

A conceptual model 1 of organochlorine fate from a combined analysis of spatial and mid/long-term trends of surface and ground water contamination in tropical areas (FWI)

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We thank S. Reichenberger and the second reviewer for their helpful comments provided in the review file hess-2018-377-RC1 and hess-2018-377-RC2. We detail in this file the response to each comment and the corresponding changes that we propose to improve the manuscript:

- each comment of the referee about the original version (OV) of the manuscript (indicated in black),
- our corresponding responses (indicated in blue)
- our proposals for amending the text (in blue and in bold-italic),

Response to the first reviewer

I think this paper is of good quality and contributes significantly to the understanding of the environmental fate of CLD on the island of Martinique. However, I have a few comments (see below) that should be addressed in order to make the paper easier to read and understand.

General comments:

1) The word “residence time” is somewhat vague. Please specify at first mention what is exactly meant (e.g. mean transit time).

Response: the groundwater residence time refers to the water age in aquifers defined as the mean transit time by Maloszewski and Zuber (1992). It would be addressed in the revised version of the manuscript.

Change in manuscript: L70 “Moreover, at depth, contrasting residence times of water (*the water age in aquifers is defined as the mean transit time (Maloszewski and Zuber, 1982)*) in aquifers, ranging from several years to several decades, partly account for the variability in GW contamination by CLD (Gourcy et al., 2009)”

2) You mention in l. 432 that the 5bCLD/CLD ratio in the commercial product Curlone was 0.0011. How was the ratio in the product Kepone that was applied before? Was the 5bCLD/CLD ratio in the commercial products constant over time, or did it vary between batches of the same product?

Response: according to Devault et al. (2016), Kepone and Curlone products had very similar ratio characterized by values of 0.00077. In our text, we specified a value from Clostre et al (2015). We propose to change to Devault et al. (2016) for consistency. Notice that Devault et al. also compared three different batches of Curlone without significant difference. Equally, we modified paragraph 3.1 according to your comment 18.

Change in manuscript: L430 “This was hypothesized by observing the distribution of 5bCLD / CLD ratios in water (median of 0.03; 1st centile of 0.006) with a far higher median and first centile value than in the commercial products *Kepon® and Curlone® used in FWI (mean ratio of 0.00077 ± 0.00027 (Devault et al., 2016))*”.

3) Can your findings / model be extrapolated to other CLD-contaminated areas in the Antilles, notably the island of Guadeloupe?

Response: indeed, our conceptual model could be generalized to Guadeloupe archipelago, where chlordecone was also applied intensively in banana plantations.

Change in manuscript : L441 end of § “ *These results on the island of Martinique could indeed be extrapolated to other CLD-contaminated areas, such as in the Guadeloupe archipelago (FWI) where CLD was also intensively applied in banana plantations.* ”

Specific comments:

Abstract:

4) p. 1, l. 28: “old geology”: I know what you mean, but maybe “old geological substrates” or “old geological formations” would be more appropriate?

Response: we agree with comment and “old geology” was replaced by “*old geological formations*” throughout the text in the revised version

5) p. 1, l.29: “theoretical leaching model”: maybe “conceptual leaching model” (as in the manuscript title)

Responses: We agree. Change OK

Introduction:

6) p. 2, l. 64-65: “acute” and “environmental” are not opposites; better use “chronic” instead of “environmental” (exposure via the environment can be acute or chronic)

Response: We agree. Change OK

7) p. 2, l. 69: “partitioning coefficient (Koc) between the sorbed part on soil organic matter”: not comprehensible ! needs to be rephrased

Response: OK. We propose to rephrase as follows, just giving the name of the coefficient:

Change in manuscript:L69 “*soil organic carbon-water partitioning coefficient (Koc)*”

8) p. 2, l. 71: “contrasting residence times”: What residence times: of water or of CLD?

Responses: it refers to residence time of water and it will be addressed in the revised version.

Change in manuscript L70 See our response to comment 1) where we propose to add the following sentence: “...contrasting residence times (*the water age in aquifers defined as the mean transit time (Małozewski and Zuber, 1982)*) in aquifers of...”

Materials and Methods:

9) p. 3, l. 111-112: “ferralitic soils (latosols) ! ferralsols”: What is the difference between the two? The WRB system contains the reference soil groups ferralsols and plinthosols, but not “ferralitic soil (latosols)”.

Response: It is a mistake. The climate sequence described in Comet-Daage et al. (1965) is latosols -> ferrisols -> vertisols. A correspondence with the WRB system is given by Delvaux and Brahy (2014) and is *ferralsol* -> *nitisol* -> *vertisol*. We corrected.

Change in manuscript : L111-112 “*ferralsol* -> *nitisol* -> *vertisol*”

10) p. 6, l. 213: “measurable”: maybe more precisely quantifiable”, since it refers to the LOQ

Response: We agree. Change OK

11) p. 6, l. 215: “data item” (or data point): How is this defined? Unique combination of water sample and compound?

Response and Change in manuscript: OK. We propose to specify in the text “data item (i.e. *water samples analyzed for CLD and 5bCLD*)”

12) p. 6, l. 230: eq. 1: explain the indices (i, j, k, l, m, t)

Response: There is a mistake in the text and in Equations 1 and 2. We corrected and specified in the text:

Change in manuscript: L 233 end of § “. *Indices i, j, t, k, l represent following factors soil x geology, hydrological sector, date, sampling point and sample replication, respectively*”

13) p. 6, l. 234: “totally correlated”: express more precisely

Response: because of a limited number of sites for groundwaters, there is not a so high spatial variability of geology and soil as observed for surface water data set. Thus, for groundwater data set (model 2), geology and soil are correlated: Andosol on recent geological formations, and ferralsols on old geological formations. To be clearer in the revised version, we modified the sentence as follows:

Change in manuscript: L234 “Soil and geological factors were *closely linked* for the GW data set (*andosols were always associated with recent geological formations and ferralsols with old geological formations*)...”

14) p. 6, l. 241: “dispersion indices”: How can this quantity be interpreted? and 15) p. 6, l. 242: “confidence coefficient”: What is this?

Response: dispersion indices can be interpreted like standard deviation. From the log values $\ln(x)$, we determined a confidence interval $[\ln_{inf}, \ln_{sup}]$ (with $\ln_{inf} = \ln(x) - sd$, $\ln_{sup} = \ln(x) + sd$, sd the standard deviation of log value). That is real values are included in the interval $[\exp(\ln_{inf}), \exp(\ln_{sup})]$. Because $\exp(sd)$ is not convenient, we defined a new dispersion index: $[\exp(\ln_{sup}) - \exp(\ln_{inf})]/2$, i.e. “half the difference between the limits of the confidence interval”.

Change in manuscript: In fact, since we no more used these dispersion indices in the revised version of the article, we propose to suppress L241 and 242.

16) p. 6, l. 248: “Sen trends”: What are Sen trends, and what do they mean statistically? (Explain in 1 or 2 sentences.)

Response: OK. We propose the following change

Change in manuscript: L248 “We calculated Sen trends (*Sen’s slope estimator, (Gilbert, 1987)* for each variable (CLD, 5bCLD and ratio) in order to compare dynamics for the two compounds. *The Sen trend of a set of two-dimensional points (x_i, y_i) and (x_j, y_j) is the median of the slopes $(y_j - y_i)/(x_j - x_i)$ determined by all pairs of sample points. The Sen trend is more robust than the least-squares estimator, because it is much less sensitive to outliers*”

17) p. 8, l. 258-262: What are the dimension and unit of the lixiviation rates TCLD and T5bCLD? This does not become entirely clear from eq.6 because of the various unit conversion factors. I end up with the unit 1/year.

Responses: Tcld and T5bclD are mass ratio. We propose to add a comment on that point in the text:

Change in manuscript: L258 “TCLD and T5bCLD are the rates of lixiviation for CLD and 5bCLD (*i.e. the ratio of lixiviated mass of CLD or 5bCLD to their respective mass in soil*)...”

Additionally we propose to modify Eq. (6) for consistency with regard to the dimensions

“Eq. (6):
$$T_{CLD} = (R \times S) / (K_{oc} \times (C/1000) \times D \times (d \times S))$$
 where K_{oc} ($L\ kg^{-1}$) is the partitioning coefficient between the sorbed part on soil organic matter and the dissolved part in water, D ($kg\ dm^{-3}$) the bulk density, C ($g\ kg^{-1}$) the soil carbon content, and R (dm) the annual amount of rainfall, S the soil surface area (dm^2) and d the soil depth (dm).”

Results:

18) p. 8, l. 279-281: It should be mentioned here how high the ratio 5bCLD/CLD was in the commercial products that were applied, and whether it was constant over time.

Response: We propose to add the following text at the end of the paragraph:

Change in manuscript: L281 end of § “*According to Devault et al. (Devault et al., 2016), these differences cannot stem from the use of different commercial products or different batches of the same product. Indeed, these authors, found no significant statistical difference between the ratio of the commercial products Kepone® and Curlone® used in FWI, no more than they did between samples from different batches of Curlone®. They found a mean ratio in commercial products of 0.00077 ± 0.00027 , i.e. ten times lower than our observations in river.*”

19) p. 9, l. 324-325: “shorter residence times were observed for more recent formations”: (Are you referring to GW only or also to SW?) This is interesting. I would rather have expected the opposite. Can you briefly explain why hydraulic residence times (mean transit times?) are

shorter in the younger geological formations of Martinique than in the older ones?

Response1: We refer to groundwaters (knowing that groundwaters are the main contributor of contamination of surface waters). We propose to add the following text in the 2.1 Section:

Change in manuscript: L134 after “old weathered formations”. “*Given that the weathering of geological formations increases with their age, it is the main cause of a global decrease in aquifer permeability, notably in volcanic regions (Lachassagne et al., 2014). Indeed, clayey alteration products by weathering constrain the soil’s physical and hydrodynamic properties by reducing porosity, and consequently permeability (Adelinet et al., 2008)*”

Response2: Thus, as we observe higher 5b/CLD ratio on younger geological formations (i.e. unweathered formations), we hypothesize that this was related to a shorter residence time. We propose to modify the sentence as follows: L324

Change in manuscript: L324 “Thereby, shorter residence times were observed for *aquifers located in* more recent *and unweathered geological* formations ”

20) p. 10, l. 344: “water CLD contents below the detection limit appeared less frequently”: meaning not entirely clear ! rephrase

Response: We propose the following text:

Change in manuscript: L343 “For the two sites showing a decrease in water CLD content, *the number of samples with* 5bCLD contents below the detection limit *decreased over time*, and *equalled zero* in the case of the Source Morne Figure site after 2011 ”

21) p. 10, l. 384: unit for bulk density is missing

Response: We corrected “*kg dm⁻³*”

22) p. 13, l. 457: “unweathered formations favour rapid transfers”; Why is that? (cf. comment on l. 324-325)

Response: please see our response to comment “19”

23) p. 13, p. 468: “we cannot assess it”: Assess what?

