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2 **Evaluation of Lacustrine Groundwater Discharge, Hydrologic**
3 **Partitioning, and Nutrient Budgets in a Proglacial Lake in**
4 **Qinghai-Tibet Plateau: Using ²²²Rn and Stable Isotopes**

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32 **Abstract**

33 Proglacial lakes are good natural laboratories to investigate groundwater and
34 glacier dynamics under current climate condition and to explore biogeochemical
35 cycling under pristine lake status. This study conducted a series of investigations of
36 ^{222}Rn , stable isotopes, nutrients and other hydrogeochemical parameters in Ximen Co
37 Lake, a remote proglacial lake in the east of Qinghai-Tibet Plateau (QTP). A radon
38 mass balance model was used to quantify the lacustrine groundwater discharge (LGD)
39 of the lake, leading to an LGD estimate of $10.3 \pm 8.2 \text{ mm d}^{-1}$. Based on the three end
40 member models of stable ^{18}O and Cl^- , the hydrologic partitioning of the lake is
41 obtained, which shows that groundwater discharge only accounts for 7.0 % of the
42 total water input. The groundwater derived DIN and DIP loadings constitute 42.9 %
43 and 5.5 % of the total nutrient loading to the lakes, indicating the significance of LGD
44 in delivering disproportionate DIN into the lake. This study presents the first attempts
45 to evaluate the LGD and hydrologic partitioning in the glacial lake by coupling
46 radioactive and stable isotopic approaches and the findings advance the understanding
47 of nutrient budgets in the proglacial lakes of QTP. The study is also instructional in
48 revealing the hydrogeochemical processes in proglacial lakes elsewhere.

49 **Keywords:** Proglacial lake; ^{222}Rn ; lacustrine groundwater discharge; hydrologic
50 partitioning; nutrient budgets.

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58 1. Introduction

59 High altitude and latitude areas are intensively influenced by the melting of
60 glaciers due to climatic warming. Of particular importance are the proglacial areas,
61 such as proglacial lakes and moraines, because they are particularly affected by
62 climatic change induced glacier retreating and thawing of permafrost (Barry 2006,
63 Heckmann et al. 2015, Slaymaker 2011). The proglacial lakes are usually located
64 close to ice front of a glacier, ice cap or ice sheet, with the vicinity to the ice front
65 sometimes defined as the areas with subrecent moraines and formed by the last
66 significant glacier advances at the end of the Little Ice Age (Barry 2006, Harris et al.

67 2009, Heckmann et al. 2015, Slaymaker 2011). These lakes are located in the
68 transition zones from glacial to non-glacial conditions, and can serve as natural
69 laboratories to explore hydrological processes, biogeochemical cycles and
70 geomorphic dynamics under current climatic conditions (Dimova et al. 2015,
71 Heckmann et al. 2015).

72 The Qinghai-Tibet Plateau (QTP), the third pole of the world, serves as the water
73 tower of most of the major rivers in Asian (Qiu 2008). Unique landscapes such as endorheric
74 lakes, permafrost, glaciers, and headwater fluvial networks are developed due to the intensive
75 interaction between the atmosphere, hydrosphere, biosphere and cryosphere (Lei et al. 2017,

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80 Yao et al. 2013, Yao et al. 2012, Zhang et al. 2017a, Zhang et al. 2017b). Distributed
81 mountainous glaciers and lakes are the most representative landscapes and are highly
82 sensitive to the climate changes. In the past decade, the lakes in the interior of the QTP show
83 overall expanding with respect to an overall increase of precipitation, accelerated glacier
84 melting and permafrost degradation (Heckmann et al. 2016, Yang et al. 2014, Zhang et al.
85 2017a, Zhang et al. 2017b, Zhang et al. 2013). Some latest studies have made effects to depict
86 the hydrologic partitioning of the majority of the lakes in the QTP based on long term
87 observation of climatological parameters, and remote sensing approaches. However, so far a
88 quantitative evaluation of the water balance and hydrologic partitioning, especially the
89 groundwater component of the lakes in the QTP is limited due to the scarcity of observational
90 data. Therefore, there is a great need to conduct refined and systematical field observation to
91 provide groundtruth dataset and tenable models to depict the water balance and hydrologic
92 partitioning of the lakes, especially proglacial lakes in the QTP (Yang et al. 2014, Zhang et al.
93 2017b).

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94 Mountainous proglacial lakes, formed by glacial erosion and filled by melting
95 glaciers, are widely distributed in the Qinghai-Tibet Plateau (QTP), especially along
96 the substantial glacier retreating areas of Himalaya Mountains (MT.), Qilian MT.,
97 Tienshan MT., etc. Characterized by higher elevations, small surface areas but
98 relatively large depths, mountainous proglacial lakes in QTP lack systematic

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105 field-based hydrological studies due to their remote locations and difficulty in
106 conducting field work (Bolch et al. 2012, Farinotti et al. 2015, Yao et al. 2012).

107 There has been extensive recognition of the importance of groundwater discharge
108 to various aquatic systems for decades (Dimova and Burnett 2011, Johannes 1980,
109 Valiela et al. 1978). Very recently, the topic of ‘lacustrine groundwater discharge
110 (LGD)’, which is comprehensively defined as groundwater exfiltration from lake
111 shore aquifers to lakes (Blume et al. 2013, Lewandowski et al. 2015, Lewandowski et
112 al. 2013, Rosenberry et al. 2015), has been introduced. LGD is analogous of in
113 submarine groundwater discharge (SGD) in coastal environments. LGD plays a vital
114 role in lake hydrologic partitioning, which is defined as the separation of groundwater
115 discharge/exfiltration, riverine inflow, riverine outflow infiltration, surface
116 evaporation and precipitation for the hydrological cycle of the lake (Good et al. 2015,
117 Luo et al. 2017). LGD also serves as an importance component in delivering solutes
118 to lakes since groundwater is usually concentrated in nutrients, CH₄, dissolved
119 inorganic/organic carbon (DIC/DOC) and other geochemical components (Belanger et
120 al. 1985, Dimova et al. 2015, Lecher et al. 2015, Paytan et al. 2015). Nutrients and
121 carbon loading from groundwater greatly influences ratios of dissolved inorganic
122 nitrogen (DIN) to dissolved inorganic phosphate (DIP) (referred as N: P ratios
123 thereafter), ecosystem structure and the primary productivity of the lake aquatic

124 system (Belanger et al. 1985, Hagerthey and Kerfoot 1998, Nakayama and Watanabe
125 2008).

126 LGD studies utilize various methods including direct seepage meters (Lee 1977,
127 Shaw and Prepas 1990), geo-tracers such as radionuclides, stable ^2H and ^{18}O isotopes
128 (Gat 1995, Kluge et al. 2007, Kraemer 2005, Lazar et al. 2008), heat and temperature
129 (Liu et al. 2015, Sebok et al. 2013), numerical modeling (Smerdon et al. 2007, Winter
130 1999, Zlotnik et al. 2009, Zlotnik et al. 2010) and remote sensing (Anderson et al.
131 2013, Lewandowski et al. 2013, Wilson and Rocha 2016). Recently, some researchers
132 started to investigate groundwater dynamics in peri- and proglacial areas, mostly
133 based on the approaches of numerical modeling (Andermann et al. 2012, Lemieux et
134 al. 2008a, Lemieux et al. 2008b, Lemieux et al. 2008c, Scheidegger and Bense 2014).
135 However, the quantification of groundwater and surface water exchange in proglacial
136 lakes is still challenging due to limited hydrogeological data and extremely seasonal
137 variability of aquifer permeability (Callegary et al. 2013, Dimova et al. 2015, Xin et al.
138 2013).

139 ^{222}Rn , a naturally occurring inert gas nuclide highly concentrated in groundwater,
140 can be more applicable in fresh aquatic systems and has been widely used as a tracer
141 to quantify groundwater discharge in fresh water lakes (Corbett et al. 1997, Dimova et
142 al. 2015, Dimova and Burnett 2011, Dimova et al. 2013, Kluge et al. 2007, Kluge et al.

143 2012, Luo et al. 2016, Schmidt et al. 2010) and terrestrial rivers and streams
144 (Batlle-Aguilar et al. 2014, Burnett et al. 2010, Cook et al. 2006, Cook et al. 2003). Of
145 particular interest are investigations of temporal ^{222}Rn distribution in lakes, since it
146 can be used to quantify groundwater discharge and reflect the locally climatological
147 dynamics (Dimova and Burnett 2011, Luo et al. 2016). Temporal radon variations
148 give high resolution estimates of groundwater discharge to lakes over diel cycles,
149 allowing evaluation of LGD and the associated chemical loadings. However, there has
150 been no study of radon-based groundwater discharge in mountainous proglacial lakes,
151 especially for those lakes in the QTP.

152 This study aims to investigate the groundwater surface water interactions for the
153 proglacial lake of Ximen Co, by estimating the LGD and evaluating the hydrologic
154 partitioning of the lake. LGD is estimated with ^{222}Rn mass balance model, and the
155 hydrologic partitioning of the lake is obtained with the three endmember model
156 coupling the mass balance of water, stable isotopes and Cl^- . ~~Moreover,~~ LGD derived
157 nutrients are estimated and the nutrient budgets of the lake are depicted. ~~This study,~~
158 to our knowledge, makes the first attempt to quantify the LGD, hydrologic partition,
159 and groundwater borne nutrients of the proglacial lake in QTP and elsewhere via the
160 approach integrating multiple tracers. This study provides insights of hydrologic
161 partitioning in a typical mountainous proglacial lake under current climate condition

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the lake water is calculated based on the
nutrient budgets.

166 and reveals groundwater borne chemical loadings in this proglacial lake in QTP and
167 elsewhere.

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169 **2. Methodology**

170 2.1 Site descriptions

171 The Nianbaoyeze MT., located at the eastern margin of the QTP and being the
172 easternmost part of NW-SW trending Bayan Har Shan, is situated at the main water
173 divide of the upper reaches of Yellow River and Yangtze River (Figure 1). With a peak
174 elevation of 5369 m, the mountain rises about 500-800 m above the surrounding
175 peneplain and displays typical Pleistocene glacial landscapes such as moraines,
176 U-shaped valleys and cirques (Lehmkuhl 1998, Schlutz and Lehmkuhl 2009,
177 Wischnewski et al. 2014). The present snow line is estimated to be at an elevation of
178 5100 m (some updated references) (Schlutz and Lehmkuhl 2009). Controlled by the
179 South Asia and East Asia monsoons, the mountain has an annual precipitation of 975
180 mm in the southern part and 582 mm in the northwestern part, with 80 % occurring
181 during May and October (Yuan et al. 2014, Zhang and Mischke 2009). The average
182 temperature gradient is about 0.55 °C per 100 m, and the closest weather station,
183 locating in Jiuzhi town (N: 33.424614 °, E: 101.485998) at the lower plains of the
184 mountain, recorded a mean annual temperature of 0.1 °C. Snowfalls occur in nearly

185 10 months of the entire year and there is no free-frost all year around (Böhner 1996,
186 2006, Schlutz and Lehmkuhl 2009). The precipitation, daily bin-averaged wind speed
187 and temperature in Aug, 2015 were recorded to be 90 mm, 0.7 m s⁻¹ and 9.5 °C from
188 Jiuzhi weather station (Figure 2). The water surface evaporation was recorded to be
189 1429.8 mm in 2015 from Jiuzhi weather station.

