Reply to Anne Van Loon

General comments

Firstly, the authors need to explain why a regional assessment of headwater drying is needed. What is the benefit of Figure 11 over Figure 5? The patterns of drying are the same, so Figure 5 would be sufficient to indicate hotspots of drying within France and temporal variability in drying.

We modified lines 88-94 to better explain that one of our objective is a temporal extrapolation of the daily drying probability (eg Fig 11) in regions, based on discrete observations (~5/years, raw data in Fig 5).

Secondly, the paper is focused on France. This in itself is not a problem, since the methodology and results are interesting and useful beyond France, but the author fail to put their findings in a broader perspective in the discussion. Literature on IRES research from outside France should be discussed and the authors should clarify what is new and interesting about this work from an international perspective. On p.19 I.443-452, the authors mention how their results are consistent with previous studies, which is great, but they should additionally point out what their study adds. If this is not done, the study would be better placed in a Journal like Journal of Hydrology – Regional Studies.

The authors agree with this remark. We modified the text (lines 465-481 and lines 544-551) to better explain the international relevance of our results.

Thirdly, I would like the authors to help the reader more in understanding the methodology. Figure 3 is helpful, but in the manuscript it is not always clear which data was used for what. Especially when explaining the equations on page 9 and 10, the authors could be clearer on which dataset was used, which time period. Also in the Results section it should be clarified when they are referring to calibration results, validation with POD data, or validation with the year 2017. For example, the first paragraph of Section 3.2.3 is quite confusing, because it discusses the performance of the models in the calibration period, which was already discussed in Section 3.2.1. Table 2 should be explained better; how is it different or similar to the information presented in Figures 7&8? Also, in the first paragraph of Section 3.3.2 the authors state that "the simulated RPoD fit well to RPoDONDE" (I.349), but wasn't that already discussed in Section 3.2.1 (Figure 7&8)?

We modified the section 2.6 to better explain our methodology. The table 2 has been revised.

We modified the Section 3.2.3 to focus on the annual performance of each model in the revised manuscript.

Fourthly, it is unclear whether natural and/or human-influenced sites are selected in this study. In Section 2.4, the authors mention that the "observed discharges were not or only slightly altered by human actions" (p.7 l.164), but they do not specify whether the other datasets, i.e. groundwater levels, ONDE and POD observations, are near-natural too. This is important, as the authors mention in the discussion, "the basins are subject of intense agriculture with important water withdrawals during summer. Abstractions greatly reduce the water availability in rivers and in aquifers which are no longer able to support the low water levels and lead to increased flow intermittence. The responses of biological communities to artificial flow intermittence is still poorly understood compared to natural IRES." (p.19 l.435-439) If near-natural and human-influenced data are mixed in the predictions, it will be very difficult to understand the reasons for the regional patterns in drying and the statements about the highest drying occurring in sedimentary plains due to the low elevation gradient and dependence on rainfall might be flawed.

And finally, it is unclear why two statistical models are used throughout the paper. If they are equally suitable from a theoretical perspective, two (or more) models could be used for testing, but then the best model should be used to simulate the final results.

Specific comments:

The regional probability of drying needs to be explained. In Section 2.6 the authors only mention that RPoD is calculated, but they never explain how this variable is calculated exactly.

The weighted average of the non-exceedance frequencies (F) needs to be explained better. According to the Discussion section discharge and groundwater levels are combined (I.411-412), but this is not explained clearly enough in the Methods section (I.202-203). How are these non-exceedance frequencies of groundwater and discharge averaged since they have such different shapes and ranges (see Figure 3). And what do the authors mean with "with respect to the relative proportions of gauging stations and piezometers" (I.203-204)?

We provided the details to better explain F in the section 2.6.

The authors conclude that "both models seem able to predict RPoD out of the calibration period" (l. 330-331), but do a NSE of 0.4 and 0.5 warrant such a statement?

This section (section 3.2.3) has been modified and the revised manuscript presents NSEs for the 2017 validation year. Table 2 has been modified and presents these additional results. Figure 10 has been

modified and shows the dispersion between predicted RPoD and drying observed at ONDE sites in
the scatter plot during the validation year 2017 (Fig. 10a and 10b) in comparison with the year 2012
which obtains the better NSE during calibration period (Fig. 10c and 10d).

A significant part of the Conclusion section discusses future work. Is that relevant for this manuscript? I would suggest leaving those paragraphs out as they distract from the main message of this paper.
The authors have shortened this part of the conclusion.

Textual comments:

All your corrections/suggestions have been taken into account. We also took into account the remark about the concept of RPoD which will be better detailed and we will present the equation to compute the values of RPoD_{ONDE} (Eq. 1; Page 6, L140-145). This formula can also be applied to derive the values of RPoD_{POC} (Page 7, L168-170).

Reply to Catherine Sefton

Specific comments

Given the small number of observations at each site, the claim that the ONDE dataset offers more accurate assessment of inter-annual variability than the gauging station network (L376-377) needs further justification. Conversely, the claim that the dataset makes it possible to capture drying events at the regional scale (L381-382) would benefit from stressing the monitoring of both upstream and downstream drying – uniquely each with national extent – in your approach.

The presentation of summer-only status data as "% drying" needs qualification, as it suggests assumptions about the status in the rest of the year. In particular, clarification would be helpful in line 242, when the context implies it means the number of sites with at least one drying each year (as in line 241), rather than the % of all observations at all sites (as in Fig 4).

The technique of constructing mean non-exceedance frequency from river flows and groundwater level is attractive and robust. However, its limitation in this regional approach of failing to capture the effect of local rainfall should be commented upon, especially given the dominance of rainfall-driven intermittency stated in section 3.3.2.

This is discussed in the revised paper (Lines 475-481).

The frequency of drying from gauging station data needs to be defined (line 272). Context suggests it is flow permanence (dry days or dry months per time period), but frequency in intermittent rivers and ephemeral streams can also mean dry spells per time period, and it also needs to be clear and justified whether it's calculated from daily means or monthly means.

We revised this section (Lines 304-306).

In section 3.2.1., the difference in performance between the two explanatory hydrological datasets is attributed to the difference in the number of gauging stations and piezometers. The pattern in Figure 8 is not as clear as the text suggests, and it would be good to comment also on the assumption of stationarity and how it might vary between HER2-HR combinations. Similarly, historical reconstructions make assumptions about stationarity that need to be acknowledged.

We revised the description of the applications (Lines 237-244).

The conclusion is a good summary of the results but would benefit from contextual comment, both with respect to the stated objective of this paper and more broadly on the contribution being made to the field.

We modified the conclusion in the revised paper in order to highlight the contribution of our study (Lines 544-551).

Textual and Figure corrections:

Figure 3, step 1: It is unclear why HR1, 4 and 6 are shown as types of monitoring site, when section 2.1 has defined them as types of hydrological regime.

We clarified this aspect on a revised Figure 3.

Figure 10: This would benefit from additional plots for a year that has good NSE, as the text is comparing years as well as model performance.

We added additional plots of the year 2012 which obtain the better NSE during the calibration period and we added observations *vs.* predictions of the full year 2017 in a revised Figure 10. The section 3.2.3 and the Table 2 have been revised in order to present these new results.

Extrapolating regional probability of drying of headwater streams

using discrete observations and gauging networks

- 3 Aurélien BEAUFORT¹, Nicolas LAMOUROUX², Hervé PELLA², Thibault DATRY² and Eric SAUQUET¹
- ¹Irstea, UR RiverLy, Hydrology and Hydraulics Research Unit, centre de Lyon-Villeurbanne, 5 rue de la
- 5 Doua CS 20244, 69625 Villeurbanne, France
- 6 ²Irstea, UR RiverLy, Laboratory Dynam, 69625 Villeurbanne, France

Abstract

1

2

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

Headwater streams represent a substantial proportion of river systems and many of them have intermittent flows due to their upstream position in the network. These intermittent rivers and ephemeral streams have recently seen a marked increase in interest, especially to assess the impact of drying on aquatic ecosystems. The objective of this paper is to quantify how discrete (in space and time) field observations of flow intermittence help to extrapolate over time the daily probability of drying (defined at the regional scale). Two empirical models based on linear or logistic regressions have been developed to predict the daily probability of intermittence at the regional scale across France. Explanatory variables were derived from available daily discharge and groundwater level data of a dense gauging/piezometer network, and models were calibrated using discrete series of field observations of flow intermittence. The robustness of the models was tested using (1) an independent, dense regional data set of intermittence observations, (2) observations of the year 2017 excluded from the calibration. The resulting models were used to extrapolate the daily regional probability of drying in France: (i) over the period 2011-2017 to identify the regions most affected by flow intermittence; (ii) over the period 1989-2017, using a reduced input dataset, to analyze temporal variability of flow intermittence at the national level. The two empirical regression models performed equally well between 2011 and 2017. The accuracy of predictions depended on the number of continuous gauging/piezometer stations and intermittence observations available to calibrate the regressions. Regions with the highest performance were located in sedimentary plains, where the monitoring network was dense and where the regional probability of drying was the highest. Conversely, worst performances were obtained in mountainous regions. Finally, temporal projections (1989-2016) suggested highest probabilities of intermittence (> 35%) in 1989-1991, 2003 and 2005. A high density of intermittence observations improved the information provided by gauging stations and piezometers to extrapolate the temporal variability of intermittent rivers and ephemeral streams.

Keywords: Intermittent rivers, headwater streams, flow regime, discrete observations, regional scale

1. Introduction

Headwater streams represent a substantial proportion of river systems (Leopold et al., 1964; Nadeau and Rains, 2007; Benstead and Leigh, 2012). From an ecological point of view, headwater catchments are at the interface between terrestrial and aquatic ecosystems and they often harbour a unique biodiversity with a very high spatial turn-over (Meyer et al., 2007; Clarke et al., 2008; Finn et al., 2011). Their contribution to the functioning of hydrographic networks is essential: sediment flows, inputs of particulate organic matter and nutrients, refugia/colonization, sources for aquatic organisms (Meyer et al., 2007; Finn et al., 2011).