Response: the effect of soil on degradation process. We modified the text accordingly.

Conclusions:

24) p. 13, 507-508: “This led to implications regarding where and how to act to reduce impacts”: Can you elaborate on this further? Is there really anything that can be done except waiting for CLD to degrade and leach from the system?

Response: We propose some examples:

Change in manuscript: L508 “(e.g. *choice of crops according to pollution levels since some plants are less sensitive to contamination than others (Clostre et al., 2015), constraints on*

water management like drinking water and irrigation, choice of priority areas to test decontamination processes, setting up compensation plans according to the risk...)”

Figures:

25) Figure 3: The figure is neat, but too small for reading the legend or for identifying much on the map. ! upscale

Response: We propose a new version of Figure 3

Change in manuscript: change of Figure 3 and relative caption: “*Distribution of water CLD content (a, c, e) and 5bCLD / CLD ratio (b, d, f) for surface water (square) and groundwater (star), according to banana cultivated areas and hydrological sectors (a and b), soils (c and d) adapted from Colmet Daage (1965), and geology (e and f) adapted from Germa et al. (2011). Large squares are relative to sample points having more than ten sampling dates and small squares having fewer than ten sampling dates*”

26) Figure 4: y-axis: The numbers are difficult to interpret. Try lg or non-logarithmised numbers instead of ln.

Response: As specified in material and methods, data were log transformed for all analysis. For Figure 4 we worked with log-transformed data. We propose to complete the caption Figure 4 specifying the correspondence between log and non-log values:

Change in manuscript: “*The y values of -6, -4 and -2 correspond to ratio values of 0.002, 0.018 and 0.135, respectively.*”

27) Figure 4: give the unit of formation age (million years) directly in the figure (e.g. 1.0-0.3 Ma)

Response: This is done

28) Figure 6: too small, needs upscaling (if it takes too much space after upscaling, consider shifting it to a SI section).

Response: We propose a new Figure

29) Figure 7: If one doesn't know what Sen trends are, the figure is not understandable.

Response: This is now explain in the text accordingly to the response to your comment N°16

Tables:

30) Table 2: Table header needs to be rephrased to improve understandability.

Response: Caption and headers were modified in the revised version. We propose the following headers that matches terms in the caption and the text:

Change in manuscript: *Simulation , target value, fixed parameter*

Best regards,

Stefan Reichenberger

References:

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Lachassagne P., B. Aunay, N. Frissant, M. Guilbert, and A. Malard, 2014. High-resolution conceptual hydrogeological model of complex basaltic volcanic islands: a Mayotte, Comoros, case study. Terra Nova, 26, 307–321, doi: 10.1111/ter.12102.

Maloszewski, P., Zuber, A., 1982. Determining the turnover time of groundwater systems with the aid of environmental tracers: I. : Models and their applicability, J. Hydrol., 57, 207-231.

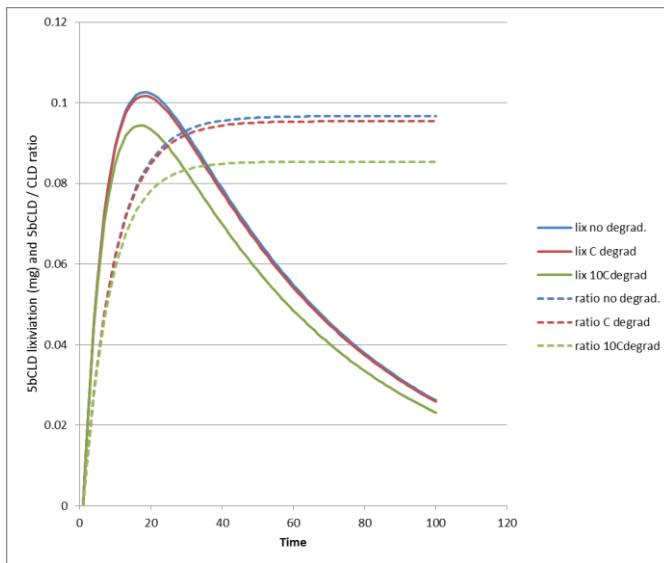
Response to the second reviewer

...key conceptual assumptions have to be thoroughly discussed and significantly improved to consider publication in HESS.

First, using a transformation product (TPs) 5bCLD as a tracer of degradation extent and associated persistence is a valuable idea, but the degradation of the TP itself is never discussed. Similarly the model seems to consider as a perfect tracer 5bCLD, i.e. without degradation. This main assumption can significantly alter the assessment of persistence done and this point is never discussed.

Response: We thank the reviewer for raising this important point. Degradation of the TP can be discussed adding a new calculation step in the model accounting for 5bCLD degradation. Eq (4) can be modified as follow:

$$5bCLD(t + 1) = 5bCLD(t) - 5bCLD(t) \times T_{5bCLD} - 5bCLD(t) \times C_{5bdegrad} + CLD(t) \times C_{degrad}$$



Then, it is possible to test 3 values of $C_{5bdegrad}$ in a wide range surrounding the one of CLD (C_{degrad}): $C_{5bdegrad} = 0$, $C_{5bdegrad} = C_{degrad}$, $C_{5bdegrad} = 10 \times C_{degrad}$. Notice these values are highly speculative since there is no experimental $C_{5bdegrad}$ value and that C_{degrad} is the result of optimization process in our paper. Results are reported in the Figure above which shows the evolution of the 5bCLD lixiviation and of the ratio for the 3 tested values. The figure shows similar dynamics of ratio evolution or of lixiviation evolution. The difference between the simulations remains weak, notably because the tested values are about 10 and 100 times lower than lixiviation rate (T_{5bCLD} equals 0.1242 here while C_{degrad} equals 0.0014). Consequently, introducing a degradation coefficient does not alter here our first conclusions. Running optimization process with this new term, we find $T_{5bCLD} = 0.1242$, $C_{degrad} = 0.0014$ and $C_{5bdegrad} = 0.0010$.

Our assumptions are also consistent with estimations of Dolfing et al. (2012) showing that the solubility is higher for transformation products of CLD.

Change in manuscript:

So, to account for the reviewer comment, we propose

- 1) To complete the current model adding a degradation term for 5bCLD.
- 2) Given the lack of knowledge and the uncertainty about degradation rate, we propose to add the following comment in the text L390 *“It should be noticed that the degradation values remained uncertain as we did not have any references for comparison. In our case, the optimization process yielded a far lower degradation rate compared to the lixiviation rate (Table 2). Consequently, the model will be less sensitive to changes in the degradation rate than in the lixiviation rate, which is the key parameter for determining the ratio in water. Additionally, there was uncertainty when comparing degradation rates for 5bCLD and CLD. The optimization process yielded degradation rates for 5bCLD and CLD of the same order of magnitude. Additional simulations showed that setting $C_{5bdegrad}$ ten times higher than C_{degrad} instead of zero reduced the 5bCLD / CLD ratio by 10 percent without changing the dynamic of the ratio and of 5bCLD lixiviation (not shown). Given that CLD transformation products are likely to be more mobile in the environment than their parent compound (Dolfing et al. 2012), we assumed that our model gave*

sufficient bases for interpreting our results.”

Do you have access to CLD/5bCLD ratio in soil to have an idea of the initial signature over time to decipher soil degradation process to those associated to surface and groundwater flowpaths?

Response: Reference of ratio in soils are in the paper of Clostre et al (2015). The median value of 0.011 in nitisols and 0.017 in andosols were used in our paper to constrain our model (see section 3.4). This does not help to speculate about ratios in water since they depend on lixiviation rates of CLD and 5bCLD. In our article, data from Cabidoche et al (2009) were used to assess CLD lixiviation rate (T_{CLD}) for andosols and nitisols. The 5bCLD lixiviation rate (T_{5bCLD}) stemming from the optimization process appears higher than T_{CLD} . This result is consistent with Devault et al (2016) who conclude for a higher mobility for 5bCLD than for CLD. Whatever, it is unlikely that CLD was leached while 5bCLD accumulated in soil profile due to the highest mobility of transformation products (Dolfing et al., 2012).

Change in manuscript: We propose to add the following sentence section 3.4 L406: “... continuously without a plateau. *This result was consistent with Devault et al (2016) who concluded on higher mobility of 5bCLD compared to CLD, and more generally with the results of (Dolfing et al., 2012), who showed that transformation products had higher mobility than CLD.*”

Second, residence time is used to explain the spatial variability of the ratio compounds/TPs. To support the discussion, the authors should provide existing reported information/simulation of these residence times:

- to discuss spatially contrasted compounds/TPs ratio delivery by soil to ground water

Response: please see our response to the comment “19)” of the First referee

- to address the question of degradation of the TP itself (especially for long residence time)

Response: longer residence time does not mean that the TP degradation is higher. In fact degradation occurs in the soil, whereas residence time in the aquifer refers to transfers in depth (below soil cover, through the unsaturated and saturated zones) where the degradation (as well as the retention) is considered as null. Groundwater residence time is generally superior to several years (up to several decades – see Gourcy et al., 2009 for instance) that is widely superior to the residence time of the infiltrated water in the soil cover (several days or months).

Third, if the sampling effort, statistical analyses and conceptual development provided a coherent approach for groundwater (slow flowpath), I have many questions on the surface water component.

Response: Global comment about flowpath, as a state of the art for our following responses to the several comments related to that aspect. This global comment justifies our approach and shows in what way the integration of the surface water component will not affect our conclusions.

First, volcanic soils in Caribbean islands have a high infiltration capacity (saturated hydraulic conductivity superior to 60 mm/h (Cattan et al., 2006; Crabit et al, 2016). Then, despite high

rainfall intensities and amounts, most of rainfall infiltrates (about 95% at the plot scale according to Cabidoche et al, (2009); more than 90% at the watershed scale according to Charlier et al., 2008; 2011) generating either subsurface or deep flows. So leaching is the main process in pesticide transport in surface water.

Second, usually, one reason to study separately pesticide transport by surface runoff is that the pesticide concentration in runoff water may vary highly according to time of pesticide application at the plot scale (Saison et al., 2008) as well as at the watershed scale (Charlier et al., 2009). It is not the case for CLD which have been applied long time ago: boundary conditions relative to pesticide concentration in soil are almost steady. Surely, during application period, agricultural practices may have affect 5bCLD/CLD ratio day by day. However our model aims to simulate the ratio evolution over a long time period. A second reason to consider separately runoff and infiltration water is that pesticide concentration in surface water at the plot scale may differ from infiltrated water. There are few references about this point for CLD. Cabidoche et al (2009) notice that CLD concentration in surface runoff was more than 3-fold lower than in drainage, while runoff volume was 10 times lower than drainage volume. They consequently neglected loads in runoff that represented less than 1/30 of those in drainage at the plot scale.

