measured seasonal interception loss. To eliminate influence of potential difference between open sky precipitation within the C2 subcatchment and sampling plot on the final results, at the end of sampling campaign, we estimated the

210 fraction of seasonal interception loss ($IL_f = 1 - TF/GP$) and multiplied that value by cumulative precipitation at the study catchment. To quantify the uncertainty of the event-based IL estimated from measurement, we grouped the gauges into 5 groups based on the ElevMADmedian canopy metric and calculate standard deviation for each group and event.

Canopy transpiration (T) was estimated using sap flux measurements. Within the EC

215 footprint area, we selected three locations (hereafter referred to as nodes) to measure canopy transpiration (Fig. S2). Within each node (25 m radius), we selected 20 trees (10 *Pinus sylvestris* and 10 *Picea abies*) that represented the diameter distribution of the entire C2 subcatchment forest stand. Although *Betula spp.* is also present within the C2 subcatchment, they contribute less than 5% of the basal area and we therefore focused on the two dominant

220 conifer species (Laudon et al., 2013).



Sap flux density (J_s , $\text{g m}^{-2}_{\text{sapwood s}^{-1}}$) was measured at breast height (1.3 m above ground) using custom-made heat dissipation-type sap flow sensors (Granier, 1987). Each pair of sensors consisted of a heated and non-heated probe made from 19-gauge hypodermic needles with metallic, sensing parts cut into 20 mm length. These sensors were installed on the selected trees with 10-15 cm spacing between probes and all sensors were covered with reflective insulation to reduce external temperature influences. To account for radial (Phillips et al., 1996; Ford et al., 2004; Oishi et al., 2008) and azimuthal (Oren et al., 1999; Lu et al., 2000; James et al., 2002; Tateishi et al., 2008) variation in J_s , we installed sensors in the north, east, south and west sides of the stems in 6 of the selected trees from all nodes ($n = 3$ per species). We also installed sensors at four 20 mm interval depths from the inner bark (i.e., 0-20 mm, 20-40 mm, 40-60 mm and 60-80 mm) in a subset of tree species. Data of temperature difference between the two probes were collected as 30-minute averages of voltage difference (ΔV , mV) using a data logger (CR1000, Campbell Scientific, Logan, UT, USA) which was set to record data every 30 s. The collected data were converted to J_s using the empirical equation (Granier, 1987)

$$J_s = 118.99 \times 10^{-6} \times \left(\frac{\Delta V_m - \Delta V}{\Delta V} \right)^{1.231} \quad (2)$$

where ΔV_m is the maximum voltage difference under zero flow conditions which occur at night and when vapor pressure deficit is low. We employed the Baseline program version



4.0 (Oishi et al., 2016) to convert the ΔV data to J_s . This accounts for nocturnal fluxes
 240 resulting from nighttime transpiration and water recharge in stems by selecting the highest
 daily ΔV to represent ΔV_m . The selection criteria for determining ΔV_m were conditions when
 (1) the average, minimum 2-hour vapor pressure deficit is less than 0.02 kPa, thus ensuring
 negligible transpiration and (2) the standard deviation of the four highest values is less than
 0.5 % of the mean of these values, therefore ensuring that water storage change above the
 245 sensor height is negligible compared to J_s .

To determine daily canopy transpiration (T , mm d^{-1}), we first integrated J_s over 24
 hours as daily J_s (J_{SD} , $\text{g cm}^{-2}_{\text{sapwood d}^{-1}}$) to avoid issues related to tree water storage and
 measurement errors (Phillips and Oren, 1998). Then, we tested J_{SD} variations within sapwood
 areas in the trees and found insignificant azimuthal variation ($p \geq 0.23$) but significant
 250 variation along sapwood depth ($p < 0.001$). Accordingly, we performed a scaling based on
 the radial variation of J_{SD} . First, we evaluated the relationship between the outermost J_{SD} at 0-
 20 mm ($J_{SD,0-20\text{mm}}$) sapwood depth and DBH and found no significant effects of stem size on
 $J_{SD,0-20\text{mm}}$ in either species ($p \geq 0.1$). Therefore, we averaged $J_{SD,0-20\text{mm}}$ across all sampled
 trees and used the data for scaling. Next, we calculated the ratios between J_{SD} at inner
 255 sapwood depths (i.e., 20-40 mm, 40-60 mm and 60-80 mm) and $J_{SD,0-20\text{mm}}$ during the study
 period. Because there was no significant relationship between the ratios and stem size ($p \geq$
 0.16), we averaged the ratios across all trees for each species in each day and used the daily



specific ratios between J_{SD} in the inner sapwood depths and the outermost J_{SD} ($J_{SD,0-20mm}$) for scaling. Sapwood area (A_S , cm^2) for each tree species (*P. sylvestris* and *P. abies*) was

