

19 **Abstract**

20 Fresh water is consumed during agricultural production. With the shortage of water resources,
21 assessing the water use efficiency is crucial to effectively manage agricultural water resources. The water
22 footprint is an improved index for water use evaluation, and it can reflect the quantity and types of water
23 usage during crop growth. This study aims to establish a method for calculating the regional scale water
24 footprint of crop production based on hydrological processes, and the water footprint is quantified in
25 terms of blue and green water. This method analyzes the water-use process during the growth of crops,
26 which includes irrigation, precipitation, groundwater, evapotranspiration, and drainage, and it ensures a
27 more credible evaluation of water use. As illustrated by the case of the Hetao Irrigation District (HID),
28 China, the water footprint of wheat, corn and sunflower were calculated using this method. The results
29 show that canal water loss and evapotranspiration were responsible for most of the water consumption
30 and accounted for 47.9% and 41.8% of the total consumption, respectively. The total water footprint of
31 wheat, sunflower and corn were $1380\text{-}2888\text{ m}^3\text{ t}^{-1}$, $942\text{-}1774\text{ m}^3\text{ t}^{-1}$, and $2095\text{-}4855\text{ m}^3\text{ t}^{-1}$, respectively,
32 and the blue footprint accounts for more than 86%. The spatial distribution pattern of the green, blue and
33 total water footprint for the three crops demonstrated that higher values occurred in the eastern part of
34 the HID, which had more precipitation and was further away from the irrigating gate. This study offers
35 a vital reference for improving the method used to calculate the crop water footprint.

36 **Key words**

37 SWAT model; Regional Scale; Water use process; Hetao Irrigation District

38 **1 Introduction**

39 Human activities and climate change have serious effects on the availability of water resources
40 (Nijssen et al., 2001; Haddeland et al., 2014). Agricultural production is major consumer of global water
41 resources and accounts for 85% of the global blue water (surface or groundwater) consumption
42 (Shiklomanov, 2000; Vörösmarty et al., 2010). In China, 63% of all water is used for agricultural
43 production each year, and the area of irrigated farmland is 39.6% of the total arable land. Irrigation is the
44 key to ensure agricultural production (NBSC, 2016). With the rapid development of China's economy,
45 the demand for water has increased in industrial production and in the lives of residents (Duh et al., 2008;
46 Liu et al., 2008; Bao and Fang, 2012). Environmental pollution reduces water availability (Jiang, 2009;
47 Schwarzenbach et al., 2010) and these changes place great pressure on regional water resources (Piao et
48 al., 2010; Wang et al., 2014); meanwhile, climate change aggravates the situation (Elliott et al., 2014).
49 With limited water resources, economic demand for water will inevitably and gradually take up the
50 agricultural water use, which is a challenge for maintaining steady agricultural production (Chen, 2007;
51 Khan et al., 2009), especially in the dry areas of northern China (Deng et al., 2006; Du et al., 2014).
52 Strengthening agricultural water management and improving water use efficiency are significant aspects
53 of handling water scarcity, and a reasonable evaluation of the water resource for crop production is the
54 premise for developing an agricultural water management plan and implementing water saving measures.
55 Therefore, how to precisely evaluate the effective utilization ratio of current agricultural water use,
56 improve the utilization efficiency, and reduce the negative impact of the reduction of available
57 agricultural water on agriculture production, are important issues that all countries need to address
58 globally, which are also of vital importance for ensuring food production and reducing the pressure on
59 water resources. The water footprint theory provides new insights and ideas to solve these problems

60 (Hoekstra, 2003). The water footprint is an indicator of freshwater use and can be used to quantify water
61 consumption throughout the production supply chain. It reflects the amount of water, the green, blue and
62 grey water that are consumed (Hoekstra et al., 2011). In the agricultural sector, it can also be used to
63 evaluate whether a crop's water footprint is reasonable and whether it varies regionally. Since green water
64 can be used in agricultural production, some measures can be taken to reduce the water footprint of crop
65 production, especially to reduce the consumption of blue water, thereby easing the demand for blue water
66 in agriculture. The accurate and precise quantification of crop production water footprint is the premise
67 to achieving the above goals.

68 Currently, based on two mainly methods proposed by Hoekstra et al. (2011), many scholars have
69 quantified various levels of crop production water footprint, such as a global level (Mekonnen and
70 Hoekstra, 2011), a national level, such as Europe (Vanham and Bidoglio, 2013) and China (Zhao et al.,
71 2009), and a regional level, such as Beijing (Sun et al., 2013a), Cremona province (Bocchiola, 2015) and
72 Hetao (Luan et al., 2018). The first is the crop water requirement method (Cao et al., 2014; Sun et al.,
73 2013c). This method simulates the actual evapotranspiration (ET) of crops under optimal conditions with
74 the potential ET calculated by the Penman-Monteith Equation (Allen et al., 1998) and the effective
75 precipitation calculation provided by the United States Department of Agriculture Soil Conservation
76 Service (USDA SCS) (Doll and Siebert, 2002). The green water consumption is the smaller value of total
77 crop actual ET and effective precipitation. The blue water consumption is obtained through the difference
78 between the total crops actual ET and effective precipitation. Finally, when combined with crop yields,
79 the crop blue and green water footprint ($\text{m}^3 \text{t}^{-1}$) can be calculated. The second is the irrigation schedule
80 method. This method is based on an empirical formula model such as the CROPWAT model (FAO, 2010;
81 Mekonnen and Hoekstra, 2011) CropSyst (Bocchiola et al., 2013), the EPIC model (Williams et al., 1989;

82 Shi et al., 2017), the GEPIC model (Liu et al., 2007), and the AQUACROP model (Pasquale et al., 2009;
83 Chukalla et al., 2015; Zhuo et al., 2016). The detail calculation process of these two methods are listed
84 in Supplementary material.

85 These methods can simulate actual ET throughout the crop growing period according to the soil
86 water balance under optimal or suboptimal conditions. The blue water consumption is the smaller value
87 of net irrigation water and the net irrigation water requirement. The green water consumption is equal to
88 the total actual ET minus blue water. Both of the above methods are based on empirical formulas. A few
89 scholars have attempted to calculate the regional scale water footprint, for example, Sun et al. (2013b)
90 used the difference between diversion and drainage to calculate the water footprint of crop production in
91 irrigated areas. However, these methods have certain shortages, which are as follows:

92 First, the applicability of the empirical method has not been determined, that is, whether the method
93 is applicable to the field scale or regional scale of water footprint calculation needs further study. These
94 methods calculated the field scale water footprint with net irrigation water considered as irrigation water,
95 and without considering water loss during transport, which definitely serves for crop growth. Therefore,
96 these methods are field scale methods, whereas a regional scale method should include the above two
97 losses. At present, irrigation water mainly refers to the net irrigation water used by crops in the field.
98 Current irrigation water analysis methods have not considered water loss during water delivery and
99 drainage. Therefore, the calculation of the water footprint at the field scale cannot be accurately applied
100 to irrigated agriculture. However, there are still few methods to calculate the water footprint on the
101 regional scale.

102 Second, the irrigation data in these methods are simulation values and not based on the actual
103 irrigation time and irrigation quota (the amount of water demanded for crop irrigation); therefore, these

104 data cannot reflect the real situation of the local water usage due to the incomplete simulation data. At
105 the same time, the traditional method does not completely analyze the water footprint components of
106 water resources in the process of water diversion, water transfer, irrigation and drainage.

107 Third, the current regional scale method has not been appropriately established. The method that
108 Sun et al. (2013b) used had these limitations which mentioned above. It included all of the water
109 consumption, but it could not distinguish the specific source of blue water from canal loss, field actual
110 ET or groundwater. Due to its low spatial resolution, only the water footprint of the entire irrigated area
111 could be calculated instead of the difference inside this area.

112 Currently, most studies focus on the field scale and lack systematic evaluation on the whole process
113 of water consumption during crop growth. To overcome this problem, this study put forward an improved
114 regional scale calculation method of the crop water footprint based on hydrological process analysis and
115 used it to quantify the crop water footprint in HID. This method simulated the hydrological cycle of the
116 region based on a physical hydrological model (SWAT). Based on the method, this study analyzed the
117 water input and output during crop production, and calculated the water consumption in crop growth,
118 field drainage and water loss during canal water transport. Combined with crop yields, the water footprint
119 of crop production at the regional scale was quantified. This method will provide comprehensive
120 information for the analysis of water consumption during crop production process and improve the spatial
121 resolution of the regional distribution of water footprint of crop production.

