1 Extrapolating regional probability of drying of headwater streams

2

using discrete observations and gauging networks

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6 Abstract

7 Headwater streams represent a substantial proportion of river systems and many of them have 8 intermittent flows due to their upstream position in the network. These intermittent rivers and 9 ephemeral streams have recently seen a marked increase in interest, especially to assess the impact 10 of drying on aquatic ecosystems. The objective of this paper is to quantify how discrete (in space and time) field observations of flow intermittence help to extrapolate over time the daily probability of 11 12 drying (defined at the regional scale). Two empirical models based on linear or logistic regressions 13 have been developed to predict the daily probability of intermittence at the regional scale across France. Explanatory variables were derived from available daily discharge and groundwater level data 14 15 of a dense gauging/piezometer network, and models were calibrated using discrete series of field 16 observations of flow intermittence. The robustness of the models was tested using (1) an 17 independent, dense regional data set of intermittence observations, (2) observations of the year 18 2017 excluded from the calibration. The resulting models were used to extrapolate the daily regional 19 probability of drying in France: (i) over the period 2011-2017 to identify the regions most affected by 20 flow intermittence; (ii) over the period 1989-2017, using a reduced input dataset, to analyze 21 temporal variability of flow intermittence at the national level. The two empirical regression models 22 performed equally well between 2011 and 2017. The accuracy of predictions depended on the 23 number of continuous gauging/piezometer stations and intermittence observations available to calibrate the regressions. Regions with the highest performance were located in sedimentary plains, 24

where the monitoring network was dense and where the regional probability of drying was the highest. Conversely, worst performances were obtained in mountainous regions. Finally, temporal projections (1989-2016) suggested highest probabilities of intermittence (> 35%) in 1989-1991, 2003 and 2005. A high density of intermittence observations improved the information provided by gauging stations and piezometers to extrapolate the temporal variability of intermittent rivers and ephemeral streams.

31 Keywords: Intermittent rivers, headwater streams, flow regime, discrete observations, regional scale

32 **1. Introduction**

Headwater streams represent a substantial proportion of river systems (Leopold et al., 1964; Nadeau and Rains, 2007; Benstead and Leigh, 2012). From an ecological point of view, headwater catchments are at the interface between terrestrial and aquatic ecosystems and they often harbour a unique biodiversity with a very high spatial turn-over (Meyer et al., 2007; Clarke et al., 2008; Finn et al., 2011). Their contribution to the functioning of hydrographic networks is essential: sediment flows, inputs of particulate organic matter and nutrients, refugia/colonization, sources for aquatic organisms (Meyer et al., 2007; Finn et al., 2011).

40 Headwater streams are generally naturally prone to flow intermittence, i.e. streams which stop 41 flowing or dry up at some point in time and space, mainly due to their upstream position in the 42 network and their high reactivity to natural or human disturbances (Benda et al., 2005; Datry et al., 43 2014b). These waterways which cease flow and/or dry are referred to as intermittent rivers and 44 ephemeral streams (IRES). The geographic extent of IRES is poorly documented due to mapping 45 limitations (digital elevation models, satellite images, aerial photos) and because of their size and 46 their location (Leopold et al., 1994; Nadeau and Rains, 2007; Benstead and Leigh, 2012; Fritz et al., 47 2013). However the proportion of IRES in hydrological networks can be very large: for example, they 48 represent 60% of the length of rivers in the United States (Nadeau and Rains, 2007) and are considered to represent probably more than 50% of the global hydrological network (Larned et al.,
2010; Datry et al., 2014b). Considering only gauging stations with continuous records may lead to
severe underestimation of their regional extent (Snelder et al., 2013; De Girolamo et al., 2015; Eng et
al., 2016).

53 Recently, IRESs have seen a marked increase in interest stimulated by the challenges of water 54 management facing the global change context (water scarcity issues, climate change impact, etc.) 55 (Acuña et al., 2014; Datry et al., 2016b). Studies have characterized the hydrological functioning of 56 IRES (Gallart et al., 2012; Costigan et al., 2016; Sarremejane et al., 2017) to assess the effects of flow 57 intermittence on aquatic ecosystems (Larned et al., 2010; Datry et al., 2016b; Leigh et al., 2016; Leigh 58 and Datry, 2017). IRES have been altered due to human actions (abstraction, hill dams, low-water 59 support, pollution, etc.) despite their high and unique biodiversity (Datry et al., 2014; Garcia et al., 2017a). In addition, some perennial streams are becoming intermittent due to global change, water 60 61 abstraction or river damming (Skoulikidis, 2009) and the extent of IRES may increase in the future 62 (Döll and Schmied, 2012; Jaeger et al., 2014; Pumo et al., 2016; Garcia et al., 2017b; De Girolamo et 63 al., 2017).

64 A better hydrological understanding of IRES is now essential and an improved management requires 65 knowing both the spatial extent and arrangement of IRES within the river network (Boulton, 2014; 66 Acuña et al., 2017). Efforts have been made to estimate the spatial distribution of IRES at the catchment scale (Skoulikidis et al., 2011; Datry et al., 2016a), at the regional scale (Gómez et al., 67 68 2005) and at the national scale (Snelder et al., 2013). In France, Snelder et al. (2013) suggested a 69 classification of IRES regimes and spatialized their distribution. Based on an analysis of the 70 continuous gauging network, they showed that the proportion of IRES accounted for 20 to 39% of the 71 hydrographic network. The accuracy of the obtained map is highly dependent on the density of the 72 flow monitoring network. The installation of additional gauging stations is expensive and headwaters 73 systems may be difficult to monitor due to active geomorphology processes or to difficult access.

74 As a promising tool to advance the mapping of IRES, citizen science creates opportunities to 75 overcome the lack of hydrological data and contributes to densify the flow state observation network (Turner and Richter, 2011; Buytaert et al., 2014; Datry et al., 2016b) and could be used for 76 77 hydrological model calibration (van Meerveld et al., 2017). In France, Datry et al. (2016a) used such 78 data to describe the spatiotemporal dynamics of aquatic and terrestrial habitats within five river 79 catchments located in the western part of France. They showed that processes resulting in flow 80 intermittence were complex at a fine scale and could vary substantially among nearby catchments. 81 However, these data were only available in a few catchments, limiting any attempt to map large-82 scale patterns of flow intermittence in river networks. Since this first attempt, new sources of 83 observational data have become available in France thanks to the ONDE network (Observatoire National des Etiages, https://onde.eaufrance.fr). This unique network in Europe provides frequent 84 85 discrete field observations (five inspections per year) of the flow intermittence across more than 3 86 300 sites throughout France and located mostly in headwater areas.

However discrete observations of intermittence do not provide any information on the persistence of dry conditions between two consecutive dates of observation. The rewetting-drying events could have significant impacts on communities whose survival is conditioned by the duration/frequency of drying. The duration of drying is of importance for ecologists, as one key driver for the composition and persistence of aquatic species (Vardakas et al., 2017; Kelso and Entrekin, 2018, Vadher et al., 2018). Temporal extrapolations of river flow regime are thus necessary to summarize the different facets of flow intermittence at various time scales, from daily to inter-annual.

The main objective of this paper is to use discrete (in space and time) field observations of flow intermittence to extrapolate over time the daily probability of drying (averaged at the regional scale). We first carried out a quantitative analysis of the ONDE network data in order to characterise the information that they contribute in comparison with the data resulting from the conventional hydrological monitoring. Then, we developed two empirical models based on linear or logistic

99 regressions to convert discontinuous series of flow intermittence observation from ONDE into 100 continuous daily probability of drying, defined at the regional scale across France. Explanatory 101 variables were derived from available continuous daily discharge and groundwater level data of a 102 dense gauging/piezometer network, and models were calibrated using the ONDE discrete 103 observations. The robustness of the models was tested using (1) an independent, dense regional data 104 set of intermittence observations and (2) observations of the year 2017 excluded from the 105 calibration. Finally, resulting models were used to extrapolate the regional probability of drying in 106 France: (i) over the period 2012-2017 to identify the regions most affected by flow intermittence; (ii) 107 over the period 1989-2017, using a reduced input dataset, to analyze temporal variability of flow intermittence at the national level. 108

109 **2. Material and Methods**

110 **2.1.** Study area

The study area is continental France and Corsica (550 000 km²). France is located in a temperate zone
characterized by a variety of climates due to the influences of the Atlantic Ocean, the Mediterranean
Sea and mountain areas.