260 estimated from allometric equations derived from > 20 tree cores taken at breast height for each tree species in 2017. Tree cores were taken from individual trees representing the full range of stem diameter distribution at the site and stained with alcohol iodine solution (Eades, 1937) to record the depth of active sapwood thereby allowing the estimation of A_S of all trees. For scaling, we first estimated weighted average J_{SD} of each species ($J_{SD,species}$; $g\ cm^{-2}\ d^{-1}$) using data from the three nodes by

$$J_{SD,species} = \frac{\sum_{i=1}^5 J_{SD,i} \times A_{S,i}}{A_{S,all}} \quad (3)$$

i is the sapwood depth from the inner bark; i.e., 0-20 mm, 20-40 mm, 40-60 mm, 60-80 mm and >80 mm, $J_{SD,i}$ is the average daily sap flux density for each layer and calculated as the product of the averaged ratios and $J_{SD,0-20mm}$, $A_{S,i}$ is sapwood area of layer i and $A_{S,all}$ is the

270 total sapwood area of all trees of the species from all nodes. Then, using this weighted average J_{SD} by species, the canopy transpiration of the C2 subcatchment (T , $mm\ d^{-1}$) was estimated using sapwood area index (SAI , $m^2_{sapwood}\ m^{-2}_{ground}$) of each species, which was derived from data from seven permanent forest inventory plots located within the C2 subcatchment.

$$275 \quad T = 10 \times (J_{SD,pine} \times SAI_{pine} + J_{SD,spruce} \times SAI_{spruce}) \quad (4)$$



where 10 is the unit conversion factor. Regarding methodological considerations, the most common criticism of the heat dissipation method for sap flux measurement, is that it underestimates the flux (Sun et al., 2012; Steppe et al., 2010). However, according to the analysis of 54 data from global pine forests in Tor-ngern et al. (2017) estimates from other
280 sap flux measurement methods showed no particular bias from those with the heat dissipation one as used in this study. In addition, it has previously been shown that radial variation of sap flux density and tree size were more important than species in scaling from single-point sap flux measurements to stand transpiration (Hernandez-Santana et al., 2015), both of which were considered in our analysis. In this study, uncertainty of daily transpiration is represented
285 by standard deviation of T within the seven permanent forest inventory plots.

2.3 Modeling ET partitioning

We also partitioned ET using a detailed ecohydrological model. We used a slightly modified version of the soil-vegetation-atmosphere transfer model APES (Launiainen et al., 2015)
290 which simulates water, energy, and carbon dioxide (CO₂) fluxes in a forest ecosystem in the vertical dimension. The forest is described by a multi-layer, multi-species tree stand, understory vegetation, and a bryophyte layer on the forest floor and thus allows partitioning water fluxes between canopy layers and species. The heat and water fluxes within the soil



profile were not solved in this study, instead we used measured soil moisture and soil

295 temperature at the depth of 0.05 m as lower boundary conditions.

As forcing variables, the model uses time-averaged (here $\frac{1}{2}$ hourly) meteorological variables at a reference level above the canopy. These include P, downwelling longwave radiation, direct and diffuse photosynthetically active and near-infrared radiation, wind speed (or friction velocity), atmospheric pressure, air temperature (T_{air}), and mixing ratios of H_2O and CO_2 . Forcing data were taken from the Svartberget ICOS station using additional
300 meteorological measurements from another ICOS station Degerö (at 15 km distance) to fill missing values. For precipitation the Degerö readings were corrected to match the daily precipitation measured at another station (at 1 km distance) before using them for gap filling.

In APES, the canopy is considered as a layered porous media characterized by leaf-
305 area density (LAD) distribution. The model solves the transfer and absorption of shortwave and longwave radiation (Zhao and Qualls, 2005, 2006) and the transport of scalars (T_{air} , H_2O , CO_2) and momentum among canopy layers (here $n=100$). Partitioning of rainfall between interception and throughfall, as well as the energy balance of wet leaves is also solved for each canopy layer (Watanabe and Mizutani, 1996). The canopy LAD distribution is further
310 divided into LAD distributions for each plant type (e.g. main tree species and understory vegetation). Each plant type is characterized by physiological properties describing phenology, photosynthetic capacity, and stomatal conductance. Leaf gas and energy



exchange is calculated separately for sunlit and shaded leaves of each plant type using well-established coupled photosynthesis–stomatal conductance theories (Medlyn et al.,

2011; Farquhar et al., 1980) and leaf energy balance. A separate forest floor component describes water, energy and CO₂ dynamics in the bryophyte layer (Kieloaho and Launiainen, 2018; Launiainen et al., 2015).

To model the ET components, we established LAD distributions for the main tree species (*Picea abies*, *Pinus sylvestris*, and *Betula pendula*) based on stand inventories from seven forest plots (10 m radius) within the C2 subcatchment. The frequency distributions of diameter at breast height for each species were derived into LAD distributions based on the models by Marklund (1988), Näslund (1936) and Tahvanainen and Forss (2008), and the specific leaf area values reported by Harkonen et al. (2015). The total one-sided leaf area index (LAI_{tot}) was scaled to match the LAI derived from the empirically measured LAI_{Licor} with LAI-2200C Plant Canopy Analyzer (LAI_{Licor} = 2.75) using a correction of 1.6–1.9 (Stenberg et al., 1994), resulting in LAI_{tot} = 4.4–5.2 m² m⁻² (Selin, 2019). The resulting normalized LAD distributions are shown in Fig. S3. In the simulations understory LAI was set to LAI_{under} = 0.4–0.8 m² m⁻², and the bryophyte layer was characterized as feather moss. A full list of model parameters can be found in supplementary Tables S1 and S2.