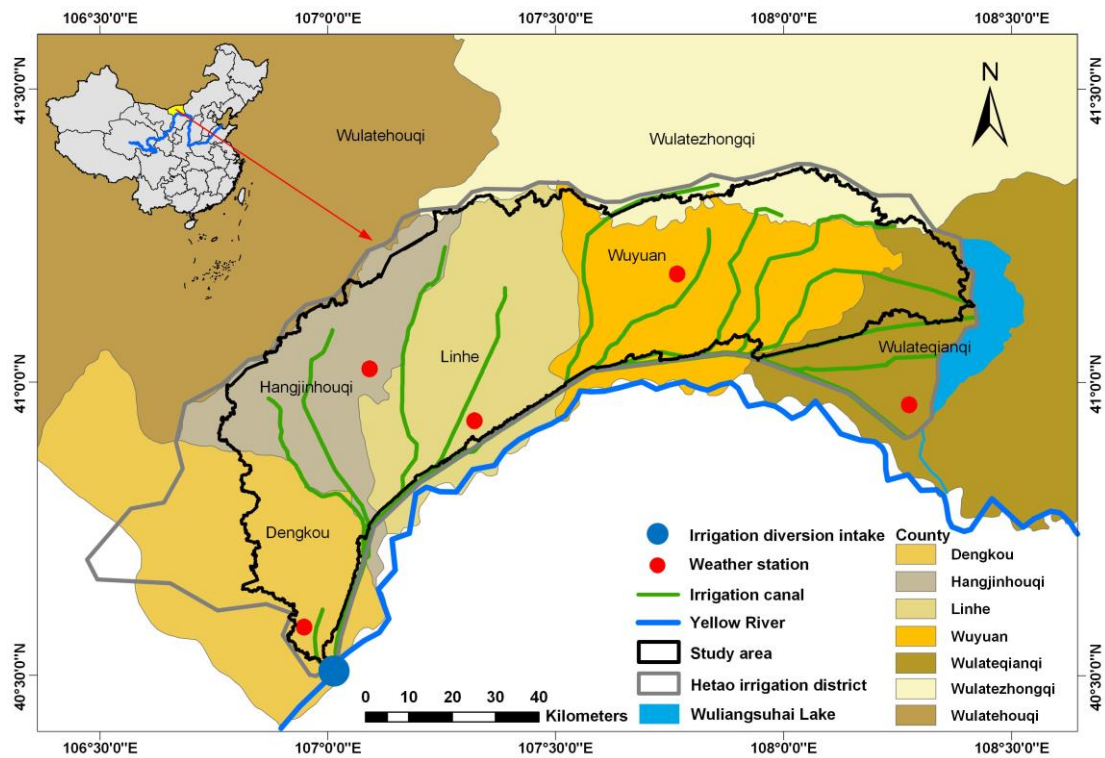
122 **2 Materials and methods**

123 **2.1 Study site**

124 The Hetao Irrigation District (HID) is located in the middle of the Yellow River basin in western
125 Inner Mongolia (Fig. 1) and is one of the three largest irrigation districts in China. The HID has a

126 continental monsoon climate with the lowest temperature in January (average -10°C) and highest
 127 temperature in July (average 23°C). The average monthly precipitation is 37.5 mm (May to September),
 128 3.4 mm (October to next year April), and the average monthly potential evaporation is 290.6 mm (April
 129 to September), 77.2 mm (October to next year March). The area of the HID is $1.12 \times 10^4 \text{ km}^2$.

130 Irrigation water is diverted from the Yellow River. The irrigation and drainage systems in the HID
 131 are composed of irrigation canals and drainage ditches; the irrigation system has a general main canal
 132 (228.9 km) and 12 main canals (total 755 km), and the drainage system has a general main ditch (227
 133 km) and 12 main ditches (total 523 km). The main crops include wheat, corn and sunflower (Fig. 1).



134
 135 **Fig. 1.** Location of the Hetao Irrigation District (HID) in China

136 **2.2 Model description**

137 The SWAT (soil and water assessment tool) model is a semi-distributed physical hydrological
 138 model. The model was developed by USDA Agricultural Research Center and it used climate, soil,
 139 topography, plants and land management practices to simulate hydrologic, sediment, crop growth and

140 nutrient cycle. The model partitions a watershed into sub-basins by topography and then partitions the
141 sub-basins into hydrologic response units (HRU) based on soil type and land use to assess soil erosion,
142 non-point pollution, and hydrologic processes (Haverkamp et al., 2002). The water balance equation
143 governed by the hydrologic component of the SWAT model (Neitsch et al., 2011) is as follows:

$$144 \quad SW_t = SW_0 + \sum_{i=1}^t (R_{day} - Q_{surf} - E_a - W_{seep} - Q_{gw}) \quad (1)$$

145 where SW_t is the final soil water content (mm H₂O), SW_0 is the initial soil water content (mm H₂O), t is
146 the time (days), R_{day} is the amount of precipitation on day i (mm H₂O), Q_{surf} is the amount of surface
147 runoff on day i (mm H₂O), E_a is the amount of actual ET on day i (mm H₂O), W_{seep} is the amount of
148 percolation and bypass flow exiting the bottom of the soil profile on day i (mm H₂O), and Q_{gw} is the
149 amount of return flow on day i (mm H₂O).

150 **2.3 Data collection**

151 The data required by the SWAT model includes a digital elevation model (DEM), soil data, land
152 use, and hydrological and climate data (Table 1). The climate data were obtained from five weather
153 stations in the HID.

154 The water efficiency of the canal system in this model was obtained from local agricultural
155 administrations (AHID, 2015). To divide the sub-basins, we defined the drainage ditch as the stream
156 (AHID, 2015) and burn-in into the DEM, and the simulation results were verified by the discharge of the
157 drainage ditch.

158 The model generated 5 outlets and 73 subbasins, and the measured data of the first outlet in the
159 study area was obtained. Therefore, this study chose the area controlled by this outlet as the study area.
160 The crops' yields (wheat, corn and sunflower) required for the calculation of the water footprint was
161 obtained from the Statistical Yearbook of the local agricultural administrations (AHID, 2015).

Table 1 Data used in the study and the resources

Dataset	Data description	Resolution	Data sources
DEM	—	30×30 m	Geospatial Data Cloud (CAS, 2009a)
Soil	Soil type map, Soil physical and chemical properties	1:1000000	China Soil Scientific Database (CAS, 2009b)
Land use	—	1:100000 (2010)	Data Center for Resources and Environmental Sciences (CAS, 2010)
Weather	Precipitation, Wind speed, Solar radiation, Maximum temperature, Minimum temperature, Relative humidity	Daily (1980-2012)	China Meteorological Data Network (NMIC, 2015) The Administration of Hetao Irrigation District (AHID, 2015)
Hydrologic	Stream map, Discharge	Monthly (2003-2012)	The Administration of Hetao Irrigation District (AHID, 2015)
Crop parameter data	Dates of plant and harvest, Dates of irrigation, Irrigation quota	—	The Administration of Hetao Irrigation District (AHID, 2015)