We defined regions as combinations of "level-2 Hydro-EcoRegions" (HER2) and classes of 114 hydrological regimes (HR). Hydro-EcoRegion (HER) corresponds to a typology developed for river 115 management in accordance with the European Water Framework Directive. The Hydro-EcoRegion 116 117 classification includes 22 "level-1 Hydro-EcoRegions " (HER1) based on geology, topography and 118 climate, and considered as the primary determinants of the functioning of water ecosystems (Wasson et al., 2002). HER2 correspond to a finer classification accounting for stream size. HER2 have 119 120 a mean drainage area of 5 000 km² (between 100 and 27 000 km²). The hydrological regimes classes 121 (HR) were identified by reference to the work carried out by (Sauquet et al., 2008) where it was 122 possible to distinguish rainfall-fed regimes, transition and snowmelt-fed river flow regimes. Overall,

we used 280 regions (that is, HER2-HR combinations) with a mean drainage area of 1 400 km²
(between 4 and 20 000 km²).

125 **2.2. ONDE dataset discrete national flow-state observations**

126 The ONDE network was set up in 2012 by the French Biodiversity Agency (AFB, formerly ONEMA) 127 with the aim of constituting a perennial network recording summer low flow levels and used to 128 anticipate and manage water crisis during severe drought events (Nowak and Durozoi, 2012).

129 There are 3 300 ONDE sites distributed throughout France (Fig. 1). ONDE sites are located on 130 headwater streams with a Strahler order strictly less than 5 and balanced across HER2 regions to take 131 into account the representativeness of the hydrological contexts (Nowak and Durozoi, 2012). The ONDE network is stable over time. Observations are made monthly (around the 25th) by trained AFB 132 133 staff, between April and September, every year since 2012. One of the statuses is assigned at each 134 observation among "visible flow", "no visible flow" and "dried out". Here, we consider two 135 intermittency statuses: "Flowing" when there is visible flow across the channel ("visible flow") and 136 "Drying" when the channel is entirely devoid of surface water ("dried out") or when there is still 137 water in the river bed but without visible flow (disconnected pools, lentic systems) ("no visible flow"). The proportion of drying sites determined on the basis of the ONDE network for each HER2-138 HR combination is considered as a good estimate of the daily Regional Probability of Drying 139 140 (RPoD_{ONDE}) of streams with a Strahler order less than 5. Observed values of RPoD_{onde} are calculated as 141 follows:

142
$$RPoD_{ONDE}(d) = \frac{(Ndrying)_{HER2-HR}}{(Nflowing+Ndrying)_{HER2-HR}}$$
(1)

where *d* denotes the observation date of the ONDE network, Ndrying and Nflowing are the number
of drying and of flowing statuses observed at ONDE sites located in a same HER2-HR combination at
the observation date *d*, respectively.

146 Figure 2 illustrates the complementary nature of the ONDE network to the already existing French river flow monitoring network HYDRO (http://www.hydro.eaufrance.fr). The ONDE sites and a set of 147 1 600 gauging stations available in the HYDRO database have been projected on the river network 148 149 RHT (Theoretical Hydrographic Network; Pella et al., 2012) and the drainage area and the elevation 150 have been estimated. A large part of ONDE sites are located on small headwater streams with 70% of 151 the sites with a drainage area of less than 50 km² while most of the gauging stations record flows of 152 catchments of medium size (between 100 and 500 km²). Only four stations display a drainage area of 153 more than 1 000 km². The distributions of elevation of the two databases look similar. The ONDE 154 sites are mostly located on rivers with an elevation below 200 m (75% of sites). The ONDE sites are 155 sparse at high elevations (95 sites located above 1 000 m). This bias is likely due to access difficulties 156 in mountainous areas.

157 2.3. POC dataset: a denser regional dataset used for independent 158 validation

A spatially denser citizen science dataset of flow-state observations in western France (Poitou-159 160 Charente region) (http://atlas.observatoire-environnement.org) has been used as validation dataset 161 to test the robustness of our models calibrated with the ONDE dataset. The POC monitoring (2011-162 2013) covered more than 4 000 km of river length across 20 catchments. Each river was entirely 163 surveyed every 1st and 15th of each month between June and October, resulting in eight observations 164 per year. Four intermittency statuses were available in the POC dataset (Datry et al. 2016a) but to allow comparisons with the ONDE network, we pooled the two "Flowing" and "Low Flow" POC 165 statuses into a single "Flowing" status and the two "No flow" and "Dry" statuses into the "Drying" 166 167 status. This dataset is available as maps with flow states assigned to the inspected streams. Values of 168 RPoD at each POC observation date is calculated in the same way as RPoD_{ONDE}. Thus RPoD_{POC} is given 169 by the ratio between the number of drying statuses and the total number of observations at each 170 inspected streams located in a same HER2-HR.

171 **2.4.** Explanatory discharge dataset

172 Two discharge datasets (continuous daily time series) were used as explanatory variables of discrete 173 intermittence observations, with the objective of extrapolating the intermittence frequency over 174 time. The two datasets included time series of daily discharge extracted from the French River discharge monitoring network ("HYDRO database", http://www.hydro.eaufrance.fr/): (i) the 2011-175 176 **2017 dataset** with full records available between the 01/01/2011 and 31/06/2017; (ii) **the 1989-2017** dataset concerning a reduced number of gauging stations and providing daily discharges between 177 178 the 01/01/1989 and 31/06/2017. According to the hydrometric services in charge of the selected 179 gauging stations, high quality of measurements was ensured and observed discharges were not or only slightly altered by human actions. 180

The 2011-2017 dataset was composed of 1 600 gauging stations distributed across France. Each stream where a HYDRO gauging station is located has been defined as IRES or perennial. Several definitions of IRES can be found in the literature (Huxter and van Meerveld, 2012, Eng et al., 2016; Reynolds et al., 2015). In this study, we considered stations as intermittent when five consecutive days with discharge less than 1 liter per second has been observed during the period of record.

186 The 1989-2017 dataset consisted of 630 gauging stations selected with less than 5% of missing data 187 (continuous or not) during the period 1989-2017. This dataset has been thereafter used to estimate 188 the regional probability of drying before the creation of the ONDE network.

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2.5. Explanatory groundwater level dataset

Because groundwater resources influence stream intermittence, we used available time series of the daily groundwater level available in the ADES database (http://www.ades.eaufrance.fr) at sites identified as involved in groundwater/surface water exchanges (Brugeron et al., 2012). Similarly to the discharge data, two sets of groundwater level data with records available over the two periods 2011-2017 and 1989-2017 have been selected. The level of alteration of groundwater levels by water withdrawal is unknown because no information is available at this scale. 196 The 2011-2017 dataset was composed by 750 piezometers with daily groundwater level data with less than 5% of missing data (continuous or not). The selection of 1989-2017 dataset was not easy 197 198 because few groundwater level measurements were available in the database before 2000. For 199 example, only five piezometers met the tolerance limit on missing values considered for the 1989-200 2017 discharge dataset. In order to extend the dataset and because groundwater levels were less 201 variable than stream discharges, the proportion of permitted gaps was fixed to 20% between 1989 202 and 2017. This led us to select 150 piezometers. Thereafter, when the missing data period was less 203 than 10 days, groundwater levels were reconstructed by linear interpolation in order to reduce the 204 proportion of missing values to less than 5% for the 150 piezometers selected.