190 Among the numerous proglacial lakes developed in the U-shaped valleys of the
191 Nianbaoyeze MT., Ximen Co lake is located at the northern margin of the mountain
192 with an elevation of 4030 m asl, and is well studied and easily accessible (Lehmkuhl
193 1998, Schlutz and Lehmkuhl 2009, Yuan et al. 2014, Zhang and Mischke 2009). The
194 lake was formed in a deep, glacially eroded basin with a catchment area of 50 km²,
195 and has a mean and a maximum depth of 40 m and 63.2 m, and a surface area of 3.6
196 km². The vegetation around the lake is dominated by pine meadows with dwarf shrubs,
197 rosette plants and alpine cushion (Schlutz and Lehmkuhl 2009, Yuan et al. 2014,
198 Zhang and Mischke 2009). Mostly recharged by the glacial and snowpack melting
199 water and regional precipitation, the lake is stratified with an epilimnion depth about
200 4.4 m in the summer time. The lake is usually covered by ice in the winter time
201 (Zhang and Mischke 2009). The superficial layer within the U-shaped valley is
202 characterized by peat, clay and fluvial gravels with a depth about 1-3.5 m.
203 Discontinuous and isolated permafrost is present at the slope of the valley above the

204 elevation of about 4150 m. The maximum frozen depth is about 1.5 m for the seasonal
205 frozen ground around the lake. The seasonal frozen ground serves as an unconfined
206 aquifer during the unfrozen months from July to October, and groundwater discharges
207 into the epilimnion of the lake (Schlutz and Lehmkuhl 2009, Wang 1997, Zhang and
208 Mischke 2009).

209

210 2.2 Sampling and field analysis

211 The field campaign to Ximen Co Lake was conducted from August, 2015, when it is
212 warm enough to take the water samples of different origins as the studied site is
213 seasonally frozen. A ^{222}Rn continuous monitoring station was setup at the southeast
214 part of the lake, which is fairly flat for setting up our tent and monitoring system.
215 Surface water samples were collected around the lake, rivers at the upstream and
216 downstream. Porewater samples were collected at one side of the lake as the other
217 side is steep and rocky. The basic water quality parameters of conductivity (EC),
218 dissolved oxygen (DO), TDS, ORP, pH in the water were recorded with the
219 multi-parameter meter (HANNA, Co.). Relative humidity was recorded with a
220 portable thermo-hydrometer (KTH-2, Co.). Lake water samples were taken with a
221 peristaltic pump into 2.5 L glass bottles for ^{222}Rn measurement with the Big Bottle
222 system (DurrIDGE, Co.). Surface water samples were filtered with 0.45 μm filters

223 (Advantec, Co.) in situ and taken into 5 ml, 15 ml, 15 ml and 50 ml Nalgene
224 centrifugation tubes for stable isotope, major anion, cation and nutrient analysis.
225 Porewater samples were taken from the lakes shore aquifers with a push point sampler
226 (M.H.E, Co.) connected to peristaltic pump (Solinst, Co.) (Luo et al. 2014, Luo et al.
227 2016). 100 ml raw surface water or porewater was titrated with 0.1 μM H_2SO_4
228 cartridge (Hach, Co.) in situ to measure total alkalinity (Hasler et al. 2016, Warner et
229 al. 2013, White et al. 2016). Porewater was filtered with 0.45 μm syringe filters
230 (Advantec, Co.) in situ and taken into 5 ml, 15 ml, 15 ml and 50 ml Nalgene
231 centrifugation tubes for stable isotope, major anion, cation and nutrient analysis. 250
232 ml porewater was taken for ^{222}Rn measurement with RAD7 H_2O (DurrIDGE, Co.)
233 Samples for major cation analysis were acidified with distilled HNO_3 immediately
234 after the sampling.

235 ^{222}Rn continuous monitoring station was set up at the northwest of the lake, close
236 to the downstream of the lake (Figure 1b). Lake water (about 0.5 m) was pumped with
237 a DC pump (12 V) driven by lithium batteries (100 Ah) and sprinkled into the
238 chamber of RAD7 AQUA with a flow rate $> 2 \text{ L min}^{-1}$, where ^{222}Rn in water vapor
239 was equilibrated with the air ^{222}Rn . The vapor in the chamber was delivered into two
240 large dry units (Drierite, Co) to remove the moisture and circulated into RAD7
241 monitor, where ^{222}Rn activities were recorded every 5 mins. A temperature probe

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243 (HOBO[®]) was insert into the chamber to record the temperature of the water vapor.
244 The monitoring was performed from 11: 31 am, Aug 22nd to 6: 30 am, Aug 24th, 2015.
245 During the period of 1:50-4:30 pm on Aug 22nd, a sudden blizzard occurred, leading
246 to an hourly precipitation about 0.6 mm to the lake area. Daily and hourly
247 climatological data such as wind speed, air temperature and precipitation were
248 retrieved from the nearest weather station in Jiuzhi town (N: 33.424614 °, E:
249 101.485998). ~~Moreover, another RAD7 was placed at the lakeshore to measure ²²²Rn~~
250 ~~in the ambient air around the lake. Due to extremely low activities, the monitoring~~
251 ~~period was conducted only for 4 hours, and the mean activity was adopted as the~~
252 ~~background radon-222 activity to be used in the mass balance model.~~ Water level and
253 temperature fluctuations were recorded with a conductivity-temperature-depth diver
254 (Schlumberger, Co.) fixed at about 20 cm below the lake surface and calibrated with
255 local atmospheric pressure recorded by a baro-diver (Schlumberger, Co.) above the
256 lake. To correct for dissolved ²²⁶Ra supported ²²²Rn, one radium sample was extracted
257 from 100 L lake water with MnO₂ fiber as described elsewhere (Luo et al. 2014,
258 Moore 1976).

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260 2.3 Chemical analysis

261 Major ions were measured with ICS-1100 (Dionex. Co.) in the Department of Earth

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267 Sciences, the University of Hong Kong. The uncertainties of the measurements are
268 less than 5 %. Nutrients, DIN and DIP were analyzed with flow injection analysis
269 equipped with auto-sampler (Lachat. Co.) in the School of Biological Sciences, the
270 University of Hong Kong. Stable ^{18}O and ^2H isotopes were measured with
271 MOA-ICOS laser absorption spectrometer (Los Gatos Research (LGR) Triple Isotope
272 Water Analyzer (TIWA-45EP)) at State Key Laboratory of Marine Geology, Tongji
273 University, Shanghai. The stable isotopic standards and the recovery test has been
274 fully described elsewhere (Luo et al., 2017). The measurement uncertainty is better
275 than 0.1 % for ^{18}O and 0.5 % for ^2H . ^{226}Ra was detected with RAD7 with the method
276 described elsewhere (Kim et al. 2001, Lee et al. 2012, Luo et al. 2018).

277

278 2.4 Radon transient model

279 Previous studies employed a steady state radon-222 mass balance model to
280 quantify LGD to lentic system such as lakes and wetlands (Dimova and Burnett 2011,
281 Luo et al. 2016). This model assumes that radon input derived from groundwater
282 inflow, diffusion and river inflow are balanced by the radon losses of atmospheric
283 evasion, decay and river outflow. However, recent studies revealed that the steady
284 state is mainly reached after 2-15 days of constant metrological conditions, and
285 mostly lentic system can be not be treated as steady state due to rapid radon-222

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287 degassing to the atmosphere driven by wind-induced turbulence (Gilfedder et al.,
288 2015; Dimova and Burnett, 2011).

289 Ximen Co lake is demonstrated to be highly stratified with an epilimnion of 4.4
290 m (Zhang and Mischke 2009). The lake was formed by glacier erosion and the
291 lakebed is characterized by granite bedrock with a thin sedimentary clay layer.
292 Previous studies have indicated that sediment with a thickness of 0.7-3.3 m has been
293 developed on the bedrock and forms the lake shore aquifer, which consists of clay,
294 soils and gravels (Schlutz and Lehmkuhl 2009). Porewater sampled in the aquifer
295 immediately behind the lake shore can well represent groundwater discharging into
296 the lake, as suggested previously (Lewandowski et al. 2015, Rosenberry et al. 2015,
297 Schafran and Driscoll 1993). LGD has been widely considered to occur within the
298 first few meters of the lake shore (Lee et al. 1980, Rosenberry et al. 2015, Schafran
299 and Driscoll 1993) and groundwater is considered to predominately discharge into the
300 epilimnion since deep groundwater flow is highly limited by the Precambrian bedrock
301 (Einarsdottir et al., 2016). Therefore, ^{222}Rn mass balance model is established to
302 quantify LGD to the epilimnion from the lake shore. Due to negligible hydrological
303 connection between the epilimnion and hypolimnion, LGD for the lake can be
304 quantified with ^{222}Rn mass balance model for the epilimnion.

305 The governing equation of radon-222 transient mass balance model within a 1 x

306 1 x z cm (where z is the depth in cm) can be expressed as (Gilfedder et al. 2015):

307
$$z \frac{\partial I_w}{\partial t} = F_{gw} + (I_{^{226}\text{Ra}} - I_w) \times z \times \lambda_{^{222}\text{Rn}} + F_{diff} - F_{atm} \quad (1)$$

308 where F_{gw} , F_{diff} , F_{atm} [$\text{Bq m}^{-2} \text{d}^{-1}$] are ^{222}Rn loadings from LGD, water-sediment

309 diffusion and water-air evasion, respectively; z [m] is the lake water level depth

310 recorded by the diver. $\lambda_{^{222}\text{Rn}}$ is the decay constant of ^{222}Rn with a value of 0.186d^{-1} .