Given the previous consideration, we then choose to focus here on lixiviation process, which affect the ratio dynamic on the long term. The reviewer ask the question of the effect of event-driven process (storm event, surface runoff, erosion, application practices) on long term trends and how they can modify CLD concentration in water and the ratio. It is a difficult issue that would require getting spatial distribution of storm event, and their contribution to river pollution. This lack of knowledge probably leads to minor CLD exportation. Indeed, most of the time (even in rainy regions), surface flow in the river is driven by baseflow from aquifer's drainage, originated from water infiltration. Knowing that groundwater concentrations are widely higher than in rivers, concentrations during storm events would lead to generate diluted concentrations in surface waters.

We propose different changes relatively to the reviewer comments. We equally propose to add a § “main assumption about CLD transfer” in discussion section

Change in the manuscript: L417 addition of the § “**4.1 main assumptions about CLD transfer.**

In our study, we focused on long-term trends for CLD and 5bCLD concentration in water, along with their ratio. We considered that the main process determining pollutant concentrations in water was relative to CLD desorption by water infiltrating the soil. We assumed this hypothesis for different reasons.

Firstly, rain water mainly infiltrates. In fact, given the high soil infiltration rate (saturated hydraulic conductivity over 60 mm/h (Cattan et al., 2006; Crabit et al., 2016), most rainfall infiltrates (about 95% on a plot scale according to Cabidoche et al, (2009); more than 90% on a watershed scale according to Charlier et al. (Charlier et al., 2008, 2011)), generating either subsurface or deep flows. Consequently, transportation by surface runoff is low.

Cabidoche et al (2009) found that CLD concentration in surface runoff was more than 3-fold lower than in drainage, while the runoff volume was 10 times lower than the drainage volume. They consequently discarded loads in surface runoff that amounted to less than 1/30 of those in drainage on a plot scale

Secondly; soils have little erodibility: Cabidoche et al (2009) found that “All the soil types in FWI are acidic, which prevents clay dispersion and sheet erosion. Hydric erosion

appears to be due only to bad soil management practices, which concentrate runoff that then forms streams that are able to carry aggregates". Thus, erosion from cultivated soils is probably not a major way of CLD transportation. Moreover, given the high contribution of erosion from river beds and from non-contaminated areas in the upstream zone (due mostly to torrential type flow of rivers in FWI), the impact of surface water contamination by sediments was considered as a minor process.

Lastly, by neglecting transport via surface runoff (since sampling mainly occurred outside storm event periods), we probably underestimated pollutant exportation. Thus, we expected that it should not have a great impact on the long-term dynamics of concentrations and ratios in rivers, which is one of the main topics of our paper.

- The representativeness of the sampling (low frequency mainly during base-flow, if I well understood the database characteristics) is not discussed taking into account percentage of Chlordecon exported during storm event associated to tropical climate.

Response: See the last paragraph of the new 4.1 Section

Change in manuscript: ...*"Finally, neglecting transport via surface runoff (since sampling mainly occurred outside storm event periods), we probably underestimated pollutant exportation"...*

- With a large Koc, the question of Chlordecon released from soil to river by erosion during runoff event is never discussed. How these pulses can contribute to spatial and temporal patterns of chlordecon in surface water?

Response: See the following sentences in the new 4.1 Section

Change in manuscript: *"Second soils have little erodibility: Cabidoche et al (2009) notice that "All the soil types in FWI are acidic, which prevents clay dispersion and sheet erosion. Hydric erosion appears to be due only to bad soil management practices, which concentrate runoff that then forms streams that are able to carry aggregates". Thus, erosion from cultivated soils is probably not a major way of CLD transportation. Moreover, given the high contribution of erosion from river beds and from non-contaminated areas in the upstream zone (due mostly to torrential type flow rivers in FWI), the impact of surface water contamination by sediments is considered as a minor process."*

- For surface water, it could be relevant to know if the CLD concentrations correspond only to the dissolved phase or if it is a "total" concentration.

Response: the CLD concentration is a total concentration. See the change we suggest for the following comment L153

- Information on the filtration and purification steps are not provided in the M&M section.

Response: there was no purification nor filtration since the suspended matter content of samples was low (less than 250 mg L⁻¹). Analyses were performed on raw water. We propose to add the following sentences section 2.2.2

Change in manuscript: L153 “Analyses were carried out on raw sampling water. Thus, the water CLD and 5bCLD contents corresponded to dissolved and particulate fractions. It should be noted that the particulate fraction of the samples was low (< 250 mg L⁻¹) due to sampling conducted mainly during periods of low flow.”

- Can contaminated sediments in river potentially be remobilized by event and alter trend assessment in surface water?

Response: see our previous response

- In the conceptual model, the surface runoff and the surface water to groundwater seem not considered. The choice targeting mainly leaching and not the other off-site transport is never discussed. The authors mentioned “hope for pollution mitigation” based on statistical model, but I wonder how fast flow in river can modify this assessment.

Response: see our previous responses

In my opinion, the paper can't be published without strengthening of these points.

Specific remarks:

L324 GW, as well as in SW fed by it. And vice et versa ?

Response: whereas infiltration from ditches towards aquifers is a likely process in such regions due to the high permeability of the shallow formations (Charlier, 2007), and even if in some cases, river infiltrations may contribute also to groundwater recharge (Charlier et al., 2011), we consider that the infiltration of surface water is a neglecting process of groundwater contamination at a global scale. In fact, in cultivated areas, surface water is generally widely less contaminated in CLD than groundwaters.

L323 The age of the main geological units was used as an indicator of hydrogeology and notably residence time in the aquifers. Could you provide evidence ? Residence time assessment form others studies ?

Response: please see our response to the comment “19)” of the first referee

L57 and L60, 1993 or 1992, I guess banned in 1992 but used until 1993. Please explain.

Response: Yes there was exemption until 1993. We propose:

Changes in the manuscript: L60 “...ban in 1992 (*there was exemption in FWI until 1993*)”

L121 “they are intergrades” ?

Response: Intergrades are defined by Colmet-Daage relative to the climatic sequence ferralsols -> vertisols for soils that are “intermediate”. Since Colmet-Daage classification is specific, we propose to suppress the last part of the sentence which is unclear “~~and they are intergrades resulting from the alteration of ferralitic soils~~”

L134 unweathered formations, to several decades for old weathered formations (provide

range for “Old”)

Change in manuscript: “between a few years for recent unweathered formations (<0.5-1My), to several decades for old weathered formations (> 1My)”

L139 routine basis with CLD. For (double space before For)

Response: OK

L150 5bCLD is the main CLD co- and alteration product of CLD: what do you mean by alteration product ? Transformation / degradation product ? Please clarify ?

Response: In fact, 5bCLD can be considered both as a co-product and as a degradation product. Consulting biochemists, the word “alteration” seemed more convenient. We propose the following change:

Change in manuscript: “*5bCLD is the main alteration product of CLD (the term “alteration” here means that 5b is both a co-product and a degradation product)* for which ...”

L151 “Reference standards for CLD and 5bCLD were purchased” : provide purity degree

Change in manuscript: L152 “... for both laboratories *with a purity degree of 96.7%.*”

L149 Between sampling and analysis, no information is provided on the filtration (raw water/filtrated water?) , purification ?, please add

Response: OK analyses were performed on raw water. See previous change for L153

L181 was not detected (i.e. 0.001 for LDA26 or 0.003 $\mu\text{g L}^{-1}$ for BRGM), and an intermediate value of 0.006 $\mu\text{g.L}^{-1}$: why is different of value provided in L177 0.003 $\mu\text{g.L}^{-1}$? Please clarify

Response: 0.003 line 177 refers to the limit of detection; 0.006 is an intermediate value between the limit of detection 0.003 and the limit of quantification 0.01 when the compound was detected but not measurable. We propose to change “measurable” by “*quantifiable*” L182

L184 double space the 5bCLD

Response: OK

L183 Factors. Not clear for me, how heterogeneity of upstream catchment for SW or drainage area for GW were integrated in metrics ?

Response: factors refer to global descriptors that do not integrate such spatial heterogeneity at a local scale. Apart soil (as it is explained in the text), each site is associated with the factor value at the sampling point.

L204. For GW, double space

[Response: OK](#)

Fig. 1. I suggest to modify this figure to add sampling point distribution (the different zoom levels can be significantly reduced)

[Response: sampling distribution are presented in figure 3](#)

L248 Kendall (MK) test. We calculated Sen trends, Sen trend ? Not defined, Instead to use Sen trend in the text, I suggest to explain the information underlined by this metric (to improve understanding for the reader)

[Response: OK, please see our response to the comment 16\) of the first Referee](#)

The section 3.2.1. looks like a figure caption (modify and interpret directly in this section)

[Response: Section 3.2.1 aims to present Figure 3 and the distribution of pollution](#)

L300 “areas since 1970, i.e. during CLD application. Surprisingly, SW and GW contamination occurred outside these banana areas” Explanation ? other dissipation Processes ? Are the contaminated areas are downstream of banana areas ?

[Response: we suggested CLD misuse L304](#)

Fig3. Legend cannot be read (too small).

[Response: We propose a new Figure with a larger legend](#)

Fig4. two small, I suggest to merge some of them or provide in SI

[Response: there are two comments for Figure 4. Perhaps this comment is relative to Fig3 ? or Fig6. We propose a new Figure 6](#)

L308 contamination level. For example, the CLD content of hydrographic subsector 1 (see Figure 3 left for locations) was different from hydrographic subsector 2 even though the points in each zone had the same contamination level. It is very descriptive, please provide explanation

[Response: We propose to rephrase L308-309](#)

Change in manuscript: “For example, although sample points of subsector 1 and 2 were very close (see Figure 3a), they did not have the same contamination level. In contrast, all the sample points of subsector 1 had the same contamination level (same for subsector 2). This suggests that the hydrographic sector, i.e. the water flows within the same hydrological unit, mainly determined the contamination level of the sample points, rather than the geographical closeness of those points.”

L 320 “This statistically confirmed”/ Missing word ?

[Response and change L320: We propose to rephrase: “This is statistical confirmation of the result mapped in Figure 3...”](#)

Figure 4. provide full name under the figure Ferr_And, Nit_And. . .

Response: OK

L375 : “duration of pollution”: persistence of pollution ?

Change in manuscript: L375 “*persistence of pollution*”

Fig 7. Legend is hiding some point: modify. I suggest to redesign the figure 7 to improve understanding of key message for the reader (add sectors/types and assessment indicator) ???

Response: We propose to keep the legend of the figures in the middle and to suppress the legend of figures on the left and right sides where points are hidden. Sen trends for others factors (hydrographic sectors and historical banana area) are not represented due to the absence of relationships.

Change in manuscript: redesign of figure 7 with and a new caption: “Sen trends of CLD vs. mean log content of CLD, 5bCLD, and 5bCLD / CLD ratio (from left to right – natural logarithm) in SW, according to a) soil, and b) geology (*for soil and geology, see legend in the middle figure*).