205 2.6. Statistical modeling of regional probability of drying

206 The parametric modeling strategy was based on 5 main steps (Fig. 3). The first step consisted in 207 selecting all ONDE sites, gauging stations and piezometers located in a same HER2-HR combination. 208 When the total number of gauging stations and piezometers was less than 5 for a HER2-HR 209 combination, we merged the HER2-HR combination with a neighboring one located in the same 210 HER1. This was done for 20 of the 280 regions. The second step consisted in calculating the RPoD_{ONDE} 211 for each observation date (5 per year) and for all selected ONDE sites. In a third step, a flow duration 212 curve was determined for each selected HYDRO gauging station. The average non-exceedance 213 frequency of the observed discharge at gauging stations was averaged for the date of observation (d) 214 at ONDE sites and the 5 days preceding the observation. The lag of six days accounted for the fact 215 that ONDE survey dates in a region could differ by 5 days, and accounted for the inertia of physical 216 processes (e.g. storage capacity). The same operation was carried out with selected piezometers. 217 Finally the hydrological conditions are described by the average (across stations) F of the non-218 exceedance frequencies of discharge (Fq) and groundwater levels (Fgw) with respect to the relative 219 proportions of gauging stations and piezometers:

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$$F(d) = \frac{\sum_{i=1}^{i=Nq} Fq_i + \sum_{j=1}^{j=Ngw} Fgw_j}{(Nq + Ngw)}$$
(2)

Where Fqi denotes the average non-exceedance frequency of discharge at the gauging station i calculated between d and d-5; Fgwj the average non-exceedance frequency of groundwater levels at the piezometer j calculated between d and d-5; Nq the number of gauging stations selected in a HER2-HR combination and Ngw the number of selected piezometers selected in the HER2-HR combination. The fourth step consisted in estimating the RPoD_{ONDE} as a function of F. Two types of regression were fitted for each HER2-HR combination across France:

a truncated logarithmic linear regression (LLR), with two parameters α_1 and β_1 :

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$$\operatorname{RPoD}_{LLR}(d) = \begin{cases} \min(1; \alpha_1 \times \ln(F(d)) + \beta_1) \text{ when } F < F0 \\ 0 \text{ when } F \ge F0 \end{cases}$$
(3)

F0 was fixed as the value of non-exceedance frequencies of discharge and groundwater levels at
which no more drying status was observed across the ONDE network (RPoD_{ONDE} = 0).

a logistic regression (LR), with two parameters α_2 and β_2 :

232
$$Logit(\operatorname{RPoD}_{LR}(d)) = ln\left(\frac{\operatorname{RPoD}_{LR}(d)}{1 - \operatorname{RPoD}_{LR}(d)}\right) = \alpha_2 \times F(d) + \beta_2 \tag{4}$$

LR is a multivariate analysis method well known for its relevance in binary classification issues (Lee,
2005). The RPoD_{LR} was then calculated as following Eq. 5:

235
$$\operatorname{RPoD}_{LR}(d) = \frac{\exp(\alpha_2 + \beta_2 F(d))}{1 + \exp(\alpha_2 + \beta_2 F(d))}$$
(5)

Models were calibrated against observations available during the same period, 2012-2016, leaving out the year 2017 for an independent validation test. However, for the continuous temporal extrapolations (one over 2011-2017, the other 1989-2017), two models were built with different piezometers and gauging stations selected as explanatory variables (see section 2.4 and 2.5). Thus there are two sets of regressions parameters specific to each dataset for both LLR and LR models leading to different prediction of RPoD. Finally, in a fifth step, a daily regional probability of drying (RPoD) could be predicted for each HER2HR combination with both models following analytical formulas (Eq. 3 and Eq. 5).

244 **2.7.** Model robustness: validation using independent data sets

We used (1) the POC independent data and (2) the 2017 ONDE year to test the robustness of the LLR and LR model to predict the intermittence frequency (1) in space and (2) over time. Note that when predicting on the POC datasets, a new model was calibrated using only ONDE sites located out of POC streams.

For both datasets (POC and ONDE 2017), the relative performance of the LLR and LR models was compared in multiple ways using both the 2011-2017 and the 1989-2017 datasets. The performance of each model was evaluated by the Nash-Sutcliffe efficiency criterion (NSE) (Nash and Sutcliffe, 1970):

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$$NSE = 1 - \frac{\sum_{i=1}^{N} (\text{RPoD}_{\text{ONDEi}} - \text{RPoD}_{\text{pri}})^{2}}{\sum_{i=1}^{N} (\text{RPoD}_{\text{ONDEi}} - \overline{\text{RPoD}_{\text{ONDEi}}})^{2}}$$
(6)

where $\text{RPoD}_{\text{ONDEi}}$ is the average proportion of drying statuses over the ONDE sites located in the HER2-HR combination at the ith observation date, RPoD_{pri} is the predicted regional probability of drying at the ith observation date, $\overline{\text{RPoD}_{\text{ONDEi}}}$ is the mean of $\text{RPoD}_{\text{ONDEi}}$ over the period and *N* is the total number of observations in the ONDE network for each HER2-HR combination.

258 **2.8. Model prediction**

Both models have been calibrated over the period 2012-2016 and were then applied in a 5th step to predict the daily RPoD in France (Fig. 3). The RPoD was firstly predicted over the period 2012-2016 in order to identify the most affected regions by flow intermittence using the 2011-2017 datasets. The second application concerned the extrapolation of RPoD in France over a longer period using the 1989-2017 dataset to analyze the temporal variability of flow intermittence at the national level. It should be noted that model predictions only concern streams with a Strahler order lower than 5 dueto the ONDE sites location.

3. Results

3.1.

Quantitative analysis

3.1.1. Inter-annual intermittence according to the raw discrete ONDE network

A total of 1 127 ONDE sites have recorded at least one drying event during the period 2012-2016 representing 35% of the 3 300 ONDE sites. From the ONDE database the probability of drying at the country scale was computed as the total number of drying statuses over France divided by the total number of ONDE observations available during statuses the same year (Fig. 4a). Between 2012 and 2016, the most critical year is 2012 with 15% of drying statuses followed by 2016 (14%) and 2015 (14%) (Fig. 4a). The years 2013 and 2014 are less affected with only 6% of drying statuses observed (Fig. 4a).

276 Drying events mainly occur between July and September but the evolution of the month's proportion 277 of drying can differ between years (Fig. 4b). In more detail, water levels in 2012 decrease in August 278 when the proportion of drying is 27% and the situation lasts until the end of September with 25% of 279 drying (Fig. 4b). In 2013, the proportion of drying is lower than in 2012 but follows the same pattern 280 with an increase at the end of July (3%) and reaching 9% in August and in September. In 2014, the 281 first peak of drying (5%) is reached early in June. Then, the proportion of drying decreases in July (3%) and increases slightly in August 4% and reaching 7% in September. In 2015, the critical period 282 283 occurs at the end of July with 19% of drying statuses and the proportion of drying decreases slightly 284 at the end of August (17%) until it reaches 9% in September. Finally, in 2016, the situation is 285 gradually deteriorates every month, reaching 20% of drying statuses in August, and 28% in 286 September.

287 Between 2012 and 2016, a proportion of drying higher than 50% is recorded on 93 ONDE sites and their spatial distribution is very patchy at the France scale (black and dark grey dots, Fig. 5a). There 288 289 are only 158 ONDE sites with at least one drying event every year and a variability of drying locations 290 can be observed across years. The south-east of France is heavily affected by rivers drying where the 291 proportion of drying can exceed 75% annually (black dots, Fig. 5b-5f). The north-western part of 292 France is less affected, although many ONDE sites show a proportion of drying observed above 50% 293 in 2014 and 2016 (Fig. 5d and 5f). Northeastern France is rather affected in 2012, 2014 and 2015 294 where several ONDE sites have more than 75% of drying statuses (Fig. 5b, 5d and 5e). The south-west 295 France is particularly affected in 2012 and 2015 (Fig. 5b and 5e).

296 3.1.2. Comparison of flow intermittence between the raw ONDE and HYDRO datasets297 The HYDRO dataset includes 90 gauging stations located on streams considered as IRES, which298 represents only 5.6% of the 1 600 gauging stations against 35% for ONDE sites. At the national scale,299 the number of IRES seems underrepresented in the south-western, central, northeastern part of300 France and Corsica in comparison with sites experiencing drying in the ONDE network (Fig. 6).

301 The number of gauging stations with at least one drying event (discharge < 1 l/s) observed between 302 May and September varies between 79 in 2012 and 47 in 2014 (Table 1). The lowest numbers of 303 gauging stations with drying events are observed in the years 2013 and 2014 while the highest 304 numbers are related to the years 2012, 2015 and 2016. This finding is consistent with the analysis of 305 the ONDE network (Fig. 5a, d). The frequency of drying, corresponding to the ratio between the 306 number of dry days and the total number of days between the 1st May and the 30th September (153 307 days), in contrast, is quite constant over the years (\sim 30%). The number of gauging stations with 308 drying event over more than 50% of the time varies little between wet years (14 in 2013) and dry 309 years (21 in 2015) unlike ONDE observations, suggesting a significant temporal variability in the 310 frequency of drying between dry and wet years (Fig. 5).