311 $\lambda_{^{222}\text{Rn}} \times I_{^{226}\text{Ra}}$ and $\lambda_{^{222}\text{Rn}} \times I_w$ account for the production and decay of ^{222}Rn [$\text{Bq m}^{-2} \text{d}^{-1}$]

312 in the water column, respectively. I_w and $I_{^{226}\text{Ra}}$ [Bq m^{-2}] represent ^{222}Rn and ^{226}Ra

313 inventories in the epilimnion, and are expressed as: $I_w = H \times C_w$ and

314 $I_{^{226}\text{Ra}} = H \times C_{^{226}\text{Ra}}$, respectively; where H [m] is the depth of the epilimnion; C_w and

315 $C_{^{226}\text{Ra}}$ is the ^{222}Rn and ^{226}Ra activity [Bq m^{-3}], respectively.

316 The model is valid under the following assumptions: 1) The epilimnion is well

317 mixed which is the actual condition for most natural boreal and high altitude glacial

318 lakes (Åberg et al. 2010, Zhang and Mischke 2009). 2) ^{222}Rn input from riverine

319 water inflow, and loss from the lake water outflow and infiltration into the lake shore

320 aquifer is negligible compared to the groundwater borne ^{222}Rn , because ^{222}Rn

321 concentration of groundwater is 2-3 orders of magnitude larger than that lake water

322 (Dimova and Burnett 2011, Dimova et al. 2013). Generally, ^{222}Rn in the epilimnion is

323 sourced from LGD and decay input from parent isotope of ^{226}Ra under secular

324 equilibrium, and is mainly lost via atmospheric evasion and radioactive decay.

325 F_{atm} is the key sinking component of the transient model and is finally a function of
326 wind speed and water temperature, both of which are temporal variant variables
327 (Supplementary information). Lake water level z is also a temporal variant variable
328 which represents the fluctuations of water volume of the epilimnion. This equation is
329 discretized by the forward finite difference method, and the groundwater flux at each
330 time step can be solved as follow

$$331 \quad [{}^{222}\text{Rn}_{t+\Delta t}] = \frac{[z \times {}^{222}\text{Rn}_t + [F_{diff} + F_{gw} - F_{atm} - {}^{222}\text{Rn}_t \times \lambda \times z] \times \Delta t]}{z} \quad (2)$$

332 where ${}^{222}\text{Rn}_{t+\Delta t}$ and ${}^{222}\text{Rn}_t$ [Bq m⁻³] is the ²²²Rn activity at current time step and at
333 the previous time steps, respectively, and Δt [min] is the time step which is set to be
334 5 min in consistence with the ²²²Rn record interval. With the inverse calculation based
335 on Equation (2), the groundwater inflow at each time step can be obtained. However,
336 large errors of the final LGD calculation will be induced by even a small amount of
337 noise in the measured ²²²Rn data due to the ${}^{222}\text{Rn}_{t+\Delta t} - {}^{222}\text{Rn}_t$ term being with the
338 measure uncertainty. To reduce the random errors of the measured ²²²Rn
339 concentrations, the time window with a width of 1 hour is proposed to smooth the
340 curve (Supplementary information).

341

342 3. Results

343 3.1 Time series data

344 Figure 2 shows the basic climatological parameters of the lake catchment during

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346 the campaign month. There are discrete rainfall events occurring throughout the
347 month with an average rainfall of 3.1 mm d^{-1} . The temperature throughout the month
348 ranges from $5.0 - 12.5 \text{ }^{\circ}\text{C}$ within an average of $9.3 \text{ }^{\circ}\text{C}$. The daily averaged wind speed
349 generally ranges from $0.7 - 2.5 \text{ m s}^{-1}$, with an average of 1.7 m s^{-1} . ^{222}Rn temporal
350 distribution and other time series data are shown in Figure 3a and listed in
351 Supplementary Table 1. Generally, ^{222}Rn concentration varies from 32.2 to 273 Bq m^{-3} ,
352 with an average of $144.2 \pm 27.7 \text{ Bq m}^{-3}$. ^{222}Rn over the monitoring period shows
353 typical diel cycle, much higher at nighttime and lower in the day time. Figures 3b-3d
354 shows the time series data of temperature (5 mins interval), nearshore lake water level
355 (1 min interval), and wind speed (1 hour interval). Temperature and lake water level
356 also show typical diel cycles, but with antiphase fluctuations with each other.
357 Temperature is higher during the daytime and lower at nighttime. However a sudden
358 decrease of temperature was recorded due to the sudden blizzard (Figure 3b). Water
359 level is higher at nighttime and lower during the daytime, with a strong fluctuation
360 due to the turbulence caused by the blizzard (Figure 3c). The variability might reflect
361 the dynamics of groundwater input and surface water inflow. The air temperature of
362 the lake area is in phase with the water temperature. Wind speed is normally higher
363 during the daytime and lower at nighttime (Figure 3d).

364 The variation of ^{222}Rn is nearly in antiphase with the fluctuations of lake water

365 temperature and air temperature, indicating that the dominated controlling factors of
366 ^{222}Rn fluctuations are water temperature and wind speed (Figure 3a). This
367 phenomenon is reasonable as lake water ^{222}Rn is predominately lost via atmospheric
368 evasion, which is the function of wind speed and water temperature (Dimova et al.
369 2015, Dimova and Burnett 2011, Dimova et al. 2013). High water temperature and
370 wind speed leads to elevated atmospheric evasion and causes the decline of ^{222}Rn
371 concentration in the lake water. However, there is a sudden reduction of radon activity
372 from 2: 00 pm to 4: 00 pm on Jul 22nd, 2015, when the snow event led to a sudden
373 decrease of water temperature, increase of wind speed, and large surface water
374 turbulence as indicated by water level fluctuations (Figures 3a-3d). ^{222}Rn in the
375 porewater is 2-3 orders of magnitude larger than ^{222}Rn in the lake water, suggesting
376 that ^{222}Rn is an ideal tracer to estimate the LGD (Supplementary Table 1). ^{222}Rn
377 concentrations in surface water range from 22.2 to 209 Bq m⁻³, with an average of
378 92.5 Bq m⁻³ (n = 12), which is in the range of ^{222}Rn continuous monitoring results,
379 suggesting reliable ^{222}Rn measurements (Supplementary Table 2).

380

381 3.2 Geochemical results

382 The results of major ions, nutrients and stable isotopes in different water end
383 members are shown in Figures 4 and 5. Cl^- ranges from 0.6 to 2.1 mg L⁻¹ in the

384 surface water (including riverine inflow water, lake water and downstream water), 0.4
385 to 2.7 mg L⁻¹ in porewater and has a much higher concentration of 5.9 mg L⁻¹ in
386 rainfall water. Na⁺ ranges from 1.6 to 3.4 mg L⁻¹ in the surface water, 1.2 to 4.4 mg
387 L⁻¹ in porewater and has a concentration of 4.4 mg L⁻¹ in rainfall water. SO₄²⁻ ranges
388 from 1.2 to 2.3 mg L⁻¹ in the surface water, 0.4 to 1.7 mg L⁻¹ in porewater and has a
389 significant low concentration of 0.01 mg L⁻¹ in rainfall water. Ca²⁺ ranges from 3.0 to
390 12.4 mg L⁻¹ in lake water, 3.4 to 12.5 mg L⁻¹ in porewater and has a significant high
391 concentration of 20.5 mg L⁻¹ in rainfall water. Other concentrations of major ions are
392 listed in Supplementary Table 2. As shown in Figure 4d and Supplementary Table 2,
393 δ¹⁸O in the lake water ranges from - 13.06 ‰ to - 12.11 ‰, with an average of -
394 12.41 ‰ (n = 7), and δ²H ranges from - 91.83 ‰ to - 87.47 ‰, with an average of -
395 89.0 ‰ (n = 7). δ¹⁸O in the riverine inflow water ranges from - 13.44 ‰ to - 13.29 ‰,
396 with an average of - 13.37 ‰ (n = 2), and δ²H ranges from - 93.25 ‰ to - 91.92 ‰,
397 with an average of - 92.59 ‰ (n = 2). δ¹⁸O in the downstream water ranges from -
398 12.51 ‰ to - 12.18 ‰, with an average of - 12.35 ‰ (n = 3), and δ²H ranges from -
399 88.96 ‰ to - 87.1 ‰, with an average of - 87.98 ‰ (n = 3). δ¹⁸O in the porewater
400 ranges from - 12.66 ‰ to - 11.52 ‰, with an average of - 11.97 ‰ (n = 8), and δ²H
401 ranges from - 91.3 ‰ to -82.87 ‰, with an average of - 85.5 ‰ (n = 8). DIN in the
402 surface water (including riverine inflow water, lake water and downstream water)

403 range from 6.6 to 16.9 μM , with an average of 10.3 μM , and DIP from 0.36 to 0.41
404 μM , with an average of 0.38 μM . The concentrations of DIN for the porewater range
405 from 0.7 to 358.8 μM , with an average of 92.8 μM , and DIP from 0.18 to 0.44 μM
406 with an average of 0.31 μM (Figure 5).

407

408 **4. Discussion**

409 4.1 Proglacial hydrologic processes and geochemical implications

410 Generally, major ion concentrations in the lake water and porewater of Ximen
411 Co lake are significantly lower than those in main rivers, streams and other tectonic
412 lakes in the QTP (Wang et al. 2010, Wang et al. 2016b, Yao et al. 2015), and are
413 similar to those of snow and glaciers (Liu et al. 2011), suggesting that the lake water
414 is mainly originated from glacier and snow melting. Ion concentrations in the lake and
415 porewater of Ximen Co lake are much lower than those of rainfall collected in Jiuzhi
416 town. This suggests that lake water is less influenced by precipitation (Figures 4a-4c).
417 The concentrations of major ions in the porewater are high compared to the lake water,
418 indicating weathering affects from the aquifer grains. The ratios of $\text{Ca}^{2+}/\text{Na}^{+}$ in the
419 porewater and groundwater is >1 , also suggesting influences of weathering digenesis
420 of major ions from the seasonal frozen ground at the lake shore aquifer (Wang et al.
421 2010, Weynell et al. 2016, Yao et al. 2015).

