

Hydrogeological controls on spatial patterns of groundwater discharge in peatlands

Danielle K. Hare^{1,2}, David F. Boutt², William P. Clement², Christine E. Hatch²,
5 Glorianna Davenport³, Alex Hackman⁴

¹AECOM Technical Services, Rocky Hill, CT 06067

²Department of Geosciences, University of Massachusetts Amherst, 611 N. Pleasant St., Amherst, MA,
01003, USA

10 ³Living Observatory at Tidmarsh Farms, 139 Bartlett Road, Plymouth, MA, 02360, USA

⁴Massachusetts Division of Ecological Restoration, 251 Causeway St., Suite 400, Boston, MA 02114

Correspondence to: Danielle K. Hare (hare.danielle@gmail.com)

15 **Abstract.** Peatland environments provide important ecosystem services including water and carbon
storage, nutrient processing and retention, and wildlife habitat. However, these systems and the services
they provide have been degraded through historical anthropogenic agricultural conversion and dewatering
practices. Effective wetland restoration requires incorporating site hydrology and understanding
20 groundwater discharge spatial patterns. Groundwater discharge maintains wetland ecosystems by providing
relatively stable hydrologic conditions, nutrient inputs, and thermal buffering important for ecological
structure and function; however, a comprehensive site-specific evaluation is rarely feasible for such
resource-constrained projects. An improved process-based understanding of groundwater discharge in
peatlands may help guide ecological restoration design without the need for invasive methodologies and
detailed site-specific investigation.

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Here we examine a kettle-pond peatland in southeast Massachusetts historically modified for commercial
cranberry farming. During the time of our investigation, a large process-based ecological restoration project
was in the assessment and design phases. To gain insight into the drivers of site hydrology, we evaluated
the spatial patterning of groundwater discharge and the subsurface structure of the peatland complex using
30 heat-tracing methods and ground penetrating radar. Our results illustrate that two groundwater discharge
processes contribute to the peatland hydrologic system: diffuse lower-flux marginal matrix seepage; and,
discrete higher-flux preferential-flow-path seepage. Both types of groundwater discharge develop through
interactions with subsurface peatland basin structure, often where the basin slope is at a high angle to the
regional groundwater gradient. These field observations indicate strong correlation between subsurface
35 structures and surficial groundwater discharge. Understanding these general patterns may allow resource
managers to more efficiently predict and locate groundwater seepage, confirm these using remote sensing
technologies, and incorporate this information into restoration design for these critical ecosystems.

40 1 Introduction

Peatlands develop in response to physical, biological, and chemical processes and feedbacks. Groundwater discharge to surface water is one of the most important physical controls on peatlands stability (Siegel et al., 1995; Watters and Stanley, 2007); yet the underlying physical hydrogeologic framework governing the development of surface seepage distribution in these systems is not well understood. Preferential flow
45 paths, hydraulic conductivity (K) anisotropy, and geologic heterogeneities likely influence the surface expression of discharge zones (Chason and Siegel, 1986; Drexler et al., 1999; Smart et al., 2012). However, these variables have been difficult to constrain due to the spatial resolution of traditional localized groundwater wetland methods (wells, boreholes, surface point measurements, etc.), and their impact on fragile flow paths. The underlying hydrologic engine of these wetlands have shown to be difficult to
50 discern on large scale systems.

Thermal dynamics of ground and surface waters also govern critical wetland functions and can be assessed in multiple ways. Surface water thermal stability, for example, is a popular research focus in ecohydrology, as this process is important for aquatic species that rely on the low variance of groundwater
55 temperature to buffer themselves from heat extremes and regulate their metabolism (Caissie, 2006; Deitchman and Loheide II, 2012). Temperature also controls chemical processes in ecosystem respiration, which in turn controls carbon processing and nutrient retention (Boulton et al., 1998; Davidson and Janssens, 2006; Demars et al., 2011; Lafleur et al., 2005), biodiversity (Parish et al., 2008), and overall species health (Verberk et al., 2011). Upwelling zones are linked to increased biogeochemical cycling
60 (Sebestyen and Schneider, 2001), and also maintain species richness through the ‘edge effect’- overlap between the thermally and chemically stable groundwater ecotone and the higher oxygen environment within the main stream channel (Brunke and Gonser, 1997; Cirkel et al., 2010). An increase in wetland temperature has been shown to stimulate methane production (McKenzie et al., 2007) as well. The underlying drivers of the thermal regime of a wetland system can be caused through varying driving
65 processes, and are important to the ecosystem services provided in the peatland.

Widespread drainage of peatlands has caused wetland degradation and loss of ecosystem services. Anthropogenic modifications such as ditching and filling create discontinuity between surface water and groundwater systems with impacts to wetland function (van Loon et al., 2009). In some parts of the world,
70 wetland restoration is attempting to address these historical impacts. Within the United Kingdom, for

example, efforts to return natural water table levels by filling drainage ditches in peat mining areas have led to disagreements as to the cost-benefits of these specific restoration designs (Grand-Clement et al., 2013). In New England (United States), where thousands of acres of historical peatlands were converted to commercial cranberry farming in the late 1800s (Garrison and Fitzgerald, 2005), wetland restoration is similarly attempting to regain natural water table levels (Price et al., 2003). An incomplete understanding of underlying hydrology and thermal regime can limit the effectiveness of such efforts.

In this research, we explore the spatial distribution of groundwater seepage through a kettle-hole peatland from the analysis of basin structure and hydraulic properties of the peatland matrix. To assist in wetland restoration design at the study site, we focus on understanding the natural processes that promote the hydrologic inputs for aquatic habitat formation and maintenance. The goals of this study are to: (1) identify groundwater discharge locations and their hydrogeologic controls, (2) determine temperature dynamics of the groundwater discharge locations, and (3) evaluate the development of these seepage patterns. Through this work, insight is gained into the hydrologic driving mechanisms of peat-based wetlands support restoration of sustainable ecosystems (e.g. process-based design) (Beechie et al., 2010; Dahl et al., 2007).

1.1 Site Description

The site 'Tidmarsh Farms' is 3 cranberry farms, and the two largest farms are separated by Beaver Dam Road (Fig. 1). The area surrounding this peatland site is characterized by outwash, kame deltas and ground moraines that show evidence of collapse features/deformation (Larson, 1982; Stone et al., 2011). These ice collapse features are typical of environments proximal of ice contact zones and can result in the formation of kettle holes, which there is extensive evidence of throughout the surrounding region. All three of the site's cranberry farms were built on kettle hole peatlands between the late 1800's and early 1900's. The cranberry peatlands on Tidmarsh East were taken out of production in 2010 and another was taken out of production in 2015. This study concerns work on Tidmarsh East (the site), which is 2.5 km² (Fig.1B).

0.3-1.5m of sand overlays native surficial soils, which was laid down as a part of a normal cranberry farming practices until the site retirement in 2010. This practice maintained a very low gradient across the site with a slight decline to the north, with minimal micro-topography.

During this research, conducted in collaboration between Living Observatory and University of Massachusetts, a restoration project involving the private landowners, governmental agencies, and non-

governmental organizations was in assessment and design phases. Project planners were specifically interested in the location of groundwater discharge across the site to help design the placement of reconstructed stream channels. In addition, the restoration design team sought to better understand the location of subsurface peat deposits, underlying site hydrology, and potential future thermal regime when considering potential restoration activities. As of 2017, the site has undergone both passive and active restoration to encourage an accelerated ecological recovery based on the conclusions of this work. In the following sections, we document our methods and findings specific to the spatial distribution of the groundwater discharge at Tidmarsh East and the implications for restoration design.

1.2 Site Hydrology

The farm is a part of the small 5 km² surficial Beaver Dam Brook Watershed, but are also a discharge location of the 360 km² Plymouth-Carver-Kingston-Duxbury (PCKD) aquifer, thus the groundwater flowpaths contribute from a much larger hydrologic system than the surficial watershed (Fig. 1A). The PCKD aquifer below the site is characterized by glacial outwash sands (Masterson et al., 2009). Surface water enters the site from four surface water bodies south of the site (Fresh Pond, Little Island Pond, the Arm Wetland, and Beaver Dam Pond headwaters), and drains northward into Beaver Dam Brook, an approximately two km reach, before discharging in Bartlett Pond and then directly into Plymouth Bay (Fig. 1B).