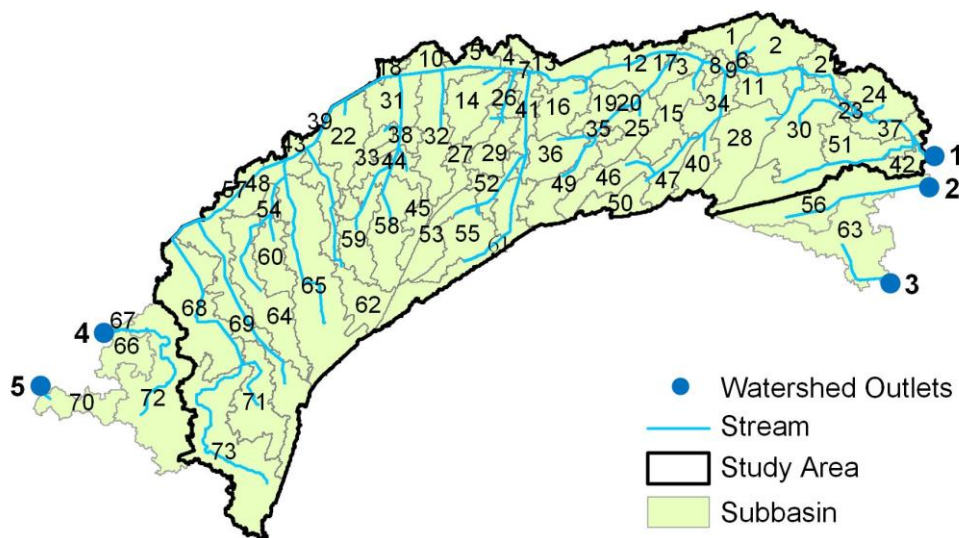


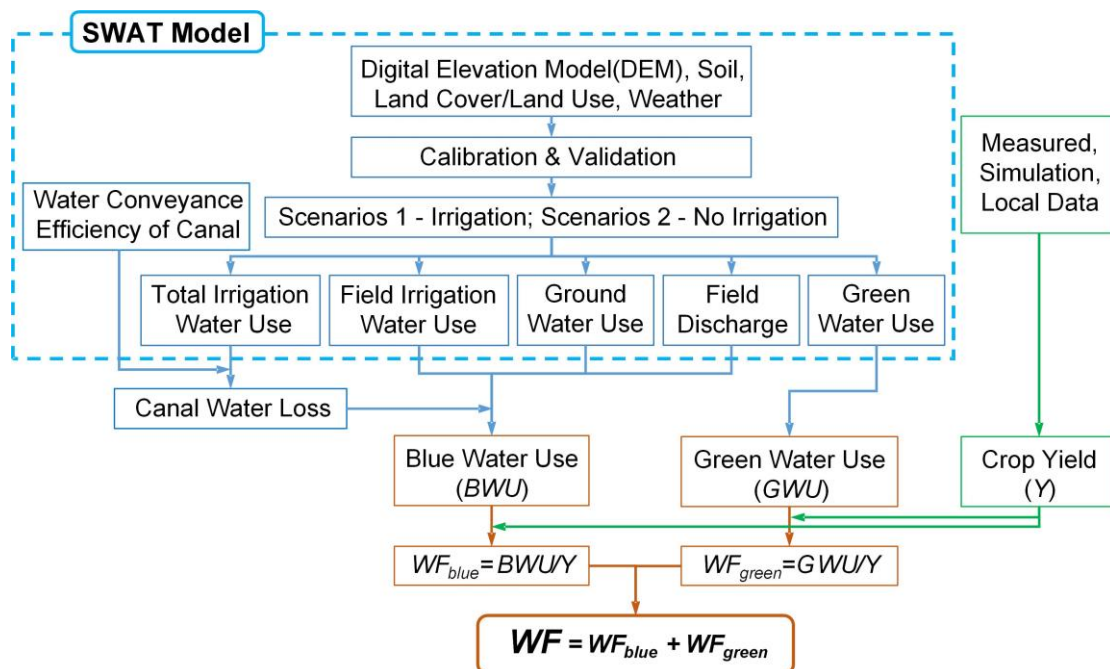
Fig. 2. Subbasins and study areas

166 **2.4 Calibration and validation**

167 The Sequential Uncertainty Fitting (SUFI-2) algorithm in SWAT-CUP was applied for calibration
 168 and validation (Abbaspour et al., 2007; Abbaspour, 2012) by comparing the simulated stream discharge
 169 from the model with the measured discharge data. The global sensitivity analysis integrated within SUFI-
 170 2 was used to evaluate the hydrologic parameters for the discharge simulation and then the optimal
 171 simulation was established by adjusting the sensitivity parameters and through multiple iterations. The
 172 calibration period was from 2006-2009, and the validation period was from 2010-2012. The result of the
 173 SWAT calibration and validation process is satisfactory, the detailed process is available in
 174 Supplementary material.

175 **2.5 The regional scale water footprint calculation method**

176 Based on the water footprint theory framework provide by Hoekstra et al. (2011), this study suggests
 177 a new way of quantifying the regional scale water footprint of crop production (Fig. 3).



178

179 **Fig. 3.** The flowchart for calculating the regional scale water footprint

180 In this study, green water consumption is the effective precipitation during crop growth process.

181 Blue water consumption includes canal water loss during delivery, the ET produced by consumption of
 182 irrigation water and groundwater for crops growth, and the drainage in the fields. To calculate the canal
 183 water loss, an extra model needs to be established according to the HID situation, and the other can be
 184 simulated and obtained by the SWAT model.

185 **2.5.1 Calculation of water consumption factors in the fields**

186 Water consumption in the fields consists of 4 parts including the actual ET of precipitation,
 187 irrigation water, groundwater utilized by crops, and field drainage. This study set up two scenarios and
 188 calculated the above water consumption by changing the sources of water in the SWAT model. In
 189 scenario 1 (S1), crop water consumption was derived from precipitation and irrigation water (irrigation
 190 systems and irrigation quotas are based on local irrigation methods), i.e., the actual situation of crop
 191 water use. In scenario 2 (S2), crop water consumption was only derived from precipitation without
 192 irrigation. The S2 was used to calculate the consumption of green water. In this study area (HID), because
 193 of less rainfall, the effective precipitation formed by precipitation is all used for crop growth. Therefore,
 194 the consumption of green water for crops is equal to the effective precipitation, which means that green
 195 water is reflected by calculating the effective precipitation stored in soil by SWAT model. The
 196 calculation formula is as follows.

$$197 \quad WF = WF_g + WF_b = \frac{W_g}{Y} + \frac{W_b}{Y} \quad (2)$$

$$198 \quad W_g = PRECIP_{s2} - SUPQ_{s2} - LATQ_{s2} \quad (3)$$

$$199 \quad W_b = Q_c + Q_f + Q_d \quad (4)$$

$$200 \quad Q_c = I_{t,s1} - I_{f,s1} = \frac{I_{f,s1}}{k_{s1}} - I_{f,s1} \quad (5)$$

$$201 \quad Q_f = ET_{s1} - W_g \quad (6)$$

$$202 \quad Q_d = WYLD_{s1} \quad (7)$$

203 where WF is the water footprint of crop production ($\text{m}^3 \text{t}^{-1}$), WF_g is the green footprint ($\text{m}^3 \text{t}^{-1}$), WF_b is
204 the blue water footprint ($\text{m}^3 \text{t}^{-1}$), W_g is the green water consumption during the crop growth period (m^3),
205 W_b is the blue water consumption during the crop growth period (m^3), Y is the crop yield (t), $PRECIP_{s2}$
206 is the precipitation during the crop growth period in Scenario 2 (m^3), $SUPQ_{s2}$ is the surface runoff during
207 the crop growth period in Scenario 2 (m^3), $LATQ_{s2}$ is the soil lateral flow during the crop growth period
208 in Scenario 2 (m^3), Q_c is the amount of water loss in the canal system (m^3), Q_f is the actual ET of field
209 irrigation water (m^3), Q_d is the field discharge (m^3), $I_{t,s1}$ is the total amount of irrigation water diversion
210 in Scenario 1 (m^3), and $I_{f,s1}$ is the actual amount of water irrigated in the field in Scenario 1 (m^3). k_{s1} is
211 the effective utilization coefficient of canal water in Scenario 1 (Obtained from the local Water resources
212 management department), ET_{s1} is the crop actual ET during the crop growth period in Scenario 1 (m^3),
213 $WYLD_{s1}$ is the total amount of water leaving the HRU in Scenario 1 (m^3). The data of parameters
214 $PRECIP_{s2}$, $SUPQ_{s2}$, $LATQ_{s2}$, $I_{t,s1}$, ET_{s1} , $WYLD_{s1}$ were obtained from the SWAT model.

215 **2.5.2 Calculation of water loss during delivery**

216 Water transfer loss is a kind of water loss in the process of channel water delivery, and it is an
217 important part of blue water consumption in crop production. For a piece of cultivated land, the water
218 loss during the process of the crop production includes the loss of water from the water source to the
219 field flowing through the canal system. In the Hetao Irrigation District, irrigation canal is composed of
220 seven grades (general main canal, main canal, sub-main canal, branch canals, lateral canals, field canals,
221 and sub-lateral canals). Because of the complex distribution of canal system and the lack of hydrological
222 data in irrigation districts (the lack of effective utilization coefficient of canal water below the main
223 canal). Therefore, in calculating the water loss of canal system during crop production process, we
224 generalized Hetao Irrigation District into a model similar to the histogram (Fig. 4).

225 We divide the total water loss of canal system into two parts. Part A is the loss of the main canal
 226 and canal, and Part B is the loss of the remaining canal system (the water loss of the sub-main canal and
 227 its sub-channels at all levels). The calculation of water loss in part A is as follows: first, the water loss of
 228 each section is calculated by dividing the main canal into equal distances (10 km). Then the water transfer
 229 loss of each section of the canal is allocated to each field downstream [Equation 10], thereby obtaining
 230 the water transfer loss in the crop production process on the field block. Therefore, the actual water loss
 231 caused by irrigation in a field is the sum of the water loss of the transfer canal and the canal in the
 232 upstream. We assign the actual water loss of the field by irrigation (Q_{ji} , equation 11) to the midpoint of
 233 each section, and use Kriging interpolation in ArcGIS to obtain the water loss distribution map of the
 234 figure a (Part A).

