



1 **Groundwater-surface water relations in regulated lowland catchments;**
2 **hydrological and hydrochemical effects of a major change in surface water**
3 **level management**

4

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12

13 **Abstract**

14 In lowland deltas with intensive land use such as The Netherlands, surface water levels are
15 tightly controlled by inlet of diverted river water during dry periods and discharge via large-scale
16 pumping stations during wet periods. The conventional water level regime in these polder
17 catchments is either a fixed water level year-round or an unnatural regime with a lower winter
18 level and a higher summer level in order to optimize hydrological conditions for agricultural land
19 use. The objective of this study was to assess the hydrological and hydrochemical effects of
20 changing the water level management from a conventional fixed water level regime to a flexible,
21 more natural regime with low levels in summer and high levels in winter between predefined
22 minimum and maximum levels.

23

24 Ten study catchments were hydrologically isolated and equipped with controlled inlet and outlet
25 weirs or pumping stations. The water level management was converted into a flexible regime.
26 We used water and solute balance modeling for catchment-scale assessments of changes in
27 water and solute fluxes.

28



1 Our model results show relevant changes in the water exchange fluxes between the polder
2 catchment and the regional water system and between the groundwater, surface water, and field
3 surface storage domains within the catchment. Compared to the reference water level regime,
4 the flexible water level regime water balance scenario showed increased surface water residence
5 times, reduced inlet and outlet fluxes, reduced groundwater-surface water exchange, and in
6 some catchments increased overland flow. The solute balance results show a general reduction
7 of chloride concentrations and a general increase in N-tot concentrations. The total phosphorus
8 (P-tot) and sulfate (SO₄) concentration responses varied and depended on catchment-specific
9 characteristics.

10

11 For our study catchments, our analyses provided a quantification of the water flux changes after
12 converting towards flexible water level management. Regarding the water quality effects, this
13 study identified the risks of increased overland flow in former agricultural fields with nutrient
14 enriched top soils and of increased seepage of deep groundwater which can deliver extra
15 nutrients to surface water. At a global scale, catchments in low-lying and subsiding deltas are
16 increasingly being managed in a similar way as the Dutch polders. Applying our water and solute
17 balance approach to these areas may prevent unexpected consequences of the implemented
18 water level regimes.

19

20 **Keywords**

21 Water level management, Hydrology, Water quality, Water and solute balance modelling

22

23 **1 Introduction**

24 Water levels in the flat, low-lying polder catchments in the western and northern parts of the
25 Netherlands are tightly regulated. Water authorities maintain either a year-round fixed water level
26 or fixed summer and winter levels. During dry periods, these water levels are maintained by
27 diverting river water into these catchments (e.g. Roelofs, 1991, Rozemeijer, 2012). This water
28 inlet prevents drought damage in agriculture and wetland nature reserves. In addition, too low
29 water levels would accelerate peat oxidation which causes land subsidence (e.g. Schothorst,



1 1977; Hoogland et al., 2012) and greenhouse gas emissions (e.g. Schrier-Uijl et al., 2014).
2 During wet conditions, the water levels are maintained to prevent water damage to agriculture,
3 roads, and buildings. Excess water is discharged towards the regional rivers and canal systems
4 via large-scale pumping stations.

5

6 The conventional water level regime in polder catchments is either a fixed water level year-round,
7 or a winter level that is lower than the summer level. These unnatural regimes optimize
8 hydrological conditions for agricultural land use. During the wet winters the drainage of
9 groundwater is enhanced by relatively low surface water levels. During the dry summer however,
10 the infiltration of surface water to replenish groundwater and soil moisture is enhanced by high
11 surface water levels due to the inlet of diverted river water (Hendriks et al., 2013).

12

13 Several water authorities consider changing the water level regimes to a more natural regime,
14 where surface water levels fluctuate freely between predefined minimum and maximum water
15 levels. Such a change in water level regime is often motivated from an ecological perspective.
16 Restoring natural water level fluctuations stimulates dispersal and germination of plant species
17 and increases vegetation diversity in the riparian zones (Sarneel et al., 2014). In a flexible water
18 level regime, temporal flooding distributes seeds over the riparian gradient. The subsequent
19 receding water levels create a variety of soil moisture conditions that meet the germination
20 requirements of many plant species (Lenssen et al., 1998; Sarneel et al., 2014). Van Leeuwen et
21 al. (2014) reported seed bank experiments in which the spring drawdown of water levels in a
22 flexible regime resulted in higher species richness and diversity compared to the conventional
23 regimes with fixed water levels.

24

25 An additional advantage of restoring water level fluctuations is the potential reduction of the
26 amount of diverted river water inlet needed from the regional water system. As long as the
27 minimum water level is not reached, no inlet water is needed. In addition summer storm
28 precipitation water does not leave the polder as long as the maximum water level is not



1 exceeded. The reduced need for inlet water enhances the regional water availability during
2 droughts and may improve water quality within the polders.

3

4 Several authors reported negative consequences of diverted river water inlet for the water quality
5 and ecology of receiving surface waters. For example, Delaune et al. (2005) and Miao et al.
6 (2011) studied eutrophication caused by the inlet of Mississippi River water into N-limited wetland
7 ecosystems. Roelofs (1991), Smolders et al. (2006), and Lamers et al. (2013) reported that the
8 inlet of alkaline (HCO_3 and SO_4 -rich) river water may also enhance the release of phosphorus
9 from sediments in the receiving surface waters.

10

11 A final potential advantage of converting to a flexible water level regime and allowing water levels
12 to rise during the wet winter period is the reduction of the amount of discharge that needs to be
13 pumped out of the polder into the regional river system. This would reduce the consumption of
14 energy and the costs for maintenance of the pumping stations and reduce the pressure on the
15 regional water system during high discharges. Introduction of a flexible water level management
16 is often postponed or hampered due to the fear for negative agricultural or recreational impacts,
17 e.g. flooding of agricultural land in winter and spring and insufficient water depth for boating in
18 summer.