To facilitate drainage and irrigation, lateral and perimeter drainage ditches exist throughout farmed areas. Parallel drainage ditches are located approximately every 18-35 m throughout the entire site, and are approximately 1 m wide and 0.5 m deep. The western agricultural cells have drainage ditches oriented east-west (Cell 3 and 4), and in the eastern cells (Cell 6 and 7) most drainage ditches are oriented north-south). When the study was conducted, the site was still predominately covered in low-lying cranberry vegetation, as well as a variety of sedges and cattails mostly adjacent to the central stream bank and marginal drainage ditches.

Flashboards in the dam creating the Beaver Dam Pond impoundment were permanently removed by the landowners in the fall of 2010, after which the southern side of the farm was allowed to return to a natural wetland state (Fig.1B). Data collection conducted for this research spanned 2012-2014, beginning two years after farming ceased, and prior to any active wetland restoration activity.

135 The site is located within the discharge zone of the large PCKD aquifer, and thus, short-term, drastic
temporal shifts are not expected in the hydrogeology or the processes described herein. We expect that our
observations from the study conducted over this 2-year period to be representative of present-day
conditions. The primary source of recharge to the PCKD aquifer is through precipitation which rapidly
infiltrates outwash plain deposits (Wareham and Carver Pitted Plains) (*Masterson, 2009*), thus changes in
140 the water table elevation can be expected. Topographic changes to base level due to isostatic rebound
and sea level rise may also contribute to water table elevation changes (*Oakley and Boothroyd, 2012*). The
regional aquifer may be sensitive to long term climatic changes (*Shuman et al., 2001; Newby et al., 2009*);
however, this question is outside the scope of this study.

145 **2 Methods**

Seepage patterns within peatlands have been difficult to constrain due to large site areas and complex,
dynamic substrates. At Tidmarsh Farms, we use multiple remote-sensing and direct-contact methods in this
environment to connect data from different scales into a process-based understanding of peatland
groundwater seepage. Ground penetrating radar (GPR) is used to evaluate the subsurface structure of the
150 peatland basin(s), and multiple thermal methods are used to locate and analyze surficial groundwater
seepage patterns. Stable water isotopes are used to describe dominant water sources supplying the seepage.

Traditional hydrogeologic methods were also implemented including well transects, seepage meters and
differential discharge gauging along Beaver Dam Brook. Figure 1B illustrates the location of field
155 measurements. Differential discharge gauging of surface water flow was performed at the site with a Marsh
McBirney Flo-Mate 2000. A low-profile seepage meter was used to quantify groundwater discharge in
accordance with the technique described by Rosenberry, 2008.

2.1 Resolving Subsurface Structure

160 GPR has been successfully used to characterize peatlands' physical structure and stratigraphy due to the
strong contrast between peat and the underlying aquifer geophysical properties (e.g. water content) (e.g.
Comas et al., 2005; Holden, 2004; Kettridge et al., 2008; Lowry et al., 2009; Slater and Reeve, 2002). The
GPR method relies on the transmission of electromagnetic (EM) waves through the subsurface then records
the time and amplitude of the returning signal (reflection) to image changes in the EM properties between

165 subsurface materials (Knight, 2001; Lowry et al., 2009). In August 2012 we collected common-offset
reflection data using Malå ProEx 100 MHz and 50 MHz antennas with a transmitter-receiver separation of
1 m and 2 m respectively. Here, we only use the 100 MHz data to generate interpolated maps of peat
thickness, as those data provide better resolution of the peat-sand interface given the electromagnetic (EM)
properties of the peat matrix and the depth of the structure of interest, which was 0-15m for this study site.
170 Nineteen GPR line surveys were completed; all surveys used 0.3-m trace spacing and ranged from 100 m to
1000 m in total length (Fig. 1B). The vertical resolution of the survey was 0.9 meters, based on the theory
that layers can only be resolved if their thickness is greater than a quarter-wavelength.

We applied a 150 MHz high-cut filter to remove the high frequency noise, and then a 100 ns automatic gain
175 control to compensate for signal loss with depth and distinguish deeper reflections by averaging over the
time window applied and adjusting the central signal strength with respect to that result. No topographic
adjustments were made, as there is negligible topographic variation both along the surveys and between the
surveys. The peat thickness was determined in each of the radargrams. Three characteristic radargrams are
shown in Figure 2.

180 We constrain the EM signal velocity through the peat for the GPR data analysis, and describe the peat's
structure with depth by collecting nine sediment cores (Fig. 1B) with a vibracorer (3) and hand corer (6).
Analysis of the cores demonstrated that the layered reflections observed in the radargrams were due to
variation in the degree of humification. We determined an average EM velocity of 0.036 m/ns (range =
185 0.030-0.040 m/ns) through the peat for the five full length cores that extended to the peat-sand interface.
This velocity range is consistent with other peatland GPR studies (0.033-0.040 m/ns) (Parsekian et al.,
2012). Using these data, a 3D interpolation of the peat-sand interface was created using kriging to estimate
the subsurface peat basin structure (Fig. 2). The second derivative of the maximum slope (profile curvature)
was calculated from the interpolated surface to identify changes in basal slope of the peat-sand interface,
190 and is shown in Figure 2.

2.2. Identifying locations of groundwater discharge to surface water using temperature

Heat can be used as a tracer to identify upwelling groundwater, as air temperature oscillations on diurnal
and annual timescales strongly influence surface waters, while deep (e.g. greater than approximately 10 m)
groundwater temperatures remain relatively constant through time (Anderson et al., 2005; Constantz,
195 2008). Local, shallow flow-paths can be more sensitive to climatic and seasonal changes in evaporation and
precipitation (Fraser et al., 2001; Kurylyk et al., 2014b; Menberg et al., 2014; Reeve et al., 2006), and may

not contribute to the thermal stability of aquatic systems to the same extent as deep (>10m), regional aquifers. This noted, during the thermal study periods, groundwater temperatures range from 10-11 °C in onsite wells below the peat.

200 **2.2.1. Fiber-Optic Distributed Temperature Sensing**

Raman spectra fiber-optic distributed temperature sensing (FO-DTS) is used for spatially extensive heat tracing in aquatic systems. Tyler et al., 2009 provides a thorough review of the details of the technology and calibration. DTS temperature data were collected with Sensor Tran Gemini HT control unit in dual-ended mode using AFL telecommunications umbilical fiber-optic cable. This FO-DTS unit allows for 1-m
205 spatial accuracy at 0.1°C precision over ~15 min integration times. Each FO-DTS deployment was operated for a minimum of 5 days to ensure multiple sufficiently strong diurnal oscillations were captured. 50-m-long calibration coils were maintained at a constant temperature with an ice-water slush bath and/or ambient bath and were compared to an independent Onset HOBO Water Temperature Pro v2 Data Logger (U22-001) (± 0.2 °C accuracy).

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In July and August of 2013 four FO-DTS deployments were installed, one within the drainage ditches of eastern peatland cells, and three within the western cells. We capitalize on the modified structure of the agricultural peatland surface, particularly the relatively evenly spaced drainage ditches, to thermally sample surface water in a distributed way which is not possible in more natural systems (e.g. Lowry et al., 2007).

215 The deployment sites were chosen based on previous infrared surveys (November 27st, 2012, discussed in Sect. 2.2.2), interviews with the farmer, and feasibility of installation. Each deployment ranged from 1000 m to 2500 m in length. Macrophyte growth was cleared during installation and continuously monitored through each deployment.

220 The arithmetic mean and standard deviation were calculated for each ~five-day time series of FO-DTS data at every meter along the fiber-optic cable to identify locations of groundwater seepage. These results can indicate the location and relative magnitude and permanence of groundwater discharge, which is not possible with other methods, such as TIR or temperature probes (Briggs et al., 2012; Hare et al., 2015; Sebok et al., 2013; Selker et al., 2006).

225 **2.2.2. Infrared Surveys**

Thermal infrared (TIR) cameras sense and quantify surface infrared (heat) radiation, and are increasingly being used to evaluate aquatic systems efficiently at large scales (Chen et al., 2009; Deitchman and

Loheide, 2009; Dugdale et al., 2016; Handcock et al., 2012; Hare et al., 2015), particularly at large sites, or sites where in-situ measurements are not possible. The hand-held TIR survey was conducted to both
230 expand the thermal survey and to compare this method to the FO-DTS data. We used a high-resolution forward-looking infrared camera (T640BX model FLIR, FLIR Systems, Inc.) with GPS and compass capabilities. The TIR method allowed for efficient spatial coverage, and allowed us to obtain thermal data unreachable with FO-DTS (Hare et al., 2015).