422 The isotopic compositions of the lake water and porewater are significantly
423 isotopic depleted, with values close to the compositions of glaciers and surface snow
424 in the QTP, suggesting the lake is dominantly recharged from snow and glacier
425 melting (Cui et al. 2014, Wang et al. 2016a, Zongxing et al. 2015). The relation of
426 $\delta^{18}\text{O}$ versus $\delta^2\text{H}$ for the lake water is $\delta^2\text{H} = 4.25 \times \delta^{18}\text{O} - 35.99$, with a slope much
427 lower than that of the global meteoric water line (GMWL) (Figure 4d), suggesting the
428 effects of lake surface evaporation. The relation of $\delta^{18}\text{O}$ versus $\delta^2\text{H}$ for the porewater
429 is $\delta^2\text{H} = 6.93 \times \delta^{18}\text{O} - 2.67$, overall on GWML (Figure 4d). Deuterium excesses is
430 defined as $\Delta\text{D} = \delta\text{D} - 8 \times \delta^{18}\text{O}$ (Dansgaard 1964). The value of ΔD is dependent on
431 airmass origins, altitude effect and the kinetic effects during evaporation (Hren et al.
432 2009). Global meteoric water has a ΔD of + 10 ‰. In the QTP, glacier/snowpack
433 melting water usually has large positive ΔD , while the precipitations derived from
434 warm and humid summer monsoon has lower ΔD (Ren et al. 2017, Ren et al. 2013).
435 In this study, ΔD of surface water, lake and porewater ranges from + 8.5 to + 11.8 ‰,
436 closed to the glacier melting water but much smaller than that of the local
437 precipitation of + 18.8 ‰. This indicates the stream and lake water are mainly
438 originated from glacial/snowpack melting rather than precipitation (Gat 1996, Lerman
439 et al. 1995, Wang et al. 2016a). The slopes of $\delta^2\text{H}$ versus $\delta^{18}\text{O}$ in lake water and
440 porewater are 4.25 and 6.93, both of which are lower than that of GMWL due to

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446 surface evaporation. Lake water is more intensively influenced by evaporation
447 compared to porewater. The plots of $\delta^{18}\text{O}$ versus Cl^- , and $\delta^2\text{H}$ versus Cl^- are well
448 clustered for porewater end member (orange area), lake water end member (blue area),
449 riverine inflow water end member (yellow area), and precipitation water (Figures 4e
450 and 4f), suggesting stable $\delta^{18}\text{O}$ and $\delta^2\text{H}$ isotopes and Cl^- can serve as tracers to
451 quantify the hydrologic partitioning of the lake by setting three endmember models.

452 The concentrations of DIN and DIP are all within the ranges of other glacial
453 melting water and proglacial lake water (Hawkings et al. 2016, Hodson 2007, Hodson
454 et al. 2005, Hudson et al. 2000, Tockner et al. 2002). Briefly, rainfall and upstream
455 lake water such as YN-4 has the highest DIN concentration, indicating the glacier
456 melting and precipitation could be important DIN sources in proglacial areas
457 (Anderson et al. 2017, Dubnick et al. 2017). DIN in porewater is overall higher
458 compared to the lake water, suggesting the porewater to be DIN effective source; and
459 DIP concentrations is higher in the lake water compared to porewater, suggesting the
460 porewater is a DIP sink (Figure 5). The N: P ratios in the lake water and porewater are
461 averaged to be 27.1 and 320.5, respectively, both much larger than the Redfield Ratio
462 (N: P = 16:1) in water and organism in most aquatic system and within the range of
463 other proglacial lakes (Anderson et al. 2017). This also suggests that the lake water
464 and porewater are under phosphate limited condition. N: P ratio in the rainfall water is

465 30.4, similar to the lake water. The average N: P ratio of porewater is much higher
466 than that of lake water, indicating DIN enrichment in the lake shore aquifers (Figure 5).
467 In pristine groundwater, NO_3^- is the predominated form of N and is highly mobile
468 within the oxic aquifers, leading to much higher DIN concentrations in the porewater;
469 DIP has high affinity to the aquifer grains, resulting in much lower DIP concentrations
470 in the porewater (Lewandowski et al. 2015, Rosenberry et al. 2015, Slomp and Van
471 Cappellen 2004). Thus, in analogous to surface runoff from glacier/snowpack melting,
472 LGD can be also regarded as an important source for the proglacial lakes. Because of
473 very high DIN and N: P ratios in the porewater, a relatively small portion of LGD
474 delivers considerable nutrients into the glacial lake, shifting the aquatic N: P ratios
475 and affecting the proglacial aquatic ecosystem (Anderson et al. 2017).

476

477 4.2 Estimation of LGD

478 Figure 6a shows all the sinks and sources of radon with the epilimnion of the lake.
479 Within ^{222}Rn transient mass balance model, the dominant ^{222}Rn loss is atmospheric
480 degassing/evasion. Generally, ^{222}Rn degassing rate is the function of the radon-222
481 concentration gradient at the water-air interface and the parameter of gas piston
482 velocity k , which is finally the function of wind speed and water temperature (Dimova
483 and Burnett 2011, Gilfedder et al. 2015). To evaluate ^{222}Rn evasion rate, this study

484 employs the widely used method proposed by MacIntyre et al. (1995) which is also
485 detailed described in Supplementary Information. Based on the field data of ^{222}Rn
486 concentration in the lake water, wind speed and temperature log, the radon degassing
487 rate is calculated in a range of 0.8 to 265.2 $\text{Bq m}^2 \text{d}^{-1}$, with an average 42.0 of Bq m^2
488 d^{-1} .

489 In addition to the atmospheric loss and sedimentary diffusion inputs, ^{222}Rn is also
490 sinked via radioactive decay, and sourced from decay of parent isotope of ^{226}Ra . The
491 decay loss of ^{222}Rn fluctuates in phase with the distribution of ^{222}Rn concentration
492 monitored by RAD 7 AQUA. The equations to estimate benthic fluxes are shown in
493 supplementary information. The decay loss is calculated to be 26.4 to 223.4 $\text{Bq m}^{-2} \text{d}^{-1}$,
494 with an average of $118.0 \pm 22.7 \text{Bq m}^{-2} \text{d}^{-1}$. ^{226}Ra concentration is 0.01Bq m^{-3} for the
495 lake water. Under secular equilibrium, the ^{226}Ra decay input can be calculated by
496 multiplying ^{226}Ra concentration in the lake water with λ_{222} (Corbett et al. 1997, Kluge
497 et al. 2007, Luo et al. 2016). ^{226}Ra decay input is calculated to be $0.83 \text{Bq m}^{-2} \text{d}^{-1}$,
498 which is significantly low compared to other ^{222}Rn sources to the epilimnion.

499 With the obtained sinks and sources of ^{222}Rn in the lake, and the constants given in
500 Table 1, LGD rate can be obtained by dividing the groundwater derived ^{222}Rn with its
501 concentration in groundwater endmember. The obtained LGD rate, ranges from -23.7
502 mm d^{-1} to 90.0mm d^{-1} , with an average of $10.3 \pm 8.2 \text{mm d}^{-1}$ (Figure 7) The LGD rate

503 range is relatively less than the daily lake water level variations (≈ 50 mm), indicating
504 that the lake water level variation could be a combined effect of surface runoff and
505 LGD (Hood et al. 2006). The negative values of LGD rate reflect the return
506 groundwater flow due to infiltration into the porewater. Normally, the dominant
507 values are positive, indicating LGD rate is significant compared to water infiltrations
508 into lakeshore aquifer. The temporal variation of LGD rate could be attributed to the
509 fluctuations of the hydraulic gradient in the proglacial areas (Hood et al. 2006, Levy
510 et al. 2015). As indicated by ΔD (mostly > 10) of surface water, the lake and the
511 upstream water is considered to be mainly recharged from glacial/snowpack melting
512 rather other precipitations.

513 To assess the magnitude of uncertainty of ^{222}Rn transient model, the sensitivity of
514 estimated LGD to changes in other variables is examined. A sensitivity coefficient f is
515 proposed to evaluate this uncertainty according to Langston et al. (2013)

$$516 \quad f = (\Delta F_{LGD} / F_{LGD}) / (\Delta y_i / y_i) \quad (3)$$

517 where ΔF_{LGD} is the amount of change in F_{LGD} from the original value. Δy_i is the
518 amount of change in the other variable of y_i from the original value. Thus, higher f
519 indicates a large uncertainty of final LGD estimate. The uncertainty mainly stems
520 from ^{222}Rn measurements in different water endmembers, the atmospheric loss and
521 water level record. The uncertainties of ^{222}Rn measurement are about 10 % and 15-20

522 % in groundwater and lake water endmember, respectively. The uncertainty of
523 atmospheric loss is derived from uncertainty of ^{222}Rn in lake water (with an
524 uncertainty of 15-20 %), temperature (with an uncertainty ≈ 5 %) and wind speed
525 (with an uncertainty ≈ 5 %). Thus, the final LGD estimate has an uncertainty of
526 35-40 %.

527

528 **4.3 Hydrologic partitioning**

529 Compared to the groundwater labeled radionuclide of ^{222}Rn , stable $^{18}\text{O}/^2\text{H}$
530 isotopes are advantageous in the investigation of evaporation processes due to their
531 fractionations from water to vapor and have been widely used to investigate the
532 hydrologic cycle of lakes in various environments (Gat 1995, Gibson et al. 1993,
533 Gonfiantini 1986, Stets et al. 2010). With the field data of stable isotopic composition
534 and Cl^- concentrations in different water end members, groundwater input, surface
535 water input, lake water outflow and infiltration, and evaporation can be partitioned by
536 coupling stable isotopic mass balance model with Cl^- mass balance model (Figure 6b).

537 The model, consisting of the budgets of stable isotopes and Cl^- , and water masses
538 for the epilimnion, is used to quantify riverine inflow, lake water outflow and
539 infiltration, and evaporation (Gibson et al. 2016, LaBaugh et al. 1995, LaBaugh et al.
540 1997). The model is valid under the following assumptions: (1) constant density of

541 water; (2) no long-term storage change in the reservoir; (3) well-mixed for the
 542 epilimnion (Gibson 2002, Gibson et al. 2016, Gibson and Edwards 2002, LaBaugh et
 543 al. 1997). The above assumptions are reasonably tenable during the short monitoring
 544 period. The model can be fully expressed as

$$545 \quad F_{in} + F_{LGD} + F_p = F_E + F_{out} \quad (4)$$

$$546 \quad F_{in} \times \delta_{in} + F_{LGD} \times \delta_{gw} + F_p \times \delta_p = F_E \times \delta_E + F_{out} \times \delta_L \quad (5)$$

$$547 \quad F_{in} \times [Cl^-]_{in} + F_{LGD} \times [Cl^-]_{gw} + F_p \times [Cl^-]_p = F_E \times [Cl^-]_L + F_{out} \times [Cl^-]_L \quad (6)$$

548 where F_{in} [mm d⁻¹] is the surface water inflow to the lake; F_{gw} [mm d⁻¹] is LGD rate.
 549 F_p [mm d⁻¹] is the mean daily rainfall rate during the sampling period. F_E [mm d⁻¹] is
 550 the lake evaporation. F_{out} [mm d⁻¹] is the lake water outflow via runoff and
 551 infiltration into the lake shore aquifer. δ_{in} , δ_{gw} , δ_E and δ_p are the isotopic compositions
 552 of surface water inflow, LGD, and evaporative flux, respectively. The values of δ_{in} ,
 553 δ_{gw} , and δ_p are obtained from field data and the composition of δ_E are calculated as
 554 shown in supplementary information. $[Cl^-]_{in}$, $[Cl^-]_{gw}$, $[Cl^-]_L$ and $[Cl^-]_p$ are the
 555 chloride concentrations in the inflow water, porewater, lake water and precipitation,
 556 respectively.

