

Sensitivity of Hydrologic and Geologic Parameters on Recharge Processes in a Highly-Heterogeneous, Semi-Confined Aquifer System

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Abstract.

Increasing reliance on groundwater resources has been observed worldwide during the past 50–70 years and has led to unsustainable groundwater abstraction in many regions, especially in semi-arid and arid alluvial groundwater basins. Managed aquifer recharge (MAR) has been promoted to replenish overdrafted groundwater basins and augment surface water supply. However, MAR feasibility in alluvial groundwater basins is complicated by complex geologic architecture that typically includes laterally-continuous, fine-texture confining units that can impede both recharge rates and regional propagation of increases in hydraulic head. Greater feasibility of MAR hinges on identifying locations where rapid, high-volume recharge that provides regional increases in pressure head are possible, but relatively little research has evaluated the factors that control MAR feasibility in alluvial groundwater basins. Here, we combine a transition probability Markov-chain geostatistical model of the subsurface geologic heterogeneity of the east side of the northern Central Valley, California, with the 3D, variably-saturated water flow code, ParFlow, to explore the variability of MAR feasibility in this region. We use a combination of computationally-efficient local and global sensitivity analyses to evaluate the relative importance of factors that contribute to MAR feasibility. A novel proxy parameter approach was used to describe the configuration and proportions of subsurface hydrofacies and water table depth for sensitivity analyses, and results suggest that recharge potential is relatively more sensitive to the variability of this proxy parameter than to the variability of individual hydrofacies hydraulic properties. Results demonstrate that large variability of MAR feasibility is typical for alluvial aquifer systems and that outsize recharge rates are possible in select locations where interconnected, coarse-texture hydrofacies occur.

1 Introduction

Geologic heterogeneity strongly affects both the movement of water in the subsurface and the exchange of water between
20 subsurface and surface stores; however, rarely are enough data available to explicitly represent heterogeneous geologic features
in groundwater models (Koltermann and Gorelick, 1996; De Marsily et al., 2005). Instead, models typically simplify and/or
upscale heterogeneity to represent subsurface flows for purposes of regional-scale water resources management (e.g., Fogg,
1986; Phillips and Belitz, 1991). Upscaling methods have been the focus of numerous studies (e.g., Renard and De Marsily,
1997; Fogg et al., 2000; Neuman and Di Federico, 2003; Fleckenstein and Fogg, 2008), and coarse-resolution models with
25 upscaled (i.e., effective) hydrologic properties are often adequate for regional-scale flow studies, but typically lack enough
detail to reliably capture some phenomena, like recharge and transport processes, that are strongly influenced by geologic
heterogeneity.

To represent the influence of geologic heterogeneity on flow and transport phenomena, many approaches have relied on
stochastic methods, like transition probability based indicator geostatistics which can represent heterogeneous features while
30 honoring measured data (Carle and Fogg, 1996; Weissmann and Fogg, 1999; Weissmann et al., 1999). These approaches
represent geologic heterogeneity with hydrogeologic facies categories, each of which is assigned effective values or probability
densities for estimates of hydraulic properties. By categorizing facies according to depositional environment rather than texture
alone, the predictable geometries (i.e., facies mean lengths, proportions, and juxtapositions) of these features can be more
accurately represented with sparse data. Studies that rely on these methods show strong influence of subsurface heterogeneity
35 on groundwater/surface-water interactions and recharge processes (Lee, 2004; Fleckenstein et al., 2006; Engdahl et al., 2010;
Liu, 2014), including managed aquifer recharge (MAR) (Maples et al., 2019), especially for instances when the mean lengths
and proportions of high-permeability facies allow for percolation, i.e., formation of connected networks (Fogg et al., 2000;
Harter, 2005).

Accurately assigning aquifer properties in models can be a challenge because they are scale dependent attributes that are
40 challenging to measure and can vary over many orders of magnitude in typical aquifer systems (e.g., Sudicky, 1986; Gelhar
et al., 1992; Weissmann and Fogg, 1999). While aquifer tests can accurately constrain estimates of hydraulic conductivity (K)
for high-permeability facies, they are typically unreliable for estimating K of low-permeability (i.e., aquitard) facies (Fogg,
1986; Fogg et al., 1998), which have been shown to influence pumping response (Fogg et al., 2000) and be important for
accommodating recharge (Maples et al., 2019). Reconciling typically sparse measurements of aquifer properties from aquifer
45 tests with the representation of effective values in models is often the source of large uncertainty because parameterization of
the properties in models is scale dependent (Sudicky and Huyakorn, 1991), and is typically achieved through model calibration.

The contaminant transport community has long recognized the strong influence of K scaling and geologic heterogeneity on
transport processes (e.g., Gelhar et al., 1992; Sudicky and Huyakorn, 1991; Koltermann and Gorelick, 1996), and recent work
has extended these concepts to assess their role on runoff generation, evapotranspiration (ET), and feedbacks between subsur-
50 face and land-surface water in integrated hydrologic models (Srivastava et al., 2014; Gilbert et al., 2016; Foster and Maxwell,
2019), but relatively little research has focused on the influence these factors for MAR processes specifically. Recent work has

highlighted the importance of connected networks of high- K facies for MAR (Maples et al., 2019), but to our knowledge the sensitivity of MAR processes to these heterogeneous geologic features as compared to other uncertain hydraulic properties has not been formally evaluated. Here, we simulate variably-saturated MAR dynamics in a highly-resolved representation of complex subsurface geologic heterogeneity of a clastic, unconsolidated sedimentary aquifer system that includes both interconnected, high- K sand and gravel deposits intermingled with silt- and clay-dominated sediments. We use a combination of local and global sensitivity analyses to provide insight into the relative importance of the subsurface geologic facies configuration and parameterization of subsurface hydraulic properties on MAR processes. This work provides insight into important factors to consider when investigating potential MAR sites and also highlights the utility of a combination of computationally-frugal local and global sensitivity analyses for computationally-intensive hydrologic models.

2 Materials and Methods

2.1 Local Hydrogeology and Domain Extent

The model domain covers about ~~1640m~~1640 m² of the east side of the northern Central Valley, California, near the convergence of the lower portions of the American and Cosumnes Rivers with the Sacramento River (Fig. 1). The domain comprises a low-angle alluvial fan complex that is typical of the Central Valley where previous studies have documented the presence of deposits that are favorable for recharge (Shlemon, 1967; Meirovitz, 2010), including massive, interconnected, highly-permeable sand and gravel deposits known as incised valley fill [IVF, (Weissmann et al., 2004, 2005)] that form from river incision and deposition events during cyclic Plio-Pleistocene Sierra Nevada glaciation. In places, multiple IVF deposits have been shown to overlap and interconnect from land surface into the deeper aquifer system, forming massive, coarse-texture, relatively high-permeability pathways for recharge that bypass local, otherwise laterally-extensive confining units. These features have been shown to accommodate recharge volumes that are orders of magnitude greater than would be possible over the rest of the landscape (Maples et al., 2019). Other studies have shown that IVF features likely occur on river fans throughout the Central Valley (Weissmann et al., 2005), and in similar glacially-influenced rivers (Pierce and Scott, 1983) but are still largely undocumented.

The local hydrostratigraphy of the area is described in detail by Meirovitz (2010) and Maples et al. (2019). In general, the two major rivers intersecting the domain, the American and Cosumnes Rivers, have markedly different depositional characteristics. The American River drains a large (>4000 km²), high-elevation catchment that extends to the Sierra Nevada crest [>3000 m above mean sea level (amsl)]. As a result, the American River was greatly influenced by cyclic plio-pleistocene glaciation that deposited IVF in the domain area. Conversely, the Cosumnes River catchment is smaller (900 km²) and lower in elevation. As a result, deposits from the Cosumnes River do not contain IVF and are typically finer in texture. In some locations in the domain area, Quaternary and Holocene channel avulsion of the American River resulted in a more southwest course that intersects the current path of the Cosumnes river, creating complex overlapping stratigraphy in that area. Cross-cutting IVF and overlapping paleochannel networks in the domain area result in an aquifer system that is typically unconfined (and sometimes perched) or semi-confined at shallow depths and increasingly confined with depth (Fleckenstein et al., 2006; Liu, 2014; Niswonger and

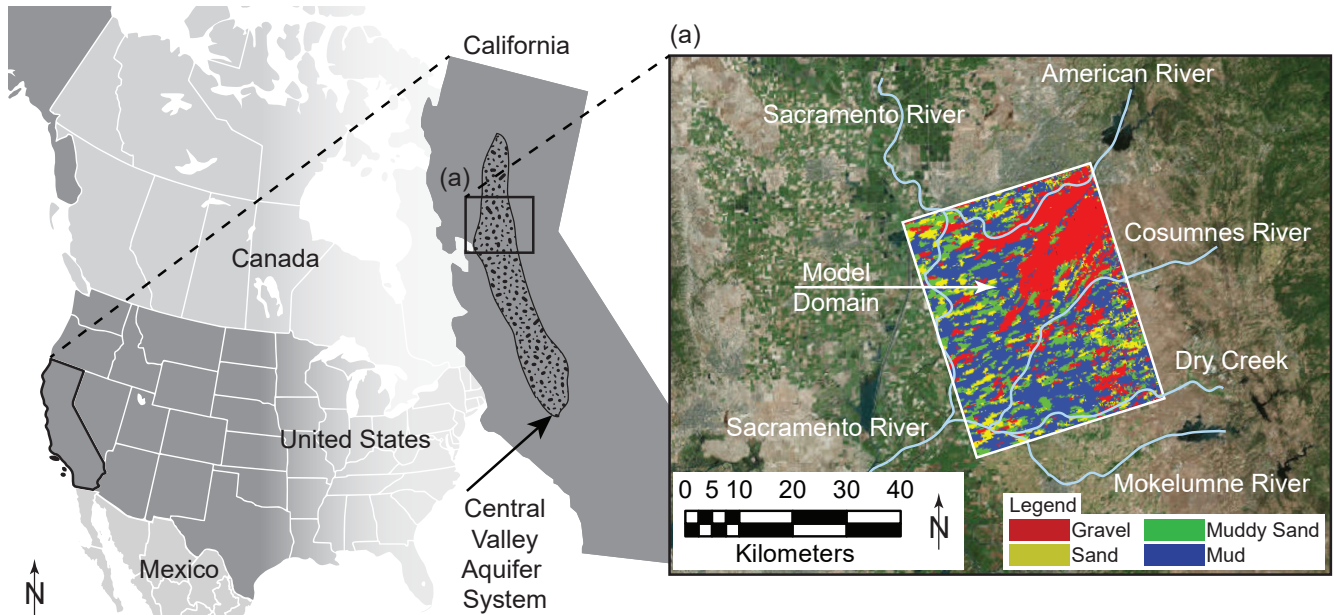


Figure 1. Location of model domain in the in the Central Valley aquifer system in California, and (a) inset of the uppermost layer of the model hydrofacies overlain over aerial imagery of the Central Valley and local river systems (Maples et al., 2019). (double-column width)

Fogg, 2008). Groundwater pumping in the region typically occurs at depths >30 m in the deeper semi-confined or confined portion of the aquifer system (Liu, 2014).