235 Due to the lack of the effective utilization coefficient of canal water and the distribution map of the
 236 canals at all levels and below, the calculation process of the water loss in Part B is as follows: the
 237 remaining canal loss in each irrigation canal is divided by the main canal irrigation and the unit area loss
 238 of the canal control area is obtained. Then, the amount of water loss per unit area within the control range
 239 of each main canal in the irrigation area (Q_j , equation 15) is obtained, and the data is brought into ArcGIS
 240 for the water loss distribution map of figure b (Part B). Finally, the figure a and the figure b are
 241 superimposed and calculated in the ArcGIS using the map algebra module of the spatial analysis tool to
 242 obtain the water loss distribution map of the canal system in HID. The formulas are as follows:

$$243 \quad Q_{ji} = W_{jn} \times \left(\frac{1}{n} + \frac{1}{n-1} + \dots + \frac{1}{n-(i-1)} \right) \quad j \in (1, 2, 3, \dots, m) \quad i \in (1, 2, 3, \dots, n) \quad (8)$$

$$244 \quad W_{jn} = \frac{W_A \times k_j}{n \times S_{jn}} \quad (9)$$

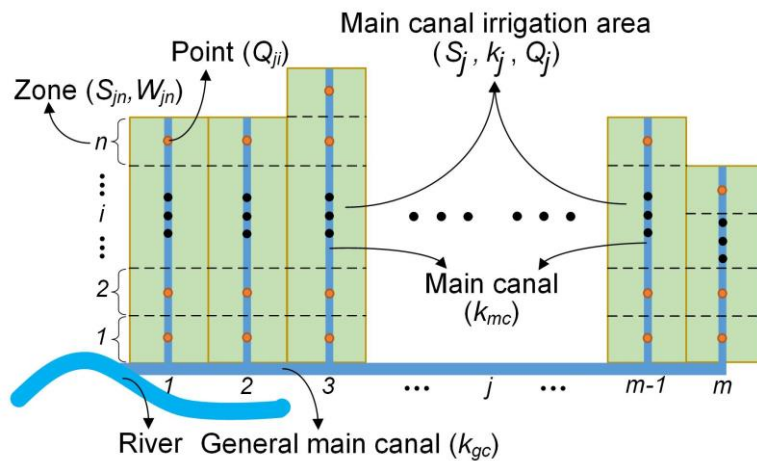
$$245 \quad W_A = I_{t,s1} \times (1 - k_{gc} \times k_{mc}) \quad (10)$$

246
$$S_{jn} = \frac{S_j}{n} \quad (11)$$

247
$$Q_j = \frac{W_B \times k_j}{S_j} \quad (12)$$

248
$$W_B = Q_c - W_A \quad (13)$$

249 where Q_{ji} is the actual amount of water loss per unit area of the i section of the j th main canal in Part
 250 A ($\text{m}^3 \text{ ha}^{-1}$), W_{jn} is the water loss per unit area of the section of the j th main canal in Part A ($\text{m}^3 \text{ ha}^{-1}$), j is
 251 the number of the main canal, i is the number of the equidistance sections in the j th main canal, n is the
 252 total number of the sections in the j th main canal, m is the total number of the main canals, W_A is the
 253 amount of water loss in Part A (m^3), k_j is the coefficient of the water distribution from the general main
 254 canal to the j th main canal, S_{jn} is the area of each sections in the j th main canal (ha), $I_{t,sl}$ is the amount of
 255 total irrigation water diversion in Scenario 1 (m^3), k_{gc} is the water conveyance efficiency of the general
 256 main canal, k_{mc} is the water conveyance efficiency of the main canal, S_j is the area controlled by the j th
 257 main canal (ha), Q_j is the water loss per unit area of the j th main canal in Part B ($\text{m}^3 \text{ ha}^{-1}$), W_B is the
 258 amount of water loss in Part B (m^3), and Q_c is the amount of water loss in the canal system (m^3).



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Fig. 4. Model for calculation of water loss in canal system

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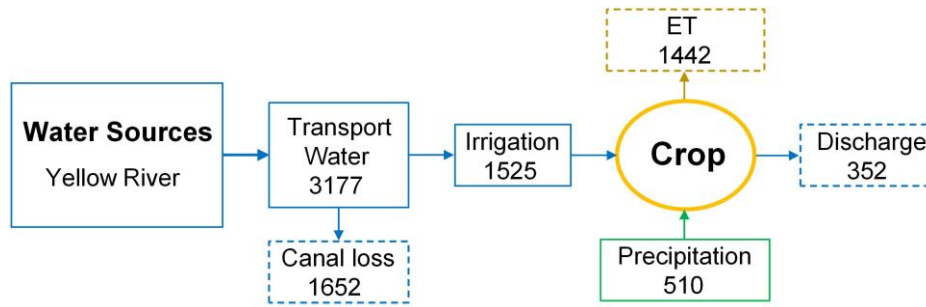
Note: S_{jn} is the area of each sections in the j th main canal, W_{jn} is the water loss per unit area of the

262 section of the j th main canal in Part A, Q_{ji} is the actual amount of water loss per unit area of the i
263 section of the j th main canal, S_j is the area controlled by the j th main canal, k_j is the coefficient of the
264 water distribution from the general main canal to the j th main canal, Q_j is the water loss per unit area of
265 the j th main canal in Part B, k_{gc} is the water conveyance efficiency of the general main canal, k_{mc} is the
266 water conveyance efficiency of the main canal, j is the number of the main canal, i is the number of the
267 equidistance sections in the j th main canal.

268 **3 Results**

269 **3.1 Analysis of the process of crop production and the quantification of hydrological elements** 270 **in each link**

271 Fig. 5 shows the average water input and consumption of the study area in the process of water
272 diversion, transportation, irrigation and drainage from 2006 to 2012. In HID, the water input for irrigation
273 for the three crops in the study area was $3177 \times 10^6 \text{ m}^3$, water loss during transportation in the canals was
274 $1652 \times 10^6 \text{ m}^3$, the actual field irrigation water was $1525 \times 10^6 \text{ m}^3$, precipitation in the farmland was 510
275 $\times 10^6 \text{ m}^3$, the actual ET of the farmland was $1442 \times 10^6 \text{ m}^3$, the field discharge was $352 \times 10^6 \text{ m}^3$, and the
276 groundwater was not considered because the consumption was small. Precipitation and irrigation are the
277 water input items in the process of crop production, and the canal water loss, field actual ET and field
278 drainage are the water output items. For water input, precipitation and irrigation accounted for 25.1%
279 and 74.9%, respectively. For water output, channel water loss, field actual ET and field drainage
280 accounted for 47.9%, 41.8% and 10.3%, respectively. Irrigation is the main water source in the irrigated
281 district, and the water loss in the canals and actual ET are the main water output in the irrigated district.



282

283

Fig. 5. The amount of water during crop growth ($\times 10^6 \text{ m}^3$)

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Green water is the precipitation used for crop growth; therefore, the green water footprint is highly

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correlated with precipitation in its growth period. Wheat's growth period is from April to July, whereas

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that of corn and sunflower is from May to September. During the growth period of wheat, the mean

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precipitation from 2006 to 2012 was 108.9 mm, and for corn and sunflower, the corresponding mean

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precipitation was 176.1 mm. The green footprint of wheat during the growth period was lower than that

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of corn and sunflower because of the lower mean precipitation in the wheat growth period. The green

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water consumption of corn was close to the value of sunflower. The average green water consumption

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of wheat, corn and sunflower were $895 \text{ m}^3 \text{ ha}^{-1}$, $1441 \text{ m}^3 \text{ ha}^{-1}$ and $1419 \text{ m}^3 \text{ ha}^{-1}$ (Fig. 6 a1, b1, c1),

292

respectively. Meanwhile, green water consumption in the high precipitation area was larger, for instance,

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the precipitation during the wheat growth period in Wuyuan reached 116.3 mm, and the green water

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consumption in this region was the largest (up to $995 \text{ m}^3 \text{ ha}^{-1}$). In the growth period of corn and sunflower,

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the precipitation in Wulateqianqi reached 199.4 mm, and the green water consumption in this area was

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again the largest, reaching $1785 \text{ m}^3 \text{ ha}^{-1}$ and $1765 \text{ m}^3 \text{ ha}^{-1}$, respectively.

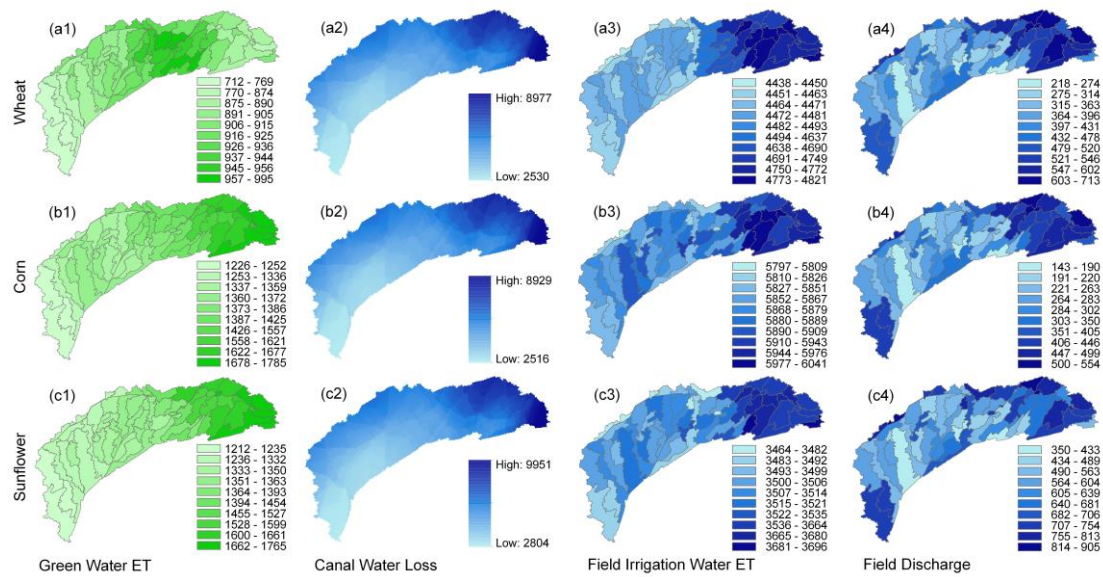


Fig. 6. Spatial distribution of the different water consumption of three crops ($\text{m}^3 \text{ha}^{-1}$)