19

20 Although water quality is an important issue in polder catchments in Netherlands, little is known
21 about the combined effects of restoring natural water level fluctuations on water and solute
22 fluxes. Small changes in nutrient concentrations in polders can result in dominance of undesired
23 species like duckweed (Lemnaceae) or floating fern (Azolla), a severe reduction in biodiversity,
24 and hypoxia (Janse and Van Puijenbroek, 1998; Hellman and Vermaat, 2012; Vermaat et al.,
25 2016). The reduced impact of water inlet suggests a positive effect of flexible water level
26 management for chemical water quality. However, other processes may induce reverse effects in
27 these hydrological complex polder systems, such as the changing interaction between
28 groundwater and surface water. For example, higher water levels in winter may enhance the
29 leaching of nutrients from the enriched upper part of the soils (Rozemeijer and Broers, 2008; Van



1 der Grift, 2016). At the same time, lower water levels in summer may enhance seepage of
2 nutrient-rich and/or saline groundwater from the peaty subsurface (e.g. Van Beek et al., 2007;
3 Delsman et al., 2017; Yu et al., 2017). In addition, lower surface water levels in summer may
4 enhance peat oxidation which causes subsidence and nutrient losses. Throughout the world, an
5 increasing amount of catchments in low-lying and subsiding deltas will be managed in a similar
6 way as the Dutch polders. Therefore, the importance of understanding of the effect of different
7 water level regimes increases.

8

9 A combination of water and solute mass balancing has been proven effective for disentangling
10 the combined effect of (changes in) multiple sources and pathways of water and solutes in
11 similar catchments. For example, Hellman and Vermaat (2012) studied the relative impact of
12 agricultural nutrient inputs by constructing water and nutrient budgets for 13 polders in The
13 Netherlands. Kieckbusch and Schrautzer (2007) used water and nutrient balances to study the
14 effects of raising water levels in two catchments in Northern Germany.

15

16 The objective of this study was to assess the combined hydrological and hydrochemical effects
17 of changing the water level management from a fixed water level regime to a flexible, more
18 natural regime with minimum and maximum water levels. Ten study catchments were partly
19 hydrologically isolated; the surface water connections with the surroundings were closed apart
20 from controllable inlet and outlet weirs or pumps. The surface water level control was changed to
21 a flexible regime with predefined minimum and maximum water levels. The catchments were
22 intensively monitored in a dense multi-scale network of monitoring locations for surface water
23 and groundwater levels and solute concentrations in soil moisture, groundwater, and surface
24 water. The reference situation, the timing of the actual change in water level management, and
25 the practical water level management varied among the study catchments. Therefore, the
26 monitoring did not provide us with adequate data for comparing the situation before and after the
27 implementation of flexible water level management. We therefore used water and solute balance
28 modelling for catchment-scale assessments of changes in water and solute fluxes caused by the
29 major change in water level management.



1 **2 Methods**

2

3 **2.1 Study area**

4 The ten investigated polder catchments are situated in the Midwestern part of The Netherlands,
5 between the cities of Amsterdam in the northwest and Utrecht in the southeast (see Figure 1). All
6 polders are within the management area of water authority Amstel, Gooi en Vecht. All studied
7 polders are below mean sea level (MSL) due to centuries of artificial drainage, land subsidence,
8 and peat extraction. The current land surface elevations range between 5.0 m and 0.6m below
9 MSL (Table 1). The subsurface consists of Pleistocene sands, covered with a layer of Holocene
10 peat and clay of variable thickness (see Table1). The original thickness of this cover layer
11 reduces from circa 8 m in the northwest (Middelpolder, Ronde Hoep) to 1 m in the southeast
12 (Oostelijke Binnenpolder, Westbroekse Zodden) near the sandy ice-pushed ridge 'Utrechtse
13 Heuvelrug'. However, in the Groene Jonker and in the lakes of Botshol, Loenderveen Oost, and
14 Muyevelde, most of this Holocene peat has been mined in the 17th and 18th century to be used
15 as fuel.

16

17 The area has a semi-humid sea climate with an average yearly precipitation of 800mm and an
18 average yearly estimated evaporation of 550 mm, resulting in an average estimated yearly
19 recharge of 250 mm. The area is also fed by groundwater seepage from the Utrechtse
20 Heuvelrug. However, whether the individual polders receive seepage or infiltrate to the regional
21 groundwater system depends on their elevation relative to neighboring polders (e.g. Oude
22 Essink, 2001). The Ronde Hoep, for example, lies at 2.3 m below MSL, but still loses water
23 towards polder Groot-Mijdrecht at ca. 6 m below MSL bordering at the south.

24

25 All investigated polders are reclaimed wetland nature reserves. The Middelpolder, Ronde Hoep,
26 and parts Muyevelde are still in extensive agricultural use (grass/reed harvest, cattle grazing).
27 Loenderveen Oost is a fallback drinking water reservoir for the city of Amsterdam. Especially
28 Muyevelde is rich in recreational activities (boating, fishing, camping). The Groene Jonker and the
29 Westbroekse Zodden are partly accessible for hiking and bird watching. Most other polders are



1 less accessible or restricted bird reserves. The studied polders differed in their percentages
2 covered by surface water; the surface water proportion varies between 4% and 98% (Table 1).
3 The polders with low surface water proportions, such as Ronde Hoep, Nieuwe Keverdijkse
4 polder, and Middelpolder, have been in intensive agricultural use and still have their artificial
5 drainage network of ditches. The catchments of Botshol, Muyevelde, and Oostelijke Binnenpolder
6 partly consist of drained fields and partly of open surface water lakes. Open surface water
7 dominates the catchments of Loenderveen Oost and Westbroekse zoden. The original ditch
8 network of the Groene Jonker catchment was largely removed in a wetland restoration project.
9 For more detailed descriptions of all studied polder catchments, we refer to Borren et al. (2012a).

10

11 The surface water levels in all catchments are human controlled through large scale pumping
12 stations, adjustable weirs, and/or culverts. The water level management in all catchments was
13 changed between 2008 and 2013 from a fixed year-round water level or fixed winter and summer
14 level regime into a flexible regime in which the water levels can fluctuate freely between a
15 minimum and a maximum level. Note that the moment of implementation of the flexible water
16 level regime varies quite a lot among the study catchments due to the progress of legal
17 processes (paperwork, stakeholder consultation) and on the planning of the infrastructural work.
18 Table 1 gives for each catchment the fixed water levels of the conventional reference regime and
19 the minimum and maximum levels of the new flexible regime. Most polders have the same or
20 similar average water levels in the new flexible regime, except Ronde Hoep and Nieuwe
21 Keverdijkse polder north. The new water levels in these polders are higher in the new situation.