2.2 Hydrofacies Model Development

Transition probability Markov-chain geostatistics (TPROGS) (Carle and Fogg, 1996, 1997; Carle, 1999), was used to simulate the subsurface distribution of hydrofacies in the domain area (Fig. 2). Model development is described in detail by Meirovitz (2010) and Maples et al. (2019). Conditioning data for the TPROGS model included about 1200 well logs, soil surveys, geologic cross-sections and mapped paleochannels. Geologic data were binned into four textural categories: gravel, sand, muddy sand, and mud (undifferentiated silt and/or clay) [Table 1 (Fleckenstein et al., 2004; Meirovitz, 2010)]. "Mud" refers to silt and clay undifferentiated, because most of the subsurface data available only identify the fine-grained sediments and are not sufficiently detailed to distinguish silt from clay. From these data the proportions for each facies were calculated directly. ~~Through geologic analysis of the data, additional~~ Additional parameters were estimated describing the mean lengths of each hydrofacies along the principal directions and the embedded transition probabilities to represent cross-correlation between different facies. Because the depositional characteristics of the American and Cosumnes fans were markedly different, individual models of each were produced and subsequently combined by Meirovitz (2010).

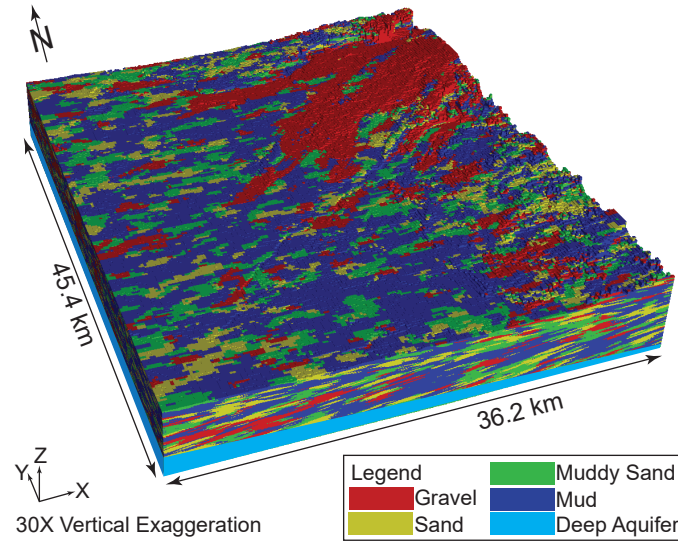


Figure 2. Three-dimensional representation of the model domain (Meirovitz, 2010; Liu, 2014).

TABLE 1 ABOUT HERE.

The model uses an orthogonal grid geometry with 181 cells \times 227 cells \times 265 cells in the x-, y-, and z- directions, respectively. The x- and y- directions of the grid were rotated 17.85 degrees counterclockwise from the cardinal directions, and the z-direction was oriented vertically. Cell sizes were 200 m \times 200 m \times 1 m. The total domain size is 36.2 km \times 45.4 km \times 0.265 km, in the x-, y- and z-directions, respectively. Cells located above land surface were designated as inactive in the model, resulting in about 7.3 million active cells in the domain area.

2.3 Hydrologic Model Development

105 2.3.1 Governing Equations

Three-dimensional, variably-saturated water flow was simulated with the hydrologic modeling code, ParFlow (Ashby and Falgout, 1996; Jones and Woodward, 2001; Kollet and Maxwell, 2006), which couples surface and subsurface flow with the 2-D diffusive or kinematic wave equation, and solves the 3-D mixed form of Richards' Equation for variably-saturated subsurface flow:

$$110 \quad S_{ss} S_{ww}(h) \frac{\partial h}{\partial t} + \phi \frac{\partial S_w(h)}{\partial t} \frac{\partial S_w(h)}{\partial t} = \nabla \cdot \mathbf{q} + q_{r,r}(x, z) \quad (1)$$

where

$$\mathbf{q} = \phi S_{ww}(h) \mathbf{v} = -\mathbf{K}_{ss}(x) k_{r,r}(h) \nabla(h + z) \quad (2)$$

In these equations S_s is specific storage [L^{-1}], S_w is relative saturation [-], h is pressure head [L], t is time [T], ϕ is porosity [-], \mathbf{q} is Darcy flux [$L T^{-1}$], q_r is a source/sink term [T^{-1}], z is elevation [L], \mathbf{v} is the subsurface flow velocity [$L T^{-1}$], $\mathbf{K}_s(\mathbf{x})$ is the saturated hydraulic conductivity tensor [$L T^{-1}$], and k_r is relative permeability [-]. The van Genuchten relations (Van Genuchten, 1980) describe S_w and k_r as a function of h in the unsaturated zone, with parameters for air entry pressure α [L^{-1}], pore size distribution n [-], and residual saturation S_{res} [-].

2.3.2 Boundary Conditions

Model boundary conditions are discussed in greater detail in Liu (2014) and Maples et al. (2019). The locations of domain boundaries were chosen to simplify the assignment of boundary conditions for the flow model. The eastern boundary roughly coincides with the Sierra Nevada foothills, and the northern, southern, and western boundaries roughly coincide with local surface water bodies (Fig. 1). A specified head boundary condition was applied for the eastern boundary to coincide with the local groundwater head distribution estimated from local monitoring well data (Liu, 2014). A general head boundary of 0 m amsl was set 1 km beyond the western boundary to approximate the Sacramento River and Sacramento-San Joaquin Delta along the northwestern, and southwestern portions of the western boundary, respectively. No-flow boundary conditions were applied along northern, southern, and bottom boundaries because the regional groundwater flow direction is generally from east to west.

Model spin up and recharge simulations used combinations of specified-flux and specified-head upper boundary conditions and are described in greater detail in subsequent sections.

2.3.3 Model Spinup and Calibration

Model spinup and calibration are described in greater detail in Liu (2014) and Maples et al. (2019). To summarize, a 16-yr simulation period was used to bring the simulated hydrology into dynamic equilibrium. Water budget components, including groundwater discharge, recharge and boundary flows along with facies hydraulic properties were estimated and adjusted manually to simulate a realistic water budget, water table configuration, and vertical hydraulic gradients during the calibration process. An initial potentiometric surface was specified using interpolated groundwater level data. Monthly estimated urban and agricultural groundwater pumping rates were applied as specified fluxes representing wells screened in lower portions of the domain that coincide with typical screened intervals of municipal and agricultural pumping wells in the region. Dominant sources of recharge for the region include stream recharge from the American River, Cosumnes River, and Deer Creek, as well as deep percolation from agricultural and urban return flows. Weekly estimates of spatially-distributed river stage for the streams were applied as specified heads along coincident land surface cells. Monthly estimates of urban and agricultural recharge volumes were applied as specified-flux boundary condition across the top of the domain to simulate deep percolation of agricultural and urban return flows and to equilibrate soil moisture conditions in the near-surface UZ cells.

2.3.3 Model Spinup and Calibration

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150 literature values for the Central Valley, California, and for similar alluvial systems (Anderson et al., 2015; Botros et al., 2009; Fleckenstein et al., 2004; Frei et al., 2009; Maserjian, 1993; Niswonger and Fogg, 2008; Sager, 2012).

TABLE 2 ABOUT HERE.

All simulations were performed using the Cheyenne high-performance cluster at NCAR's Computational and Information Systems Laboratory (doi:10.5065/D6RX99HX). The numerical problem was distributed on 540 cores for each simulation.
155 Approximately 450 model evaluations were required for the exploratory simulations, local sensitivity analyses, and global sensitivity analyses described in subsequent sections, which required approximately 400,000 core-hours of computing time in total. The large computational expense for each simulation (890 core-hours per simulation, on average) required that computationally resources be allocated efficiently.

2.4 Exploratory Simulations

160 2.4.1 Site Selection

One hundred 1 km² recharge sites, each encompassing 25 upper-boundary cells, were chosen to approximate hypothetical MAR infiltration basins (Fig. 3a). Each site was randomly selected from a 910 km² region within the domain that excluded locations within 5 km of lateral domain boundaries to minimize the influence of boundary conditions. The 100 exploratory sites encompass roughly 6% of the total domain area, which was deemed sufficient to sample the variability of site characteristics
165 observed across the domain. The size of each site was chosen to reflect a regional-scale MAR site, which range from large networks of basins >25 km² in size (e.g., Kern Water Bank Authority, 2018) to individual infiltration basins over several hectares or smaller (e.g., Beganskas and Fisher, 2017) in California.

2.4.2 Site Characteristics

Maples et al. (2019) highlighted that the (1) relative proportions, and degree of vertical interconnection, of coarse-texture facies
170 (sand and gravel) and (2) the unsaturated-zone thickness beneath recharge sites are important factors for recharge feasibility. In this study, we sought to develop site characteristics to describe recharge feasibility at the 100 exploratory sites across the domain, ~~including approximations of how effective (i.e., upscaled) vertical K_s varies spatially. Here, First,~~ we use a heuristic approach of simple averages to bound the expected range of ~~upscaled vertical K_s , effective (i.e., upscaled) vertical K_s at all~~

