

1 **Characterizing the Potential for Drought Action from**
2 **Combined Hydrological and Societal Perspectives**
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14
15 **Abstract**

16 Drought is a function of both natural and human influences, but fully
17 characterizing the interactions between human and natural influences on drought remains
18 challenging. To better characterize parts of the drought feedback loop, this study
19 combines hydrological and societal perspectives to characterize and quantify the potential
20 for drought action. For the hydrological perspective, we examine historical groundwater
21 data, from which we determine the decadal likelihoods of exceeding hydrologic
22 thresholds relevant to different water uses. Stakeholder interviews yield data about how
23 people rate the importance of water for different water uses. We combine these to
24 quantify the Potential Drought Action Indicator (PDAI). The PDAI is demonstrated for a
25 study site in south-central Oklahoma, where water availability is highly influenced by
26 drought and management of water resources is contested by local stakeholders. For the
27 hydrological perspective, we find that the historical decadal likelihood of exceedance for
28 a moderate threshold associated with municipal supply has ranged widely: from 23% to
29 75%, which corresponds well with natural drought variability in the region. For the
30 societal perspective, stakeholder interviews reveal that people value water differently for
31 various uses. Combining this information into the PDAI illustrates that potential drought
32 action increases as the hydrologic threshold is exceeded more often; this occurs as
33 conditions get drier and when water use thresholds are more moderate. The PDAI also
34 shows that for water uses where stakeholders have diverse views of importance, the
35 PDAI will be diverse as well, and this is exacerbated under drier conditions. The
36 variability in stakeholder views of importance is partially explained by stakeholders'
37 cultural worldviews, pointing to some implications for managing water when drought
38 risks threaten. We discuss how the results can be used to reduce potential disagreement
39 among stakeholders and promote sustainable water management, which is particularly
40 important for planning under increasing drought.

41
42 **Key words**

43 Drought feedbacks, hydrologic drought, climate variability, social perception, cultural
44 worldviews, water management

46 1. Introduction

47 Drought can pose significant challenges to meeting the water needs of society and
48 ecosystems, which has led to increased interest in understanding and managing drought
49 risk now and into the future (e.g., Georgakakos et al. 2014). There are many definitions
50 of drought, with the classic definitions including meteorological, hydrological,
51 agricultural, and socioeconomic (Wilhite and Glantz 1985). Similarly, many different
52 drought indices have been developed (Mishra and Singh 2010). The main driver of
53 drought in most definitions and indices of drought is natural climate variability (Van
54 Loon 2016a), which is where efforts to improve prediction and modeling have focused
55 (see Mishra and Singh 2011 and references therein). Even with advances in drought
56 prediction, drought remains one of the most expensive hazards affecting the United States
57 (NCDC 2015), reinforcing the idea that social factors must also be considered for drought
58 planning (Wilhite and Buchanan-Smith 2005; Bachmair et al. 2016).

59 The need for more proactive drought planning has led to increased interest in the
60 development of drought management plans (e.g., Wilhite et al. 2000, Wilhite et al. 2005,
61 Knutson et al. 1998). Drought risk management requires identifying drought indicators
62 and triggers (Steinemann and Hayes 2005), which can be developed and evaluated using
63 stakeholder processes to make them useful for decision-making (Steinemann and
64 Cavalcanti 2006, Steinemann et al. 2015). Further, the need to better link drought indices
65 with impacts has been recognized (Bachmair et al. 2016). Frameworks to link drought
66 indicators directly with impacts are emerging (Bachmair et al. 2016; Stagge et al. 2015 ;
67 Towler and Lazrus 2016), though there is still a need for more systematic impacts
68 monitoring (Lackstrom et al. 2013). Ostrom (1990) found that assessments that can
69 account for how people value, perceive, and make decisions about resources such as
70 water, particularly when water is scarce, are critical for guiding policies that meet
71 management goals and stakeholder needs, and thus promote sustainable management of
72 water resources. Dessai and Sims (2010) explored public perceptions of drought and
73 climate change to understand barriers to action and paths towards sustainable
74 management. Lazrus (2016) examined how stakeholders perceive drought and how
75 drought intersects with their cultural processes.

76 Recent work has highlighted how the natural and human causes of drought are
77 intertwined, and that researchers must consider both in any examination of drought (Van
78 Loon 2016a). This general notion has been echoed in the hydrologic science literature
79 (Wagener et al. 2010), as well as the natural hazard (Jones and Preston 2011) and climate
80 change literature (Oppenheimer et al. 2014). This has also motivated the new science of
81 socio-hydrology, which explores the dynamics and co-evolution of human and water
82 systems (Sivapalan 2012). Van Loon et al. (2016b) describe a new framework that
83 explicitly acknowledges the human dimension of drought. They outline several research
84 gaps, including a gap in our understanding of the human feedbacks on drought.

85 Understanding human feedbacks on drought is important, but has not been well
86 studied, partially because of its complexity and potential for nonlinear feedbacks (Van
87 Loon et al. 2016b). Drought feedbacks can be influenced by many factors, for example,
88 through science and technology (Polsky and Cash 2005), historical lessons learned
89 (McLeman et al. 2014), and management strategies (Maggioni 2015). Further, feedbacks
90 may be positive, i.e., the drought is made worse, or negative, the drought condition is
91 alleviated (Pulwarty 2003; Tisdeman et al 2018). In addition, these interactions and

92 feedbacks can result in changing the normal drought reference baseline (Van Loon et al.
93 2016b). However, fully characterizing the feedback loops between human and natural
94 influences on drought remains challenging.

95 The goal of this paper is to provide an experimental methodology towards a better
96 characterization of several components of the drought feedback loop, specifically to gain
97 understanding on how and why people might take action in response to drought. To this
98 end, we develop an indicator to characterize how natural influences on drought inform
99 *potential* human actions on drought. We use the term “potential”, since in this study, we
100 do not have the data to validate whether or not human actions were actually taken as a
101 result of these natural drought influences. In this investigation, we characterize the
102 natural influences by taking a hydrological perspective on drought (Van Loon et al.
103 2016b); specifically, we examine the exceedance of relevant thresholds from historical
104 hydrologic data. For the societal perspective, we examine stakeholder input from
105 interviews, specifically how stakeholders rated the importance of water for different uses.
106 In our attempt to better characterize the potential for drought action, we combine the data
107 from the hydrological and societal perspectives, developing a new, derivative product
108 that we call the Potential Drought Action Indicator (PDAI). Here, by “action”, we
109 generally mean some effort towards drought mitigation. Though we do not directly
110 validate the PDAI, we are able to interpret the findings to provide insights to water
111 management policy using additional interview data on stakeholder worldviews and social
112 science theory; this is unique in that it allows us to investigate the theoretical
113 underpinnings that are not typically explored in drought risk studies or stakeholder
114 processes.

115 We demonstrate the PDAI through a place-based assessment of drought risk in south-
116 central Oklahoma, where water availability is highly influenced by drought and
117 management of water resources is contested by local stakeholders; we provide some
118 background and describe this study site in section 2. Section 3 outlines the methodology:
119 sections 3.1 and 3.2 outline the details of the methods used to assess the hydrological and
120 social perspectives, respectively. Details about how the PDAI is developed is provided in
121 section 3.3. In section 3.4, we describe additional interview data on the stakeholder
122 worldviews and provide an overview of the social science theory. Results for our study
123 site are shown in section 4.

124 125 2. Background and Study Site

126 The goal of this paper is to gain insights into the potential for human action on drought,
127 and one suggested way to do this is to study a particular water system in detail (Sivapalan
128 2012). As such, the PDAI is developed and demonstrated for the Arbuckle-Simpson
129 Aquifer (ASA), a groundwater resource that underlies an area of about 520 square miles
130 (1350 square kilometers) in south-central Oklahoma. This area is part of Oklahoma
131 Climate Division 8 (Karl and Koss, 1984), which is one of the 344 climate divisions that
132 the United States is divided into for reporting purposes, based on climate as well as
133 several other considerations (Guttman and Quayle 1996). The ASA provides water for
134 municipal supply, ranching, and mining, and is also the source of local springs and
135 streams that support wildlife, recreation, and tourism. Drought is part of the region’s
136 history (Silvis et al. 2014), and the ASA is recharged by rainfall, thus making it
137 susceptible to climate variability and change. The ASA has been the center of a water

138 management dispute that arose in 2002 when landowners began negotiations to sell their
139 groundwater to an area outside of the ASA, near Oklahoma City. The landowners’
140 actions were quickly contested by a local environmental group, the Citizens for the
141 Protection of the Arbuckle-Simpson Aquifer (CPASA; Shriver and Peaden 2009; Lazrus
142 2016), which led to a moratorium in 2003 that suspended any activities to remove water
143 from the basin until a hydrological study could be conducted. The study included a water
144 balance, hydrogeological study, and groundwater model of the aquifer; the study shows
145 that although water is extracted, groundwater use from the aquifer is relatively small, and
146 that the groundwater-fed streamflow discharge is mostly related to rainfall recharge
147 (Christenson et al. 2011). This was followed by a ruling that reduced the amount of water
148 that could be removed from the aquifer annually by an order of magnitude. This further
149 exacerbated the tensions between the landowners who see the decision as an
150 encroachment on their individual property rights, and CPASA and other community
151 members who see the reduction as a way to protect local water resources (Lazrus 2016).
152 The ASA’s susceptibility to drought, as well as its diverse community and contentious
153 management issues, make it an ideal site for exploring the potential for feedback on
154 drought.