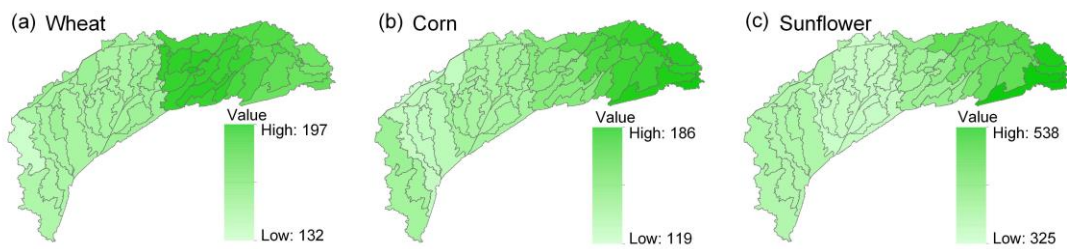
Blue water is the surface water used for crop growth in this study. In blue water consumption, the farther away from the watershed inlets the longer the canal was and the larger the water loss of the three crops. Northeast of the irrigation area (parts of Wuyuan and Wulateqianqi) and due to the far distance from watershed inlets, canal water loss in these places was much higher than that in other areas, and the maximum canal water loss of wheat, corn and sunflower reached $8977 \text{ m}^3 \text{ha}^{-1}$, $8929 \text{ m}^3 \text{ha}^{-1}$ and $9951 \text{ m}^3 \text{ha}^{-1}$, respectively.

The actual ET and the discharge of the three crops was higher in the east than in the west, which was due to the higher evaporation in the east than in the west. Meanwhile, Fig. 6 shows that the actual ET in the field was complementary with discharge. The higher the actual ET, the smaller the discharge and vice versa.

3.2 The regional green water footprint of crop production

The spatial difference of the green water footprint of wheat, corn and sunflower in HID was obvious (Fig. 7). It can be seen from the figure that the overall distribution of the green water footprint of the three crops was higher in the east than it was in the west. However, the distribution of green water

313 footprint was somewhat different for each crop. Wheat had the largest green water footprint in Wuyuan
 314 ($197 \text{ m}^3 \text{ t}^{-1}$) and the lowest in Dengkou ($132 \text{ m}^3 \text{ t}^{-1}$). Corn had the largest green water footprint in
 315 Wulateqianqi ($186 \text{ m}^3 \text{ t}^{-1}$) and the lowest in Hangjinhouqi ($119 \text{ m}^3 \text{ t}^{-1}$), but in Dengkou, it was
 316 approximate to that in Linhe, ranging from 133 to $139 \text{ m}^3 \text{ t}^{-1}$. Sunflower had the largest green water
 317 footprint in Wulateqianqi ($538 \text{ m}^3 \text{ t}^{-1}$) and the lowest in Linhe ($325 \text{ m}^3 \text{ t}^{-1}$).

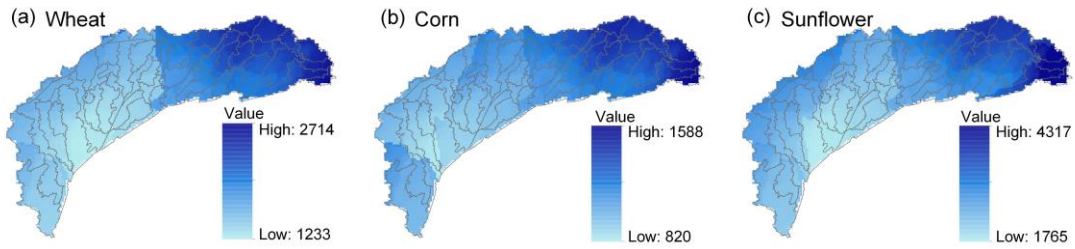


318

319 **Fig. 7.** The spatial distribution of the green water footprint of crop production in the HID ($\text{m}^3 \text{ t}^{-1}$)

320 **3.3 The regional blue water footprint of crop production**

321 The blue water footprint of the crops is produced by blue water that is consumed during crop growth.
 322 The blue water consumption during crop growth mainly includes the loss during transportation, actual
 323 ET and field drainage. Fig. 8 shows the spatial variability of wheat, corn, and sunflower in HID. The
 324 overall distribution of the total water footprint of the three crops was higher in the east than in the west
 325 and higher in the north than in the south. However, the specific distribution was somewhat different for
 326 each crop. Wheat had the largest blue water footprint in Wulateqianqi ($2714 \text{ m}^3 \text{ t}^{-1}$) and the lowest in
 327 southern Linhe ($1233 \text{ m}^3 \text{ t}^{-1}$). Corn had the largest blue water footprint in northern Wulateqianqi (1588
 328 $\text{m}^3 \text{ t}^{-1}$) and the lowest in southern Hangjinhouqi ($820 \text{ m}^3 \text{ t}^{-1}$). Sunflower had the largest blue water
 329 footprint in northern Wulateqianqi ($4317 \text{ m}^3 \text{ t}^{-1}$) and the lowest in southern Linhe ($1765 \text{ m}^3 \text{ t}^{-1}$).

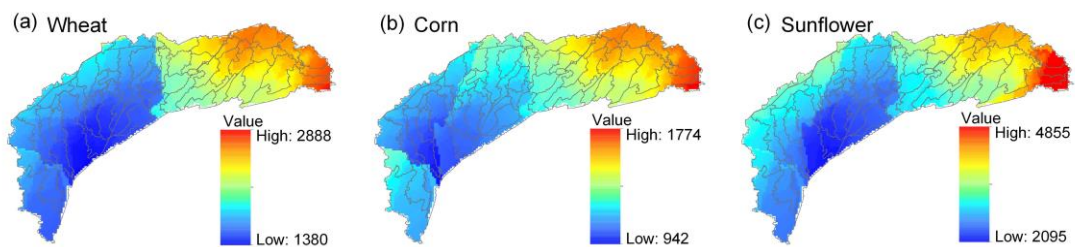


330

331 **Fig. 8.** The spatial distribution of the blue water footprint of crop production in the HID ($\text{m}^3 \text{t}^{-1}$)

332 **3.4 The regional total water footprint of crop production**

333 The total water footprint of crop production consists of both blue and green water footprint during
 334 the crop growth period. Fig. 9 shows the total water footprint of crop production and spatial variability
 335 of wheat, corn, and sunflower in HID. The overall distribution of the total water footprint of the three
 336 crops was higher in the east (Wulateqianqi and Wuyuan) than it was in the west (Dengkou), followed by
 337 the central region (Hangjinhouqi and Linhe) and was higher in the north than in the south. However, the
 338 specific distribution was somewhat different for each crop. Wheat had the largest total water footprint in
 339 the east (Wulateqianqi, $2888 \text{ m}^3 \text{t}^{-1}$) and the lowest in southern Linhe ($1380 \text{ m}^3 \text{t}^{-1}$). Corn had the largest
 340 total water footprint in the east (Wulateqianqi, $1774 \text{ m}^3 \text{t}^{-1}$) and the lowest in southern Hangjinhouqi (942
 341 $\text{m}^3 \text{t}^{-1}$). Sunflower had the largest total water footprint in the east (Wulateqianqi, $4855 \text{ m}^3 \text{t}^{-1}$) and the
 342 lowest value was in southern Linhe ($2095 \text{ m}^3 \text{t}^{-1}$). The total water footprint of crop production also varied
 343 across crops. The largest of the average total water footprint in the HID was sunflower, followed by
 344 wheat and corn. The blue water footprint of wheat, corn and sunflower accounted for 89%, 87% and 86%
 345 of the total water footprint, respectively.



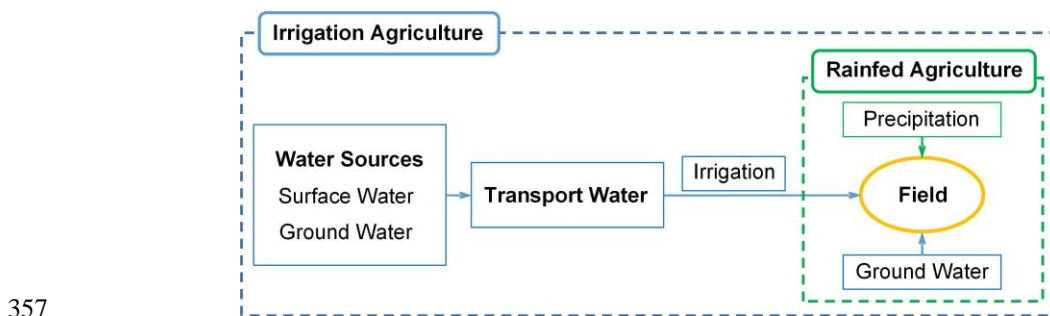
346

347 **Fig. 9.** The spatial distribution of the total water footprint of crop production in the HID ($\text{m}^3 \text{t}^{-1}$)

348 **4 Discussion**

349 **4.1 The regional scale and field scale methods for calculating crop production water footprint**

350 In this paper, the calculation method for calculating crop production water footprint is divided into
351 the field scale and regional scale method, according to the calculation boundary of water consumption in
352 crop growth process. The field scale water footprint is composed of the transpiration of crops and the
353 evaporation of soil, and the water loss during transportation is not included. The regional scale water
354 footprint calculation method considers all of the water consumption related to crop growth from the water
355 source to the field. It not only includes the ET from the field but also the water loss during transportation
356 in the canal system and the water loss discharged out of the region.