Headwater streams are generally naturally prone to flow intermittence, i.e. streams which stop flowing or dry up at some point in time and space, mainly due to their upstream position in the network and their high reactivity to natural or human disturbances (Benda et al., 2005; Datry et al., 2014b). These waterways which cease flow and/or dry are referred to as intermittent rivers and ephemeral streams (IRES). The geographic extent of IRES is poorly documented due to mapping limitations (digital elevation models, satellite images, aerial photos) and because of their size and their location (Leopold et al., 1994; Nadeau and Rains, 2007; Benstead and Leigh, 2012; Fritz et al., 2013). However the proportion of IRES in hydrological networks can be very large: for example, they

represent 60% of the length of rivers in the United States (Nadeau and Rains, 2007) and are considered to represent probably more than 50% of the global hydrological network (Larned et al., 2010; Datry et al., 2014b). Considering only gauging stations with continuous records may lead to severe underestimation of their regional extent (Snelder et al., 2013; De Girolamo et al., 2015; Eng et al., 2016). Recently, IRESs have seen a marked increase in interest stimulated by the challenges of water management facing the global change context (water scarcity issues, climate change impact, etc.) (Acuña et al., 2014; Datry et al., 2016b). Studies have characterized the hydrological functioning of IRES (Gallart et al., 2012; Costigan et al., 2016; Sarremejane et al., 2017) to assess the effects of flow intermittence on aquatic ecosystems (Larned et al., 2010; Datry et al., 2016b; Leigh et al., 2016; Leigh and Datry, 2017). IRES have been altered due to human actions (abstraction, hill dams, low-water support, pollution, etc.) despite their high and unique biodiversity (Datry et al., 2014; Garcia et al., 2017a). In addition, some perennial streams are becoming intermittent due to global change, water abstraction or river damming (Skoulikidis, 2009) and the extent of IRES may increase in the future (Döll and Schmied, 2012; Jaeger et al., 2014; Pumo et al., 2016; Garcia et al., 2017b; De Girolamo et al., 2017). A better hydrological understanding of IRES is now essential and an improved management requires knowing both the spatial extent and arrangement of IRES within the river network (Boulton, 2014; Acuña et al., 2017). Efforts have been made to estimate the spatial distribution of IRES at the catchment scale (Skoulikidis et al., 2011; Datry et al., 2016a), at the regional scale (Gómez et al., 2005) and at the national scale (Snelder et al., 2013). In France, Snelder et al. (2013) suggested a classification of IRES regimes and spatialized their distribution. Based on an analysis of the continuous gauging network, they showed that the proportion of IRES accounted for 20 to 39% of the hydrographic network. The accuracy of the obtained map is highly dependent on the density of the

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

70

71

flow monitoring network. The installation of additional gauging stations is expensive and headwaters systems may be difficult to monitor due to active geomorphology processes or to difficult access.

As a promising tool to advance the mapping of IRES, citizen science creates opportunities to overcome the lack of hydrological data and lead to densify the flow state observation network (Turner and Richter, 2011; Buytaert et al., 2014; Datry et al., 2016b) and could be used for hydrological model calibration (van Meerveld et al., 2017). In France, Datry et al. (2016a) used such data to describe the spatiotemporal dynamics of aquatic and terrestrial habitats within five river catchments located in the western part of France. They showed that processes resulting in flow intermittence were complex at a fine scale and could vary substantially among nearby catchments. However, these data were only available in a few catchments, limiting any attempt to map large-scale patterns of flow intermittence in river networks. Since this first attempt, new sources of observational data have become available in France thanks to the ONDE network (Observatoire National des Etiages, https://onde.eaufrance.fr). This unique network in Europe provides frequent discrete field observations (five inspections per year) of the flow intermittence across more than 3 300 sites throughout France and located mostly in headwater areas.

However discrete observations of intermittence do not provide any information on the persistence of dry conditions between two consecutive dates of observation. The rewetting-drying events could have significant impacts on communities whose survival is conditioned by the duration/frequency of drying. The duration of drying is of importance for ecologists, as one key driver of the composition and persistence of aquatic species (Vardakas et al., 2017; Kelso and Entrekin, 2018, Vadher et al., 2018). Temporal extrapolations of river flow regime are thus necessary to summarize the different facets of flow intermittence at various time scales, from daily to inter-annual.

The main objective of this paper is to use discrete (in space and time) field observations of flow intermittence to extrapolate over time the daily probability of drying (averaged at the regional scale). We first carried out a quantitative analysis of the ONDE network data in order to characterise the

information that they contribute in comparison with the data resulting from the conventional hydrological monitoring. Then, we developed two empirical models based on linear or logistic regressions to convert discontinuous series of flow intermittence observation from ONDE into continuous daily probability of drying, defined at the regional scale across France. Explanatory variables were derived from available continuous daily discharge and groundwater level data of a dense gauging/piezometer network, and models were calibrated using the ONDE discrete observations. The robustness of the models was tested using (1) an independent, dense regional data set of intermittence observations and (2) observations of the year 2017 excluded from the calibration. Finally, resulting models were used to extrapolate the regional probability of drying in France: (i) over the period 2012-2017 to identify the regions most affected by flow intermittence; (ii) over the period 1989-2017, using a reduced input dataset, to analyze temporal variability of flow intermittence at the national level.

2. Material and Methods

2.1. Study area

The study area is continental France and Corsica (550 000 km²). France is located in a temperate zone characterized by a variety of climates due to the influences of the Atlantic Ocean, the Mediterranean Sea and mountain areas.

We defined regions as combinations of "level-2 Hydro-EcoRegions" (HER2) and classes of hydrological regimes (HR). Hydro-EcoRegion (HER) corresponds to a typology developed for river management in accordance with the European Water Framework Directive. The Hydro-EcoRegion classification includes 22 "level-1 Hydro-EcoRegions" (HER1) based on geology, topography and climate, and considered as the primary determinants of the functioning of water ecosystems (Wasson et al., 2002). HER2 correspond to a finer classification accounting for stream size. HER2 have a mean drainage area of 5 000 km² (between 100 and 27 000 km²). The hydrological regimes classes

(HR) were identified by reference to the work carried out by (Sauquet et al., 2008) where it was possible to distinguish rainfall-fed regimes, transition and snowmelt-fed river flow regimes. Overall, we used 280 regions (that is, HER2-HR combinations) with a mean drainage area of 1 400 km² (between 4 and 20 000 km²).

2.2. ONDE dataset discrete national flow-state observations

The ONDE network was set up in 2012 by the French Biodiversity Agency (AFB, formerly ONEMA) with the aim of constituting a perennial network recording summer low flow levels and used to anticipate and manage water crisis during severe drought events (Nowak and Durozoi, 2012).

There are 3 300 ONDE sites distributed throughout France (Fig. 1). ONDE sites are located on headwater streams with a Strahler order strictly less than 5 and balanced across HER2 regions to take into account the representativeness of the hydrological contexts (Nowak and Durozoi, 2012). The ONDE network is stable over time. Observations are made monthly (around the 25th) by trained AFB staff, between April and September, every year since 2012. One of the statuses is assigned at each observation among "visible flow", "no visible flow" and "dried out". Here, we consider two intermittency statuses: "Flowing" when there is visible flow across the channel ("visible flow") and "Drying" when the channel is entirely devoid of surface water ("dried out") or when there is still water in the river bed but without visible flow (disconnected pools, lentic systems) ("no visible flow"). The proportion of drying sites determined on the basis of the ONDE network for each HER2-HR combination is considered as a good estimate of the daily Regional Probability of Drying (RPOD_{ONDE}) of streams with a Strahler order less than 5. Observed values of RPOD_{onde} are calculated as follows:

$$RPoD_{ONDE}(d) = \frac{(Ndrying)_{HER2-HR}}{(Nflowing+Ndrying)_{HER2-HR}}$$
(1)

where *d* denotes the observation date of the ONDE network, Ndrying and Nflowing are the number of drying and of flowing statuses observed at ONDE sites located in a same HER2-HR combination at the observation date *d*, respectively.

Figure 2 illustrates the complementary nature of the ONDE network to the already existing French river flow monitoring network HYDRO (http://www.hydro.eaufrance.fr). The ONDE sites and a set of 1 600 gauging stations available in the HYDRO database have been projected on the river network RHT (Theoretical Hydrographic Network; ; Pella *et al.*, 2012) and the drainage area and the elevation have been estimated. A large part of ONDE sites are located on small headwater streams with 70% of the sites with a drainage area of less than 50 km² while most of the gauging stations record flows of catchment of medium size (between 100 and 500 km²). Only four stations display a drainage area of more than 1 000 km². The distributions of elevation of the two databases look similar. The ONDE sites are mostly located on rivers with an elevation below 200 m (75% of sites). The ONDE sites are sparse at high elevations (95 sites located above 1 000 m). This bias is likely due to access difficulties in mountainous areas.

2.3. POC dataset: a denser regional dataset used for independent validation

A spatially denser citizen science dataset of flow-state observations in western France (Poitou-Charente region) (http://atlas.observatoire-environnement.org) has been used as validation dataset to test the robustness of our models calibrated with the ONDE dataset. The POC monitoring (2011-2013) covered more than 4 000 km of river length across 20 catchments. Each river was entirely surveyed every 1st and 15th of each month between June and October, resulting in eight observations per year. Four intermittency statuses were available in the POC dataset (Datry *et al.* 2016a) but to allow comparisons with the ONDE network, we pooled the two "Flowing" and "Low Flow" POC statuses into a single "Flowing" status and the two "No flow" and "Dry" statuses into the "Drying" status. This dataset is available as maps with flow states assigned to the inspected streams. Values of RPoD at each POC observation date is calculated in the same way as RPoD_{ONDE}. Thus RPoD_{POC} is given by the ratio between the number of drying and the total number of observations at each inspected streams located in a same HER2-HR.

2.4. Explanatory discharge dataset

Two discharge datasets (continuous daily time series) were used as explanatory variables of discrete intermittence observations, with the objective of extrapolating the intermittence frequency over time. The two datasets included time series of daily discharge extracted from the French River discharge monitoring network ("HYDRO database", http://www.hydro.eaufrance.fr/): (i) the 2011-2017 dataset with full records available between the 01/01/2011 and 31/06/2017; (ii) the 1989-2017 dataset concerning a reduced number of gauging stations and providing daily discharges between the 01/01/1989 and 31/06/2017. According to the hydrometric services in charge of the selected gauging stations, high quality of measurements was ensured and observed discharges were not or only slightly altered by human actions.

The 2011-2017 dataset was composed of 1 600 gauging stations distributed across France. Each stream where a HYDRO gauging station is located has been defined as IRES or perennial. Several definitions of IRES can be found in the literature (Huxter and van Meerveld, 2012, Eng et al., 2016; Reynolds et al., 2015). In this study, we considered stations as intermittent when five consecutive days with discharge less than 1 liter per second has been observed during the period of record.

The 1989-2017 dataset consisted of 630 gauging stations selected with less than 5% of missing data (continuous or not) during the period 1989-2017. This dataset has been thereafter used to estimate the proportion of drying before the creation of the ONDE network.

2.5. Explanatory groundwater level dataset

Because groundwater resources influence stream intermittence, we used available time series of the daily groundwater level available in the ADES database (http://www.ades.eaufrance.fr) at sites identified as involved in groundwater/surface water exchanges (Brugeron et al., 2012). Similarly to the discharge data, two sets of groundwater level data with records available over the two periods 2011-2017 and 1989-2017 have been selected. The level of alteration of groundwater levels by water withdrawal is unknown because no information is available at this scale.