We simulated the period from May to October 2016, and included parameter uncertainty through parameter ranges for LAI_{tot}, LAI_{under}, maximum carboxylation rate



(V_{cmax}) at 25°C and interception capacity (see Tables S1 and S2). To assess model performance, model results were evaluated at ½ hourly time interval against ecosystem fluxes (net shortwave and longwave radiation, latent heat, sensible heat and gross primary productivity) observed at the ICOS-Svartberget EC tower (Chi et al., 2019). Performance test against the simulation results for the center of the parameter space showed a good agreement between modelled and measured variables (Fig. S4). Net shortwave and longwave radiation were predicted with good accuracy, which is crucial for further modeling steps. Sensible heat flux was slightly overestimated and latent heat flux slightly underestimated. Model results of ET components were further analyzed on a daily or rain event based time interval and compared against corresponding estimates derived from empirical measurements.

3 Results

Meteorological conditions during the snow free vegetation season of 2016 (Fig. 2) were similar to long term averages. The highest daily mean temperatures were in the middle of July (ca. 20 °C) followed by a gradual decrease to around 0 °C at the end of October. As observed for air temperature, photosynthetically active radiation (PAR) peaked at the end of July and then decreased to less than 20 W m⁻² at the end of October. During the study period daily vapor pressure deficit (VPD) ranged between 0 and 1.5 kPa, with a notable peak in the middle of July, which also corresponded to a peak in air temperature. Total precipitation over



the study period was 226 mm, with a strong peak in early August and another at the end of September. These events are also observed as peaks in stream runoff (Fig. 2c).

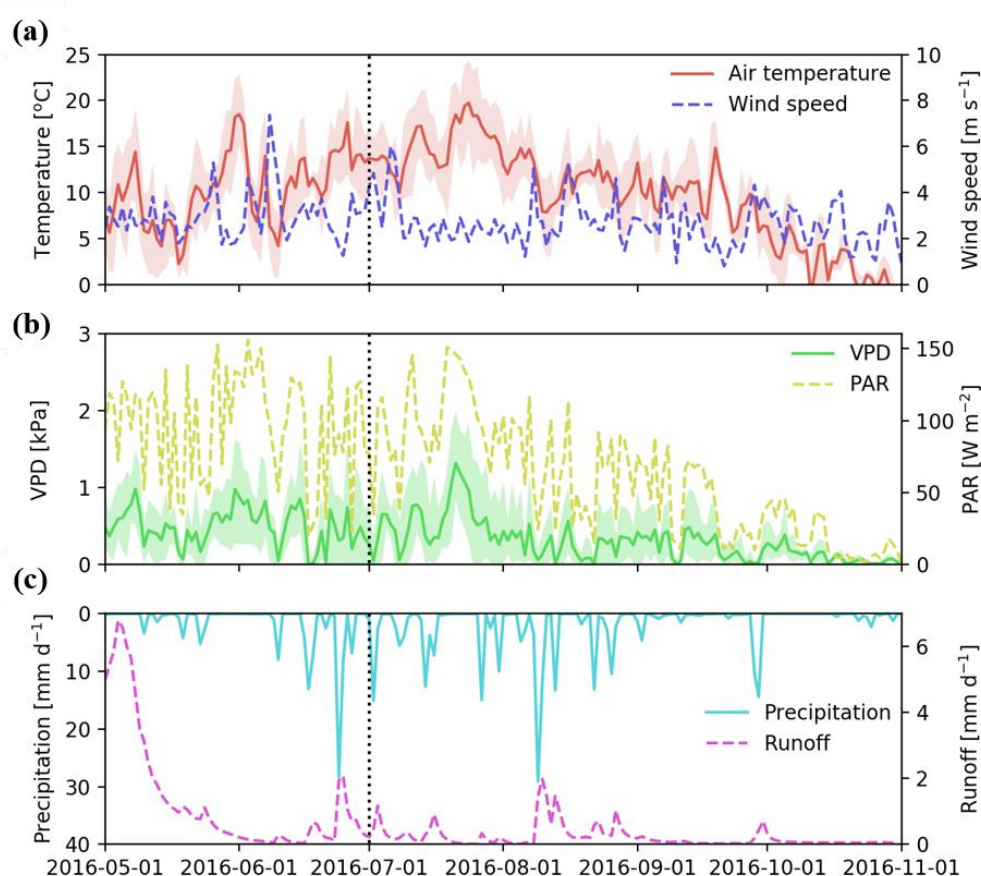


Figure 2. Mean daily hydro-meteorological variables at the Krycklan catchment during the
 2016 growing season: air temperature and wind speed (a); vapor pressure deficit, VPD and
 photosynthetically active radiation, PAR (b); precipitation and stream runoff (c). Beginning
 of study period is marked with a vertical dotted line. Shaded areas for air temperature and
 VPD show minimum and maximum values during a day.



