


14 **Abstract**

15 Nitrate is considered the most common non-point pollutant in groundwater. It is often
16 attributed to agricultural management, when excess application of nitrogen fertilizer
17 leaches below the root zone and is eventually transported as nitrate through the
18 unsaturated zone to the water table. A lag time of years to decades between processes
19 occurring in the root zone and their final imprint on groundwater quality prevents
20 proper decision-making on land use and groundwater-resource management. This
21 study implemented the vadose monitoring system (VMS) under a commercial crop-
22 field. Data obtained by the VMS for of 6 years allowed, for the first time known to us,
23 a unique detailed tracking of water percolation and nitrate migration from the surface
24 through the entire vadose zone to the water table at 18.5 m depth. A nitrate
25 concentration time series, which varied with time and depth, revealed—in real time—
26 a major pulse of nitrate mass propagating down through the vadose zone from the root
27 zone toward the water table. Analysis of stable nitrate isotopes indicated that manure
28 is the prevalent source of nitrate in the deep vadose zone and nitrogen transformation
29 processes have little effect on nitrate isotopic signature. The total nitrogen mass
30 calculations emphasized the nitrate mass migration towards the water table.
31 Furthermore, the simulated pore-water velocity through analytical solution of the
32 convection–dispersion equation shows that nitrate migration time from land surface to
33 groundwater is relatively rapid, approximately 5.9 years. Ultimately, agriculture land
34 uses, which are constrained to high nitrogen application rates and coarse soil texture,
35 are prone to induce substantial nitrate leaching.

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91 groundwater recharge behavior and tendency in the deep vadose zone of two
92 agricultural settings, a grapefruit orchard and a crop field (Turkeltaub et al., 2014).
93 Unsaturated flow models were calibrated to the water content observation and were
94 used for groundwater recharge fluxes simulations.

95 The objective of the present study was to demonstrate the water flow and
96 nitrate transport through the deep vadose zone underlying the crop field, with respect to
97 rain patterns as well as the agricultural and fertilization setup. Continuous data on
98 variations in the sediment water content and nitrate concentrations were collected
99 from the entire vadose zone for over 6 years. The nitrate concentration time series,
100 which included variation of nitrate in time and at multiple depths, revealed, in real
101 time, a major pulse of nitrate mass propagating down through the vadose zone toward
102 the water table. These results indicate that nitrate fluxes in the unsaturated zone
103 underlying agriculture land-uses were associated with high nitrogen application rates and
104 coarse texture soils. Furthermore, pollution events originated from agriculture land-
105 uses can be monitored in their early stages, long before pollution accumulates in the
106 aquifer water.

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108 2 Methods


109


110 2.1 Study area


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112 A commercial crop field site was selected as a representative prevalent
113 agriculture setting in the southern part of the coastal plain of Israel (34°41'13" E;
114 31°49'42" N) and is part of an array of VMSs that were installed under different
115 representative land-uses situated above the southern part of the phreatic coastal aquifer

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116 (Dahan et al., 2014, Baram et al., 2013, 2014, Turkeltaub et al., 2014,2015a, 2015b).
117 The study was conducted between 09/2009 and 04/2015. Mediterranean climate
118 prevails in this area, with hot, dry summers (May–September) and rainy winters
119 (October–April), with an average annual rainfall of 512 mm and average temperatures
120 of 31.2 °C (August) and 17.8 °C (January) in the hottest and coldest months,
121 respectively (Israeli Meteorological Service, 2015). Reference evapotranspiration
122 rates calculated according to the Penman–Monteith method (suggested by the Food
123 and Agriculture Organization) range from 1.5 mm day⁻¹ (January) to 5.7 mm day⁻¹
124 (July) (Israeli Meteorological Service, 2015).