The 2011-2017 dataset was composed by 750 piezometers with daily groundwater level data with less than 5% of missing data (continuous or not). The selection of 1989-2017 dataset was not easy because few groundwater level measurements were available in the database before 2000. For example, only five piezometers met the tolerance limit on missing values considered for the 1989-2017 discharge dataset. In order to extend the dataset and because groundwater levels were less variable than stream discharges, the proportion of permitted gaps was fixed to 20% between 1989 and 2017. This led us to select 150 piezometers. Thereafter, when the missing data period was less than 10 days, groundwater levels were reconstructed by linear interpolation in order to reduce the proportion of missing values to less than 5% for the 150 piezometers selected.

2.6. Statistical modeling of regional probability of drying

The parametric modeling strategy was based on 5 main steps (Fig. 3). The first step consisted in selecting all ONDE sites, gauging stations and piezometers located in a same HER2-HR combination. When the total number of gauging stations and piezometers was less than 5 for a HER2-HR combination, we merged the HER2-HR combination with a neighboring one located in the same HER1. This was done for 20 of the 280 regions. The second step consisted in calculating the RPOD_{ONDE} for each observation date (5 per year) and for all selected ONDE sites. In a third step, a flow duration curve was determined for each selected HYDRO gauging station. The average non-exceedance frequency of the observed discharge at gauging stations was averaged for the date of observation (d) at ONDE sites and the 5 days preceding the observation. The lag of six days accounted for the fact that ONDE survey dates in a region could differ by 5 days, and accounted for the inertia of physical processes (e.g. storage capacity); it was chosen after a few trials. The same operation was carried out with selected piezometers. Finally the hydrological conditions are described by the average (across stations) F of the non-exceedance frequencies of discharge (Fq) and groundwater levels (Fgw) with respect to the relative proportions of gauging stations and piezometers:

$$F(d) = \frac{\sum_{i=1}^{i=Nq} Fq_i + \sum_{j=1}^{j=Ngw} Fgw_j}{(Nq + Ngw)}$$
(2)

Where Fqi denotes the average non-exceedance frequency of discharge at the gauging station is calculated between d and d-5; Fgwj the average non-exceedance frequency of groundwater levels at the piezometer j calculated between d and d-5; Nq the number of gauging stations selected in a HER2-HR combination and Ngw the number of selected piezometers selected in the HER2-HR combination. The fourth step consisted in estimating the RPoD_{ONDE} as a function of F. Two types of regression were fitted for each HER2-HR combination across France:

a truncated logarithmic linear regression (LLR), with two parameters α_1 and β_1 :

$$RPoD_{LLR}(d) = \begin{cases} \min(1; \alpha_1 \times \ln(F(d)) + \beta_1) \text{ when } F < F0 \\ 0 \text{ when } F \ge F0 \end{cases}$$
(3)

- FO was fixed as the value of non-exceedance frequencies of discharge and groundwater levels at which no more drying was observed across the ONDE network (RPoD_{ONDE} = 0).
- a logistic regression (LR), with two parameters α_2 and β_2 :

$$Logit(RPoD_{LR}(d)) = ln\left(\frac{RPoD_{LR}(d)}{1 - RPoD_{LR}(d)}\right) = \alpha_2 \times F(d) + \beta_2$$
(4)

234 LR is a multivariate analysis method well known for its relevance in binary classification issues (Lee, 235 2005). The RPoD_{LR} was then calculated as following Eq. 5:

236
$$RPoD_{LR}(d) = \frac{\exp(\alpha_2 + \beta_2 F(d))}{1 + \exp(\alpha_2 + \beta_2 F(d))}$$
 (5)

Models were calibrated against observation available during the same period, 2012-2016, leaving out the year 2017 for an independent validation test. However, for the continuous temporal extrapolations (one over 2011-2017, the other 1989-2017), two models were built with different piezometers and gauging stations selected as explanatory variables (see section 2.4 and 2.5). Thus there are two set of regressions parameters specific to each dataset for both LLR and LR models leading to different prediction of RPoD.

Finally, in a fifth step, a daily regional probability of drying (RPoD) could be predicted for each HER2-HR combination with both models following analytical formulas (Eq. 3 and Eq. 5).

2.7. Model robustness: validation using independent data sets

We used (1) the POC independent data and (2) the 2017 ONDE year to test the robustness of the LLR and LR model to predict the intermittence frequency (1) in space and (2) over time. Note than when predicting on the POC datasets, a new model was calibrated using only ONDE sites located out of POC streams.

For both datasets (POC and ONDE 2017), the relative performance of the LLR and LR models was compared in multiple ways using both the 2011-2017 and the 1989-2017 datasets. The performance of each model was evaluated by the Nash-Sutcliffe efficiency criterion (NSE) (Nash and Sutcliffe, 1970):

$$NSE = 1 - \frac{\sum_{i=1}^{N} (RPoD_{ONDEi} - RPoD_{pri})^{2}}{\sum_{i=1}^{N} (RPoD_{ONDEi} - \overline{RPoD_{ONDEi}})^{2}}$$
(6)

where $RPoD_{ONDEi}$ is the average proportion of drying over the ONDE sites located in the HER2-HR combination at the i^{th} observation date, $RPoD_{pri}$ is the predicted regional probability of drying at the i^{th} observation date, $\overline{RPoD_{ONDEi}}$ is the mean of $RPoD_{ONDEi}$ over the period and N is the total number of observations in the ONDE network for each HER2-HR combination.

2.8. Model prediction

Both models have been calibrated over the period 2012-2016 and were then applied in a 5th step to predict the daily RPoD in France (Fig. 3). The RPoD was firstly predicted over the period 2012-2016 in order to identify the most affected regions by flow intermittence using the 2011-2017 datasets. The second application concerned the extrapolation of RPoD in France over a longer period using the 1989-2017 dataset to analyze the temporal variability of flow intermittence at the national level. It

should be noted that model predictions only concern streams with a Strahler order lower than 5 due to the ONDE sites location.

3. Results

3.1. Quantitative analysis

3.1.1. Inter-annual intermittence according to the raw discrete UNDE network
A total of 1127 ONDE sites have recorded at least one drying during the period 2012-2016
representing 35% of the 3 300 ONDE sites. From the ONDE database the proportion of drying at the
country scale was computed as the total number of drying over France divided by the total number
of ONDE observations available during the same year (Fig. 4a). Between 2012 and 2016, the most
critical year is 2012 with 15% of drying followed by 2016 (14%) and 2015 (14%) (Fig. 4a). The years
2013 and 2014 are less affected with only 6% of drying observed (Fig. 4a).
Dryings mainly occur between July and September but the evolution of the month's proportion of
drying can differ between years (Fig. 4b). In more detail, water levels in 2012 decrease in August
when the proportion of drying is 27% and the situation lasts until the end of September with 25% of
drying (Fig. 4b). In 2013, the drying proportion is lower than in 2012 but follows the same pattern
with an increase at the end of July (3%) and reaching 9% in August and in September. In 2014, the
first peak of drying (5%) is reached early in June. Then, the drying proportion decreases in July (3%)
and increases slightly in August 4% and reaching 7% in September. In 2015, the critical period occurs
at the end of July with 19% of drying and the proportion of drying decreases slightly at the end of
August (17%) until it reaches 9% in September. Finally, in 2016, the situation is gradually deteriorates
every month, reaching 20% of drying in August, and 28% in September.
Between 2012 and 2016, a proportion of drying higher than 50% is recorded on 93 ONDE sites and
their spatial distribution is very patchy at the France scale (black and dark grey dots, Fig. 5a). There

are only 158 ONDE sites with at least one drying every year and a variability of drying locations can

be observed across years. The south-east of France is heavily affected by rivers drying where drying proportion can exceed 75% annually (black dots, Fig. 5b-5f). The north-western part of France is less affected, although many ONDE sites show a drying proportion observed above 50% in 2014 and 2016 (Fig. 5d and 5f). Northeastern France is rather affected in 2012, 2014 and 2015 where several ONDE sites have more than 75% of drying (Fig. 5b, 5d and 5e). The south-west France is particularly affected in 2012 and 2015 (Fig. 5b and 5e).

3.1.2. Comparison of flow intermittence between the raw ONDE and HYDRO datasets

The HYDRO dataset includes 90 gauging stations located on streams considered as IRES, which represents only 5.6% of the 1 600 gauging stations against 35% for ONDE sites. At the national scale, the number of IRES seems underrepresented in the south-western, central, northeastern part of France and Corsica in comparison with sites experiencing drying in the ONDE network (Fig. 6).

The number of gauging stations with at least one drying (discharge < 1 l/s) observed between May and September varies between 79 in 2012 and 47 in 2014 (Table 1). The lowest numbers of gauging stations with drying are observed in the years 2013 and 2014 while the highest numbers are related to the years 2012, 2015 and 2016. This finding is consistent with the analysis of the ONDE network (Fig. 5a, d). The frequency of drying, corresponding to the ratio between the number of dry days and the total number of days between the 1st May and the 30th September (153 days), in contrast, is quite constant over the years (~30%). The number of gauging stations with drying over more than 50% of the time varies little between wet years (14 in 2013) and dry years (21 in 2015) unlike ONDE observations, suggesting a significant temporal variability in the frequency of drying between dry and wet years (Fig. 5).

3.2. Validation of the predicted regional probability of drying

3.2.1. Regression results

LLR and LR models, calibrated over the period 2012-2016, perform well with the 2011-2017 dataset with a mean NSE of 0.8 with LR model against 0.7 with LLR model (Fig. 7a and b). With the LR model,

50% of the HER2-HR combinations obtain a NSE greater than 0.8, representing a coverage of 65% of the French territory, while 33% of HER2-HR combinations display a NSE higher than 0.8 (50% of France coverage) with the LLR model. Regions with the highest performances are located in sedimentary plains, in the south-east of France and in the Pyrenees Mountains. Conversely, the worst performances are obtained in the mountainous regions of Alps as well as in the Massif Central. In these regions the size of the HER2 is rather small and the number of ONDE sites, gauging stations and piezometers per HER2-HR combinations are certainly too few to derive reliable relations. Despite pooling, estimating RPoD remains impossible for 9 HER2-HR combinations (4.5% of France coverage) because the number of ONDE sites, gauging stations and piezometers sites is insufficient (less than 5) to perform the regression analysis.

The performance level is lower when the 1989-2017 dataset is used in models: the mean NSE with the LR and LLR models is 0.7 and 0.6, respectively (Fig. 7c and d).