22

23 No reference water level regime is given for the Groene Jonker, because the landscape was
24 severely reshaped before introducing the flexible water level regime.

25

26 **2.2 Monitoring**

27 All studied polders were intensively monitored in 2011 and 2012 in a dense multi-scale network
28 of measurement locations for surface water and groundwater levels and solute concentrations.
29 This monitoring was mainly implemented for getting a more detailed overview of water and solute



1 fluxes in the study catchments. We did not implement the monitoring for comparing the situation
2 before and after the implementation of flexible water level management. This was not possible
3 within the timeframe of the monitoring period (2011-2012), because the timing of the actual
4 change in water level management varied among the study catchments (2008-2013).

5

6 Piezometers were installed to measure phreatic and deep groundwater heads and surface water
7 levels. The amounts of piezometers varied from 15 in the relatively small Oostelijke Binnenpolder
8 to 47 in Muyevelde. Part of the piezometers were distributed along the expected regional scale
9 hydrological gradients and the other part were placed in field scale (5-50m) transects from the
10 surface water ditches or lakes, through the riparian zones into the fields. Surface water and
11 groundwater levels were registered at hourly time intervals using pressure sensors (van Essen
12 instruments, Delft, the Netherlands). More details on the water level monitoring are described by
13 Borren et al. (2012a).

14

15 Concentrations of nutrients and other macro-parameters were monitored through monthly
16 sampling of surface water, groundwater, and soil moisture (Smolders et al., 2012). In each
17 polder, surface water was sampled at the inlet and outlet of the catchment and at several
18 locations distributed over the polder. Both shallow (phreatic) and deep groundwater from below
19 the Holocene top layer were sampled from piezometers using peristaltic pumps. Soil moisture
20 was sampled from porous ceramic suction cups that were installed close to the piezometers at
21 depths of 25, 50, and 100 cm below the land surface. The surface water, groundwater, and soil
22 moisture samples were analysed for pH (pH sensor), alkalinity (titration), and elements (e.g. Na,
23 K, NH₄, NO₃, PO₄, SO₄, Cl, Ca, Mg, S, P, Fe, Mn, Si, Zn) using Auto-Analyzers (AA) and
24 Inductively Coupled Plasma Atomic Emission Spectroscopy (ICP-OES). Details on the water
25 quality monitoring are reported in Smolders et al. (2012).

26

27 Water quality in the regional river water that was diverted into the polders during dry periods was
28 monitored monthly by the regional water authority Amstel, Gooi en Vecht. Amstel, Gooi en Vecht
29 also monitored the inlet and outlet water fluxes of the catchments, either by registering the



1 pumping hours or by monitoring water levels at weirs. Solute concentrations in precipitation water
2 were monitored by the National Institute for Public Health and the Environment (e.g. RIVM,
3 2005).

4

5 **2.3 Water and solute balance modelling**

6 To assess the changes in water and solute fluxes caused by flexible water level management,
7 water and solute balances were set up for all polder catchments. To facilitate future application
8 by the water authority staff, the water and solute balances were built in MS Excel. The water
9 storage in the catchment was divided in a groundwater, surface water, and field storage domain
10 (Figure 2). All exchange fluxes between these storages and between the catchment and its
11 surroundings were quantified at daily time steps in the water balance model. The only time-
12 variable inputs were the measured precipitation and evaporation time series from nearby weather
13 stations operated by the Royal Netherlands Meteorological Institute (KNMI). All other fluxes were
14 calculated following the model formulations in Supplement A. This approach allowed for uniform
15 scenario analyses for all catchments, assuming that the change in water level management did
16 not have a large influence on precipitation and evaporation.

17

18 The water level management was represented in the water balance by fixed year-round or fixed
19 summer and winter water levels for the reference conventional regime (Ref) and by minimum and
20 maximum water levels for the flexible water level management scenarios (Flex). Exceeding these
21 fixed or minimum and maximum levels induced water inlet or outlet to and from the polders (see
22 Supplement A). In case of a pumping station outlet, a maximum outlet discharge was imposed.

23

24 The solute balance modelling focused on the inputs and outputs of Cl, P-tot, N-tot, and SO₄ to
25 and from the surface water domain. Average measured concentrations were assigned to all input
26 fluxes towards the surface water (overland flow, groundwater exfiltration, seepage, precipitation,
27 water inlet; see Figure 2). The concentrations in overland flow were not directly measured, but
28 were assumed to be equal to the average measured soil moisture concentrations at 25 cm below
29 the surface. For groundwater exfiltration, separate concentrations were assigned to groundwater



1 flow from the topsoil (upper 30 cm) and groundwater flow from below the topsoil. The average
2 measured soil moisture concentrations at 25 cm below surface were assigned to groundwater
3 exfiltration from the topsoil. Groundwater exfiltration from below the topsoil was assigned the
4 average concentrations of samples from suction cups at 50 and 100 cm below the surface and
5 from the upper groundwater piezometers. The solute concentrations measured in samples from
6 the deep groundwater piezometers were assigned to the seepage flux. The concentrations
7 assigned to precipitation water came from the national monitoring network for precipitation
8 composition (RIVM, 2005), while concentrations in the diverted river water were provided by
9 water authority Amstel, Gooi en Vecht.

10

11 The water and solute balances were manually calibrated towards an optimal representation of
12 the dynamics in measured surface water levels, groundwater levels, Cl concentrations, and, if
13 available, discharge at the inlet and outlet of the polder catchments. The most sensitive
14 parameters were the surface water and land area percentages (A_i and A_w in Supplement A), the
15 infiltration or seepage fluxes (F_{si} or K_v , d , and h_d depending on the lower boundary definition),
16 and/or the exfiltrating fraction of groundwater towards surface water (f_i). As an example, Figure 3
17 shows the model performance on simulating the surface water levels, groundwater levels, and
18 chloride concentrations for the Middelpolder. The same information for the other polder
19 catchments is given in Supplement B. For more details on the water and solute balance models
20 we refer to Borren et al. (2012b).



