

Impact of snow deposition on major and trace element concentrations and elementary fluxes in surface waters of Western Siberian Lowland across a 1700-km latitudinal gradient

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Abstract. Towards a better understanding of chemical composition of snow and its impact on surface water hydrochemistry in poorly studied Western Siberia Lowland (WSL), the surface layer of snow was sampled in February 2014 across a 1700-km latitudinal gradient (c.a. 56.5 to 68°N) in essentially pristine regions. The concentration of dissolved+colloidal (< 0.45 µm) Fe, Co, Cu, As, La increased by a factor of 2 to 5 north of 63°N compared to southern regions. The pH and dissolved Ca, Mg, Sr, Mo and U in snow water increased with the increase in concentration of particulate fraction (PF). Principal Component Analyses of major and trace element concentration in both dissolved and particulate fractions revealed 2 factors not linked to the latitude. A hierarchical cluster analysis yielded several group of elements originated from alumino-silicate mineral matrix, carbonate minerals and marine aerosols or belonging to volatile atmospheric heavy metals, labile elements from weatherable minerals and nutrients. The main sources of mineral components in PF are desert and semi-desert regions of central Asia.

Comparison of major and trace elements in dissolved fraction of snow with lakes and rivers of western Siberia across the latitudinal gradient revealed significant atmospheric input of a number of trace elements to the inland waters of the WSL. The snow water concentration of DIC, Cl, SO₄, Mg, Ca, Cr, Co, Ni, Cu, Mo, Cd, Sb, Cs, W, Pb and U exceeded or were comparable with spring-time concentration in thermokarst lakes of the permafrost-affected WSL zone. The spring-time river fluxes of DIC, Cl, SO₄, Na, Mg, Ca, Rb, Cs, metals (Cr, Co, Ni, Cu, Zn, Cd, Pb), metalloids (As, Sb), Mo and U in the discontinuous to continuous permafrost zone (64-68°N) can be explained solely by melting of accumulated snow. The impact of snow deposition on riverine fluxes of elements strongly increased northward, in discontinuous and continuous permafrost zones of frozen peat bogs. This was consistent with the decrease of the impact of rock lithology on river chemical composition in the permafrost zone of WSL, relative to the permafrost-free regions. Therefore, the present study demonstrates significant and previously underestimated atmospheric input of many major and trace elements to their riverine fluxes during spring flood. A broader impact of this result is that current estimations of river water fluxes response to the climate warming in high latitudes may be unwarranted without detailed analysis of winter precipitation.

38 1 Introduction

39 The snow cover exhibits a number of properties making it unique natural archive and indicator of the ecosystem status
40 (Baltrėnaitė et al., 2014; Bokhorst et al., 2016; Callaghan et al., 2011; Caritat et al., 1998, 2005; Garbarino et al., 2002;
41 Guéguen et al., 2016; Kashulina et al., 2014; Lisitzin, 2002; Niu et al., 2016; Ross and Granat, 1986; Singh et al., 2011;
42 Siudek et al., 2015; Van de Velde et al., 1999; Walker et al., 2003). The snow washes out insoluble aerosols particles from
43 the atmosphere as well as soluble compounds, including various pollutants (Telmer et al., 2004; Barrie, 1986; Tranter et al.,
44 1986, 1987). Unlike rain, the snow remains at the soil surface and thus records all atmospheric input during the glacial
45 period of the year. In boreal and subarctic regions, both dissolved and particulate fraction of snow water reflect the
46 chemistry of winter atmosphere, when the land is covered by snow and the water surfaces are covered by ice. During
47 winter, the input of mineral compounds from adjacent regions is minimal and the main factor controlling chemical
48 composition of snow is long-range, hundred and thousand km, atmospheric transfer (Franzén et al., 1994; Huang et al.,
49 2015; Shevchenko, 2003, Shevchenko et al., 2000, 2010; Welch et al., 1991; Zdanowicz et al., 1998, 2006; Zhang et al.,
50 2015).