The LR and LLR models lead to similar performance range. However, the LR model outperforms the LLR model in terms of number of HER2-HR combinations with NSE greater than 0.8 (Fig. 7c and d). The performance is sensitive to the dataset. As expected, the best results are obtained with the denser network. A decrease in NSE by more than 0.2 is identified for 5% of the French territory when the 1989-2017 dataset is used (black areas; Fig. 7e and f). The regions with the most degraded values of NSE are small HER2-HR combinations located in eastern France (Fig. 7e and f).

The decrease in performance is mainly due to the difference in number of gauging stations and piezometers between the two datasets (Fig. 8). The most degraded NSEs correspond to HER2-HR combinations where the number of gauging stations and piezometers considered in regressions is the most reduced, i.e. with a loss higher than 50% of stations (black and dark grey dots; Fig 8a and b). However, the decrease in performance remains low (difference in NSE is below 0.1 for 75% and 64% of HER2-HR combinations with LLR and LR model, respectively).

3.2.2. Comparison to the POC database

The observed proportion of drying RPoD_{PoC} is rather well simulated by both LLR and LR models with the 2011-2017 explanatory dataset (NSE > 0.7 except for the year 2011, Fig. 9). In addition, the models are able to capture small fluctuations of RPoD_{PoC} during the summer period. The best results during the year 2011 are obtained with the LLR model (black curve; Fig. 9) and the LR model overestimates RPoD_{PoC} by 3% (dashed grey curve; Fig. 9). In 2012, the decline in water levels is more gradual than in 2011 and a marked peak is reached in September with 40% of RPoD_{PoC} (Fig. 9). This pattern is well reproduced by both models with a good fit to all observation points (Fig. 9). The year 2013 is less affected by drying occurrence and the maximum RPoD_{PoC} does not exceed 20% (Fig. 9). Curves of both models fit to observations well until the end of August. Note that the LR model is slightly closer to the observations around the peak in September compared to the LLR model. However the LR model overestimates the RPoD_{PoC} at the end of September and in October.

When the 1989-2017 dataset is used as explanatory variables, the simulations of RPoD are weakly degraded with both models (Fig. 9d, e, f). However the simulated pattern is similar to the observed one. The LLR model outperforms the LR model during the three years of validation with the 1989-2017 dataset (black curve; Fig. 9d, e, f).

3.2.3. Temporal patterns assessment of models between 2012 and 2017

During the calibration period, the LLR and LR models tend to better simulate the RPoD during dry years 2012 and 2016 (NSE = 0.8 with LLR and LR models; Tab. 2) than during wet years (e.g. 2014 with NSE < 0.7). The NSEs are lower during the months of May and June when few drying events are observed while NSEs are much better during the driest months of August and September.

During the validation year of 2017, both models obtain a similar performance over the year independent of datasets (NSE = 0.7).

Monthly NSEs in 2017 follow the same trend as monthly NSEs of the calibration period with lower NSEs in May (NSEs < 0.4) and June (NSEs = 0.5) and higher NSEs in July, August and September (NSEs

= 0.6) with both models independent of datasets. Figure 10 shows the dispersion between predicted RPoD and drying observed at ONDE sites in the scatter plot during the validation year 2017 (Fig. 10a and 10b) in comparison with the year 2012 which obtains the better NSE during calibration period (Fig. 10c and 10d). The NSEs obtained in 2017 are 0.72 with the LLR model and 0.68 with the LR model against 0.83 and 0.81 in 2012, respectively. The performance is slightly lower in 2017 but remains acceptable with NSEs close to 0.7 and both models seem able to predict RPoD out of the calibration period.

3.3. Application of regional models

3.3.1. Modeling of intermittencies severity between 2012 and 2016

Both models have been applied using the 2011-2017 dataset. Figure 11 displays the maximum number of consecutive days ($D_{RPoD>20\%}$) with RPoD higher than 20% simulated by both LLR and LR models. The most affected regions are located in the south-east of France and in the sedimentary plains which are consistent with the spatial pattern obtained from the ONDE observations (Fig. 5). The most impacted year followed the same hierarchy: the year 2012 is the most critical year with 30% of France displaying $D_{RPoD>20\%}$ higher than 60 days followed by the year 2015 (20% of France with $D_{RPoD>20\%} > 60$ days) and 2016 (15% of France with $D_{RPoD>20\%} > 60$ days) (Fig. 11). The years 2013 and 2014 are weakly affected with 5% and 6% of the France with $D_{RPoD>20\%}$ higher than 60 days, respectively.

The LR model tends to simulate shorter periods of drying, particularly in HER2-HR combinations located in the South-East France in 2013 and 2014 (Fig. 11). However, there is an overall agreement between RPoD simulated by both models in terms of spatial and temporal extent of dry streams.

3.3.2. Reconstitution of historical regional probability of drying

The trend temporal patterns of RPoD predicted by the two models, considering the 1989-2017 dataset, look similar between 1989 and 2016 and the simulated RPoD fit well to RPoD_{ONDE} (Fig. 12).

The proportion of drying is highly variable over the total simulation period, with alternating dry (1989 to 1991, 2003 to 2006, 2009 to 2012) and wet (1994 to 1995, 2000 to 2002; 2013 to 2014) phases. In spite of interannual variability, peaks of RPoD occur regularly between August and September, whether in dry years or wet years. This finding is consistent with the preeminence of rainfall fed river flow regime with low flows in summer, in France.

The highest values of RPoDs (above 35% over France) are observed in 1989, 1990, 1991, 2003 and 2005 (black curve, Fig. 12a and b). The RPoDs simulated during these dry years are out of the range of the observed values over the calibration period (2012-2016). Estimations are thus uncertain. However, the high values of RPoD are consistent with observations reported in previous studies (e.g. Larue and Giret, 2004; Snelder et al., 2013; Caillouet et al., 2017). Conversely, the years less affected by drying are simulated in 1994, 2001 and 2014 with an average RPoD below 15% throughout the year (black curve, Figs. 12a and b).

Results obtained with the LLR model are more contrasted in terms of extreme values than those obtained with the LR model (Fig. 12b).

4. Discussion

ONDE network complementarity with conventional flow monitoring network

The analysis of the ONDE observations shows that the proportion of rivers undergoing drying is significantly higher (35%) than that observed with the conventional monitoring (HYDRO database, 8%). This proportion although related to a short period of records 2012 and 2016 is consistent with the percentage of 39% of river segments classified as intermittent by Snelder at al. (2013). This analysis confirms the under-representation of IRES in the French HYDRO database, and probably others in other countries (flow are often uncontrolled in IRES). Without gauging stations located on headwaters, Snelder et al. (2013) were unable to predict IRES in eastern France (see Fig. 9, pp. 2694). The high density of ONDE sites makes it possible to improve the detection of drying and lead to

better understand the spatial distribution of IRES located at the upstream extent of the hydrographic network. The ONDE network encompasses various hydrological conditions which provides a more accurate assessment of inter-annual variability, differentiating between dry years (2012, 2015 and 2016) and wet years (2013, 2014) with clearly few drying occurrences.

The validation of the LR and LLR models against the spatially dense POC database also demonstrates the spatial representativeness of the ONDE network. Thanks to the qualitative information provided and to models such as statistical models developed here, it is now possible to capture drying event at the regional scale.

The ONDE sites are located on small headwater streams which can be very reactive to external disturbances (rainfall deficit, change in air temperature, increase in water withdrawals, etc.) and by nature are more likely to be IRES. The gauging stations available in the HYDRO database are located on larger streams and their hydrologic response to changes in external factors (environmental or human) is slower and drying occurred with greater inertia under temperate climate. Their uneven distribution across France does not allow to accurately characterize the inter-annual variability of drying development. Overall, the ONDE network provides very complementary information to conventional flow monitoring, leading to a better understanding of the processes of drying in upstream catchments.

Dependency on spatial gauging networks density

The performance obtained with the LR and LLR models is slightly better with the 2011-2017 dataset (mean NSE = 0.75) than those obtained with the 1989-2017 dataset (mean NSE > 0.65), whose network is less dense. HER2-HR combinations are the most degraded where the number of monitoring stations is the most decreased between the two datasets. The accuracy of the predictions is dependent on the number of gauging stations, ONDE sites and piezometers available to calibrate the regressions. Highest NSEs are obtained in western sedimentary plains and southeastern of France where a significant number of streams have dryings regardless of years (Fig. 5). The dominant river

flow regime in these regions is mainly influenced by precipitation and the lowest water levels are reached in August and September, which corresponds to the monitoring period of the ONDE database. They benefit from a dense monitoring network (gauging stations, ONDE sites, piezometers), which allows a better representation of the hydrological functioning of streams located within the same HER2. Conversely, performance was poor in mountainous areas such as in the Alps or the Massif Central (NSE < 0.4) where river flow regimes are diversified combining rainfall and snowmelt influences. By construction, the area of HER2-HR combination in mountains is reduced, which leads to a limited number of monitoring stations, certainly not sufficient to fit the models. Moreover, the observation period for ONDE sites was limited between May and September and dryings can be missed, particularly for streams influenced by snow or ice melting with potential drying periods in winter. In regions potentially concerned by drying events out of the May-September period, the actual ONDE monitoring strategy needs to be adapted to provide reliable temporal observations and extrapolations of drying frequencies.

We have chosen to average the non-exceedance frequencies of flows and groundwater levels in order to increase the monitoring network. If models had been calibrated using only gauging stations, performance will have been globally similar, or slightly better, in some HER2-HR combinations (Fig. 13). Therefore, we could not validate the real gain of using groundwater level data in addition to discharge data. This is certainly due to the dominant proportion of the gauging stations compared to the piezometers. Indeed, in the 2011-2017 dataset, the proportion of gauging stations is greater than 75% for more than 70% of HER2-HR combinations whereas the proportion of piezometers exceeds 70% in only 5% of HER2-HR combinations. Groundwater level data thus have small weight in regressions for this dataset. However, in the 1989-2017 dataset, the proportion of piezometer is greater than 70% in more than 30% of HER2-HR combinations. The presence of piezometers increases the density of the monitoring network in HER2-HR combinations with few available gauging stations. Thanks to groundwater level data, RPOD can be predicted on more HER2-HR combinations.