557 The components of the mass balance model can be obtained from the field data of
 558 isotopic composition and Cl⁻ concentrations in different water endmembers. The
 559 average ¹⁸O composition -13.37 ‰ of riverine inflow water is taken as the value of

560 the input parameter δ_{in} , $\delta^{18}\text{O}$ and $\delta^2\text{H}$ in the groundwater endmember and lake water
561 end member are calculated to be -12.41 ‰ and -87.18 ‰, respectively. $\delta^{18}\text{O}$ and $\delta^2\text{H}$
562 in the rainfall are measured to be -5.47 ‰ and -24.98 ‰, respectively. With the
563 measured values of δ_L , h , δ_{in} , and the estimated ε and δ_a , the isotopic composition
564 of δ_E is calculated to be -35.11 ‰, which is in line with the results of alpine and
565 arctic lakes elsewhere (Gibson 2002, Gibson et al. 2016, Gibson and Edwards 2002).
566 The values of $[\text{Cl}^-]_{in}$, $[\text{Cl}^-]_{gw}$, and $[\text{Cl}^-]_L$ are calculated to be 0.91 mg L⁻¹, 1.48
567 mg L⁻¹ and 1.02 mg L⁻¹, respectively. All the parameters used in the model are shown
568 in Table 2.

569 According to Equations 4-6, the uncertainties of calculations of F_{in} , F_{out} and E are
570 mainly derived from the uncertainty of F_{LGD} and the compositions of Cl⁻, δD and $\delta^{18}\text{O}$
571 in different water endmembers as suggested in previous studies (Genereux 1998,
572 Klaus and McDonnell 2013). The compositions of Cl⁻, δD and $\delta^{18}\text{O}$ in surface water,
573 groundwater endmembers have an uncertainty of 5 %. The uncertainty of δ_E is
574 reasonably assumed to be ≈ 20 %. Thus, considering the uncertainty propagation of
575 all the above parameters, the uncertainties of F_{in} , F_{out} and E would be scaled up to
576 70-80 % of the final estimates.

577

578 4.4 The hydrologic partitioning of the glacial lake

579 Based on the three endmember model of ^{18}O and Cl^- , the riverine inflow rate was
580 calculated to be $135.6 \pm 119.0 \text{ mm d}^{-1}$, and the lake outflow rate is estimated to be
581 $141.5 \pm 132.4 \text{ mm d}^{-1}$; the evaporation rate is calculated to be $5.2 \pm 4.7 \text{ mm d}^{-1}$. The
582 summary of the hydrologic partitioning of the lake is shown in Figure 8a. Generally,
583 the proglacial lake is mostly recharged by the riverine inflow from the snowpack or
584 the glacier melting. The groundwater discharge contributes about only 7.0 % of the
585 total water input to the lake, indicating groundwater input does not dominate water
586 input to the proglacial lake. The recent review on LGD rate by Rosenberry et al.
587 (2015) suggests that the median of LGD rate in the literatures is 7.4 mm d^{-1} (0.05 mm
588 d^{-1} to 133 mm d^{-1}), which is about 2/3 of LGD rate in this study. This difference may
589 be due to the hydrogeological setting of the lake shore aquifer. This aquifer is formed
590 by grey loam, clayey soil and sand (Lehmkuhl 1998, Schlutz and Lehmkuhl 2009),
591 which is with relatively high permeability. Previous studies have indicated that
592 groundwater forms a key component of proglacial hydrology (Levy et al. 2015).
593 However, there have been limited quantitative studies of groundwater contribution to
594 hydrologic budget of proglacial areas. This study further summarizes the groundwater
595 discharge studies over the glacial forefield areas. Based on long term hydrological and
596 climatological parameter monitoring on the Nam Co lake in the QTP, Zhou et al.
597 (2013) estimated the LGD to be 5-8 mm d⁻¹, which is comparable to the surface

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surface water outflow and infiltration to
lake shore aquifers. The evaporation
constitutes relatively small ratio (≈ 3.5
of total water losses. The annual
evaporation rate was recorded to be 142
mm (equivalent to 3.92 mm d^{-1}) in 2015
the Jiuzhi weather station, lower than the
obtained evaporation in this study. This
may be due to much higher evaporation
August during the monitoring period.

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2013)

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611 | runoff input and LGD of this study. Brown et al. (2006) investigated the headwater
612 streams at the proglacial areas of Taillon Glacier in French and found that
613 groundwater contributes 6-10 % of the stream water immediate downwards of the
614 glacier. Using water mass balance model, Hood et al. (2006) shows that groundwater
615 inflow is substantial in the hydrologic partitioning of the proglacial Lake O'Hara in
616 front of Opabin Glacier in Canada and comprised of 30 -74 % of the total inflow. Roy
617 and Hayashi (2008) studied the proglacial lakes of Hungabee lake and Opabin lake at
618 glacier forefield of Opabin Glacier and found that groundwater component is
619 predominant water sources of the lakes and consisted of 35-39 % of the total water
620 input of the lakes. Langston et al. (2013) further investigated a tarn immediate in front
621 of Opabin Glacier and indicated the tarn is predominantly controlled by groundwater
622 inflow/outflow, which consisted of 50-100 % of total tarn volume. Magnusson et al.
623 (2014) studied the streams in the glacier forefield of Dammagletscher, Switzerland
624 and revealed that groundwater contributed only 1-8 % of the total surface runoff.
625 Groundwater contribution in this study is similar to those obtained the mountainous
626 proglacial areas in Europe, but much lower than those obtained in the proglacial areas
627 of polar regions. It is concluded that proglacial lakes/streams in front of mountainous
628 glaciers are mainly recharged by surface runoff from glacier/snowpack melting. This
629 might be due to well-developed stream networks and limited deep groundwater flow

630 (Brown et al. 2006, Einarsdottir et al. 2017, Magnusson et al. 2014). However,
631 proglacial tarns and lakes in the polar areas are predominantly controlled by
632 groundwater discharge, due to less connectivity of surface runoff and high shallow
633 and deep groundwater connectivity (Hood et al. 2006, Langston et al. 2013, Roy and
634 Hayashi 2008).

635 The evaporation constitutes relatively small ratio ($\approx 3.5\%$) of total water losses. The
636 annual evaporation rate was recorded to be 1429.8 mm (equivalent to 3.92 mm d^{-1}) in
637 2015 by the Jiuzhi weather station, lower than the obtained evaporation in this study.

638 This may be due to much higher evaporation in August during the monitoring period.

639 The estimation of evaporation in this study generally represents the upper limit of the
640 lake, as the sampling campaign was conducted during the summer time when the

641 highest evaporation might occur. The lake surface evaporation derived from the pan

642 evaporation in the QTP ranges from $\sim 700\text{ mm yr}^{-1}$ in the eastern QTP to over 1400

643 mm yr^{-1} in the interior lakes of the QTP (Ma et al. 2015, Yang et al. 2014, Zhang et al.

644 2007). The evaporation of this study is rather in line with the previous evaporation

645 observation in the eastern QTP, stressing the tenability of evaporation in this study.

646 The runoff input is predominated recharge component ($> 90\%$) compared to other

647 components, with an area normalized value comparable to previous studies of runoff

648 input in other glacial melting dominant lakes in the QTP (Biskop et al. 2016, Zhang et

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659 al. 2011, Zhou et al. 2013). The runoff input and the lake evaporation of the study
660 area, however, are subject to highly daily, seasonal and inter-annual variability as
661 indicated by previous studies in the QTP (Lazhu et al. 2016, Lei et al. 2017, Ma et al.
662 2015, Zhou et al. 2013). Therefore, further investigations of long term and high
663 resolution climatological and isotopic data are required to provide precise constraints
664 of hydrologic partitioning of the lakes in the QTP.
665 4.5 LGD derived nutrient loadings, nutrient budget and ecological implications

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666 Compared to extensive studies of SGD derived nutrient loadings in the past decade
667 (Luo and Jiao 2016, Slomp and Van Cappellen 2004), studies of LGD derived nutrient
668 loadings have received limited attention, even given the fact that groundwater in lake
669 shore aquifers is usually concentrated in nutrients (Lewandowski et al. 2015,
670 Rosenberry et al. 2015). Even fewer studies focus on chemical budgets in the
671 proglacial lakes which are often difficult to access for sampling. Groundwater borne
672 DIN and DIP across the sediment-water interface in this study are determined with an
673 equation coupling the advective or LGD-derived, and diffusive solute transport
674 (Hagerthey and Kerfoot 1998, Lerman et al. 1995)

675
$$F_j = -nD_j^m \frac{dC_j}{dx} + v_{gw} C_j \quad (7)$$

676 where $-nD_j^m \frac{dC_j}{dx}$ is the diffusion input and $v_{gw} C_j$ is the LGD derived fluxes, F_j

681 $[\mu\text{M m}^{-2} \text{ d}^{-1}]$ is the mol flux of nutrient species j (representing DIN or DIP). n is the
682 sediment porosity. D_j^m is the molecular diffusion coefficient of nutrient species j ,
683 which is given to be $4.8 \times 10^{-5} \text{ m}^2 \text{ d}^{-1}$ for DIP (Quigley and Robbins 1986), and $8.8 \times$
684 $10^{-5} \text{ m}^2 \text{ d}^{-1}$ for DIN (Li and Gregory 1974), respectively. $C_j [\mu\text{M}]$ is the concentration
685 of nutrient species j . $x[\text{m}]$ is the sampling depth. v_{gw} is LGD rate estimated by ²²²Rn
686 mass balance model and has a value of $10.3 \pm 8.2 \text{ mm d}^{-1}$. $\frac{dC_j}{dx}$ is the concentration
687 gradient of nutrient species j across the water-sedimentary interface.