155 In this paper, we explicitly combine hydrological and societal perspectives, but
156 historically, these two perspectives would likely be examined in isolation. In fact, this
157 work builds upon and extends two previous studies that focused on the same ASA case
158 study, but were disciplinary in nature: Lazrus (2016) and Towler and Lazrus (2016).
159 Lazrus (2016) describes results of stakeholder interviews collected for the ASA; it offers
160 an anthropological perspective, examining how stakeholders perceive drought and how
161 those perceptions intersect with their cultural processes. Lazrus (2016) was motivated by
162 the hydrological context of the ASA, but did not engage directly with any quantitative
163 meteorological or hydrological analysis. On the other hand, Towler and Lazrus (2016)
164 take a hydrological perspective, developing a generalized framework that links
165 meteorological drought indices with hydrologic threshold exceedances that are relevant to
166 ASA stakeholders. To identify some of the hydrological thresholds and provide social
167 context, Towler and Lazrus (2016) draw on qualitative insights gathered from the
168 interviews, but do not directly incorporate any of the quantitative interview results into
169 the analysis. In this paper, we extend these two studies to offer a novel, quantitative,
170 interdisciplinary approach, that results in a derivative product, adding value to the
171 preceding studies. Although the PDAI is experimental, conducting this type of study is
172 critical, given the grand challenge of engaging in interdisciplinary research at the climate-
173 water-society interface (McNeeley et al. 2011).

174

175 3. Methodology

176

177 Figure 1 provides the conceptual overview of the study methodology, which is
178 detailed in the subsequent sections.

179

180 3.1 Hydrologic Perspective: Threshold Exceedance

181 To characterize natural influences on drought, we examine drought from a
182 hydrological perspective. Taking a hydrological, rather than meteorological, perspective
183 is advocated by Van Loon et al (2016b), given the closer connection of surface water and

184 groundwater with societal use and management. Here, we use a groundwater (GW) well
185 that has relevance to the community (Towler and Lazrus 2016), has a long available
186 record, and is monitored by water managers in the community: the USGS Fittstown well
187 (USGS 343457096404501). We use data from the beginning of the GW monitoring
188 record through the year the interviews were conducted, which corresponds to 1959-2012.
189 Details of this dataset can be found in Towler and Lazrus (2016).

190 To connect the hydrologic perspective with human action, we examine the historical
191 groundwater data in terms of decision relevant thresholds (Jones 2001). From Towler and
192 Lazrus (2016), we identify two main thresholds relevant to water uses asked about in the
193 interviews (see section 3.2). The first threshold is called a “moderate” threshold: This is a
194 groundwater level of 111 feet below the surface, which is decision relevant because it is
195 when the aquifer begins to be closely monitored because of potential impacts to
196 municipal supply. The second threshold is the “severe” threshold: this is when the
197 groundwater level lowers further, to 117 feet below the surface, which is the level at
198 which artesian springs in the area have minimal flow or stop flowing altogether, affecting
199 uses such as wildlife and recreation. For illustrative purposes, we also look at an
200 “extreme” threshold of groundwater levels to 120 feet below the surface, which have
201 been experienced in the aquifer and further the likelihood of minimal or stopped spring
202 flows (see Figure 2 in Towler and Lazrus 2016).

203 To quantify the threshold exceedance, we calculate the percent frequency of
204 exceedance¹ for each threshold in the historical record. To calculate the exceedance
205 frequency, a time window needs to be selected; we initially examined 5-, 10-, 15-, and
206 20- year windows. Specifically, we calculate the number of months during each x-year
207 running window that the threshold was exceeded across the available record. For
208 example, for the 10-year window, it would be 1959-1968, 1960-1969, etc., all the way to
209 2003-2012. Henceforth, we refer to this as the groundwater drought likelihood. For the
210 analysis, we look at the three most recent decades (i.e., 1983-1992, 1993-2002, and 2003-
211 2012), as well as the highest, median, and lowest decades in terms of the groundwater
212 drought likelihood.

213 We also calculate the Pearson’s correlation coefficient (r) values between the decadal
214 likelihoods and several drought indices for the area. Specifically, we correlate the decadal
215 likelihoods with 10-year running averages of several drought indicators from different
216 categories. As a measure of agricultural drought, we use the well-known Palmer Drought
217 Severity Index (PDSI; Palmer, 1965) that is based on a water balance of precipitation,
218 soil moisture, potential evapotranspiration, and runoff. We also look at the Standardized
219 Precipitation Index (SP), which only considers the effect of precipitation variability on
220 drought (McKee et al., 1993). The SP can be calculated to consider different time scales:
221 for example, the 1-month SP (SP01) considers short-term conditions, and the 24-month
222 SP (SP24) considers longer-term conditions (i.e., precipitation from the last 2 years). We
223 use the 6-month SP (SP06). To measure hydrological drought, we use the Palmer
224 Hydrological Drought Index (PDHI; Palmer 1965), which is a modification of the
225 original PDSI to account for longer-term dryness that affects water storage, streamflow
226 and groundwater. NOAA’s National Climatic Data Center provides this historical data for

¹ We note that groundwater threshold levels are negative; so here we define “exceedance” as going below (more negative) than the threshold.

227 United States climate divisions; we downloaded monthly data from 1959 to 2012 for
228 Oklahoma Climate Division 8 from
229 <http://www7.ncdc.noaa.gov/CDO/CDODivisionalSelect.jsp>.

230

231 3.2. Social Perspective: Stakeholder Importance Ratings

232 To understand how community members in the ASA region might respond to natural
233 influences drought, we use stakeholder interview data from a previous investigation
234 (Lazrus 2016). Stakeholder interviews (n=38) were conducted in the summer of 2012,
235 following a significant drought in 2011. Interviewee selection followed a targeted
236 snowball sampling strategy whereby interviewees were selected based on their
237 involvement in the ASA water management negotiations, their dependence on or
238 engagement with water resources – for example, in ranching or recreation operations –
239 and recommendations from other interviewees.

240 For this study, we examined the question: “How do people perceive the importance
241 of water for various uses?” We make the assumption that the more important water is
242 perceived to be for a particular use, the greater the potential will be for taking action - in
243 this case, conserving water for that use.

244 To understand the importance of water for various uses, interviewees were asked
245 how important (on a Likert scale of 1-5, 5 being very important) water resources are in
246 their community for: a) People’s livelihoods, b) Recreational activities, c) Spiritual
247 fulfillment, d) Cultural practices, e) Habitat for plants and animals, and f) Availability of
248 drinking water. Data from these questions was used directly and called “importance
249 ratings”, which were integrated into the PDAI (see section 3.3).

250

251 3.3. Creating the Potential Drought Action Indicator (PDAI)

252 We express the PDAI as a function, f , of (i) the decadal probability (P) of the
253 groundwater level, Z , exceeding the hydrologic threshold, z , and (ii) the importance
254 ratings (I):

$$255 \text{PDAI} = f(P(Z < z), I) \quad \text{(Equation 1)}$$

256 Here, we define f as the product (i.e., multiplication) of the two explanatory terms:

$$257 \text{PDAI} = P(Z < z) \times I \quad \text{(Equation 2)}$$

258 Multiplying two terms to create a new indicator is based on a frequently used definition
259 of risk, which combines the likelihood of an event and its consequence (Jones and
260 Preston 2011). However, we point out that the form of f is flexible, e.g., it could be
261 additive, etc.