51 Numerous studies of particulate fraction of snow have been conducted in different regions including western Siberia
52 (Boyarkina et al., 1993; Ermolov et al., 2014; Kashulina et al., 2014; Makarov, 2014; Moskovchenko and Babushkin,
53 2012; Mullen et al., 1972; Salo et al., 2016; Shevchenko et al., 2002, 2010, 2016; Talovskaya et al., 2014; Topchaya et al.,
54 2012; Walker, 2005; Xu et al., 2016). The dissolved ($< 0.45 \mu\text{m}$ or $< 0.22 \mu\text{m}$) fraction of snow was traditionally studied in
55 European subarctic (Caritat et al., 1998; Chekushin et al., 1998; Kashulina et al., 2014; Reimann et al., 1999; Reinosdotter
56 and Viklander, 2005) but the data on trace elements in snow water collected in boreal, arctic and subarctic regions are
57 limited. In contrast to numerous studies of trace element geochemistry of the snow cover of high altitude zones of Asia and
58 northern China (Dong et al., 2015; Kang et al., 2007; Lee et al., 2008; Wang et al., 2015; Zhang et al., 2013), glaciers of
59 Greenland (Barbante et al., 2003; Boutron et al., 2011; Candelone et al., 1996) and Alaskan and Canadian High Arctic
60 (Douglas and Sturm, 2004; Garbarino et al., 2002; Krachler et al., 2005; Snyder-Conn et al., 1997), the trace element
61 geochemistry of dissolved and particulate fraction of Siberian snow remains at the beginning of exploration. This is
62 especially true for large and geographically homogeneous territories of western Siberia, presenting relatively similar level
63 of snow deposition during winter seasons (i.e., from 100 mm of water in the south to 140–150 mm of water in the north)
64 without any pronounced influence of large industrial centers, mountain regions and marine aerosols over the territory close
65 to 1.5 million km^2 (Resources, 1972, 1973; Boyarkina et al., 2013).

66 The originality of the present study consists in *i*) sampling of substantial ($\sim 1700 \text{ km}$) latitudinal transect in relatively
67 pristine zones comprising forest, forest tundra and tundra within the permafrost-free, discontinuous and continuous
68 permafrost regions; *ii*) assessment of both dissolved+colloidal and particulate forms of major and trace elements in snow
69 samples. Given the scarcity of available measurements of snow chemical and particulate composition in Western Siberia,
70 we aimed at addressing the following specific issues: (1) characterizing the effect of the latitude on major and trace
71 element concentration in dissolved ($< 0.45 \mu\text{m}$) and particulate ($> 0.45 \mu\text{m}$) fractions of snow; (2) testing the link between
72 dissolved and particulate fractions of elements and the impact of particle mineralogy on snow chemical composition; (3)
73 comparing dissolved concentrations of major and trace elements in snow to those in lakes and rivers across the latitudinal
74 gradient of WSL and (4) assessing the share of snow deposition on seasonal and annual export of dissolved elements by
75 western Siberian rivers. Via addressing quantitatively the abovementioned issues using unified methodology in
76 unprecedentedly large geographical coverage (56 to 68°N) of orographically flat low populated terrain, we anticipate to
77 enhance our knowledge of the winter atmospheric deposition in western Siberia, in the absence of direct influence of

78 marine aerosols and large industrial centers. This should eventually allow to evaluate the impact of snow deposition on
79 chemical composition and elementary fluxes of subarctic inland waters across a large latitudinal gradient of climate and
80 permafrost parameters.

81

82 2. Study site, materials and methods

83 2.1. Geographic settings

84 Western Siberia Lowland, located between the Ural mountains and the Yenisei River, extends over 2000 km
85 from south to north and presents highly homogeneous, from physico-geographical point of view, taiga, forest-tundra
86 and tundra landscapes comprising bogs and mires in the permafrost-free zone and thermokarst lakes developed on flat
87 peat bogs (palsa) in the permafrost-bearing zone. Detailed physico-geographical description, hydrology, lithology and
88 soils can be found in earlier works (Botch et al., 1995; Smith et al., 2004; Frey and Smith, 2007; Beilman et al., 2009)
89 and in our recent geochemical studies (Manasyrov et al., 2014; Stepanova et al., 2015; Pokrovsky et al., 2015). Because
90 of its exceptionally flat orographic context, extensive vegetation cover and relative remoteness from the Arctic Coast
91 (except the north of the Gyda and the Yamal peninsulas), the atmospheric precipitates in winter are likely to bear the
92 signature of remote desert and semi-desert regions of Central Asia. The anthropogenic impact is not expected to be
93 strongly pronounced because of *i*) low population density (average 6 people/km² but only 0.5–2 people/km² in the
94 northern half of WSL) and *ii*) moderate local pollution from the gas burning in oil wells mostly in permafrost-free
95 zone, south of the Surgut town. The part of WSL north of 64°N contains essentially gas exploration facilities (no gas
96 burning) and minimally impacts the environment. Taken together, the latitudinal profile of the WSL presents a unique
97 opportunity to study the chemistry of atmospheric deposits within highly homogeneous physico-geographical context
98 and relatively low local anthropogenic impact.