Fig 8. time unit ? Years. . . As discussed in the main comments, all the model predictions seem to be dependant of persistence of the 5bCLD, how the results could be altered by considering TPs degradation.

Response: see response above

L437 0.1 $\mu\text{g L}^{-1}$ 437 during baseflow periods (flood flow periods being rarely sampled) given a current concentration of 0.5 $\mu\text{g L}^{-1}$ 438 on average. I don't understand your assumption?

Response: baseflow periods refers to periods without flood flows (or storm flows). Please see also our response to your previous comments on sampling

L499 “catchment scale”, you used watershed during all the manuscript why changed now? “The residence time - estimated by the water apparent age: not discussed or characterized before?”

Response and change: catchment is replaced by watershed. Regarding the residence time, it was discussed in L452-458 of the submitted version

L388 “they should lie”: sentence ?

Response and change L462: “...considering *the ratios should lie*...”

Additional references

Cattan, P., Y.-M. Cabidoche, J.-G. Lacas, and M. Voltz. 2006. *Effects of tillage and mulching on runoff under banana (Musa spp.) on a tropical Andosol*. *Soil Tillage Res.* 86:38–51.

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Dolfing J., I. Novak, A. Archelas, and H. Macarie, 2012. *Gibbs Free Energy of Formation of Chlordecone and Potential Degradation Products: Implications for Remediation Strategies and Environmental Fate*. *Environ. Sci. Technol.*, 2012, 46 (15), pp 8131–8139. DOI: 10.1021/es301165p

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1 A conceptual model of organochlorine fate from a combined 2 analysis of spatial and mid/long-term trends of surface and 3 ground water contamination in tropical areas (FWI)

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17 **Abstract.** In this study, we investigated the management of long-term environmental pollution by organic
18 pollutants such as organochlorine pesticides. We set out to identify conditions that are conducive to reducing
19 pollution levels for these persistent molecules and then propose a conceptual model of organochlorine fate in
20 water. Our approach looked at spatio-temporal changes in pollutant contents in surface water (SW) and
21 groundwater (GW) on a large scale, in order to decipher the respective roles of soil, geology, hydrology and past
22 treatment practices. The case of chlordecone (CLD) on the island of Martinique (1,100 km²) was selected given
23 the sampling campaigns carried out since 2007 over more than 150 sites. CLD, its metabolite chlordecone-5b-
24 hydro (5bCLD) and the metabolite/parent compound ratio were compared. As regards the spatial variability of
25 water contamination, our results showed that banana cropping areas explained the location of contaminated SW
26 and GW, whereas the combination of soil and geology factors explained the main spatial variability in the
27 5bCLD / CLD ratio. For temporal variability, these conditions defined a high diversity of situations in terms of
28 the duration of pollution, highlighting two groups: water draining old geological formations & ferralsols or
29 vertisols *vs.* recent geology & andosols. A conceptual leaching model provided some key information to help
30 interpret downward trends in CLD and 5bCLD observed in water. Lastly, a conceptual model of organochlorine
31 fate is proposed to explain the diversity of the 5bCLD/CLD ratio in water. Our conclusions highlight the
32 combined role of soil and groundwater residence time for differentiating between conditions that are more
33 conducive, or not, to the disappearance of CLD from the environment. This paper presents a model that provides
34 an overall perception of organochlorine pesticide fate in the environment.

Supprimé: geology

Supprimé: theoretical

36 **Keywords**

37 Pesticide; Surface water; Groundwater; Temporal dynamics; Chlordecone

40 **1 Introduction**

41 The pollution of rivers and aquifers by persistent organic pollutants (POPs) and organochlorine pesticides is a
42 global issue (Gonzalez et al., 2012; Masih et al., 2014; Montuori et al., 2014; Zhang et al., 2004). Their long-
43 term persistence after application (i.e. several decades to several centuries) raises the question of what is polluted
44 and to what level, and how to manage and live with pollution. Moreover, the environment is not uniformly
45 contaminated. Interactions between human pesticide application practices and environmental conditions lead to
46 high variability in the contamination level of environmental compartments. This variability can be perceived by
47 observing surface water (SW) and groundwater (GW) contamination.

48 Globally, changes in pesticide applications over several decades have resulted in downward and upward trends
49 for pesticide concentrations in SW (Ryberg and Gilliom, 2015; Stone et al., 2014). This is also the case for GW,
50 for which contamination trends have illustrated the leaching of pesticides from soils towards aquifers on a
51 regional scale (Bexfield, 2008; Kolpin et al., 2004; Lapworth et al., 2006). Quality in SW is highly correlated to
52 that in GW, due to strong interactions between aquifers and rivers on a watershed scale. Surprisingly, there is a
53 lack of studies combining both SW and GW observations in order to characterize pollution in all the
54 compartments (shallow and deep) of the hydrological cycle. Thus, this article addresses the issue of the
55 conditions and processes that determine the spatial distribution of a persistent pollutant in water on a regional
56 scale, investigating the case of chlordecone contamination in the French West Indies.

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57 Chlordecone (CLD, C₁₀Cl₁₀O; CAS number 143-50-0; 491 g mol⁻¹) is an organochlorine classified as a POP (U.
58 S. Environmental Protection Agency, 2012; UNEP, 2007). Numerous issues stem from CLD use in the French
59 West Indies (FWI íslands of Martinique and Guadeloupe) (Lesueur Jannoyer et al., 2017). CLD was used from
60 1970 to 1993 to control the black weevil (*Cosmopolites sordidus*) in banana plantations. Application intensity
61 greatly depended on the farmers (Cabidoche et al., 2009; Della Rossa et al., 2017; Levillain et al., 2012) and
62 introduced high spatial variability in soil contamination. Despite its worldwide ban in 1992 (there was an
63 exemption in FWI until 1993), CLD continues to contaminate aquatic ecosystems in different parts of the world
64 (Coat et al., 2011; Luellen et al., 2006). As a consequence, CLD-polluted soils in FWI go on to contaminate GW
65 (Arnaud et al., 2017; Gourcy et al., 2009) and rivers (Bocquene and Franco, 2005; Coat et al., 2011; Crabit et al.,
66 4 2 3 8 = " O q v v g u " g v " c n 0 . " 4 2 3 7 = " Q d u g t This pollution rises concerns, as G c w " f g " n c " O c t v k k s w
67 CLD causes adverse effects on health, from both acute and chronic exposure (Cannon et al., 1978; Cordier et al.,
68 2017; Multigner et al., 2015).

Supprimé: environmental

69 The persistence of pesticides in soils and their transfer to percolation water depend on various processes, such as
70 degradation and sorption, influenced by molecule properties, as well as the soil and climate context (Arias-
71 Estévez et al., 2008). For CLD, adsorption on soil aggregates, hence the risk of water pollution, greatly depends
72 on soil type, as indicated by the soil organic carbon-water partitioning coefficient (K_{oc}), which varies from 2.5
73 to 20 m³ kg⁻¹ (Cabidoche et al., 2009; Woignier et al., 2012). Moreover, at depth, contrasting residence times
74 (the water age in aquifers is defined as the mean transit time * O c € q u | g y u m k ") in aquifers, w d g t . . 3 ; : 4 +
75 ranging from several years to several decades, partly account for the variability in GW contamination by CLD
76 (Gourcy et al., 2009).

Supprimé: partitioning coefficient

Supprimé: between the sorbed part on soil organic matter

82 Recent studies highlighted the fact that degradation can occur for this molecule (Fernández-Bayo et al., 2013;
83 Mouvet et al., 2017). CLD-5b-hydro (5bCLD, C₁₀Cl₉HO; CAS number 53308-47-7; 456 g mol⁻¹) is a mono-
84 hydrochlordecone, which can be produced as an impurity during CLD manufacturing (Cabidoche et al., 2009;
85 Fernández-Bayo et al., 2013). It has also been obtained experimentally by degradation of CLD through
86 photolysis and microbial degradation (Orndorff and Colwell, 1980; Wilson and Zehr, 1979). Orndorff and
87 Colwell (1980) interpreted the in situ value of 5bCLD content as an indicator of the degradation process.
88 Studying the fate of both the parent and metabolite compounds, or their ratio, provides a more complete
89 understanding of the transportation of the molecule (Farlin et al., 2017; Gassmann et al., 2013; Kolpin et al.,
90 2004). 5bCLD has been found in soils, water and food webs, along with CLD, but at much lower levels (Borsetti
91 and Roach, 1978; Clostre et al., 2015; Coat et al., 2011; Devault et al., 2010; Martinique et al., 2012).

92
93 To sum up, in FWI human practices and the physical environment lead to high variability conditions for CLD
94 and 5bCLD that may impact the environment. Our aim was to identify the conditions that are conducive to a
95 decrease in pollution levels, in order to propose a conceptual model of organochlorine fate in water. We focus
96 here on river contamination, which is driven by all the environmental compartments, thus being an integrative
97 survey site for land-use, soil variability, and aquifer contributions. Based on the sampling campaigns in
98 Martinique (FWI) since 2007, we explored river contamination trends over time and the relationships between
99 surface and underground CLD rates in water. Spatial and temporal distributions of contamination were
100 interpreted according to soil and geology mapping, hydrology and past CLD treatment practices. This work will
101 lead on to identifying areas with a low or high impact on water pollution, in order to manage polluted areas more
102 effectively.

103 **2 Material and methods**

104 **2.1 Study site**

105 **Location and climate.** The study area covered the volcanic island of Martinique (1,100 km²) in the French West
106 Indies in the Caribbean (Figure 1). The climate is tropical, hot and humid. Annual rainfall is almost a linear
107 function of altitude (0 to 1,500 m ASL) and ranges from 2,500 to 10,000 mm on the east coast, and 1,000 to
108 10,000 mm on the west coast.

109 **Geology.** Eight volcanic units (grouped into 3 simplified types according to the age of the volcanic arcs) have
110 been identified (Germa et al., 2010, 2011) (see Figure 3 for an overview of the geological map): (1) Basal
111 Complex and Sainte Anne Series (24.8±0.4 to 0.8±0.4 Ma) for the older arc; (2) Vauclin-Pitault Chain
112 (16.1±0.2 to 4.4±0.12 Ma) and (3) South-western Volcanism (9.18±0.16 to 1.10±0.10 Ma) for the intermediate arc;
113 (4) Morne Jacob volcano (5.14±0.07 to 1.54±0.03 Ma), (5) Trois Ilets Volcanism (2.358±0.034 Ma and 346±27
114 ka), (6) Carbet Complex (998±14 to 322±6 ka), (7) Mount Conil (543±8 to 127±2 ka) and (8) Mount Pelée
115 (126±2 ka to present) for the recent arc. The volcanism is andesitic with predominantly explosive volcanoes.
116 Geological formations are thus composed by ash flows, lava flows, and reworked formations (e.g. lahars and

117 debris flows) channelled in peripheral valleys, and atmospheric fallout on a larger scale. Such geology generates
118 a high spatial variability of lithology strata and contrasting weathering levels between geological units.