235 At Tidmarsh Farms East three TIR surveys were completed: July 30-31st 2013; March 21, 2014; and one reconnaissance survey on November 27th, 2012. The July survey was used to make comparisons to the FO-DTS data as it was taken during the same time period; the March survey was used to compare seasonal variability in seepage patterns. Surveys were conducted in the morning and evening to minimize reflection interference, and all temperature collection practices and considerations for this site are described in detail
240 in Hare et al. (2015). To create a spatial site map comprised of all TIR images, a single temperature (color-contoured pixel) from an aquatic point of interest was selected, and used to color an icon on the map. This allowed for georeferenced TIR data to be used quantitatively to evaluate seepage patterns by location. The relative magnitude of seepage rate is estimated based on how similar the observed temperature is to the regional groundwater temperatures.

245 **2.2.3. One-Dimensional Vertical Temperature Profiles**

The depth to which the surface diurnal temperature signal penetrates saturated near-surface sediments depends on the period of the signal, the fluid flow velocity and direction, and the physical properties of the fluid-saturated sediment (Goto et al., 2005; Hatch et al., 2006; Irvine et al., 2016; Stallman, 1965). With
250 depth, the diurnal heat signal variation decreases in amplitude and its shifts forward in time. Much of the heat transport not explained by pure conduction is attributable to advective fluxes, which can be solved for from thermal time-series at multiple depths using simple analytical solutions to the one dimensional heat transport equation with specified boundary conditions (Hatch et al., 2006; Rau et al., 2014; Schmidt et al., 2007; Silliman et al., 1995; Stallman, 1965).

255 We analyzed four 1D vertical temperature profiles to understand the vertical subsurface fluid flux patterns at the site. Maxim iButton temperature loggers (0.0625 °C resolution; 1°C accuracy) were attached to cavities drilled into a wooden dowel, and placed into the ground such that the logger locations were -2.5, -5.0, -10.0, -25.0 cm depth below the ground surface and +2.5 cm above the surface. We coated each iButton with silicon sealant to prevent leaking/sensor damage; however, a 25% sensor failure rate was still

260 experienced. A 10-minute sampling interval was used for a minimum of 7 days during July and August of
2013 for each temperature time series.

Installation locations chosen represented the two types of seepage observed with the FO-DTS, and 1D
vertical temperature data were collected synchronously with DTS deployments. Two additional control
265 deployments of 1D temperature profiles were installed within/below drainage ditches. We assume that
under low surficial flow conditions the system is at quasi-steady-state, allowing us to estimate (upward)
seepage flux from measured surface water, groundwater, and intermediate-depth temperatures using the
analytical solution to the heat transport equation derived by Turcotte and Schubert (1982) and modified by
Schmidt et al. (2007). A flux value was calculated for each collected data time step, and was averaged for
270 each profile for the final reported flux value. Flux values were calculated four times for each profile using
the range of peat porosity and range of thermal conductivity values. The thermal parameters utilized for
the 1D heat transport equation are shown on Table 2.

2.3 Assessment of environmental isotopes to infer groundwater flow paths

To trace the source of the groundwater flow paths contributing to discharge, we use $\delta^{18}\text{O}$ and $\delta^2\text{H}$ to
275 distinguish between local recharge (short flow paths) and regional recharge (long flow paths). The isotopic
composition ($\delta^2\text{H-H}_2\text{O}$, $\delta^{18}\text{O-H}_2\text{O}$) of hydrogen and oxygen of the water molecule was analyzed for water
samples collected from surface water (monthly), shallow ground water (seasonally), deep groundwater
(seasonally), groundwater seepage (August 2013) and pore waters (October 2013). The four pore water
samples were acquired through a manual press of samples from Russian peat cores 0-1 meter below the
280 ground surface, and subsequently filtered for analysis.

$\delta^2\text{H-H}_2\text{O}$ and $\delta^{18}\text{O-H}_2\text{O}$ was measured by wavelength scanned cavity ring-down spectrometry on un-
acidified samples with a Picarro L-1102i WS-CRDS analyzer (Picarro, Sunnyvale, CA). Samples were
vaporized at 110°C. International reference standards (IAEA, Vienna, Austria) were used to calibrate the
285 instrument to the VSMOW-VSLAP scale and working standards were used with each analytical run. Three
standards that isotopically bracket the sample values are run alternately with the samples. Secondary
laboratory reference waters (from Boulder, Colorado; Tallahassee, Florida; and Amherst, Massachusetts)
were calibrated with Greenland Ice Sheet Precipitation (GISP), Standard Light Antarctic Precipitation
(SLAP) and Vienna Standard Mean Ocean Water (VSMOW) from the IAEA. The isotopic composition
290 results use a rolling calibration, which calculates each samples error by the three standards run closest in

time to the sample. Long-term averages of internal laboratory standard analytical results yield an instrumental precision of 0.51 ‰ for $\delta^2\text{H-H}_2\text{O}$ and 0.08 ‰ for $\delta^{18}\text{O-H}_2\text{O}$.

295 The USGS wells were sampled for groundwater isotopic compositions within the PCKD aquifer, providing regional groundwater values for the aquifer and defining the expected annual range of isotopic values for local precipitation (Table 1). The regional groundwater trend line was generated by fitting a linear regression through the USGS well isotope data from the regional PCKD aquifer.

3 Results

300 As an initial evaluation of the groundwater contribution to the site, we conducted differential discharge gauging measurements on September 15th, 2013. The locations of these measurements are indicated by the purple circles on Figure 1B. The stream gained 6 L s⁻¹ discharge through Cell 7 from the Arm Pond input to the confluence with Beaver Dam Brook (1.5 km), equal to an average of 0.004 L s⁻¹ per m of river length (Fig. 1B). Cell 3 and 4 gained 113 L s⁻¹ from the Beaver Dam Pond input to the confluence with the East side river (1 km), equal to an average of 0.113 L s⁻¹ per m of river length. At
305 other wetland sites seepage flux magnitudes and directions have shown to be temporally transient (*Fraser et al.*, 2001; *Sebestyen and Schneider*, 2001); however due to the consistent high hydraulic gradient in the regional aquifer and the small watershed, we assume that temporal dynamics are insignificant within our data set and sufficiently static to describe the present day conditions. This assumption is supported by the two seasonally distinct infrared surveys resulting with similar seepage distribution results.

310 3.1 Resolving Peatland Basin Structure

The interpolation of the basal surface, or the peat-sand contact beneath the peat from GPR data, indicates four isolated peat depressions at the site, two depressions in Cell 6 and Cell 7 and two Cell 3 and Cell 4. Cells 6 and 7 have a maximum peat thickness of ~7 m and a gradual curvature of the peat-sand interface than the western cells, Cell 3 and Cell 4 (Fig. 2). The western cells show a maximum peat thickness of ~10
315 m, and relatively high curvature values. The basin structure of the western cells is also more complex than Cell 6 and 7, as Cell 3 and Cell 4 have pronounced undulations in the basal peat-sand contact surface, creating dramatic changes in basin shape. Particularly, there is a notably high curvature of the basal peat-sand interface along the western edge approximately 30 m from the margin. The GPR profiles illustrate multiple series of normal faults beneath the peat body that are consistent with ice melt-out and/or collapse
320 features (Fig. 2C) typical of kettle pond origin (*Kruger et al.*, 2009).

3.2 Thermal evaluation of groundwater seepage

Surface water temperatures in the main channel and ambient drainage ditch environments generally show high standard deviation, indicative of a coupling between these surface waters and air temperatures, and mean water temperatures closely tied to the seasonal surface temperature average, also indicative of surface water dominance. FO-DTS surveys were designed to detect low standard deviation and consistent mean temperature anomalies from these background conditions, which is indicative of groundwater inflows. The temperature results of both these surveys are presented in Hare et al. (2015). Results from both TIR and FO-DTS identified two categories of thermal anomalies: type 1 anomalies manifest as temperatures with relatively low standard deviation through time, and an anomalous heat signature that is seasonally warmer or cooler than regional groundwater temperature by approximately $\pm 3-5$ °C; and type 2 thermal anomalies also have a low standard deviation, but temperatures more closely resemble regional groundwater temperatures (10-11 °C). Figure 3 shows time series data collected with the FO-DTS and illustrates each of the major thermal signatures shown on site: temperatures of groundwater, the main channel, a drainage ditch, and the two thermal anomalies. We interpret these two anomalies to correspond to two modes of seepage, type 1 thermal anomalies correspond to matrix seepage, and type 2 thermal anomalies correspond to preferential flow path (PFP) seepage. The two seepage types are clearly differentiated through thermal signatures, and can be isolated using the average and standard deviation of temperatures with time. The TIR surveys also revealed these two distinct types of seepage, which were present in both the summer and winter surveys (Fig. 4).