3.1 Daily variability of ET and its components

360 Daily ET over the study period varied between 0 and 4 mm d⁻¹ depending on the weather conditions (Fig. 3a). Except for the very short period following a 29 mm rain event on August 9, ET was always higher than Q. In general, there was good agreement between empirical and modeled estimates of ET ($R^2 = 0.79$; $P < 0.001$; Fig. 3a). Yet during a one-week period in July modeled estimates of ET were 30 % higher than measurement ET, which also
 365 corresponded to the time period of high IL (Fig. 3d).

The amount of water loss via T was substantial, and during 88% of the study period was higher than Q (Fig. 3b). Maximum daily values of T were reached during the latter half of July and occasionally exceeded 2 mm d⁻¹ (Figure 3b). During summer months (JJA) and the first half of September, daily T was on average 0.93 mm d⁻¹ but later substantially
 370 decreased to <0.2 mm d⁻¹. Overall, modelled estimates of T were tightly correlated with T based on sap flow measurements ($R^2 = 0.89$; $p < 0.001$), although the patterns of modelled and measured T diverged during a one-week in July (Fig. 3b).

Total evaporation together with understory transpiration (IL + ET_u) had the highest variability (Fig. 3c) mainly because it includes interception evaporation (Fig. 3d) which is
 375 related to the frequency of rain events and only secondarily to other weather conditions like temperature or VPD (Fig. 3d). Modeled IL + ET_u followed similar pattern to observed, which here is computed as a residual of ET and T (Fig. 3c). During rainy periods total evaporation exceeded T while the opposite was true for rain-free periods, e.g. during the second part of July the contribution of T to total ET was close to 80 %.