688 Substituting the constants and the field data of DIN and DIP in to Equation 6, LGD
689 derived nutrient loadings are calculated to be $954.3 \mu\text{mol m}^{-2} \text{ d}^{-1}$ and $3.2 \mu\text{mol m}^{-2} \text{ d}^{-1}$
690 for DIN and DIP, respectively. Riverine inflow brings $1195.0 \mu\text{mol m}^{-2} \text{ d}^{-1}$ DIN, 52.9
691 $\mu\text{mol m}^{-2} \text{ d}^{-1}$ DIP into the lake. Lake water outflow derived nutrient loss is estimated
692 to be $1439.9 \mu\text{mol m}^{-2} \text{ d}^{-1}$ and $54.7 \mu\text{mol m}^{-2} \text{ d}^{-1}$ for DIN and DIP, respectively.
693 Nutrients in the lake can be also sourced from atmospheric deposit (mostly in form of
694 precipitation). With the nutrient concentrations in the rain water during the monitoring
695 period, the wet deposit is calculated to be $76 \mu\text{mol m}^{-2} \text{ d}^{-1}$ and $2.5 \mu\text{mol m}^{-2} \text{ d}^{-1}$, for
696 DIN and DIP, respectively. The loadings of DIN to the lakes are mainly from surface
697 runoff and LGD, which comprised of 42.9 % and 53.7 % of the total DIN loadings.
698 Groundwater derived DIP input, however, constitutes only 6.3 % of the total DIP
699 inputs to the lake, indicating groundwater borne DIP is less contributive to the

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701 nutrient budget of the lake compared to DIN. Very recent studies on polar regions
702 have indicated that the glacier/snowpack water is the main N sources to the proglacial
703 lakes (Anderson et al. 2013, Dubnick et al. 2017). However, they do not consider the
704 contribution of groundwater borne N, in spite of the high groundwater connectivity in
705 the proglacial areas (Roy and Hayashi 2008). This study stresses that groundwater
706 borne DIN could be comparable to the surface runoff derived DIN.

707 Based on nutrient results, the lake is considered to be an oligotrophic lake, similar
708 to other glacier melting dominant lakes in the QTP (Liu et al. 2011, Mitamura et al.
709 2003). Phytoplankton is good dissolved organic phosphate (DOP) recyclers and will
710 overcome inorganic P limitation though DOP cycling in most template lakes (Hudson
711 et al. 2000). However, this may be not applicable for the glacial melting water and the
712 peri/pro-glacial lake water. Previous studies show that phosphate nutrients are
713 dominated by DIP and particulate phosphate, and the DOP contributes less than 10 %
714 of the dissolved phosphate (Cole et al. 1998, Hawkins et al. 2016, Hodson 2007).
715 Thus DOP recycling is not likely to low N: P ratio under these conditions. Thus, the
716 primary production (PP) is therefore considered to be controlled by the DIP loadings.
717 The sum of DIN and DIP inputs minus the calculated DIN and DIP outputs leads to
718 surpluses of 785.4 $\mu\text{mol m}^{-2} \text{d}^{-1}$ and 3.9 $\mu\text{mol m}^{-2} \text{d}^{-1}$ for DIN and DIP, respectively.
719 The surpluses are expected to be consumed by the phytoplankton and converted into

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Moved down [1]: Thus, the primary production (PP) is therefore considered to be controlled by the DIP loadings. The surplus of DIN and DIP inputs minus the sum of calculated DIN and DIP outputs leads to surpluses of 785.4 $\mu\text{mol m}^{-2} \text{d}^{-1}$ and 3.9 $\mu\text{mol m}^{-2} \text{d}^{-1}$ for DIN and DIP, respectively

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747 the PP under phosphate limited conditions. As primary producers in the fresh
748 lacustrine system consume the nutrient under variant N: P ratios (7.1 to 44.2, mean:
749 22.9) (Downing and McCauley 1992), the biological uptake of DIN is roughly
750 estimated to be $89.3 \mu M m^2 d^{-1}$. Therefore, the nutrient budgets for DIN and DIP
751 can be finally conceptualized in Figures 8b and 8c. ▾

752 4.6. Implications, prospective and limitations

753 Mountainous proglacial lakes are readily developed in glacier forefields of QTP and
754 other high mountainous glacial such as Europe Alps and Pamir at central Asian
755 (Heckmann et al. 2016). The proglacial lakes are always trapping system of sediment
756 and sinks for water and chemical originated from glacier/snowpack melting and
757 groundwater. In analogous to cosmogenic isotopes such as ^{10}Be serving as a tool to
758 quantify the sediment sources, approaches integrating ^{222}Rn and stable isotopes
759 provides both qualitatively and quantitatively evaluations of groundwater
760 contributions and hydrologic partitioning in these remote and untapped lacustrine
761 systems. Thus, it is expected that the multiple aqueous isotopes is considered to be
762 effective tools to investigate the LGD and hydrologic partitioning in other proglacial
763 lakes. This study is mainly limited by the relatively short sampling and monitoring
764 period. As a special hydrologic regime, the lake shore aquifers of the proglacial lakes
765 are experiencing frozen-unfrozen transition seasonally, and the dominant recharge of

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Deleted: The estimated primary productivity is lower than most temperate eutrophicated and oligotrophic lakes (Cetler et al. 1998, Smith 1979), and comparable to some high latitude or altitude lakes (Richerson et al. 1986, Sterner 2010).

775 glacial melting could be fluctuated significantly due to air temperature variation.
776 Therefore, future groundwater and hydrological studies can be extended to longtime
777 sampling and monitoring of stable isotopes and ^{222}Rn in different water endmembers
778 to reveal the seasonally hydrological and hydrogeological dynamics and their impacts
779 on local biogeochemical cycles and ecological systems. Special concerns would be
780 placed on how surface/groundwater interactions and the associated biogeochemical
781 processes in response to the seasonal frozen ground variations and glacier/snowpack
782 melting intensity.

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784 **5. Conclusion**

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785 A ^{222}Rn continuous monitoring is conducted at Ximen Co Lake, a proglacial lake
786 located at the east QTP. A dynamic ^{222}Rn mass balance model constrained by radium
787 mass balance and water level fluctuation is used to quantify temporal distribution of
788 LGD of the lake. The obtained LGD over the monitoring time ranges from -23.7
789 mm d^{-1} to 80.9 mm d^{-1} , with an average of $10.3 \pm 8.2 \text{ mm d}^{-1}$. Thereafter, a three
790 endmember model consisting of the budgets of water, stable isotopes and Cl^- is used
791 to depict the hydrologic partitioning of the lake. Riverine inflow, lake water outflow
792 via surface runoff, and surface evaporation are estimated to be 135.6 mm d^{-1} , 141.5
793 mm d^{-1} and 5.2 mm d^{-1} , respectively. LGD derived nutrient loading is estimated to be

Deleted: Upon depicting nutrient budget within the lake, the primary productivity estimated to be $0.41 \text{ mol C m}^{-2} \text{ d}^{-1}$.

796 785.4 $\mu\text{mol m}^{-2} \text{ d}^{-1}$ and 3.2 $\mu\text{mol m}^{-2} \text{ d}^{-1}$ for DIN and DIP, respectively. This study also
797 implicates that LGD constitutes relatively small portion of the proglacial hydrologic
798 partitioning, however, delivers nearly a half of the nutrient loadings to the proglacial
799 lake.

800 This study presents the first attempt to quantify LGD and the associated nutrient
801 loadings to the proglacial lake of QTP. To our knowledge, there is almost no study on
802 the groundwater-lake water interaction in the high altitude proglacial lakes in QTP.

803 This study demonstrates that ^{222}Rn based approach can be used to investigate the
804 groundwater dynamics in the high altitude proglacial lakes. The method is

805 instructional to similar studies in other proglacial lakes in the QTP and elsewhere. For

806 a comprehensive understanding the hydrological and biogeochemical dynamics in the

807 QTP, interdisciplinary and multi-approach integrated studies are in great need. Of

808 particular importance are the lake hydrology and groundwater surface water

809 interaction studies based on multiple approaches such as remote sensing products,

810 long term and high resolution observation of climatological parameters and isotopic

811 data.

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829 Supporting data are included as in the files of supplementary information 2 and 3;
830 Climatological data are purchased through <http://www.weatherdt.com/shop.html>; any
831 additional data may be obtained from L.X. (email: xinluo@hku.hk);

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1220 **Figure captions**

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1222 **Figure 1** The geological and topographic map of the Yellow River Source Region,
1223 Nianbaoyeze glacial mountains (a), and the sampling settings of the Ximen Co Lake
1224 (b), with the bathymetry map of the lake (d). (c) Photograph of the Ximen Co Lake
1225 and the surrounding geomorphic settings looking northeast direction on 22 Aug 2015,
1226 showing the late-laying snowpack in the U-shaped valleys of the north part of
1227 Nianbaoyeze MT.

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1229 **Figure 2** The climatological parameters (wind speed, air temperature, and
1230 precipitation) in the Aug, 2015 recorded from Jiuzhi weather station.

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1232 **Figure 3** The temporal distributions of ^{222}Rn (a), water temperature (b), water level
1233 fluctuation recorded by the divers (c), and hourly wind speed and air temperature
1234 recorded in Jiuzhi weather station (d).

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1236 **Figure 4** The cross plots of Cl^- versus Na^+ (a), SO_4^{2-} versus Cl^- (b), Ca^{2+} versus Cl^-
1237 (c); The relations of ^2H versus ^{18}O (d), Cl^- versus ^2H (e), and Cl^- versus ^{18}O (f).

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1239 **Figure 5** Cross plots of ^{222}Rn versus DIN (a) and DIP (b).

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1241 **Figure 6** The conceptual model of ^{222}Rn transient model (a), and three endmember
1242 model (b).

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1244 **Figure 7** The results of the final LGD derived from ^{222}Rn transient model.

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1246 **Figure 8** The hydrologic partition of the proglacial lake of Ximen CO (a), and the
1247 budgets of DIN (b) and DIP (c).