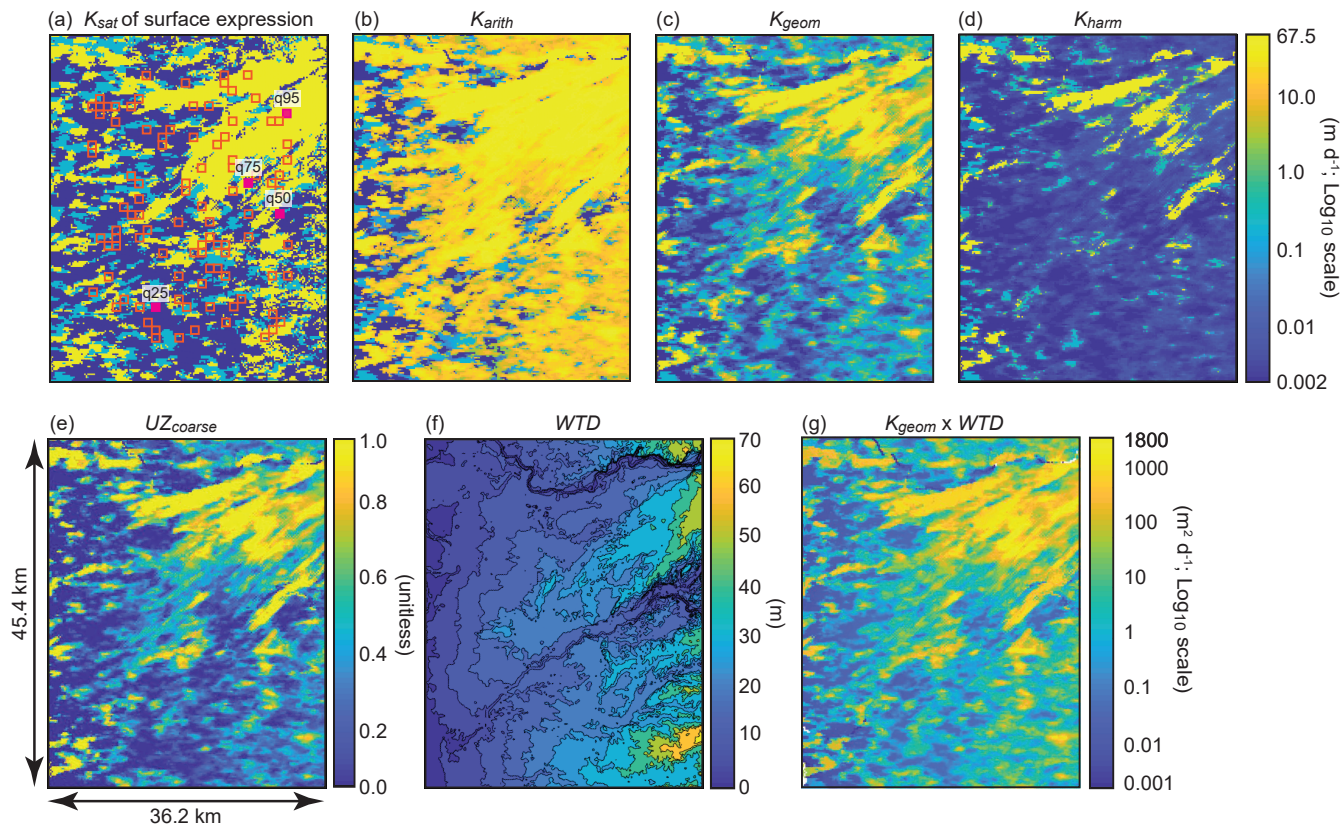


Figure 3. Plan view of the model domain, where the first row shows (a) K_{sat} of the uppermost model layer (i.e., the surface expression) overlaid with the locations of 100 randomly-sampled 1-km² exploratory recharge sites (red squares) randomly chosen from 910 potential locations as well as the 4 sites chosen for sensitivity analyses (q95, q75, q50, and q25; solid violet squares), along with (b) arithmetic, (c) geometric, and (d) harmonic mean of vertical K_{sat} for unsaturated zone (UZ) facies (i.e., K_{geom} , K_{arith} , K_{harm} , respectively). The second row shows (e) the coarse-texture (gravel and sand) fraction of UZ facies (UZ_{coarse}), (f) simulated initial depth-to-water (WTD), and (g) K_{geom} multiplied by WTD ($K_{\text{geom}} \times WTD$). Locations >5 km from the lateral domain boundaries were excluded from the potential sites to avoid interference with boundary conditions. (*double-column width*)

175 41,087 x,y cell locations across the domain, where the arithmetic and harmonic mean (K_{arith} and K_{harm}) are the upper and lower bounds, respectively, and the geometric mean (K_{geom}) is an intermediate value. K_{arith} and K_{harm} are typically used to approximate groundwater flow parallel and perpendicular to layering, respectively, in anisotropic systems (Freeze and Cherry, 1979). This concept has been generally been extended to variably-saturated flow (Mualem, 1984; Assouline and Or, 2006) (Mualem, 1984; Yeh et al., 1985a, b; Assouline and Or, 2006). Fogg et al. (2000) showed that vertical groundwater flow in

180 systems with vertically-connected networks of permeable facies tends toward values between K_{arith} and K_{geom} . ~~Values for For~~
~~each x,y cell location across the domain.~~ K_{arith} , K_{geom} , and K_{harm} ~~for n vertically-coincident cells~~ are given as:

$$K_{arith} = \frac{K_1 + K_2 + \dots + K_n}{n} \quad (3)$$

$$K_{geom} = \sqrt[n]{K_1 \times K_2 \times \dots \times K_n} \quad (4)$$

$$K_{harm} = \frac{n}{\frac{1}{K_1} + \frac{1}{K_2} + \dots + \frac{1}{K_n}} \quad (5)$$

185 ~~Site characteristics are described in detail in Table 3 and include upscaled K_s approximations K_{arith} , K_{geom} , and K_{harm} ,
along with where n is the number of vertically-coincident cells from the land surface to the initial water table depth (i.e., the
unsaturated-zone cells) for each x,y cell location. In addition, the initial unsaturated-zone thickness (WTD), and proportion
of coarse-texture sand and gravel unsaturated-zone facies (UZ_{coarse}), and proportion of coarse-texture facies at land surface
($Surf_{coarse}$). Each site characteristic was calculated for a control volume encompassing vertically-coincident cells from the land
surface to the initial water table depth for each site. For the purposes of this work, the interface of the deeper aquifer system was
190 designated as the initial water table depth (WTD). Additional site characteristics were developed by combining existing site
characteristics at each x,y cell location across the domain were included as metrics. Additional metrics were developed for each
x,y cell location by combining individual metrics, i.e., WTD was used as a multiplier for K_{arith} , K_{geom} , K_{harm} , $Surf_{coarse}$, and
 UZ_{coarse} . Spatial distributions of select ~~site characteristics metrics at all x,y cell locations~~ are shown in Fig. 3.~~

TABLE 3 ABOUT HERE.

195 ~~Site characteristics were calculated for the 100 1 km² exploratory sites from these metrics by determining the average and
maximum value of K_{arith} , K_{geom} , and K_{harm} for the 25 x,y cell locations encompassing each site. In addition, the proportion
of coarse-texture facies at land surface ($Surf_{coarse}$) were calculated for each site. Site characteristics are described in detail
in Table 3. Each site was further evaluated according to whether there was vertical connectivity (i.e., percolation) of coarse-
texture facies from land surface into the deeper to the aquifer system. Intereconnection was Percolation was evaluated for a
200 control volume encompassing all cells from the land surface to the initial water table depth (i.e., unsaturated-zone cells) at
the 25 x,y cell locations encompassing each site. Percolation was defined within each site control volume by a 6-connectivity
metric (Pardo-Iguzquiza and Dowd, 2003), in which neighboring gravel and sand cells are said to be connected if they intersect
along a face. Coarse-texture facies were said to percolate if any combination of gravel and sand facies were interconnected
vertically within the control volume from land-surface to the initial water table.~~

205 2.4.3 Recharge Scenarios and Model Post-Processing

To evaluate the system response to recharge stress and recovery, 90-day recharge scenarios were run individually at each site, wherein recharge was simulated during the initial 30-day period followed by a 60-day recovery period. Surface ponding was approximated by a specified head boundary condition representing 10 cm of ponding at the 25 upper-boundary cells coincident with each recharge site, with no recharge specified for the remaining upper-boundary cells. An additional simulation was run
210 in which no recharge was specified for all upper-boundary cells, i.e., as a no-recharge scenario. The initial condition for all recharge and no-recharge scenarios was the h distribution from the end of the model spin up. Results were output at 5-day intervals for all simulations.

Following Maples et al. (2019), recharge responses were isolated from other model stimuli by differencing h , and total subsurface water storage (TSS) from each colocated cell at each timestep, in each recharge and no-recharge simulation.
215 In this way, perturbations in h and TSS result from the recharge stress alone, while other stimuli, including transient model response to regional boundary condition effects, are eliminated. For each simulation, h perturbations were evaluated at a 10-cm threshold. Domain-wide perturbations in h and TSS from recharge stress were evaluated for all 100 [recharge simulations sites](#) at $t = 30$ days to calculate the volumetric extent of subsurface pressure perturbation, P_{30d} [L^3], and the effective recharge rate, R_{30d} [$L T^{-1}$], respectively, at the end of the 30-day recharge stress period for each site. Similarly, domain-wide perturbations
220 in TSS were evaluated at $t = 90$ days and were further delineated according to whether the change-in-storage occurred in fine-texture (muddy sand and mud) or coarse-texture (sand and gravel) facies, so that the proportion of the total recharge volume accommodated by fine-texture facies, $V_{\text{fines}, 90d}$ [–], could be evaluated. Previous work has highlighted the importance of fine-texture facies for accommodating recharge, especially during late time (e.g., Maples et al., 2019). All model outputs used for subsequent analyses are shown in Table 3.

225 2.4.4 Relations between Site Characteristics and Recharge Potential

To better understand the relationships between site characteristics ([variablespredictors](#)) and model outputs ([predictorsvariables](#)), correlations (Pearson's r , Spearman's rho, and Kendall's tau) were evaluated between all variable and predictor pairs across all 100 sites. Variables and predictors are described in Table 3. The purpose of evaluating correlations between variables and predictors was to determine if any site characteristics could be used to reasonably predict model outputs with empirical relations.
230 \log_{10} data transformations were selectively performed on variables and predictors to improve normality prior to calculation of Pearson's r . Transformations were not performed for Spearman's rho and Kendall's tau, as neither require normal distributions for prediction.

2.4.5 Development of a Geologic Proxy Parameter for Recharge Potential

To incorporate descriptions of geologic configuration in sensitivity analyses of recharge potential, development of a geologic proxy parameter (GPP) was required. Correlations between select variables (R_{30d} , P_{30d} , and $V_{\text{fines}, 90d}$) and predictor pairs
235 described in section 2.4.4 were ranked, and a GPP was determined for each by developing empirical regression relations

between those variables and the highest ranked predictor. Unlike other parameters, the *GPP* cannot be directly varied at each site, and instead was approximated either with linear regression of recharge response and *GPP*, or by relocating the recharge site within the domain to a location with the corresponding *GPP*. Previous studies have shown the importance of geologic heterogeneity on MAR feasibility (e.g., Maples et al., 2019), and the novel proxy parameter methodology described here is analogous to a transfer function (e.g., Wösten et al., 2001), in that it describes the influence of complex geologic heterogeneity on recharge processes with relatively easily-derived site characteristics. By using this approach in sensitivity analyses, we are able to both capture this geologic complexity and also reduce the overall computational expense, albeit with some predictive uncertainty related to the regression relations.

2.5 Sensitivity Analyses using Fit-Independent Statistics

Realistic ranges of model parameters describing eight six hydraulic properties for four facies types ($n = 24$ model parameters) are shown in Table 4 along with an estimated range of *GPP*. Ranges for model parameters were chosen from literature values for the Central Valley California, and for similar alluvial systems (Anderson et al., 2015; Botros et al., 2009; Fleckenstein et al., 2004; Frei et al., 2009; Maserjian, 1993; Niswonger and Fogg, 2008; Sager, 2012). Model parameters were assumed to be distributed uniformly within each of these ranges for simplicity. The range of *GPP* was determined from the range observed from the 100 exploratory sites described in section 2.4. The distribution *GPP* was observed to be approximately log-normal, so a Log_{10} data transformation was performed for subsequent sensitivity analyses.

TABLE 4 ABOUT HERE.

All sensitivity scenarios were initialized with the h distribution from the end of the model spin up and were simulated and post-processed following the approach outlined in section 2.4.3. In this way, each scenario required two simulations be run with the same parameter sets (i.e., a recharge and no-recharge simulation) which were then differenced to isolate recharge stresses from other model stimuli, including transient model responses to changes in parameter values.