1 **3** **Results**

2

3 The results of the water balance modelling for the reference and the flexible water level
4 management for all catchments are summarized in Table 2, 3, and 4. Many different output
5 results can be obtained from the water and solute balance models (Borren et al., 2012b). We
6 focus here on the differences in water fluxes and solute concentrations between the reference
7 and the flexible water level scenarios.

8

9 Table 2 gives the calculated average residence times and the exchange fluxes over 2003-2011
10 between the catchments and the regional surface water system via inlet and outlet and between
11 the catchments and the deep groundwater system via infiltration and/or seepage. Relative to the
12 reference scenario, flexible water level management results in longer surface water residence
13 times in all catchments. The residence times increase with 8-204%, with the largest relative
14 increases in the catchments with the lowest surface water area percentages (Ronde Hoep,
15 Nieuwe Keverdijkse polder north and south). This increase in residence time is related to the
16 decrease in inlet and outlet volumes, while the total water volume in the polders does not change
17 much. In polders with relatively small water volumes (Ronde Hoep, Nieuwe Keverdijkse polder
18 north and south), the decrease in inlet and outlet volumes have a relatively large impact on the
19 residence time.

20

21 The increase in surface water residence times is directly related to the decrease in both the
22 amount of inlet of diverted river water and the outlet to the regional water system (Table 2), while
23 the total water volume in the polders does not change much. The inlet and outlet volumes are
24 lower in all catchments in the flexible water level scenario. The Nieuwe Keverdijkse polder north
25 is an exception; the water inlet increases due to the higher water levels that are maintained in the
26 flexible water level scenario (see Table 1). As a consequence, this polder infiltrates more to the
27 deep groundwater system. At the same time, seepage from the deep groundwater system
28 decreases to 0 mm d⁻¹ in the flexible water level regime (see Table 2). While the inlet flux in
29 Nieuwe Keverdijkse polder north increases in the flexible water level scenario, the average



1 residence time still increases like in the other polders. The residence time still increases because
2 the total in- and outflux to and from the surface water domain are substantially reduced after
3 introducing the flexible water level regime.

4

5 The 2003-2011 average exchange fluxes between the surface water domain and the
6 groundwater domain (exfiltration and infiltration) and the field storage domain (overland flow and
7 inundation) for the reference and flexible water level regimes are presented in Table 3. In most
8 catchments, the exchange between surface water and groundwater reduces after the
9 introduction of flexible water level management or remains approximately the same. Exception is
10 again the Nieuwe Keverdijkse polder north, where infiltration increases due to the relatively high
11 surface water levels of the flexile water level scenario.

12

13 Overland flow and inundation are minor components in the long term average water balances of
14 most catchments. Due to the high groundwater and surface water levels in winter, the overland
15 flow contribution to surface water increases substantially in some catchments with low surface
16 water area percentages. Especially in Middelpolder and Ronde Hoep overland flow becomes and
17 important flow route in the flexible water level scenario.

18

19 To summarize the results from the solute balance modelling, Table 4 gives the 2003-2011
20 average surface water concentrations of Cl, N-tot, P-tot, and SO₄ for the reference and the
21 flexible water level management scenarios. The Cl concentrations decrease in most catchments,
22 while N-tot concentrations generally increase. For P-tot and SO₄ the results vary between the
23 polders. P-tot concentrations reduce in four catchments, but increase in three catchments. SO₄
24 concentrations reduce in six catchments, but increase in three catchments.

25



1 **4 Discussion**

2

3 This research aimed at assessing the combined hydrological and hydrochemical effects of
4 changing the water level management from a conventional fixed water level regime to a flexible,
5 more natural regime with minimum and maximum water levels for 10 study polder catchments.
6 Our water balance modelling results show significant changes in the water exchange fluxes
7 between the polder catchment and the regional water system and between the groundwater,
8 surface water, and field storage domains within the catchment. Overall, compared to the
9 reference water level regime, the flexible water level regime scenario showed increased surface
10 water residence times, reduced inlet and outlet fluxes, reduced groundwater-surface water
11 exchange, and in some catchments increased overland flow. The solute balance results shows a
12 reduction of chloride concentrations in all catchments, a reduction in P-tot and SO₄
13 concentrations in most catchments, and an increase in N-tot concentrations in most catchments.

14

15 **4.1 Hydrological effects**

16 The water balance modelling results help to understand the hydrological effects of the conversion
17 towards flexible water level management. Although the water balance models drastically simplify
18 the real situation, they performed well in simulating the surface water and groundwater levels
19 (see Figure 3 and Supplement B). Deviations between modelled and measured water levels can
20 have multiple causes. For example, the actual water level management is not always just based
21 on water levels and weather variations; in some cases human decisions such as changing the
22 water levels for maintenance or reed harvesting play a role. Some polders (e.g. Muyevelde,
23 Loenderveen) also receive excess water from neighboring polders, which may lead to higher
24 water levels and fluxes than expected. In addition, we modelled all polders without taking
25 account of the spatial differences within the polder. In most polders, the seepage or infiltration
26 rates vary due to higher or lower water levels in neighboring polders.

27

28 The results of the water balance modelling show that the introduction of flexible water level
29 management causes major changes in the water fluxes in all studied catchments. These



1 changes can be interpreted from the perspective of groundwater-surface water interactions
2 (Figure 4). In a conventional water level regime (Figure 4a), surface water levels in winter are
3 kept low, while groundwater levels are relatively high (Figure 4b). The relatively large difference
4 between the high groundwater levels and low surface water levels induces a relatively large
5 water flux from groundwater to surface water (Table 3). A relatively large outlet flux (Table 2) is
6 needed to compensate for the large groundwater input and maintain the low surface levels. In
7 summer, the high surface water levels and low groundwater levels (Figure 4c) induce a relatively
8 large infiltration flux from the surface water to the groundwater (Table 3). To compensate for this
9 infiltration flux and maintain the high surface water levels, a relatively large amount of inlet of
10 diverted river water is needed (Table 2). Due to the large in and out fluxes to and from the
11 surface water, residence times in the surface water are relatively low (Table 2). In the flexible
12 water level regime, the groundwater and surface water level covariate; both are relatively high in
13 winter (Figure 4e) and relatively low in summer (Figure 4f). The relatively small differences
14 between groundwater and surface water levels result in smaller exchange fluxes between
15 groundwater and surface water (Table 3). This also brings lower inlet fluxes in summer, lower
16 outlet fluxes in winter and longer residence times in surface water (Table 2).