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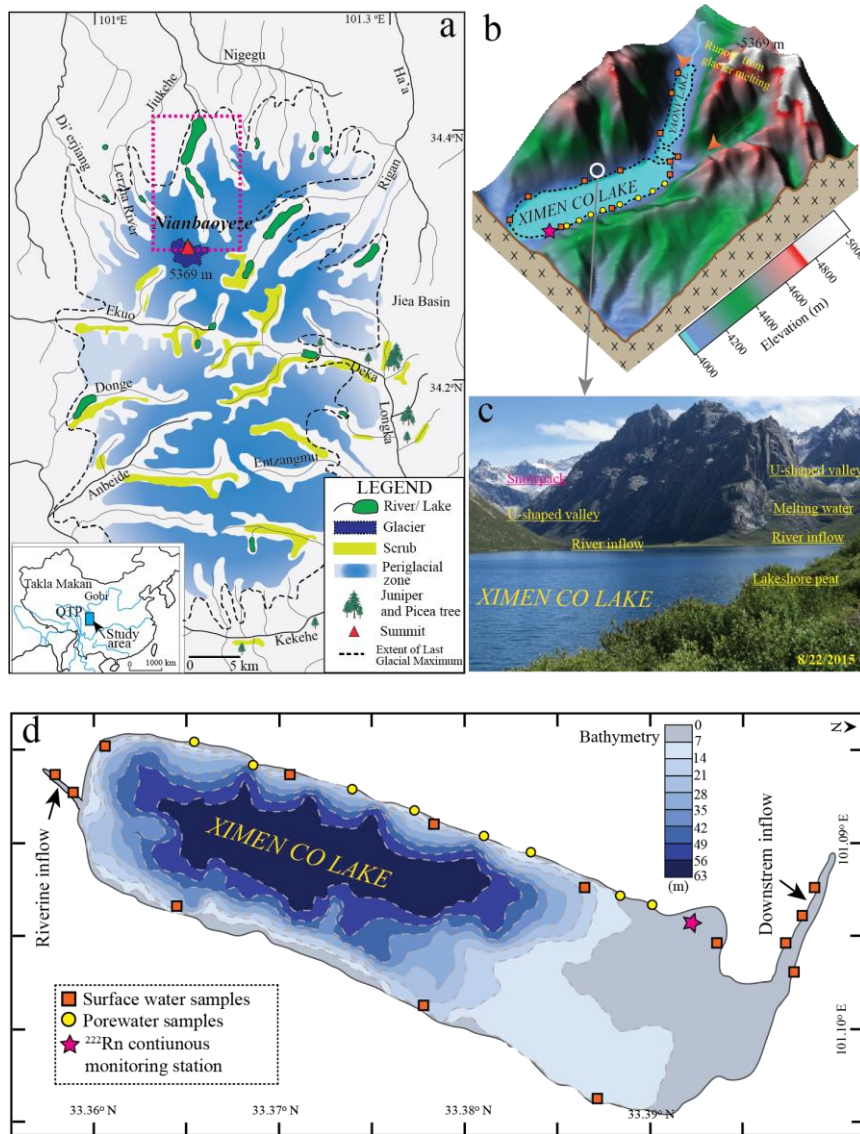
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1258 Figure 1



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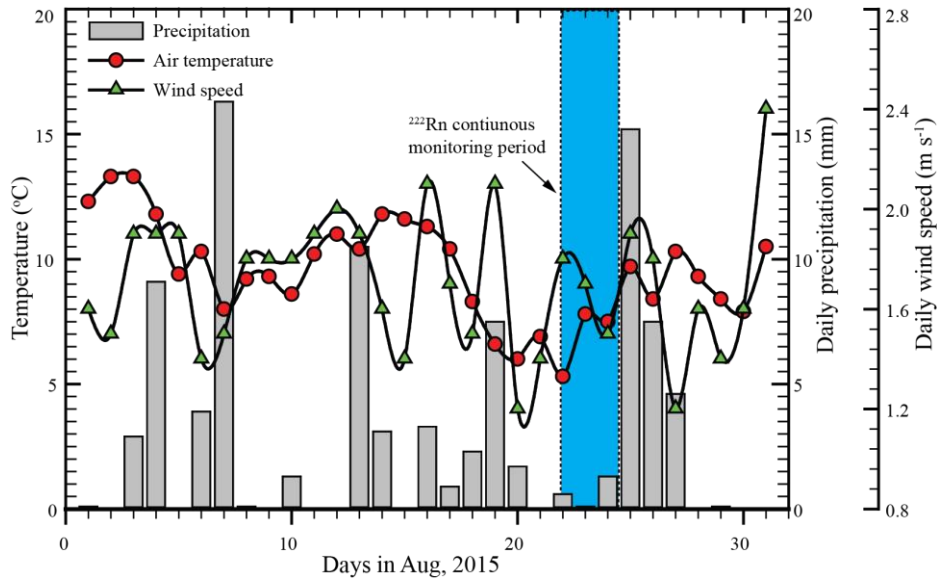
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1264 Figure 2



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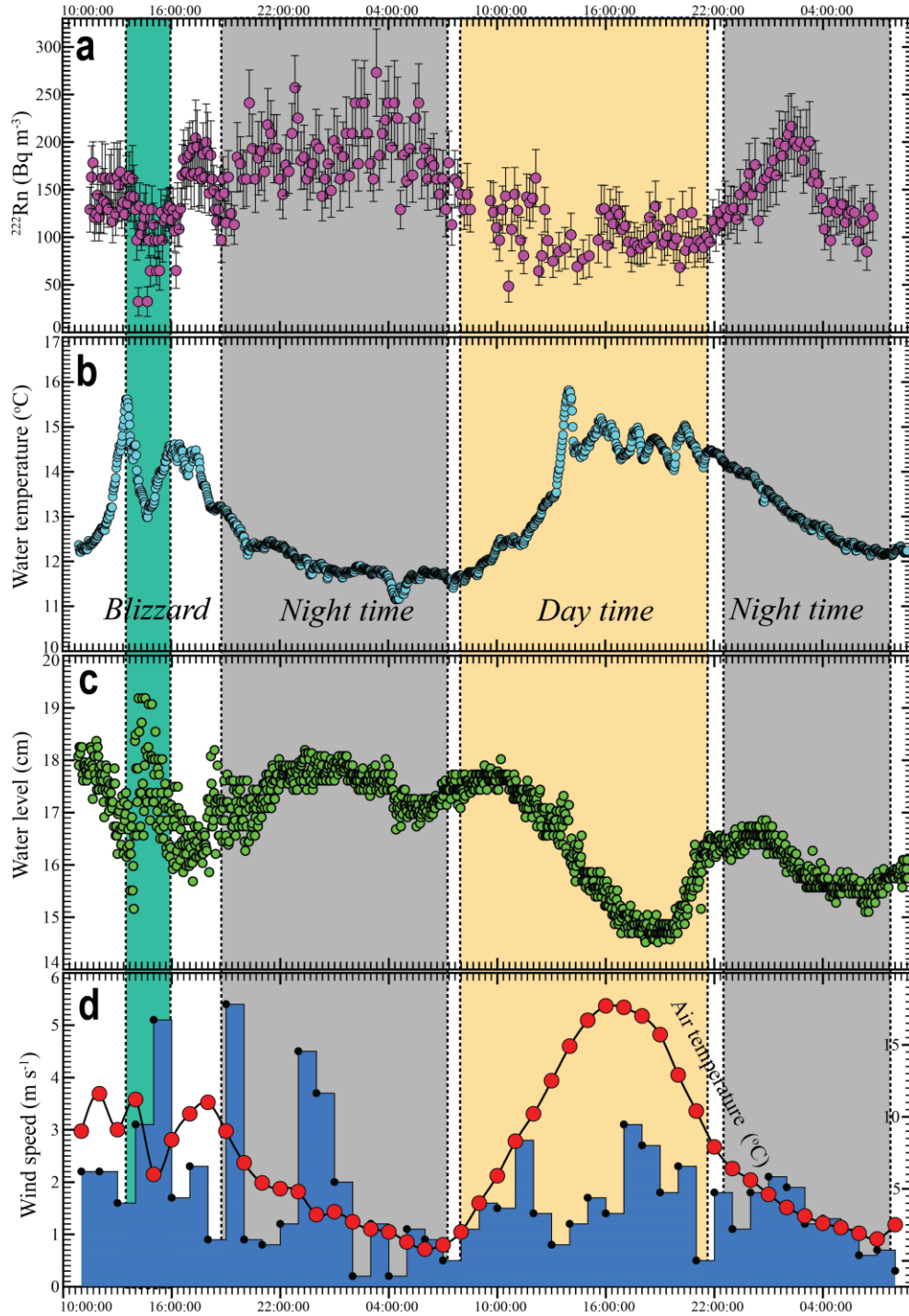
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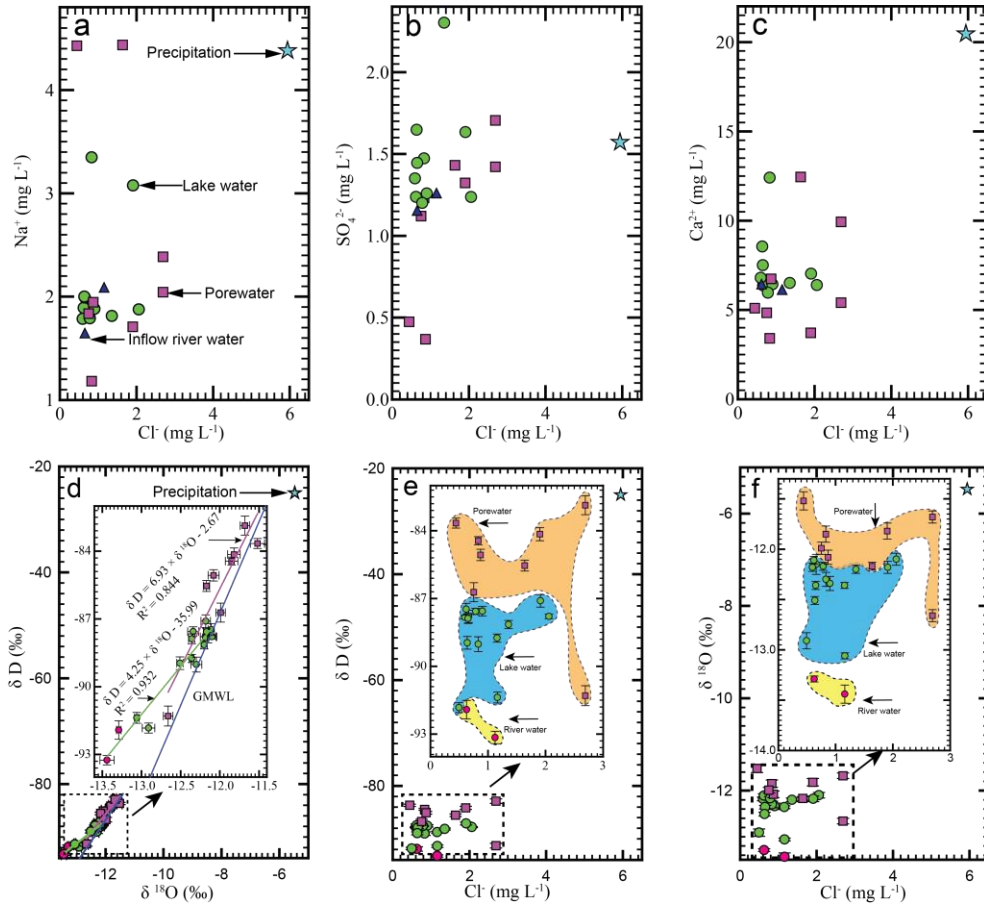
1278 Figure 3