119 **Soils.** Two climate sequences of soils (IUSS Working Group WRB, 2014) are found in Martinique according to
120 Colmet-Daage et al. (Colmet-Daage et al., 1965) (see Figure 3 for an overview of the soil map: (1) ferralsols ->
121 nitisols -> vertisols and (2) andosols -> nitisols. All primary minerals of andesitic rocks are weathered, so that
122 soils have a high content of secondary minerals: halloysite for nitisols, halloysite and Fe-oxihydroxides for
123 ferralsols, and allophane for andosols. In addition, Martinique has skeletal andosols and young raw soil
124 containing pumice gravels, deriving from recent pyroclasts. All these soil types are acidic. Carbon contents are
125 unusually high for tropical soils, in particular for untilled andosols, and range from 10 to 140 g kg⁻¹ according to
126 Cabidoche et al. (2009) and Brunet et al. (Brunet et al., 2009). These features may induce large differences in
127 pesticide fate in soils. Since the soil types poorly developed soil on ash and pumice and andosol are similar
128 and rich in allophanes, in this study they were grouped with ferralsols and nitisols. Ferralsols and nitisols
129 soils and ferralsols were grouped with nitisols and andosols. Ferralsols and nitisols
130 they are both rich in kaolinites (Colmet-Daage et al., 1965; Quantin et al., 1991).

Supprimé: ferralitic
Supprimé: soils (latosols)
Supprimé: ferralsols

131 **Hydrology, hydrogeology and contamination.** High rainfall intensities during tropical storms generate flash
132 floods with a torrential regime in the rivers of Martinique. Permeable soils in the Lesser Antilles favour
133 infiltration and aquifer recharge (Charlier et al., 2008). As a consequence, hydrological studies on a watershed
134 scale showed that the water budget on an annual scale is mainly controlled by underground processes, limiting
135 surface runoff contributions (Charlier et al., 2008, 2011). Stream flows are greatly influenced by SW/GW
136 interactions, suggesting that GW drainage is a major process of river contamination (Arnaud et al., 2017;
137 Charlier et al., 2009; Morgenstern et al., 2015; Mottes et al., 2015). At depth, most of the volcanic aquifers are
138 small, a few km² at most, as a result of the complex geological structure, which has undergone several phases of
139 volcanism, erosion and weathering (Lachassagne et al., 2014; Vittecoq et al., 2015). As shown by Charlier et al.
140 (Charlier et al., 2015), who compared the hydrogeological functioning of aquifers with contrasting lithologies
141 and age formations, the groundwater residence time is highly variable, between a few years for recent
142 unweathered formations (0.5-1 My), to several decades for old weathered formations (> 1 My). Given that the
143 weathering of geological formations increases with their age, it is the main cause of a global decrease in aquifer
144 permeability, notably in volcanic regions (Lachassagne et al., 2014). Indeed, clayey alteration products by
145 weathering constrain the soil's physical and hydrodynamic properties by reducing porosity, and consequently
146 permeability (Adelinet et al., 2008). It may result in various levels of river contamination by CLD linked to the
147 hydrogeological context of the watershed.

Supprimé: and they are intergrades resulting from the alteration of ferralitic soils

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148 2.2 Building up the database

149 2.2.1 CLD and 5bCLD sampling in water

150 The study period ran from the end of 2009/early 2010 to 2014. Since 2009-2010, 5bCLD has been analysed on a
151 routine basis with CLD. For SW, we used data from a programme monitoring water quality carried out by the
152 Martinique Water Office throughout Martinique and from a research programme implemented by CIRAD
153 (CIRAD, F-97285 Le Lamentin, Martinique, France) in the Galion watershed in Martinique. Sampling was

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162 carried out manually according to standard NF EN ISO 5667-3 and the FD T 90-523-1 guideline. For GW, we
163 used data from a programme monitoring groundwater quality carried out by BRGM throughout Martinique. The
164 sampling methodology was based on standard NF EN ISO 5667-3, and the FD T 90-523-3 and FD X31-615
165 guidelines. Before sampling in wells, at least three purge volumes were pumped with a submersible pump until
166 stabilization of the chemical groundwater parameters. Samples were stored at 5°C and shipped in ice coolers to
167 the BRGM analytical laboratory in Orléans, France.
168

169 2.2.2 CLD and 5bCLD analysis

170 7 d E N F " k u " v j g " o c k p " c n v g t c v k q p " r t q f w e v " q h " p n d i c ' a n d v j g " v g t o " o c n v g t c v k q
171 a degradation product), for which a commercial analytical standard is available. Reference standards for CLD and
172 5bCLD were purchased from Dr. Ehrenstorfer GmbH (Augsburg, Germany) for both laboratories with a purity
173 degree of 96.7%.

Supprime: 5bCLD is the main CLD co-
and alteration product of CLD

174 For SW, samples were analysed at the LDA26 laboratory. Analyses were carried out on raw sampling water.
175 Thus, the water CLD and 5bCLD contents corresponded to dissolved and particulate fractions. It should be noted
176 that the particulate fraction of the samples was low (< 250 mg L⁻¹) due to sampling conducted mainly during
177 periods of low flow. CLD and 5bCLD sample analyses were carried out by liquid/liquid extraction
178 (Dichloromethane and ethyl acetate 80/20) followed by Ultra High-Performance liquid chromatographic
179 separation and mass spectrometric identification. An Ultra High-Performance liquid chromatography tandem
180 mass spectrometry analysis was performed with a Thermo electron system (TSQ Quantum Ultra) or ABSciex
181 system (API4000 or API4000 Q-Trap). The compounds were separated on an Alltima C18 (5µm-150 x 2.1mm).
182 Two transitions were monitored 506.7 > 426.5 and 506.7 > 424.5 for CLD and 472.6 > 392 and 472.6 > 454.5
183 for 5bCLD. 2.4D d3 was used as the internal standard for calibration. The key parameters of the method
184 (linearity, repeatability, interday precision, specificity, extraction efficiency and limit of quantification) were
185 validated in accordance with the NF T 90-210 standard method (AFNOR 2009). The CLD and 5bCLD limits of
186 quantification were determined by spiking natural surface water samples.

187 For GW, samples were analysed at the BRGM laboratory in Orléans, France. A gas chromatography tandem
188 mass spectrometry analysis was carried out with a Bruker system (Marne la Vallée, France) composed of a
189 GC450 gas chromatography apparatus equipped with a 1177 injector, a Combi Pal (CTC) autosampler and a
190 300MS triple quadrupole mass spectrometer. The injector was equipped with a 4 × 6.3 × 78.5 mm liner with
191 fibre i n c u u " c p f " U m { Î " f g c e v k x c v k q p 0 " N M S g 3 0 q o d e q w p f w ' b y k F g " u g h d t ' c o g f " q p " c
192 column from Restek (Lisses, France). CLD and 5bCLD analyses of water samples were carried out by
193 liquid/liquid extraction followed by gas chromatographic separation and mass spectrometric identification. The
194 key parameters of the method (linearity, repeatability, interday precision, specificity, extraction efficiency and
195 limit of quantification) were validated in accordance with the NF T 90-210 standard method (AFNOR, 2009).
196 The CLD and 5bCLD limit of quantification were determined by spiking natural water samples.

197 Both the LDA26 and BRGM laboratories are accredited for pesticide analysis and are involved in proficiency
198 testing schemes organized by ANSES (French Agency for Food, Environmental and Occupational Health, and

201 Safety), thereby ensuring the quality and coherence of the results. The limits of CLD and 5bCLD quantification
202 in water were different for LDA26 and BRGM: 0.01 and 0.03 $\mu\text{g L}^{-1}$ respectively. By convention, the limits of
203 detection were set at one third of the limits of quantification, i.e. 0.003 and 0.01 $\mu\text{g L}^{-1}$ for LDA26 and BRGM,
204 respectively.

205 2.2.3 Value assessment and factors

206 **Value assessment.** For calculation, a value of 10% of the quantification limit was assigned when the compound
207 was not detected (i.e. 0.001 for LDA26 or 0.003 $\mu\text{g L}^{-1}$ for BRGM), compared to the value assigned when the compound
208 was assigned when the compound was detected but not quantifiable at LDA26.

Supprimé: measurable

209 **Factors.** The statistical analysis set out to assess the effect of various environmental factors - soils, geology,
210 hydrological sectors, historical banana area, and time - on CLD and 5bCLD concentrations and on the 5bCLD /
211 CLD ratio, determined at each sampling point. For the soil factor, as the water at one sampling site originated
212 from a watershed possibly draining various soil types, we associated each sampling point with the main soil type
213 of the watershed drained by the sampling point according to the soil map of Colmet-Daage et al. (1965). For the
214 other factors, each sampling point was associated with the factor value at the sampling point.

215 2.3 Selection of data and statistical analysis

216 2.3.1 Range of contamination values

217 The relevance of contamination was assessed according to the GW "Directive 98/83/EC" and SW "Directive 98/63/EC"
218 (European Union, 1998, 2000) and their transposition into French law (French government,
219 2001). Three thresholds of water contamination classes stemmed from these directives: 0.1, 0.5 and 2.0 $\mu\text{g L}^{-1}$.
220 The first two regulatory thresholds apply to the mean annual content in tap water intended for human
221 consumption: 0.1 $\mu\text{g L}^{-1}$ is the threshold for each pesticide (threshold applying to CLD), and 0.5 $\mu\text{g L}^{-1}$ is the
222 threshold for the sum of all pesticides. Raw water exceeding these thresholds needs to be treated for human
223 consumption. The third value, 2.0 $\mu\text{g L}^{-1}$, is the threshold beyond which, according to the regulation, water can
224 no longer be termed drinkable even after treatment. The threshold values of 0.1 and 0.5 $\mu\text{g L}^{-1}$ are also chosen to
225 define good environmental status.

226 2.3.2 Data selection

227 **Global data set.** For SW, the data set consisted of 1,866 analyses from 136 sampling points and 76 rivers. The
228 analyses were not evenly distributed. Most of the sampling points had a low measurement frequency (105 had
229 fewer than 5 analyses) and only 18 sampling points had more than 50 analyses covering the entire 2009-2014
230 period. However, the number of analyses per complete year varied between 188 and 352. For GW, the data set
231 consisted of 282 analyses from 21 sampling points and 6 water bodies. Basically, sampling occurred twice a year
232 at each sampling point. At three sampling points, sampling occurred monthly in some years.

233 **Data selection for statistical models.** For statistical analysis, we discarded data where CLD concentrations were
234 below detection limits (and consequently 5bCLD concentrations too, as 5bCLD concentrations are always lower
235 than CLD), as they would have led to an inappropriate ratio value (ratio of 1 according to the value assessment

237 rule described below). Additionally, although we gathered data from contaminated areas, some of the water
 238 samples were contaminated with CLD, but no 5bCLD was detectable. For the statistical analysis, we kept all the
 239 data (with and without quantifiable 5bCLD) from sampling points for which at least half the samples had
 240 quantifiable 5bCLD (i.e. water samples analysed for CLD and 5bCLD). This avoided overestimating the concentration for the
 241 sampling point, which would have been the case if we had discarded all the data with no quantifiable 5bCLD.
 242 For SW, we selected 963 data items. This SW data set
 243 covered 38 sampling points out of a total of 136. For GW, we selected 123 data items. This GW data set came
 244 from 7 sampling points.