311 **3.2.** Validation of the predicted regional probability of drying

312 **3.2.1. Regression results**

313 LLR and LR models, calibrated over the period 2012-2016, perform well with the 2011-2017 dataset 314 with a mean NSE of 0.8 with LR model against 0.7 with LLR model (Fig. 7a and b). With the LR model, 315 50% of the HER2-HR combinations obtain a NSE greater than 0.8, representing a coverage of 65% of 316 the French territory, while 33% of HER2-HR combinations display a NSE higher than 0.8 (50% of 317 France coverage) with the LLR model. Regions with the highest performances are located in 318 sedimentary plains, in the south-east of France and in the Pyrenees Mountains. Conversely, the 319 worst performances are obtained in the mountainous regions of Alps as well as in the Massif Central. 320 In these regions the size of the HER2 is rather small and the number of ONDE sites, gauging stations 321 and piezometers per HER2-HR combinations are certainly too few to derive reliable relations. Despite 322 pooling, estimating RPoD remains impossible for 9 HER2-HR combinations (4.5% of France coverage) 323 because the number of ONDE sites, gauging stations and piezometers sites is insufficient (less than 5) 324 to perform the regression analysis.

The performance level is lower when the 1989-2017 dataset is used in models: the mean NSE with the LR and LLR models is 0.7 and 0.6, respectively (Fig. 7c and d).

The LR and LLR models lead to similar performance range. However, the LR model outperforms the LLR model in terms of number of HER2-HR combinations with NSE greater than 0.8 (Fig. 7c and d). The performance is sensitive to the dataset. As expected, the best results are obtained with the denser network. A decrease in NSE by more than 0.2 is identified for 5% of the French territory when the 1989-2017 dataset is used (black areas; Fig. 7e and f). The regions with the most degraded values of NSE are small HER2-HR combinations located in eastern France (Fig. 7e and f).

The decrease in performance is mainly due to the difference in number of gauging stations and piezometers between the two datasets (Fig. 8). The most degraded NSEs correspond to HER2-HR

combinations where the number of gauging stations and piezometers considered in regressions is
the most reduced, i.e. with a loss higher than 50% of stations (black and dark grey dots; Fig 8a and b).
However, the decrease in performance remains low (difference in NSE is below 0.1 for 75% and 64%
of HER2-HR combinations with LLR and LR model, respectively).

339

3.2.2. Comparison to the POC database

340 The observed proportion of drying RPoD_{POC} is rather well simulated by both LLR and LR models with 341 the 2011-2017 explanatory dataset (NSE > 0.7 except for the year 2011, Fig. 9). In addition, the 342 models are able to capture small fluctuations of RPoD_{POC} during the summer period. The best results 343 during the year 2011 are obtained with the LLR model (black curve; Fig. 9) and the LR model 344 overestimates RPoD_{POC} by 3% (dashed grey curve; Fig. 9). In 2012, the decline in water levels is more gradual than in 2011 and a marked peak is reached in September with 40% of RPoD_{POC} (Fig. 9). This 345 346 pattern is well reproduced by both models with a good fit to all observation points (Fig. 9). The year 347 2013 is less affected by drying occurrence and the maximum RPoD_{POC} does not exceed 20% (Fig. 9). 348 Curves of both models fit to observations well until the end of August. Note that the LR model is 349 slightly closer to the observations around the peak in September compared to the LLR model. 350 However the LR model overestimates the RPoD_{POC} at the end of September and in October.

When the 1989-2017 dataset is used as explanatory variables, the simulations of RPoD are weakly degraded with both models (Fig. 9d, e, f). However the simulated pattern is similar to the observed one. The LLR model outperforms the LR model during the three years of validation with the 1989-2017 dataset (black curve; Fig. 9d, e, f).

355 **3.2.3. Temporal patterns assessment of models between 2012 and 2017**

During the calibration period, the LLR and LR models tend to better simulate the RPoD during dry years 2012 and 2016 (NSE = 0.8 with LLR and LR models; Tab. 2) than during wet years (e.g. 2014 with NSE < 0.7). The NSEs are lower during the months of May and June when few drying events are observed while NSEs are much better during the driest months of August and September. 360 During the validation year of 2017, both models obtain a similar performance over the year 361 independent of datasets (NSE = 0.7).

362 Monthly NSEs in 2017 follow the same trend as monthly NSEs of the calibration period with lower 363 NSEs in May (NSEs < 0.4) and June (NSEs = 0.5) and higher NSEs in July, August and September (NSEs 364 = 0.6) with both models independent of datasets. Figure 10 shows the dispersion between predicted 365 RPoD and drying statuses observed at ONDE sites in the scatter plot during the validation year 2017 366 (Fig. 10a and 10b) in comparison with the year 2012 which obtains the better NSE during calibration period (Fig. 10c and 10d). The NSEs obtained in 2017 are 0.72 with the LLR model and 0.68 with the 367 368 LR model against 0.83 and 0.81 in 2012, respectively. The performance is slightly lower in 2017 but 369 remains acceptable with NSEs close to 0.7 and both models seem able to predict RPoD out of the 370 calibration period.

371

3.3. Application of regional models

372 **3.3.1.** Modeling of intermittencies severity between 2012 and 2016

373 Both models have been applied using the 2011-2017 dataset. Figure 11 displays the maximum 374 number of consecutive days (D_{RPoD>20%}) with RPoD higher than 20% simulated by both LLR and LR 375 models. The most affected regions are located in the south-east of France and in the sedimentary 376 plains which are consistent with the spatial pattern obtained from the ONDE observations (Fig. 5). 377 The most impacted year followed the same hierarchy: the year 2012 is the most critical year with 378 30% of France displaying $D_{\text{RPoD}>20\%}$ higher than 60 days followed by the year 2015 (20% of France with 379 $D_{RPoD>20\%} > 60$ days) and 2016 (15% of France with $D_{RPoD>20\%} > 60$ days) (Fig. 11). The years 2013 and 380 2014 are weakly affected with 5% and 6% of the France with $D_{RPoD>20\%}$ higher than 60 days, 381 respectively.

382 The LR model tends to simulate shorter periods of drying, particularly in HER2-HR combinations located in the South-East France in 2013 and 2014 (Fig. 11). However, there is an overall agreement 383 384 between RPoD simulated by both models in terms of spatial and temporal extent of dry streams.

385

3.3.2. Reconstitution of historical regional probability of drying

386 The trend temporal patterns of RPoD predicted by the two models, considering the 1989-2017 387 dataset, look similar between 1989 and 2016 and the simulated RPoD fit well to RPoD_{ONDE} (Fig. 12).

388 The proportion of drying is highly variable over the total simulation period, with alternating dry (1989 389 to 1991, 2003 to 2006, 2009 to 2012) and wet (1994 to 1995, 2000 to 2002; 2013 to 2014) phases. In 390 spite of interannual variability, peaks of RPoD occur regularly between August and September, 391 whether in dry years or wet years. This finding is consistent with the preeminence of rainfall fed river 392 flow regime with low flows in summer, in France.

393 The highest values of RPoDs (above 35% over France) are observed in 1989, 1990, 1991, 2003 and 394 2005 (black curve, Fig. 12a and b). The RPoDs simulated during these dry years are out of the range 395 of the observed values over the calibration period (2012-2016). Estimations are thus uncertain. 396 However, the high values of RPoD are consistent with observations reported in previous studies (e.g. 397 Larue and Giret, 2004; Snelder et al., 2013; Caillouet et al., 2017). Conversely, the years less affected 398 by drying are simulated in 1994, 2001 and 2014 with an average RPoD below 15% throughout the 399 year (black curve, Figs. 12a and b).

400 Results obtained with the LLR model are more contrasted in terms of extreme values than those 401 obtained with the LR model (Fig. 12b).

4. Discussion 402

ONDE network complementarity with conventional flow monitoring network 403

404 The analysis of the ONDE observations shows that the proportion of rivers undergoing drying is 405 significantly higher (35%) than that observed with the conventional monitoring (HYDRO database, 406 8%). This proportion although related to a short period of records 2012 and 2016 is consistent with 407 the percentage of 39% of river segments classified as intermittent by Snelder at al. (2013). This 408 analysis confirms the under-representation of IRES in the French HYDRO database, and probably 409 others in other countries (flow are often uncontrolled in IRES). Without gauging stations located on 410 headwaters, Snelder et al. (2013) were unable to predict IRES in eastern France (see Fig. 9, pp. 2694). 411 The high density of ONDE sites makes it possible to improve the detection of drying events and lead 412 to better understand the spatial distribution of IRES located at the upstream extent of the 413 hydrographic network. The ONDE network encompasses various hydrological conditions which 414 provides a more accurate assessment of inter-annual variability, differentiating between dry years 415 (2012, 2015 and 2016) and wet years (2013, 2014) with clearly few drying occurrences.

The validation of the LR and LLR models against the spatially dense POC database also demonstrates the spatial representativeness of the ONDE network. Thanks to the qualitative information provided and to models such as statistical models developed here, it is now possible to capture drying event at the regional scale.