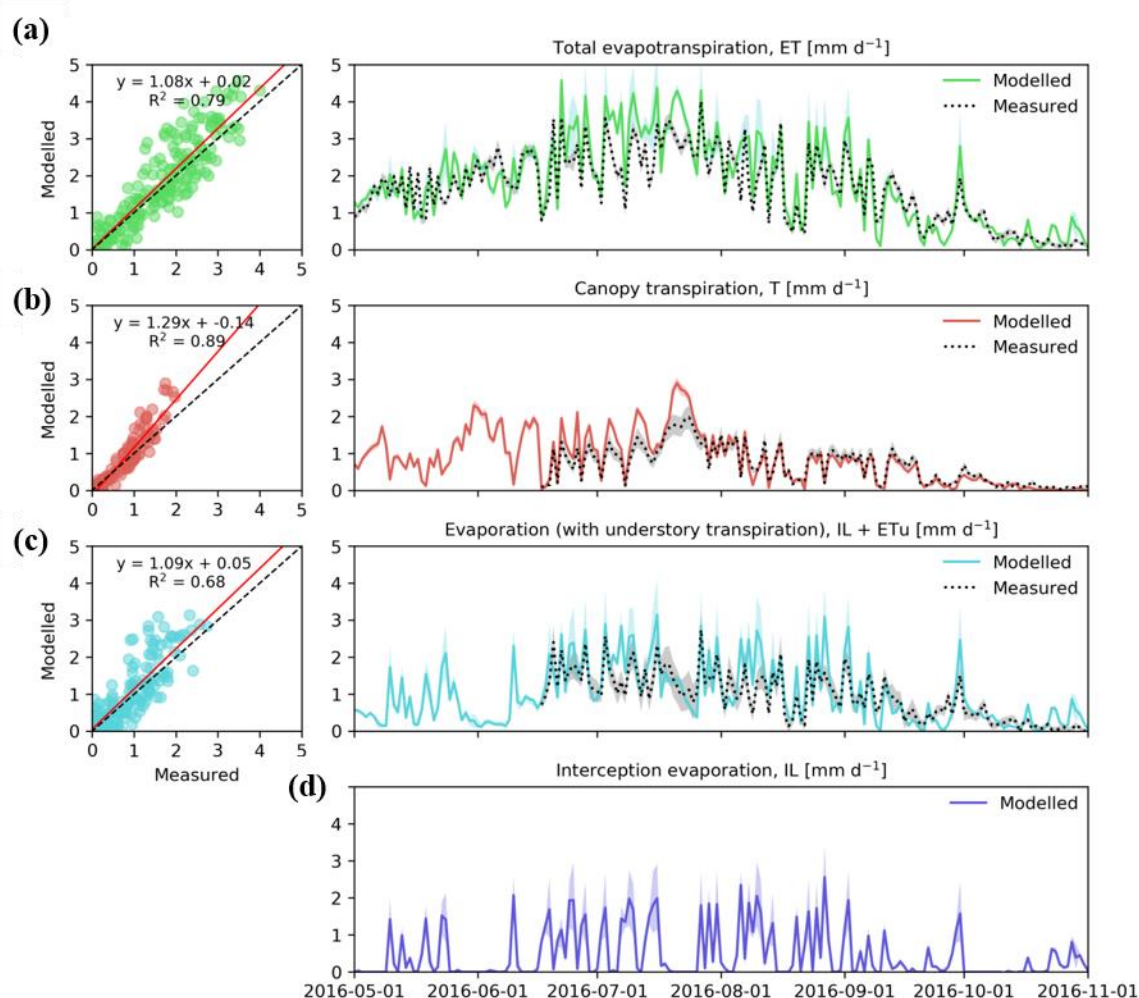


Figure 3. Evapotranspiration (a) and its component flux components: canopy transpiration, T (b), total evaporation with understory transpiration, IL + ETu (c) and canopy interception evaporation, IL (d) in a boreal forest catchment during the 2016 growing season. Small panels on a left side show correlation between daily modelled and measured values. Colored shaded areas show simulation results for whole parameter space and gray shaded areas represents uncertainty in measurements.

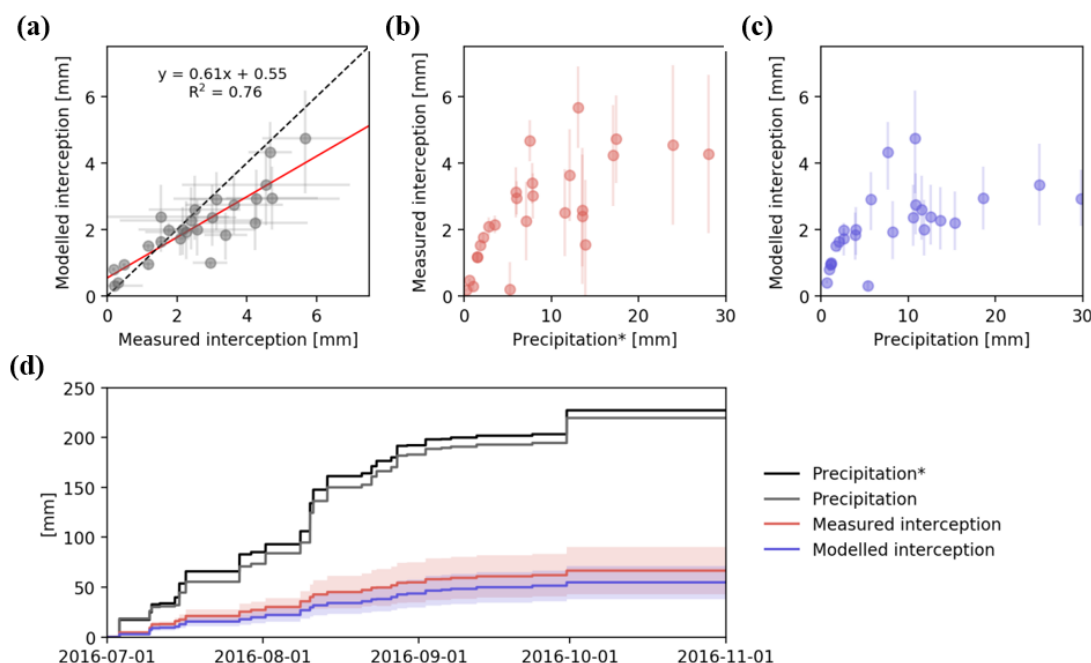


Figure 4. Comparison of measured and modelled event-based interception (IL) (a),

relationship between precipitation and measured interception (b) and modelled interception (c). Cumulative plot of precipitation and interception based on the two different approaches (d). Error bars and shaded areas show simulation results for whole parameter space and uncertainty range in measurements. * refers to precipitation measured in open area at the same location as throughfall (see methods section).

Comparison of IL obtained by empirical measurements and modeling on event based time scale, which is the minimum temporal scale allowed by empirical sampling, showed high correlation, $R^2=0.76$ (Fig. 4a). However, modelled IL values were higher than measured for small rain events whereas the opposite was true for large rain events (Fig. 4a).



400 Uncertainty of both measured and modelled IL increased with the amount of precipitation
(Fig. 4b, c). It is worth noting that precipitation used as reference for the throughfall
measurements was from an open area close to the throughfall collectors, not from the ICOS
station where precipitation was used in the model forcing. The difference in total
precipitation from these two different sources was 7 mm, or 3 % over the entire study period
405 (Fig. 4d).

3.2 Water balance and ET partitioning

During the growing season, the C2 subcatchment received 226 mm of precipitation and
released only 28 mm of water as a stream runoff (Q). Based on EC measurements, ET
410 represented 86 % of precipitation during the study period (194 ± 16 mm), which was similar
to model estimated that showed 96 % of precipitation (217 ± 18 mm) was lost via ET during
the study period (Fig. 5). When partitioning ET, it was found in both approaches that tree
canopy transpiration (T) represents the largest flux of ET (44 % based on empirical
measurements and 41 % based on model prediction). The amount of water lost via IL
415 represents roughly 34 % and 28 % of total ET for the measured and modeling approach,
respectively. When combining T and IL, trees were responsible for 78 % of ET when using
empirical data and 69 % based on the model approach. Estimates of the relative importance
of ET_u was slightly higher when using the modeling approach compared to measure



estimated (31 % and 22 %; respectively). Higher estimates of the relative importance of ETu
 420 in the modeling approach may be the result of reduced IL in the model compared to what was
 measured.