262 Here, $P(Z < z)$ is the decadal likelihood of exceeding a particular threshold (e.g.,
263 moderate or severe), per section 3.2. This is multiplied by the importance of water for a
264 particular use, which is directly derived from the stakeholder ratings (section 3.1).
265 Essentially, the importance ratings are used as a weight function to modulate the
266 likelihood of exceedance. In this definition, a lower PDAI equates to less potential for
267 action and higher PDAI indicates greater likelihood of action. The PDAI was calculated
268 for all 6 of the water uses asked about in Section 3.2.

269

270 3.4. Social Perspective: Stakeholder Worldviews

271 We are also interested in exploring *why* people perceive the importance of water
272 for various uses differently. For this, we again interrogate the interview data using a

273 social science theory called The Cultural Theory of Risk (CTR; Douglas 1966; McNeeley
274 and Lazrus 2014). According to CTR, people hold different cultural worldviews about
275 how society should be organized and how society and nature should interact. CTR
276 predicts that people will perceive risks and consequences from hazards when their
277 worldview is challenged. According to this understanding, perceptions are as much about
278 social organization as they are about the physical hazard. Their worldview will also guide
279 their preference for different risk management strategies, or in this case drought actions,
280 making it relevant to our PDAI results. Two of the worldviews described by CTR are
281 individualism and egalitarianism. These represent idealized categories and are useful
282 heuristics, but in reality, people may adhere to some elements of the cultural worldviews
283 more than others. People with individualist views favor weak social bonds and have little
284 need for social structure, preferring individual competition and market-based transaction
285 strategies. For them, nature is a bountiful resource robust to human uses and therefore
286 may not need to be managed for conservation. People with egalitarian views favor strong
287 social bonds and collective decision-making processes. For them, nature is fragile and
288 easily impacted by humans and so must be carefully managed to avoid catastrophe
289 (Thompson et al. 1990). By identifying the cultural processes that lead people to
290 recognize risks and perceive consequences, CTR also helps to diagnose why
291 disagreements arise over risk management; that is, disagreements may arise between
292 constituent groups holding different worldviews when management strategies do not
293 reflect elements of each constituent's predominant worldview (Verweij et al. 2006).

294 To this end, we examined how peoples' importance ratings from section 3.2 were
295 related to their worldviews. If so, it would help us to understand how the PDAI could be
296 operationalized – that is, might people respond more favorably to water management
297 strategies that reflected their own management preferences based on their cultural
298 worldviews? For the CTR, interview questions about worldview used previously tested
299 measures for individualism and egalitarianism developed by Smith and Leiserowitz
300 (2014) as well as additional questions informed by CTR that reflected the particular water
301 management context of the ASA; all questions can be seen in Tables 1 and 2 of Lazrus
302 (2016). These questions asked people whether they strongly agreed, agreed, neither
303 agreed nor disagreed, disagreed, or strongly disagreed (on a 5 point Likert scale) to a
304 series of statements. Responses were summed for each interviewee to determine a value
305 for individualism or egalitarianism. Follow-up open-ended questions allowed
306 interviewees to elaborate on their worldview preferences and importance ratings.

307

308 4. Results

309

310 4.1. Threshold Exceedance Likelihood

311 Figure 2 shows the historical monthly groundwater time series, including the
312 moderate threshold (111 feet below the surface) and severe threshold (117 feet below the
313 surface) introduced in Section 3.2. Groundwater drought likelihood is calculated as the
314 number of months within each 5-, 10-, 15-, and 20-year running window that the level
315 went below a particular threshold. Drought likelihoods for the selected time windows (5-,
316 10-, 15-, and 20-years) are shown in Figure 3. Results for each time window follow
317 similar patterns, though as expected, the shorter the time window, the greater variability
318 in the likelihood. We selected the 10-year running window for calculating the PDAI (e.g.,

319 Figure 3b), as it strikes a balance between shorter time windows that have high variability
320 (e.g., 5-year windows) and longer time windows (e.g., 15-, and 20-years) where much of
321 the variability is smoothed out. Figure 3b shows the decadal likelihood for the moderate
322 and severe threshold. As expected, the higher the threshold, the higher the likelihood of
323 exceedance (i.e., a moderate threshold is exceeded more often than the severe threshold).
324 Further, the likelihoods are correlated ($r=.94$) and significant at the 99% confidence level.
325 We also point out the very close association between the groundwater threshold
326 exceedance likelihoods and selected drought indices for the region (i.e., Oklahoma south-
327 central climate division 8): Table 1 shows that for meteorological (SP06), agricultural
328 (PDSI), and hydrological (PDHI) drought indices, the correlations with the moderate
329 threshold exceedance is $>-.9$ and with the severe threshold is $>-.8$. This underscores the
330 notion that for this case study, the hydrological perspective is a good indicator of the
331 natural influences on drought. This is the case because water extraction in the area is
332 relatively low and the groundwater levels are very closely related to rainfall recharge
333 (Christenson et al., 2011). However, we note that that this may not be the case for other
334 groundwater aquifers that are more affected by human extraction (e.g., Tarhule and
335 Bergey 2006) or aquifers with different properties (e.g., slower hydrologic responses due
336 to increased propagation times). This point is further discussed in the Discussion (section
337 4.5 Future Enhancements).

338 Table 2 shows the exceedance likelihoods of select decades from the historical record
339 for both moderate and severe. First, we look at the three most recent decades (i.e., 1983-
340 1992, 1993-2002, and 2003-2012), in which relatively wet, average, and dry conditions
341 occurred. For 2003-2012, the moderate threshold was exceeded 61% of the time, which
342 we refer to as the “dry/recent” decade. In the next most recent decades, the exceedance
343 likelihood decreased to 35% (1983-1992) and 31% (1993-2002), which we refer to as
344 “average/recent” and “wet/recent” decades, respectively. To put into context, for the
345 moderate threshold, the decade with the lowest exceedance likelihood was 23% (1985-
346 1994), which we call the “wettest” decade, and highest exceedance was 75% (1959-
347 1968), or “driest” decade. Results follow similar patterns for the severe threshold (Table
348 2).

349

350 4.2. Stakeholder Importance Ratings

351 Stakeholder interviews reveal that there is more consensus on the importance of water
352 for some water uses than others (Figure 4). On average, water was deemed most
353 important for drinking water, followed closely by habitat for wildlife, and supporting
354 livelihoods. The importance of water for these uses was similar for most stakeholders
355 interviewed, as evident by the tightness of the box plot (Figure 4). On the other hand,
356 there was a spread in responses for recreation, cultural practices, and spiritual fulfillment.
357 Some of the spread in responses on these measures may be due to how interviewees
358 interpreted the water uses (Lazrus 2016).

359 The spread in responses indicates that different stakeholders place different levels of
360 importance on some water uses, such as water for recreation which shows a broader
361 spread than water for drinking water, habitat, or livelihood. For example, one interviewee
362 underscored the importance of water, describing that “Murray County is one of the top
363 tourist attractions with Arbuckle Lake and Chickasaw National Recreation area. So
364 water is the absolute key” (Interview 1). Demonstrating a very different perspective,

365 another interviewee noted that “Recreation and water are not critical to me. I mean in
366 this part of the world, they don’t necessarily go hand-in-hand because it’s a relatively dry
367 place, and there are not that many places to really go and play in the water” (Interview 5).
368

369 4.3 Potential Drought Action Indicator (PDAI)

370 The PDAI is calculated for all of the water uses (Figure 5). Here, the top row shows
371 results for water uses using the moderate threshold (Figure 5a-c) and the bottom row
372 shows results for water uses using the severe threshold (Figure 5d-f). Because the results
373 across the rows are quite similar, we will focus on the results for drinking water (Figure
374 5a) and then recreation (Figure 5d).

375 First, we focus on drinking water, which is an example of a water use which exhibited
376 more consensus among interviewees. For drinking water, to calculate the PDAI, we use
377 the moderate threshold, since this is the threshold at which municipal supply is monitored
378 (see Section 3.1). Figure 5a shows the PDAI for drinking water for the different drought
379 conditions (e.g., wet/recent, dry/recent, etc) from Table 2. Results are shown as empirical
380 Cumulative Density Functions (eCDFs) to reflect the discrete nature of the importance
381 ratings. In the eCDFs, the vertical lines represent the PDAI values, and the horizontal
382 lines represent the percentage of data that are equal to or less than that value. In Figure
383 5a, as the eCDF moves across drought conditions from wettest to driest, the PDAI shifts
384 towards higher values, reflecting the increased potential for action under drier conditions.
385 Specifically, the wettest decade has an average PDAI value of 1.1, and the driest decade
386 has an average PDAI value of 3.7. Given the stakeholder consensus on the importance
387 for drinking water, for each drought condition there is very little range – that is, the
388 eCDFs are fairly vertical. Results are similar when the Moderate threshold is used for the
389 other two water uses, habitat and livelihood, that showed strong consensus (Figure 5b and
390 5c).

