

Implications of water management representations for watershed hydrologic modeling in the Yakima River Basin

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Abstract. Water management substantially alters natural regimes of streamflow through modifying retention time and water exchanges among different components of the terrestrial water cycle. Accurate simulation of water cycling in intensively managed watersheds, such as the Yakima River Basin (YRB) in the Pacific Northwest of the U.S., faces challenges in reliably characterizing influences of management practices (e.g., reservoir operation and cropland irrigation) on the watershed hydrology. Using the Soil and Water Assessment Tool (SWAT) model, we evaluated streamflow simulations in the YRB based on different reservoir operation and irrigation schemes. Simulated streamflow with the reservoir operation scheme optimized by the RiverWare model better reproduced measured streamflow than the simulation using default SWAT reservoir operation scheme. Scenarios with irrigation practices demonstrated higher water losses through evapotranspiration (ET), and matched benchmark data better than the scenario that only considered reservoir operations. Results of this study highlight the importance of reliably representing reservoir operations and irrigation management for credible modeling of watershed hydrology. The methods and findings presented here hold the promise to apply to other intensively managed watersheds to enhance water resources assessment.

Keywords: Reservoir operation; Irrigation; Managed watershed; RiverWare; SWAT

1. Introduction

Ever-intensifying human activities have profoundly affected terrestrial water cycling across the globe (Jackson et al., 2001), particularly at the watershed scale (Vörösmarty and Sahagian, 2000; Yang et al., 2015; Yang et al., 2014). Water management substantially alters natural regimes of streamflow through modifying retention time and water exchanges among different components of the terrestrial water cycle (Haddeland et al., 2007). Hydrologic consequences of management activities should

30 be explicitly investigated for effective water resource management (Siebert et al., 2010), especially for watersheds striving to
31 maintain sustainable water supply for multiple users. Accurate simulation of water cycling in intensively managed watersheds
32 faces challenges in reliably characterizing influences of management practices (e.g., reservoir operations and cropland
33 irrigation) on the hydrologic cycling (Wada et al., 2017). Explicit analyses of how model representations of water
34 impoundments and withdrawals would affect hydrologic modeling are needed to advance knowledge of water cycling in
35 managed watersheds.

36 Construction of dams and reservoirs has substantial influences on the magnitude and variability of downstream runoff
37 (Lu and Siew, 2006; Vicente-Serrano et al., 2017). For example, reservoir operations reduced 9% - 25% of summer runoff to
38 the Pacific Ocean in western U.S. and Mexico (Haddeland et al., 2007). In heavily dammed regions, reduction of streamflow
39 following dam construction even reached 100% (Graf, 1999). Reservoir operations affect the temporal variability of
40 streamflow at multiple temporal scales in different regions across the globe (Huang et al., 2015; Zajac et al., 2017). Regulated
41 streamflow from reservoirs to downstream areas contributes to attenuating flood peaks and volumes, but could increase
42 baseflow in dry seasons (Batalla et al., 2004).

43 Reliable representation of reservoir operations in hydrological models is critical for credible simulation of water
44 cycling (Coerver et al., 2018). To characterize impacts of reservoir operations on watershed hydrology, multiple methods have
45 been developed to simulate reservoir releases. These models include mathematical tools which optimize water release for
46 achieving management objectives, simulation models which consider physical processes of water cycling in reservoirs to allow
47 users to evaluate impacts of different management alternatives on reservoir storages and releases, and a combination of these
48 two types of models for reservoir planning and management (Branets et al., 2009; Dogrul et al., 2016; Yeh, 1985). Among
49 these models, the RiverWare model and models developed based on RiverWare consider both management policies and
50 physical processes (Zagona et al., 2001), and have proven capability of simulating reservoir storages and downstream flows.
51 However, how reservoir operations affect watershed hydrology is still not explicitly examined.

52 In addition to reservoir operations, cropland irrigation also affects watershed hydrology. Water withdrawal for
53 irrigation has been widely adopted to increase crop production in arid and semi-arid regions. Water redistribution through
54 irrigation enhances water and energy fluxes between soils and the atmosphere (Rost et al., 2008; Sacks et al., 2009), and results
55 in elevated water loss through evapotranspiration (Hao et al., 2015; Malek et al., 2017; Polo and Losada, 2016), and depletion
56 of water resources (Aeschbach-Hertig and Gleeson, 2012) in different regions of the world. To better simulate impacts of
57 irrigation, numerical models have been developed to quantify water fluxes among soils, vegetation, and water bodies induced
58 by irrigation (Leng et al., 2013; Santhi et al., 2005). Impacts of irrigation on watershed hydrology should be further evaluated
59 to application of this tool for effective management of water resources in basins with competing demands for water.

60 The Soil and Water Assessment Tool (SWAT) has been widely used to simulate water cycle dynamics in response to
61 management practices across the watershed and regional scales (Arnold et al., 1998). Previous studies indicated that the default
62 SWAT reservoir operation scheme which simulates water release based on target storages may either overestimate reservoir
63 storages in no-flood seasons (Lv et al., 2016), or underestimate water releases when actual reservoir storages are lower than
64 target storages (Wu and Chen, 2012). SWAT simulates water withdrawal for irrigation from different water sources (e.g.,
65 reservoirs, streams, and groundwater aquifers). Multiple efforts have employed SWAT to evaluate impacts of different
66 irrigation practices on watershed hydrology (Ahmadzadeh et al., 2016; Chen et al., 2017; Maier and Dietrich, 2016), and
67 emphasized the importance of balancing water supply and irrigation demands in hydrologic simulations. However, applicability
68 of SWAT in watersheds with interacting reservoir operations and irrigation has not been well studied, and thus deserves further
69 investigation to inform effective water resource management.

70 The Yakima River Basin (YRB) in the Pacific Northwest of the U.S. has been regulated for regional hydropower,
71 flood control, fishery, crop cultivation, and drinking water supply. Water supply for irrigation is one of the most important
72 water resource management objectives in the YRB (USBR, 2012). The Yakima River Reservoir system supplies water to
73 180,000 hectares of cropland through the operation of five reservoirs which store ca. 30% of the mean annual runoff of the
74 basin (Vano et al., 2010). Reservoir operations and cropland irrigation in the YRB altered historical streamflow regimes,
75 resulted in severe low flow, and elevated flow events. Since the 1990s, increasing demands for irrigation, municipal water
76 consumption, and critical environmental flow for conserving wildlife habitats in the context of climate change have challenged
77 water resource management in the basin. Thus, there is an urgent need to reliably simulate water cycling in the basin to provide
78 a solid basis for policy formulation and management actions which strive to achieve a balance among water demands for
79 different purposes (Poff et al., 2003).

80 In recognition of the challenges in modeling hydrology in heavily managed watersheds, this study investigated impacts
81 of water management on streamflow modeling in the YRB. Using the YRB as a testbed, we evaluated streamflow simulations
82 with different model representations of management activities. Objectives of this study are to (1) examine how different
83 representations of reservoir operations influence watershed streamflow simulations, and (2) assess impacts of cropland
84 irrigation on watershed hydrology. Methods and findings derived from this study hold the promise to provide valuable
85 information for improving hydrologic modeling in intensively managed basins across the globe.

86 **2. Materials and methods**

87 **2.1. Study area**

88

89 [Figure 1]

90 The Yakima River Basin (Figure 1) is located in central Washington, U.S. (45.98 ~47.60° N, 121.53~119.20° W). The basin
 91 has a semi-arid climate with a Mediterranean precipitation pattern. Winters are cold, with a mean temperature of -2.1 °C.
 92 Annual average precipitation is ca. 675 mm, with an average snowfall of 550 mm, occurring mainly in December and January.
 93 Rangeland, forest, and cropland are the primary land uses in the basin, and cover 36%, 33%, and 28% of the study area (Vaccaro
 94 and Olsen, 2007), respectively. Dams were built throughout the basin for the irrigated agriculture. There are five big reservoirs
 95 in the YRB, including Keechelus, Kachess, Cle Elum, Bumping, and Rimrock (Figure 1). Malek et al., (2016) reported that the
 96 YRB experienced major droughts in 20% of the years between 1980 and 2010, and the frequency may double in the future. It
 97 is expected that the increasing competition for water from multiple users, especially for irrigation, fishery, and wildlife habitats,
 98 may escalate in the coming decades (Miles et al., 2000).