99

100 2.2. Snow sampling

101 The snow of the WSL was sampled along the latitudinal transect S → N, from the vicinity of the Tomsk city (zone
102 of southern taiga) to the eastern coast of the Ob estuary (tundra zone) from 19.02.2014 to 5.03.2014 (Fig. 1). The possible
103 sources of snow deposition and the pathways of aerosols transport to the WSL were reconstructed by analyzing
104 meteorological maps and by calculating back trajectories of air transport to the observation points using NOAA's
105 HYSPLIT model (Draxler and Rolf, 2003). In order to assess a snapshot of snow deposition across 1700-km latitudinal
106 profile and collect the freshest snow that was subjected to minimal transformation, we have chosen to sample only the
107 upper layer of the snow cover. This technique, in contrast to traditional sampling of full snow column (i.e., Guéguen et al.,
108 2016; Niu et al., 2017) allows adequate representation of the upper fresh snow layer that had minimal transformation at the
109 soil, and frequently used in remote regions (Kang et al., 2007; Zhang et al., 2013). The isotope composition of collected
110 snow proved its fresh character, not subjected to any metamorphism (Vasil'chuk et al., 2016).

111 The upper 0–5 cm of snow was sampled in 39 locations (Fig. 1). All sampling points were located far than 500 m
112 from the winter road. The sampling was performed using metal-free technique, in protected environment, using pre-
113 cleaned plastic shovel and vinyl single-used gloves. Approximately 30 L of snow was collected into single-used
114 polyethylene bags. These polyethylene bags were thoroughly washed with 1 M HCl and abundant MilliQ water in the
115 clean room class A 10,000. In the laboratory, the snow was melted at ambient temperature, and filtered through pre-

116 weighted acetate cellulose filters (Millipore, 47 mm diameter) of 0.45 μm poresize. The storage of unfiltered snow
117 water samples was less than 1 h at 4°C.

118

119 2.3. Particle analyses

120 The sizes and morphology of particles on filters and elemental composition of individual particles were studied
121 using scanning electron microscope VEGA 3 SEM (Tescan) with a microprobe attachment INCA Energy (Oxford
122 Instruments). The mineralogical composition of particulate fraction on selected filters was studied by X-ray powder
123 diffractometric method on the D8 ADVANCE (Bruker AXS) X-ray diffractometer equipped with the LYNXEYE linear
124 detector (Lisitzin et al., 2015). The uncertainty of the relative proportion of mineral composition was 1–2% and the
125 detection limit was 1%.

126 Freshly melted snow water was filtered through pre-weighted 0.45 μm acetate cellulose (Millipore) filters. These
127 filters were placed in the Petri dishes, dried at 60°C in an oven and digested using microwave acid attack which comprised
128 6.5 mL concentrated HNO_3 , 3.5 mL concentrated HCl and 0.5 mL concentrated HF , HNO_3 and HCl were bi-distilled in the
129 clean room and HF was commercial ultra-pure quality (Fluka). The filters were reacted 30 min in ultrasonic bath prior full
130 digestion using Mars 5 microwave digestion system (CEM, France). For this, 10 samples of filters, 1 certified 2711a
131 Montana II Soil standard and 1 blank filter sample were loaded into Teflon reactors subjected to treating at 150°C during
132 20 min. After completing the digestion, the content of reactors was transferred to 30 mL Savilex vials and evaporated at
133 70°C. The residue was dissolved in 10 mL of 10% HNO_3 and diluted by 2% HNO_3 prior to the analyses. For the analysis of
134 snow particles on filters, the blanks (digestion of initial filters) were a factor of 10 to 100 lower than the filters with
135 particles after 0.5–1.0 L of snow water filtration. The concentration of major and trace elements (TE) in filter digestion
136 products was measured using an ICP-MS Agilent 7500 ce with $\sim 3 \mu\text{g/L}$ of indium and rhenium as internal standards. 4 in-
137 house external standards were analyzed every 10 samples. Necessary corrections for oxide and hydroxide ion interferences
138 were made for rare earth elements (REE) and metals (Ariés et al., 2000). Based on replicate analyses of in-house standards
139 and certified materials, the uncertainty for TE measurement ranged from 5 % at 0.1–100 $\mu\text{g/L}$ to 10 % at 0.001–0.01 $\mu\text{g/L}$.
140 Analyses of low concentrations of Hf, Ge, Cs, Ga, and W (e.g., on the order of 0.001 $\mu\text{g/L}$, comparable with detection
141 limits) was possible with a minimal estimated uncertainty of 20%.