17

18 As a result of the described changes in water fluxes, the contribution of different sources of
19 surface water changes as well. As a typical example, Figure 4 presents the surface water source
20 proportions for the Middelpolder catchment for the reference and the flexible water level regime
21 scenarios. The proportions are shown for a dry summer (2003), a wet summer (2007), the
22 average of the 2003-2011 summers, a dry winter (2004), a wet winter (2006), and the average of
23 the 2003-2011 winters. Figure 5 shows that the reduction in diverted river water inlet flux due to
24 the conversion to flexible water level management also results in a decreasing proportion of inlet
25 water in the catchment. The lower inlet water fractions are compensated by larger proportions of
26 direct precipitation water. This increase in the precipitation water proportion at the expense of the
27 inlet water proportion is a general consequence of the introduction of flexible water level
28 management in all polders.



1 Figure 5 also shows larger proportions of overland flow in the flexible water level management
2 scenario for the Middelpolder. In wet periods, surface water levels and groundwater levels are
3 higher in the flexible water level scenario compared to the reference scenario. In polders with
4 high surface water percentages, the discharged groundwater and overland flow proportions are
5 low in both scenarios. Inlet water and precipitation water are the dominant water sources in those
6 catchments.

7

8 Deep groundwater seepage comes up as a relatively small proportion after introducing the
9 flexible water level regime in Loenderveen Oost, Oostelijke Binnenpolder, and Westbroekse
10 Zodden. Up to 10% of the water volume in the Oostelijke Binnenpolder consists of deep
11 groundwater seepage in the dry summer of 2003. For Loenderveen Oost the deep groundwater
12 seepage proportion is up to 5% and for Westbroekse Zodden up to 1%. Although the proportions
13 are relatively low, the water quality effects of this new water source can be significant depending
14 on the chemistry of the seepage water (see also Yu et al., 2017).

15

16 **4.2 Water quality effects**

17 The changes in fluxes and the related changes in surface water source contributions help to
18 interpret the concentration changes due to conversion to flexible water level management
19 predicted by the solute balance model (Table 4). These solute balance results, however, should
20 be interpreted with care. The predicted concentrations are the combined contribution of all fluxes
21 to and from the surface water, and biochemical processes within the surface water were not
22 accounted for. This is not critical for the inert transport of Cl. Therefore, the Cl concentration
23 measurements can be used in the calibration as a check on the proportions of surface water from
24 different sources (Borren et al., 2012b; Hellman and Vermaat, 2012) (see for example Figure 3
25 for the Middelpolder). N-tot, P-tot, and SO₄ concentrations, however, are also influenced by
26 biochemical processes such as denitrification, adsorption, sedimentation, and plant uptake (e.g.
27 Vermaat et al, 2016; Van der Grift et al., 2014, 2016; Yu et al., 2017). The solute balance model
28 also assumes that the concentrations that were assigned to the fluxes towards the surface water
29 do not depend on the change in water level regime. However, the predicted longer residence



1 times in the groundwater domain may also cause a change in the concentrations of the
2 groundwater exfiltration flux towards surface water. In addition, the lower summer water levels in
3 a flexible water level regime may enhance peat oxidation, which could also result in higher
4 nutrient fluxes to surface water. The difference between the predicted concentrations for the
5 reference and the flexible water level scenarios should therefore be regarded as an indication for
6 the direction of change rather than an exact quantification.

7

8 The solute balance model predicted a general decrease in Cl concentrations and a general
9 increase in N-tot concentrations after the conversion to flexible water level management. The
10 general reduction in predicted Cl concentrations is related to the reduced proportions of inlet
11 water. Cl concentrations in the inlet water from the regional water system are relatively high,
12 partly due to the discharge from several deep polders with a large brackish groundwater seepage
13 flux. The Cl concentrations did not change much in the polders where the Cl concentrations in
14 inlet water are low (Loenderveen Oost, Nieuwe Keverdijkse polder).

15

16 Although the reduced inlet also reduced the input of N-tot into the polders, the solute balance
17 model predicted a general increase in N-concentrations. This is related to the longer residence
18 times and the larger relative impact of evaporation on the water balance, which amplifies the
19 effects other sources. For N-tot, atmospheric input is an important remaining input for all
20 catchments. In some polders, higher groundwater levels in winter increased mobilization of N-tot
21 from the enriched upper part of the soil. The Middelpolder was the only catchment for which the
22 combination of changes did not cause an increase in N-tot concentrations. The decrease in N-tot
23 loading from inlet water and groundwater discharge in The Middelpolder was large enough to
24 compensate for the larger relative impact of evaporation.

25

26 For P-tot and SO₄, the solute balance results varied between the polders. For most polders, the
27 solute balance model predicted a reduction in P-tot and SO₄ concentrations after the introduction
28 of flexible water level management. Similar to the reduction of Cl, this is mainly related to the
29 reduced proportion of inlet water with relatively high P-tot and SO₄ concentrations. In some



1 polders however, the effect of remaining P-tot or SO₄ sources is amplified and counteracts the
2 reduced inlet. The increase in P-tot concentrations in Loenderveen Oost and the Westbroekse
3 Zodden is caused by increased groundwater inputs induced by the lower water levels in summer.
4 The increased P-tot and SO₄ concentrations in the Middelpolder and the increased SO₄
5 concentrations in the Ronde Hoep are caused by an increased mobilization from the upper part
6 of the soil, which is still enriched due to past agricultural activities and peat oxidation. The higher
7 SO₄ concentrations in the Nieuwe Keverdijkse polder south are caused by the longer residence
8 times and the larger relative impact of evaporation. Like in the other polders, the SO₄ loading
9 towards the surface water is substantially reduced. However, this effect counteracted in the
10 Nieuwe Keverdijkse polder by the large increase in residence time and associated large relative
11 impact of evaporation.