2.5.1 Local Sensitivity Analyses

Parameter sensitivities were evaluated locally using the dimensionless-scaled sensitivity (*DSS*) and composite-scaled sensitivity (*CSS*) metrics, which are computationally-frugal screening methods used to compare the relative importance of different parameters to the estimation of a simulated model output (Hill and Tiedeman, 2007). *DSS* for simulated output i and parameter j are calculated as

$$DSS_{ij} = \left(\frac{\partial y_i'}{\partial b_j} \right) \Big|_{\mathbf{b}} |b_j| \omega_{ii}^{1/2} \quad (6)$$

where y_i' is the i th simulated output, b_j is the j th estimated parameter, $\partial y_i' / \partial b_j$ is the derivative (i.e., the sensitivity) of the simulated output with respect to the j th parameter, \mathbf{b} is the vector of parameter values at which sensitivities are evaluated, and

$\omega_{ii}^{1/2}$ is the weight of the i th simulated output. For this work, simulated outputs P_{30d} , R_{30d} , $V_{\text{fines}, 90d}$ were weighted equally at unity.

Composite-scaled sensitivities (*CSS*) were calculated to estimate the total amount of sensitivity provided by each parameter across multiple sites and for multiple model outputs

$$270 \quad CSS_j = \sum_{i=1}^n \left[(DSS_{ij})^2 \Big|_{\mathbf{b}} / n \right]^{1/2} \quad (7)$$

where DSS_{ij} is from equation 6, and n is the total number of simulated outputs i associated with parameter j .

DSS were estimated for select model outputs R_{30d} , P_{30d} and $V_{\text{fines}, 90d}$ (Table 3) by perturbing each hydraulic-property parameter ($n = 24$) by 10% of its total range (Table 4). Results from the exploratory simulations show that recharge response is highly dependent on site choice, so DSS was evaluated at four representative sites which span a large range of recharge
275 potential. Each of the four representative sites were chosen to correspond with the 25th, 50th, 75th and 95th percentile of recharge potential, as estimated by *GPP*. These sites are hereto referred to as q25, q50, q75, and q95, respectively. A total of 96 model evaluations were required to estimate DSS on the three model outputs (R_{30d} , P_{30d} and $V_{\text{fines}, 90d}$) for all 24 parameters (Table 4) at these four sites. To incorporate *GPP* in DSS analyses, an approach was developed using the predictive regression relation between *GPP* and R_{30d} . For example, DSS require perturbation of a parameter (i.e., ∂b_j) by a percentage of the
280 parameter range (e.g., by 10%). A corresponding 10% perturbation in R_{30d} (i.e., $\partial y'_i$) was approximated using the predictive regression relation between *GPP* and R_{30d} rather than by performing an additional model evaluation.

CSS were calculated for R_{30d} , P_{30d} and $V_{\text{fines}, 90d}$ (Table 3) by combining DSS estimates for each model output across sites q25, q50, q75, and q95 for each of 24 model parameters. The same approach was used to calculate CSS for *GPP*, but was only estimated for R_{30d} and not for P_{30d} and $V_{\text{fines}, 90d}$.

285 2.5.2 Global Sensitivity Analyses

A measure of global sensitivity was provided by the method of Morris (1991), which relies on the calculation of elementary effects, i.e., local derivatives sampled one-at-a-time (OAT) on a grid that covers the parameter space. The method of Morris creates a trajectory through the parameter space by perturbing each parameter x_j along a grid by a step Δ_j . A sequence of p perturbations is required to obtain a one trajectory for a model with p parameters. For each trajectory, the elementary effect for
290 a single parameter, $EE_{\underline{j}}$, is calculated as the ratio of the perturbation in model output to the perturbation of the parameter

$$EE_{\underline{j}} = \frac{f(x_1, \dots, x_j + \Delta_j, \dots, x_p) - f(x)}{\Delta_j} \frac{f(x_1, \dots, x_j + \Delta_j, \dots, x_p) - f(x)}{\Delta_j} \quad (8)$$

where $f(x)$ is the evaluation of the function at the prior point in the trajectory. Calculating the elementary effects for p parameters using a single trajectory requires $p + 1$ model evaluations. Because the elementary effect for any single trajectory does not account for interactions between parameters and depends strongly on the location of the initial point, x , in the parameter space,

295 the method of Morris performs the OAT approach over multiple trajectories, N , within the parameter space using a factorial sampling approach. A variation of the original approach was employed to resolve issues related to opposite signs of elementary effects affecting the calculation of total-order sensitivity (Campolongo et al., 2007), in which the total-order sensitivity of each parameter, μ_j^* , was calculated as the mean of the absolute values of N elementary effects

$$\mu_j^* = \frac{1}{N} \sum_{k=1}^N |EE_j^k| \quad (9)$$

300 Unlike *DSS* approaches, the method of Morris required significantly greater computational resources for an equivalent number of parameters. ~~so a sub-set of three parameters were included in the Morris approach. Results from local sensitivity analyses suggest that K_s and *GPP* are relatively important parameters for recharge potential.~~ Computational expense was further reduced by pairing parameters, i.e., K_s of gravel and sand, and of muddy-sand and mud, effectively reducing these four parameters into two describing 'coarse-' and 'fine-texture' facies, respectively. By pairing parameters, K_s of gravel and sand
 305 (and of muddy sand and mud) are perturbed within their respective parameter ranges together, ~~reducing the total number of parameters from five to three. In total, three parameters were included in the Morris approach, including K_s of coarse-texture facies, K_s of fine-texture facies, and *GPP*.~~ Sensitivity indices were calculated using a sample size of $N = 20$, resulting in a total of 80 model evaluations. Herman et al. (2013) demonstrated that the method of Morris with $N = 20$ trajectories produced similar sensitivity results to the Sobol' method (Sobol, 2001) with >2 orders-of-magnitude fewer model evaluations.

310 To further reduce the computational expense, the total simulation time was reduced from 90 to 10 days, during which recharge was applied for the entire simulation. Sensitivity indices, Morris μ_j^* , were only evaluated with respect to the effective recharge rate at the end the 10-day simulation period, R_{10d} [cm d^{-1}]. Morris μ_j^* was not evaluated with respect to other model outputs describing pressure perturbation or volume of recharge accommodated by fines because *GPP* was determined to be an inadequate predictor of these model outputs.

315 To incorporate *GPP* in the Morris framework, a novel approach was developed in which the location of the sampling site was varied to correspond with the requisite *GPP* parameter choice. For example, if a hypothetical sensitivity analysis required evaluation of the model with *GPP* at the 50th quantile (q50; i.e., the median value), the model would be run using the site with the nearest corresponding *GPP* value from the 100 exploratory sites described in section 2.4. In this way, the variability of *GPP* as identified in the exploratory simulations can be sampled directly by simply varying the location of the recharge site
 320 within the domain. The Morris approach was implemented with open-source library developed by Herman and Usher (2017).

3 Results and Discussion

3.1 Exploratory Simulations

Results from the exploratory simulations at the 100 selected sites show a wide range of R_{30d} , P_{30d} , and $V_{\text{fines}, 90d}$ across sites (Fig. 4). R_{30d} varied over 2 orders of magnitude and were non-normally distributed, with a maximum, minimum, and mean

325 value of 66.4 cm d^{-1} , 0.5 cm d^{-1} , and 8.6 cm d^{-1} . P_{30d} were similarly non-normally distributed and also showed a large range, varying over 4 orders of magnitude. Maximum, minimum, and mean P_{30d} were $1.6 \times 10^5 \text{ m}^3$, 33 m^3 , and $1.9 \times 10^3 \text{ m}^3$, respectively. These results highlight that a small number of sites have outside recharge potential compared with most of the landscape. R_{30d} and P_{30d} were positively correlated ($r > 0.70$), indicating that these recharge benefits are physically related. The proportion of recharge accommodated by fine-texture facies ($V_{\text{fines}, 90d}$) also showed large variability across sites, ranging
330 from 0.13 to 1.00, with a mean value of 0.69. The high proportion of $V_{\text{fines}, 90d}$ observed here is consistent with previous findings that suggest that fine-texture facies are the largest reservoir for MAR in this aquifer system (Maples et al., 2019). $V_{\text{fines}, 90d}$ was negatively correlated with both R_{30d} and P_{30d} ($r < -0.70$), which indicates that when interconnected, coarse-texture pathways are present, a greater proportion of MAR is accommodated in the coarse-texture aquifer system.

3.1.1 Influence of Coarse-Texture Connectivity

335 Of the 100 exploratory sites, 23 were shown to have interconnected coarse-texture gravel and sand facies from land surface to the initial water table depth. R_{30d} , P_{30d} , and $V_{\text{fines}, 90d}$ were parsed according to whether they were interconnected (Fig. 4). Results show that mean R_{30d} and P_{30d} were 2.2 \times and 2.3 \times greater, respectively, for interconnected sites than for non-interconnected sites (14.7 cm d^{-1} vs. 6.7 cm d^{-1} , and $1.6 \times 10^4 \text{ m}^3$ vs. $6.9 \times 10^3 \text{ m}^3$, respectively). Mean $V_{\text{fines}, 90d}$ were 1.3 \times greater for non-interconnected sites than for interconnected sites. Distributions of R_{30d} and P_{30d} for interconnected
340 and non-interconnected sites differed significantly according to the two-sample Kolmogorov-Smirnov test. Interconnected and non-interconnected distributions of $V_{\text{fines}, 90d}$ were not significantly different.

These results indicate that sites with interconnected coarse-texture facies have greater R_{30d} and P_{30d} potential. However, this metric is not entirely diagnostic of recharge potential. As shown in Fig. 4a,b, some interconnected sites exhibited low R_{30d} and P_{30d} . This is likely because some interconnected sites with shallow water table depths have limited unsaturated
345 pore volume to accommodate large recharge volumes. In addition, the interconnection metric described herein only describes vertical interconnection of coarse-texture facies for unsaturated-zone cells that are vertically coincident with the recharge site, and does not consider whether these coarse facies connect with the greater aquifer network outside of the unsaturated-zone control volume. Results also show that some seemingly disconnected sites have large recharge potential. Indeed, the interconnection metric described here does not account for any lateral interconnection from land surface to the greater aquifer
350 network, which could explain this behavior. In reality, the simplified estimator of connectivity used here likely underestimates the number of interconnected sites.