TIR surveys and FO-DTS data indicate that most groundwater input likely occurs along the western edge of the Cell 3 and Cell 4, where peat is thinner or where there is strong sand/peat contact curvature in peat basin shape (Fig. 5). Isolated locations of consistent temperatures similar to groundwater temperatures and anomalously low standard deviations exist along the linear location of highest peat/sand contact curvature near the western edge of the cells, as well as along edge areas with the thinnest peat. The isolated, unique locations of PFP seepage that occur within the deeper peat represent a distinct seepage process from matrix seepage and PFP seeps along the edge of the peat.

During the March infrared survey, a high density of ~1-5 cm diameter flowing macropores within the peat was discovered in Cell 3. The water discharging from these macropores exhibited groundwater seepage temperatures (Fig. 6), and led us to term this mode of PFP seepage. This observation is similar to the peat macropores or 'peat pipes' described in previous peatland research (e.g. Briggs et al., 2016; Cunliffe et al., 2013; Holden, 2004; Smart et al., 2012; Vandenbohede et al., 2014), but the concentration of macropores in

355 this singular location makes the northwest cell macropores observation unique. We measured high 3.0 L
min⁻¹ flux PFP seeps with a seepage meter. Despite the very few locations of PFPs, their high fluxes have
the potential to contribute significantly to the groundwater gain across the site (Poulsen et al., 2015). The
peat thickness map (Fig. 5) indicates that the zone of high macropore density is an area of peat thinning
reaching a minimum peat thickness of 3 m, and also a location of high curvature (center of cell 3). Rossi et
al. (2012) describes similar correlation to peat thinning at a site in Finland.

360 3.3 1D vertical temperature profiles

The two seepage types and two ambient drainage ditch locations were monitored with 1D vertical
temperature profiles for seven to ten days. We expected to observe significant upwelling at this site, which
we could easily identify by a rapid attenuation of the diurnal signal with depth coupled with a characteristic
convex upward shape of mean temperature with depth (e.g. Schmidt et al., 2007). Temperatures from all
365 four 1D vertical temperature profiles are distinct from one another; however, all the temperature profiles,
including the “ambient” drainage ditches, are consistent with upwelling of groundwater (convex upward
shape of mean temperature with depth in Fig.7). The surface temperature of the ambient drainage ditches
(temperature profiles 3 and 4) is similar to the diurnal temperature cycles measured with FO-DTS, and
were used as background data for the heat signature of the site. The 1D fluid flux calculations of the
370 temperature time series of the two drainage ditch locations yielded a range of -0.028 to -0.031 m d⁻¹ and -
0.067 to -0.074 m d⁻¹.

Temperature profiler 1 was installed at a location with a surficial temperature of 13-14 °C in August 2013.
The total peat thickness at this location is 50 cm, and consistent with groundwater upwelling, minimal
375 diurnal signal propagates to depth, and surface water exhibits relatively low variance in temperatures over
time. Thermal time series estimates of flux show a modest -0.146 to -0.163 m d⁻¹ upwelling through the
peat at this seepage location.

Finally, temperature profiler 2 was installed in a location with a surficial temperature consistent with
380 groundwater temperatures of 10-11 °C in August 2013, and temperatures with depth exhibit a groundwater
thermal signal throughout the entire profile. Even close to the bed interface, the streambed thermistor (2.5
cm) shows slight thermal shifts ($\sigma = 0.096$ °C), which are near to the resolution of the instrument (0.0625
°C). This unique temperature profile is indicative of high upward flux rates, as the diurnal signal cannot be
resolved and there is essentially no downward conduction from above; therefore, we were unable to use the
385 steady state analytical solution to estimate a flux rate. However, in July 2015, we deployed a seepage meter

at this location and measured fluxes in excess of 3 m d^{-1} , rates which exceed the limits for analytical flux calculations.

3.4 Groundwater Discharge Source Areas

390 Groundwater discharge to the wetland complex is a mixture of shallow and deep regional flow
paths. Isotopic analyses of waters from wells in the up-gradient portion of the PCKD aquifer (blue circles
in Figure 8) fall along a regional groundwater trend line. We interpret this regional trend line to be
characteristic of the annual isotopic composition of recharge water to the region as well as local
groundwater recharge in the topographic watershed of Tidmarsh. These upgradient groundwater isotopic
values plot left of the global meteoric water line (GMWL)(Craig, 1961), which reflects local and regional
395 vapor recycling and a characteristic mixture of vapor sources (Koster et al., 1993). The one exception to
this line is the USGS well MA-PWW 494 in Plymouth, MA which is similar to Tidmarsh in that it is
downgradient of the recharge area of the PCKD aquifer. This water falls to the right of the regional
groundwater trend line. Discharging and shallow groundwaters at the wetland site plot close to but off of
the regional groundwater trend line. The blue diamonds (Fig. 8) represent a monthly sampling of wetland
400 surface waters that depict a significant clustering to the right of the regional groundwater trend and evolve
along a line tangent to this intersecting the deep TM groundwater. Uncharacteristically, the deepest
sampled groundwater at the site ($>15 \text{ m}$) falls to the right of the GMWL (orange circle), suggesting this
water has experienced a significant enrichment in the heavy isotopes due to evaporation
processes. Repeated sampling of this water reveals a consistent isotopic composition that suggests the deep
405 groundwater beneath Tidmarsh is isotopically enriched due to evaporation from open water bodies in
upgradient kettle ponds. The headwater seepage area and a the strong discharge seepage area (large pink
and red triangles in Figure 8) in the interior of the wetland complex fall along a line that represents either a
mixture of this evaporated water and the regional groundwater trend (finely dashed line) or itself is simply
an evaporatively evolved water. Both interpretations suggest that the source of water to the shallow
410 groundwater wells and the large volume springs in the interior of the wetland complex are distinct. This
indicates that the local flow path from the southwest to the northeast is the large-scale hydraulic gradient
that dominates the observed seepage patterns. The orientation of peatland basin slope break and the
regional groundwater gradient also intercept the southwest corner of the peatland where numerous high-
flux groundwater seeps are located.

415 4 Discussion

4.1 Groundwater discharge types

Two types of groundwater discharge (or seepage) were identified using thermal methods, as detailed in Section 3.2. PFP discharge areas that have regional groundwater temperature (e.g. 10-11 °C). Matrix seepage locations are groundwater discharge with temperatures that are offset ($\pm 3-5^\circ\text{C}$) from regional
420 groundwater temperature, but have very low variance compared to expected diurnal variations and are also significantly distinct from local surface water temperatures. Both seepage types appear to strongly buffer stream temperatures, illustrated by low variance when examined through time (FO-DTS data). A low variance could have also been caused mobile sediment (Sebok et al., 2015); however, within this peatland environment this process is not expected, nor was observed. The identification of these two distinct seepage
425 types using multiple methods and during distinct seasons indicates different mechanisms for generation of each of these seepage patterns. Figure 5 combines both matrix and PFP seepage observed with either FO-DTS or TIR to evaluate spatial patterning and consistencies, and shows how the two types are related to one another as well as to patterns of high basal curvature.

430 Consistent (low standard deviation) and groundwater-like temperatures (10-11°C) of the PFP seepage indicate very high flux ($>3 \text{ m d}^{-1}$ was confirmed with seepage meter measurements). Given the low vertical K of peat matrices, sustaining such high fluxes would require seemingly implausible hydraulic gradients, certainly far above the vertical hydraulic gradients observed on site. Therefore, it is highly likely that this seepage does not occur as flow through the peat matrix, but instead focused, high discharge, conduit flow,
435 consistent with “short-circuit discharges” described by Conant Jr. (2004). Focused flow in conduits through the peat was observed in the field at Tidmarsh Farms (Fig. 6), and by Briggs et al. (2016), and has been documented through visual descriptions of peat pipes, or macropores at other locations (Baird, 1997; Beckwith et al., 2003; Cunliffe et al., 2013; Holden, 2004; Smart et al., 2012; Wallage and Holden, 2011). However, the spatial extent of these preferential flow zones has not been previously demonstrated. Due to
440 their high flux, physical isolation, and focused nature, we refer to this type of seepage as PFP seeps.