99 **2.2. Management Schemes in SWAT and RiverWare model**

100 **2.2.1. Reservoir operation schemes**

101

102

[Table 1]

103 Settings of the five reservoirs, including locations, height, storage capacity, operating purpose, and surface area were compiled
 104 and added to SWAT input files (Table 1). We use three scenarios (R0, R1, and R2) to evaluate reservoir operation simulations
 105 in the YRB. Scenario R0 does not simulate reservoir operations and we use it as a baseline scenario. Scenario R1 uses the
 106 SWAT model's built-in reservoir management schemes which specifies monthly target volumes for managed reservoirs
 107 (Neitsch et al., 2011). Under the R2 scenario, the SWAT model uses reservoir releases calculated by the RiverWare model as
 108 the outflow from these reservoirs to downstream reaches.

109 The SWAT model calculates water balance for a reservoir on a daily scale as follows:

$$110 \quad V_{net} = V_{stored} + V_{flowin} - V_{flowout} + V_{pcp} - V_{evap} - V_{seep} \quad (1)$$

111 where V_{net} is net volume changes of a reservoir on a given day (m^3 water); V_{stored} is the water stored in a reservoir at the
 112 beginning of a day (m^3 water); V_{flowin} is the water entering a reservoir in one day (m^3 water); $V_{flowout}$ is the amount of water
 113 release to downstream reaches of a reservoir (m^3 water); V_{pcp} is the amount of water falling to a reservoir in one day (m^3
 114 water); V_{evap} is the water loss through evaporation from a reservoir (m^3 water); V_{seep} is the amount of water loss through
 115 seepage in a reservoir (m^3 water).

116 Under the R1 scenario, the target release approach calculates reservoir storage using the following equations:

$$117 \quad V_{targ} = V_{em}, \text{ if } mon\ fld, beg < mon < mon\ fld, end \quad (2)$$

118

$$V_{targ} = V_{pr} + \frac{\left(1 - \min\left[\frac{SW}{FC}, 1\right]\right)}{2} \cdot (V_{em} - V_{pr}), \text{ if } mon \leq mon_{fld,beg} \text{ or } mon \geq mon_{fld,end} \quad (3)$$

119

120

121

122

where V_{targ} is the target reservoir storage of a given day (m^3 water); V_{em} is the volume of reservoir for filling to the emergency spillway (m^3 water); mon is the month of the year; $mon_{fld,beg}$ is the beginning month of a flood season; $mon_{fld,end}$ is the ending month of the flood season; V_{pr} is the reservoir volume when filled to the principal spillway (m^3 water); SW is average soil water content (mm) on a given day, and FC is field capacity (mm).

123

124

With the target volume is determined, the reservoir outflow ($V_{swat_flowout}$, m^3/day) in default SWAT for a given day is calculated as follows:

125

$$V_{swat_flowout} = \frac{V_{stored} - V_{targ}}{ND_{targ}} \quad (4)$$

126

127

where V_{stored} is the volume of water stored in the reservoir on a given day; and ND_{targ} is the number of days required for the reservoir to reach the target storage.

128

129

Under the R2 scenario, outflow from a reservoir is calculated based on the estimated daily release provided by the RiverWare model as follows:

130

$$V_{RiverWare_flowout} = 86400 \cdot q_{out} \quad (5)$$

131

132

where $V_{RiverWare_flowout}$ is the volume of water flowing out of a reservoir in one day (m^3) and q_{out} is the outflow rate estimated by RiverWare (m^3/s).

133

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138

RiverWare simulates operations and scheduling of reservoir management objectives, including hydropower production, flood control, and irrigation (Zagona et al., 2001). RiverWare can model a variety of physical processes for reservoirs with computational time steps ranging from one hour to one year. In RiverWare simulations, the solver is based on operating rules or operating policies that provide instructions for operation decisions such as reservoir releases (Zagona et al., 2001). The rules are strictly prioritized, with high priority rules requiring that reservoir release should not be less than the minimum flow for downstream reaches; whereas a low priority rule requires that reservoir storage should fit a seasonal guide

139 curve value. Conflicts are resolved by giving higher priority rules precedence. This model has been applied to the YRB to
140 simulate outflow from the reservoirs (USBR, 2012).

141 **2.2.2. Irrigation representation in the SWAT model**

142 SWAT irrigation schemes consider multiple water sources including reservoirs, streams, shallow aquifers, and sources outside
143 the watershed. Irrigation can be triggered by a water stress threshold (a fraction of potential plant growth). In SWAT, water
144 stress is simulated as a function of actual and potential plant transpiration:

$$145 \quad wstr = 1 - \frac{E_{t,act}}{E_t} = 1 - \frac{w_{actualup}}{E_t} \quad (6)$$

146 where $wstr$ is the water stress; E_t is the potential plant transpiration (mm/day); $E_{t,act}$ is the actual amount of transpiration
147 (mm/day) and $w_{actualup}$ is the total plant water uptake (mm/day). The plant water uptake is a function of the maximum plant
148 transpiration, a water-use distribution parameter, the depth of the soil layer and the depth of plant root. In the SWAT auto
149 irrigation algorithm, irrigation is applied when the water stress factor falls below a predefined threshold. Irrigation will increase
150 soil moisture to field capacity, if irrigation water sources could provide enough water. We conducted two additional simulations
151 by assuming that irrigation water was withdrawn from reservoirs and streams (R2S1), or groundwater (R2S2), based on the
152 simulations with RiverWare reservoir schemes (R2).

153 **2.3. Model setup, sensitivity analyses, and simulations**

154

155 [Table 2]

156

157 We used a plethora of geospatial datasets to parameterize and drive hydrological simulations in the YRB (Table 2). Topography
158 information was derived from U.S. Geological Survey (USGS) National Elevation Dataset (NED) (<https://lta.cr.usgs.gov/NED>)
159 with a spatial resolution of 30 meters. The U.S. Department of Agriculture (USDA) Cropland Data Layer (CDL)
160 (<https://nassgeodata.gmu.edu/CropScape/>) with a spatial resolution of 30 meters was used to obtain land covers including
161 shrubland, forestland, grassland, developed land and barren land, cultivated land and orchard in the YRB (Figure 1). We
162 derived daily climate data for the period of 1980-2012 from North America Land Data Assimilation System (NLDAS)
163 (<https://ldas.gsfc.nasa.gov/nldas/NLDAS2forcing.php>). In addition, we obtained nitrogen and phosphorus fertilizer application
164 rates (USDA-ERS, 2018), tillage intensity (CTIC, 2008), and planting and harvesting (USDA, 2010) for crop management.
165 When defining hydrologic response units (HRUs), we used thresholds of 20%, 10%, and 10% for land use types, soil classes,

166 and slop groups, respectively. The SWAT model divides the YRB into 181 subbasins and 1950 HRUs. Streamflow simulations
167 in four subbasins (Figure 1) with long-term observations were explicitly examined to evaluate how different schemes affected
168 model performances. To evaluate SWAT ET simulations, we compiled the annual Moderate Resolution Imaging
169 Spectroradiometer (MODIS) evapotranspiration (ET) data for the study area. The MODIS ET data were produced using the
170 Penman-Monteith equation and remotely sensed land cover/ Leaf Area Index (LAI) information, with a spatial resolution of 1
171 km (Mu et al., 2011).

172 We quantified parameter sensitivities with a global sensitivity method described by (Abbaspour et al., 2017), which
173 employs model runs driven by randomly sampled parameter sets, a multi-regression approach, and a T-test to identify and rank
174 sensitive parameters. Sensitivity analysis for SWAT simulations in the YRB is computationally expensive. For each scenario,
175 we spent about three weeks to run SWAT 10000 times (Zhang et al., 2009a; Zhang et al., 2009b) to understand parameter
176 sensitivity and minimize the discrepancy between simulations and observations under different scenarios. We used the Nash–
177 Sutcliffe efficiency coefficient (*Ens*) (Nash and Sutcliffe, 1970) and correlation coefficient (*R*) (Legates and McCabe, 1999) as
178 the metrics to evaluate model performance.