142

143 2.4. Melted snow analyses

144 The pH and specific conductivity were measured on unfiltered snow water samples using Hanna portable
145 instruments. The dissolved ($< 0.45 \mu\text{m}$) fraction of snow water was obtained via filtration using a polycarbonate
146 Nalgene vacuum filter unit, and a PVC-made Mityvac hand vacuum pump. This fraction included colloidal and truly
147 dissolved (ionic) forms. Blanks of MilliQ water were also placed in polyethylene bags for the same time as melting
148 snow ($\leq 1 \text{ h}$ at 4°C) and processed via filtration similar to snow samples. The filtrates were divided into two parts; one
149 was acidified with double distilled HNO_3 acid and stored in pre-cleaned HDPE vials for ICP MS analysis, the second
150 part was stored in HDPE bottles without acidification, for dissolved organic and inorganic carbon analysis (DOC and
151 DIC), respectively, and anion analysis.

152 The major anion concentrations (Cl^- , SO_4^{2-}) in the $< 0.45 \mu\text{m}$ fraction were measured using ion chromatography
153 (HPLC, Dionex ICS 2000i), with an uncertainty of 2%, estimated from the replicate analyses of PERADE and RAIN
154 international certified materials. The DOC and DIC in this fraction were analyzed using a Carbon Total Analyzer
155 (Shimadzu TOC-VSCN) with an uncertainty of 5% and a detection limit of 0.1 and 0.05 mg/L , respectively.

156 Filtered snow water samples were analyzed with an Element XR ICP MS allowing for much better precision of
157 the analyses of highly diluted samples and avoiding many interferences compared to Agilent 7500 ce. The uncertainty of

158 the Element XR analysis was $\pm 5\%$, while its detection limit was a factor of 100 lower than the traditional (Agilent)
159 instrument. The Element XR operated in three modes depending on the elements measured: low resolution for B, Rb, Sr,
160 Zr, Mo, Cd, Sb, Cs, Ba, REEs, Hf, W, Pb, Th, U; medium resolution for Na, Mg, Al, Si, P, Ca, Ti, V, Cr, Mn, Fe, Co, Ni,
161 Cu, Zn, Ga, Sr, and high resolution for K and As. The agreement between two instruments for most elements was within 10
162 %. The international geostandards SLRS-5 (Riverine Water References Material for Trace Metals certified by the National
163 Research Council of Canada) was used to assess the validity and reproducibility of the analyses.

166 2.5. River fluxes and snow storage

167 The mass balance calculation of the degree of snow melt influence on element fluxes in WSL rivers was
168 performed taken into account *i*) the water stock in snow (in mm snow water accumulated during winter), fairly well
169 known for Western Siberia (Karnatzevitch and Khruschev, 2014; Resources, 1972, 1973; Zakharova et al., 2011) and *ii*)
170 the spring-time river runoff (in mm during May and June) calculated based on hydrological parameters. For water stock
171 calculation, we used the available mean multi-annual daily and monthly discharges of WSL rivers across the latitudinal
172 profile (Resources, 1972 and 1972 and recently compiled in the database R-ActicNET (www.r-arcticnet.sr.unh.edu)).
173 The WSL territory is covered by Russian Hydrological Survey (RHS) gauging stations which allowed to calculate the
174 discharge during May-June as described elsewhere (Pokrovsky et al., 2015). The most recent thorough hydrological
175 measurements on small and medium size rivers of permafrost – affected part of WSL (Novikov et al., 2009) were used
176 together with RHS database to calculate the spring flood fluxes of individual rivers and snow water stock for three
177 latitudinal zones, 56-60°N, 60-64° and 64-68°N. Note that a comparison between the elementary snow stock and the
178 river elementary discharge could not be performed for individual river watersheds, since no snow water chemical data
179 are available with necessary spatial resolution. Therefore, we compared the winter snow stock with riverine spring flood
180 fluxes of major and trace element for three latitudinal zones. For this, both spring flood flux of individual rivers and
181 snow water stock were averaged for each latitudinal zone.

184 2.6. Statistical methods

185 Statistical analysis of the average, median and geometric mean values and the link between element concentration
186 in suspended and dissolved fraction as well as comparison of different sampling sets (snow water and snow particles)
187 included ANOVA, H-criterion of the Kruskal–Wallis and Mann–Whitney U tests. These tests allowed evaluating the
188 difference between two sets of data separately for each TE following the approaches developed for lakes and rivers of
189 western Siberia (Manasypov et al., 2014, 2015; Pokrovsky et al., 2015, 2016). The multiple regressions were performed
190 for quantifying the relationship between dissolved and particulate concentration of TE and the latitudinal trends of
191 concentrations and enrichment factors. More thorough statistical treatment of both log-transformed and non-transformed
192 major and TE concentration in dissolved and particulate fraction of snow samples in each location included a normed PCA
193 analysis using the ADE-4 R package (Thioulouse et al., 1997; Chessel et al., 2004) using the methods for scores and
194 variables (De la Cruz and Holmes, 2011).