Data represented by matrix seepage show that surface water diurnal temperatures are also buffered in these zones and are distinct from most ambient surface temperatures. This observation could indicate shallow aquifer groundwater discharge, which is more influenced by atmospheric temperatures than deeper regional
445 flow (Kurylyk et al., 2014a; Menberg et al., 2014). However, consistent temperatures in the site’s shallow groundwater wells and 1D temperature profiles indicate that these seepage temperatures are controlled by a

lower flux rather than distinct atmospheric-influenced shallow flow paths. These matrix seeps indicate that while vertical upwelling fluxes are present, they are much smaller than PFP discharge zones, and must be controlled by a different mechanism. Thermal profilers yielded vertical flux rates consistent with a low to moderate upwelling through porous media according to *Conant Jr. (2004)*; which would be typical of the hydraulic properties associated with peat, and thus, which is the reason we refer to locations with this signature as “matrix” seeps. The two seepage types, PFP and Matrix seepage, are similar to the “point” and “diffuse” peat seepage categories defined by *Rossi et al. (2012)*, but rather than focusing on the area of influence, instead highlight the physical structure that governs the process which ultimately generates seepage in these peatland seepage zones.

4.2 Subsurface Structural Control on the Spatial Distribution of Seepage Types

Matrix seeps were plentiful within approximately 30 m of the peatland edge (Fig. 5), consistent with margin seepage observed in lake environments (*Rosenberry et al., 2010; Sebestyen and Schneider, 2004; Sebok et al., 2013; Winter, 2001*) and other wetlands (*Freeze, 1988; Labaugh et al., 1998*). The peat is 0.1-3.0 m thick along the margin where matrix seepage occurs (Fig. 3), which is generally significantly thinner than locations of observed interior PFP seepage. Matrix seeps generally occur in the thinnest peat zones, and typically decrease rapidly with distance from the peatland edge toward the interior slope change, after which no thermally distinct groundwater discharge points are observed (Fig. 5). While evidence for PFP seepage does occur as well in these shallow areas, matrix seepage is more consistent within this shallow peat environment. This is shown as a conceptual model in Figure 9, based on temperature data collected proximal to GPR line 7.1 (radargram shown in Fig.2C). Similar landscape-scale observations have been made within lakes and wetlands (e.g. *Cherkauer and Zager, 1989; Sebok et al., 2013*), and as kettle hole peatlands typically form from initially open water bodies, there are logical similarities in basic processes between the two environments.

Discrete seepage zones may reflect zones of higher effective K than the surrounding peat matrix, which could be explained by littoral-zone migration in the lake to wetland evolution as the water table fluctuates and migrates. In lake environments, diffuse matrix seepage occurs because of an increase in K at the edge of the lake caused by “erosional deposition,” whereby focused wave and current action disrupt and erode sediments, particularly mobilizing the finest sediments elsewhere, and concentrating larger particles, indicative of these higher-energy environments in these locations. Preferentially stronger flow paths are thus concentrated at the break in land surface slope (*Blume et al., 2013; Casson et al., 2010; Cherkauer and McKereghan, 1991; McBride and Pfannkuch, 1975; Rosenberry et al., 2010; Winter, 1981*). Previous work

proposes that seepage flux decreases exponentially with distance from shore of a lake (Cherkauer and Zager, 1989; McBride and Pfannkuch, 1975), which is qualitatively confirmed by our data. Paleoclimate
480 reconstructions have demonstrated that the regional water table around Tidmarsh has been increasing in elevation since the Laurentide ice sheet retreated ~10 ka ago, with 2-3 significant low stands (Newby et al., 2000, 2009). Therefore, we hypothesize that the extent of the matrix seepage observed along the western edge of the peatland is a result of this lake transgression and coincident decrease in deposition of organic material. Here the lower K of the peat matrix intersects with shallow groundwater flowpaths, strongly
485 affecting lateral hydraulic gradients and driving upward flux; a process which likely generates much of the observed matrix seepage (Fig.9). This observation is supported by similar seepage processes observed in riverine systems (Sophocleous, 2002), wetland (Larsen et al., 2007), lake (Bakker and Anderson, 2002; Winter, 1981) and hillslope environments (Shaw et al., 2017; Winter et al., 1998).

In contrast to the matrix seepage, PFP seepage was less common and spatially disconnected from similar
490 flux seeps (Fig. 5). Similar to matrix seepage, PFP seepage exhibits low standard deviation of temperature (Fig. 3), but PFP seep temperatures were much closer to average regional groundwater temperature. This indicates that PFP seepage waters have very short residence times within peatland sediments, which may have important implications for nutrient transformations within them. At some PFP seeps the peat is generally thicker and located more toward the interior of the peatland rather than along the margin where
495 matrix seepage zones are found in addition to being found between the peatland edge and the area with high basal curvature values (Fig. 5 and Fig.9). Typical interior PFP flow path lengths from the sandy aquifer below the peat to the surface should be much greater than for matrix seeps; however, the thermal signature seems to contradict this; therefore, PFP seepage zones must be generated through a unique hydraulic process from matrix seeps. Since PFP seeps at Tidmarsh Farms correlate with significant slope
500 changes, or locations of high curvature, these isolated seepage zones must be generated by an abrupt change in horizontal K, and the PFP seep locations closer to the edge may be a result of zones of inherent matrix weaknesses such as varying degrees of humification caused by vegetative difference and water level, or other disruptions in the peat matrix including plant rooting and desiccation 'cracks' as proposed by (Smart et al., 2012) (Fig.9).

505 An abrupt change from high to low K has long been known to promote the transition from horizontal to vertical flow (Freeze and Witherspoon, 1967). Lowry et al., (2009) hypothesized this process to explain developed seepage within the interior of a peatland through using 3D numerical groundwater flow models. As horizontally flowing regional groundwater encounters a low-conductivity peatland, it is forced to go through or around it, causing pressure to increase where the abrupt change in the K from the sand to

510 catotelm peat matrix occurs (Fig.9). PFP seeps develop as a fast-pathway to the surface, a pressure-relief valve, where these localized increases in aquifer pressure at the base of the peat matrix translate into strong, sustained discharge of unaltered regional groundwater to the surface.

Rosenberry et al. (2010) note that in lake bottoms, a significant upward seepage velocity can maintain a locally high K as the upward force may suspend smaller particles within the water column. Particulate
515 organic matter and lacustrine sediment have a very low settling velocity, therefore if the upward force that groundwater seepage induces is greater than the settling velocity, only organic matter with a high mass will be able to accumulate over these lake seepage locations. This would cause the peat matrix to have a relatively high porosity and a high permeability compared to its surrounding very low permeability matrix. These locations will continue to be zones of weakness through the formation of the peatland. Thus, we
520 hypothesize that high-flux PFP seepage zones persist through the transition from lake to peatland environment due to the inability of fine sediments and organic matter to accumulate over these high flux locations. Still, these locations of consistently high hydraulic gradient will also continually take advantage of inherent matrix weaknesses.. However, the underlying mechanics of PFP seepage in the interior/deeper peat are caused by the interception of the regional groundwater gradient and high curvature peat subsurface
525 structure (Fig.9).

The orientation of peatland basin slope break (high basin profile curvature) and the southwest to northeast regional groundwater gradient dictates the observed pattern of strong seepage along the western boundary, which is supported by isotopic analysis. PFP and matrix seep waters both exhibit isotopic signatures consistent with a mixture of local groundwater and regional recharge signature (Fig. 8). This observation is
530 further reinforced by the increase in net groundwater gain through the western cells, as well as a large number of PFP seeps in the southwestern portion of the site (Fig. 5).

5 Conclusion

Subsurface basin shape exhibits significant control on the spatial distribution of groundwater discharge within peatland environments. As horizontal groundwater flow intercepts the peat matrix, two types of
535 seepage develop: matrix and preferential flow path seepage. Matrix seepage is defined by a low standard deviation in temperature and surface temperature similar to groundwater $\pm 3-5$ °C, consistent with relatively low-flux seepage. Low fluxes are produced where the regional groundwater flow paths intercept the low-K peat at the basin 'shoreline', inducing upward flow through relatively thin (0.1-3.0 m) peat. The second type of observed discharge, PFP seepage, has a surface temperature essentially indistinguishable from deep

540 regional groundwater temperature. This indicates very strong upwelling fluxes at these locations and little
time for conductive heat losses/gains. Locations of PFP seeps appear along the periphery of the peatland,
but more notably also correlate with high rates of basal peat slope change (curvature) of the peat basin
(Fig.9). These seeps develop where the regional groundwater flow path intercepts a secondary slope change
and where there is a stark change in K between the high-K sand aquifer material and the low-K peat.
545 Together, these physical features generate large pressures, induce localized zones of high vertical hydraulic
gradient and drive large seepage fluxes upward. Because PFP seeps occur typically in locations with
thicker peat and yet maintain close to groundwater temperatures, they must have a much higher vertical
hydraulic gradient and/or higher effective K than the matrix seeps. Through multiple lines of evidence, we
conclude that the development and spatial distribution of minerotrophic peatland seepage is strongly
550 controlled by the interaction between the subsurface basin structure, physical process within the peat
structure and hydraulic gradient.