179 3. Results

180 3.1. Parameter sensitivity under different scenarios

181

182 [Table 3]

183

184 Table 3 shows the ranking of parameter sensitivity under different scenarios. In general, selected parameters demonstrated
185 similar sensitivities among all scenarios, particularly for the ten most sensitive parameters, indicating snow melting (SMFMX,
186 SFTMP and SMTMP), soil water dynamics (CN2, SOL_k, and SOL_Z), and water routing (CH_N2 and SLSUBBSN) are
187 critical for water cycling in the basin (Tables 3 and S1). For all scenarios, the most sensitive parameters are CN2 and the snow
188 factors, including SFTMP, SMTMP, SMFMX, SMFMN, and TAMP, indicating that snowmelt is the key hydrological process
189 in the YRB. SWAT uses the Soil Conservation Service curve number method (SCS-CN) to predict runoff. As a result, parameter
190 CN2 affects the partition of water between surface runoff and infiltration, and has significant impacts on streamflow estimates.
191 We also observed that sensitivities of several parameters were different among the five scenarios. Specifically, parameters
192 relevant to reservoir operations or irrigation management, including the RES_K and NDTARGR, played important roles in
193 simulations with reservoir operations. The differences could be attributed to the inclusion of reservoir operation and irrigation
194 schemes, and further suggest that significant impacts of the management activities on water cycling should be considered in
195 hydrologic modeling. Note that although the inclusion of management activities altered the sensitivity of reservoir and irrigation

196 related parameter, snow melting and soil water dynamics may still play the fundamental role in water cycling, as evidenced by
197 the high sensitivity of CN2 and SFTMP.

198 **3.2. Streamflow simulations under different reservoir operation scenarios (R0, R1, and R2)**

199

200 [Figure 2]

201 [Figure 3]

202 [Figure 4]

203 Without considering impacts of reservoir operations and water withdrawals on water cycling, the R0 scenario demonstrated
204 poor performance in streamflow simulations (Figure 2, Table S2). Streamflow simulations in R1 and R2 were significantly
205 improved when reservoir operation schemes were added to SWAT, which further confirmed the importance of considering
206 reservoir operations in hydrologic modeling in the YRB. Note that reservoirs either increase or reduce streamflow, as reservoirs
207 could increase water release in dry seasons, or retain upstream water for flood control in wet seasons. In addition, streamflow
208 simulated in the R2 scenario (average correlation coefficient of 0.59) showed a better agreement with measured flow than that
209 of the R1 scenario (average correlation coefficient of 0.57). R2 exhibits better *Ens* in three of the four subbasins than R1 (Table
210 S2), indicating that reservoir outflow estimated by RiverWare more accurately simulated water releases than the default
211 reservoir operation scheme in SWAT. The streamflow simulations in subbasins 67 and 99 were more sensitive to the different
212 reservoir schemes, as evidenced by greater improvements in the *Ens* and *R* values than those of the other two downstream
213 subbasins (Figures 3 and 4).

214 [Figure 5]

215 We also compared ET simulations of the YRB under the three scenarios (R0, R1, and R2). Specifically, ET estimates
216 increased in May and June, but decreased in winter for R1 and R2 simulations (Figure S1). In addition, annual ET increased
217 by 7.83% and 8.05% for R1 and R2 simulations relative to the R0 simulation, respectively (Figure 5). The changes could be
218 attributed to increased evaporation from reservoirs.

219 **3.3. Streamflow and ET simulations under the two irrigation operation scenarios (R2S1 and R2S2)**

220 **3.3.1. Streamflow and ET**

221

222 [Figure 6]

223 [Figure 7]

224 Settings of scenario R2S1, which used reservoirs and streams as water sources for irrigation, are consistent with the actual
225 irrigation practices in the YRB where surface water is the primary irrigation water source (Figure 6). For the R2S2 scenario,
226 shallow groundwater was assumed to be the water source for irrigation (Figure 7). Consequently, streamflow simulations under
227 the scenario R2S1 matched observations better than that in R2S2. Compared with the R2 scenario, the simulated flow decreased
228 by 24.87% and 31.29% in R2S1 and R2S2, respectively.

229 [Figure 8]

230 ET is an important component of terrestrial water cycling and this variable is used in the calculation of irrigation
231 demand in SWAT simulations. Figure 8 compares simulated monthly ET of the irrigation scenarios (R2S1 and R2S2) with the
232 RiverWare reservoir operation scenario (R2) which did not consider irrigation. The mean monthly ET rates of the irrigation
233 scenarios (R2S1 and R2S2) were significantly higher (85% and 63% for R2S1 and R2S2, respectively) than simulations without
234 irrigation, particularly during March-July, when irrigation was applied to support crop growth.

235 [Figure 9]

236 We further compared the simulated annual ET in the R2S1 and R2 scenarios (Figure 9). We observed low cropland
237 ET in the R2 scenario relative to the R2S scenario. Specifically, when irrigation was included in our simulation, SWAT ET
238 estimates increased by ca. 85% at the annual scale. Monthly scale comparison showed that increases in ET mainly occurred in
239 growing seasons (April to August, Figure S2). The comparison demonstrated that inclusion of irrigation schemes achieved
240 better estimates of water losses during irrigation, and contributed to enhancing streamflow simulations (Figure 6). In addition
241 to magnitude, the irrigation scenario (R2S1) also simulated well interannual variability of ET, as evidenced by the high
242 coefficient of determination in the scatter plot against ET estimates based on remote sensing data (Supplementary Material
243 Figure S3).

244 3.3.2. Irrigation water consumption

245 The mean annual irrigation depth for the irrigation scenarios of R2S1 and R2S2 was 480.66 mm/year and 228.46 mm/year,
246 respectively. Under the R2S1 scenario, water for irrigation was provided by the five reservoirs in the corresponding subbasins;
247 in subbasins without reservoirs, irrigation water was withdrawn from local streams. Average irrigation water was higher in the
248 R2S1 scenario than that of R2S2. There are notable differences in irrigation depths for different crop species between the two
249 irrigation scenarios. In general, the irrigation water consumption for all crops was higher in the R2S1 scenario than that of the
250 R2S2 scenario.

251 3.4. Management impacts on watershed hydrology

252 As indicated by the improved *Ens* and *R* values, streamflow simulations under scenarios simulating both reservoir operations
253 and irrigation schemes (R2S1 and R2S2) are more comparable with observations than those of the baseline scenario (R0) which

254 does not consider water management activities in the simulation. Reservoirs have contributed to streamflow increases in dry
255 periods and streamflow reduction in wet seasons by regulating water storage and release. Compared with the baseline scenario
256 (R0), we found reductions in simulated streamflow in the scenarios that consider reservoir and irrigation operations, indicating
257 that water withdrawal for irrigation tends to reduce streamflow as a result of enhanced water loss through ET.

258 ET in the composite scenarios (R2S1 and R2S2) was higher than the R0 scenario, which can be attributed to the
259 elevated evaporation from reservoirs and irrigated cropland. Direct evaporation from reservoirs increased by 7% - 8% over the
260 study period (1980 to 2010) due to improved simulation of reservoir surface areas in the R1 and R2 simulations relative to the
261 R0 simulation. Irrigation practices led to more pronounced increases in ET in R2S1 and R2S2 simulations as compared with
262 that of R2 (Figure 8). These results indicate that irrigation may have more pronounced impacts on ET through stimulating ET
263 than reservoir operations in the study area.

264 **4. Discussion**

265 **4.1. SWAT simulation of water cycling in response to management activities**

266 In recent decades, water users of the YRB passed the Yakima River Integrated Water Management Plan, which is a
267 comprehensive agreement that advances water infrastructures and management (USBR, 2012). Enhanced hydrologic modeling
268 provided by this study will provide valuable information for goals of the Integrated Plan, which requires accurate streamflow
269 information to manage water resources to meet ecological objectives as well as to secure water supply for domestic uses.

270 Although previous investigations highlighted the importance of irrigation and reservoir management to water balance
271 and availability (Hillman et al., 2012; Malek et al., 2014), joint impacts of these two water management practices on watershed
272 hydrology have not been fully understood. In recognition of this challenge, we enhanced SWAT representations of the two
273 critical water management activities, including reservoir operations and irrigation, to constrain uncertainties in hydrologic
274 simulations. We achieved improved model performances through including the two activities in the SWAT modeling
275 framework. The simulated streamflow was generally lower in simulations with management activities than the baseline
276 simulation (R0). Without including reservoir management and irrigation, SWAT may overestimate streamflow due to the
277 unreasonably estimated water loss through ET.