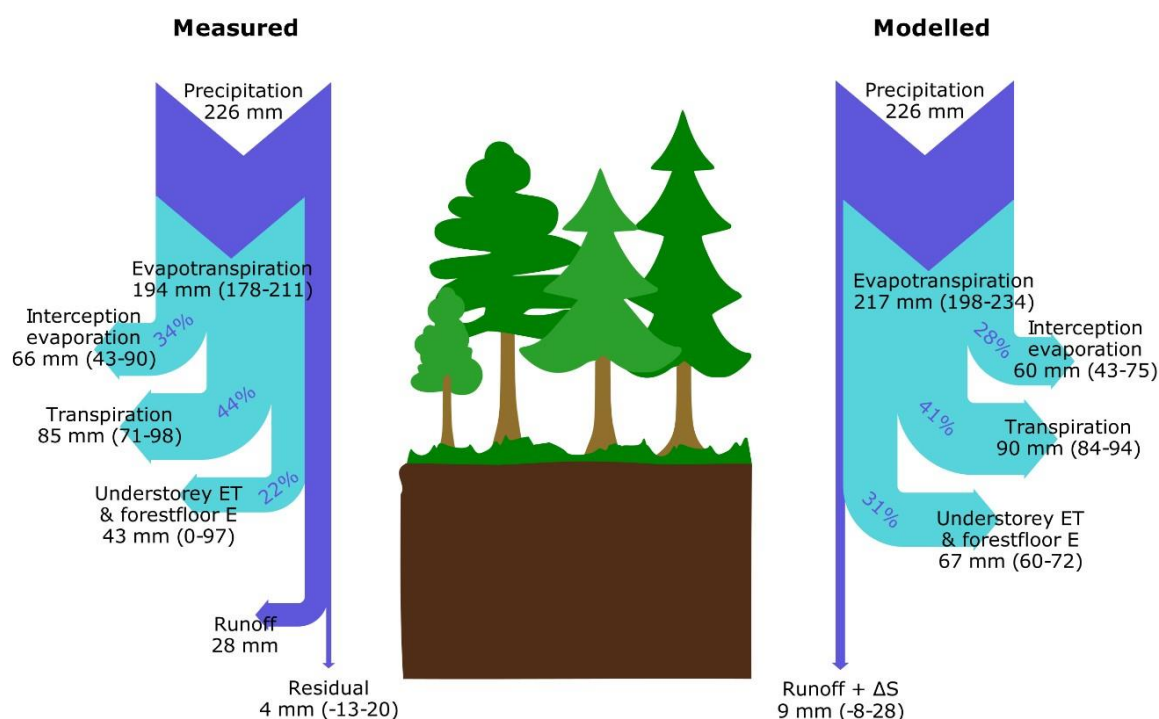


Figure 5. Partitioning of water fluxes based on empirical measurements (left side) and model
 simulation (right side) in a small coniferous boreal catchment during the 2016 growing
 425 season (July-October). Values for each flux component are presented as averages with upper
 and lower boundaries in parenthesis.



430 4. Discussion and Conclusion

In this study, we used both empirical measurements and a process-base model to partitioning ET into its individual flux components and assessed how these different fluxes varied during the course of the growing season in a northern boreal catchment. Both the empirical results and model estimates highlighted the importance of ET losses during the growing season, with
435 ET representing *ca.* 85 % of the incoming precipitation during the study period. Moreover, the results demonstrated that canopy trees are the main driver of ET losses, as canopy transpiration (T) together with canopy interception losses (IL) represented 69-78 % of ET depending on the approach used. Being able to comparing the amount of water lost via different ET flux components to other water loss pathways within the subcatchment further
440 emphasizes the important role canopy trees play in the hydrological cycle during the growing season and stresses the need to better understand the effect of trees on the catchment-scale water budget in response to forest management practices and a changing climate.

It is well known that there is strong seasonal variation in the relative importance of different water loss pathways in northern latitude catchments, with stream runoff being the
445 main water loss pathway during snow melt in spring. Within the Krycklan Catchment, roughly 40 % of annual stream runoff occurs as a response to snowmelt (Ågren et al., 2012), when trees are relatively inactive (Tor-Ngern et al., 2017). In contrast, in this study we found that ET becomes the dominant water loss pathway after spring flood has ceased, and became