391 Next, we focus on the PDAI for Recreation, a water use that shows diverse
392 importance ratings from stakeholders (Figure 5d). For recreation, to calculate the PDAI,
393 we use the severe threshold, since that is the threshold at which artesian springs have
394 minimal flow or no longer flow (see Section 3.1). Figure 5d shows the PDAI for
395 recreation for the select decadal drought conditions, using the severe threshold
396 likelihoods from Table 2. Similar to drinking water, we see that as we move from wetter
397 to drier decades, the PDAI also increases; for example, from wet/recent to dry/recent, the
398 average PDAI values are 0.3 and 1.5, respectively. However, given the stakeholder
399 diversity in importance ratings, as we move towards drier conditions, the PDAI becomes
400 more diffuse, spanning a great range of values: in the wet/recent, the PDAI spans from
401 .08 to .4, or for 0.32 units of the PDAI scale, and in the dry/recent it spans from 0.4 to
402 1.9, or 1.5 units on the PDAI scale, indicating a wide range in stakeholder appetite for
403 potential action. Interestingly, the wet/recent decade (1993-2002) was also the wettest
404 decade on record, with the groundwater threshold only being exceeded 8% of the time.
405 Results are similar when the Severe threshold is used for the other two water uses that
406 showed diverse ratings, i.e., cultural practices and spiritual fulfillment (Figure 5e and 5f).

407 In Figure 6, we also looked at recreation under the possibility of a new “normal”
408 drought baseline (Van Loon 2016b). It has been suggested that human adaptation to new
409 drought normals can be illustrated by changing thresholds (Vidal et al. 2012; Wanders et
410 al. 2015); here, we show how this could influence the PDAI. To this end, we look at a

411 more extreme threshold that has been identified for Recreation (i.e., GW levels below
412 120 feet, see section 3.1), under the dry/recent period: the eCDF curve shifts back to the
413 left, towards lower action potential, with average PDAI of 0.9, reflecting this new
414 normal. This is relevant given climate change projections that suggest that the ASA will
415 likely become drier in the future (Towler et al. 2016; Liu et al. 2012).

416 Finally, in Figure 7, we narrow our focus to the most recent decade (i.e., dry/recent,
417 2003-2012), and compare both drinking water and recreation with the moderate and
418 severe thresholds, respectively. From Figure 7, we see that drinking water has a higher
419 action potential than recreation: the average PDAI for drinking water is 3, while it is
420 about 1.5 for recreation. This is an artifact of the thresholds selected for each respective
421 water use (i.e., moderate for drinking water and severe for recreation). This makes sense
422 from a human standpoint, since drinking water is a primary consumptive use, and
423 recreation is a more discretionary use. However, this could be more subjective for other
424 water uses (e.g., spiritual fulfillment). Although it may seem counterintuitive at first, we
425 purposely pair the moderate threshold with the primary use to indicate this hierarchy, but
426 this does not mean that exceedance of the severe threshold would not also prompt action
427 (or further action) to ensure adequate drinking water supplies. However, it does make the
428 assumption that for a more discretionary use, like recreation, action would not be
429 prompted until this severe threshold was exceeded.

430 Another key point from Figure 7 is that drinking water spans a smaller range on the
431 PDAI scale than recreation, which is more diffuse. Specifically, for drinking water, the
432 eCDF only falls between 2.4 and 3.5; this is due to the agreement across respondents on
433 the importance of water to this use (i.e., Figure 4). On the other hand, the recreation
434 PDAI eCDF covers of a larger range of values – here it spans from 0.4 to 1.9, similarly
435 reflecting the range of stakeholder responses. This shows that for water uses where values
436 are diverse, the appetite for potential action will be diverse as well.

437 In summary, the key points from these results: the PDAI increases with (1) drier
438 decadal drought conditions and (2) water use thresholds that are exceeded more often.
439 Further, it shows that for water uses where perceived importance is diverse among
440 stakeholders, the PDAI will be diverse as well, and this is exacerbated under drier
441 conditions.

442 4.4. Management Implications based on Worldviews

443 To understand the management implications, we need to look at the results alongside
444 of CTR. Results from the CTR questions show that both individualist and egalitarian
445 worldviews were represented by the interviewees (Figure 8) and that some of the spread
446 in the importance responses can be explained by worldview (Table 3). Although the
447 correlations are relatively low, eight out of twelve are statistically significant at the 90th
448 percentile or higher. Further, the sign of each correlation coefficient is opposite between
449 the egalitarianism and individualism measures, indicating that people holding each
450 worldview have opposing importance ratings (Table 3). The water use that showed the
451 highest correlation with worldview was recreation: $r=0.45$ for Egalitarianism and $r=0.40$
452 for Individualism. These correlations provide initial insight about the role of worldview
453 in how people assess the importance of water and, by extension, their appetite for
454 potential drought action.

455 Results from the CTR questions, along with the PDAI, point to some implications for
456 water management policy. CTR posits that disagreement over resource management

457 strategies may arise among constituents with diverse worldviews for two
458 reasons (McNeeley and Lazrus 2014): first, as demonstrated in Table 3, worldviews
459 explain some of the variance in how important people think that local water resources are
460 for different activities - and thus presumably *whether or not* maintaining water for those
461 activities should be prioritized by water management. For example, in recreation, because
462 of the large spread in importance ratings, which can partially be explained by CTR, there
463 is an increase in the PDAI categories from the wet/recent to average/recent to dry/recent
464 decades; this implies that people will disagree on whether or not water should be
465 managed for recreation, potentially leading to disagreements that could hinder sustainable
466 water management. Second, is *how* water should be managed, even when people agree on
467 its importance. In drinking water, there is consensus on importance – even among people
468 with different worldviews – presumably indicating that people agree that water needs to
469 be managed for drinking water. However, because of the different worldviews, there is
470 still potential for disagreement over how it should be managed. That is, those with
471 egalitarian preferences advocate for management that is collectively debated,
472 implemented, and enforced whereas those with individualist preferences favor
473 management that is individually enacted and market-based. We see this in our qualitative
474 data: for example, one interviewee with individualist preferences said: “we have to have a
475 set of rules that everyone understands. And once those rules are set you can’t have a
476 bunch of water Nazis trying to make judgment calls about how someone’s using their
477 water. So, if I can use a certain amount - tell me what that amount is, and then stay the
478 hell out of my business” (Interview 2). The finding shows that disagreement is not solely
479 due to threats to water resources – such as more frequent drought – but rather that it can
480 also arise from disagreement about the strategies designed to manage water and address
481 drought.

482 4.5. Future Enhancements

483 We develop and demonstrate this methodology as a step towards closing the drought
484 feedback loop, but note that there are caveats and limitations that warrant discussion. A
485 conceptual overview of the contribution of our study to the drought feedback loop is
486 shown in Figure 9, and we use this figure to identify five places where there is scope for
487 future enhancements; each number below corresponds to a place in the drought feedback
488 loop in Figure 9:

- 489 1) For the natural influence on drought, we examine the probability of groundwater
490 drought. In our case, the groundwater levels are closely related to rainfall
491 recharge, which is a function of natural climate variability. We recognize that this
492 is not the case for many groundwater aquifers, where human activities, such as
493 groundwater extraction, may trump the natural climate signal (e.g., Tarhule and
494 Bergey 2006), often leading to water scarcity, rather than a natural phenomenon
495 of temporary water deficiency. In many systems a full water balance would need
496 to be examined to understand the relative contributions of extraction versus
497 moisture deficit to the likelihood of going below a relevant hydrologic threshold.
498 Further, other aquifers may have different properties; for example, some aquifer’s
499 natural response may be different and the levels may not closely resemble rainfall.
- 500 2) Our interviews were conducted following a drought event, and we recognize that
501 the timing of the interviews will likely affect the responses, possibly introducing a
502 bias. For instance, interviews conducted during wet or average conditions might

503 elicit less polarized responses, since drought impacts haven't been recently
504 experienced. We note that our approach of applying the interview responses
505 across different climate conditions (i.e., wettest to driest) makes the assumption
506 that the importance of water uses and management preferences are stationary. We
507 acknowledge that different climate conditions, as well as cultural change,
508 technological innovation, climate adaptation, and other processes are likely to
509 influence the cultural factors we investigated here and may mediate how people
510 interact with their environments. Future work could investigate how responses
511 change with different climate conditions over time, and the subsequent
512 implications for drought action. However, hazards and disasters research is almost
513 always conducted immediately after an event, so this is a wise-spread
514 epistemological issue with both pros and cons in terms of what we learn from
515 post-disaster research.