278 Water management activities have altered natural hydrological cycling and posed challenges to reliable simulation of
279 watershed hydrology. The YRB is a typical watershed that is regulated to support agricultural production. Maintaining
280 sustainable water supply in basins like the YRB calls for sound understanding of hydrological impacts of management activities.
281 Management schemes developed and evaluated in this study will be transferable and applicable to future SWAT and other
282 watershed models applications for investigating water cycling that is influenced by reservoir operations and water withdrawal
283 for irrigation across broader spatial scales.

284 **4.2. Water cycling under reservoir operation scenarios**

285 Reservoir operations have both direct and indirect impacts on streamflow. Water release from reservoirs directly affects the
286 magnitude and variability of streamflow in downstream reaches. Dam and water diversion operations determine the amount
287 and timing of water discharge to downstream river channels. As a result, reservoir operations may either attenuate flood peaks
288 in wet seasons, or increase streamflow in dry years in compliance with minimum instream flow policies (Yoder et al., 2017).
289 In addition, multiple hydrological processes, such as vertical flow in surface or subsurface waters, water routing, evaporation,
290 precipitation and microclimate, are also responsive to reservoir operations (Lv et al., 2016). Our simulations suggested that
291 reservoir operations altered both streamflow and ET in the YRB.

292 Most precipitation in the YRB occurs in winter as snowfall. Snowpack serves as a water reservoir for spring and
293 summer streamflow. Consequently, streamflow is high in spring but low in summer. As shown in Table 1, most of the reservoirs
294 were built to support cropland irrigation. Presence of reservoirs positively contributed to water availability in dry periods.
295 Water storage management in reservoirs is one adaptation strategy particularly applicable to snowmelt-dominant watersheds
296 like the YRB which experiences water scarcity during the summer irrigation season (Yoder et al., 2017), and thus alters natural
297 flow regimes. Without representing reservoir regulations, SWAT simulations failed to reasonably reconstruct temporal
298 variability in streamflow (R0 scenario). Results of this study indicated that reservoir algorithms based on RiverWare (R2) were
299 relatively more realistic compared with the default reservoir operation algorithms in SWAT (R1), as evidenced by the improved
300 model performances. Enhanced model performances in the R1 and R2 scenarios further corroborated the significant impacts
301 of reservoir operations on seasonal patterns of streamflow (Adam et al., 2007).

302 Compared with the baseline scenario (R0), R1 and R2 simulations showed that the ET rates increased considerably
303 from April to September due to reservoir operation. Direct evaporation from reservoirs increased under the R1 and R2 scenarios
304 because of improved estimates of reservoir surface areas. Consideration of such an impact on ET in the R1 and R2 scenarios
305 also contributed to enhanced model performances relative to the baseline scenario (R0).

306 **4.3. Impacts of irrigation on water cycling**

307 Water withdrawal for irrigation has increased pressures on maintaining sustainable water resources in the YRB (Malek et al.,
308 2017). Insufficient water supply for agricultural production, drinking water supply, and environmental flows has raised
309 concerns on the local economy and ecosystem integrity (Hillman et al., 2012). Due to the significant impacts on soil moisture
310 and plant growth, the amount and timing of irrigation have influences on ET losses and watershed hydrology (Maier and
311 Dietrich, 2016). As a result, the irrigation impacts on streamflow should be evaluated to provide reliable estimates of streamflow
312 in basins like the YRB to help balance the water supplies and demands for effective water resource management.

313 As reported in previous studies, most of the water for agricultural irrigation was provided by surface water and one-
314 third was from groundwater in the YRB (USBR, 2012). Under the R2S1 scenario, our assumption that irrigation water was

315 from the reservoirs and streams generally agreed with the actual water uses for irrigation in the basin. The less satisfactory
316 model performances in the R2S2 scenario may stem from the unrealistic assumption of water sources, irrigation efficiencies,
317 and return flow of irrigation. In addition, SWAT simulates streamflow based on water balance among multiple water pools,
318 including shallow groundwater which is recharged by subsurface runoff (Shadkam et al., 2016). Under the R2S2 scenario,
319 water withdrawal from the shallow renewable groundwater was used in our simulation. This simplification did not consider
320 water withdrawal from deep nonrenewable aquifers. As a result, water availability based on shallow groundwater for irrigation
321 and groundwater recharge, may have been unreasonably estimated, and partially contributed to unsatisfactory model
322 performances under this scenario (R2S2).

323 To better investigate hydrological consequences of water management, future studies should further constrain
324 uncertainties in streamflow simulations by incorporating additional reservoir management and irrigation information. Including
325 of observed reservoir release will help improve model representations water discharge from reservoirs. In addition, model
326 representation of irrigation should be improved in the future. Note that model performances of the R2S1 scenario were not
327 substantially improved relative to the R2 scenario. The irrigation operation scheme that used surface water as the single source
328 may have introduced uncertainties to streamflow simulations, since groundwater is also an important water source for irrigation,
329 particularly in dry years in the YRB. Future simulations need to incorporate explicit irrigation information about irrigated areas,
330 the source, amount, and timing of groundwater withdrawals into hydrologic modeling to better simulate agricultural hydrology.
331 We observed different seasonal patterns of ET under the five scenarios. How management activities affected water and energy
332 exchanges between soil and the atmosphere should also be investigated in the future.

333 As most reservoirs were built for irrigation in the YRB, impacts of reservoirs should be assessed jointly with the
334 accelerating development of irrigated agriculture in the basin. Presence of reservoirs positively contributed to water availability
335 for irrigation, particularly for dry seasons. In general, the combination of reservoir operations and irrigation have reduced
336 streamflow in the YRB when compared with the baseline scenario (R0). This is attributable to the large amounts of water loss
337 through ET in irrigation and additional water storage in reservoirs.

338 **4.4. Caveats in model selection**

339 Among the multiple modeling scenarios, we found that linking RiverWare reservoir model with SWAT achieved
340 better performance than those model structures that rely on simplified reservoir operations, as evidenced by the relatively
341 higher correlation coefficient and *Ens*. However, it is worth noting that these statistical metrics are calculated based on a limited
342 set of hydrological variables (e.g. streamflow), but cannot guarantee other hydrological processes are well represented (Zhang
343 et al. 2013). Therefore, we further used MODIS estimated ET and reported irrigation water demand data to justify the favorable
344 performance of the combined SWAT-RiverWare watershed model configuration.

345 Our model evaluation process follows the widely accepted procedures for model calibration and evaluation (Moriassi
346 et al. 2007; Arnold et al. 2012). We also would like to point out that the complexity difference between the SWAT-RiverWare
347 and other watershed model configurations was not explicitly considered in model evaluation. Previous research notes that
348 model complexity is an important factor in selecting the most robust model configuration that can fulfill a specific purpose.
349 For example, Höge et al. (2018) reviewed existing methods and laid the foundation for a comprehensive framework for
350 understanding the critical role of model complexity in model selection. The lack of reliable prior knowledge of the model
351 structure and associated model parameters makes it difficult to directly consider model complexity here. However, the
352 framework laid out by Höge et al. (2018) deserve further exploration in comparing the performance of different watershed
353 model configurations in the future.

354 **5. Conclusions**

355 Reservoir operations and irrigation have substantial impacts on water cycling globally. Hydrologic simulation in the managed
356 basins faces challenges in reliably characterizing water management activities. This study assessed the hydrological impacts
357 of reservoir systems and irrigation practices through numerical model experiments with SWAT. Reservoir operation
358 representations by coupling the RiverWare model and SWAT significantly improved streamflow simulations. We achieved
359 reasonable model performances in the scenario using reservoirs and streams as the water sources for irrigation, since these
360 assumptions are consistent with the actual irrigation practices in the basin. Model simulations suggested that reservoir
361 operations and irrigation water withdrawal generally reduced streamflow by enhancing water loss through ET in the study
362 area. Results of this study demonstrated importance of incorporating water management activities into hydrologic modeling.
363 Both SWAT and RiverWare are community models that have been widely tested and applied in diverse regions across the
364 globe, as evidenced by the numerous peer-reviewed publications in the fields of reservoir operation and watershed modeling
365 (https://www.card.iastate.edu/swat_articles/). The knowledge discovered through our numerical experiments is
366 expected to help understand uncertainties in water cycling simulations resulted from water management representations in
367 hydrological models. Methods and findings derived from this study are expected to help enhance future hydrologic modeling
368 in managed watersheds with intensive reservoir and irrigation activities.

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Tables and Figures

Table 1 Reservoir information of the YRB's five reservoirs (Locations are marked in Figure 1).