3.2 Recharge Metrics

3.2.1 Correlation Matrices

A matrix of correlations (absolute value of Pearson's r) of pairs of site characteristics and simulated outputs for the 100
355 exploratory simulations was generated to better understand the relationships between variables (Fig. 5). Simulated outputs R_{10d} , R_{30d} , and P_{30d} were all positively correlated with all simulated outputs, while $V_{\text{fines}, 90d}$ was generally negatively correlated with

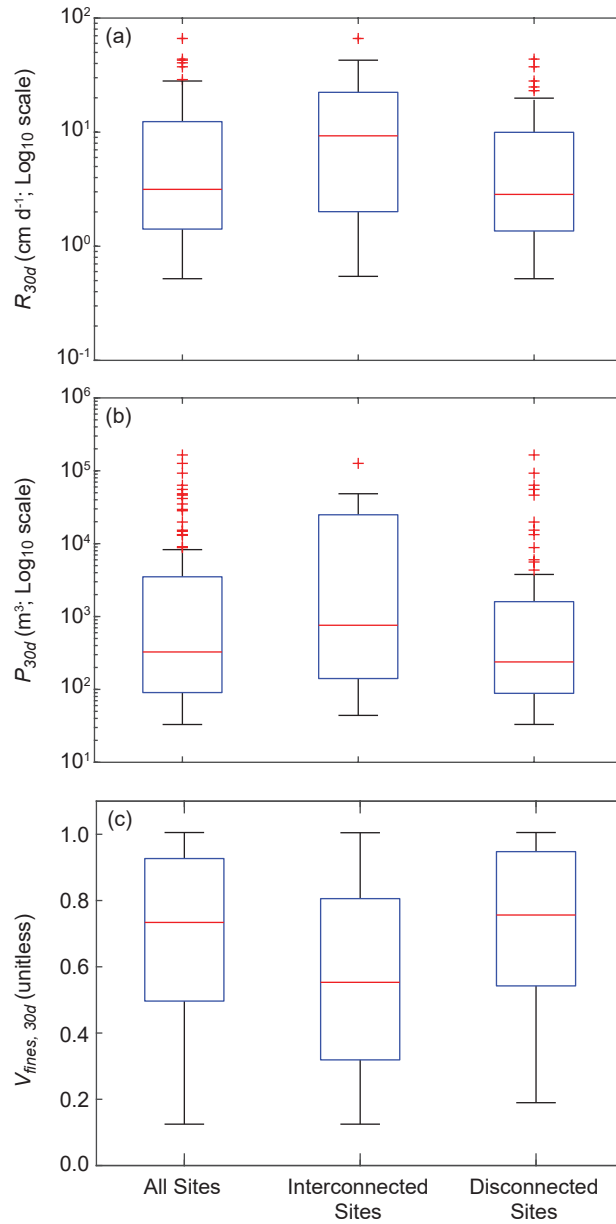


Figure 4. Box plots of the (a) 30-day average recharge rate, R_{30d} and (b) 30-day pressure perturbation area of influence, P_{30d} for all exploratory simulations ($n = 100$). Additionally, sites were parsed according to whether there was vertical interconnection of coarse-texture facies from land surface to the initial water table depth (i.e., interconnected sites, $n = 23$), or whether sites did not have interconnection of coarse-texture facies (i.e., disconnected sites, $n = 77$).

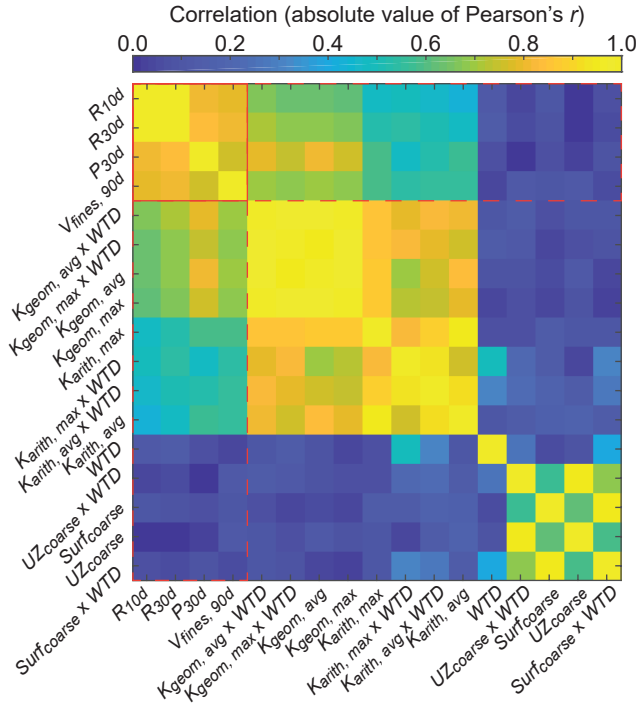


Figure 5. Correlations ([absolute value of Pearson's \$r\$](#)) for all combinations of site characteristics and model outputs. Correlations among site characteristics are bounded by a solid red box, and correlations between site characteristics and model outputs are bounded by a dashed red box.

[simulated outputs](#). Strong correlation ($r > 0.70$ | $|r| > 0.70$) was observed for 6 of 52 pairs of site characteristics and simulated outputs. Strong correlation was also observed among many site characteristics and among the majority of simulated outputs (i.e, collinearity), which can make the choice of an optimal proxy parameter more challenging. Site characteristics that include K_{harm} were not shown in the correlation matrix because we were not able to improve normality of the distribution these data with a Log_{10} data transformation; however, additional correlation metrics (Fig. 6) indicate that site characteristics that include K_{harm} may also be strongly correlated.

3.2.2 Ranked Correlations

Additional correlation metrics (Pearson's r , Spearman's rho, and Kendall's tau) between R_{30d} and site characteristics were ranked and are shown in Fig. 6. Results show that site characteristics that include some variation of K_{arith} , K_{geom} , or K_{harm} were, in general, more correlated with R_{30d} than site characteristics that only include WTD , UZ_{coarse} , and $Surf_{\text{coarse}}$. $K_{\text{geom}} \times WTD$ was, on average, most correlated with R_{30d} .

In general, site characteristics that included K_{geom} and K_{harm} were slightly more correlated with R_{30d} than site characteristics that included K_{arith} . We speculate that this behavior is related to the dominantly vertical flow direction of recharge across

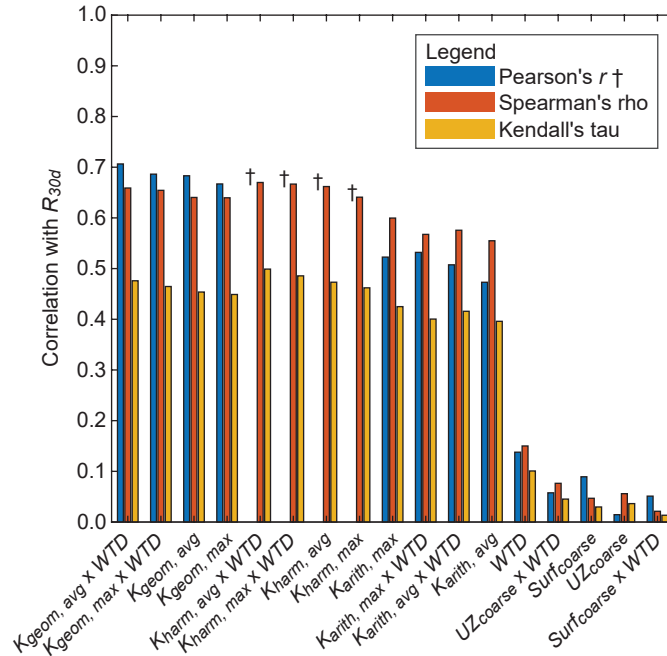


Figure 6. Ranked correlations (Pearson's r , Spearman's Rho, and Kendall's Tau) of site characteristics with 30-day average recharge rate (R_{30d}). †Pearson's r was not evaluated for site characteristics where the normality of the distribution could not be improved with a Log_{10} data transformation.

370 typically horizontal facies configurations. Previous work has shown that K_{geom} and K_{harm} best describe upscaled K for these flow configurations in this domain (Yunjie Liu, personal communication; Fogg, 1986).

Interestingly, site characteristics that included only WTD , UZ_{coarse} , and $Surf_{coarse}$ were poorly correlated ($r < 0.20$) with R_{30d} . This finding has important implications for determining MAR site suitability because many GIS-derived indices of recharge suitability rely solely on soil and/or surface geology to determine geologic suitability for recharge. These results suggest that even more detailed geologic descriptions that estimate deeper fractions of coarse-texture facies may not fully capture recharge potential. Instead, metrics that include some description of upscaled vertical K appear to be most diagnostic of recharge potential.

375

3.3 Recharge Extrapolation

The relation between site-averaged $K_{geom} \times DTW$ and R_{30d} was determined to be the best predictor and was used to predict R_{30d} for subsequent sensitivity analyses by treating $K_{geom} \times DTW$ as a GPP (Fig. 7a). The linear regression relation between $K_{geom} \times DTW$ and R_{30d} was highly significant ($p < 0.01$), and correlation coefficients (r^2) showed that empirical regression explained 70% of the variation in the data. Linear regression relations for $K_{geom} \times DTW$ and P_{30d} and $V_{fines, 90d}$ were deemed insufficient for prediction ($r^2 < 0.40$) and were not incorporated in sensitivity analyses.

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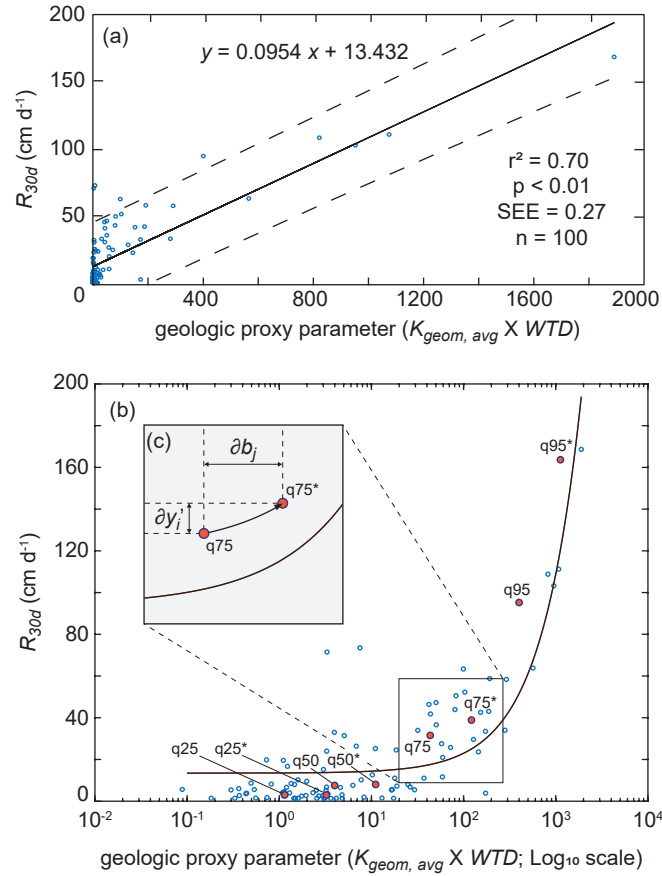


Figure 7. (a) Relation-Regression relation between the geologic proxy parameter ($K_{geom} \times WTD$) and the 30-day average recharge rate (R_{30d}) for all exploratory simulations shown with a solid black line, where dashed lines indicate the upper and lower 95% confidence intervals. (b) The relation is shown with $K_{geom} \times WTD$ on a Log_{10} scale, where red circles indicate the original and perturbed sites at which dimensionless scaled sensitivity (DSS) was estimated. (c) The inset illustrates the procedure for estimating the perturbed site (e.g., $q75^*$) from the original site (e.g., $q75$) for DSS , using the regression relation, where $\partial y'_i$ is the change in $K_{geom} \times WTD$ and ∂b_j is the estimated corresponding change in R_{30d} .