516 3) We use stakeholders' importance ratings as a proxy for their willingness to take
517 action in relation to particular water uses, where by "action", we generally mean
518 some effort towards drought mitigation. The interviews included questions about
519 the importance of different water uses to test the application of the Cultural
520 Theory of Risk (usually applied in a more global sense) to a specific water
521 management issue, which had not been done before (Lazrus 2016). For the
522 purposes in this paper, multiplying the importance ratings by the probability
523 served as a way to make an objective characterization of drought subjective, that
524 is, we wanted to modulate the groundwater drought probability by each individual
525 stakeholder's lens.

526 4) The formulation of the PDAI strongly affects the conclusions drawn. Our
527 formulation of the PDAI follows from other precedents in risk management that
528 take the product of the likelihood of an event and its importance (Jones and
529 Preston 2011; Oppenheimer et al. 2014). However, the functional form of the
530 PDAI is flexible, allowing it to be tailored to other locations. As such, we note
531 that the PDAI, as well as the best data to use to calculate it, will depend on the
532 needs of the community, as well as the water system context.

533 5) We use social science theory to interpret our results, and to better understand the
534 theoretical underpinnings of how and why people take action in response to
535 drought. However, we note that empirical validation is important for indicator
536 development and refinement. We recommend that future project designs include
537 a validation component in the methodology. This could take the form of follow-
538 up interviews, such as direct feedback from stakeholders on if the indicator
539 reflects their willingness to take action for certain water uses at certain drought
540 levels. Methods, including stakeholder processes, for developing and evaluating
541 drought indicator effectiveness have been put forth in the drought community
542 (Steinemann and Hayes 2005 Steinemann and Cavalcanti 2006, Steinemann et al.
543 2015). Other options for validation can be indirect, such as looking at historical
544 data, like government and local reports, media, and/or other collected response
545 information, e.g., in the United States the US Drought Impact Reporter (DIR;
546 Wilhite et al. 2007). Tjrdemen et al (2018) examined the relationship between
547 drought indicators and impact data from the DIR; however, it has been noted that
548 the DIR would benefit from a more systematic and coordinated collection effort

549 (Lackstrom et al. 2013), which presents challenges for its interpretation.
550 Promising methods for mining social media, such as Twitter, have also been
551 developed (Demuth et al. 2018) and could be adapted for evaluative purposes.
552

553 Related to the points above is the question about how the PDAI could connect with
554 existing operational products and its transferability to other locations. In our case,
555 groundwater threshold exceedance was linked with water use impacts. Ideally, the PDAI
556 could be modified to incorporate an operational drought indicator that is associated with
557 impacts; however, evaluations of the connection between monitored indicators and
558 impacts has been limited (Bachmair et al 2016). In terms of the transferability of the
559 social perspective, the idea behind the cultural theory of risk worldview measures is that
560 they are loosely universal, that is, they should apply fairly generally to any context within
561 the broad culture for which they were initially put together – in this case, the United
562 States (Smith and Leiserowitz 2014). However, worldview measures can also be tailored
563 to a particular context (Lazrus 2016), which might need to be revised for other
564 applications.
565

566 5. Conclusions

567 Our study implements a conceptual methodology combining hydrological and
568 societal perspectives to understand drought action potential. Results from stakeholder
569 interviews in the study site reveal that people perceive the relative importance of water
570 for various uses differently, as shown by the notable variability that existed across certain
571 water uses. A retrospective analysis of groundwater threshold exceedance shows that in
572 recent decades, stakeholders experienced a wide range of likelihoods of exceeding
573 relevant thresholds, and these corresponded to drought indices. These pieces of
574 information are brought together through the PDAI. We find that for a given water use,
575 drier conditions increase the frequency of exceeding the groundwater threshold, and
576 hence increase the PDAI. The PDAI is tied to the threshold selected for each water use:
577 we find that the PDAI is higher for more moderate thresholds, i.e., thresholds that are
578 exceeded more often. And conversely, as thresholds become more extreme, which can
579 illustrate human adaptation to new drought normal, the PDAI decreases. Finally, we find
580 that for water uses where stakeholder values are diverse, the PDAI will be diverse as
581 well, and this is exacerbated under drier conditions.

582 We can also ask why values might be diverse, and what that might mean about
583 how people are affected by water scarcity and how they will respond. To this end, the
584 study also examined worldview, as measured by the CTR, which can help to diagnose
585 *why* disagreement may arise over water management and point to some implications for
586 water management policy. In the stakeholder sample, we found a diverse range of
587 worldviews on the individualist/egalitarian spectrum. Further, for some water uses, the
588 importance people attribute to water can be partially explained by worldview. This
589 implies that there are two potential sources for disagreement over water management:
590 first, where there is variability in people's perception of importance, there may be
591 disagreement over whether or not a water resource needs to be managed (e.g., with water
592 for recreation). Second, even where there is consensus on people's perceived
593 importance, there is still potential for disagreement over how these water resources
594 should be managed according to different preferences of worldviews (e.g., with drinking

595 water). We are careful to say *potential* disagreement because (i) our analysis only
596 investigates CTR as one of the many factors explaining importance and (ii) by
597 understanding stakeholder worldviews, potential disagreement across sectors can be
598 predicted and ideally avoided. The latter finding suggests that water management policies
599 will be more successful if they follow a strategy whereby elements of each worldview are
600 represented in the solution (Verweij et al. 2006).

601 Although reducing disagreement is always important for promoting sustainability, it
602 is particularly important for management planning under potentially increasing drought
603 due to climate change, as has been predicted for this area (Towler et al. 2016; Liu et al.
604 2012). We examined this by examining possible adaptation to a new normal, where we
605 illustrate how a more extreme threshold lowers the PDAI.

606 Although the methodology to develop the PDAI is experimental, we posit that
607 explicit efforts to combine natural and human perspectives is critical to gaining a deeper
608 and more nuanced understanding of drought feedbacks, and this paper provides a novel
609 contribution to this end.

610

611 Data Availability

612 Groundwater data from the USGS Fittstown well (USGS 343457096404501) is available
613 from the USGS National Water Information System Web Interface
614 <https://nwis.waterdata.usgs.gov/nwis/gw>. Drought index data for Oklahoma Climate
615 Division 8 is available from: <http://www7.ncdc.noaa.gov/CDO/CDODivisionalSelect.jsp>
616 Inquiries on the stakeholder data from the interviews can be sent to hlazrus@ucar.edu

617

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Tables

Table 1. Correlation Between Select Drought Indices* and the Likelihood (P) of Groundwater (GW) Level Going Below Moderate (Mod) and Severe (Sev) Thresholds

Drought Index		Correlation	
Type	Name	P(GW<Mod)	P(GW<Sev)
Agricultural	Palmer Drought Severity Index (PDSI)	-0.92**	-0.83**
Hydrological	Palmer Hydrological Drought Index (PDHI)	-0.95**	-0.84**
Meteorological	Standardized Precipitation Index - 6 month (SP06)	-0.94**	-0.82**

* Drought indices for Oklahoma Climate Division 8 downloaded from:

<http://www7.ncdc.noaa.gov/CDO/CDODivisionalSelect.jsp>

** = Significant at the 99th percentile

820 Table 2. Decadal Likelihood (P) of Groundwater (Z) Level Going Below Moderate (Mod) and Severe (Sev) Thresholds for Recent
 821 Decades, as Well as Respective Driest, Median, and Wettest Decades.
 822

P(Z<Mod) (%)	P(Z <Sev) (%)	Comment	Decade
61	38	Dry/recent decade; most recent decade to interviews	2003-2012
35	14	Average/recent decade; third most recent decade	1983-1992
31	8	Wet/recent decade; second most recent decade	1993-2002
75	40	Driest decade; highest exceedance likelihood	1959-1968(Mod); 1964-1973(Sev); 1972-1981(Sev)
50	25	Median decade; median exceedance likelihood	1999-2008(Mod); 1979-1988(Mod); 1969-1978(Sev); 1980-1989(Sev)
23	8	Wettest decade; lowest exceedance likelihood	1985-1994(Mod); 1993-2002(Sev)

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825 Table 3. Correlation and Statistical Significance of Worldviews, as Quantified by the
826 Egalitarian and Individualist Measures, with Importance Ratings for Each Water Use.