Reservoir name	River	Completion year	Dam height (m)	Active Capacity (10^6 m^3)	Surface area (km^2)
Bumping	Bumping River	1909	19	42	5.3
Keechelus	Yakima River	1916	39	195	12.8
Kachess	Kachess River	1911	35	295	18.6
Cle Elum	Cle Elum River	1932	50	539	19.5
Rimrock	Tieton River	1924	97	244	10.2

Table 2 Dataset used in the SWAT simulations.

Data type	Spatial/Temporal Resolution/scale	Data description
Topography	30 m	Elevation
Land use	30 m	Land use classifications
Soils	1:250,000	Soil physical and chemical properties
Weather	Daily data in a one-eighth grid resolution	Precipitation, maximum and minimum air temperature, relative humidity, wind speed, and solar radiation.
Hydrological data	Daily	Streamflow
Dam	N/A	Locations, completion year, height, normal and maximal storage capacity, operating purpose, and surface area

Table 3 Parameter sensitivity analysis under various scenarios.

Parameters	Description	Lower limit	Upper limit	³ Parameter modification	Sensitivity rank ¹ of five scenarios ²				
					R0	R1	R2	R2S1	R2S2
SFTMP	Snowfall temperature (°C)	-20	20	V	2	2	2	14	2
CN2	Initial SCS runoff curve number for moisture condition	-0.9	1.2	R	1	1	1	1	1
SMFMX	Maximum melt rate for snow during year (occurs on summer solstice) (mm H ₂ O/°C/day)	0	20	V	4	5	7	24	6
SMTMP	Snow melt base temperature (°C)	-20	20	V	5	3	3	18	4
CH_N2	Manning's "n" value for the main channel	0	0.30	V	7	16	5	19	11
SMFMN	Minimum melt rate for snow during the year (occurs on winter solstice) (mm H ₂ O/°C/day)	0	20	V	15	13	28	17	15
SLSUBBSN	Average slope length (m)	10	150	V	3	6	4	2	3
CH_N1	Manning's "n" value for the tributary channels	0.01	30	V	23	23	17	22	25
SOL_K	Saturated hydraulic conductivity (mm/hr)	-0.8	0.8	R	8	12	8	3	7
GW_REVAP	Groundwater "revap" coefficient	0.02	0.20	V	14	18	12	13	14
CANMX	Maximum canopy storage (mm H ₂ O)	0	100	V	26	25	19	27	28
HRU_SLP	Average slope steepness (m/m)	0	1	V	16	10	23	6	19
RES_K	Hydraulic conductivity of the reservoir bottom (mm/hr)	0	1	V	11	11	26	4	22
GW_DELAY	Groundwater delay (days)	0	500	V	12	19	18	25	9
EVRSV	Lake evaporation coefficient	0	1	V	17	8	20	12	18
TIMP	Snow pack temperature lag factor	0	1	V	27	27	16	28	24
ESCO	Soil evaporation compensation coefficient	0	1	V	24	15	24	15	23

Table 3
(continued)

GWQMN	Threshold water level in the shallow aquifer for the base flow (mm)	0	5000	V	22	20	15	16	27
PLAPS	Precipitation lapse rate (mm H ₂ O/km)	-10	10	R	21	7	6	8	13
OV_N	Manning's "n" value for overland flow	0.01	30	V	9	24	22	11	8
REVAPMN	Threshold depth of water in the shallow aquifer for "revap" to occur (mm)	0	500	V	25	26	21	21	26
SOL_AWC	Available water capacity of the soil layer (mm H ₂ O/mm soil)	0	1	V	28	14	27	23	16
NDTARGR	Number of days to reach target storage from current reservoir storage	1	200	V	13	22	11	9	20
ALPHA_BF	Baseflow alpha factor (1/day)	0	1	V	20	21	14	10	17
SOL_Z	Depth from soil surface to the bottom of the layer (mm)	-1	1	R	6	9	9	5	5
TLAPS	Temperature lapse rate (°C/km)	-10	10	R	19	4	13	7	21
SURLAG	Surface runoff lag coefficient	0.05	24	V	18	28	25	26	10
EPCO	Plant uptake compensation factor	0	1	V	10	17	10	20	12

¹ The sensitive parameters were identified using the Global sensitivity analysis method (Abbaspour, 2007).

² R0 represents the scenario without any reservoir operations; R1 represents the scenario that used the target release approach for the simulation of reservoir outflow in the SWAT model; R2 represents the scenario that used the output of RiverWare model as the daily outflow of the five reservoirs in the SWAT model; R2S1 represents the scenario with irrigation operation that withdraws water from the reservoirs and streams based on the R2 scenario; R2S2 represents the scenario using groundwater as the water source for irrigation based on the R2 scenario.

³ This column indicates how parameters were modified in calibration. V indicates that existing values were replaced with values in the provided range; R indicates relative changes in parameters by multiplying existing values with (1+ calibrated parameter values in the range).

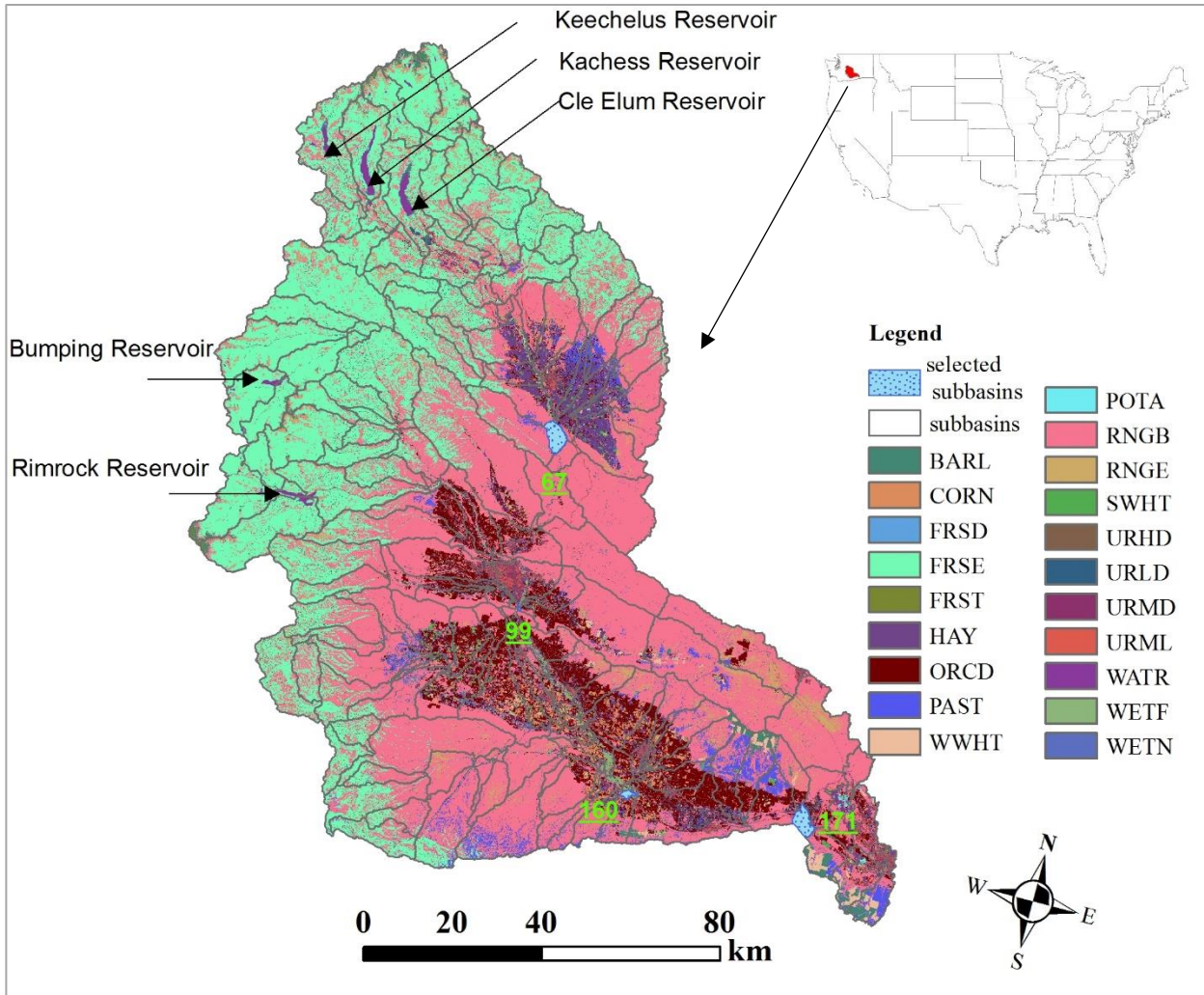


Figure 1 Location and land use of the Yakima River Basin (67, 99, 160, and 171 are subbasins used for streamflow calibration and validation. BARL: Spring Barley; CORN: Corn; FRSD: Deciduous forest; FRSE: Evergreen forest; FRST: Mixed forest; HAY: Hay; ORCD: Orchard; PAST: Pasture; POTA: Potato; RNGB: Range-bush; RNGE: Range-grasses; SWHT: Spring wheat; URHD: Residential-high Density; URLD: Residential-Low Density; URMD: Residential-Medium Density; WATER: Water; WETF: Wetland-forested; WETN: Wetland-non-forested; WWHT: Winter wheat).