195 To identify the group of elements that behaved in a similar way in snow water and snow particles, we applied a
196 complementary hierarchical cluster analysis (HCA) (e.g. Hartigan, 1975; Kaufman and Rousseeuw, 2005) which is widely
197 adopted in geochemical interpretations of element concentration data (e.g. Bini et al., 2011; Levitan et al., 2015; Schot and
198 van der Wal, 1992; Moragues-Quiroga (2017)). We used the Ward's method (Ward, 1963) for the linkages rule, following
199 previous studies (Gourdol et al., 2013; Lin et al., 2014). The Pearson correlation distance was used for the linkage distance,

200 which is frequently used for cluster variables (Reimann et al., 2008). These choices are in agreement with the group search
201 of the PCA loadings.

202 For the assessment of element enrichment factor in snow particles, a normalization to both general upper Earth
203 crust and the local geological background (soil, peat and moss) was used. The reason for this is that, unlike in studies of the
204 local pollution tracing in the European arctic (e.g., within the Kola Ecogeochemistry project, see de Caritat et al., 1997;
205 Reimann and de Caritat, 2000; Reimann et al., 2000) or small-scale stream bed sediments or soils (N'guessan et al., 2009;
206 Moragues-Quiroga et al., 2017; Levitan et al., 2015) where the normalization to the local soil or bedrock was necessary, the
207 present study essentially deals with winter-period long-range atmospheric transport of soluble and mineral forms of
208 elements. As such, following the common practice in this field, the normalization to upper Earth crust allowed assessing
209 the true enrichment/depletion of the atmospheric aerosols. However, in order to better represent the elementary features of
210 snow particles, the concentration of elements in PF was also compared with western Siberian mineral soils, peat and moss.

211
212

213 3. Results

214 3.1. Soluble fraction of the snow water

215 For all major and most trace elements, the concentrations in the blanks were below or comparable with analytical
216 detection limits (≤ 0.1 ng/L for Cd, Ba, Y, Zr, REEs, Hf, Pb, Th, U; 1 ng/L for Ga, Ge, Rb, Sr, Sb; ~ 10 ng/L for Ti, V, Cr,
217 Mn, Fe, Co, Ni, Cu, Zn, As). These values were at least 5 times lower than the average concentration of trace elements in
218 snow samples. Most TE presented in this work exhibited $\leq 15\%$ -agreement between the certified or recommended values
219 and our measurements. The TE for which certified or recommended data were not available were considered only for the
220 cases where we obtained good analytical reproducibility (i.e., the relative standard deviation based on our standard
221 measurements was $\leq 10\%$).

222 The latitude-averaged concentrations of dissolved and particulate fraction of snow samples are listed in **Table 1**.
223 A full data set of major and TE concentration in snow water is given in the "Data availability" section. Examples of the
224 effect of latitude on dissolved (< 0.45 μm) element concentrations are shown in **Fig. 2**. Fe and Cu demonstrated a 2 to
225 5-fold increase in dissolved concentration north of 63°N (at $p < 0.05$). Zn and Pb did not exhibit any systematic effect of
226 latitude, and Sb, Cd and Ni demonstrated a single maximum at c.a. 63 - 65°N . As exhibited two maxima, at 63.5 and
227 67.5°N with overall 2 to 3-fold decreasing trend northward. All other major and trace elements demonstrated a lack of
228 systematic variation of concentrations as a function of latitude (not shown).

229 The PCA treatment of soluble fraction suggested that at least two factors are interpretable. The PC1 x PC2
230 correlation circle revealed two large groups of variables (**Fig. 3A**). The first group is composed of Al, Fe, Cr, Zr, Pb,
231 REEs corresponding to lithogenic, poorly soluble trace elements. The second group is composed of DOC, K, Rb, Cs,
232 Mn, Co, Ba, Sb, Co, Mo, Mg, Si, Sr, Ca, pH. These highly mobile elements presumably reflect the marine aerosols and
233 leaching from soluble soil minerals such as carbonates as well as plant biomass. Similar factors determine chemical
234 composition of snow water regardless of the latitude of the sampling and no specific conditions or limiting factors
235 depended on geographical location.