Supprimé: measurable
 Supprimé: measurable
 Supprimé: measurable

245 **Data selection for temporal analysis on specific rivers.** In order to highlight differences between pesticide
 246 trends depending on the sampling point, we chose rivers for which the analysis covered the entire 2009-2014
 247 period. This led to the selection of 14 sampling points, all having more than 50 analyses. As stated above, we
 248 discarded analyses where CLD and 5bCLD contents were below detection limits.

249 **2.3.3 Statistical analysis**

250 **Models.** To ensure that the residue distribution of the analysis of variance (ANOVA) model followed the
 251 assumptions of equal variance and normality, we used log transformed (natural log) data. We analysed our SW
 252 and GW data sets by a multi-way analysis of variance using the MIXED procedure in SAS software (SAS
 253 Institute Inc, 2002). The effects to be taken into account in the models were chosen by comparison of the AIC
 254 (Akaike Information Criterion).

255 Model 1 was used on the SW data set to test different effects on the CLD content, the 5bCLD content and the
 256 ratio of the 5bCLD content to the CLD content in SW. The soil and geology factors were dependent on each
 257 other. For this reason, only combinations of these 2 factors were considered in the model.

258 Eq. (1)

259 where Y_{ijklm} is the observation (i.e. $\ln(5bCLD)$, $\ln(CLD)$ or $\ln(5bCLD/CLD)$), μ is the overall mean, α_j is the
 260 (soil x geology) type effect, β_{ij} the hydrological sector effect for each (soil x geology) type, γ_t is the date effect,
 261 D_{ijk} the random effect of the sampling point for each (soil x geology) type and ϵ_{ijkl} is the residual error. Indices i,
 262 j, t, k, l represent following factors for soil x geology, hydrological sector, date, sampling point and sample
 263 replication, respectively.

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264 Model 2 was used on the GW data set. Soil and geological factors were closely linked for the GW data set
 265 (andosols were always associated with recent geological formations and ferralsols with old geological
 266 formations), making it impossible to distinguish the soil effect from geology; likewise for groundwater basins
 267 and hydrographic sectors. Consequently, only soil and hydrographic sectors were tested for model 2:

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268 Eq. (2)

269 where Y_{ijklm} is the observation (i.e. $\ln(5bCLD)$, $\ln(CLD)$ or $\ln(5bCLD/CLD)$), μ is the overall mean, α_j is the
 270 soil type effect, β_{ij} the hydrological sector effect for each soil, γ_t is the date effect, D_{ijk} the random effect of the
 271 sampling point for each soil and ϵ_{ijkl} is the residual error. Indices i, j, t, k, l represent following factors for soil,
 272 hydrological sector, date, sampling point and sample replication, respectively

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281 The significance of the sampling point effect was assessed by comparison of - 2 log-likelihood from the models
282 with and without the sampling point as the random effect, as this difference followed a chi-square distribution
283 under the null hypothesis.

Supprimé: As our data set was log transformed, dispersion indices were calculated as half the difference between the limits of the confidence interval (confidence coefficient: 0.68).¶

284 **Trend analysis.** For SW, to study temporal trends, we selected estimated means of the time series for each date.
285 Autocorrelations were assessed with the Durbin-Watson test and monotonic trends were assessed with the Mann-
286 Kendall (MK) test. We calculated Sen trends (Sen θ u " u n q r g (Gilbert, 1987) for each variable (CLD, 5bCLD and ratio) in order to compare dynamics for the two compounds. The Sen trend of a set of two-dimensional points (xi,yi) and (xj,yj) is t j g " o g f k c p " q h " v j g " u n q r g u " * f l " " { k + l * z l " " z k + " f g
287 sample points. The Sen trend is more robust than the least-squares estimator, because it is much less sensitive to
288 outliers.

291 2.3.4 Conceptual model of CLD fate

292 A simple iterative leaching model was developed to assess the evolution of CLD, 5bCLD and the 5bCLD / CLD
293 ratio over time. This model expressed that the 5bCLD / CLD ratio in water equally depended on degradation and
294 transfer rates as well as the remaining storage of CLD and 5bCLD in soils. The governing equations are given
295 below:

296 Eq. (3): CLD storage in soil

297 Eq. (4): 5bCLD storage in soil

299 Eq. (5): ratio in water

300 T_{CLD} and T_{5bCLD} are the rates of lixiviation for CLD and 5bCLD (i.e. the ratio of lixiviated mass of CLD or
301 5bCLD to their respective mass in soil), respectively, C_{degrad} the rate of CLD degradation into 5bCLD, $C_{5bdegrad}$
302 the rate of 5bCLD degradation, and t the time. CLD and 5bCLD are expressed in units of mass. According to
303 data reported by Cabidoche et al. (2009), considering an area of 1 m² and that pollutants are distributed within
304 the first 3 dm of soil, T_{CLD} is expressed as follows:

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305 Eq. (6):

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306 where K_{oc} (L kg⁻¹) is the partitioning coefficient between the sorbed part on soil organic matter and the
307 dissolved part in water, D (kg dm⁻³) the bulk density, C (g kg⁻¹) the soil carbon content, R (dm) the annual
308 amount of rainfall, S the soil surface area (dm²) and d the soil depth (dm).

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309 The calculation steps are given below:

- 310 - the initial CLD and 5bCLD stocks were set to 100 and 0 units of mass respectively
- 311 - calculation of leached CLD quantities (Eq. (3))
- 312 - calculation of degraded CLD quantities, i.e. transformed in 5bCLD (Eq. (3))
- 313 - calculation of remaining CLD quantities (Eq. (3))
- 314 - calculation of leached 5bCLD quantities (Eq. (4))
- 315 - calculation of degraded 5bCLD quantities (Eq. (4))
- 316 - calculation of remaining 5bCLD quantities (Eq. (4))

326 - calculation of mass ratio in water (Eq. (5)), which accounts for the concentration ratio since the two
327 compounds are leached with the same water quantities.

328 **3 Results**

329 **3.1 Variability of CLD contamination and its relationships with 5bCLD**

330 Figure 2 shows the relationship between the means of 5bCLD and CLD in rivers at each sampling point. We
331 found that the water 5bCLD content was at least tenfold lower than the water CLD content. However, there was
332 not a unique relationship between 5bCLD and CLD. The frequency distribution of the means of the 5bCLD to
333 CLD ratio in SW and GW clearly showed that a threshold of 0.07 divided the data set into two groups: a low and
334 a high ratio around 0.02 and 0.1, respectively. According to Devault et al. (Devault et al., 2016), these
335 differences cannot stem from the use of different commercial products or different batches of the same product.
336 Indeed, these authors, found no significant statistical difference between the ratio of the commercial products
337 Kepone® and Curlone® used in FWI, no more than they did between samples from different batches of
338 Curlone®. They found a mean ratio in commercial products of 0.00077 ± 0.00027 , i.e. ten times lower than our
339 observations in river.

340 **3.2 Spatial analysis**

341 **3.2.1 General distribution**

342 Figure 3 presents the CLD concentrations (top) and the 5bCLD/CLD ratio (bottom) for SW (square/triangle) and
343 GW (star) throughout Martinique, according to hydrological sectors (left), soil (middle), and geology. The top of
344 Figure 3 shows that the most challenging areas relative to CLD contamination were mainly situated in the
345 northern Atlantic and central part of Martinique. The distribution for the 5bCLD / CLD ratio was different. The
346 bottom of Figure 3 shows that the group with the high ratio (>0.07) was mainly located either in the highly
347 contaminated northern areas, or in some parts of the low-contamination areas in southern and western
348 Martinique.

349 We observed overall consistency between the distribution of SW and GW contamination: the higher the CLD
350 content or 5bCLD / CLD ratio for SW, the higher the CLD content or 5bCLD / CLD ratio for GW. However, the
351 west coast displayed some exceptions, since we observed contaminated GW (primarily low contamination) while
352 CLD was not detected in the rivers in the neighbourhood. Similarly, the 5bCLD / CLD ratio for GW belonged to
353 the high value group ($>0.07 \mu\text{g L}^{-1}$), while the 5bCLD / CLD ratio for SW belonged to the low value group, or
354 was not available because of no contamination.

355 **3.2.2 Impact of physical conditions**

356 *Land-use practices: high level of contamination in historical banana areas*

357 Globally, for the water CLD content, the SW and GW contamination sites matched with the historical banana
358 areas since 1970, i.e. during CLD application. Surprisingly, SW and GW contamination occurred outside these
359 banana areas. This was mostly with low concentrations under $0.1 \mu\text{g l}^{-1}$ and rarely with the higher levels (one

360 point in the South-West for GW, far from the banana area.). Most of these isolated points had a high 5bCLD /
361 CLD ratio, leading the 5bCLD / CLD ratio not to match banana field distribution, suggesting past CLD misuse.

362 **Hydrographic sector: a functional relationship between measurement points**

363 Introducing hydrographic subsectors made it possible to establish a functional relationship between measurement
364 point data. Notably, this helped to explain why some points close to each other did not have the same
365 contamination level. For example, although sample points of subsector 1 and 2 were very close (see Figure 3a),
366 they did not have the same contamination level. In contrast, all the sample points of subsector 1 had the same
367 contamination level (same for subsector 2). This suggests that the hydrographic sector, i.e. the water flows
368 within the same hydrological unit, mainly determined the contamination level of the sample points, rather than
369 the geographical closeness of those points. However, some differences were found on the north-east coast. This

Supprimé: For example, the CLD content of hydrographic subsector 1 (see Figure 3 left for locations) was different from hydrographic subsector 2 even though the points in each zone had the same contamination level.

370 was encountered in zone 3, where the contamination levels seemed to be linked to the altitudinal gradient.
371 Contamination increased downwards in coherence with banana field distribution along the coast at the lowest
372 altitudes. The statistical results summarized in Table 1 confirm this interaction between hydrographic sectors and
373 soil /geology for CLD and GW. However, no effect was found for the 5bCLD content and the 5bCLD / CLD
374 ratio.

375 **Soil type: a factor explaining some ratio variations in SW**

376 Table 1 shows significant differences in GW CLD contamination according to the soil/geology pair: GW on
377 nitisols, which are associated with old formations (older than 1 My), was more contaminated than on andosols
378 associated with recent formations (1My to present). This did not result in any significant difference for SW.
379 However, for SW, we observed significant differences for the 5bCLD / CLD ratio, opposing a low ratio for
380 nitisols to a higher ratio for andosols (Figure 4). We also noted a higher ratio for vertisols. This is statistical
381 confirmation of the result mapped in Figure 3, showing high 5bCLD / CLD ratios on vertisols in southern
382 Martinique.