Interest in reconstructing the dynamic regional probability of drying

461

462

463

464

465

466

467

468

469

470

471

472

473

474

475

476

477

478

479

480

481

482

483

484

485

Spatio-temporal simulation of the probability of drying is crucial for advancing our understanding of IRES ecology and management. Some aquatic species can persist in dry reach for a few days, weeks or months, while some are highly sensitive to desiccation (Datry, 2012; Storey and Quinn, 2013; Stubbington and Datry, 2013). Estimating the total duration of days with drying at the reach scale is therefore needed to understand biological patterns in river networks (Kelso and Entrekin, 2018). To our knowledge, no study has proposed to reconstruct daily flow states time series of headwater streams at the country scale as France (> 500 000 km²) using discrete observations in time and space. In the literature, studies at national scale remain focused on the detection and the mapping of IRES because these rivers are historically poorly investigated and their proportion in existing hydrographic networks remains inaccurate or misunderstood (Nadeau and Rains, 2007; Snelder et al., 2013). Recently, several studies proposed alternative methodologies in order to estimate metrics in ungauged IRES (Gallart et al., 2016) or to predict daily streamflow in river basin experiencing flow intermittence (De Girolamo et al., 2017b) but remain applicable at local scale. This study provides a first regional approach to use discrete data obtained from regular observations. The average non-exceedance frequency is a global hydrological statistic that only captures the hydrological conditions at the regional scale in modelling the RPoD. For rainfall-driven river flow regime, the effect of rainfall events on flow intermittence at the HER2-HR scale is probably indirectly reflected by the daily discharge and groundwater levels used to calculate the average nonexceedance frequency. However, when more observation data are available, it is likely that including more detailed descriptors of rainfall events and local geology could improve our approach without. In France, based on the 2011-2017 dataset, both models suggest highest values of RPoD along the Mediterranean coast (D_{RPoD>20%}> 100 days each year). Rivers in this region are subject to a predominantly pluvial regime (Class 7; Sauquet et al., 2008), i.e. hot and dry summers follow by intense rainfall events in autumn, leading to high flows in November (Skoulikidis et al., 2017b). The catchments in this region are small and particularly reactive to environmental changes, making them highly sensitive to flow intermittence. Rivers located in the sedimentary plain in western France are also very impacted by flow intermittence. The regime is also influenced by precipitation and for the basins subject to intense agriculture significant water abstractions during summer in this region reduce water availability in rivers and in aquifers which are no longer able to support the low water levels and which lead to increased flow intermittence. Regarding alteration issues in our datasets, we do not have access to the exact location and the volumes of water withdrawal for irrigation purposes. However, due to their upstream location, water availability is expected to be low, which may limit potential withdrawals and as consequence flow alteration at ONDE sites. The alteration of groundwater levels is unknown because no information is available. However, in sedimentary plains where agricultural crops dominate the landscape, we are not sure that no human action affects low flows. It is important to note that the responses of biological communities to artificial flow intermittence is still poorly understood compared to natural IRES (Datry et al., 2014b, Skoulikidis et al., 2017a).

Validity of historical regional probability of drying during severe low-flow period

The second application aimed at reconstructing historical RPoD over the period 1989-2016. Both models suggest highest values of mean RPoD (> 35%) in 1989-1991, 2003 and 2005. During these dry years, predicted values of RPoD result from extrapolation but are consistent with published studies (Mérillon and Chaperon, 1990, Moreau, 2004). For example, Mérillon (1992) estimated that for the whole of France, 11 000 km of rivers were dried at the end of summers of 1989 and 1990. Caillouet et al (2016) found that the low-flow event observed in 1989-1990 was particularly severe in terms of duration and affected 95% of France. Snelder et al. (2013) showed from 628 gauging stations that the years 1989-1991, 2003 and 2005 had witnessed particularly high values of duration and frequency of drying. They found that regions with the highest probability of drying were located along the Mediterranean and Atlantic coasts, which is consistent with ONDE observations and with our results.

Both models suggest the same sequence of dry and wet years. However the application of the LLR model lead to less contrasted RPoD than the LR model (Fig. 12).

To illustrate these differences, the RPoD has been simulated by both models with an extreme F of 1% (Fig. 14). The RPoD_{LLR} is significantly higher and exceeds 80% in 30% of the study area against only 5% of the area with the RPoD_{LR}. On the other hand, models simulate low RPoD in HER2-HR combinations where the RPoD_{ONDE} is very low between 2012-2016, even when F was 1% because this situation never occurred during the calibration period (Fig. 14). The logistic function of the LR model takes an S-shape which induced a decrease of the slope of the curve toward extreme values observed during the calibration period (2012-2016). The truncated logarithmic function of the LLR model is not bounded and RPoD can reach 100% during extreme low flow events by extrapolation. Since the ONDE network monitoring period does not include a period with drought as severe as in the 1990s, it is not currently possible to appreciate the relative performance of the two models. Refining extrapolated values requires additional information on headwater collected during more severe droughts than those observed during the last five years and then gives support to the pursuit of the ONDE network.

5. Conclusion

This paper investigates the spatial and temporal dynamics of the regional probability of drying (RPoD) of headwater streams by taking benefit from qualitative and discontinuous data provided by the ONDE network. Two models based on linear or logistic regressions have been developed and succeeded to reconstruct the temporal dynamics of RPoD. They are based on a strong relationship between the non-exceedance frequencies of discharges and groundwater levels as a function of the proportion of drying observed at ONDE sites per HER2-HR combination. LLR and LR models show similar performance and perform well between 2011 and 2017. The accuracy of predictions is dependent on the number of gauging stations, ONDE sites and piezometers available to calibrate the regressions. Regions with the highest performance are located in the sedimentary plains, where the

monitoring network is dense and where the RPoD is the highest. Conversely, the worst performances are obtained in the mountainous regions. Finally, both models have been used to reconstruct historical RPoD between 1989 and 2016 and suggest highest values of mean RPoD (> 35%) in 1989-1991, 2003 and 2005. This is consistent with other published studies but the high density of ONDE sites makes it possible to improve the detection of drying and lead to better capturing of the spatial distribution of IRES located at the upstream extent of the hydrographic network. Moreover, the duration of drying is of importance for ecologists and the prediction of a daily RPoD provides one key driver of the composition and persistence of aquatic species.

From a methodological point of view, our method relating discrete drying observation obtained by citizen science networks to continuous daily gauging data seems robust across the highly diverse (climate and topography) regions of France, and provides good predictions in an independent region excluded from the calibration process (PoC). These two results suggest a potential application of our approach in other countries. Citizen science creates opportunities to overcome the lack of hydrological data and lead to densify the flow state observation network (Turner and Richter, 2011; Buytaert et al., 2014) and remains less expensive than the installation of additional gauging stations to survey flow intermittence. The next step will be to use this regional approach to simulate the RPoD in future periods by taking into account effects of climate change through predicted discharge and groundwater level data. This would allow quantification of the evolution of the probability of drying between the current period and the different climate projections provided by the latest IPCC Report (IPCC 2014a, 2014b) and would assist decision makers in defining protocols for restoring flows with appropriate measures to preserve aquatic ecosystems (Woelfle-Erskine, 2017).

Secondly, further work is needed to develop an approach capable of reconstructing the drying dynamics locally by differentiating each stream. Our approach remains spatially valid to estimate RPoDs at the scale of HER2-HR combinations but does not allow characterizing the variability of the probability of drying occurrence between nearby streams within these regions. Statistical tools such

as neural networks (Breiman, 2001) have shown good ability to assess both the occurrence and extent of perennial and temporary segments (González-Ferreras and Barquín, 2017) and could be investigated as an alternative method to reconstruct locally the temporal variability of drying.

6. Acknowledgment

The authors wish to thank A. van Loon and C. Sefton for their valuable comments, suggestions and positive feedback on the manuscript. The research project was partly funded by the French Agency for Biodiversity (AFB, formerly ONEMA). This study is based upon works from COST Action CA15113 (SMIRES, Science and Management of Intermittent Rivers and Ephemeral Streams, www.smires.eu), supported by COST (European Cooperation in Science and Technology).

7. References

- 571 Acuña, V., Datry, T., Marshall, J., Barceló, D., Dahm, C. N., Ginebreda, A., McGregor, G., Sabater, S.,
- 572 Tockner, K. and Palmer, M. A.: Why should we care about temporary waterways?, Science,
- 573 343(6175), 1080–1081, 2014.
- 574 Acuña, V., Hunter, M. and Ruhí, A.: Managing temporary streams and rivers as unique rather than
- second-class ecosystems, Biological Conservation, 211, 12–19, doi:10.1016/j.biocon.2016.12.025,
- 576 2017.

- 577 Benda, L., Hassan, M. A., Church, M. and May, C. L.: Geomorphology Of Steepland Headwaters: The
- 578 Transition From Hillslopes To Channels1, Journal of the American Water Resources Association,
- 579 41(4), 835, 2005.
- Benstead, J. P. and Leigh, D. S.: An expanded role for river networks, Nature Geoscience, 5(10), 678-
- 581 679, 2012.
- 582 Boulton, A. J.: Conservation of ephemeral streams and their ecosystem services: what are we
- 583 missing?: Editorial, Aquatic Conservation: Marine and Freshwater Ecosystems, 24(6), 733–738,
- 584 doi:10.1002/aqc.2537, 2014.
- 585 Breiman, L.: Random forests, Machine learning, 45(1), 5–32, 2001.
- 586 Brugeron, A., Allier, D., Klinka, T.: Approche exploratoire des liens entre référentiels hydrogéologique
- 587 et hydrographique : Premières identifications des piézomètres potentiellement représentatifs d'une
- relation nappe/rivière et contribution à leur valorisation. Rapport final BRGM/RP-61047-FR. 241 p,
- 589 2012.
- 590 Buytaert, W., Zulkafli, Z., Grainger, S., Acosta, L., Alemie, T. C., Bastiaensen, J., De Bi??vre, B., Bhusal,
- 591 J., Clark, J., Dewulf, A., Foggin, M., Hannah, D. M., Hergarten, C., Isaeva, A., Karpouzoglou, T.,
- Pandeya, B., Paudel, D., Sharma, K., Steenhuis, T., Tilahun, S., Van Hecken, G. and Zhumanova, M.:
- 593 Citizen science in hydrology and water resources: opportunities for knowledge generation,
- 594 ecosystem service management, and sustainable development, Frontiers in Earth Science, 2,
- 595 doi:10.3389/feart.2014.00026, 2014.
- Clarke, A., Mac Nally, R., Bond, N. and Lake, P. S.: Macroinvertebrate diversity in headwater streams:
- 597 a review, Freshwater Biology, 53(9), 1707–1721, doi:10.1111/j.1365-2427.2008.02041.x, 2008.
- 598 Costigan, K. H., Jaeger, K. L., Goss, C. W., Fritz, K. M. and Goebel, P. C.: Understanding controls on
- flow permanence in intermittent rivers to aid ecological research: integrating meteorology, geology
- and land cover: Integrating Science to Understand Flow Intermittence, Ecohydrology, 9(7), 1141-
- 601 1153, doi:10.1002/eco.1712, 2016.
- 602 Datry, T.: Benthic and hyporheic invertebrate assemblages along a flow intermittence gradient:
- 603 effects of duration of dry events: River drying and temporary river invertebrates, Freshwater Biology,
- 604 57(3), 563–574, doi:10.1111/j.1365-2427.2011.02725.x, 2012.
- Datry, T., Larned, S. T., Fritz, K. M., Bogan, M. T., Wood, P. J., Meyer, E. I. and Santos, A. N.: Broad-
- scale patterns of invertebrate richness and community composition in temporary rivers: effects of
- flow intermittence, Ecography, 37(1), 94–104, doi:10.1111/j.1600-0587.2013.00287.x, 2014a.
- 608 Datry, T., Larned, S. T. and Tockner, K.: Intermittent Rivers: A Challenge for Freshwater Ecology,
- 609 BioScience, 64(3), 229–235, doi:10.1093/biosci/bit027, 2014b.