Domain-wide $K_{geom} \times WTD$ was converted to R_{30d} using the predictive relation described above (Fig. 8). Results show that
 385 84% of the domain has R_{30d} potential $<10 \text{ cm d}^{-1}$, while 6% of the domain has R_{30d} potential $>25 \text{ cm d}^{-1}$, and a small portion of
 the domain has R_{30d} potential $>150 \text{ cm d}^{-1}$. These results show a large contrast between locations with high recharge potential
 and those with low recharge potential which supports previous findings indicating that a small fraction of the landscape has
 recharge potential that is orders-of-magnitude greater than the rest of the landscape (Maples et al., 2019; Fleckenstein et al.,
 2006). Deposition of IVF within the domain area has been documented by Meirovitz (2010) and explains the presence of these
 390 high recharge potential locations.

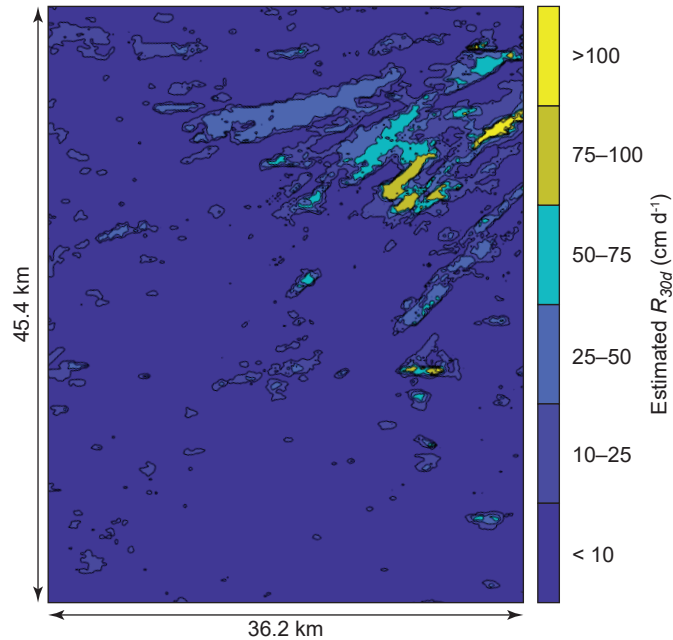


Figure 8. Domain-wide estimated distribution of 30-day average recharge rate R_{30d} .

3.4 Sensitivity Analyses

3.4.1 Local Sensitivity Analyses

GPP perturbations to estimate DSS for sites q25, q50, q75, and q95 using predictive regression relations are illustrated in Fig. 7b,c. DSS and CSS results for each model parameter and GPP with respect to R_{30d} is shown in Fig. 9. DSS results (Fig. 9a) show that for low recharge potential sites q25 and q50, K_s of mud and muddy sand facies were the most sensitive parameters with respect to R_{30d} . For high recharge potential sites q75 and q95, GPP was the most sensitive parameter. These findings demonstrate that K_s of fine-texture facies is the dominant driver of recharge potential for low recharge potential sites and the configuration of facies and water table depth is relatively less important. However, for high recharge potential sites, which presumably have a higher proportion of coarse-texture facies, the configuration of facies and water depth becomes the dominant driver of recharge potential. In general, DSS of all parameters were greater for sites with higher recharge potential than for sites with low recharge potential.

CSS results for R_{30d} (Fig 9b) show that, in general, GPP was the most sensitive parameter for R_{30d} when aggregated across all 4 sites. In general, K_s and ϕ were also sensitive with respect to R_{30d} . It is unsurprising that ϕ is sensitive to R_{30d} because specific yield (S_y), which is not explicitly parameterized in ParFlow, is closely related to ϕ . Moreover, Maples et al. (2019) showed that the majority of recharge volume in this alluvial system is accommodated by filling unsaturated-zone pore volume, which is controlled primarily by S_y , and by association in this model, by ϕ . Empirical fitting parameters describing

unsaturated-zone texture and soil water retention, α , n , and S_{res} , were relatively insensitive, especially for sites q75 and q95. This suggests that while unsaturated pore volume is important for recharge, the unsaturated flow processes are not particularly important, at least when considering infiltration of ponded water, which typically allows for rapid wetting-front advancement through the unsaturated zone, especially for high recharge potential sites. Results suggest that saturated storage properties, i.e., S_s , were also relatively unimportant. This likely because most recharge volume is accommodated by filling unsaturated pore volume, and is thus more dependent on ϕ (and S_y) than on S_s .

Normalized DSS for sites q25, q50, q75, and q95, and normalized CSS for all sites are shown in Fig. 10. DSS were scaled to the range [0,1] (i.e., normalized) for each group of parameters for a given site and a given model output. For example, all DSS values at q25 for R_{30d} were normalized to the maximum value of DSS for that group of parameters. CSS were similarly scaled for each group of parameters for given model output. Because DSS and CSS values are influenced by the units of each model output, normalization allows for comparison of their relative magnitudes between model outputs. Results show similar sensitivity importance for each model output, wherein K_s and ϕ are generally the most sensitive parameters, while S_s , α , n , and S_{res} are all relatively unimportant, except for n of q95 for $V_{fines, 90d}$, which was shown to be important. DSS and CSS of GPP were not evaluated for P_{30d} and $V_{fines, 90d}$ because regression relations between site characteristics and these outputs were generally poor compared to those for R_{30d} , as noted in section 3.3.

DSS and CSS results for GPP demonstrate the novel usage of empirical regression relations in a local sensitivity analysis framework. By perturbing GPP in this way, constancy of other parameters can be maintained in a way that would be otherwise difficult if GPP was perturbed by changing the location of the recharge site. Performing local sensitivity analyses at multiple sites spanning a range of recharge potential allowed for comparison of DSS sensitivities across sites and highlights differences of parameter sensitivities for low- and high-recharge potential sites. Our findings demonstrate that (1) facies permeability and unsaturated-zone storage properties are important factors for recharge potential, and (2) the configuration of subsurface geology and water table depth is particularly important for the total recharge volume that can be accommodated at a particular site, especially for high recharge potential sites.

3.4.2 Global Sensitivity Analyses

Results from global sensitivity analyses are shown in Fig. 11. Morris μ^* values indicate that GPP is the most sensitive parameter when compared with K_{sat} of coarse- and fine-texture facies. These results are consistent with findings from local sensitivity analyses which also showed that GPP was the most important parameter with respect to R_{30d} . Unlike DSS and CSS results, which compared GPP against model parameters for each facies, Morris analysis combined K_s parameters for coarse- and fine-texture facies which, in turn, increased the influence of those parameters on R_{30d} relative to GPP . Even so, results indicate that GPP is the most important parameter with respect to R_{30d} . These results further highlight the importance of the configuration of subsurface geology and water table depth for groundwater recharge potential.

Morris results demonstrate a novel incorporation of GPP within a global sensitivity analysis framework, and was unique as compared to incorporation of GPP in local sensitivity analyses described in section 3.4.1. Unlike the local methods, which used an empirical relation to incorporate an estimate of GPP sensitivity, the method used for the Morris approach directly

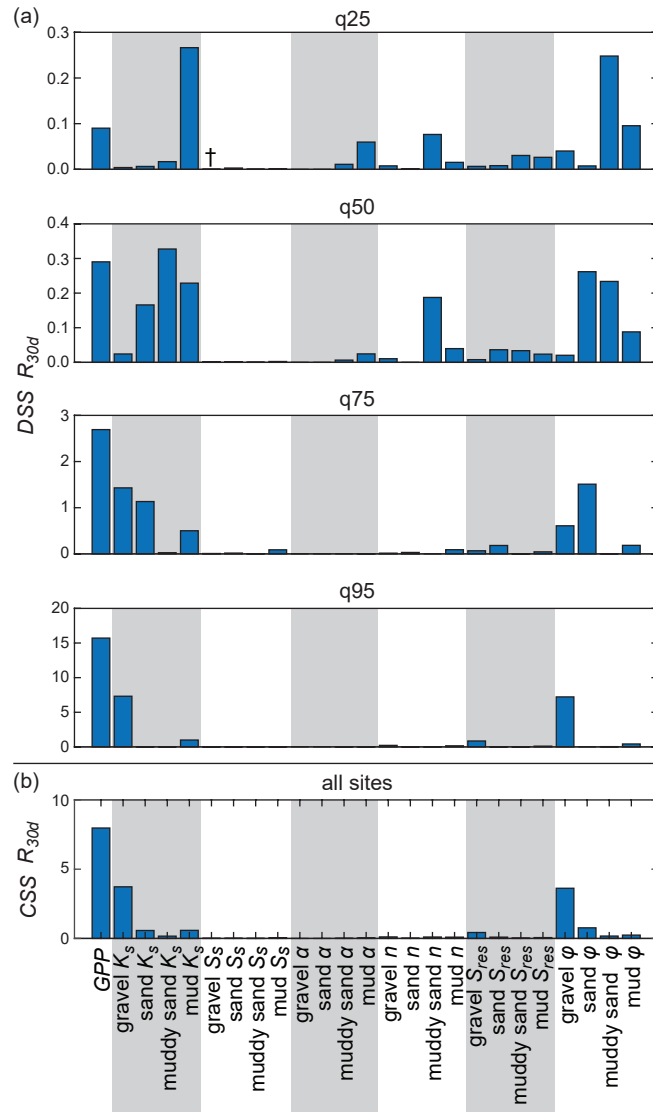


Figure 9. (a) Dimensionless scaled sensitivities (DSS) evaluated for each model parameter and model output R_{30d} at sites q25, q50, q75, and q95, and (b) composite scaled sensitivities (CSS) evaluated for each parameter and model output at all sites. DSS and CSS of parameters were scaled to the range [0,1] (i.e., normalized). † DSS and CSS values below 0.001 are not shown.

varied GPP within the parameter space by moving the recharge site to the location with the requisite GPP parameter value. Unlike the local approaches, which required constancy among all other parameters as each parameter is perturbed and thus required usage of an empirical relation to perturb GPP , the Morris approach varies all parameters globally, which allowed for GPP to be included explicitly within the approach. Consistency of the results of the local and global approaches despite these methodological differences for incorporating GPP highlights the robustness of these findings.