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383 **Geology: a factor explaining ratio variations in SW and GW**

384 The age of the main geological units was used as an indicator of hydrogeology, and notably residence time in the
385 aquifers, which is linked to pesticide transfer kinetics in GW, as well as in SW fed by it. Thereby, shorter
386 residence times were observed for aquifers located in more recent and unweathered geological formations. It can
387 be seen in Figure 3 that the highest CLD contents in water matched with recent geological formations in the
388 banana cropping area (northern half of the island). Medium and low CLD contents were observed in other older
389 geological units, or outside banana cropping areas. As regards the 5bCLD / CLD ratio, the highest values were
390 only observed in the most recent units (0.5 My to present), for the most contaminated water bodies in the North
391 Atlantic area (not shown).

392 It is interesting to note that the soil effect depended on geology. Figure 4 illustrates this, presenting the mean
393 ratio for each soil type according to the age of the geological formations. For andosols and ferralsols/andosols,
394 the ratio appeared to be significantly higher for recent geology.

395 To sum up, banana cropping areas explained the location of contaminated SW and GW, whereas the
396 combination of soil and geology factors explained the main spatial variability of the 5bCLD / CLD ratio, with

404 the highest values in the North associated with recent geological units and the highest values in the South
405 associated with vertisols.

406 3.3 Temporal analysis

407 3.3.1 Pesticides evolve differently in GW

408 Figure 5 illustrates pesticide trends in GW for the three longest available time series. The mean CLD content
409 globally decreased for two sites (Chalvet and Source Morne Figure) and remained stable for Lelene, while the
410 5bCLD content had a more erratic evolution, probably due to the greater influence of hydrological conditions
411 (climatic seasonality). As pointed out by Arnaud et al. (2016), these contamination periods correspond to rising
412 and falling groundwater levels, and therefore to periods of aquifer recharge. For the two sites showing a decrease
413 in water CLD content, the number of samples with 5bCLD contents below the detection limit decreased over
414 time, and equalled zero in the case of the Source Morne Figure site after 2011. This was consistent with an
415 increase in 5bCLD content, or at least with a more regular occurrence of positive values. Lastly, despite the
416 impossibility of generalizing behaviour with the limited sampling sites and available period series, the
417 groundwater data sets showed an interesting evolution pattern with, in some cases, a decrease in CLD content
418 associated with an increase in water 5bCLD content.

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419 3.3.2 In SW: the pesticide concentration and ratio globally decreased

420 From all the available data, we observed a highly significant downward trend in mean river concentrations for
421 the CLD content, 5bCLD content and the 5bCLD / CLD ratio in water (a slope of -0.008, -0.028, -0.018,
422 respectively). It is interesting to note that the decreasing trend for the 5bCLD content was about threefold higher
423 than for the water CLD content.

424 More specifically, Figure 6 shows the evolution of water CLD content for the 14 rivers with the highest
425 measurement frequency. Globally, the mean Sen trend was -0.008 for the log, meaning that the CLD content was
426 halved after 7.5 years. Although most of the rivers showed a significant decrease in water CLD content, some of
427 them were characterized by a constant level of contamination (Saint Pierre, Pont RN Rouge) and even one by a
428 slight increase (Camping Matouba). Independently, we noted a high variation in the level of contamination.

429 A further analysis of temporal evolution (Figure 7) highlighted a relationship between Sen trends for CLD and
430 the mean water 5bCLD contents (regression p-value =0.06): the lower the water 5bCLD content, the greater the
431 decrease in water CLD content. A similar trend was observed for the 5bCLD / CLD ratio (regression p-
432 value=0.05), while the relationship was not significant for mean water CLD content. This indicated that the
433 decrease intensity did not depend on water CLD content. Figure 7a and 7b (left: Sen CLD vs. mean CLD) shows
434 favourable situations at the bottom left, where strong decreases in water CLD content were associated with a low
435 water CLD content in SW, which gives hope for pollution mitigation. Adversely, in the situations at the top right
436 of the figure, the pollution level is likely to last for a long time.

437 Additionally, Figure 7b shows that the smallest decreases in water CLD content were partly associated with
438 recent (0.1 My to present) geological formations and that the largest decreases were associated with older ones.

442 Lastly, regarding soils, Figure 7a shows that while andosols were distributed over the entire range of Sen trends,
443 ferralsols and vertisols characterized large decreases in water CLD content.

444 To sum up, high water CLD contents decreased with low water 5bCLD contents and low 5bCLD / CLD ratios
445 were encountered for basins situated on old geological formations, and mostly ferralsols or vertisols. On andosols
446 and recent geological formations, the water CLD content did not vary over the study period, and the water
447 5bCLD content and 5bCLD / CLD ratio were high. These conditions define a high diversity of situations with
448 regard to the persistence of pollution.

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449 3.4 Model simulation

450 In order to grasp the complex fate of CLD and 5bCLD, we used the simple model presented in Sect. 2.3.4. It is
451 an iterative leaching model investigating the theoretical fate of CLD and 5bCLD in water, accounting for
452 CLD and 5bCLD lixiviation rates (T_{CLD} and T_{5bCLD}), as well as the rate of CLD degradation into 5bCLD (C_{degrad}).

453 Table 2 gives the results of the optimization processes in order to assess T_{5bCLD} , C_{degrad} and $C_{5bdegrad}$ from realistic
454 values of T_{CLD} and the 5bCLD / CLD ratios. Thus, according to Eq. 6 (see Sect. 2.3.4), T_{CLD} may vary from
455 0.017 for an andosol (And model) to 0.15 for a nitisol (Nit model), considering the respective values given by
456 Cabidoche et al., (2009) of 20,000 and 2,000 L kg⁻¹ for Koc, 0.55 and 1.1 kg dm⁻³ for bulk density D, 70 and 20
457 g kg⁻¹ for soil carbon content C, and 4,000 and 2,000 mm for annual rainfall R. We targeted the 5bCLD / CLD
458 ratios of 0.1 and 0.025 in water (cases And1, Nit1 and And2, Nit2, respectively), which corresponded to the
459 median 5bCLD / CLD ratios of SW for the two groups identified in Sect. 3.1. We applied a constraint on the
460 5bCLD / CLD ratios in soil, considering that the ratios should lie between 0.01 and 0.017, referring to the
461 median value encountered for andosols and nitisols, respectively (Clostre et al., 2015).

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462 It should be noticed that the degradation values remained uncertain as we did not have any references for
463 comparison. In our case, the optimization process yielded a far lower degradation rate compared to the
464 lixiviation rate (Table 2). Consequently, the model will be less sensitive to changes in the degradation rate than
465 in the lixiviation rate, which is the key parameter for determining the ratio in water. Additionally, there was
466 uncertainty when comparing degradation rates for 5bCLD and CLD. The optimization process yielded
467 degradation rates for 5bCLD and CLD of the same order of magnitude. Additional simulations showed that
468 setting $C_{5bdegrad}$ ten times higher than C_{degrad} instead of zero reduced the 5bCLD / CLD ratio by 10 percent
469 without changing the dynamic of the ratio and of 5bCLD lixiviation (not shown). Given that CLD transformation
470 products are likely to be more mobile in the environment than their parent compound (Dolfing et al. 2012), we
471 assumed that our model gave sufficient bases for interpreting our results.

472 Figure 8 shows the results of two simulations: simulation And2 corresponds to an andosol situation with high
473 soil retention, and simulation Nit1 to a nitisol situation with low soil retention (Table 2). It should be noted that,
474 according to Eq. (3) and (4), Figure 8 shows the leached quantities of CLD and 5bCLD, not the concentration.
475 However, as the two compounds were lixiviated with the same quantities of water, the shape of the concentration
476 curve and quantity curve did not differ.

477 The simulation results showed that the ratio increased with time over the entire period up to a plateau (see Figure
478 8). A decrease in the ratio was not simulated, although a global trend was noted for our observed data on the

483 whole. At one sampling point, such a decrease could occur with an increase in lixiviation conditions (not
 484 shown), which may have been linked to land use changes. More likely, it could have been an artefact due the
 485 difficulty in determining low values near the quantification threshold.

486 CLD decreased exponentially in the modelling approach. The current decrease mainly observed in SW marched
 487 these dynamics (linear decrease in log scale, Figure 6). Interestingly, we found that the decrease rate for andosols
 488 (simulation And2 - Figure 8) was far lower than for nitisols (simulation Nit1). This matched the andosol
 489 situation, where no significant decrease in the river was observed.

490 5bCLD first increased and then decreased at the same time as CLD. This may explain why we found a 5b CLD /
 491 CLD ratio increase, whereas a 5bCLD decrease was observed. Our simulations also showed that T_{5bCLD} must be
 492 higher than T_{CLD} otherwise the ratio increased continuously without a plateau. This result was consistent with
 493 Devault et al (2016) who concluded on higher mobility of 5bCLD compared to CLD, and more generally with
 494 the results of (Dolfing et al., 2012), who showed that transformation products had higher mobility than CLD.
 495 Optimization processes also gave a higher value for T_{5bCLD} (Table 2), given that high ratios are unlikely when
 496 T_{CLD} is high (0.15) since it yields a T_{5bCLD} of 1 (meaning that all 5bCLD is leached).

497 Lastly, despite difficulties in predicting what would happen for each location, our simulations gave interesting
 498 insights for a better understanding of the global dynamics of the 5bCLD / CLD ratio and explained some of the
 499 observations in water.

500 **4 Discussion**

501 Our results showed high spatial and temporal variability for water CLD content in SW and GW contamination.
 502 By relating water CLD content to its metabolite compound, 5bCLD, we highlighted physical conditions relative
 503 to soils and geology that may explain its variability in water, but also in the dynamics of pollution trends. We
 504 summarized our conclusions in a conceptual scheme presented below. But first, let us specify the interpretation
 505 framework.

506 **4.1 Main assumptions about CLD transfer**

507 In our study, we focused on long-term trends for CLD and 5bCLD concentration in water, along with their ratio.
 508 We considered that the main process determining pollutant concentrations in water was relative to CLD
 509 desorption by water infiltrating the soil. We assumed this hypothesis for different reasons.