- Datry, T., Pella, H., Leigh, C., Bonada, N. and Hugueny, B.: A landscape approach to advance
- 611 intermittent river ecology, Freshwater Biology, 61(8), 1200–1213, doi:10.1111/fwb.12645, 2016a.
- 612 Datry, T., Fritz, K. and Leigh, C.: Challenges, developments and perspectives in intermittent river
- 613 ecology, Freshwater Biology, 61(8), 1171–1180, doi:10.1111/fwb.12789, 2016b.
- De Girolamo, A. M., Lo Porto, A., Pappagallo, G., Tzoraki, O. and Gallart, F.: The Hydrological Status
- 615 Concept: Application at a Temporary River (Candelaro, Italy): EVALUATING HYDROLOGICAL STATUS
- 616 IN TEMPORARY RIVERS, River Research and Applications, 31(7), 892–903, doi:10.1002/rra.2786,
- 617 2015.
- 618 De Girolamo, A. M., Bouraoui, F., Buffagni, A., Pappagallo, G. and Lo Porto, A.: Hydrology under
- climate change in a temporary river system: Potential impact on water balance and flow regime,
- 620 River Research and Applications, doi:10.1002/rra.3165, 2017a.
- 621 De Girolamo, A. M., Barca, E., Pappagallo, G. and Lo Porto, A.: Simulating ecologically relevant
- 622 hydrological indicators in a temporary river system, Agricultural Water Management, 180, 194–204,
- 623 doi:10.1016/j.agwat.2016.05.034, 2017b.
- 624 Döll, P. and Schmied, H. M.: How is the impact of climate change on river flow regimes related to the
- 625 impact on mean annual runoff? A global-scale analysis, Environmental Research Letters, 7(1),
- 626 014037, doi:10.1088/1748-9326/7/1/014037, 2012.
- 627 Eng, K., Wolock, D. M. and Dettinger, M. D.: Sensitivity of Intermittent Streams to Climate Variations
- 628 in the USA: Sensitivity of Intermittent Streams, River Research and Applications, 32(5), 885–895,
- 629 doi:10.1002/rra.2939, 2016.
- 630 Finn, D. S., Bonada, N., M?rria, C. and Hughes, J. M.: Small but mighty: headwaters are vital to stream
- 631 network biodiversity at two levels of organization, Journal of the North American Benthological
- 632 Society, 30(4), 963–980, doi:10.1899/11-012.1, 2011.
- Fritz, K. M., Hagenbuch, E., D'Amico, E., Reif, M., Wigington, P. J., Leibowitz, S. G., Comeleo, R. L.,
- 634 Ebersole, J. L. and Nadeau, T.-L.: Comparing the Extent and Permanence of Headwater Streams From
- 635 Two Field Surveys to Values From Hydrographic Databases and Maps, JAWRA Journal of the
- 636 American Water Resources Association, 49(4), 867–882, doi:10.1111/jawr.12040, 2013.
- Gallart, F., Prat, N., García-Roger, E. M., Latron, J., Rieradevall, M., Llorens, P., Barberá, G. G., Brito,
- D., De Girolamo, A. M., Lo Porto, A., Buffagni, A., Erba, S., Neves, R., Nikolaidis, N. P., Perrin, J. L.,
- 639 Querner, E. P., Quiñonero, J. M., Tournoud, M. G., Tzoraki, O., Skoulikidis, N., Gómez, R., Sánchez-
- Montoya, M. M. and Froebrich, J.: A novel approach to analysing the regimes of temporary streams
- in relation to their controls on the composition and structure of aquatic biota, Hydrology and Earth
- System Sciences, 16(9), 3165–3182, doi:10.5194/hess-16-3165-2012, 2012.
- 643 Gallart, F., Llorens, P., Latron, J., Cid, N., Rieradevall, M. and Prat, N.: Validating alternative
- methodologies to estimate the regime of temporary rivers when flow data are unavailable, Science
- of The Total Environment, 565, 1001–1010, doi:10.1016/j.scitotenv.2016.05.116, 2016.
- 646 Garcia, C., Gibbins, C. N., Pardo, I. and Batalla, R. J.: Long term flow change threatens invertebrate
- diversity in temporary streams: Evidence from an island, Science of The Total Environment, 580,
- 648 1453–1459, doi:10.1016/j.scitotenv.2016.12.119, 2017a.

- 649 Garcia, C., Amengual, A., Homar, V. and Zamora, A.: Losing water in temporary streams on a
- 650 Mediterranean island: Effects of climate and land-cover changes, Global and Planetary Change, 148,
- 651 139–152, doi:10.1016/j.gloplacha.2016.11.010, 2017b.
- 652 Gómez, R., Hurtado, I., Suárez, M. L. and Vidal-Abarca, M. R.: Ramblas in south-east Spain:
- threatened and valuable ecosystems, Aquatic Conservation 15, 387–402, doi:10.1002/aqc.680, 2005.
- 654 González-Ferreras, A. M. and Barquín, J.: Mapping the temporary and perennial character of whole
- 655 river networks: MAPPING FLOW PERMANENCE IN RIVER NETWORK, Water Resources Research,
- 656 doi:10.1002/2017WR020390, 2017.
- 657 Huxter, E. H. H. and (Ilja) van Meerveld, H. J.: Intermittent and Perennial Streamflow Regime
- 658 Characteristics in the Okanagan, Canadian Water Resources Journal / Revue canadienne des
- ressources hydriques, 37(4), 391–414, doi:10.4296/cwrj2012-910, 2012.
- 660 Jaeger, K. L., Olden, J. D. and Pelland, N. A.: Climate change poised to threaten hydrologic
- 661 connectivity and endemic fishes in dryland streams, Proceedings of the National Academy of
- Sciences, 111(38), 13894–13899, doi:10.1073/pnas.1320890111, 2014.
- 663 Kelso, J. E. and Entrekin, S. A.: Intermittent and perennial macroinvertebrate communities had similar
- richness but differed in species trait composition depending on flow duration, Hydrobiologia, 807(1),
- 665 189–206, doi:10.1007/s10750-017-3393-y, 2018.
- 666 Larned, S. T., Datry, T., Arscott, D. B. and Tockner, K.: Emerging concepts in temporary-river ecology,
- 667 Freshwater Biology, 55(4), 717–738, doi:10.1111/j.1365-2427.2009.02322.x, 2010.
- 668 Lee, S.: Application of logistic regression model and its validation for landslide susceptibility mapping
- 669 using GIS and remote sensing data, International Journal of Remote Sensing, 26(7), 1477–1491,
- 670 doi:10.1080/01431160412331331012, 2005.
- 671 Leigh, C. and Datry, T.: Drying as a primary hydrological determinant of biodiversity in river systems:
- a broad-scale analysis, Ecography, 40(4), 487–499, doi:10.1111/ecog.02230, 2017.
- 673 Leigh, C., Boulton, A. J., Courtwright, J. L., Fritz, K., May, C. L., Walker, R. H. and Datry, T.: Ecological
- 674 research and management of intermittent rivers: an historical review and future directions,
- 675 Freshwater Biology, 61(8), 1181–1199, doi:10.1111/fwb.12646, 2016.
- 676 Leopold, L. B.: A View of the River. Harvard University Press, Cambridge, Massachusetts, USA, 1994.
- 677 Leopold, L. B., Wolman, M. G. and Miller, J. P.: Fluvial Processes in Geomorphology. Dover
- 678 Publications, New York, USA, 1964.
- 679 Meyer, J. L., Strayer, D. L., Wallace, J. B., Eggert, S. L., Helfman, G. S. and Leonard, N. E.: The
- 680 Contribution of Headwater Streams to Biodiversity in River Networks1: The Contribution of
- 681 Headwater Streams to Biodiversity in River Networks, JAWRA Journal of the American Water
- 682 Resources Association, 43(1), 86–103, doi:10.1111/j.1752-1688.2007.00008.x, 2007.
- 683 Nadeau, T.-L. and Rains, M. C.: Hydrological Connectivity Between Headwater Streams and
- 684 Downstream Waters: How Science Can Inform Policy1: Hydrological Connectivity Between
- 685 Headwater Streams and Downstream Waters: How Science Can Inform Policy, JAWRA Journal of the
- 686 American Water Resources Association, 43(1), 118–133, doi:10.1111/j.1752-1688.2007.00010.x,
- 687 2007.

- Nash, J. E. and Sutcliffe, J. V.: River flow forecasting through conceptual models part I A discussion
- of principles, Journal of Hydrology, 10(3), 282–290, doi:10.1016/0022-1694(70)90255-6, 1970.
- 690 Nowak, C. and Durozoi, B.: Observatoire National Des Etiages, Note technique, ONEMA., 2012.
- 691 Pella, H., Lejot, J., Lamouroux, N. and Snelder, T.: Le réseau hydrographique théorique (RHT) français
- et ses attributs environnementaux, Géomorphologie: relief, processus, environnement, 18(3), 317-
- 693 336, 2012.
- Pumo, D., Caracciolo, D., Viola, F. and Noto, L. V.: Climate change effects on the hydrological regime
- of small non-perennial river basins, Science of The Total Environment, 542, 76–92,
- 696 doi:10.1016/j.scitotenv.2015.10.109, 2016.
- Reynolds, L. V., Shafroth, P. B. and LeRoy Poff, N.: Modeled intermittency risk for small streams in the
- 698 Upper Colorado River Basin under climate change, Journal of Hydrology, 523, 768-780,
- 699 doi:10.1016/j.jhydrol.2015.02.025, 2015.
- 700 Sarremejane, R., Cañedo-Argüelles, M., Prat, N., Mykrä, H., Muotka, T. and Bonada, N.: Do
- 701 metacommunities vary through time? Intermittent rivers as model systems, Journal of Biogeography,
- 702 44(12), 2752–2763, doi:10.1111/jbi.13077, 2017.
- 703 Sauquet, E., Gottschalk, L. and Krasovskaia, I.: Estimating mean monthly runoff at ungauged
- locations: an application to France, Hydrology Research, 39(5–6), 403, doi:10.2166/nh.2008.331,
- 705 2008.
- 706 Skoulikidis, N. T.: The environmental state of rivers in the Balkans—A review within the DPSIR
- 707 framework, Science of The Total Environment, 407(8), 2501–2516,
- 708 doi:10.1016/j.scitotenv.2009.01.026, 2009.
- 709 Skoulikidis, N. T., Vardakas, L., Karaouzas, I., Economou, A. N., Dimitriou, E. and Zogaris, S.: Assessing
- 710 water stress in Mediterranean lotic systems: insights from an artificially intermittent river in Greece,
- 711 Aquatic Sciences, 73(4), 581–597, doi:10.1007/s00027-011-0228-1, 2011.
- 712 Skoulikidis, N. T., Vardakas, L., Amaxidis, Y. and Michalopoulos, P.: Biogeochemical processes
- 713 controlling aquatic quality during drying and rewetting events in a Mediterranean non-perennial river
- 714 reach, Science of The Total Environment, 575, 378–389, doi:10.1016/j.scitotenv.2016.10.015, 2017a.
- 715 Skoulikidis, N. T., Sabater, S., Datry, T., Morais, M. M., Buffagni, A., Dörflinger, G., Zogaris, S., del Mar
- Sánchez-Montoya, M., Bonada, N., Kalogianni, E., Rosado, J., Vardakas, L., De Girolamo, A. M. and
- 717 Tockner, K.: Non-perennial Mediterranean rivers in Europe: Status, pressures, and challenges for
- 718 research and management, Science of The Total Environment, 577, 1–18,
- 719 doi:10.1016/j.scitotenv.2016.10.147, 2017b.
- 720 Snelder, T. H., Datry, T., Lamouroux, N., Larned, S. T., Sauquet, E., Pella, H. and Catalogne, C.:
- Regionalization of patterns of flow intermittence from gauging station records, Hydrology and Earth
- 722 System Sciences, 17(7), 2685–2699, doi:10.5194/hess-17-2685-2013, 2013.
- 723 Storey, R. G. and Quinn, J. M.: Survival of aquatic invertebrates in dry bed sediments of intermittent
- 724 streams: temperature tolerances and implications for riparian management, Freshwater Science,
- 725 32(1), 250–266, doi:10.1899/12-008.1, 2013.