<u>Water Use</u>	<u>Egalitarian</u>	<u>Individualist</u>
Drinking Water	-0.20	0.27*
Habitat	0.24*	-0.25*
Livelihood	0.13	-0.09
Recreation	0.45**	-0.40**
Cultural Practices	0.42**	-0.29*
<u>Spiritual Fulfillment</u>	<u>0.18</u>	<u>-0.23*</u>

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* = Significant at the 90th percentile

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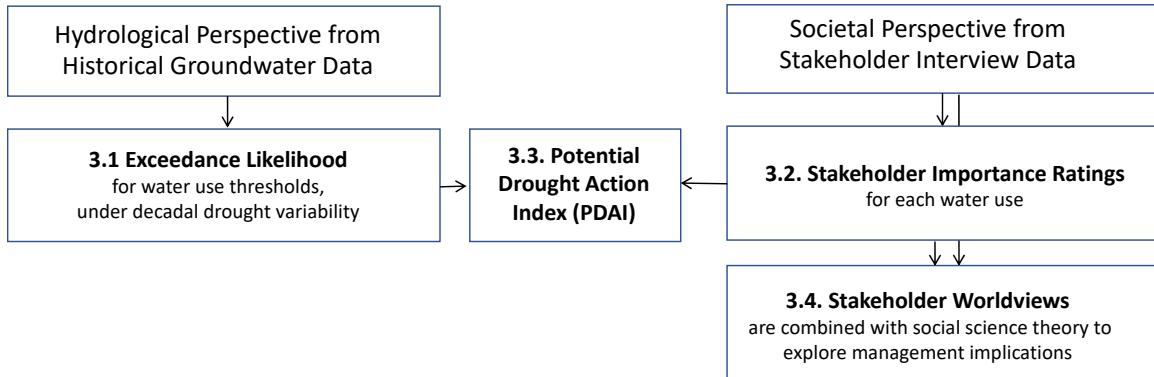
**= Significant at the 99th percentile

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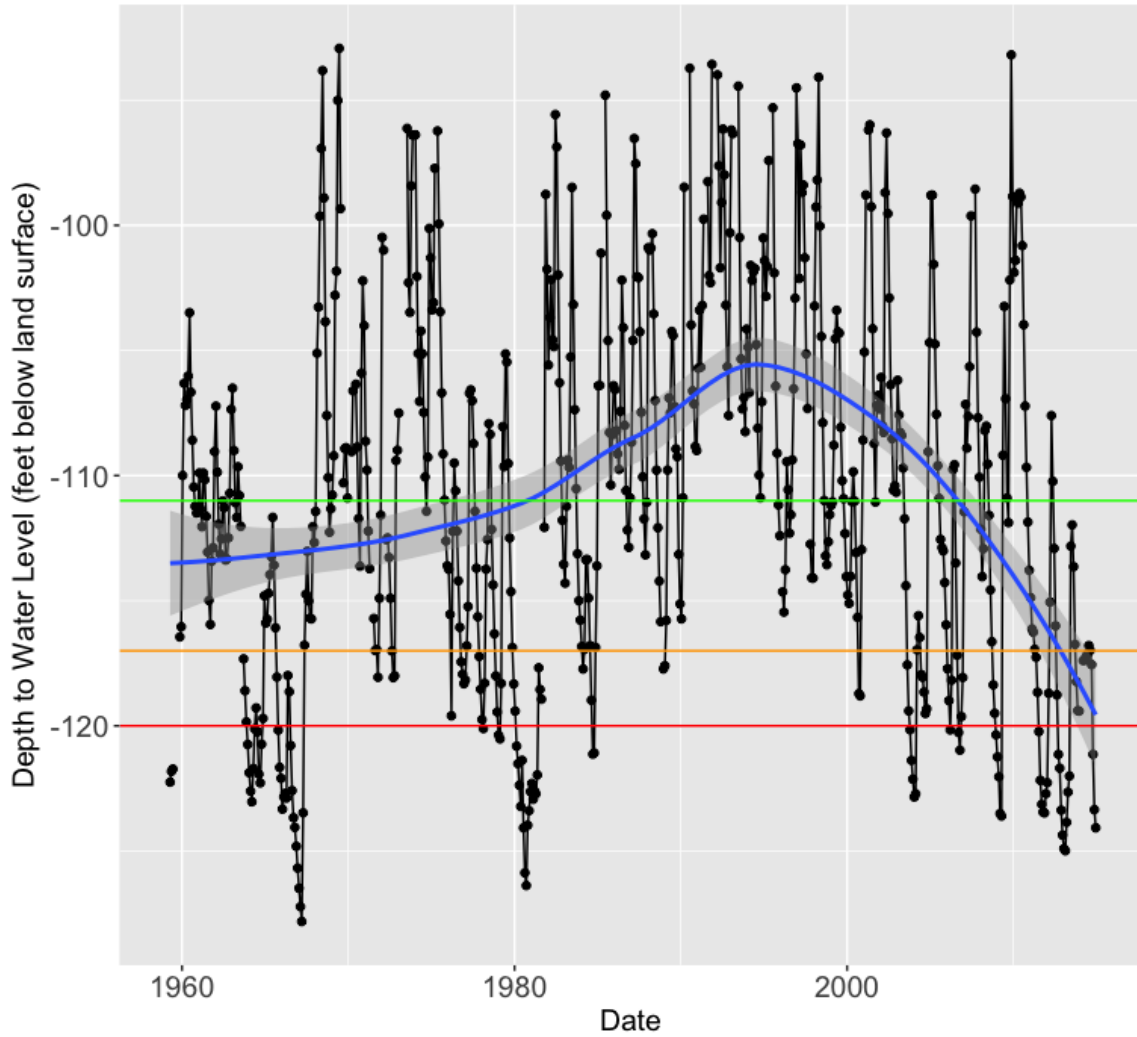
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Figures



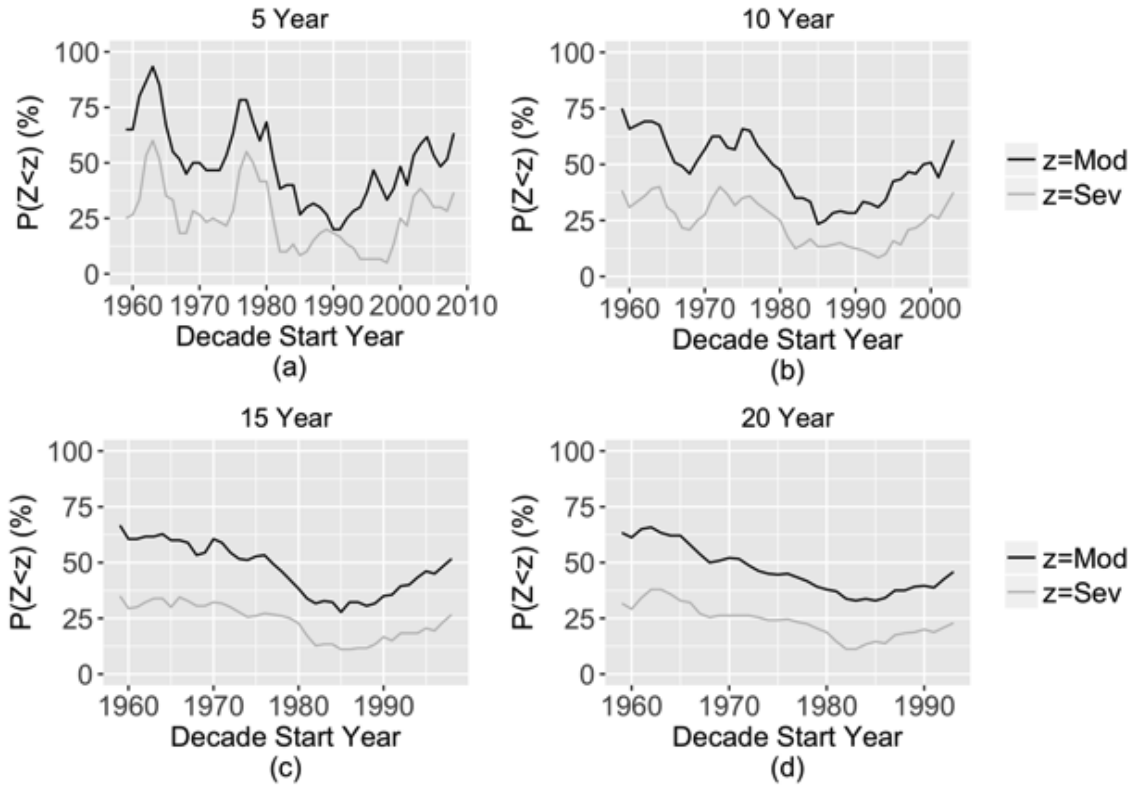
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Figure 1. Conceptual overview of the methodology that combines a hydrological perspective from historical groundwater data with a societal perspective from stakeholder interview data to quantify the Potential Drought Action Indicator (PDAI); stakeholder worldviews from the interviews and social science theory are used to explore management implications.