236 The HCA analysis was conducted on the basis of first two factors of the PCA. The criterion of non-intersection
237 between the groups allowed partitioning the chemical elements of the dissolved part into 6 specific groups presented in
238 **Fig. 3B**. These groups characterize the elements according to their general chemical properties, ability to mobilize in
239 aqueous solution from the solid minerals, affinity to the biota or their presence in the contaminated particles of
240 industrial activity. Thus, the first two group of the dissolved fraction shown in **Fig. 3B** and encircled in **Fig. 3A**
241 comprise low mobile elements likely originated from alumino-silicate mineral matrix (Al, Cr, REE, Ti, Zr, Fe, V) as

242 well as some volatile heavy metals typically present in the solid aerosol particles (Cu, Cd, Pb). The 4th group includes
243 major constituents of carbonate or marine aerosols matrix (elevated pH, Mg, Ca and Na). The 5th group is represented
244 by typical macro- and micronutrients (K, Rb, Mn, Co, Ba). Finally, the last 6th group of elements comprises both labile
245 elements linked to weatherable minerals (Sr, Sb, Si, Ni) and nutrients such as Sr, Ni, Si, DOC and Mo. Three of these
246 elements are strongly enriched in snow particles relative to the Earth crust (Sr, Sb, Mo, see section 3.3 below), thus
247 suggesting their possible leaching from atmospheric dust into the soluble fraction of snow. We could not find a
248 straightforward explication of the common group of Zn and U in soluble snow fraction (Fig. 3B)

249 The effect of particulate fraction on dissolved element composition in snow is illustrated in **Fig. 4** where the
250 value of pH (**4 A**), Sr (**4 B**), Al (**4 C**) and Pb (**4 D**) in dissolved fraction are plotted as a function of total particle
251 concentration in snow water. The elements of 4th and 5th group (Ca, Mg, Sr, Mn, and Co) increase their concentration in
252 snow water by ca. an order of magnitude with the increase in particle concentration by 2 orders of magnitude. The
253 insoluble hydrolysates (Fe, Al, light REEs, Zr, Cu and Pb) belonging to 1st and 2nd HCA group decrease their
254 concentration (less than a factor 10) when the particle concentration increases by 2 orders of magnitude. Other elements
255 in < 0.45 µm fraction exhibit the variations within an order of magnitude (DOC, DIC, Na, Cl, SO₄, K, Si, Cr, V, Ni, Cu,
256 Zn, As, Sb, Rb, Cd, Cs, Ba, heavy REE and U) or two orders of magnitude (Ti, Ga, Mo, W) and do not demonstrate any
257 significant (at p < 0.05) link with particle concentration.

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260 **3.2. Possible impact of snow deposition on major and TE in lakes and rivers.**

261 3.2.1. Snow water in comparison to lake and river water

262 The concentrations of dissolved major and trace element in snow water fraction can be compared with those in
263 thermokarst (thaw) lakes of western Siberia measured in 2013-2014. These lakes are shallow (0.5-1.5 m depth) water
264 bodies representing the largest reservoir of surface waters in western Siberia, north of 62°N. The average concentration
265 of major and TE in thermokarst lakes of various size (Manasypov et al., 2014) can be compared with those in snow
266 water collected in this study across the same latitudinal gradient. Because the size of thermokarst lakes of WSL ranges
267 from few m² to several km², 4 representative ranges of lake diameters are used for this comparison (0-10, 11-100, 101-
268 500 and > 500 m). Given that the spring-time lake concentrations across the latitudinal gradient are not available, the
269 summer-time elementary compositions of lakes were taken as most representative for the open period of the year. The
270 concentrations of low-soluble elements in lakes are well above their concentrations in snow (Fig. S1, A, B for Al and
271 Fe). At the same time, many trace elements exhibited snow-water concentrations that were comparable or significantly
272 higher (p < 0.05) than the concentrations in lakes. Examples of Zn, Cu, Cd, Pb, Sb and Mo are given in Fig. 5. The
273 excess of snow water concentration over summer lake concentration did not follow any particular latitudinal pattern.
274 Note, however, the similarity of the shape of latitudinal pattern in snow and lakes for Zn and Sb.

275 Because the main source of water in shallow lakes of WSL in spring is melted snow (Manasypov et al., 2015),
276 we could compare the mean concentrations of snow water with spring-period lake water concentration for one particular
277 region of discontinuous permafrost zone (town of Nojabrsk, Khanymey site) for which high-resolution seasonal
278 observations on lakes of various size are available. For two classes of lake size (< 0.5 km² and > 0.5 km²), the following
279 three groups of elements could be distinguished. The concentrations of dissolved Na, Mn, Zn, As, Rb and Sr in snow
280 water are similar (within a factor of 2) to lake water concentrations. Concentrations of DIC, Cl, SO₄, Mg, Ca, Cr, Co,
281 Ni, Cu, Mo, Cd, Sb, Cs, W, Pb and U in snow are close or higher (p < 0.05) than those in lakes. And finally,
282 concentrations of DOC, Al, Si, K, Ti, V, Fe, Ga, Zr, Ba, and REEs in snow water are significantly lower than the lakes'
283 concentrations. There was no distinction of elements belonging to individual groups of the HCA and this classification.