125 The crop field cultivation history includes alternation between rainfed
126 agriculture, as wheat, and irrigated agriculture as watermelon for seeds, and cotton as
127 summer crop (personal communication). From 2005 to 2013, the crop field site was
128 cultivated with rainfed winter crops—spring wheat (*Triticum aestivum* L.) and pea
129 (*Pisum sativum* L.) (Fig. 1). Then for 1 year (2013/2014), the field was uncultivated.
130 The crops were sown at the beginning of the wet season (November) and grew into
131 the spring (April). After harvest, disk plow and roller practices were implemented.
132 Since 2005, the main fertilization application to the field was dairy-farm slurry
133 manure, which was distributed over the 10 ha field for 60 days during May and June
134 (Fig. 1). The total nitrogen concentration in the dairy slurry is 900 mg L⁻¹ (Water
135 Authority, 2012). In September 2014, jojoba (*Simmondsia chinensis*) shrubs were
136 planted and irrigation systems were installed.

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138 2.2 Monitoring


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264 year⁻¹ (Herridge et al., 2008), which is about 43% of the nitrogen applied by the dairy
265 slurry. Thus, application of dairy farm slurry combined with a legume crop (pea)
266 seemed to have enriched the top soil with excess nitrogen, as compared to cultivation
267 of cereal-type crops (Fig. 3a).


268 Progression of the nitrate migration deeper into the vadose zone can be
269 divided into two periods. In the first period, October 2010 to January 2013, at depths
270 of 2.7, 4.2, 9.5 and 15.6 m (Fig. 3b,c,e,g), the increase in nitrate concentration was
271 moderate and continuous; whereas, at depths of 6.3 and 18 m, there was no major
272 change in nitrate concentrations (Fig. 3b-d). In the second period, starting from July
273 2013 following the rainy winter of 2012/13, substantial nitrate breakthroughs were
274 noticeable throughout most of the vadose zone cross section (marked with arrows in
275 Fig. 3). This rapid nitrate progression to the deeper parts of the vadose zone could be
276 related to the soil's physical characteristics. In the top 3 m, the soil comprised of fine-
277 textured layers (sandy-loam and loamy sand), and from 3 to 18.5 m (water table), the
278 soil consisted of a coarser sand-textured layer (Turkeltaub et al., 2014). Thus, as a
279 consequence of substantial water percolation, which induced intensive water flux
280 across the coarse-textured soil, nitrate transport could be detected at deeper depths of
281 the vadose zone.


282 Here, as well in previous studies in literature, nitrate fluxes in the unsaturated
283 zone underlie agriculture land-uses were associated with nitrogen application rates
284 and soil physical properties (Green et al., 2008; Botros et al., 2012; Turkeltaub et al.,
285 2015b). Therefore, to attenuate nitrate leaching to aquifers, search should be dedicated
286 to locate the 'hot spots' where these conditions prevailed. Liao et al., 2012).


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288 3.2 Nitrate sources


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site characterization efforts

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conditions favor higher rates of transport

289 The $\delta^{15}\text{N}$ values clearly showed that manure is the main source of nitrate in the
290 vadose zone pore water (Fig. 4). Nitrate isotope composition in the vadose zone pore
291 water depends on nitrogen sources and transformation processes (Böhlke, 2002).
292 Examination of the isotopes values suggested that transformation processes such as
293 denitrification and mineralization of soil nitrogen sources have little effect on nitrate
294 isotopic signature. As discussed in the previous section, the relatively rapid nitrate
295 transport downward to deeper parts of the vadose zone is controlled by soil properties
296 and nitrogen application rates. These factors reduce the potential for transformation
297 processes and plant uptake to occur (Liao et al., 2012). Moreover, various studies
298 conducted under similar conditions (soil types and agriculture land use) as in the
299 current study, presented insignificant nitrogen transformation processes and ability of
300 attenuating nitrate within the deeper vadose zone (Green et al., 2008; Burow
301 et al., 2010; Gautam and Iqbal 2010; Dann et al., 2013; Zhang et al., 2014; Turkeltaub
302 et al., 2015b). Yet, other studies suggested contrasting conclusions. Salazar et al. (2012)
303 reported on low nitrate leaching rates in spite of high nitrogen application rates and
304 Lockhart et al. (2013) claimed that depth to groundwater provided a significant
305 control on nitrate concentration in groundwater regardless of soil type or crop type.
306 Thus, a holistic approach comprises several potential factors that control nitrate fluxes to
307 groundwater should be held to identify the dominant ones.