- 726 Stubbington, R. and Datry, T.: The macroinvertebrate seedbank promotes community persistence in
- 727 temporary rivers across climate zones, Freshwater Biology, 58(6), 1202-1220,
- 728 doi:10.1111/fwb.12121, 2013.
- 729 Turner, D. S. and Richter, H. E.: Wet/Dry Mapping: Using Citizen Scientists to Monitor the Extent of
- 730 Perennial Surface Flow in Dryland Regions, Environmental Management, 47(3), 497-505,
- 731 doi:10.1007/s00267-010-9607-y, 2011.
- Vadher, A. N., Millett, J., Stubbington, R. and Wood, P. J.: Drying duration and stream characteristics
- 733 influence macroinvertebrate survivorship within the sediments of a temporary channel and exposed
- gravel bars of a connected perennial stream, Hydrobiologia, doi:10.1007/s10750-018-3544-9, 2018.
- van Meerveld, H. J. I., Vis, M. J. P. and Seibert, J.: Information content of stream level class data for
- 736 hydrological model calibration, Hydrology and Earth System Sciences, 21(9), 4895–4905,
- 737 doi:10.5194/hess-21-4895-2017, 2017.
- 738 Vardakas, L., Kalogianni, E., Economou, A. N., Koutsikos, N. and Skoulikidis, N. T.: Mass mortalities
- and population recovery of an endemic fish assemblage in an intermittent river reach during drying
- 740 and rewetting, Fundamental and Applied Limnology / Archiv für Hydrobiologie,
- 741 doi:10.1127/fal/2017/1056, 2017.

746

- 742 Wasson, J.-G., Chandesris, A., Pella, H. and Blanc, L.: Typology and reference conditions for surface
- 743 water bodies in France: the hydro-ecoregion approach, TemaNord, 566, 37–41, 2002.
- 744 Woelfle-Erskine, C.: Collaborative Approaches to Flow Restoration in Intermittent Salmon-Bearing
- 745 Streams: Salmon Creek, CA, USA, Water, 9(3), 217, doi:10.3390/w9030217, 2017.

	Stations with at least one drying event	Stations with drying > 50%	Frequency of discharge < 1 l/s
2012	79	19	32.7
2013	47	14	37.9
2014	54	15	32.9
2015	76	21	31.1
2016	71	19	28.6

Table 1. Annual statistics on flow intermittence calculated on HYDRO gauging stations between the $1^{\rm st}$ May and the $30^{\rm th}$ September

		2011-2017 dataset					1989-2017 dataset						
		Calibration				Valid.	Calibration					Valid.	
		2012	2013	2014	2015	2016	2017	2012	2013	2014	2015	2016	2017
	May	0.2	0.0	0.5	0.5	0.6	0.4	0.2	0.0	0.3	0.0	0.7	0.2
	June	0.6	0.3	0.8	0.5	0.8	0.5	0.6	0.3	0.5	0.3	0.8	0.5
LLR	July	0.7	0.5	0.6	0.6	0.8	<mark>0.7</mark>	0.7	0.5	0.5	0.4	0.8	<mark>0.6</mark>
model	August	0.8	0.6	0.7	0.7	0.8	<mark>0.6</mark>	0.7	0.5	0.5	0.5	0.8	<mark>0.6</mark>
	Sept.	0.7	0.8	0.6	0.6	0.7	<mark>0.6</mark>	0.6	0.7	0.5	0.5	0.6	<mark>0.6</mark>
	May - Sept	0.8	0.8	0.7	0.7	0.8	<mark>0.7</mark>	0.8	0.7	0.5	0.6	0.8	<mark>0.7</mark>
	May	0.2	0.0	0.5	0.1	0.6	0.3	0.3	0.0	0.3	0.0	0.7	0.2
	June	0.6	0.5	0.8	0.5	0.8	0.4	0.6	0.4	0.5	0.3	0.7	0.4
LR	July	0.7	0.6	0.5	0.6	0.8	<mark>0.6</mark>	0.7	0.4	0.5	0.4	0.8	<mark>0.6</mark>
model	August	0.7	0.6	0.7	0.6	0.7	<mark>0.6</mark>	0.6	0.4	0.5	0.4	0.7	<mark>0.5</mark>
	Sept.	0.6	0.8	0.6	0.7	0.7	<mark>0.6</mark>	0.5	0.6	0.4	0.5	0.6	<mark>0.6</mark>
	May - Sept	0.8	0.8	0.7	0.7	0.8	<mark>0.7</mark>	0.8	0.7	0.5	0.6	0.8	<mark>0.7</mark>

Table 2. NSE criteria obtained between 2012 and 2017 with the LLR and LR models calibrated over the period 2012-2016.

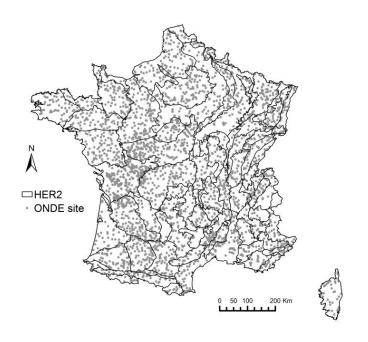


Figure 1. Location of the 3 300 ONDE sites and partition into HER2.

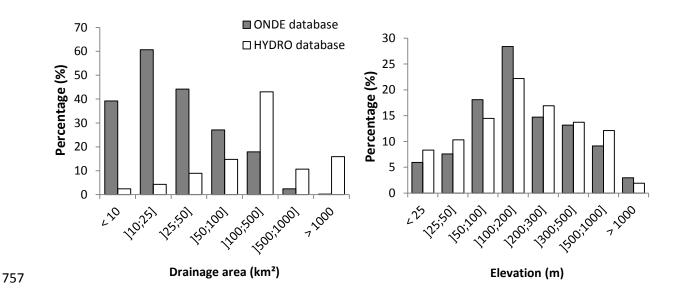


Figure 2. Distribution of the 3 300 ONDE sites and of the 1 600 gauging stations available in the HYDRO database against: (a) drainage area and (b) elevation.

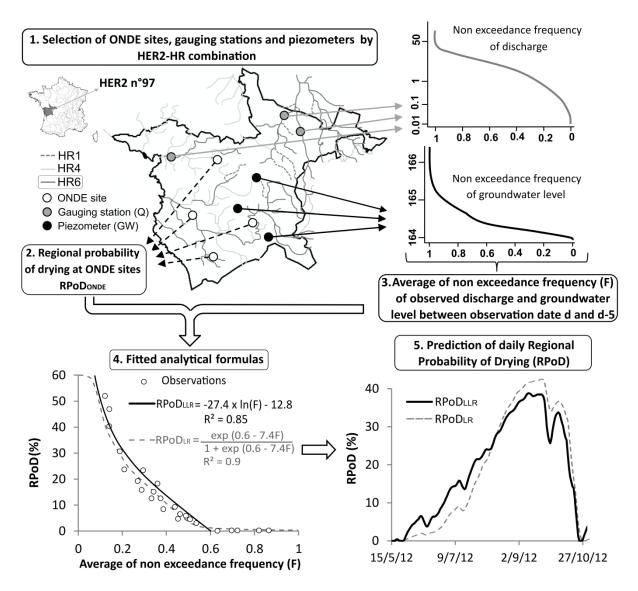


Figure 3. Strategy of parametric modeling (step 1-4) developed to predict (step 5) the regional probability of drying (RPoD) by HER2-HR combination in France.

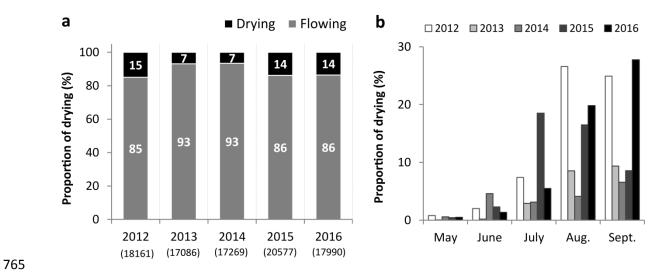


Figure 4. (a) Distribution of yearly proportion of drying observed with the ONDE network with the total yearly number of ONDE observations written in brackets and (b) distribution of proportions of drying per year and per month.

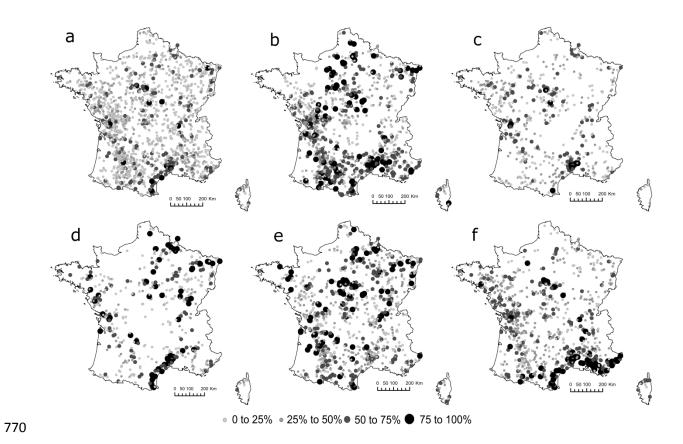


Figure 5. Distribution of the percentages of drying observed at ONDE sites for the years: (a) 2012-2016, (b) 2012, (c) 2013, (d) 2014, (e) 2015 and (f) 2016.