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1281 **Figure 4**



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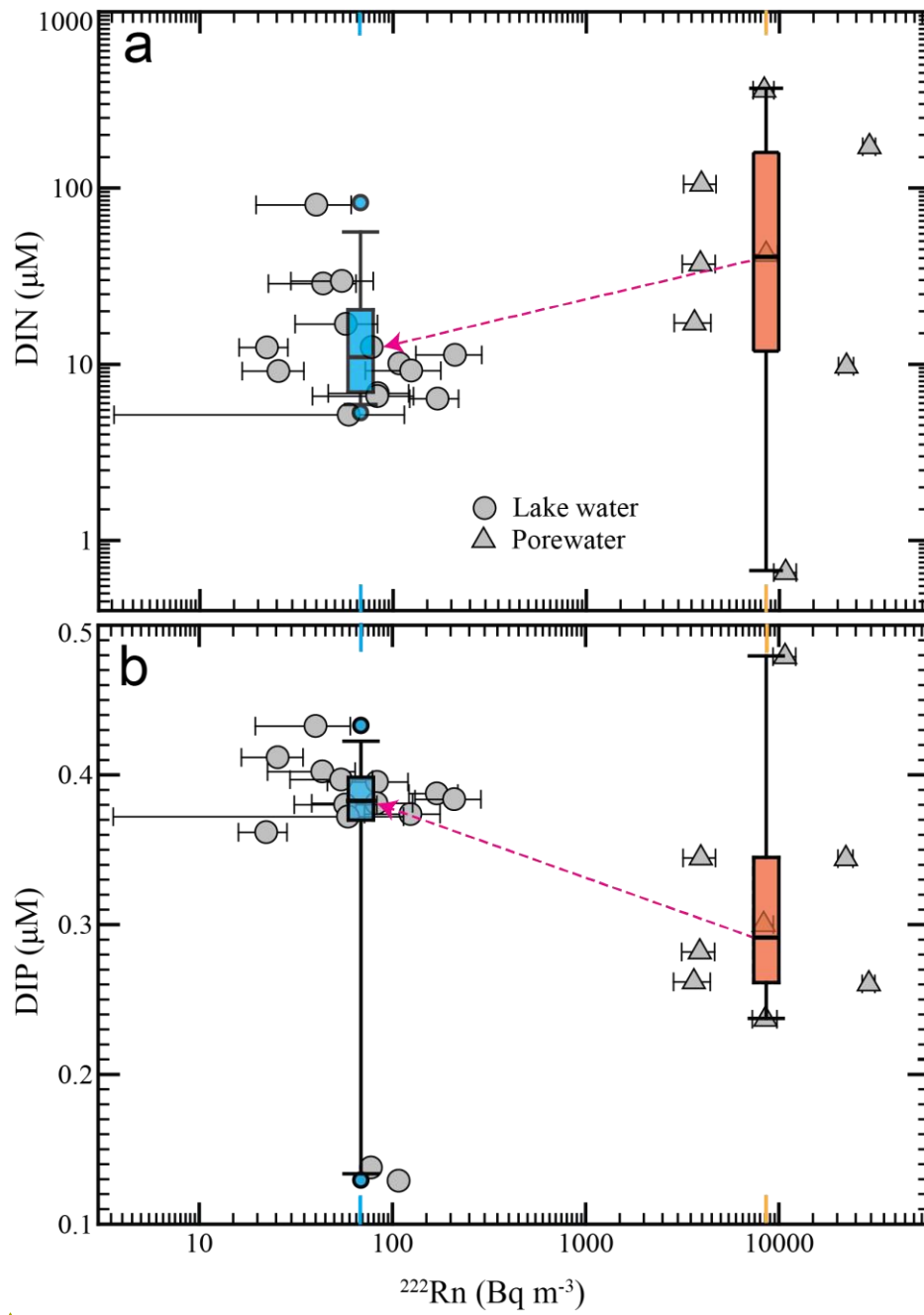
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1290 **Figure 5**



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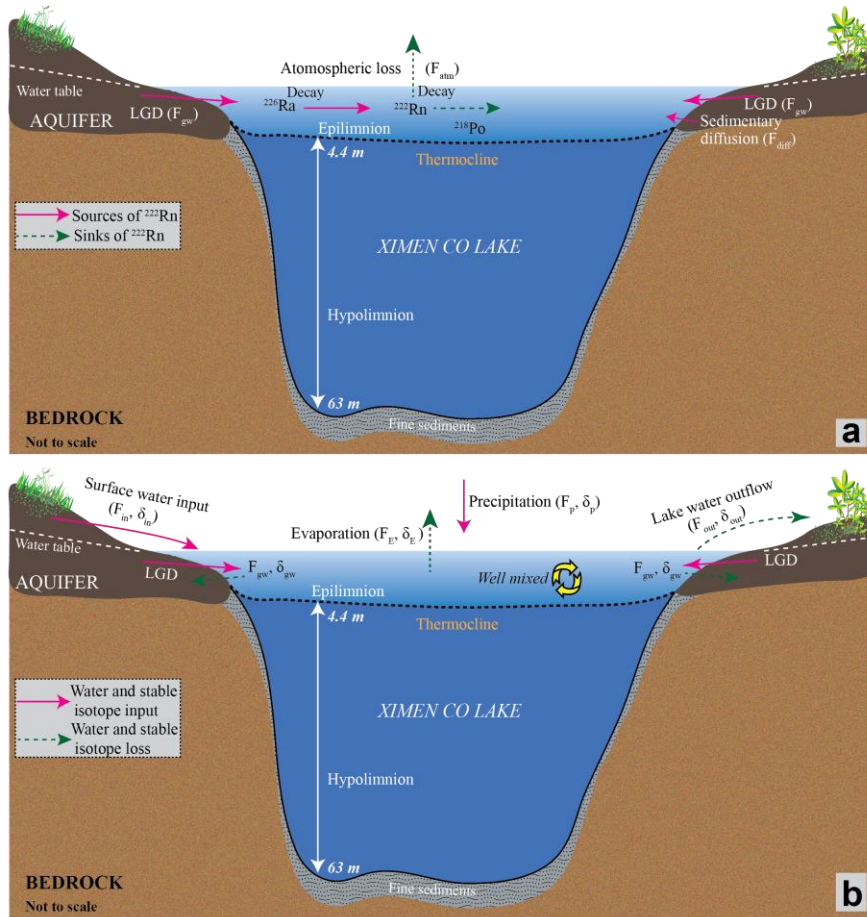
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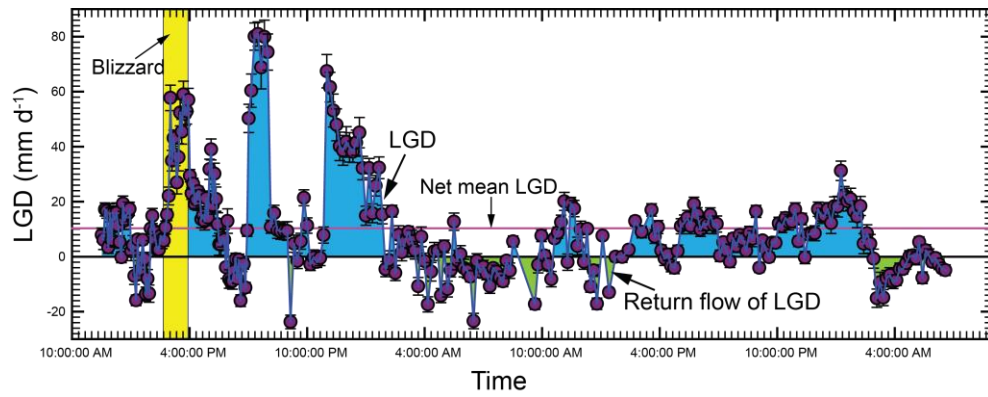
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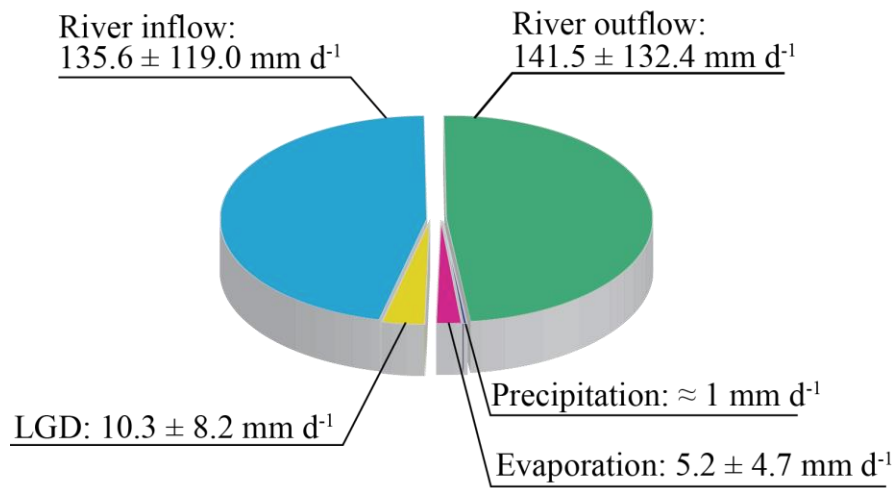
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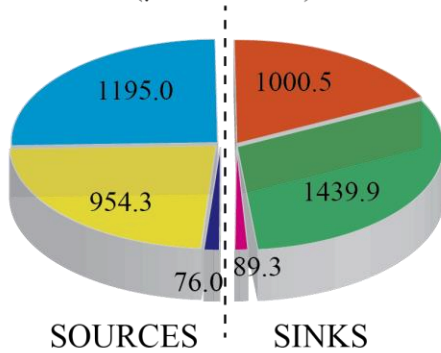
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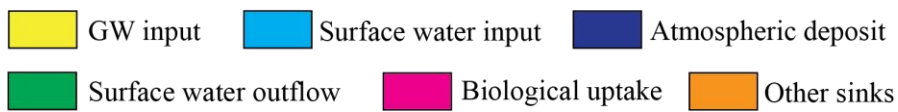
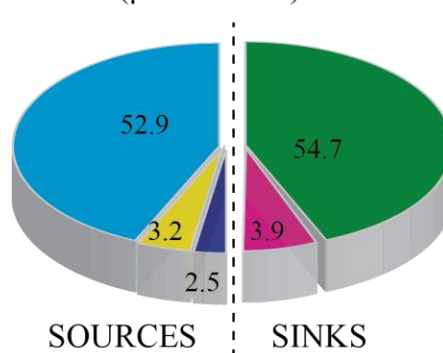
a Hydrologic partitioning



b DIN ($\mu\text{mol m}^{-2} \text{d}^{-1}$)



c DIP ($\mu\text{mol m}^{-2} \text{d}^{-1}$)



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