Through our results, we establish a predictable pattern of seepage, consistent across the coastal site that is
explained by knowledge of basin shape and regional hydraulic gradient. This information provides
555 valuable insight for water resource managers to better understand the natural forces driving groundwater
seepage. This knowledge, in turn, may be used in the restoration design of degraded peatland systems.
Knowing where seepage is expected to occur naturally across a site allows for the development of more
sustainable restoration designs that work with the land, and not against it. In retired cranberry farms, for
example, channels may be relocated to intercept springs to maintain cooler water temperatures. This
560 knowledge can also guide the location of targeted intensive grading. For example, as was done at Tidmarsh
Farms, the dense cranberry mat can be broken up mechanically to encourage groundwater expression on
former dry farm surfaces and access native seed banks below. Incorporating this data into a restoration
design will greatly aid the ability to predict and achieve desired ecosystem outcomes, making restoration
projects more efficient, both ecologically and monetarily.

565 This research provides a process-based investigation of the subsurface hydrodynamics within a peatland.
While a peat matrix exhibit strongly heterogeneous and anisotropic tendencies, large-scale patterns occur
and can be predicted. These patterns are dependent on basin shape, peat accumulation history, and
underlying aquifer flow paths. The importance of groundwater flow paths surrounding the peatland and
570 resulting seepage patterns emphasizes that peatlands are not isolated entities from the groundwater system
and cannot be treated as such. Observed large-scale seepage patterning provides insight that may help
explain vegetation patterning, macropore development, and other localized peat dynamics that have been

unidentified in the past, and greatly aid peatland management and restoration to establish more naturally sustainable, efficient practices.

575 **Disclaimer**

The views and opinions expressed in this article are those of the authors and do not necessarily reflect the official policy or position of AECOM Technical Services.

Acknowledgments

This research would like to acknowledge the support of Evan Shulman for site access, field support and
580 good humor, Henry Eshbaugh, Eric Van Dam, Steve Hurley and the many volunteers who support the data
collection efforts site co-ordination. Much gratitude is given to John Lane and Martin Briggs of the U.S.
Geological Survey, Branch of Hydrogeophysics for equipment rental and technical support. Also, we thank
Jon Woodruff of the University of Massachusetts, Amherst for use of his laboratory equipment, data
processing software, student support and technical guidance.

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USGS Well ID	$\delta^{18}\text{O}$	$\delta^2\text{H}$	Latitude (WGS 84)	Longitude (WGS 84)
MA-PWW 494, Plymouth	-6.15	-37.38	41.8713889	-70.6586111
MA-EBW 30, East Bridgewater	-8.07	-46.33	42.0155556	-70.9658333
MA-WFW 51 Wareham	-6.83	-37.45	41.7550000	-70.7325000
MA-D4W 80 Duxbury	-8.13	-47.42	42.0547222	-70.7247222
MA-XGW 2, Weymouth	-8.55	-50.33	42.1650000	-70.9458333
MA-NGW 116, New Bedford	-7.55	-43.42	41.6736111	-70.9577778
MA-F3W 23, Freetown	-7.86	-44.02	41.7847222	-71.0813889

Table 1: USGS groundwater wells $\delta^2\text{H-H}_2\text{O}$, $\delta^{18}\text{O-H}_2\text{O}$ isotopic data used to establish the regional groundwater trend.

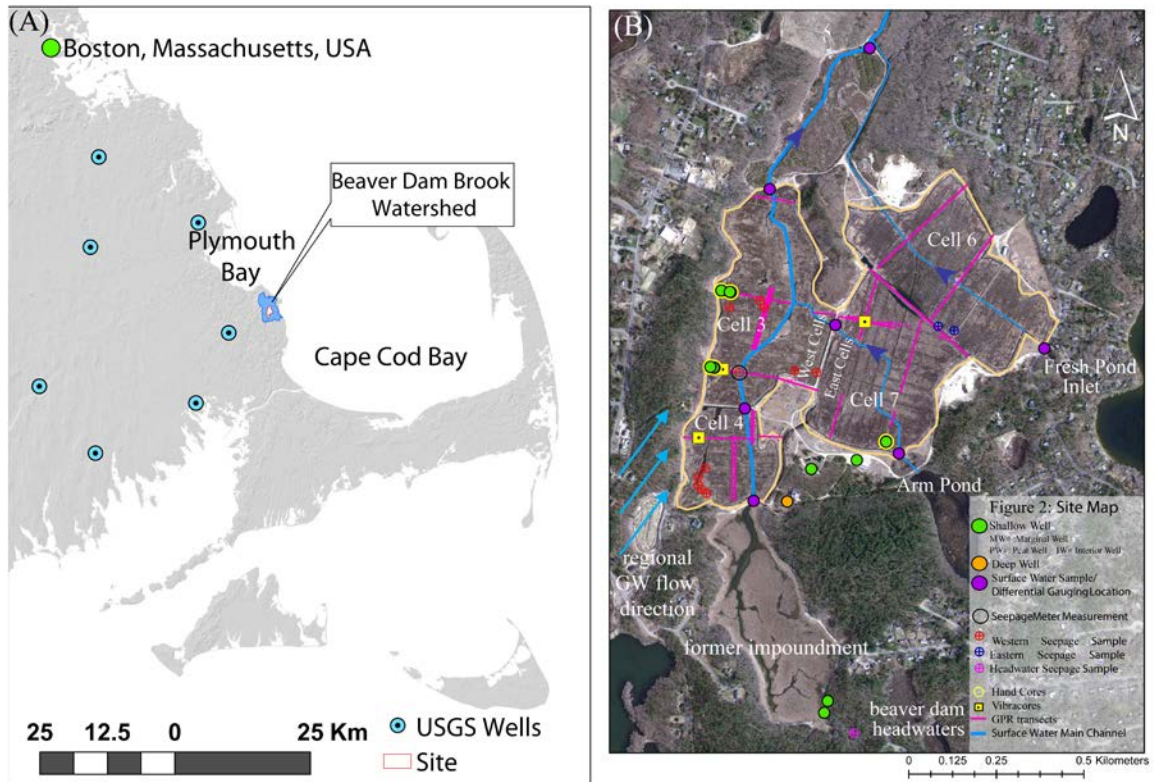
840

K_s [$J s^{-1} m^{-1} K^{-1}$]	$0.4^a-0.6^b$
K_f [$J s^{-1} m^{-1} K^{-1}$]	0.6^c
n	$0.5^a-0.8^d$
ρ_f at $10^\circ C$ [$kg m^{-3}$]	999.7
c_f at $10^\circ C$ [$kJ kg^{-1} K^{-1}$]	4193

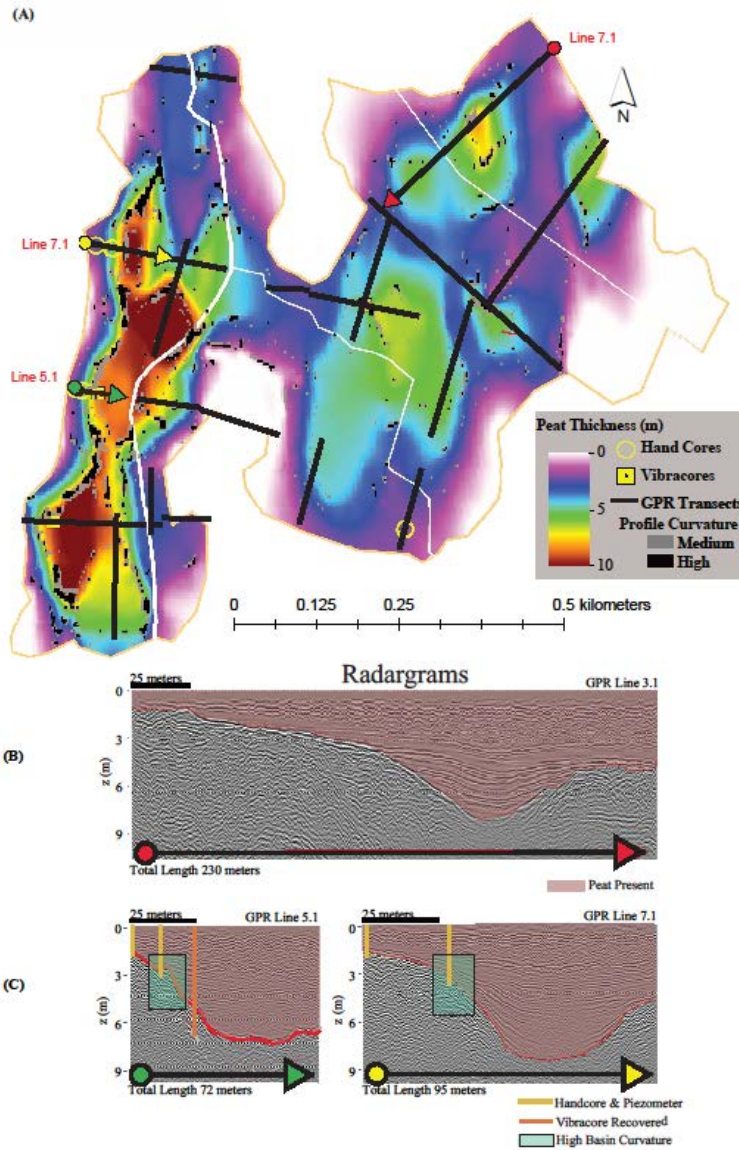
^a McKenzie et al., 2007 ^bLetts et al., 2000; ^c Schmidt et al., 2007
^dRezanezhad et al. 2016