seven times greater than stream runoff during the growing season (Fig. 2c, 3a). Moreover, the
450 amount of water lost via ET during a three-week period in the middle of the growing season
was comparable to the amount of water lost through stream runoff during spring peak flow
(59.3 and 89.3 mm; respectively). Such asynchrony in relative importance of different water
loss pathways might be even more pronounced in a future climate when higher air
temperatures and less frequent, albeit more intense, precipitation events can be expected
455 (IPCC, 2018). One future scenario is earlier snow melt, because of higher air temperatures
(Byun et al., 2019), which would result in more water being lost as stream runoff thereby
reducing the annual amount of water available for tree growth during the growing season
(Barnett et al., 2005). In our study, combining precipitation with modelled estimates of ET
and measured stream runoff results in a negative water balance ($P < ET+Q$) during the
460 growing season. This is in agreement with other studies in boreal forests that have found a
negative water balance during the growing season (Sarkkola et al., 2013b; Wang et al., 2017)
which, in turn, could have cascading effects on forest productivity (Barber et al., 2000; Silva
et al., 2010), tree mortality (Peng et al., 2011) and the overall carbon balance in boreal forests
(Ma et al., 2012).

465 Our results further suggest that transpiration was the largest individual water flux
during the growing season, representing *ca.* 40 % of incoming precipitation. Our cumulative
T estimates during the study period (85-90 mm) was similar in magnitude to what has been



observed in other boreal forests (e.g. Sarkkola et al., 2013b; Grelle et al., 1997). When compared to ET, T represented *ca.* 45 % of total ET loss (Fig. 5), which is consistent with other studies in boreal forests that have spanned between 39-85 % (39-48 % - Sarkkola et al. (2013b), 45-85 % - Cienciala et al. (1997), 47 % - Wang et al. (2017), 54 % - Ohta et al. (2001), 65 % - Grelle et al. (1997)). Meanwhile, our estimate of the relative importance of T to total ET, is slightly lower than the global average of *ca.* 60 % (Schlesinger and Jasechko, 2014; Wei et al., 2017). However, it is known that the ratio of T/ET varies considerably among different ecosystems as well as within the same ecosystems (Evaristo et al., 2015; Wei et al., 2017; Peel et al., 2010). Such variation in T/ET may be the result of differences in study location and duration, its spatial scale, forests stand structure, climatic conditions as well as the method used (Schlesinger and Jasechko, 2014). It is important to point out that the two approaches (*i.e.*, empirical measurements and modelling) gave similar estimates of T, both in term of overall magnitude (Fig. 5) and seasonal dynamics (Fig. 3b), thereby giving us confidence in our measurements of T.

There was good agreement on cumulative IL between empirical measurements and model estimates (66 and 60 mm; respectively), which corresponded to *ca.* 30 % of incoming precipitation during the study period. The importance of IL is however not surprising since Barbier et al. (2009) has previously shown that IL accounts for more than 30 % of seasonal precipitation in a wide range of temperate and boreal coniferous forest stands. When



compared to other ET flux components, IL was the second largest ET flux component, representing *ca.* 35 % of total ET. In our study, IL was calculated for each rain event and it is important to point out that the fraction of P lost via IL (*i.e.*, IL/P) during a single rain event varies in response to the magnitude and intensity of P (Gash, 1979; Linhoss and Siebert, 2016; Rutter et al., 1971; Zeng et al., 2000). The highest IL/P are expected to occur during light rainfall events in a dry canopy, whereas IL/P decreases with increasing rain amount and intensity as well as when water storage capacity in the canopy is reduced by intercepted water from previous precipitation events. Given the strong interactions between the amount and intensity of P and IL, projected climate change in northern latitude ecosystems (IPCC, 2014) could drastically affect the amount of water lost via IL that, in turn, could strongly affect the amount of water available to plants, stream runoff and other downstream processes.

Understory evapotranspiration (ET_u) represented 20 to 30 % of total ET depending on the method used, which is consistent with other studies in boreal forests where seasonal ET_u/ET has been shown to range between 10 % and 50 % (Constantin et al., 1999; Iida et al., 2009; Kelliher et al., 1998; Suzuki et al., 2007; Launiainen et al., 2005; Launiainen, 2010). Although ET_u was in general less important than T and IL during the entire study period, it is worth pointing out that ET_u was the largest ET flux component in late autumn when the amount of water lost via T in canopy trees is reduced. Using the APES model we were able to further partition ET_u into forest floor evaporation and understory transpiration. During the



study period, cumulative water loss via forest floor evaporation was 57 mm, representing 85 % of total ET_u, suggesting that evaporative water losses from the moss layer may play an important role in the boreal hydrological cycle, especially in late autumn (Bond-Lamberty et al., 2011; Suzuki et al., 2007). However, it is worth pointing out that ET_u was the component
 510 flux that showed the greatest difference between the two approaches, which stress the need for additional studies to better quantify ET_u and its individual flux components.