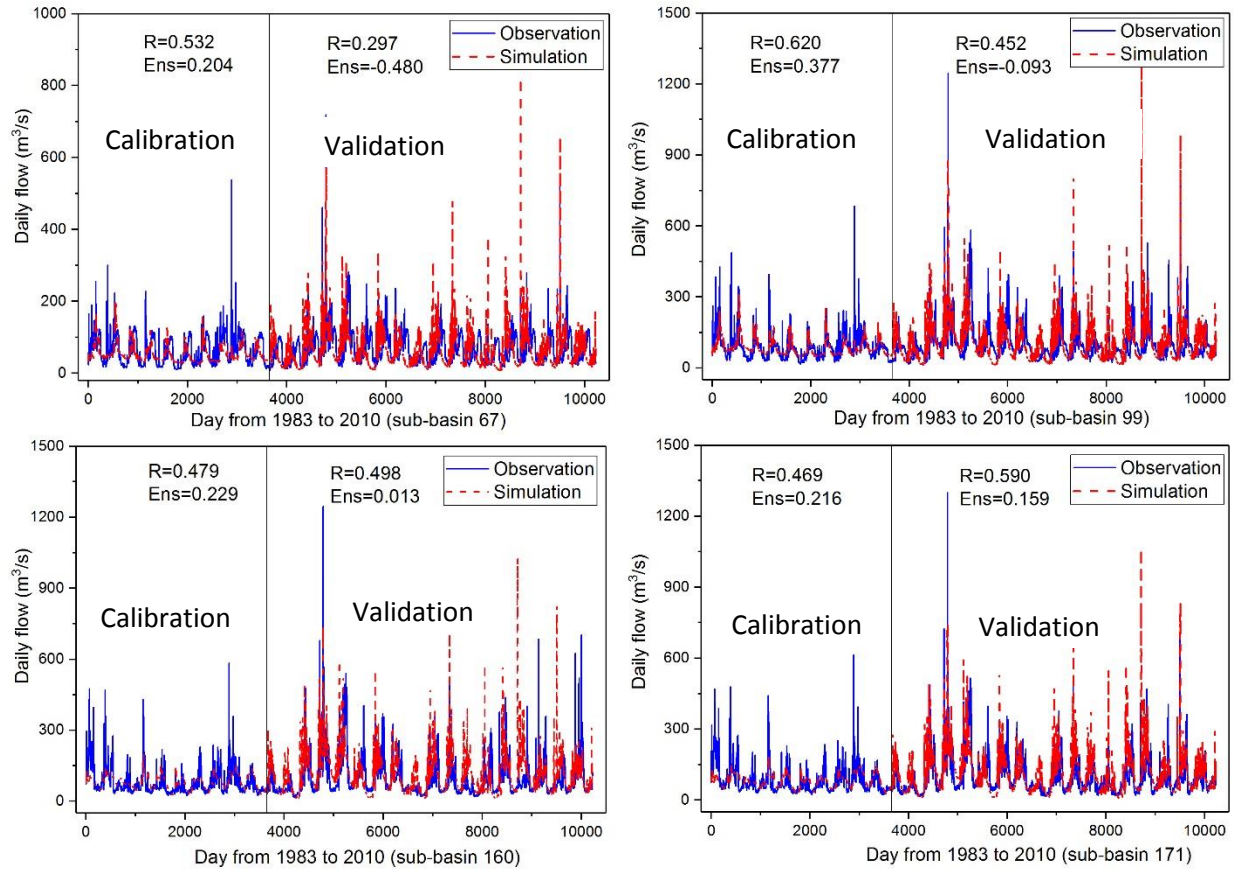


Figure 2 Calibration and validation results in four subbasins under the R0 scenario (baseline simulation does not consider management activities).

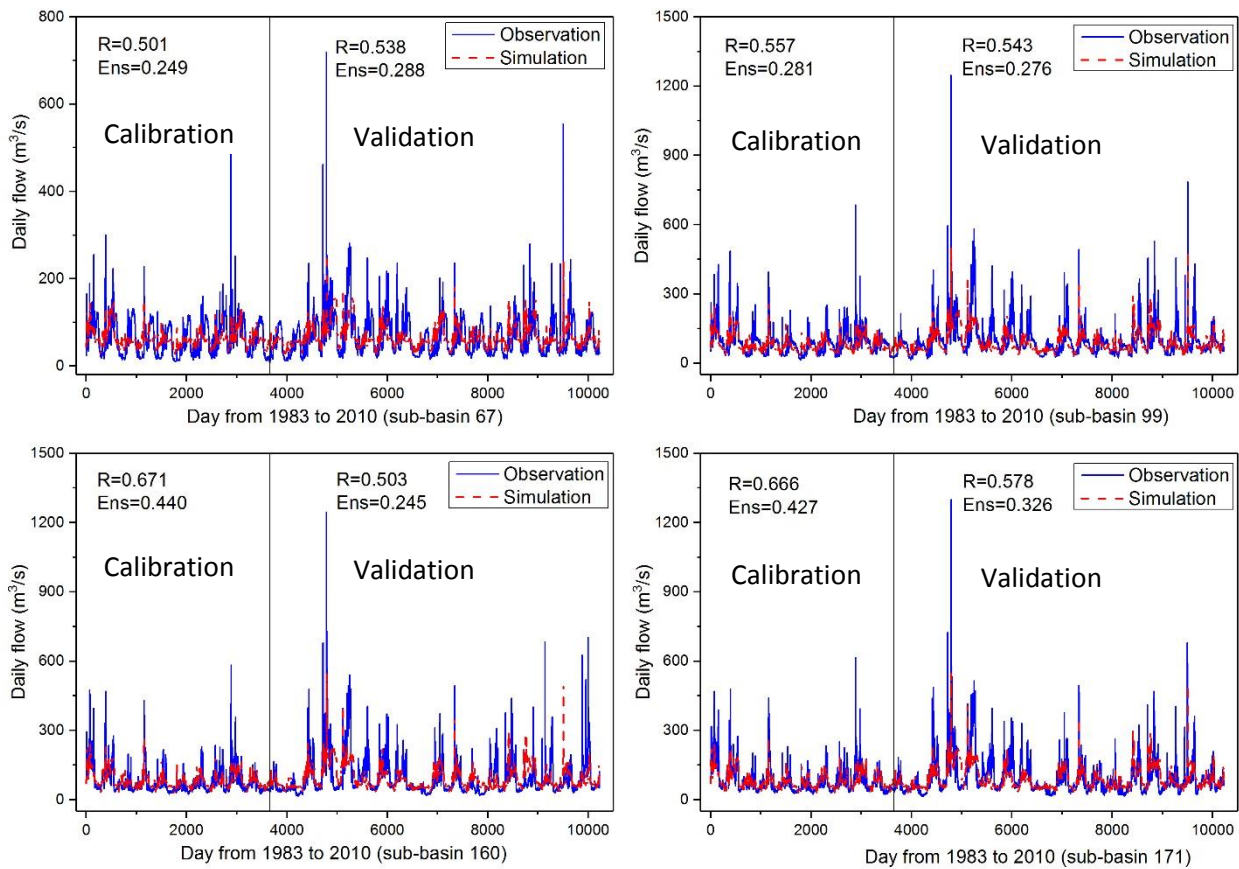


Figure 3 Calibration and validation results under the R1 scenario (Default SWAT schemes for reservoir operations)