357

358 **Fig. 10.** Irrigation agriculture and rainfed agriculture

359 Currently, irrigated farmland occupied 39.6% of the total arable land in China (NBSC, 2016).
360 Globally, irrigated area accounted for 20.6% of all arable land (FAO, 2016). Overall, the yields of
361 irrigation agriculture are much higher than that of rainfed agriculture. Fig. 10 illustrates the water sources
362 and use conditions of two types of agriculture. In irrigated agriculture, water (blue water) goes through
363 the following processes from water source to field: water diversion, water transportation (canal system
364 or pipeline), different methods (surface irrigation, sprinkler irrigation, drip irrigation, etc.) to irrigate
365 crops, and excess water discharged from the field. In irrigated agricultural production, especially in areas

366 where water is transported through channels for irrigation, a large amount of water is lost (canal leakage
367 or water evaporation) during the transportation process, which is indirectly used for crop production. The
368 transportation process generates a large amount of cost (energy, machinery, facilities, management, etc.
369 for water diversion). Therefore, this water loss is also a part of the crop production water footprint. In
370 China, the irrigation water consumption was $360 \times 10^9 \text{ m}^3$, and the effective utilization coefficient of
371 irrigation water was 0.53 (MWR, 2015), which indicated that about $169.2 \times 10^9 \text{ m}^3$ of water resources
372 were lost in the process of transportation and irrigation. It is necessary to include the amount of blue
373 water loss during irrigation into the crop production water footprint. Fig. 11 is the calculation range of
374 the regional scale and field scale method of crop production water footprint. Consequently, the
375 calculation method on regional scale can comprehensively calculate all the water consumption in the
376 crop production process, and the calculated results of water footprint are more accurate, while the field
377 scale method only calculates part of the water consumption in the process (Zhao et al., 2009; Bocchiola
378 et al., 2015). At the same time, the calculation method on regional scale proposed in this study improves
379 the resolution of the water footprint results. It can also reflect the variation of the water footprint within
380 the region, and more effectively discover the water footprint hotspots, and avoid the shortcomings that
381 can only reflect the mean value of the regional results due to low resolution (Vanham and Bidoglio, 2013;
382 Zhuo et al., 2016). These two advantages of the regional scale approach can help local authorities to
383 develop more rational water allocation and management policies.

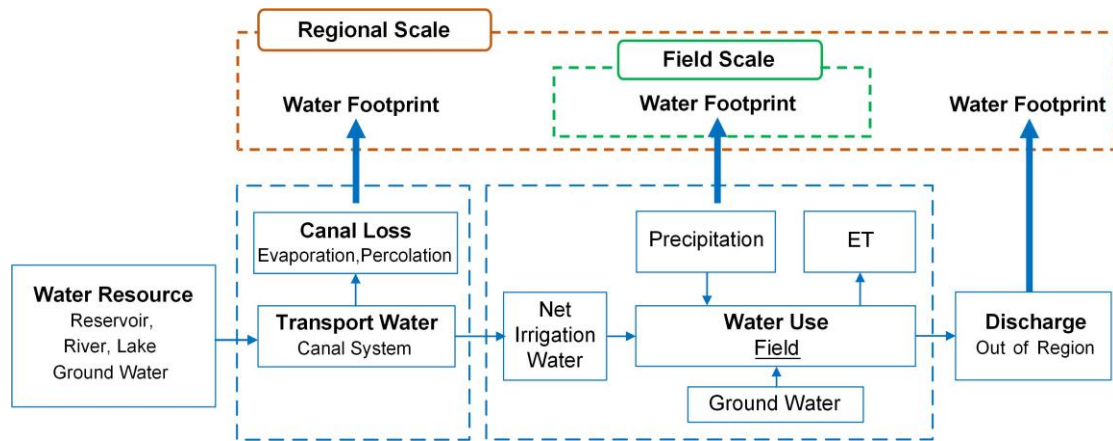


Fig. 11. The different scales of calculating water footprint

4.2 Comparison of the results of two methods

For the field scale method, the calculated value was less than the actual value because it did not consider the loss of water during transportation or discharge, and the actual water footprint of irrigation agriculture cannot be precisely assessed. At present, most studies use field scale method (e.g. CROPWAT model) (Lovarelli et al., 2016), so these studies mainly focus on agricultural water use at field scale, lacking an analysis of the entire process of agricultural production water use, which is also the shortcomings of the current research on the crop production water footprint. Therefore, using the regional scale method to calculate the crop water footprint, especially in irrigation agriculture, is the basis for a comprehensive and accurate evaluation of a crop production water footprint in China and other regions or countries.

In HID, the water footprint of three crops (wheat, corn and sunflower) calculated by regional scale method were 1380-2888 $\text{m}^3 \text{t}^{-1}$, 942-1774 $\text{m}^3 \text{t}^{-1}$, and 2095-4855 $\text{m}^3 \text{t}^{-1}$, respectively. These values were higher than the results calculated by the field scale method. Cao et al. (2014) calculated the average crop water footprint of irrigation agriculture in China from 1998 to 2010 and the average value of many crops in the Inner Mongolia autonomous region (including HID) was 1556 $\text{m}^3 \text{t}^{-1}$. Sun et al. (2013b) calculated the average water footprint of HID by using regional scale method and water balance principle and the

402 result was $3910 \text{ m}^3 \text{ t}^{-1}$. The proportion of blue water and green water was 90.9% and 9.1% respectively.
403 This result was the average water footprint of many crops, and the value was approximate to our results.
404 Qin et al. (2016) calculated the water footprint of sunflower in Jilin province by using field scale method
405 and found that the water footprint of sunflower in this area from 2006 to 2008 were $1280 \text{ m}^3 \text{ t}^{-1}$, 1684 m^3
406 t^{-1} and $1726 \text{ m}^3 \text{ t}^{-1}$, respectively, which was smaller than this study. This is because Jilin's water footprint
407 is mainly composed of green water footprint, which reached 95%, and its blue water footprint is smaller.
408 In addition, these studies all showed the water footprint of the region, which cannot distinguish the spatial
409 distribution of the crop production water footprint within the region, and has a limited impact on the
410 improvement of local water resources management.

411 This method also has limitations. The method requires more data types (e.g. DEM, land use, soil
412 data, climate data, hydrological data, and crop management data), and higher data resolution. Therefore,
413 the method is not applicable to areas where the above data are lacking.

414 **4.3 Strategies for adjusting the crop production water footprint**

415 The water footprint of crop production is affected by crop species. Different crops have different
416 water use characteristics and different growth periods. Therefore, adjusting the crop planting structure
417 can change the water supply in the region (Fasakhodi et al., 2010), which in turn affects the water
418 footprint of crop production. At the same time, changing the crop pattern, planting crops which growth
419 periods are consistent with the precipitation period can increase the utilization of green water, reduce the
420 consumption of blue water, and reduce the pressure on local water resources (Liu et al., 2018). This study
421 found that in the HID, the growth period of sunflower is basically the same as the precipitation period.
422 Consequently, expanding the planting area of sunflower can make better use of local precipitation
423 resources and reduce the use of blue water.

424 Crop yield is an important factor affecting the water footprint of crop production. Selecting crop
425 varieties with high yields and improving agricultural management measures play an important role in
426 increasing crop yields. Sun et al. (2013b) found that improving agricultural management measures is an
427 important factor to increase crop yield and reduce water footprint of crop production. Liu et al. (2014,
428 2015) discussed the water use situation and virtual water flow in Hetao Irrigation District and found that
429 crop yield had an important impact on the water footprint of crop production, and with the increasing of
430 crop yield per unit area, the water footprint of crop production had declined.

431 The efficiency of irrigation system is affected by the way of water transportation, the condition of
432 canal system, the irrigation technology and so on. Therefore, the water use efficiency of the regional
433 irrigation system can be improved by changing the water delivery method (from the channel to the
434 pipeline) and the irrigation method (such as dropper, sprinkler and other advanced irrigation
435 technologies). For the study area, the results show that more than half of the water resources were lost
436 during the process of canal water transport and irrigation. Therefore, adopting anti-seepage measures to
437 reduce the leakage of canal system, and adopting advanced irrigation technology to reduce the amount
438 of irrigation water will help to reduce the water footprint of crop production in this region.

439 **5 Conclusions**

440 In this study, we proposed an improved regional scale method for calculating crop production water
441 footprint. This method was based on the hydrological model (SWAT model), combined the irrigation
442 parameters of the irrigation area (water conveyance efficiency of canal), and calculated the crop
443 production water footprint.