308

309 3.3 Nitrate storage in the vadose zone

310 The yearly nitrate mass calculations (Eq. 2) displayed an increase from 2009
311 to 2010 (Fig. 5), at the same time as NO_3 concentration increased in the upper part of
312 the vadose zone (Fig. 3a). Subsequently, the highest increase in nitrate mass was
313 calculated for 2011 following the combination of cultivation of the pea crop and

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314 excessive application of dairy slurry (Fig. 5). It seems that the yearly fluctuations in
315 calculated nitrate mass can be explained by the lag time in the transport process
316 between the sampling points. Hence, the peak in nitrate mass observed in the upper
317 parts during 2011 remained in the vadose cross section and eventually reached the
318 deeper parts of the vadose zone ¹ as a breakthrough type (Fig. 5).


319

320 3.4 Nitrate transport model

321 Using nitrate time series obtained from deeper part of the vadose zone for
322 model simulations allowed avoiding the highly dynamic nature of the root zone.
323 Furthermore, transport calculations are less effected by mass balance uncertainties as
324 according to previous section, ² nitrate attenuation processes are insignificant in deep
325 vadose zone.

326 The results indicated relatively good agreement between observed and
327 simulated nitrate concentration trends (Fig. 6). Nevertheless there were discrepancies
328 in the absolute values and with the simulated nitrate concentrations increasing before
329 the observed concentrations at the 6.3 and 18 m depths (Fig. 6a, d). These gaps could
330 be explained by the assumptions that are intrinsic to the CDE model (Eq. 1) —
331 homogeneous medium and average velocity—along with the assumption of even
332 distribution of the nitrogen source on the surface. Nevertheless, the CDE provided an
333 approximation that could be compared with earlier numerical modeling results (van
334 Genuchten et al., 2012). The calculated hydrodynamic dispersion coefficient was 81
335 $\text{cm}^2 \text{day}^{-1}$ and the pore-water velocity was $0.836 \text{ cm day}^{-1}$, which is about 305
336 year^{-1} . Multiplying the velocity by the weighted average water content, $0.060 \text{ cm}^3 \text{ cm}^{-3}$
337 ³ (Fig. 2c-h), the Darcian flux equaled $18.3 \text{ cm year}^{-1}$, which is very similar to earlier
338 average flux estimation of $19.9 \text{ cm year}^{-1}$ averaged for 24 years (Turkeltaub et al.,

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339 2014). If neglecting the diffusion term in the hydrodynamic dispersion coefficient, the
340 estimated longitudinal dispersivity (D/v) is 97 cm. The calculated dispersivity value is
341 relatively large compared with reported values from earlier solute transport
342 investigations in sandy texture soils (e.g. Toride et al., 2003; Dann et al., 2010).
343 However, it was showed that dispersivity increases with travel distance (Vanderborght
344 and Vereecken, 2007).


345 The calculated nitrate transport time from land surface to groundwater is
346 approximately 5.9 years. Yet, the increase in nitrate concentration at the 18 m depth
347 occurred in July 2013, which is 8 years after the first slurry application. Olson et al.
348 (2009) reported that there was a threshold amount of slurry application before nitrate
349 accumulated in the soil. Hence, the gap of 2 years between the first application and
350 nitrate arrival to 18 m depth might be related to the period before critical amount of
351 manure was applied to the field.