845

Table 2: Parameters within the 1D heat transport equation derived by Turcotte and Schubert (1982) and modified by Schmidt et al. (2007). K_s is the thermal conductivity of the solid, K_f is the thermal conductivity of the fluid, and n is the porosity of the matrix. The density of the fluid and heat capacity of the fluid multiplied together are the volumetric heat capacity of the fluid ($\rho_f c_f$, $J m^{-3} K^{-1}$).



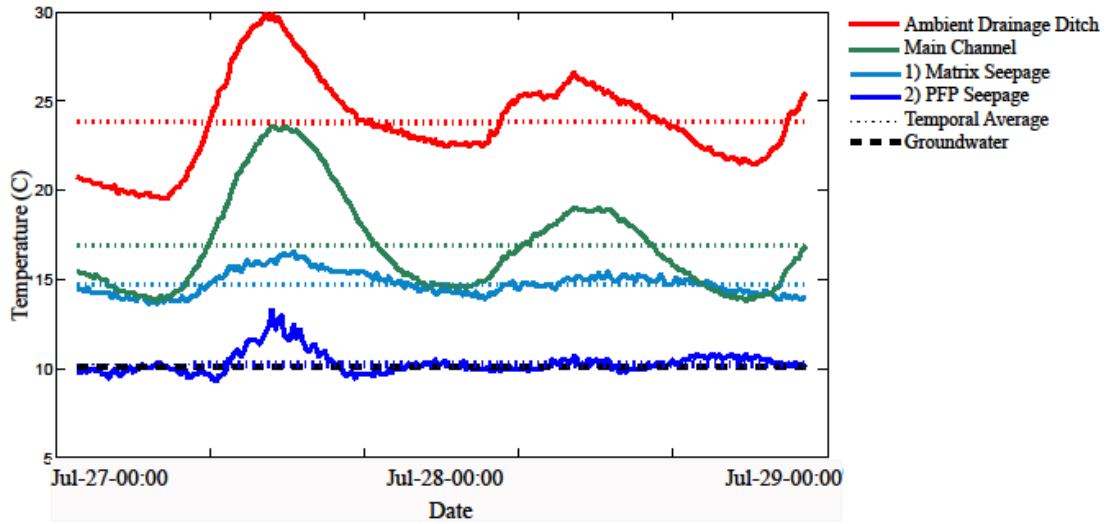
850 **Figure 1: a) Site map of the Tidmarsh Farms regional peatland showing the study area and watershed boundary; Plymouth County, Massachusetts, and PCKD USGS wells used for regional groundwater isotopic data (Table 1). b) Detail of the Tidmarsh Farms study site showing the major waterways and flow direction in blue, site groundwater wells, isotopic sample locations, and GPR transects. Beaver Dam Brook flows north into Plymouth Bay.**



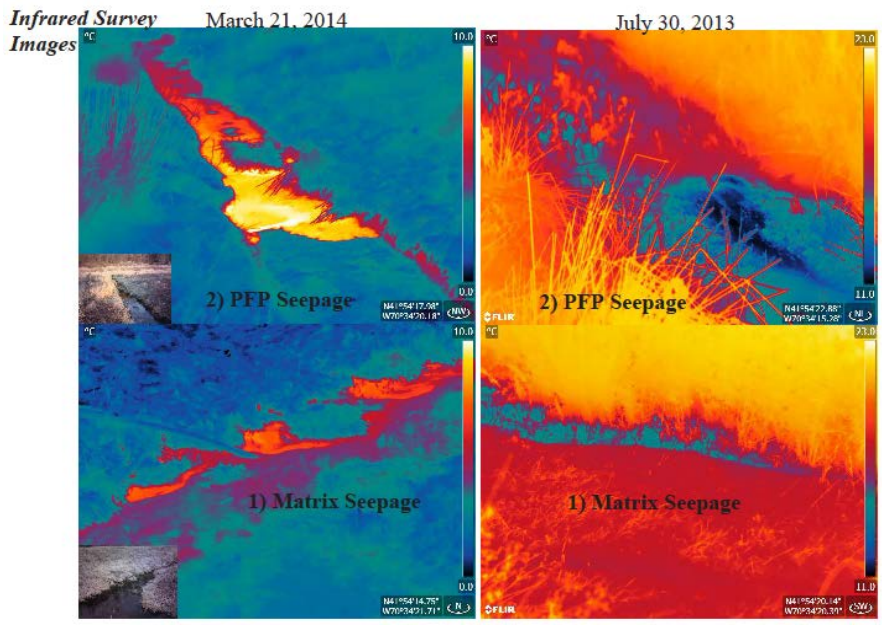
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Figure 2: a) Map of total peat thickness beneath Tidmarsh Farms based on GPR data. GPR data collected along linear transects shown here (black lines; pink lines on Figure 1B) were interpolated and contoured to show peat thickness (colors) on the 2D surface map. Zones of medium and high curvature (the 2nd derivative of the thickness) of the peat-sand interface are shown as grey and black pixels, respectively. b) and c) Three example cross sectional profiles, or radargrams, illustrate a distinct reflector at the basal peat-sand contact. Peat is shaded red. Sediment cores shown as yellow lines (hand cores) and orange lines (vibracore) were used to constrain the GPR velocity data. High curvature is highlighted in green boxes.

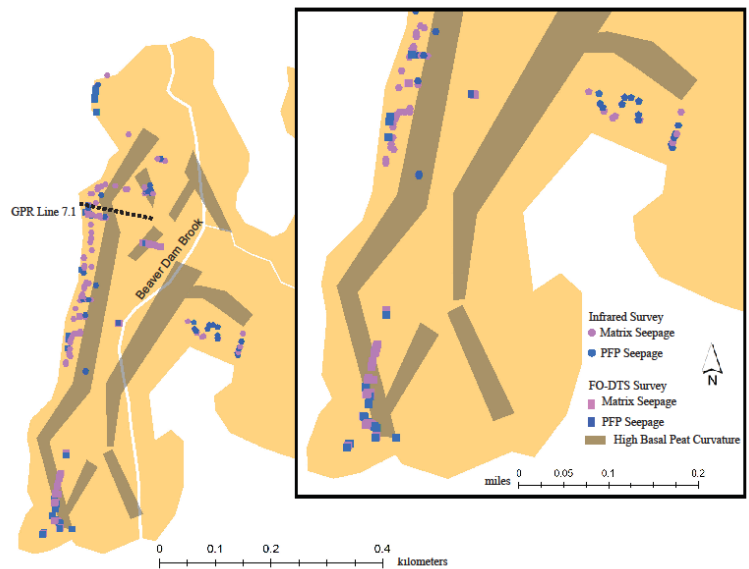


865 Figure 3: Fiber-optic Distributed Temperature Sensing (FO-DTS) temperature time series from four 1-m
 segments of cable to illustrate the characteristic thermal signatures at Tidmarsh Farms. The greatest amplitude
 and variability occurs in the drainage ditches with little flow and significant solar heating (red), followed by the
 main channel of Beaver Dam Brook (green). Two seepage types are also plotted over 2.5 days: matrix (type 1)
 seepage, with very low variability (low standard deviation) over time and a mean temperature a few degrees
 870 higher than groundwater (light blue) and preferential flow path (type 2) seepage with a mean temperature
 nearly equal to groundwater (dark blue).