12

13 The effects of converting to flexible water level management on solute concentrations are
14 governed by the change in solute loading via water inlet, the change in residence times, and the
15 change in flow route contributions. Although other studies for similar catchments did not
16 investigate the effects of a change from fixed water levels to flexible water level management,
17 their results partly support the findings from this study. For example, Hellman and Vermaat
18 (2014) found increased nitrogen concentrations due to reduced outlet and increased evaporation
19 as a result of climate change. This meets with our suggestion that the longer residence times and
20 larger relative impact of evaporation on the water balance may cause an increase in N-tot
21 concentrations. Based on field-scale 2-D water transport simulations for a similar polder, Van
22 Beek et al. (2007) reported larger groundwater fluxes through the upper subsurface towards the
23 surface water as a result of an increase in surface water levels. This supports our finding that the
24 higher water levels in winter in a flexible water level regime may increase overland flow and
25 solute mobilization from the enriched upper part of the soil.

26



1 **5 Conclusions**

2 Changing the water level management in polders from fixed levels to more a natural, flexible
3 water level regime induces relevant changes in water and solute fluxes to and from the
4 catchment and between the groundwater, surface water, and field surface domains within the
5 catchment. When supported by measurements of groundwater levels, surface water levels and
6 fluxes, and solute concentrations, water and solute balance modelling is a feasible approach to
7 study the combined catchment scale effects of changing water and solute fluxes after a major
8 change in water level management.

9

10 Parts of the modelled effects depend on catchment specific characteristics like elevation,
11 seepage flux, surface water proportion, and solute concentrations. However, we found a general
12 increase of surface water residence times, reduction of inlet and outlet fluxes, reduction of
13 groundwater-surface water exchange, and in some catchments increase of overland flow. The
14 water quality effects of introducing flexible water level management are governed by the change
15 in solute loading via water inlet, the change in residence times, and the change in flow route
16 contributions. For Cl, the reduced input via inlet water led to lower concentrations in all polders.
17 However, N-tot concentrations generally increased in spite of the reduced loads via inlet water.
18 The effects of other N-tot sources, such as atmospheric deposition and groundwater exfiltration,
19 were amplified by the longer residence times and the larger relative contribution of evaporation
20 on the water balance. The change in P-tot and SO₄ concentrations varied between the
21 catchments as the general effect of reduced loads via inlet water was counteracted in some
22 catchments by increased inputs from groundwater. The modelled P-tot concentrations increased
23 in polders with a low surface water area percentage and P enriched top soils due to the
24 increased loads from overland flow towards surface water, combined with longer residence times
25 and increased evaporation.

26

27 This study for the first time quantified the hydrological effects and identified potential water
28 quality risks of changing a controlled surface water level regime. Whereas an increasing amount



- 1 of catchments are being managed in a similar way as the Dutch polders, our approach and
- 2 findings are relevant for many low-lying and often subsiding delta's around the world.
- 3
- 4 **Acknowledgements.** The authors acknowledge the entire 'Flexpeil' project team for a fruitful
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2

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1 Tables

2

3 Table 1: Proportion of surface water area (relative to land surface area), land surface elevations,

4 approximate thickness of the Holocene peat and clay layer, reference conventional water

5 regimes and new flexible water level regimes for all polder catchments. Land surface elevations

6 and water levels in m relative to mean sea level (MSL).

Catchment	Proportion of surface water (%)	Land surface elevation (m)	Thickness of Holocene peat and clay layer (m)	Reference conventional water level regime	Flexible water level regime
Botshol	64	-2.4	4.5-8	Summer: -2.45, winter: -2.70	Min: -2.65, Max: -2.45
Groene Jonker	29	-5.0	Ca. 4	NA	Min: -5.60, Max: -5.10
Loenderveen Oost	98	-1.0	Ca. 1	Fixed: -1.15 +/- 0.05	Min: -1.30, Max: -1.00
Middelpolder	18	-2.15	Ca. 10	Summer: -2.40, winter: -2.45	Min: -2.55, Max: -2.25
Muyeveld	62	-0.6	0-5	Summer: -1.15, winter: -1.10	Min: -1.20, Max: -1.05
Nieuwe Keverdijkse polder north	8	-1.2	1.5-3	Fixed: -1.65 +/- 0.05	Min: -1.5, Max: -1.2
Nieuwe Keverdijkse polder south	8	-1.2	1.5-3	Fixed: -1.65 +/- 0.05	Min: -1.7, Max: -1.4
Oostelijke Binnenpolder	50	-0.9	0-1	Summer: -1.25, winter: -1.35	Min: -1.40, Max: -1.20
Ronde Hoep	4	-2.3	Ca. 8	Summer: -2.85, winter: -3.00	Min: -2.80, Max: -2.45
Westbroekse Zodden	90	-0.7	0-1	Summer: -1.00, winter: -1.05	Min: -1.1, Max: -0.95

7

8

9 Table 2: The 2003-2011 average modelled residence times in surface water and exchange fluxes

10 between the surface water domain and the regional water system (inlet and outlet) and the deep

11 groundwater system (seepage and infiltration) for all research catchments for the reference

12 conventional water level regime (Ref) and the flexible water level regime (Flex). Positive fluxes

13 are into the surface water domain, negative fluxes leave the surface water domain.

Catchment	Residence time (d)		Regional Exchange				Deep groundwater exchange			
	Ref	Flex	Inlet (mm d ⁻¹)		Outlet (mm d ⁻¹)		Seepage (mm d ⁻¹)		Infiltration (mm d ⁻¹)	
	Ref	Flex	Ref	Flex	Ref	Flex	Ref	Flex	Ref	Flex
Botshol	275	329	2.91	1.85	-0.95	0.00	0.00	0.00	-1.12	-1.08
Groene Jonker	NA	275	NA	0.00	NA	-0.08	NA	0.07	NA	-0.04
Loenderveen Oost	806	874	0.50	0.11	-0.18	0.00	0.00	0.14	-0.67	-0.61
Middelpolder	56	82	1.99	0.70	-4.33	-2.88	0.00	0.00	-0.39	-0.41
Muyeveld	435	524	1.54	0.87	-1.13	-0.44	0.00	0.00	-0.75	-0.76
Nieuwe Keverdijkse polder north	60	175	0.58	1.77	-8.21	-0.10	0.13	0.00	-0.08	-1.07
Nieuwe Keverdijkse polder south	29	78	0.00	0.00	-11.43	-3.99	0.59	0.19	0.00	-0.35
Oostelijke Binnenpolder	126	209	2.70	0.00	-1.47	-0.04	0.00	0.27	-0.58	-0.46
Ronde Hoep	25	76	2.54	1.36	-16.73	-6.45	0.00	0.00	-0.43	-0.53
Westbroekse Zodden	365	511	1.03	0.00	-1.39	0.00	0.00	0.03	-0.19	-0.31

14



- 1 Table 3: The 2003-2011 average modelled exchange fluxes between the surface water domain
2 and the groundwater domain (exfiltration and infiltration) and the field storage domain (overland
3 flow and inundation) for all research catchments for the reference conventional water level
4 regime (Ref) and the flexible water level regime (Flex). Positive fluxes are into the surface water
5 domain, negative fluxes leave the surface water domain.