284 The concentrations of elements in snow water could be also compared with river water concentrations measured
285 during spring flood 2014 across the full latitudinal profile, since such data for rivers of different size are available
286 (Pokrovsky et al., 2015, 2016a). Examples of elements whose concentrations in snow water are higher or comparable
287 with those in rivers during spring flood are illustrated in Fig. S2. Generally, the effect of snow melt is mostly
288 pronounced north of 64°N. During this period, when the rivers are essentially fed by melted snow, the atmospheric
289 deposition exhibited comparable or higher ($p < 0.05$) concentrations of SO₄, Cr, Co, Ni, Cu, Zn, Mo, Cd, Sb, Cs, W and
290 Pb than those in rivers. These elements belonged to 5 dominant groups of HCA treatment. The concentrations of all
291 other elements in WSL rivers cannot be explained by solely snow water concentration.

292 Note that, by filtering the snow and the lake/river water to 0.45 μm, the dissolved fraction includes a colloidal
293 load, which can play a crucial role in the concentration of trace elements (Pokrovsky et al., 2016b). However, with
294 typical concentration of DOC in snow water around 1-2 mg/L, the share of colloidal forms of metals will be an order of
295 magnitude lower than that in river and lakes of the WSL having 10 to 30 mg/L of DOC.

297 3.2.2. Comparison of river fluxes in spring and snow water stock

298 Considering the mass balance calculation of snow melt influence on element fluxes in WSL rivers, the ratios of
299 river fluxes in May-June to snow stock can be presented in the form of histograms (Fig. S3 of Supplement). These
300 ratios systematically decrease with the increase in the latitude. In the southern, permafrost-free zone, Zn, Cd, Pb, Ga,
301 Cs, W, Sb and Cl fluxes in rivers can be provided essentially by snow melt. The riverine fluxes of DIC, Cl⁻, SO₄²⁻, Na,
302 Mg, Ca, Sr, Rb, Cs, Zn, Cu, Cr, Ni, Cu, Pb, As, Sb, Mo, W and U are strongly (i.e., $\geq 50\%$ at $p < 0.05$) affected by snow
303 melt in the discontinuous and continuous permafrost zones, north of 60-62°N.

304 According to the evolution of the ratio [river flux] / [snow stock] with the latitude, three group of element can be
305 distinguished: (i) elements that steadily decrease this ratio suggesting an increase in the impact of snowmelt northward:
306 DOC, SO₄, Al, Ti, V, Cr, Rb, Sr, Cd, Sb, Cs, La, Ce, W, Pb; (ii) elements for which this ratio decreases abruptly to
307 62±2°N and then remains constant further northward: DIC, Na, Mg, Si, K, Ca, Ni, Cu, As, Mo and U; (iii) elements
308 exhibiting non-systematic variation of the ratio with latitude but having strong (> 50%) impact of snowmelt on river
309 fluxes (Cl, Co, Zn, Ga) and (iv) elements having negligible (< 10 %) impact of snowmelt on river fluxes (Mn, Fe, Zr
310 and Ba). The histograms of elements whose riverine fluxes in spring are affected by snow deposition are given in Fig. 6
311 for 3 latitudinal zones. Overall, the impact of snow melt on river export fluxes in spring strongly increases northward
312 for DIC, Cl⁻, SO₄²⁻, Na, Mg, Ca, Cd, Pb, Sb, Cr, Cu, Ni, As, Mo, Rb, U. Although these elements belong to all 5 major
313 groups of cluster analysis (Fig. 3 B), they can be characterized as soluble (highly labile) elements, originated either
314 from marine aerosols or from leaching from soluble minerals such as carbonates, and also include volatile constituents
315 of the atmospheric aerosols (Cd, Pb, Sb, As).

318 3.3. Particle concentration and TE in particulate fraction of snow

319 Concentration of particulate fraction (PF) of snow and its elementary composition are available in the “Data
320 availability” section. The mineralogical composition of most representative snow samples is given in Table S1 of the
321 Supplement. The dominant minerals are quartz (37%), albite (13%), K-feldspar (13%), phlogopite (10%), chrysotile
322 (8%), illite (7%), and chlorite (5%). The concentration of dolomite and calcite ranges from 1 to 48 and 1 to 19%,
323 respectively. Although mineral components dominated the composition of particulate fraction, the PF also contained
324 organic fibers, diatom frustules, pollens and particles produced during fuel burning (fly ash and black carbon). The
325 concentration of particles in snow water ranged from 0.4 to 67 mg/L. The highest values are encountered in the vicinity

326 of the Tomsk city (No SF 1) and around towns of Surgut (No SF 54, 14), Nojabrsk (SF 36, SF 38) and Gubkinsky (SF
327 33). Although the proportion of fly ash and black carbon in these samples is significant and higher than in the rest of
328 samples as follows from SEM observation, the mineral particles (1-25 μm size) still dominate. Note that high content of
329 fly ash and fuel burning spheres **was** not linked ($p > 0.05$) to high particulate and dissolved elements. The lowest
330 concentrations of particles ($< 5\text{-}10\text{ mg/L}$) **were** recorded north of 65°N , the region of gas industry, and between 58 and
331 61°N **corresponding to the winter road** along the Ob River with very low population density.