420 The ONDE sites are located on small headwater streams which can be very reactive to external 421 disturbances (rainfall deficit, change in air temperature, increase in water withdrawals, etc.) and by 422 nature are more likely to be IRES. The gauging stations available in the HYDRO database are located 423 on larger streams and their hydrologic response to changes in external factors (environmental or 424 human) is slower and drying occurred with greater inertia under temperate climate. Their uneven 425 distribution across France does not allow to accurately characterize the inter-annual variability of 426 drying development. Overall, the ONDE network provides very complementary information to 427 conventional flow monitoring, leading to a better understanding of the processes of drying in upstream catchments. 428

The performance obtained with the LR and LLR models is slightly better with the 2011-2017 dataset 430 431 (mean NSE = 0.75) than those obtained with the 1989-2017 dataset (mean NSE > 0.65), whose 432 network is less dense. HER2-HR combinations are the most degraded where the number of 433 monitoring stations is the most decreased between the two datasets. The accuracy of the predictions 434 is dependent on the number of gauging stations, ONDE sites and piezometers available to calibrate 435 the regressions. Highest NSEs are obtained in western sedimentary plains and southeastern of France 436 where a significant number of streams have dryings regardless of years (Fig. 5). The dominant river 437 flow regime in these regions is mainly influenced by precipitation and the lowest water levels are 438 reached in August and September, which corresponds to the monitoring period of the ONDE 439 database. They benefit from a dense monitoring network (gauging stations, ONDE sites, 440 piezometers), which allows a better representation of the hydrological functioning of streams 441 located within the same HER2. Conversely, performance was poor in mountainous areas such as in 442 the Alps or the Massif Central (NSE < 0.4) where river flow regimes are diversified combining rainfall 443 and snowmelt influences. By construction, the area of HER2-HR combination in mountains is 444 reduced, which leads to a limited number of monitoring stations, certainly not sufficient to fit the 445 models. Moreover, the observation period for ONDE sites was limited between May and September 446 and dryings can be missed, particularly for streams influenced by snow or ice melting with potential 447 drying periods in winter. In regions potentially concerned by drying events out of the May-September 448 period, the actual ONDE monitoring strategy needs to be adapted to provide reliable temporal 449 observations and extrapolations of drying frequencies.

We have chosen to average the non-exceedance frequencies of flows and groundwater levels in order to increase the monitoring network. If models had been calibrated using only gauging stations, performance will have been globally similar, or slightly better, in some HER2-HR combinations (Fig. 13). Therefore, we could not validate the real gain of using groundwater level data in addition to

discharge data. This is certainly due to the dominant proportion of the gauging stations compared to 454 455 the piezometers. Indeed, in the 2011-2017 dataset, the proportion of gauging stations is greater than 456 75% for more than 70% of HER2-HR combinations whereas the proportion of piezometers exceeds 457 70% in only 5% of HER2-HR combinations. Groundwater level data thus have small weight in 458 regressions for this dataset. However, in the 1989-2017 dataset, the proportion of piezometer is 459 greater than 70% in more than 30% of HER2-HR combinations. The presence of piezometers 460 increases the density of the monitoring network in HER2-HR combinations with few available gauging 461 stations. Thanks to groundwater level data, RPoD can be predicted on more HER2-HR combinations.

462 Interest in reconstructing the dynamic regional probability of drying

463 Spatio-temporal simulation of the probability of drying is crucial for advancing our understanding of 464 IRES ecology and management. Some aquatic species can persist in a dry reach for a few days, weeks 465 or months, while some are highly sensitive to desiccation (Datry, 2012; Storey and Quinn, 2013; 466 Stubbington and Datry, 2013). Estimating the total duration of days with drying at the reach scale is 467 therefore needed to understand biological patterns in river networks (Kelso and Entrekin, 2018). To 468 our knowledge, no study has proposed to reconstruct daily flow states time series of headwater 469 streams at the country scale as France (> 500 000 km²) using discrete observations in time and space. 470 In the literature, studies at national scale remain focused on the detection and the mapping of IRES 471 because these rivers are historically poorly investigated and their proportion in existing hydrographic networks remains inaccurate or misunderstood (Nadeau and Rains, 2007; Snelder et al., 2013). 472 473 Recently, several studies proposed alternative methodologies in order to estimate metrics in 474 ungauged IRES (Gallart et al., 2016) or to predict daily streamflow in river basin experiencing flow 475 intermittence (De Girolamo et al., 2017b) but remain applicable at local scale.

This study provides a first regional approach to use discrete data obtained from regular observations. The average non-exceedance frequency is a global hydrological statistic that only captures the hydrological conditions at the regional scale in modelling the RPoD. For rainfall-driven river flow 479 regimes, the effect of rainfall events on flow intermittence at the HER2-HR scale is probably indirectly reflected by the daily discharge and groundwater levels used to calculate the average non-480 481 exceedance frequency. However, when more observation data are available, it is likely that including 482 more detailed descriptors of rainfall events and local geology could improve our approach. In France, 483 based on the 2011-2017 dataset, both models suggest highest values of RPoD along the Mediterranean coast ($D_{RPoD>20\%}$ > 100 days each year). Rivers in this region are subject to a 484 485 predominantly pluvial regime (Class 7; Sauquet et al., 2008), i.e. hot and dry summers follow by 486 intense rainfall events in autumn, leading to high flows in November (Skoulikidis et al., 2017b). The 487 catchments in this region are small and particularly reactive to environmental changes, making them 488 highly sensitive to flow intermittence. Rivers located in the sedimentary plain in western France are 489 also very impacted by flow intermittence. The regime is also influenced by precipitation and for the 490 basins subject to intense agriculture significant water abstractions during summer in this region 491 reduce water availability in rivers and in aquifers which are no longer able to support the low water 492 levels, leading to increased flow intermittence. Regarding alteration issues in our datasets, we do not 493 have access to the exact location and the volumes of water withdrawal for irrigation purposes. 494 However, due to their upstream location, water availability is expected to be low, which may limit 495 potential withdrawals and as a consequence flow alteration at ONDE sites. The alteration of 496 groundwater levels is unknown because no information is available. However, in sedimentary plains 497 where agricultural crops dominate the landscape, we are not sure that no human action affects low 498 flows. It is important to note that the responses of biological communities to artificial flow 499 intermittence is still poorly understood compared to natural IRES (Datry et al., 2014b, Skoulikidis et 500 al., 2017a).

501 Validity of historical regional probability of drying during severe low-flow period

502 The second application aimed at reconstructing historical RPoD over the period 1989-2016. Both 503 models suggest highest values of mean RPoD (> 35%) in 1989-1991, 2003 and 2005. During these dry

504 years, predicted values of RPoD result from extrapolation but are consistent with published studies 505 (Mérillon and Chaperon, 1990, Moreau, 2004). For example, Mérillon (1992) estimated that for the 506 whole of France, 11 000 km of rivers were dried at the end of summers of 1989 and 1990. Caillouet 507 et al (2016) found that the low-flow event observed in 1989-1990 was particularly severe in terms of 508 duration and affected 95% of France. Snelder et al. (2013) showed from 628 gauging stations that the 509 years 1989-1991, 2003 and 2005 had witnessed particularly high values of duration and frequency of 510 drying events. They found that regions with the highest probability of drying were located along the 511 Mediterranean and Atlantic coasts, which is consistent with ONDE observations and with our results.

512 Both models suggest the same sequence of dry and wet years. However the application of the LLR 513 model lead to less contrasted RPoD than the LR model (Fig. 12).

514 To illustrate these differences, the RPoD has been simulated by both models with an extreme F of 1% 515 (Fig. 14). The RPoD_{LLR} is significantly higher and exceeds 80% in 30% of the study area against only 5% 516 of the area with the RPoD_{LR}. On the other hand, models simulate low RPoD in HER2-HR combinations 517 where the RPoD_{ONDE} is very low between 2012-2016, even when F was 1% because this situation 518 never occurred during the calibration period (Fig. 14). The logistic function of the LR model takes an 519 S-shape which induced a decrease of the slope of the curve toward extreme values observed during 520 the calibration period (2012-2016). The truncated logarithmic function of the LLR model is not 521 bounded and RPoD can reach 100% during extreme low flow events by extrapolation. Since the ONDE network monitoring period does not include a period with drought as severe as in the 1990s, it 522 523 is not currently possible to assess the relative performance of the two models. Refining extrapolated 524 values requires additional information on headwater collected during more severe droughts than 525 those observed during the last five years and then gives support to the pursuit of the ONDE network.