Catchment	Groundwater exchange				Field storage exchange			
	Exfiltration (mm d ⁻¹)		Infiltration (mm d ⁻¹)		Overland flow (mm d ⁻¹)		Inundation (mm d ⁻¹)	
	Ref	Flex	Ref	Flex	Ref	Flex	Ref	Flex
Botshol	0.03	0.01	-1.28	-1.18	0.01	0.00	0.00	0.00
Groene Jonker	NA	0.17	NA	-0.35	NA	0.00	NA	0.00
Loenderveen Oost	0.02	0.02	-0.02	-0.02	0.00	0.00	0.00	0.00
Middelpolder	3.37	1.50	-1.23	-0.75	0.16	1.41	0.00	0.00
Muyeveld	0.40	0.34	-0.41	-0.35	0.00	0.00	0.00	0.00
Nieuwe Keverdijkse polder north	7.54	0.53	-0.38	-1.60	0.00	0.06	0.00	0.00
Nieuwe Keverdijkse polder south	10.55	4.07	-0.13	-0.37	0.00	0.04	0.00	0.00
Oostelijke Binnenpolder	0.40	0.44	-1.47	-0.62	0.00	0.00	0.00	0.00
Ronde Hoep	15.91	3.58	-1.68	-1.66	0.00	3.48	0.00	-0.19
Westbroekse Zodden	0.22	0.25	-0.03	-0.32	0.00	0.00	0.00	0.00

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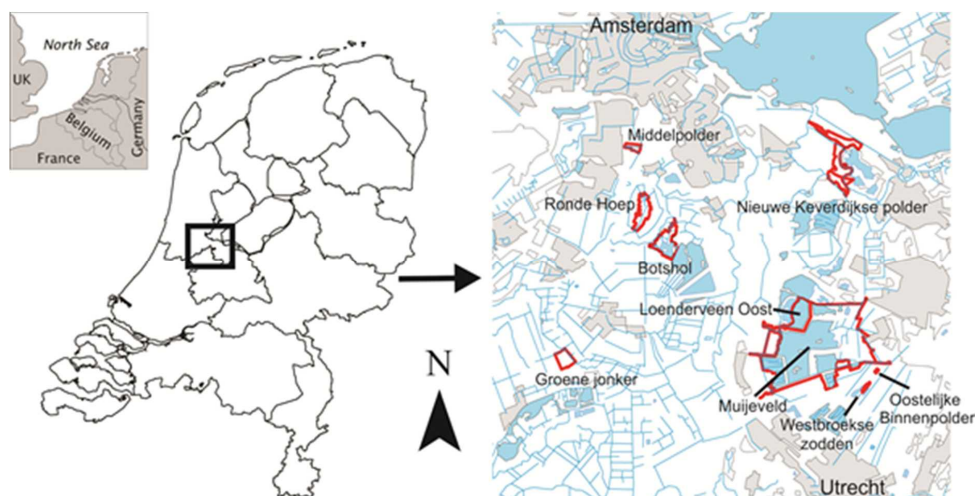
- 9 Table 4: The 2003-2011 average modelled concentrations in surface water of Chloride (Cl), Total
10 Phosphorus (P-tot), Total Nitrogen (N-tot), and Sulfate (SO₄) for all research catchments for the
11 reference conventional water level regime (Ref) and the flexible water level regime (Flex).

Catchment	Cl (mg L ⁻¹)		P-tot (mg L ⁻¹)		N-tot (mg L ⁻¹)		SO ₄ (mg L ⁻¹)	
	Ref	Flex	Ref	Flex	Ref	Flex	Ref	Flex
Botshol	836	804	0.06	0.06	3.0	3.4	39	34
Groene Jonker	NA	43	NA	0.14	NA	12.6	NA	277
Loenderveen Oost	35	35	0.07	0.14	4.2	5.0	18	12
Middelpolder	270	200	1.05	1.29	3.8	3.7	109	126
Muyeveld	61	55	0.07	0.07	3.2	3.7	30	25
Nieuwe Keverdijkse polder north	115	112	0.13	0.06	1.3	2.1	99	68
Nieuwe Keverdijkse polder south	132	138	0.17	0.15	1.4	1.8	114	137
Oostelijke Binnenpolder	38	13	0.05	0.02	2.3	4.1	21	8
Ronde Hoep	231	186	0.33	0.30	4.0	4.3	127	149
Westbroekse Zodden	34	13	0.12	0.22	3.3	5.9	21	12



1 **Figures**

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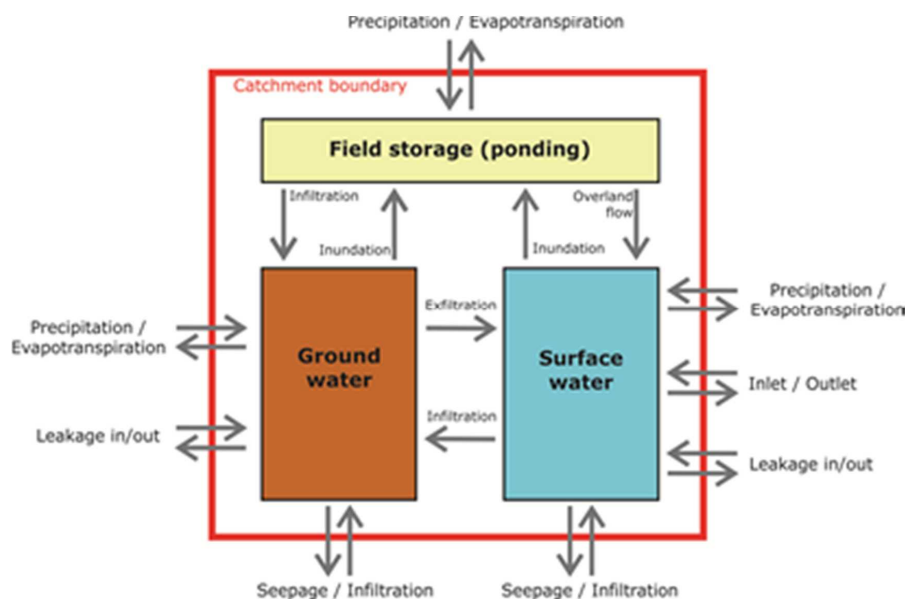
5 Figure 1: Study area and locations of the investigated polder catchments in The Netherlands

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between Amsterdam and Utrecht.