875 **Figure 4: Thermal infrared (TIR) images recorded in July 30-31, 2013 (Summer) and March 21, 2014 (Winter) at Tidmarsh Farms. Visible light images are shown in the bottom left of March images, but not July, as these surveys were conducted at night to limit issues associated with reflectance. TIR images illustrate the two types of seepage in both seasons: type 1 preferential flow path seepage that is characterized by discrete discharge points very close to groundwater temperature with high-flux, and type 2 matrix seeps that are diffuse, 3-5 °C warmer or cooler than groundwater and lower flux.**

880



885 **Figure 5: Map of seepage at Tidmarsh Farms determined with fiber-optic distributed temperature sensing (FO-**
DTS, squares) and thermal infrared (TIR) surveys (circles). Background shaded region(s) match the bounded
area from Figure 1B, and darker background shading delineates zones of high curvature (the 2nd derivative of
the thickness) of the peat-sand interface (Fig. 2). For both methods, light purple to pink symbols indicate matrix
(type 1) seepage, and dark blue indicates locations of PFP (type 2) seepage. From FO-DTS data, a location was
tagged as seepage if the standard deviation was less than 1.5 and the temperature was less than 15 °C for matrix
and less than or equal to 11 °C for PFP seepage. From TIR surveys, seepage was distinguished by temperatures
 890 **of 9-11 °C for interior seepage, and 11-15 °C for matrix seepage. The location of GPR line 7.1 is shown on this**
figure to reference data for the conceptual model in Figure 9.

895



Figure 6: Thermal infrared (TIR) image from March 21, 2014 at Tidmarsh Farms illustrating PFP (type 2) seepage. Many macropores are observed in both the infrared (slightly smaller) and the visual image. These seeps are located in the middle of cell 3 (Fig. 1B), where peat is ~3m thick and dramatically thinning.

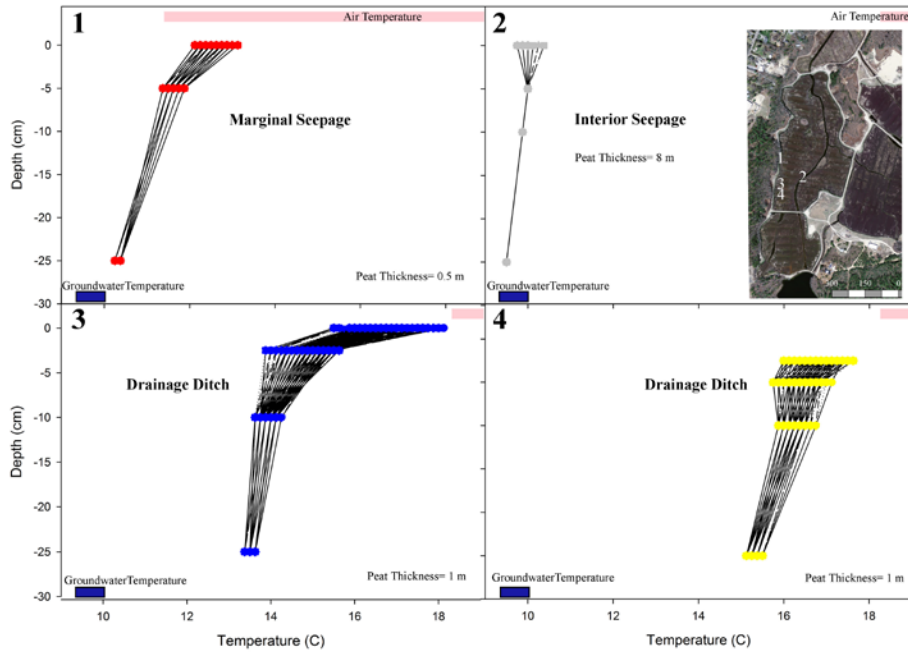
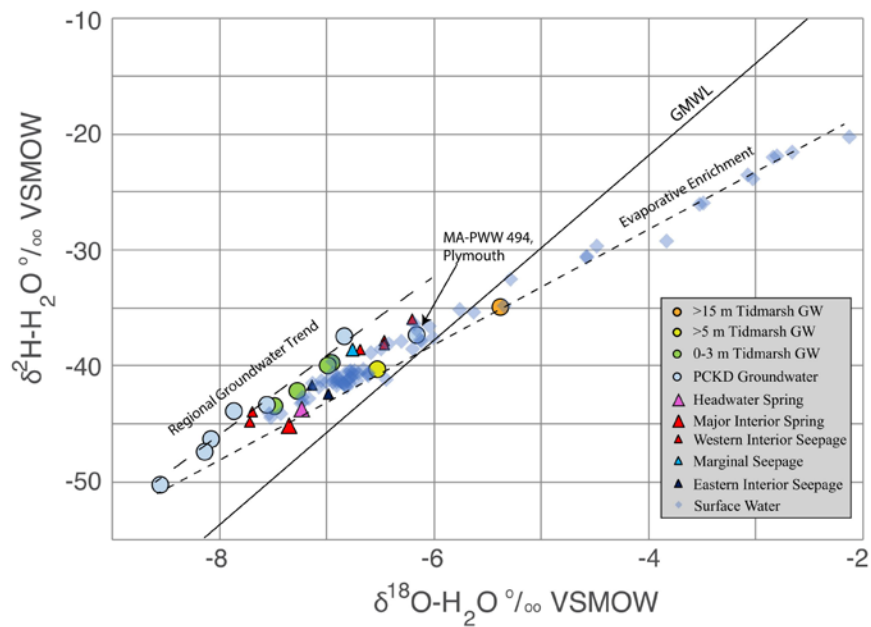
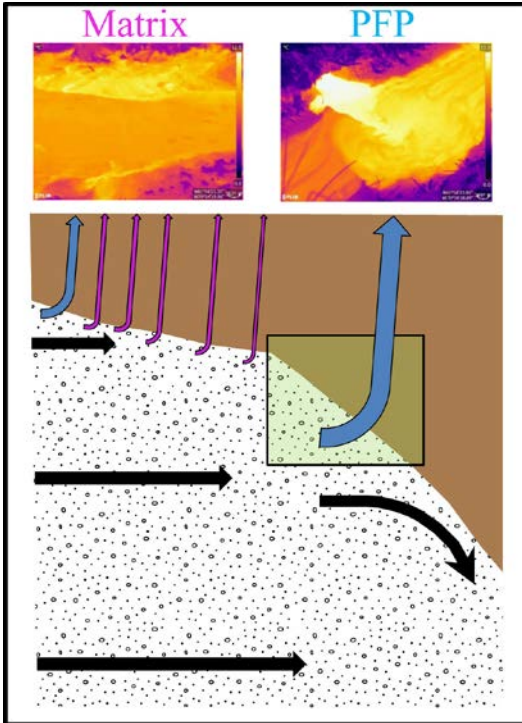


Figure 7: Temperature profiles vs. depth at Tidmarsh Farms recorded in July 30-31, 2013. For each profile, the range of air temperatures and groundwater (GW) temperatures are shown as bands of pink (air) and dark blue (GW). Locations 1 and 2, (profiles 1 and 2) show the influence of upwelling groundwater; expressed as type 1 preferential flow path (PFP) seepage (profile 2) and type 2 matrix seepage (profile 1). The convex upward shape of temperature-depth profiles 3 and 4 is also consistent with upwelling seepage.



910 **Figure 8: Plot of the stable isotopes $\delta^2\text{H-H}_2\text{O}$ and $\delta^{18}\text{O-H}_2\text{O}$ from the Tidmarsh Farms area surface water (diamonds), groundwater (circles), shallow well, and seep sources (triangles). The regional groundwater trend line was derived from samples from relatively shallow, regional USGS wells (blue dots), consistent with a relatively humid climate at the site.**



915 Figure 9: Conceptual model illustrating the mechanism for development of matrix (type 1) seepage (pink
 arrows) and preferential flow path (PFP) or interior (type 2) seepage (blue arrows), corresponding to locations
 in winter TIR images. Thick black lines represent groundwater flow direction, and the yellow-green box
 920 indicates the region of high basin curvature. The brown basin represents peat in a typical basin shape based on
 GPR line 7.1 (Figure 2C). Conceptual matrix and PFP seepage locations are based on the temperature data
 recorded proximal to GPR line 7.1 and winter TIR images from this same transect. PFP seeps found in the
 thicker peat are associated with locations of high basin curvature where strong vertical gradients drive focused,
 higher-flux seepage through pre-existing weaknesses in the peat matrix.