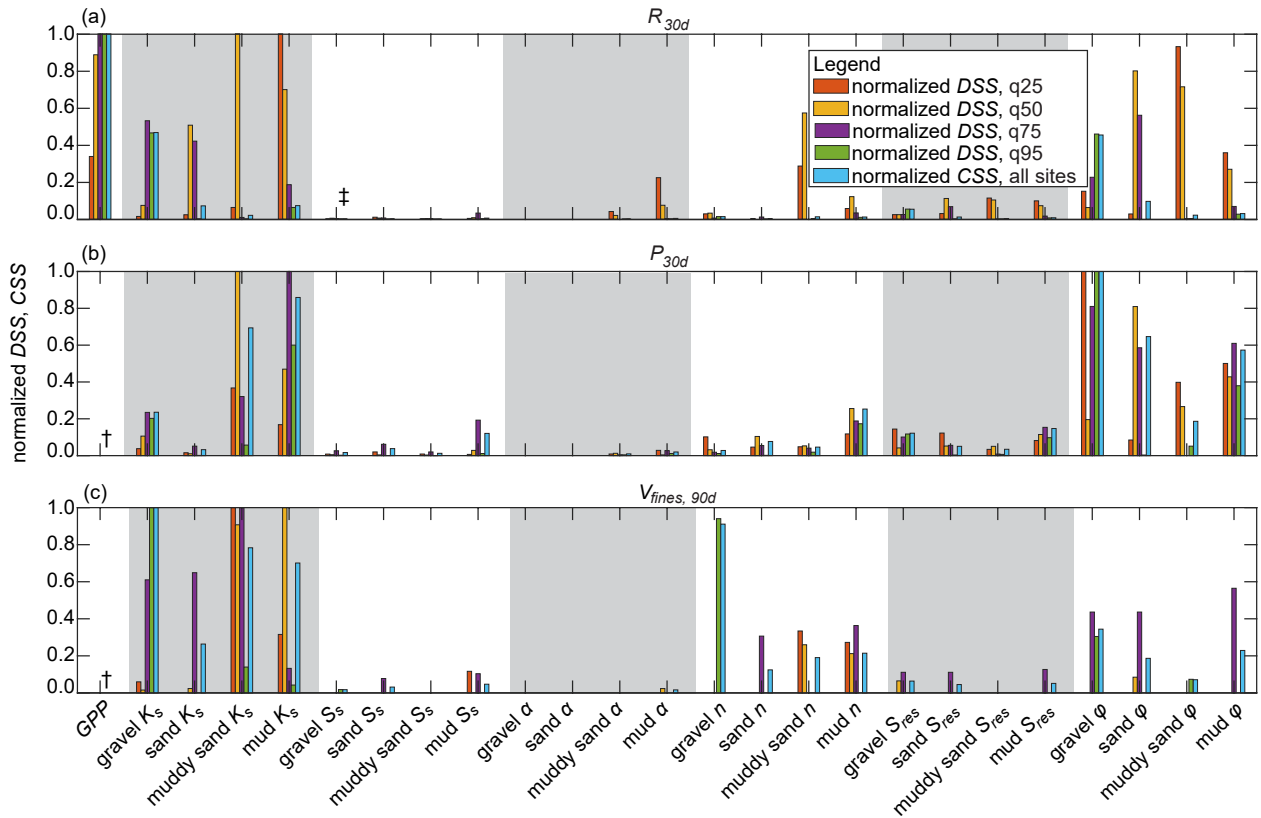


Figure 10. Normalized dimensionless scaled sensitivities (DSS) evaluated for each model parameter and model outputs (a) R_{30d} , (b) P_{30d} , and (c) $V_{fines, 90d}$ at sites q25, q50, q75, and q95, and normalized composite scaled sensitivities (CSS) evaluated for each parameter and model output at all sites. DSS and CSS of parameters were scaled to the range [0,1] (i.e., normalized). † DSS and CSS of GPP were not evaluated for P_{30d} and $V_{fines, 30d}$. ‡ DSS and CSS values below 0.01 are not shown. (double-column width)

4 Discussion

Sensitivity analyses are a fundamental diagnostic tool to provide insight into the relative importance of the parameterization of aquifer properties among other inputs in complex hydrologic models (Saltelli et al., 2004). Sensitivity analyses can be broadly categorized as local or global methods, where local methods provide sensitivity evaluation at a single location in the parameter space (Hill and Tiedeman, 2007), while global approaches explore sensitivities throughout a multi-dimensional parameter space (Saltelli et al., 2008). Many studies have shown the diagnostic utility of local approaches (e.g., Foglia et al., 2009); however, local approaches are generally less robust than global approaches, especially for non-linear models (Saltelli et al., 2008). On the other hand, global methods are typically orders of magnitude more computationally expensive than local approaches. Many studies have evaluated sensitivity of diffuse recharge in hydrologic and landscape models (e.g., Hartmann et al., 2017; McCallum et al., 2010). Other studies have evaluated sensitivities related to subsurface heterogeneity and per-

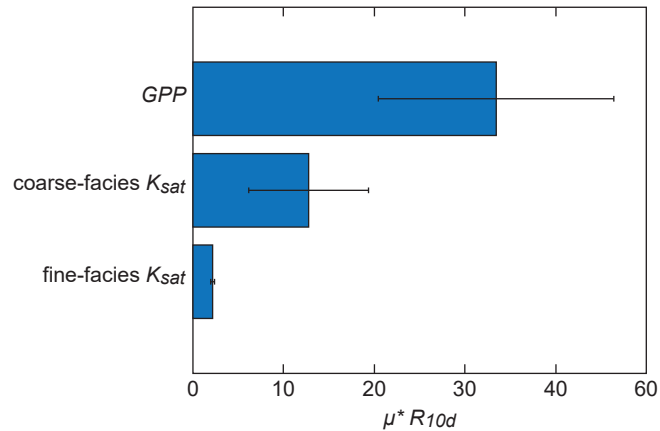


Figure 11. Morris μ^* value of R_{10d} for GPP and K_{sat} of coarse-texture (gravel and sand) and fine-texture (muddy sand and mud) facies, where bars represent each μ^* estimate, and whiskers represent the respective 95% confidence interval of the estimate.

meability upscaling in variably-saturated flow models (e.g., Gilbert et al., 2016; Foster and Maxwell, 2019; Srivastava et al., 2014) and on MAR specifically (e.g., Rahman et al., 2013; Heilweil et al., 2015), but to our knowledge, this is the first study to use a 3-dimensional variably-saturated water flow code with a detailed representation of geologic heterogeneity to evaluate the importance of hydraulic properties and geologic configuration on MAR dynamics with a combination of local and global sensitivity analyses.

Our approach includes novel incorporation of geologic architecture as a geologic proxy parameter (i.e., GPP). Of the many approaches to develop a GPP of recharge potential from descriptions of subsurface geologic and hydrologic characteristics, our results show that a GPP which combines metrics related to upscaled vertical K_s and unsaturated zone thickness was most diagnostic of recharge potential. In addition, results from local and global sensitivity analyses indicate that this GPP is equally or more important as characterizing the hydraulic properties of any particular facies for recharge potential. Consistency among results for both local and global approaches shows that these findings are reasonably robust and highlights the importance of accurately characterizing the subsurface configuration of coarse-texture facies in clastic sedimentary aquifer systems. While a GPP was shown to be the most important parameter with both approaches, we also show that parameters related to unsaturated-zone storage and facies permeability (i.e., ϕ and K_s , respectively) were also important for MAR. In contrast, we show that parameters related to unsaturated-zone geologic texture and soil water retention, along with saturated-zone storage properties (i.e., α , n , S_{res} , and S_s) were relatively unimportant. We speculate that these parameters are relatively unimportant because our simulations typically showed that surface ponding initiated rapid downward wetting-front advancement through the unsaturated zone, quickly developing fully saturated conditions from land surface to the water table. In systems dominated by diffuse recharge, these parameters may be more sensitive.

Findings presented here for a semi-confined alluvial aquifer system show large spatial variability of recharge rates that are dependent primarily on subsurface geologic configuration. we show that select locations in the domain area are capable of

accommodating orders-of-magnitude greater recharge benefit than would be possible over the rest of the landscape. These findings are consistent with previous studies that indicate that favorable site characteristics, including connect networks of coarse-texture IVF, are present in the American-Cosumnes River area of the Central Valley, California (Meirovitz, 2010; 480 Maples et al., 2019), but likely occur over a small fraction of the domain area. Other studies have shown that IVF deposits occur elsewhere in California's Central Valley (e.g., Weissmann et al., 2005) and in other major river fans that drain high-elevation, glacially-influenced catchments (e.g., Pierce and Scott, 1983). Identifying sites that can accommodate large MAR volumes during short windows is especially valuable in places like California, where excess surface water available for recharge typically occurs from a few precipitation events (Dettinger et al., 2011) during short (< 10 day) windows (Kocis and Dahlke, 485 2017).

Our results show that cursory investigations of soil or surficial geology are likely insufficient to adequately characterize MAR favorability. Instead, our findings indicate that more thorough investigations of subsurface geologic architecture and aquifer configuration are needed to accurately characterize MAR feasibility. We found that metrics that consider the geologic configuration of facies and provide some measure of upscaled vertical K_s are the best predictors of recharge feasibility. We 490 show that connectivity metrics that determine whether coarse-texture facies interconnect from land surface to the saturated zone are also helpful, but not fully diagnostic of recharge potential. Interestingly, our results show that metrics describing unsaturated-zone thickness, fraction of coarse-texture facies at land surface, and fraction of coarse-texture unsaturated-zone facies are insufficient when each is considered alone. This finding has important implications because several GIS-derived metrics of recharge potential describing recharge suitability of surficial soils have been developed for California and elsewhere 495 (O'Geen et al., 2015; Adham et al., 2010; Ghayoumian et al., 2007). We consider these products as valuable, albeit incomplete metrics that are likely complemented by more detailed investigations of deeper subsurface geologic architecture.

Existing MAR siting efforts in California have mostly relied on surficial mapping rather than characterization of deeper three-dimensional subsurface geologic architecture because these complex subsurface investigations are typically prohibitively labor- and cost-intensive. In the past, these subsurface investigations relied primarily on gathering and interpreting geologic texture from typically sparse well-log data. However, emerging geophysical approaches using airborne time-domain electromagnetic methods (AEM; Knight et al., 2018) and ground-based transient towed electromagnetic methods (tTEM; Behroozmand et al., 2019; Auken 500 are promising new approaches to directly image subsurface geologic texture. Pilot studies have shown the efficacy of these approaches for mapping geologic texture up to 500 m in depth to efficiently identify regional aquifer configuration and potential MAR locations in the Central Valley (Knight et al., 2018) and mapping favorable MAR locations in high-resolution at the field scale (Behroozmand et al., 2019). The authors present this work, in part, to motivate the scientific community to develop and adopt new, cost-effective approaches for identifying favorable geology for MAR.

Importantly, no single *GPP* described herein was a fully diagnostic metric of recharge potential at all sites. This result is not surprising given the complexity of geologic architecture and variability of aquifer configuration sampled across sites in the domain, which are challenging to fully ~~captured~~-capture with a single metric. For example, all site characteristics described 510 here were developed only for those model cells that are vertically-coincident with each site footprint, and do not account for possible preferential pathways in adjacent cells outside of the immediate site footprint. We acknowledge that further research

into this phenomena could provide additional insight into developing site-specific *GPP*, but is outside the immediate scope of this work. We also acknowledge some limitations of our sensitivity analyses. For example, reliance on imperfect empirical regression relations to include measures of geologic configuration in local methods likely introduced uncertainty to *DSS* and *CSS* estimates for this parameter. In addition, inclusion of all model parameters describing facies hydraulic properties in the Morris approach would have been valuable, but was infeasible given the computational resources for the simulations required. In addition, parameter-range uncertainty contributes some uncertainty to rankings of parameter importance.