444 The method provided whole hydrological processes analysis for water use during crop production,
445 including water diversion, irrigation/precipitation, field evapotranspiration and drainage, Therefore, the

446 method contributed to establish a more comprehensive calculation of water consumption during the crop
447 growth period and more precisely quantify crop production water footprint. The method can be applied
448 to calculate the crop production water footprint at both the field and regional scale. In HID, the main
449 water consumption occurs during the crop growth period; the canal water loss was $1652 \times 10^6 \text{ m}^3$, and
450 actual ET in the field was $1442 \times 10^6 \text{ m}^3$, which accounted for 47.9% and 41.8% of the total used,
451 respectively.

452 The regional climate, the condition of irrigation system and the crop yield are the main factors that
453 affect the water footprint of crop production. The area with higher crop yield per unit area, higher
454 efficiency of irrigation water use, less irrigation water loss and closer to source of water has lower crop
455 production water footprint. Water loss during transportation increased with the increasing distance of the
456 canals, and the farther away from the watershed inlets they were, the more water was lost, the values
457 were higher in the east than they were in the west in the study area.

458 Due to special climatic conditions, crops in Hetao Irrigation District mainly depend on irrigation
459 water in the production process. Overall, in the composition of water footprint in Hetao Irrigation District,
460 blue water footprint accounts for more than 86%. Therefore, applying water-saving irrigation technology,
461 increasing channel lining rate and reducing the loss of irrigation water are the main ways to adjust and
462 control the water footprint of crop production in this area.

463 Based on the SWAT model, this paper analyzed the utilization and consumption of water
464 resources during crop production in irrigated areas, which provided a hydrological mechanism for
465 quantifying the water footprint of crop production. However, the SWAT model does not consider the
466 relation of groundwater flow between different subbasins. At the same time, the shallow groundwater
467 evaporation is based on the soil as a medium and directly into the atmosphere, the model cannot

468 accurately quantify the recharge of shallow groundwater to soil water. Consequently, the SWAT model
469 cannot accurately simulate the shallow groundwater consumption of crops. Therefore, combining the
470 groundwater model, analyzing the flow of water in the process of regional agricultural production, and
471 then quantifying the water footprint of crop production is the direction of further research.

472 **Acknowledgements:**

473 This work is jointly supported by the National Natural Science Foundation of China (51409218;
474 51609063), Innovative Talents Promotion Project in Shaanxi Province of China (2018KJXX-053), the
475 Open Research Fund of the State Key Laboratory of Simulation and Regulation of Water Cycle in River
476 Basin at the China Institute of Water Resources and Hydropower Research (IWHR-SKL-201601) and
477 Young Scholar Project of Cyrus Tang Foundation.

478 **Author Contributions:** Pute Wu, Shikun Sun and Yubao Wang designed the study. Xiaobo Luan, Yali
479 Yin, Xuerui Gao and Jing Liu did the literature search and data collection. Xiaobo Luan, Shikun Sun and
480 Yali Yin managed and analyzed the data. Xiaobo Luan and Suikun Sun drew the figures and wrote the
481 paper. All authors discussed and commented on the manuscript.

482 **References:**

- 483 AHID.: The Administration of Hetao Irrigation District (AHID), Bayannaoer Department of Water, Inner
484 Mongolia Autonomous Region, China, available at: <http://www.htgq.gov.cn/>, 2015.
- 485 Allen, R. G., Pereira, L. S., Raes, D., and Smith, M.: Crop evapotranspiration: Guidelines for computing
486 crop water requirements. FAO Irrigation and Drainage Paper 56, Rome, 1998.
- 487 Abbaspour, K. C.: SWAT-CUP 2012: SWAT Calibration and Uncertainty Programs - A User Manual,
488 Eawag: Swiss Federal Institute Science and Technology, 2012.
- 489 Abbaspour, K. C., Vejdani M., Haghightat S.: SWAT-CUP calibration and uncertainty programs for
490 SWAT, Modsim 2007: International Congress on Modelling and Simulation: Land, Water and
491 Environmental Management: Integrated Systems for Sustainability, Christchurch, New Zealand,
492 2007.
- 493 Bao, C., Fang, C.: Water Resources Flows Related to Urbanization in China: Challenges and Perspectives
494 for Water Management and Urban Development, *Water Resour. Mang.*, 26, 531-552, doi:
495 10.1007/s11269-011-9930-y, 2012.
- 496 Bocchiola, D.: Impact of potential climate change on crop yield and water footprint of rice in the Po
497 valley of Italy, *Agr. Syst.*, 139, 223-237, doi: 10.1016/j.agry.2015.07.009, 2015.
- 498 Bocchiola, D., Nana, E. and Soncini, A.: Impact of climate change scenarios on crop yield and water
499 footprint of maize in the Po valley of Italy, *Arg. Water Manage.*, 116, 50-61, doi:
500 10.1016/j.agwat.2012.10.009, 2013.
- 501 Cao, X., Wu, P., Wang, Y. and Zhao, X.: Water Footprint of Grain Product in Irrigated Farmland of
502 China, *Water Resour. Manag.*, 28, 2213-2227, doi: 10.1007/s11269-014-0607-1, 2014.
- 503 CAS: Geospatial Data Cloud site (GSCloud), Computer Network Information Center, Chinese Academy

504 of Sciences, available at: <http://www.gscloud.cn>, 2009a.

505 CAS: China Soil Scientific Database (CSDB), Soil Research Center, Institute of Soil Science, Chinese
506 Academy of Sciences, available at: <http://www.soil.csdb.cn/>, 2009b.

507 CAS: Data Center for Resources and Environmental Sciences (RESDC), Chinese Academy of Sciences,
508 available at: <http://www.resdc.cn>, 2010.

509 Chen, J.: Rapid urbanization in China: a real challenge to soil protection and food security. *Catena*, 69,
510 1-15, doi: 10.1016/j.catena.2006.04.019, 2007.

511 Chukalla, A. D., Krol, M. S. and Hoekstra, A. Y.: Green and blue water footprint reduction in irrigated
512 agriculture: effect of irrigation techniques, irrigation strategies and mulching, *Hydrol. Earth Syst.*
513 *Sc*, 12, 6945-6979, doi: 10.5194/hess-19-4877-2015, 2015.

514 Deng, X. P., Shan, L., Zhang, H. and Turner, N. C.: Improving agricultural water use efficiency in arid
515 and semiarid areas of China, *Arg. Water Manage.*, 80, 23-40, doi: 10.1016/j.agwat.2005.07.021,
516 2006.

517 Doll, P. and Siebert, S.: Global modeling of irrigation water requirements, *Water Resour. Res.*, 38, 1037-
518 1048, doi: 10.1029/2001WR000355, 2002.

519 Du, T., Kang, S., Zhang, X., and Zhang, J.: China's food security is threatened by the unsustainable use
520 of water resources in North and Northwest China, *Food Energy Secur.*, 3, 7-18, doi: 10.1002/fes3.40,
521 2014.

522 Duh, J., Shandas, V., Chang, H., and George, L. A.: Rates of urbanisation and the resiliency of air and
523 water quality, *Sci. Total Environ.*, 400, 238-256, doi: 10.1016/j.scitotenv.2008.05.002, 2008.

524 Elliott, J., Deryng, D., Muller, C., Frieler, K., Konzmann, M., Gerten, D., Glotter, M., Flörke, M., Wada,
525 Y., Best, N., Eisner, S., Fekete, B. M., Folberth, C., Foster, I., Gosling, S. N., Haddeland, I.,

526 Khabarov, N., Ludwig, F., Masaki, Y., Olin, S., Rosenzweig, C., Ruane, A. C., Satoh, Y., Schmid,
527 E., Stacke, T., Tang, Q. H., and Wisser, D.: Constraints and potentials of future irrigation water
528 availability on agricultural production under climate change, *P. Natl. Acad. Sci. USA.*, 111, 3239-
529 3244, doi: 10.1073/pnas.1222474110, 2014.

530 FAO. Food and Agriculture Organization of the United Nations, Land and Water Development Division,
531 CROPWAT model, Rome, Italy, [http://www.fao.org/land-water/databases-and-software/cropwat/e](http://www.fao.org/land-water/databases-and-software/cropwat/en/)
532 [n/](http://www.fao.org/land-water/databases-and-software/cropwat/en/), 2010

533 FAO. AQUASTAT website, Food and Agriculture Organization of the United Nations, (website last
534 access: 20 September 2017), 2016.

535 Fasakhodi, A. A., Nouri, S. H., and Amini, M.: Water resources sustainability and optimal cropping
536 pattern in farming systems: a multi-objective fractional goal programming approach, *Water Resour.*
537 *Manag.*, 24, 4639-4657, doi: 10.1007/s11269-010-9683-z, 2010.

538 Haddeland, I., Heinke, J., Biemans, H., Eisner, S., Flörke, M., Hanasaki, N., Konzmann, M., Ludwig, F.,
539 Masaki, Y., Schewe, J., Stacke, T., Tessler, Z. D., Wada, Y., and Wisser, D.: Global water resources
540 affected by human interventions and climate change, *P. Natl. Acad. Sci. USA.*, 111, 3251-3256, doi:
541 10.1073/pnas.1222475110, 2014.