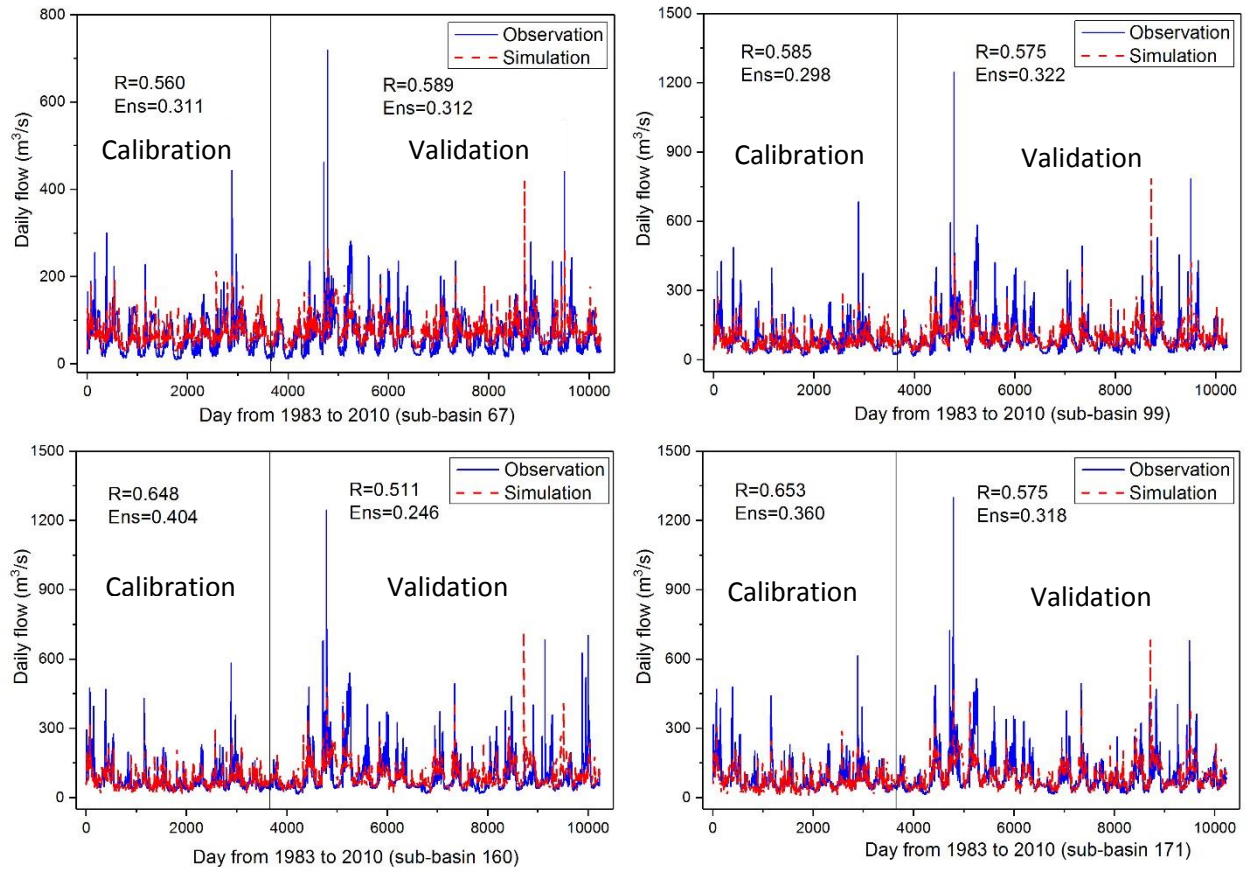


Figure 4 Calibration and validation results under the R2 scenario (RiverWare for reservoir operations).

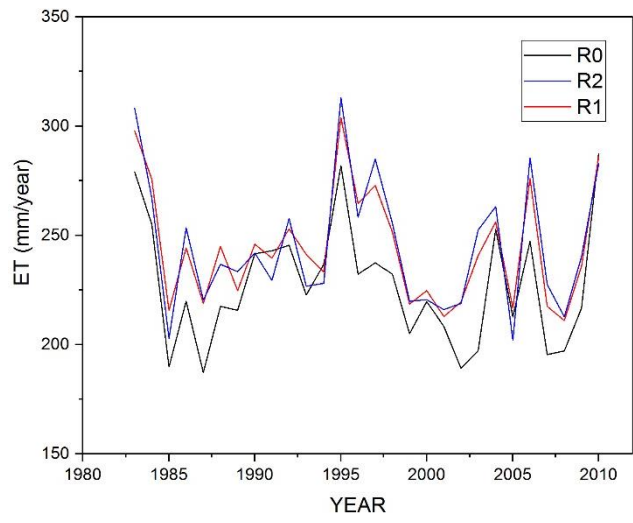


Figure 5 Annual ET simulated under reservoir operation only scenarios (R0, R1, and R2).

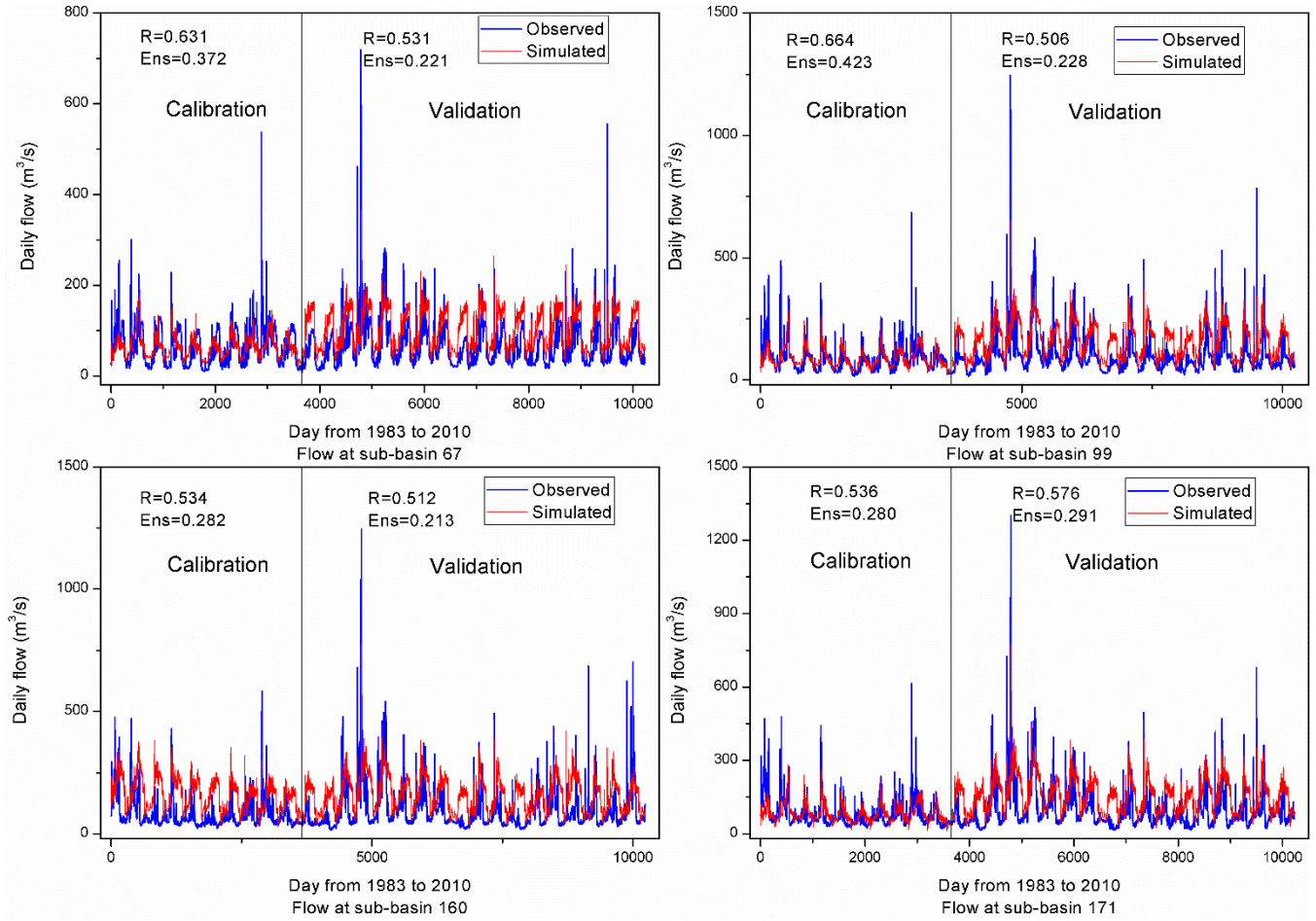


Figure 6 Calibration and validation results under the R2S1 scenario (RiverWare for reservoir operation and surface water as the water source for irrigation)