Our simulations also do not consider some subsurface geologic conditions that influence MAR. Clastic sedimentary aquifer systems are typically replenished naturally over longer timescales (Taylor et al., 2013) because even productive aquifer systems are commonly composed mostly of fine-texture sediments (e.g., Fogg, 1986; Fogg et al., 2000), that form nearly ubiquitous, multiple confining layers that inhibit direct recharge of the interconnected sand and gravel body networks that comprise the aquifer system. The presence of laterally-continuous aquitard facies have been well documented portions of the southern Central Valley (Phillips and Belitz, 1991; Faunt et al., 2009), and in other unconsolidated alluvial aquifer systems in California (e.g., Fisher, 1964). While not present within the domain area, these features have been shown to uniformly impede recharge to confined aquifer systems where they are present. In addition, we do not consider some surface conditions that affect real-world MAR, like topographic site limitations, evaporative losses, and clogging effects (Bouwer, 2002). We emphasize that this study is not a thorough site investigation of the American-Cosumnes area. The TPROGS approach is inherently stochastic and conditioning data to inform the model are sparse in places (Maples et al., 2019). In addition, the single TPROGS realization used for our simulations provides only a single representation of possible facies distributions within the domain. Our findings are presented as a proof-of-concept to explore the importance of geologic heterogeneity on MAR in a hypothetical but physically-realistic domain. Finally, we acknowledge some potential limitations when applying the differencing approach used to isolate recharge stresses from other model stimuli for the non-linear model presented here. The potential for errors with this approach for non-linear models have been noted in several studies (e.g., Reilly et al., 1984; Nadler et al., 2018). However, spurious recharge stresses or unrealistic model noise were not encountered when isolating recharge stresses with the differencing approach in this study.

Our findings have important implications for assessing MAR feasibility and for understanding MAR processes in clastic alluvial aquifer systems in California and globally, where accelerating groundwater overdraft and increasing water scarcity are observed (Scanlon et al., 2012; Famiglietti et al., 2011; Wada et al., 2011). Our results highlight the importance of identifying and cataloging locations with favorable geology for recharge, especially in light of recently-passed groundwater management legislation in California that mandates limiting both the "chronic lowering of groundwater levels" and "significant and unreasonable reductions in groundwater storage" (Kiparsky et al., 2016). While studies have shown that implementation of MAR can lead to more sustainable groundwater management (e.g., Niswonger et al., 2017), widespread adoption of of MAR is still hampered by a number of challenges, including institutional barriers to water-rights transference and water accounting uncertainty (Asano, 2016), infrastructure limitations, including land acquisition and water conveyance costs (Gailey, 2018), and water quality considerations (Hartog and Stuyfzand, 2017). Our approach, which combines a detailed representation of subsurface geology with physically-realistic water flow physics in a sensitivity analysis framework, can (1) help guide site

investigations and data collection methods for proposed MAR projects, and (2) improve representation of recharge processes in management-focused, typically coarse-resolution groundwater models.

5 Conclusions

550 This research explores the sensitivity of hydraulic properties and subsurface geologic architecture on MAR processes with the variably-saturated water flow code, ParFlow, in a highly heterogeneous geologic domain that reflects the complex, unconsolidated alluvial geologic architecture of the northern Central Valley, CA that is consistent with many alluvial aquifer systems. This work comprises two fundamental components. First, exploratory simulations were performed at 100 randomly-sampled sites across the domain to evaluate the correlation between 17 geologic and hydrologic site characteristics and simulated
555 recharge benefits. Results from the exploratory simulations show that site characteristics representing subsurface geologic configuration by upscaling vertical K can produce good correlations with the average 30-day recharge rate (R_{30d}). Regression relations between site-averaged $K_{geom} \times WTD$ and R_{30d} were shown to be the most correlated ($r = 0.70$, $p < 0.01$, $r^2 = 0.70$). Conversely, site characteristics describing unsaturated-zone thickness (WTD), fraction of coarse-texture unsaturated-zone (UZ_{coarse}), and fraction of coarse-texture surface facies ($Surf_{coarse}$) alone were all poorly correlated with R_{30d} . These results highlight the value of characterizing subsurface geologic configuration through K upscaling. For subsequent sensitivity
560 analyses, $K_{geom} \times WTD$ was designated as a geologic proxy parameter, GPP , for recharge potential using aforementioned predictive regression relation.

Subsequent local and global sensitivity analyses were performed for model hydraulic properties and the GPP to evaluate the relative importance of these parameters on recharge potential. Results from local sensitivity analyses indicated that GPP is
565 the most sensitive parameter for R_{30d} , more so than any parameters describing hydraulic properties of each facies. Sensitivity analyses also indicated that permeability and unsaturated-zone pore volume (i.e., K_s and ϕ , respectively) were relatively more important than other hydraulic properties, including unsaturated-zone geologic texture, soil water retention, and saturated-zone storage properties (i.e., α , n , S_{res} , and S_s) for R_{30d} . Results from global sensitivity analyses were consistent with local sensitivity analyses, indicating that GPP is relatively more important than K_s of coarse- and fine-texture facies for R_{30d} . Agreement
570 of local and global approaches regarding the importance of GPP shows a degree of robustness of these findings. The results presented here demonstrate the importance of thoroughly characterizing subsurface geologic configuration when considering recharge feasibility. To our knowledge, this study is the first of its kind to incorporate ~~of~~ a measure of geologic configuration with a geologic proxy parameter in formal sensitivity analyses. Our approach outlines a novel combination of subsurface site characterization with simulations of variably-saturated water flow physics within a sensitivity analysis framework to (1) im-
575 prove understanding the role of geologic heterogeneity on MAR processes and (2) provide insight into potential strategies to characterize subsurface geologic heterogeneity when considering recharge feasibility.

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Table 1. Textural classification of hydrofacies designations (Fleckenstein et al., 2004)

Hydrofacies	Geologic Interpretation	Texture
Gravel	Channel deposits	Gravel and coarse sand
Sand	Near channel/levee	Sand (fine to coarse)
Muddy Sand	Proximal floodplain	Silty and clayey sand, sandy clay, and silt
Mud	Floodplain	Clay, silty clay, shale

Table 2. [Calibrated](#) Hydrofacies Hydraulic Properties (Liu, 2014)

Facies Designation	K_s (m d ⁻¹)	S_s (m ⁻¹)	ϕ (unitless)	α (unitless)	n (unitless)	S_{res} (unitless)
Gravel	67.5	4.00E-05	0.35	3.55	3.16	0.1
Sand	41.2	8.00E-05	0.35	3.55	3.16	0.1
Muddy Sand	0.2	1.00E-04	0.40	2.69	2.00	0.1
Mud	0.0017	1.00E-03	0.45	1.62	2.00	0.2
Deep Aquifer	45.0	4.80E-04	0.35	3.55	3.16	0.1

Table 3. Descriptions of site characteristics (variables) and model outputs (predictors)

Observation type	Name	units	Description
site characteristics (<u>variables</u> <u>predictors</u>)	$K_{arith, max}$	$m d^{-1}$	Maximum arithmetic mean of vertical K_s for cells above the initial water table for each site
	$K_{geom, max}$	$m d^{-1}$	Maximum geometric mean of vertical K_s for cells above the initial water table for each site
	$K_{harm, max}$	$m d^{-1}$	Maximum harmonic mean of vertical K_s for cells above the initial water table for each site
	$K_{arith, avg}$	$m d^{-1}$	Average arithmetic mean of vertical K_s for cells above the initial water table for each site
	$K_{geom, avg}$	$m d^{-1}$	Average geometric mean of vertical K_s for cells above the initial water table for each site
	$K_{harm, avg}$	$m d^{-1}$	Average harmonic mean of vertical K_s for cells above the initial water table for each site
	UZ_{coarse}	unitless	Proportion of cells above the initial water table that are coarse-texture facies (gravel and sand) for each site
	$Surf_{coarse}$	unitless	Proportion of surface cells that are coarse-texture facies (gravel and sand) for each site
	WTD	m	Average initial water table depth for each site
model outputs (<u>predictors</u> <u>variables</u>)	R_{10d}	$cm d^{-1}$	Effective recharge rate (0–10 day average)
	R_{30d}	$cm d^{-1}$	Effective recharge rate (0–30 day average)
	P_{30d}	m^3	Volumetric extent of pressure perturbation after 30 days (10-cm threshold)
	$V_{90d, fines}$	unitless	Proportion of total recharge volume accommodated by fine-texture facies after 90 days

Table 4. Parameter ranges, along with baseline and *DSS* perturbed parameter values for hydraulic properties and the geologic proxy parameter (*GPP*)

Parameter	Facies Designation	Parameter Range [†]	Baseline Parameter [‡]	<i>DSS</i> Perturbed Parameter
K_s (m d ⁻¹)	Gravel	14.42 – 144.2	67.5	80.48
	Sand	5.4 – 54.0	41.2	46.06
	Muddy Sand	0.089 – 0.89	0.20	0.28
	Mud	0.0023 – 0.023	0.0017	0.0038
S_s (m ⁻¹)	Gravel	1.0×10^{-6} – 1.0×10^{-4}	4.0×10^{-5}	4.9×10^{-5}
	Sand	1.0×10^{-6} – 1.0×10^{-4}	8.0×10^{-5}	8.9×10^{-5}
	Muddy Sand	0.0001 – 0.001	0.0001	0.0002
	Mud	0.0001 – 0.001	0.0010	0.0011
α (unitless)	Gravel	3.55 – 3.55	3.55	3.55
	Sand	3.55 – 3.55	3.55	3.55
	Muddy Sand	2.69 – 3.55	2.69	2.78
	Mud	0.35 – 0.45	1.62	1.81
n (unitless)	Gravel	2.00 – 3.16	3.16	3.044
	Sand	1.89 – 3.16	3.16	3.033
	Muddy Sand	1.44 – 2.00	2.00	1.94
	Mud	1.32 – 2.00	2.00	1.93
S_{res} (unitless)	Gravel	0.10 – 0.14	0.10	0.104
	Sand	0.10 – 0.14	0.10	0.104
	Muddy Sand	0.10 – 0.25	0.10	0.12
	Mud	0.16 – 0.23	0.20	0.21
ϕ (unitless)	Gravel	0.25 – 0.35	0.35	0.34
	Sand	0.25 – 0.35	0.35	0.34
	Muddy Sand	0.35 – 0.45	0.40	0.39
	Mud	0.35 – 0.45	0.45	0.44
<i>GPP</i> (m ² d ⁻¹)	-	0.08 – 1891.6	-	-

[†] Ranges of hydraulic properties for each facies category were derived from literature values for the Central Valley, California, and for similar alluvial systems (Anderson et al., 2015; Botros et al., 2009; Fleckenstein et al., 2004; Frei et al., 2009; Maserjian, 1993; Niswonger and Fogg, 2008; Sager, 2012).

[‡] Baseline hydraulic properties were calibrated manually by Liu (2014)