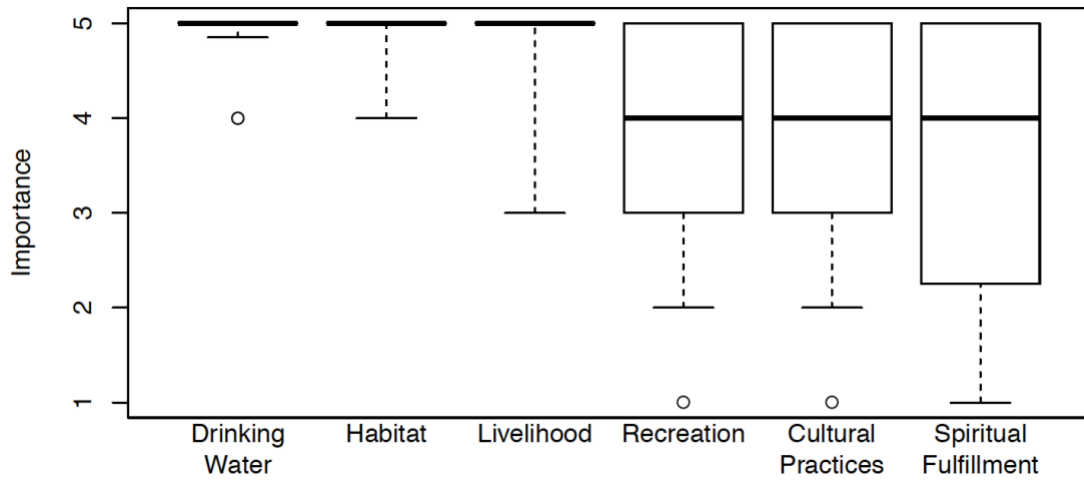
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Figure 2. Monthly groundwater time series; blue line is a local smoother average, green line is the moderate threshold (= -111 feet) and the orange line is severe threshold (= -117 feet); the red line is an extreme threshold (= -120 feet) that is used to illustrate a possible new normal drought threshold.



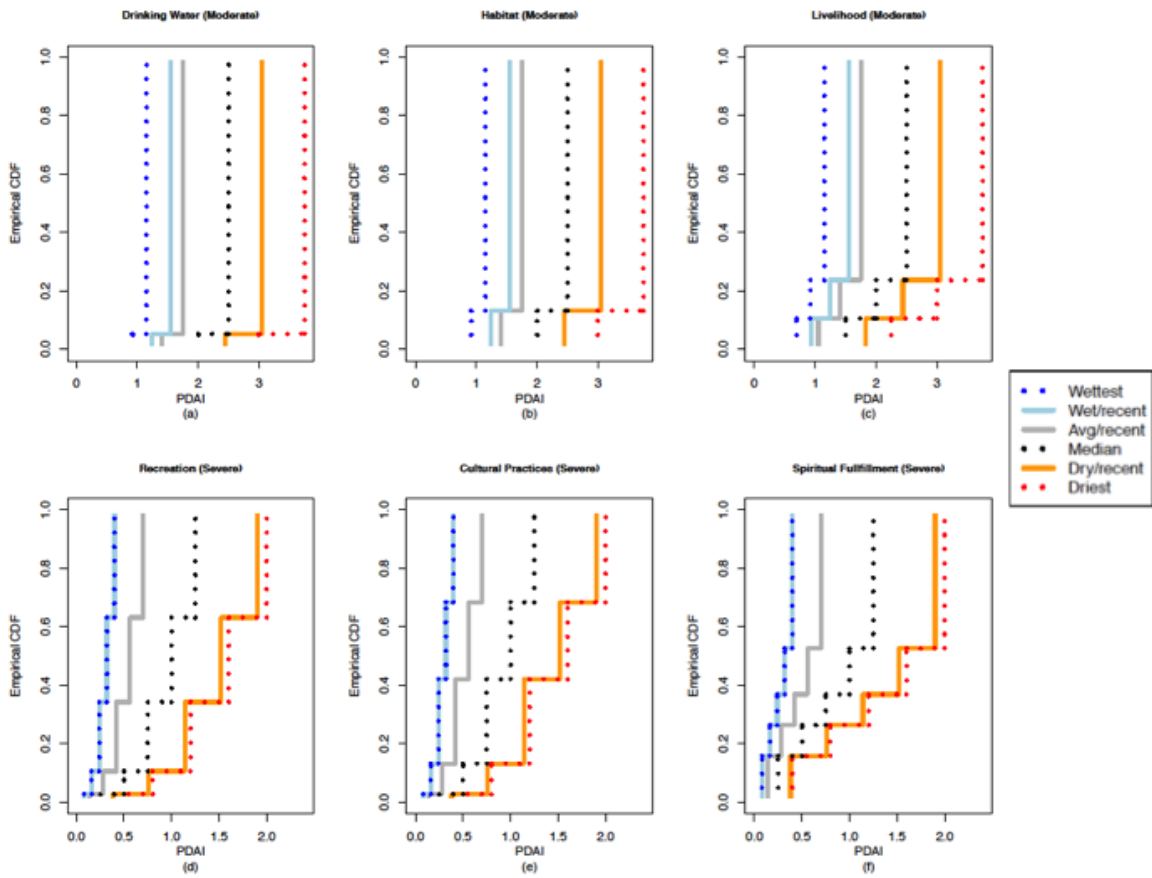
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 851 Figure 3. Groundwater drought likelihood (P) of the depth to groundwater level (Z) going
 852 below the moderate (Mod, $z = -111$ ft) and severe thresholds (Sev, $z = -117$ ft) for time
 853 windows from 5 to 20 years.
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859 Figure 4. Rated importance of water for each water use from stakeholder surveys (N=38)
860 on a Likert scale of 1-5, 5 being very important. Responses are shown as box plots, where
861 the box represents the 25th and 75th percentile, the line is the median, and the whiskers
862 are the 5th and 95th percentile. Outliers are shown as points outside the box and
863 whiskers.