542 Haverkamp, S., Srinivasan, R., Frede, H. G., and Santhi, C.: Subwatershed spatial analysis tool:
543 discretization of a distributed hydrologic model by statistical criteria, *J. Am. Water Resour. As.*, 38,
544 1723-1733, doi: 10.1111/j.1752-1688.2002.tb04377.x, 2002.

545 Hoekstra, A. Y. (ed): Virtual water trade: Proceedings of the International Expert Meeting on Virtual
546 Water Trade, 12-13 December 2002, Value of Water Research Report Series No 12, UNESCO-IHE,
547 Delft, Netherlands, <http://waterfootprint.org/media/downloads/Report12.pdf>, 2003.

548 Hoekstra, A. Y., Chapagain, A. K., Aldaya M. M., and Mekonnen M. M.: The water footprint assessment
549 manual-setting the global standard, London • Washington, 2011.

550 Jiang, Y.: China's water scarcity, *J. Environ. Manage.*, 90, 3185-3196, doi: 10.1016/j.jenvma
551 n.2009.04.016, 2009.

552 Khan, S., Hanjra, M. A., and Mu, J. X.: Water management and crop production for food security in
553 China: a review, *Agr. Water Manage.*, 96, 349-360, doi: 10.1016/j.agwat.2008.09.022, 2009.

554 Liu, J. G., Williams J. R., Zehnder, A. J. B., and Hong, Y.: GEPIC-modelling wheat yield and crop water
555 productivity with high resolution on a global scale, *AGR. SYST.*, 94, 478-493, doi:
556 10.1016/j.agsy.2006.11.019, 2007.

557 Liu, J., Cao, X., Li, B., and Yu, Z.: Analysis of blue and green water consumption at the irrigation district
558 scale, *SUSTAINABILITY-BASEL*, 10, 305, doi: 10.3390/su10020305, 2018.

559 Liu, J., Sun, S., Wu, P., Wang, Y., and Zhao, X.: Evaluation of crop production, trade, and consumption
560 from the perspective of water resources: A case study of the Hetao irrigation district, China, for
561 1960-2010, *SCI. TOTAL ENVIRON.*, 505, 1174-1181, doi: 10.1016/j.scitotenv.2014.10.088, 2015.

562 Liu, J., Wu, P., Wang, Y., Zhao, X., Sun, S., and Cao, X.: Impacts of changing cropping pattern on virtual
563 water flows related to crops transfer: a case study for the Hetao irrigation district, China, *J. SCI.
564 FOOD AGR.*, 94, 2992-3000, doi: 10.1002/jsfa.6645, 2014.

565 Liu, J., Yang, H., and Savenije, H. H.: China's move to higher-meat diet hits water security, *NATURE*,
566 454, 397-397, doi: 10.1038/454397a, 2008.

567 Lovarelli, D., Bacenetti, J., and Fiala, M.: Water footprint of crop productions: a review, *SCI. TOTAL
568 ENVIRON.*, 236-251, doi: 10.1016/j.scitotenv.2016.01.022, 2016.

569 Luan, X. B., Wu, P. T., Sun, S. K., Wang, Y. B., Gao, X. R.: Quantitative study of the crop production

570 water footprint using the SWAT model, *ECOL. INDIC.*, 89, 1-10, doi: 10.1016/j.ecolin d.2018.0
571 1.046, 2018.

572 Mekonnen, M. M., and Hoekstra, A. Y.: The green, blue and grey water footprint of crops and derived
573 crop products, *HYDROL. EARTH SYST. SC.*, 15, 1577-1600, doi: 10.5194/hess-15-1577-2011,
574 2011.

575 MWR (Ministry of Water Resources People's Republic of China): China water resources bulletin 2014,
576 China Water and Power Press, Beijing, 2015.

577 NBSC (National Bureau of Statistics of China): China Statistical Yearbook 2016, China Statistics Press,
578 Beijing, 2016.

579 Neitsch, S. L., Arnold, J. G., Kiniry, J. R., and Williams, J. R.: Soil and Water Assessment Tool:
580 Theoretical Documentation, Version 2009, Texas Water Resources Institute, 2011.

581 Nijssen, B., Odonnell, G. M., Hamlet, A. F., and Lettenmaier, D. P.: Hydrologic Sensitivity of Global
582 Rivers to Climate Change, *CLIMATIC CHANGE*, 50, 143-175, doi: 10.1023/A:1010616428763,
583 2001.

584 NMIC.: China meteorological data network (CMA), National Meteorological Information Center, China
585 (<http://data.cma.cn/>), 2015.

586 Pasquale, S., Theodorec, H., Dirk, R., and Elias, F.: Aquacrop--the FAO crop model to simulate yield
587 response to water: I. concepts and underlying principles, *AGRON. J.*, 101, 448-459, doi:
588 10.2134/agronj2008.0139s, 2009.

589 Piao, S., Ciais, P., Huang, Y., Shen, Z., Peng, S., Li, J., Zhou, L., Liu, H., Ma, Y., Ding, Y., Friedingstein,
590 P., Liu, C., Tan, K., Yu, Y., Zhang, T., Fang, J.: The impacts of climate change on water resources
591 and agriculture in China, *NATURE*, 467, 43-51, doi: 10.1038/nature09364, 2010.

592 Qin, L., Jin, Y., Duan, P., and He, H.: Field-based experimental water footprint study of sunflower growth
593 in a semi-arid region of China, *J. SCI. FOOD AGR.*, 96, 3266-3273, doi: 10.1002/jsfa.7726, 2016.

594 Schwarzenbach, R. P., Egli, T., Hofstetter, T. B., Von Gunten, U., and Wehrli, B.: Global water pollution
595 and human health. *ANNU. REV. ENV. RESOUR.*, 35, 109-136, doi: 10.1146/ann urev-environ-
596 100809-125342, 2010.

597 Shi, R., Ukaew, S., Archer, D. W., Lee, J. H., Pearlson, M. N., Lewis, K. C., and Shonnard, D. R.: Life
598 Cycle Water Footprint Analysis for Rapeseed Derived Jet Fuel in North Dakota, *ACS SUSTAIN.*
599 *CHEM. ENG.*, 5, 3845-3854, doi: 10.1021/acssuschemeng.6b02956, 2017.

600 Shiklomanov, I. A.: Appraisal and assessment of world water resources, *WATER INT.*, 25, 11-32, doi:
601 10.1080/02508060008686794, 2000.

602 Sun, S. K., Wu, P. T., Wang, Y. B., and Zhao, X. N.: Temporal variability of water footprint for maize
603 production: the case of Beijing from 1978 to 2008. *WATER RESOUR. MANAG.*, 27, 2447-2463,
604 doi: 10.1007/s11269-013-0296-1, 2013a.

605 Sun, S. K., Wu, P. T., Wang, Y. B., and Zhao, X. N.: The virtual water content of major grain crops and
606 virtual water flows between regions in China, *J. SCI. FOOD AGR.*, 93, 1427-37, doi:
607 10.1002/jsfa.5911, 2013c.

608 Sun, S. K., Wu, P. T., Wang, Y. B., Zhao, X. N., Liu, J., and Zhang, X.: The impacts of inter-annual
609 climate variability and agricultural inputs on water footprint of crop production in an irrigation
610 district of China, *SCI. TOTAL ENVIRON.*, 444, 498-507, doi: 10.1016/j.scitoten v.2012.12.016,
611 2013b.

612 Vanham, D., and Bidoglio, G.: A review on the indicator water footprint for the EU28. *ECOL. INDIC.*,
613 26, 61-75, doi: 10.1016/j.ecolind.2012.10.021, 2013.

614 Vörösmarty, C. J., McIntyre, P. B., Gessner, M. O., Dudgeon, D., Prusevich, A., and Green, P., et al.:
615 Global threats to human water security and river biodiversity, *NATURE*, 467, 555-561, doi:
616 10.1038/nature09440, 2010.

617 Wang, Y. B., Wu, P. T., Zhao, X. N., and Engel, B. A.: Virtual water flows of grain within China and its
618 impact on water resource and grain security in 2010, *ECOL. ENG.*, 69, 255-264, doi:
619 10.1016/j.ecoleng.2014.03.057, 2014.

620 Williams, J. R., Jones, C. A., Kiniry, J. R., and Spanel, D. A.: The EPIC crop growth-model, *T. ASABE*,
621 32, 497-511, doi: 10.13031/2013.31032, 1989.

622 Zhao, X., Chen, B., and Yang, Z. F.: National water footprint in an input-output framework - a case study
623 of China 2002, *ECOL. MODEL.*, 220, 245-253, doi: 10.1016/j.ecolmodel.2008.09.016, 2009.

624 Zhuo, L., Mekonnen, M. M., and Hoekstra, A. Y.: Benchmark levels for the consumptive water footprint
625 of crop production for different environmental conditions: a case study for winter wheat in China,
626 *HYDROL. EARTH SYST. SC.*, 20, 4547-4559, doi: 10.5194/hess-20-4547-2016, 2016.