510 Firstly, rain water mainly infiltrates. In fact, given the high soil infiltration rate (saturated hydraulic conductivity
 511 over 60 mm/h (Cattan et al., 2006; Crabit et al., 2016), most rainfall infiltrates (about 95% on a plot scale
 512 according to Cabidoche et al. (2009); more than 90% on a watershed scale according to Charlier et al. (Charlier
 513 et al., 2008, 2011)), generating either subsurface or deep flows. Consequently, transportation by surface runoff is
 514 low. Cabidoche et al (2009) found that CLD concentration in surface runoff was more than 3-fold lower than in
 515 drainage, while the runoff volume was 10 times lower than the drainage volume. They consequently discarded
 516 loads in surface runoff that amounted to less than 1/30 of those in drainage on a plot scale

517 Secondly, soils have little erodibility: Cabidoche et al (2009) found v j c v " ð C n n " v j g " e acıđık, n " v { r g u " k p " H Y K " c t
 518 which prevents clay dispersion and sheet erosion. Hydric erosion appears to be due only to bad soil management

519 Thus, erosion from cultivated soils is probably not a major way of CLD transportation. Moreover, given the high contribution of
520 erosion from river beds and from non-contaminated areas in the upstream zone (due mostly to torrential type
521 flow of rivers in FWI), the impact of surface water contamination by sediments was considered as a minor
522 process.
523 Lastly, by neglecting transport via surface runoff (since sampling mainly occurred outside storm event periods),
524 we probably underestimated pollutant exportation. Thus, we expected that it should not have a great impact on
525 the long-term dynamics of concentrations and ratios in rivers, which is one of the main topics of our paper.
526

527 **4.2 CLD is degraded and contamination decreases**

528 First of all, the CLD content in SW tallied with the areas where CLD had been applied, i.e. in banana cropping
529 areas, irrespective of geology and soils. This was consistent with a global link between the location of
530 contaminated soil areas and the location of contaminated rivers, as shown on a watershed scale by Della Rossa
531 (Della Rossa et al., 2017). Surprisingly, we found that, overall, the soil type had no significant effect on water
532 CLD content in SW, although large differences in CLD content were usually encountered in soils (Clostre et al.,
533 2015; Devault et al., 2016). This paradoxical result was consistent with previous work showing that the most
534 contaminated soils are not the most contaminant for water, owing to their different capacity to retain the
535 molecule (Cabidoche et al., 2009; Levillain et al., 2012; Woignier et al., 2012). In other words, two types of soils
536 with different CLD contents may release the same quantity of CLD into water. However, our simulations
537 showed (see Figure 8) that over a long time scale, CLD contents in a river will quickly decrease for basins
538 draining soils such as nitisols, due to their low capacity to retain CLD.

539 In this environment, our results were in line with CLD degradation, being visible over a decadal time period
540 despite its strong persistence in the environment. This was hypothesized by observing the distribution of 5bCLD
541 / CLD ratios in water (median of 0.03; 1st centile of 0.006) with a far higher median and first centile value than
542 in the commercial products Kepon[®] and Curlone[®], used in FWI (mean ratio of 0.00077 ± 0.00027, (Devault et
543 al., 2016)). This was consistent with the result obtained by Devault et al. (2016), who found high 5bCLD / CLD
544 ratios in soils and, in particular, larger amounts of 5bCLD than should have been applied using commercial
545 formulations.

546 The water CLD content in SW decreased as well as the water 5bCLD content and the 5bCLD / CLD ratio. Given
547 the mean Sen trends of about -0.008 for CLD (see Sect. 3.3.2), it takes about 40 years to yield the threshold of
548 0.1 µg L⁻¹ during baseflow periods (flood flow periods being rarely sampled) given a current concentration of
549 0.5 µg L⁻¹ on average. This trend was higher than that expected by Cabidoche et al. (2009), maybe because the
550 authors underestimated the degradation process, which is still not greatly documented. However, it was
551 consistent with the results obtained by Crabit (Crabit et al., 2016) based on a storage approach that assessed the
552 duration of CLD pollution of a river of a watershed at 60 years. These results on the island of Martinique could
553 indeed be extrapolated to other CLD-contaminated areas, such as in the Guadeloupe archipelago (FWI) where
554 CLD was also intensively applied in banana plantations.

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557 4.3 Hypothesis relative to leaching processes

558 One of the main questions in this paper was what the 5bCLD / CLD ratio represents. To answer this sensitive
559 issue, we differentiated between three dimensions. A temporal dimension, because the 5bCLD / CLD ratio is
560 assumed to increase over time as degradation progresses. A spatial dimension, since the 5bCLD / CLD ratio may
561 depend on local degradation conditions. A dynamic dimension, since the 5bCLD / CLD ratio may depend on the
562 mobility properties of both molecules, CLD and 5bCLD.

563 The temporal dimension was firstly related to the long application period (from 1970 to 1993 for CLD), given
564 that land-use changes led to different application phases in the 70s and 80s and that land-use changes are
565 correlated with soil contamination levels (Desprats et al., 2004). Secondly, comparing simulation results to
566 measurement time series, the temporal dimension could also be grasped by observing GW, if we consider that
567 the residence time within the aquifer gives a temporal window on the water infiltration conditions (Gourcy et al.,
568 2009; Tesoriero et al., 2007). The residence time - estimated by the water apparent age - depends on
569 hydrogeological properties, and thus to the geological context (type of lithology and its weathering level,
570 geometry of the geological deposits, etc.). For example, we observed that high 5bCLD / CLD ratios were mainly
571 located in the waters of northern Martinique, where rivers drain recent geological formations. In that area,
572 unweathered formations favour rapid transfers and thus low GW residence times of several years (Arnaud et al.,
573 2017; Gourcy et al., 2009). Thus, in that area, GW is young and probably today mainly composed of waters that
574 percolated in the last decade with a 5bCLD / CLD ratio close to the current 5bCLD / CLD ratio in soil leaching
575 waters. Conversely, the higher groundwater residence times in more weathered geological formations probably
576 characterize older GW (residence time of several decades) where the 5bCLD / CLD ratio may reflect an earlier
577 5bCLD / CLD ratio in soil leaching waters - closer to the ratio in the commercial product - during periods of
578 application or just several years after, leading to lower 5bCLD / CLD ratios in water.

579 The spatial dimension is hard to grasp since some of the variability can be attributed to the spatio-temporal
580 variability of land-use changes over the application period. Considering that soil might be an important factor,
581 the results from Clostre et al. (2015) show that the distribution of the 5bCLD / CLD ratio differs little from one
582 soil to another, with a median value of around 0.011 [0.002 0.077] in andosols and 0.017 [0.007 0.081] in
583 nitisols. This does not mean that degradation does not depend on soil, but it does mean that we cannot assess the
584 effect of soil on degradation. It is interesting to note that the simulations accounting for nitisols and andosols in
585 Table 2 give close values of 0.14% and 0.16% for the degradation rate, respectively. The soil factor could
586 therefore not be considered decisive in explaining spatial degradation intensity.

587 For the dynamic dimension, our theoretical leaching model helped to represent how contamination evolved. On
588 the whole, the simulations accounting roughly for andosol and nitisol conditions tallied well with our
589 observations or with results from the literature: i) a large decrease in CLD was associated with a low 5bCLD /
590 CLD ratio, and ii) nitisol situations were more conducive to a contamination decrease than andosol situations,
591 considering pollution duration as noted by Cabidoche (Cabidoche et al., 2009).

592 Lastly, this discussion shows that the combined role of geology and soils together may explain 5bCLD / CLD
593 ratio levels. In a comprehensive way, we derived a conceptual scheme of water contamination on a regional
594 scale.

596 4.4 A conceptual scheme of water contamination on a regional scale

597 We propose a conceptual scheme in Figure 9 to explain differences in 5bCLD / CLD ratios in water. We first
598 assumed that degradation occurs in soils. This process, which is combined with other processes determining
599 CLD and 5bCLD fate in soil, results in a general increase in water 5bCLD content and in the 5bCLD/CLD ratio,
600 which is more or less pronounced depending on the soil. Hydrogeology teaches us that SW today could either be
601 a signal of ancient infiltrations and transfers underground, several decades ago, when 5bCLD/CLD ratios in soils
602 were low (long residence time), or a signal of recent percolations, several years ago, when 5bCLD/CLD ratios in
603 soils were high (short residence time). Thus, soil properties and residence times both contribute to explaining the
604 current impact on water quality in SW. This explanation is consistent with high 5bCLD/CLD ratios in northern
605 Martinique on recent geological formations, and low 5bCLD/CLD ratios elsewhere. For high 5bCLD/CLD ratios
606 in the South on vertisols, we can speculate that the degradation process was greater in this soil type (like soil 2 in
607 Figure 9) because lixiviation is lower in the southern area with a lower rainfall rate. This may explain the higher
608 5bCLD/CLD ratios in SW, as simulated by a previous model, despite a longer residence time in the aquifers.
609 All of these results identify a set of conditions that favour the disappearance of CLD from the environment,
610 namely ferralsols with low retention properties on older geological formations, while others - notably andosols
611 with high retention rates on recent formations - are more risky.

612 5 Conclusion

613 The aim of this paper was to identify conditions that are conducive to a decrease in organochlorine pollution
614 levels in Martinique (FWI). We adopted an unusual approach that accounted, on the one hand, for the
615 interactions between aquifers and rivers on a watershed scale and, on the other hand, for the fate of CLD and its
616 compound 5bCLD. This approach was fruitful and led to the proposal of a global scheme of water contamination
617 on a regional scale accounting for physical conditions relative to soils and geology. This scheme coherently links
618 the various amounts of chlordecone (CLD) and its metabolite 5bCLD in SW and GW. It explains their variability
619 in water, but also in the dynamics of pollution trends.

620 Our results have several implications for evaluating diffuse pollution of agricultural origin. The spatial analysis
621 of metabolite/parent compounds provided some interesting information for identifying risky areas, or areas
622 where persistent pollutants are more likely decreasing. This also provided some insights into key parameters that
623 control these conditions and environmental vulnerability to agricultural pollution. It led to implications regarding
624 where and how to act to reduce impacts (e.g. choice of crops according to pollution levels, since some plants are
625 less sensitive to contamination than others (Clostre et al., 2015), constraints on water management, such as
626 drinking water and irrigation, choice of priority areas to test decontamination processes, setting up compensation
627 plans according to the risk, etc.). Another implication is to promote continuous long-term observations as
628 opposed to one-off sampling, completing modelling approaches: in our case, long CLD time series revealed a
629 faster decrease than that expected by previous model predictions. Lastly, such a spatial and temporal overview is
630 required on a large scale to help stakeholders manage pollution on a territory scale, accounting for the main

633 characteristics of the landscape. This is the main challenge for the OPA-C Observatory in FWI (Cattan et al.,
634 2017).

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640 **6 References**

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Table 1: Effects of physical conditions on the contamination level of surface water (model 1) and groundwater (model 2), showing probability levels of tested factors

806

	CLD	5b	ratio
Model 1: surface water			
Soil x geology	0.7210	0.5989	<0.0001
Soil x geology x hydrographic sector	0.9077	0.1377	<0.0001
Date	<0.0001	<0.0001	<0.0001
Model 2 : groundwater			
Soil (or geology)	0.0228	0.8143	0.1209
Soil (or geology) x hydrographic sector	<u>0.0674</u>	0.2811	0.6333
Date	<0.0001	<0.0001	<0.0001

Bold: statistically significant at the 0.05 probability level

Underlined italics: statistically significant at the 0.10 probability level

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