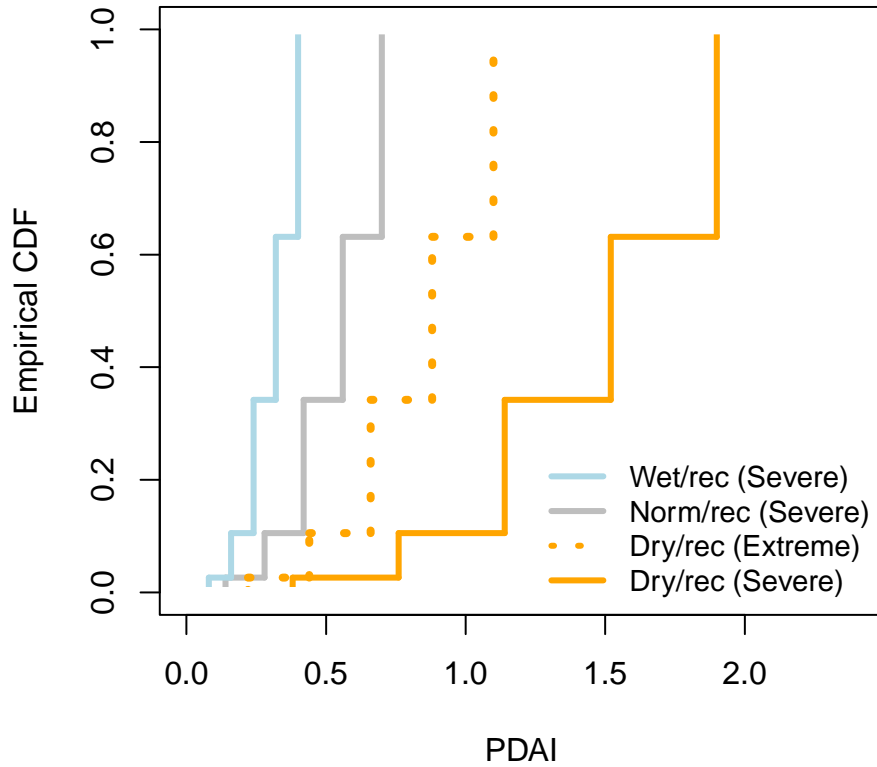
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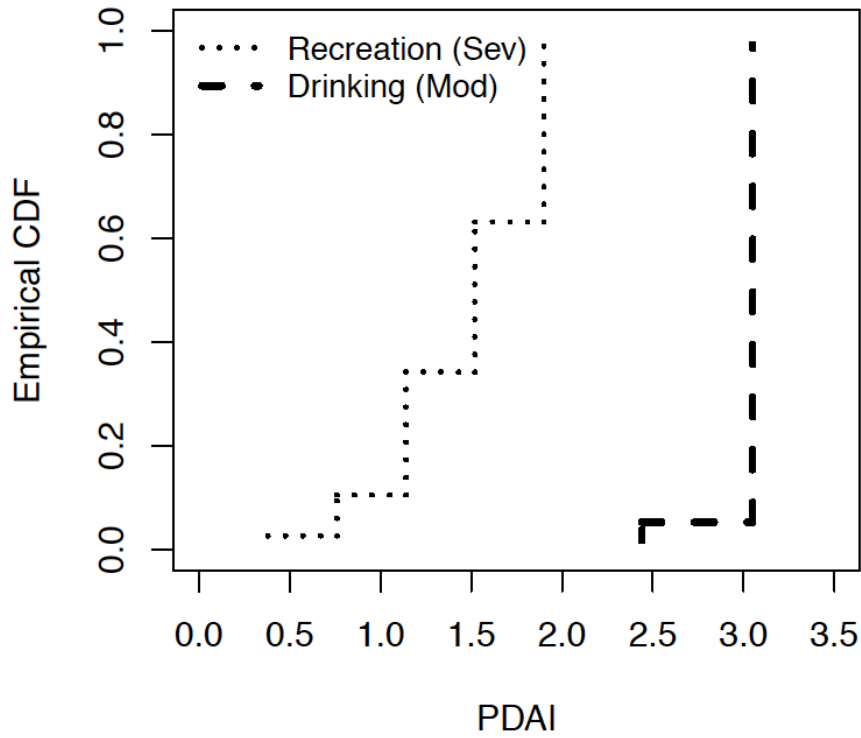
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Figure 5. Empirical cumulative density functions (eCDFs) for the PDAI (Potential Drought Action Indicator) for water uses using the moderate threshold (a-c) and for water uses using the severe threshold (d-f), under the wettest, wet/recent, average/recent, median, dry/recent, and driest historical decades.

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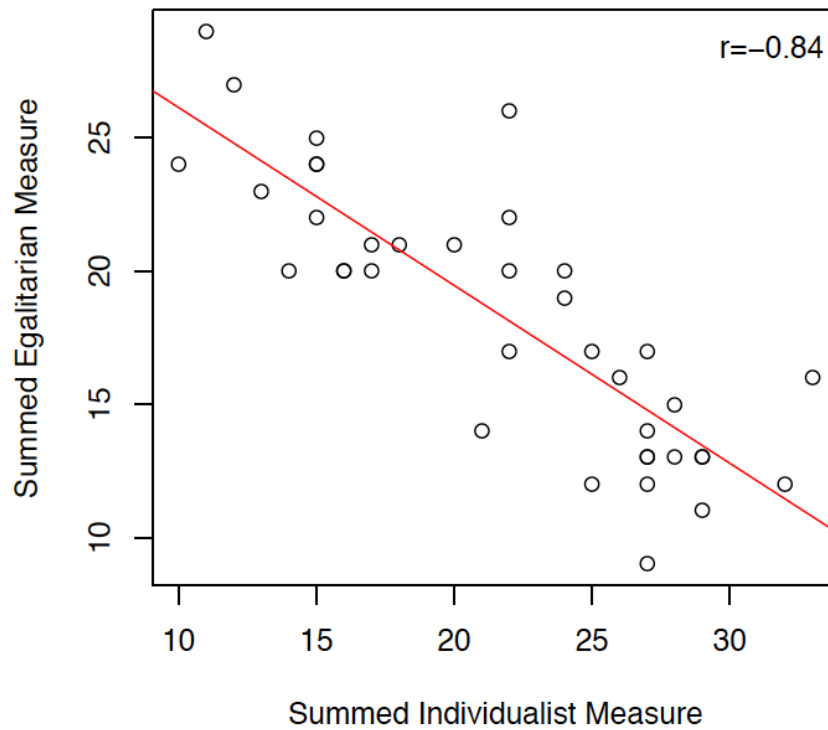


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894 Figure 6. Empirical cumulative density functions (eCDFs) for the PDAI (Potential
895 Drought Action Indicator) for recreation under the wet/recent (1993-2002), normal/recent
896 (1983-1992), and dry/recent (2003-2012) for the severe threshold, as well as the
897 dry/recent for the extreme threshold.
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 902 Figure 7. Empirical cumulative density functions (eCDFs) for the PDAI (Potential
 903 Drought Action Index) for recreation using the severe (Sev) threshold and for drinking
 904 water using the moderate (Mod) threshold for the dry/recent decade (2003-2012).
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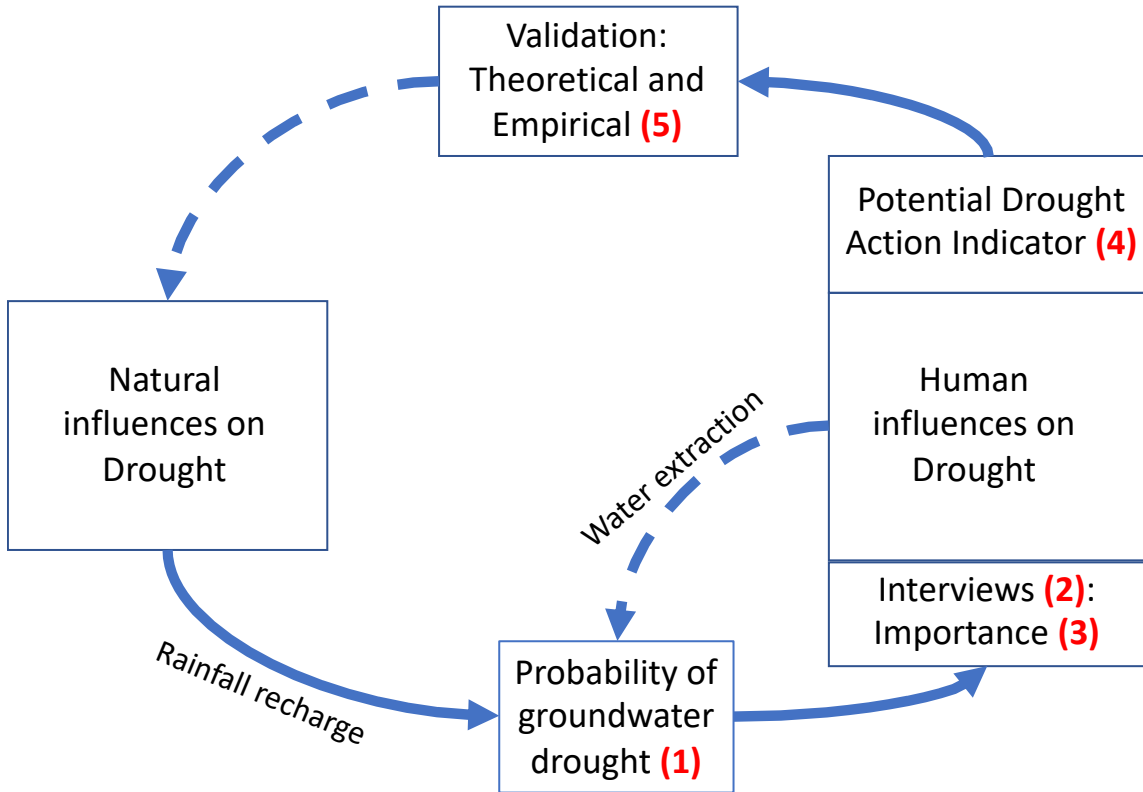
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908 Figure 8. Summed responses for individualism versus egalitarianism for each interviewee
909 (n=38) show that both individualist and egalitarian worldviews were represented by the
910 interviewees. The egalitarianism and individualism measures were strongly inversely
911 correlated ($r=-0.84$).

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 914 Figure 9. Conceptual map of drought feedback loop components addressed in this study
 915 (blue solid lines) and remaining gaps (blue dashed lines). Numbers correspond to
 916 discussion points in Conclusions.