526 **5. Conclusion**

527 This paper investigates the spatial and temporal dynamics of the regional probability of drying (RPoD) 528 of headwater streams by taking benefit from qualitative and discontinuous data provided by the 529 ONDE network. Two models based on linear or logistic regressions have been developed and 530 succeeded to reconstruct the temporal dynamics of RPoD. They are based on a strong relationship 531 between the non-exceedance frequencies of discharges and groundwater levels as a function of the 532 proportion of drying statuses observed at ONDE sites per HER2-HR combination. LLR and LR models 533 show similar performance and perform well between 2011 and 2017. The accuracy of predictions is 534 dependent on the number of gauging stations, ONDE sites and piezometers available to calibrate the 535 regressions. Regions with the highest performance are located in the sedimentary plains, where the 536 monitoring network is dense and where the RPoD is the highest. Conversely, the worst performances 537 are obtained in the mountainous regions. Finally, both models have been used to reconstruct 538 historical RPoD between 1989 and 2016 and suggest highest values of mean RPoD (> 35%) in 1989-539 1991, 2003 and 2005. This is consistent with other published studies but the high density of ONDE 540 sites makes it possible to improve the detection of drying events and lead to better capturing of the 541 spatial distribution of IRES located at the upstream extent of the hydrographic network. Moreover, 542 the duration of drying is of importance for ecologists and the prediction of a daily RPoD provides one 543 key driver for the composition and persistence of aquatic species.

From a methodological point of view, our method relating discrete drying observation obtained by citizen science networks to continuous daily gauging data seems robust across the highly diverse (climate and topography) regions of France, and provides good predictions in an independent region excluded from the calibration process (PoC). These two results suggest a potential application of our approach in other countries. Citizen science creates opportunities to overcome the lack of hydrological data and contributes to densify the flow state observation network (Turner and Richter, 2011; Buytaert et al., 2014) and remains less expensive than the installation of additional gauging

stations to survey flow intermittence. The next step will be to use this regional approach to simulate the RPoD in future periods by taking into account effects of climate change through predicted discharge and groundwater level data. This would allow quantification of the evolution of the probability of drying between the current period and the different climate projections provided by the latest IPCC Report (IPCC 2014a, 2014b) and would assist decision makers in defining protocols for restoring flows with appropriate measures to preserve aquatic ecosystems (Woelfle-Erskine, 2017).

557 Secondly, further work is needed to develop an approach capable of reconstructing the drying 558 dynamics locally by differentiating each stream. Our approach remains spatially valid to estimate 559 RPoDs at the scale of HER2-HR combinations but does not allow characterizing the variability of 560 drying occurrence between nearby streams within these regions. From a methodological point of 561 view, statistical tools such as neural networks (Breiman, 2001) have shown good ability to assess 562 both the occurrence and extent of perennial and temporary segments (González-Ferreras and 563 Barquín, 2017) and could be investigated as an alternative method to reconstruct locally the 564 temporal variability of drying.

565 **6. Acknowledgment**

The authors wish to thank A. van Loon and C. Sefton for their valuable comments, suggestions and positive feedback on the manuscript. The research project was partly funded by the French Agency for Biodiversity (AFB, formerly ONEMA). This study is based upon works from COST Action CA15113 (SMIRES, Science and Management of Intermittent Rivers and Ephemeral Streams, www.smires.eu), supported by COST (European Cooperation in Science and Technology).

571 **7. References**

Acuña, V., Datry, T., Marshall, J., Barceló, D., Dahm, C. N., Ginebreda, A., McGregor, G., Sabater, S.,
Tockner, K. and Palmer, M. A.: Why should we care about temporary waterways?, Science,
343(6175), 1080–1081, 2014.

Acuña, V., Hunter, M. and Ruhí, A.: Managing temporary streams and rivers as unique rather than
second-class ecosystems, Biological Conservation, 211, 12–19, doi:10.1016/j.biocon.2016.12.025,
2017.

Benda, L., Hassan, M. A., Church, M. and May, C. L.: Geomorphology Of Steepland Headwaters: The
Transition From Hillslopes To Channels1, Journal of the American Water Resources Association,
41(4), 835, 2005.

581 Benstead, J. P. and Leigh, D. S.: An expanded role for river networks, Nature Geoscience, 5(10), 678– 582 679, 2012.

583 Boulton, A. J.: Conservation of ephemeral streams and their ecosystem services: what are we 584 missing?: Editorial, Aquatic Conservation: Marine and Freshwater Ecosystems, 24(6), 733–738, 585 doi:10.1002/aqc.2537, 2014.

586 Breiman, L.: Random forests, Machine learning, 45(1), 5–32, 2001.

587 Brugeron, A., Allier, D., Klinka, T.: Approche exploratoire des liens entre référentiels hydrogéologique 588 et hydrographique : Premières identifications des piézomètres potentiellement représentatifs d'une 589 relation nappe/rivière et contribution à leur valorisation. Rapport final BRGM/RP-61047-FR. 241 p, 590 2012.

Buytaert, W., Zulkafli, Z., Grainger, S., Acosta, L., Alemie, T. C., Bastiaensen, J., De Bi??vre, B., Bhusal,
J., Clark, J., Dewulf, A., Foggin, M., Hannah, D. M., Hergarten, C., Isaeva, A., Karpouzoglou, T.,
Pandeya, B., Paudel, D., Sharma, K., Steenhuis, T., Tilahun, S., Van Hecken, G. and Zhumanova, M.:
Citizen science in hydrology and water resources: opportunities for knowledge generation,
ecosystem service management, and sustainable development, Frontiers in Earth Science, 2,
doi:10.3389/feart.2014.00026, 2014.

597 Clarke, A., Mac Nally, R., Bond, N. and Lake, P. S.: Macroinvertebrate diversity in headwater streams:
598 a review, Freshwater Biology, 53(9), 1707–1721, doi:10.1111/j.1365-2427.2008.02041.x, 2008.

599 Costigan, K. H., Jaeger, K. L., Goss, C. W., Fritz, K. M. and Goebel, P. C.: Understanding controls on 600 flow permanence in intermittent rivers to aid ecological research: integrating meteorology, geology 601 and land cover: Integrating Science to Understand Flow Intermittence, Ecohydrology, 9(7), 1141– 602 1153, doi:10.1002/eco.1712, 2016.

Datry, T.: Benthic and hyporheic invertebrate assemblages along a flow intermittence gradient:
effects of duration of dry events: River drying and temporary river invertebrates, Freshwater Biology,
57(3), 563–574, doi:10.1111/j.1365-2427.2011.02725.x, 2012.

Datry, T., Larned, S. T., Fritz, K. M., Bogan, M. T., Wood, P. J., Meyer, E. I. and Santos, A. N.: Broadscale patterns of invertebrate richness and community composition in temporary rivers: effects of
flow intermittence, Ecography, 37(1), 94–104, doi:10.1111/j.1600-0587.2013.00287.x, 2014a.

Datry, T., Larned, S. T. and Tockner, K.: Intermittent Rivers: A Challenge for Freshwater Ecology,
BioScience, 64(3), 229–235, doi:10.1093/biosci/bit027, 2014b.