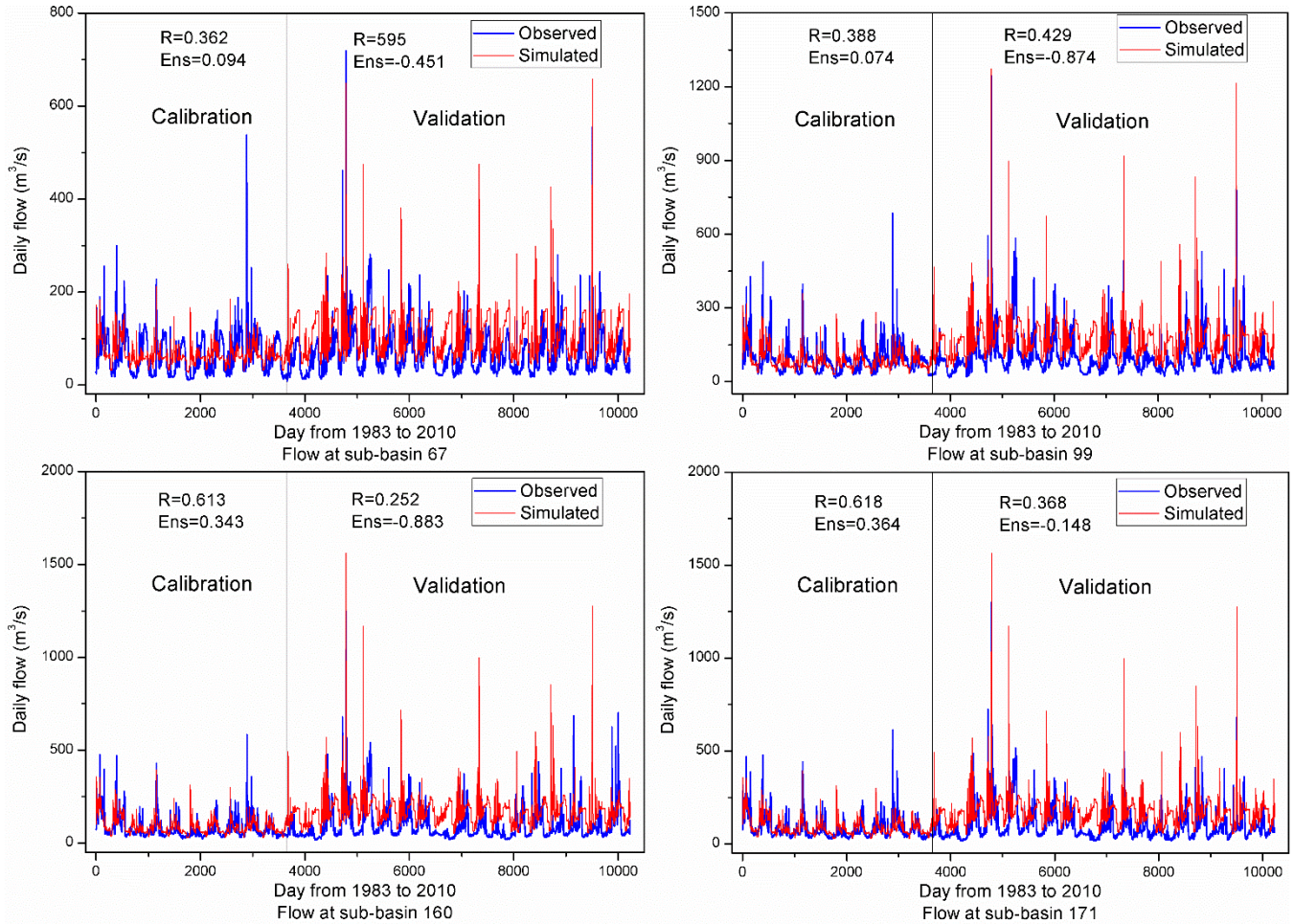


Figure 7 Calibration and validation results under the R2S2 scenario (RiverWare for reservoir operation and groundwater as the water source for irrigation)

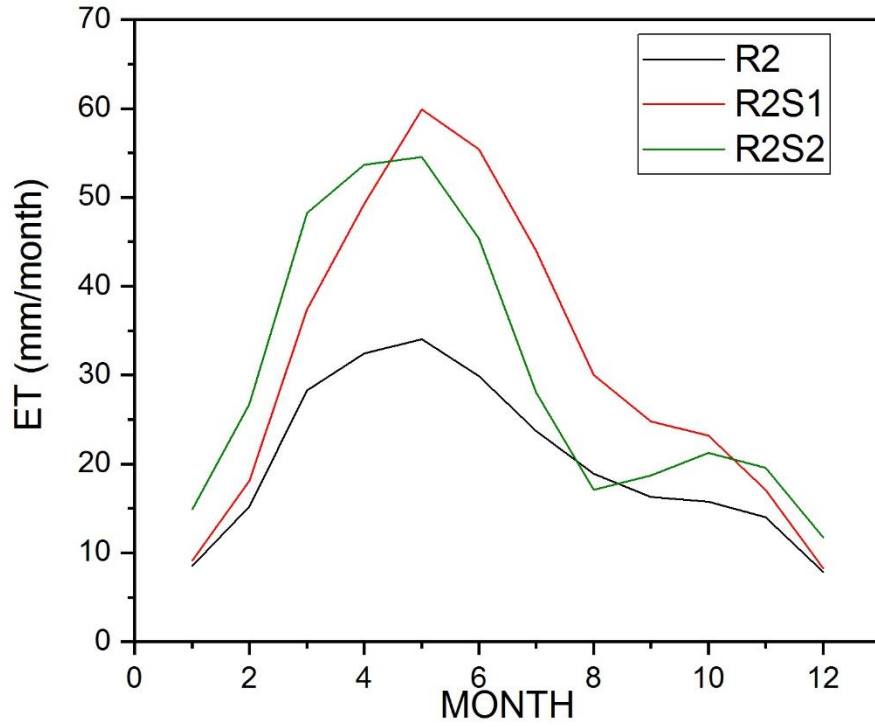


Figure 8 Monthly ET simulated under the irrigation operation scenarios (R2S1 and R2S2) relative to the reservoir operation-only scenario (R2).

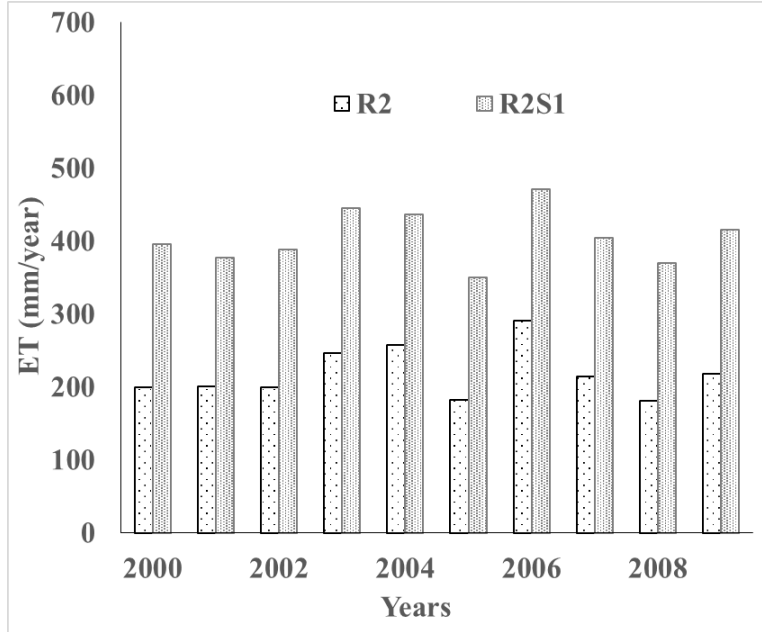


Figure 9 Comparison of ET simulations for cropland during 2000-2009 under the R2 and R2S1 scenarios