352

353 **3.5 Practical implications of vadose-zone monitoring**


354 To prevent a long-term gradual degradation in groundwater quality, the link
355 between sources of pollution on the surface and their migration pattern in the
356 unsaturated zone should be understood long before their final cumulative imprint in
357 the aquifer water. Herein, the application of a VMS under an agricultural field
358 enabled, for the first time known to us, real-time tracking of water flow and nitrate
359 transport from the surface through the entire ¹ deep vadose zone. ² Accordingly similar
360 monitoring concepts for the vadose zone can be used as an alert apparatus for
361 pollution events in their early stages while pollution is still migrating in the
362 unsaturated zone, and long before accumulation in the aquifers water.


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that was sampled by our VMS unit


363 This study demonstrates how nitrate concentrations in the vadose zone exceed
364 the local standard for disqualified drinking-water wells and threaten the groundwater
365 quality. Hence, agro-hydrologically sustainable manure application rates, i.e.
366 sufficient crop production and minimizing nitrate leaching, could be satisfied by
367 suitable regulation or adjustments to meet crop requirements (Olson et al. 2010). To
368 optimize the efficiency of the manure distribution methodology, estimations should
369 include the controlling factors as soil properties, crop type, season, nitrogen
370 attenuation processes and the critical amount of manure application before nitrate
371 accumulation in the soil occurs. Considering only part of the factors could lead to the
372 opposite result. For example, the manure application in this study occurred during the
373 beginning of the dry period, May and June (there are no rain events till October) to
374 prevent nitrogen leaching due to rain events. However, the distributed nitrogen was
375 retained in the soil till winter time and did not undergo significant attenuation
376 processes. The incorrect assumption of manure distribution during the dry period
377 resulted in intensive nitrate leaching. Furthermore, according to the observations
378 presented in this study, the manure application should be reduced following legume
379 crop type. Yet, in many cases, there is a surplus amount of manure to be disposed.
380 Therefore, alternative methods for waste management have to be utilized, coincided¹
381 with regulating² manure application (Westerman and Bicudo, 2005; van Grinsven et
382 al., 2012).

383 Nitrate transport from land surface to water table through a relatively thick
384 vadose zone occurred within less than a decade. This is a considerably rapid pollutant
385 migration when considering remediation strategies. Moreover, the nitrate observations
386 obtained by the VMS and the isotopic signature analysis indicated that nitrate
387 attenuation processes are insignificant. Hence, agriculture sites constrained to similar

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388 conditions as in this study, most of the nitrate mass that leaches under the root zone
389 will eventually reach groundwater.

390


391 **4 Summary and Conclusions**

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393 An intensive nitrate leaching beyond root zone was attributed to soil properties
394 and nitrogen application rates. The implementation of a vadose zone monitoring
395 system (VMS) under an agricultural field enabled real-time tracking of water flow and
396 migration of a nitrate plume from the surface through the deep vadose zone to the
397 water table at 18.5 m depth. Isotopic composition of nitrate-nitrogen in the water
398 samples indicated that manure is the main nitrogen source for nitrate in the vadose-
399 zone pore water. Nitrogen transformation processes seem to have only little effect
400 under an intensively fertilized crop field. Total nitrate mass estimations displayed the
401 nitrate mass advancement toward the deep vadose zone. Moreover, according to the
402 simulated pore-water velocity, nitrate arrival to water table occurred within less than a
403 decade.

404 As in this study, an array of VMSs was installed under other representative
405 agriculture land-uses situated above the southern part of the Israeli costal aquifer. ¹The
406 findings from each site are combined to generate a comprehensive perspective on
407 dominant factors controlling groundwater quality and quantities. Subsequently, these
408 conclusions will be examined with a regional scale aquifer transport model.

409 Protection of groundwater from potential pollution originating from
410 agricultural land uses has to include effective and continuous monitoring of the
411 vadose zone. Pollution events can be monitored in their early stages, long before
412 pollution accumulates in the aquifer water.

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This seems misleading because it implies that in this study, data from an array of VMSs was used; but my understanding is that the data in this paper are only from ONE VMS. I think you need to replace "are combined" to "will be combined in the future".