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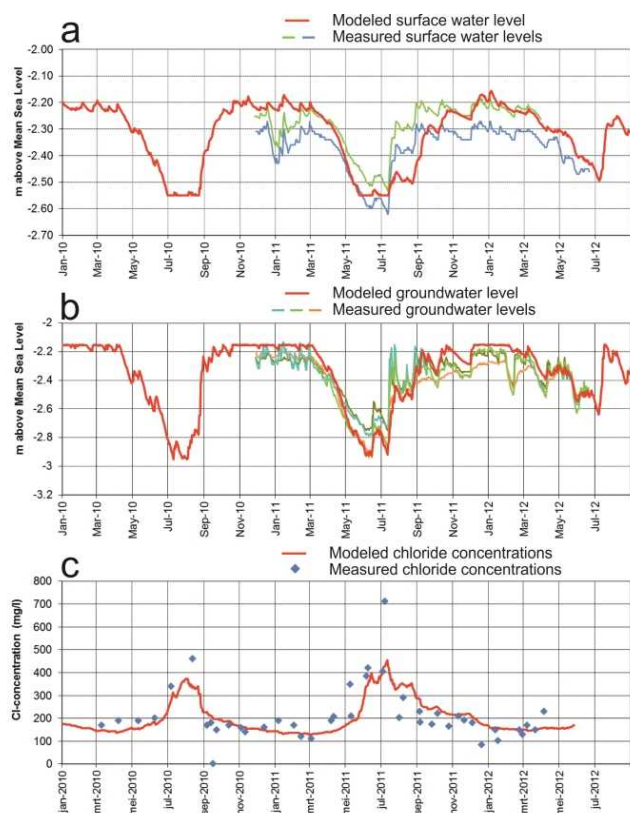
11 Figure 2: Visualisation of the three water and solute balance domains (field storage,

12

groundwater, and surface water) and all water fluxes to and from these domains. See

13

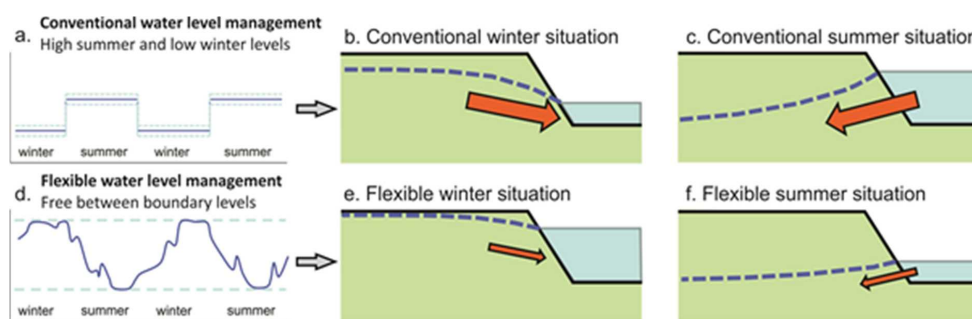
Supplement A for the water balance model formulations.



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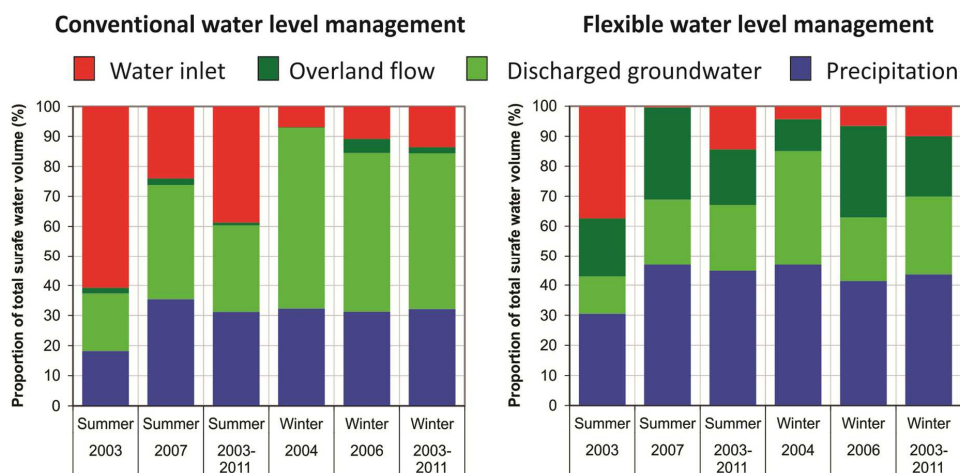
3 Figure 3: Model performance for the Middelpolder; modelled and measured surface water levels
4 (a), groundwater levels (b), and chloride concentrations (c). See Supplement B for other polders.

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Figure 4: Groundwater-surface water interaction in conventional and flexible water level management regimes. Fig. 4a visualises a conventional water level regime. Fig 4b and 4c show the large fluxes from groundwater to surface water in winter (b) and from surface water to groundwater in summer (c). Fig. 4d visualises a flexible water level regime. Fig 4e and 4f show the relatively small fluxes from groundwater to surface water in winter (4e) and from surface water to groundwater in summer (4f).



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Figure 5: Proportions of water sources of surface water in the Middelpolder catchment for the conventional water level management and the flexible water level management scenarios for a dry summer (2003), a wet summer (2007), the average of the 2003-2011 summers, a dry winter (2004), a wet winter (2006), and the average of the 2003-2011 winters.