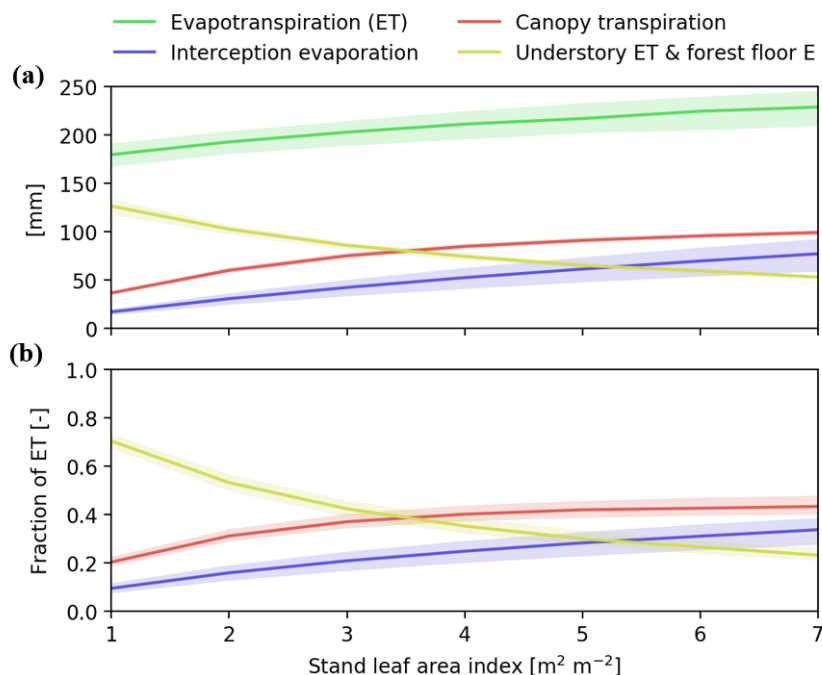


Figure 6. Modelling results for evapotranspiration and its components during July-October
 515 2016 as a function of stand leaf area index in absolute amounts (a) and as a fractions of ET
 (b).



By combining T and IL, we are able to show that trees are directly responsible for *ca.* 70 % of water losses from a northern boreal catchment during the snow free growing season. Similarly, Gu et al. (2018) reported that T and IL together represent 55 to 83 % (average 63 %)

520 %) of total ET across a range of needle-leaved evergreen forests in the boreal and temperate region. Taken together, there is increasing evidence highlighting the important role trees play in the boreal hydrological cycle. Consequently, forest management practices that alter forest stand structure could have large cascading effects on the way water moves through these landscapes (Greiser et al., 2018). For instance, thinning reduces basal area and

525 correspondingly results in a reduction in the LAI of the remaining forest stand, whereas nitrogen fertilization promotes greater aboveground carbon allocation leading to an increase in LAI in northern forested ecosystems (Lim et al., 2015). To assess how forest management practices may influence the overall magnitude of ET as well as the relative importance of the different ET flux components we ran the APES model with a range of canopy LAI values;

530 from 1 to 7 m² m⁻². This resulted in an increase of total ET by *ca.* 50 mm, or 27 %, with increasing LAI (Fig. 6a). The outcome of these simulations also enabled us to identify threshold in canopy LAI where there are changes in the dominant ET flux component. For example, for sparse stands with LAI less than 3 m² m⁻² understory evapotranspiration is the dominant water loss pathway, whereas in forest stands with an LAI greater than 3 m² m⁻²

535 transpiration becomes the dominant ET flux component. (Fig. 6b). Understanding how LAI



influences total ET and its individual flux components thereby provides an opportunity to assess how different forest management practices may affect the movement of water in forested landscapes, and, in turn, could assist in the development of more sustainable management practices (Stenberg et al., 2018; Sarkkola et al., 2013a).

540 In conclusion, this study is unique in that it combines empirical data and a model approach to assess both the vertical and lateral fluxes of water in a northern boreal catchment. Both our empirical measurements and model estimates clearly showed that canopy trees play a central role in the water cycle of northern boreal forests, representing *ca.* 70 % of the water loss during the growing season. Forest management practices that alter forest stand structure, 545 such as commercial thinning, continuous cover forestry, and clear cutting, are therefore likely to have large cascading effects on the way water moves through these forested landscapes. Additionally, it is important remember that this study was conducted during a single growing season, and it is reasonable to assume that changes in climatic conditions could also alter the magnitude and relative importance of different water loss pathways. Thus, further studies are 550 needed to better understand how forest management practices and environmental conditions influence total ET and its individual flux components in order to identify more sustainable forest management practices in a changing climate.



555 **Code and data availability**

Sapflux data is archived in the sapfluxnet data base (<https://github.com/sapfluxnet/sapfluxnet-public/wiki>). Data on greenhouse gas, water and energy fluxes as well as meteorological and environmental data used for model forcing are available through the ICOS portal, Svartberget station (www.icos-sweden.se/station_svartberget.html). Model source code is available upon
560 request.

Author Contributions

N.K., N.J.H., P.T., R.O., and H.L. worked on the conceptualization of the research goals.

N.K., N.J.H. and P.T. installed, collected and analyzed the sapflux data; K.H. and S.L.

565 performed the modelling; J.C and M.P. were responsible for processing the eddy covariance data; E.M.H. and J.W. provided the forest canopy data that was acquired by airborne laser scanning. N.K. and N.J.H. wrote the paper with contributions from all other others.

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Competing interests

The authors declare that they have no conflict of interest.

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