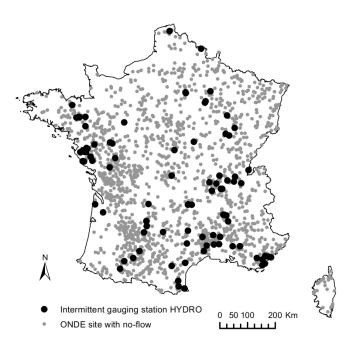


Figure 6. Map of ONDE sites and HYDRO gauging stations having at least one drying.

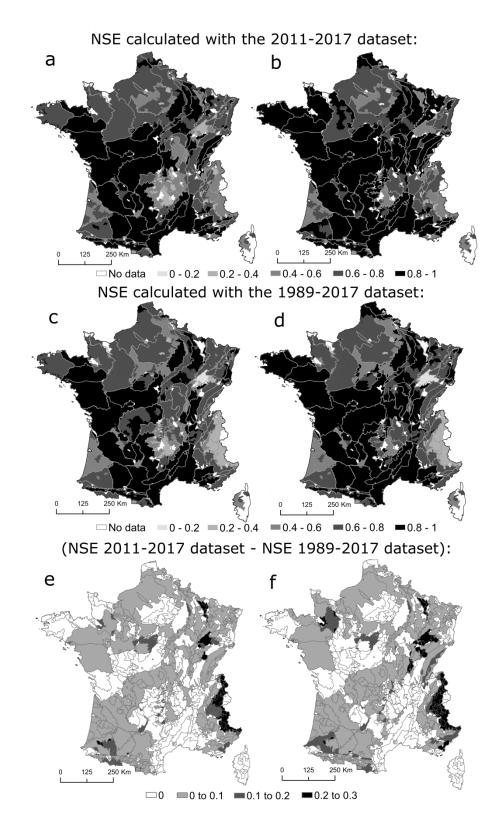


Figure 7. Map of Nash-Sutcliffe criteria (NSE) obtained for each HER2-HR combination between 2012 and 2016 with the 2011-2017 and 1989-2017 datasets according to: (a) and (c) a log-linear regression (LLR) model; (b) and (d) a logistic regression (LR) model. NSE differences between the 2011-2017 dataset and the 1989-2017 dataset are represented for: (e) LLR model and (f) LR model.

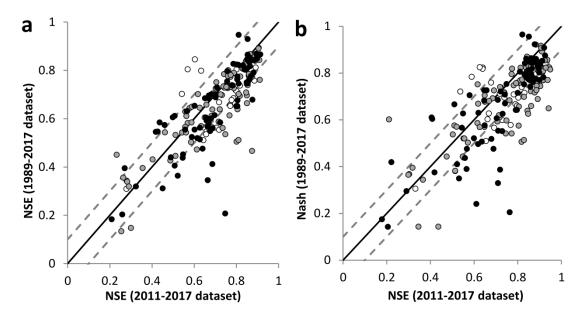


Figure 8. NSE calculated for each HER2-HR combination between 2012 and 2016 with the 1989-2017 dataset as a function of NSE calculated with 2011-2017 dataset with respectively: (a) the LLR model and (b) the LR model. The color of dots represents the proportion of gauging station and piezometers lost between the 2011-2017 database and the 1989-2017 database: losses < 50% (white); losses between 50% and 75% (grey); losses > 75% (black).

RPoD simulated with the 2011-2017 dataset:

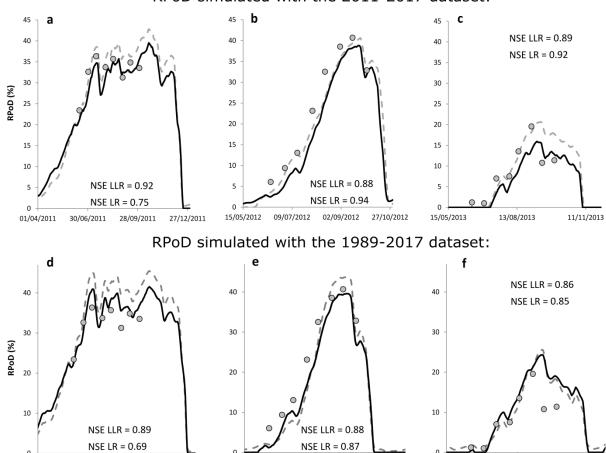


Figure 9. Comparison between observed proportion of drying RPoD_{POC} and RPoD predicted by the LLR and LR models with the 2011-2017 dataset in: (a) 2011, (b) 2012 (c) 2013 and with the 1989-2017 dataset in: (d) 2011, (e) 2012 (f) 2013.

13/08/2012

---- LR model

15/05/2012

LLR model

17/04/2011

790

791

792 793

794

16/07/2011

14/10/2011

11/11/2012

15/05/2013

Obs - POC data

11/11/2013

13/08/2013

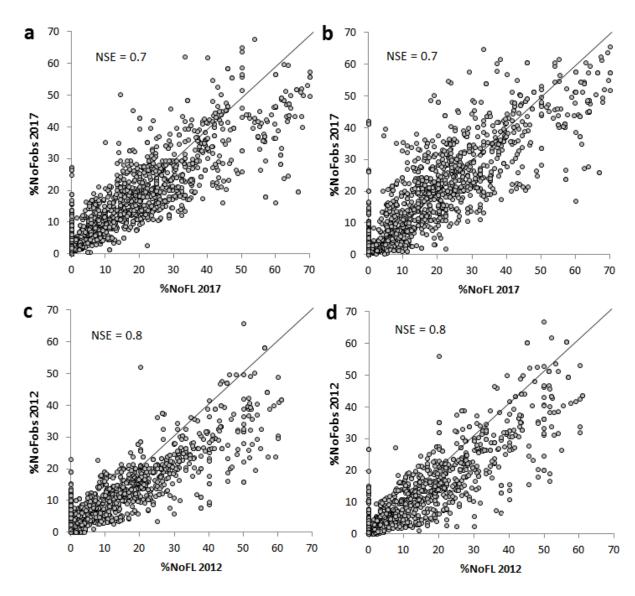


Figure 10. Scatter plot of the predicted RPoD (x axis) and drying observed at ONDE sites (y axis) in 2017 and 2012 simulated with the 2011-2017 dataset by: (a) and (c) the LLR model and (b) and (d) the LR model.

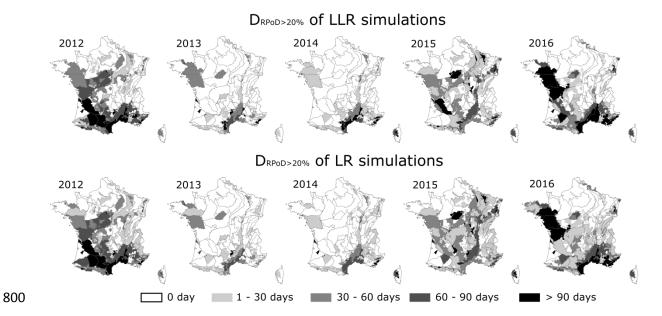


Figure 11. Maximum duration of consecutive days with RPoD higher than 20% simulated with LLR and LR model.

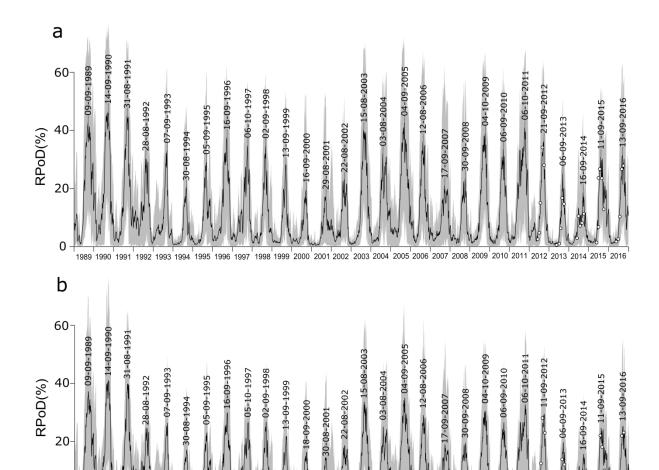


Figure 12. RPoD simulated between 1989 and 2016 the 1989-2017 dataset with: (a) the LR model and (b) the LLR model. The grey area represents the RPoD between the 90th percentile and the 10th percentile simulated on HER2-HR combination, the black curve represents the average RPoD simulated by HER2-HR combination and white dots represent the mean RPoD_{ONDE} for each observation dates. Dates mentioned correspond to the day of the maximum average RPoD simulated by HER2-HR combination (black curve) of each year.

1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015 2016

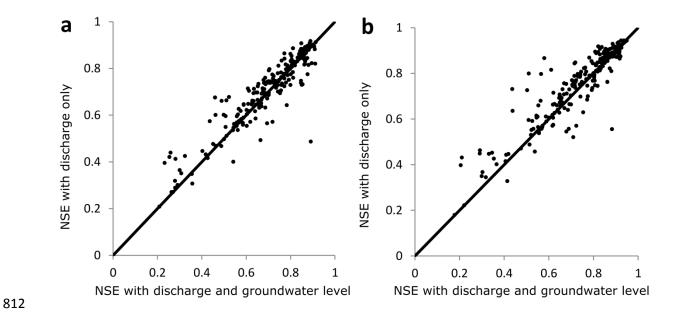


Figure 13. Comparison of NSE obtained with regression including only discharge variable as a function of NSE obtained with including discharge and groundwater level variables in the 2011-2017 dataset with: (a) LLR model and (b) LR model.

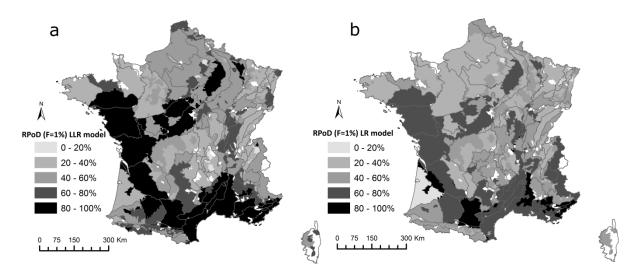


Figure 14. Regional probability of drying simulated with F = 1% predicted with: (a) the LLR model and (b) the LR model.