- Datry, T., Pella, H., Leigh, C., Bonada, N. and Hugueny, B.: A landscape approach to advance intermittent river ecology, Freshwater Biology, 61(8), 1200–1213, doi:10.1111/fwb.12645, 2016a.
- Datry, T., Fritz, K. and Leigh, C.: Challenges, developments and perspectives in intermittent river ecology, Freshwater Biology, 61(8), 1171–1180, doi:10.1111/fwb.12789, 2016b.
- De Girolamo, A. M., Lo Porto, A., Pappagallo, G., Tzoraki, O. and Gallart, F.: The Hydrological Status
 Concept: Application at a Temporary River (Candelaro, Italy): EVALUATING HYDROLOGICAL STATUS
 IN TEMPORARY RIVERS, River Research and Applications, 31(7), 892–903, doi:10.1002/rra.2786,
 2015.
- De Girolamo, A. M., Bouraoui, F., Buffagni, A., Pappagallo, G. and Lo Porto, A.: Hydrology under
 climate change in a temporary river system: Potential impact on water balance and flow regime,
 River Research and Applications, doi:10.1002/rra.3165, 2017a.
- De Girolamo, A. M., Barca, E., Pappagallo, G. and Lo Porto, A.: Simulating ecologically relevant
 hydrological indicators in a temporary river system, Agricultural Water Management, 180, 194–204,
 doi:10.1016/j.agwat.2016.05.034, 2017b.
- Döll, P. and Schmied, H. M.: How is the impact of climate change on river flow regimes related to the
 impact on mean annual runoff? A global-scale analysis, Environmental Research Letters, 7(1),
 014037, doi:10.1088/1748-9326/7/1/014037, 2012.
- Eng, K., Wolock, D. M. and Dettinger, M. D.: Sensitivity of Intermittent Streams to Climate Variations
 in the USA: Sensitivity of Intermittent Streams, River Research and Applications, 32(5), 885–895,
 doi:10.1002/rra.2939, 2016.
- Finn, D. S., Bonada, N., M?rria, C. and Hughes, J. M.: Small but mighty: headwaters are vital to stream
 network biodiversity at two levels of organization, Journal of the North American Benthological
 Society, 30(4), 963–980, doi:10.1899/11-012.1, 2011.
- Fritz, K. M., Hagenbuch, E., D'Amico, E., Reif, M., Wigington, P. J., Leibowitz, S. G., Comeleo, R. L.,
 Ebersole, J. L. and Nadeau, T.-L.: Comparing the Extent and Permanence of Headwater Streams From
 Two Field Surveys to Values From Hydrographic Databases and Maps, JAWRA Journal of the
 American Water Resources Association, 49(4), 867–882, doi:10.1111/jawr.12040, 2013.
- Gallart, F., Prat, N., García-Roger, E. M., Latron, J., Rieradevall, M., Llorens, P., Barberá, G. G., Brito,
 D., De Girolamo, A. M., Lo Porto, A., Buffagni, A., Erba, S., Neves, R., Nikolaidis, N. P., Perrin, J. L.,
 Querner, E. P., Quiñonero, J. M., Tournoud, M. G., Tzoraki, O., Skoulikidis, N., Gómez, R., SánchezMontoya, M. M. and Froebrich, J.: A novel approach to analysing the regimes of temporary streams
 in relation to their controls on the composition and structure of aquatic biota, Hydrology and Earth
 System Sciences, 16(9), 3165–3182, doi:10.5194/hess-16-3165-2012, 2012.
- 644 Gallart, F., Llorens, P., Latron, J., Cid, N., Rieradevall, M. and Prat, N.: Validating alternative 645 methodologies to estimate the regime of temporary rivers when flow data are unavailable, Science 646 of The Total Environment, 565, 1001–1010, doi:10.1016/j.scitotenv.2016.05.116, 2016.
- Garcia, C., Gibbins, C. N., Pardo, I. and Batalla, R. J.: Long term flow change threatens invertebrate
 diversity in temporary streams: Evidence from an island, Science of The Total Environment, 580,
 1453–1459, doi:10.1016/j.scitotenv.2016.12.119, 2017a.

- Garcia, C., Amengual, A., Homar, V. and Zamora, A.: Losing water in temporary streams on a
 Mediterranean island: Effects of climate and land-cover changes, Global and Planetary Change, 148,
 139–152, doi:10.1016/j.gloplacha.2016.11.010, 2017b.
- 653 Gómez, R., Hurtado, I., Suárez, M. L. and Vidal-Abarca, M. R.: Ramblas in south-east Spain: 654 threatened and valuable ecosystems, Aquatic Conservation 15, 387–402, doi:10.1002/aqc.680, 2005.
- 655 González-Ferreras, A. M. and Barquín, J.: Mapping the temporary and perennial character of whole 656 river networks: MAPPING FLOW PERMANENCE IN RIVER NETWORK, Water Resources Research, 657 doi:10.1002/2017WR020390, 2017.
- Huxter, E. H. H. and (Ilja) van Meerveld, H. J.: Intermittent and Perennial Streamflow Regime
 Characteristics in the Okanagan, Canadian Water Resources Journal / Revue canadienne des
 ressources hydriques, 37(4), 391–414, doi:10.4296/cwrj2012-910, 2012.
- Jaeger, K. L., Olden, J. D. and Pelland, N. A.: Climate change poised to threaten hydrologic
 connectivity and endemic fishes in dryland streams, Proceedings of the National Academy of
 Sciences, 111(38), 13894–13899, doi:10.1073/pnas.1320890111, 2014.
- Kelso, J. E. and Entrekin, S. A.: Intermittent and perennial macroinvertebrate communities had similar
 richness but differed in species trait composition depending on flow duration, Hydrobiologia, 807(1),
 189–206, doi:10.1007/s10750-017-3393-y, 2018.
- Larned, S. T., Datry, T., Arscott, D. B. and Tockner, K.: Emerging concepts in temporary-river ecology,
 Freshwater Biology, 55(4), 717–738, doi:10.1111/j.1365-2427.2009.02322.x, 2010.
- Lee, S.: Application of logistic regression model and its validation for landslide susceptibility mapping
 using GIS and remote sensing data, International Journal of Remote Sensing, 26(7), 1477–1491,
 doi:10.1080/01431160412331331012, 2005.
- Leigh, C. and Datry, T.: Drying as a primary hydrological determinant of biodiversity in river systems:
 a broad-scale analysis, Ecography, 40(4), 487–499, doi:10.1111/ecog.02230, 2017.
- Leigh, C., Boulton, A. J., Courtwright, J. L., Fritz, K., May, C. L., Walker, R. H. and Datry, T.: Ecological
 research and management of intermittent rivers: an historical review and future directions,
 Freshwater Biology, 61(8), 1181–1199, doi:10.1111/fwb.12646, 2016.
- 677 Leopold, L. B.: A View of the River. Harvard University Press, Cambridge, Massachusetts, USA, 1994.
- Leopold, L. B., Wolman, M. G. and Miller, J. P.: Fluvial Processes in Geomorphology. Dover
 Publications, New York, USA, 1964.
- 680 Meyer, J. L., Strayer, D. L., Wallace, J. B., Eggert, S. L., Helfman, G. S. and Leonard, N. E.: The 681 Contribution of Headwater Streams to Biodiversity in River Networks1: The Contribution of 682 Headwater Streams to Biodiversity in River Networks, JAWRA Journal of the American Water 683 Resources Association, 43(1), 86–103, doi:10.1111/j.1752-1688.2007.00008.x, 2007.
- Nadeau, T.-L. and Rains, M. C.: Hydrological Connectivity Between Headwater Streams and
 Downstream Waters: How Science Can Inform Policy1: Hydrological Connectivity Between
 Headwater Streams and Downstream Waters: How Science Can Inform Policy, JAWRA Journal of the
 American Water Resources Association, 43(1), 118–133, doi:10.1111/j.1752-1688.2007.00010.x,
 2007.

- Nash, J. E. and Sutcliffe, J. V.: River flow forecasting through conceptual models part I A discussion
 of principles, Journal of Hydrology, 10(3), 282–290, doi:10.1016/0022-1694(70)90255-6, 1970.
- 691 Nowak, C. and Durozoi, B.: Observatoire National Des Etiages, Note technique, ONEMA., 2012.

Pella, H., Lejot, J., Lamouroux, N. and Snelder, T.: Le réseau hydrographique théorique (RHT) français
et ses attributs environnementaux, Géomorphologie: relief, processus, environnement, 18(3), 317–
336, 2012.

Pumo, D., Caracciolo, D., Viola, F. and Noto, L. V.: Climate change effects on the hydrological regime
of small non-perennial river basins, Science of The Total Environment, 542, 76–92,
doi:10.1016/j.scitotenv.2015.10.109, 2016.

Reynolds, L. V., Shafroth, P. B. and LeRoy Poff, N.: Modeled intermittency risk for small streams in the
Upper Colorado River Basin under climate change, Journal of Hydrology, 523, 768–780,
doi:10.1016/j.jhydrol.2015.02.025, 2015.

Sarremejane, R., Cañedo-Argüelles, M., Prat, N., Mykrä, H., Muotka, T. and Bonada, N.: Do
metacommunities vary through time? Intermittent rivers as model systems, Journal of Biogeography,
44(12), 2752–2763, doi:10.1111/jbi.13077, 2017.

Sauquet, E., Gottschalk, L. and Krasovskaia, I.: Estimating mean monthly runoff at ungauged
locations: an application to France, Hydrology Research, 39(5–6), 403, doi:10.2166/nh.2008.331,
2008.

707Skoulikidis, N. T.: The environmental state of rivers in the Balkans—A review within the DPSIR708framework, Science of The Total Environment, 407(8), 2501–2516,709doi:10.1016/j.scitotenv.2009.01.026, 2009.

Skoulikidis, N. T., Vardakas, L., Karaouzas, I., Economou, A. N., Dimitriou, E. and Zogaris, S.: Assessing
water stress in Mediterranean lotic systems: insights from an artificially intermittent river in Greece,
Aquatic Sciences, 73(4), 581–597, doi:10.1007/s00027-011-0228-1, 2011.

Skoulikidis, N. T., Vardakas, L., Amaxidis, Y. and Michalopoulos, P.: Biogeochemical processes
controlling aquatic quality during drying and rewetting events in a Mediterranean non-perennial river
reach, Science of The Total Environment, 575, 378–389, doi:10.1016/j.scitotenv.2016.10.015, 2017a.

716 Skoulikidis, N. T., Sabater, S., Datry, T., Morais, M. M., Buffagni, A., Dörflinger, G., Zogaris, S., del Mar 717 Sánchez-Montoya, M., Bonada, N., Kalogianni, E., Rosado, J., Vardakas, L., De Girolamo, A. M. and 718 Tockner, K.: Non-perennial Mediterranean rivers in Europe: Status, pressures, and challenges for 719 research management, Science of The Total Environment, 1-18, and 577, 720 doi:10.1016/j.scitotenv.2016.10.147, 2017b.

Snelder, T. H., Datry, T., Lamouroux, N., Larned, S. T., Sauquet, E., Pella, H. and Catalogne, C.:
Regionalization of patterns of flow intermittence from gauging station records, Hydrology and Earth
System Sciences, 17(7), 2685–2699, doi:10.5194/hess-17-2685-2013, 2013.

Storey, R. G. and Quinn, J. M.: Survival of aquatic invertebrates in dry bed sediments of intermittent
streams: temperature tolerances and implications for riparian management, Freshwater Science,
32(1), 250–266, doi:10.1899/12-008.1, 2013.

Stubbington, R. and Datry, T.: The macroinvertebrate seedbank promotes community persistence in
temporary rivers across climate zones, Freshwater Biology, 58(6), 1202–1220,
doi:10.1111/fwb.12121, 2013.

Turner, D. S. and Richter, H. E.: Wet/Dry Mapping: Using Citizen Scientists to Monitor the Extent of
Perennial Surface Flow in Dryland Regions, Environmental Management, 47(3), 497–505,
doi:10.1007/s00267-010-9607-y, 2011.

Vadher, A. N., Millett, J., Stubbington, R. and Wood, P. J.: Drying duration and stream characteristics
influence macroinvertebrate survivorship within the sediments of a temporary channel and exposed
gravel bars of a connected perennial stream, Hydrobiologia, doi:10.1007/s10750-018-3544-9, 2018.

van Meerveld, H. J. I., Vis, M. J. P. and Seibert, J.: Information content of stream level class data for
hydrological model calibration, Hydrology and Earth System Sciences, 21(9), 4895–4905,
doi:10.5194/hess-21-4895-2017, 2017.

Vardakas, L., Kalogianni, E., Economou, A. N., Koutsikos, N. and Skoulikidis, N. T.: Mass mortalities
and population recovery of an endemic fish assemblage in an intermittent river reach during drying
and rewetting, Fundamental and Applied Limnology / Archiv für Hydrobiologie,
doi:10.1127/fal/2017/1056, 2017.

Wasson, J.-G., Chandesris, A., Pella, H. and Blanc, L.: Typology and reference conditions for surface
water bodies in France: the hydro-ecoregion approach, TemaNord, 566, 37–41, 2002.

- Woelfle-Erskine, C.: Collaborative Approaches to Flow Restoration in Intermittent Salmon-Bearing
 Streams: Salmon Creek, CA, USA, Water, 9(3), 217, doi:10.3390/w9030217, 2017.
- 747

	Stations with at least one	Stations with drying >	Frequency of
	drying event	50%	discharge < 1 l/s
2012	79	19	32.7
2013	47	14	37.9
2014	54	15	32.9
2015	76	21	31.1
2016	71	19	28.6

Table 1. Annual statistics on flow intermittence calculated on HYDRO gauging stations between the
 750 1st May and the 30th September

		2011-2017 dataset					1989-2017 dataset						
		Calibration				Valid.	Calibration				Valid.		
		2012	2013	2014	2015	2016	2017	2012	2013	2014	2015	2016	2017
	May	0.2	0.0	0.5	0.5	0.6	0.4	0.2	0.0	0.3	0.0	0.7	0.2
	June	0.6	0.3	0.8	0.5	0.8	0.5	0.6	0.3	0.5	0.3	0.8	0.5
LLR	July	0.7	0.5	0.6	0.6	0.8	0.7	0.7	0.5	0.5	0.4	0.8	0.6
model	August	0.8	0.6	0.7	0.7	0.8	0.6	0.7	0.5	0.5	0.5	0.8	0.6
	Sept.	0.7	0.8	0.6	0.6	0.7	0.6	0.6	0.7	0.5	0.5	0.6	0.6
	May - Sept	0.8	0.8	0.7	0.7	0.8	0.7	0.8	0.7	0.5	0.6	0.8	0.7
	May	0.2	0.0	0.5	0.1	0.6	0.3	0.3	0.0	0.3	0.0	0.7	0.2
	June	0.6	0.5	0.8	0.5	0.8	0.4	0.6	0.4	0.5	0.3	0.7	0.4
LR	July	0.7	0.6	0.5	0.6	0.8	0.6	0.7	0.4	0.5	0.4	0.8	0.6
model	August	0.7	0.6	0.7	0.6	0.7	0.6	0.6	0.4	0.5	0.4	0.7	0.5
	Sept.	0.6	0.8	0.6	0.7	0.7	0.6	0.5	0.6	0.4	0.5	0.6	0.6
	May - Sept	0.8	0.8	0.7	0.7	0.8	0.7	0.8	0.7	0.5	0.6	0.8	0.7



Table 2. NSE criteria obtained between 2012 and 2017 with the LLR and LR models calibrated over
the period 2012-2016.







Figure 1. Location of the 3 300 ONDE sites and partition into HER2.



Figure 2. Distribution of the 3 300 ONDE sites and of the 1 600 gauging stations available in the
 HYDRO database against: (a) drainage area and (b) elevation.



Figure 3. Strategy of parametric modeling (step 1-4) developed to predict (step 5) the regional
 probability of drying (RPoD) by HER2-HR combination in France.



Figure 4. (a) Distribution of yearly proportion of drying observed with the ONDE network with the
 total yearly number of ONDE observations written in brackets and (b) distribution of proportions of
 drying per year and per month.



Figure 5. Distribution of the percentages of drying observed at ONDE sites for the years: (a) 20122016, (b) 2012, (c) 2013, (d) 2014, (e) 2015 and (f) 2016.





Figure 6. Map of ONDE sites and HYDRO gauging stations having at least one drying.



Figure 7. Map of Nash-Sutcliffe criteria (NSE) obtained for each HER2-HR combination between 2012
 and 2016 with the 2011-2017 and 1989-2017 datasets according to: (a) and (c) a log-linear regression
 (LLR) model; (b) and (d) a logistic regression (LR) model. NSE differences between the 2011-2017
 dataset and the 1989-2017 dataset are represented for: (e) LLR model and (f) LR model.



Figure 8. NSE calculated for each HER2-HR combination between 2012 and 2016 with the 1989-2017
 dataset as a function of NSE calculated with 2011-2017 dataset with respectively: (a) the LLR model
 and (b) the LR model. The color of dots represents the proportion of gauging station and piezometers
 lost between the 2011-2017 database and the 1989-2017 database: losses < 50% (white); losses
 between 50% and 75% (grey); losses > 75% (black).



Figure 9. Comparison between observed proportion of drying RPoD_{POC} and RPoD predicted by the LLR
 and LR models with the 2011-2017 dataset in: (a) 2011, (b) 2012 (c) 2013 and with the 1989-2017
 dataset in: (d) 2011, (e) 2012 (f) 2013.



Figure 10. Scatter plot of the predicted RPoD (x axis) and drying observed at ONDE sites (y axis) in
 2017 and 2012 simulated with the 2011-2017 dataset by: (a) and (c) the LLR model and (b) and (d)
 the LR model.







Figure 13. Comparison of NSE obtained with regression including only discharge variable as a
 function of NSE obtained with including discharge and groundwater level variables in the 2011-2017
 dataset with: (a) LLR model and (b) LR model.



Figure 14. Regional probability of drying simulated with F = 1% predicted with: (a) the LLR model and
(b) the LR model.