332 In order to assess the degree of element fractionation in snow particles, Al-normalized TE enrichment factor
333 (EF) with respect to the average upper part of continental earth crust (Rudnick and Gao, 2003) was calculated according
334 to:

$$335 \quad \text{EF} = \frac{[TE]/[Al]_{\text{sample}}}{[TE]/[Al]_{\text{crust}}}$$

336 The enrichment coefficient ranged from $\sim 1\text{-}5$ (Ga, REEs, Fe) to > 100 (Mo, W, As, Sb, Ni, Cu, Pb, Mg, Ca, Na)
337 as illustrated in **Fig. 7**. The highest enrichment ($\text{EF} \geq 1000$) is observed of Sb, Zn and Cd. The variation of the
338 enrichment factor as a function of latitude is shown for elements most enriched in particulate fraction in **Fig. S4 of**
339 **Supplement**. For Mg, Ca, Sr, Ba, Fe, Mn, Co, Ni, K, Rb, Cs, V, Cr, As, Cd, W the EF exhibits a maximum around 63-
340 64.5°N . This maximum coincides with the maximum of particulate fraction concentration (not shown).

341 The majority of chemical elements are present in particulate rather than dissolved form in snow meltwater
342 samples. This is illustrated by a histogram of the ratios averaged over full latitudinal profile (**Fig. 8**). Although the
343 variations of this ratio for different snow water samples across the WSL achieve ± 0.5 order of magnitude, the average
344 values shown in this figure illustrate the importance of particulate deposition of Al, Fe, Ga, REEs, Cr, V, Ti, Zr, Mo and
345 W. For other elements, particulate and dissolved inputs in the form of snow are within the same order of magnitude.
346 Some soluble elements such as Na, Cd, Ca, Sr, Ba, K, As and Zn exhibit the dominance of dissolved transport in snow.

347 **Although the use of average crust for assessment of element enrichment in snow particles is justified by long-**
348 **range transfer of snow components, it is known since the works of group of Reimann and de Caritat in NW Europe that**
349 **the “average crust” is unlikely to represent the local background and the use of the “upper crust” average value can**
350 **introduce a 2 to 3 order of magnitude uncertainty to any calculated EF (de Caritat et al., 1997; Reimann and de Caritat,**
351 **2000; Reimann et al., 2000). As such, western Siberia moss, peat and clay/loam horizons were used to assess relative**
352 **enrichment of elements in snow particles. It can be assumed that the leaching of soluble forms of elements from these**
353 **solid phases in winter is highly unlikely. The specificity of western Siberia is that the mineral (“geological”) local**
354 **substrate is completely frozen, even in summer, since the active (unfrozen) layer depth does not exceed the peat**
355 **thickness, and in that case, the use of “organic” substrates is most relevant. All three WSL reference substances (“local”**
356 **moss, peat and clays) represent latitudinal-averaged values based on large (> 50) number of samples collected in**
357 **previous studies across the 1700-km latitudinal gradient.**

358 **The elementary ratios of snow particles to that in mineral soil, peat and moss of the WSL are illustrated in Fig. 9**
359 **A, B, and C, respectively.** Given significant uncertainties on the latitude-averaged values of element concentration in
360 snow particles, mineral, peat and moss of soil column, the deviation of the ratios from unity is significant if it exceeds a
361 factor of 2 to 3. Compared to mineral soil of WSL, the snow particles are strongly ($\geq 10\times$) enriched in Sb, Zn, Ni and
362 Cd and in a lesser degree ($\geq 5\times$) in Mg, Ca, Pb, Mo, and As (**Fig. 9 A**). Note that western Siberian soils, developed on
363 sand and clay (silt) deposits (Vasil’evskaya et al., 1986), are quite poor in Ca and Mg, especially in the permafrost-
364 bearing zone north of 62°N . The enrichment of snow particles relative to peat is observed for all elements, being
365 particularly high ($> 50\times$) for Ni, Cr, Pb, Cu, Zn, Mg, Na and Sb (**Fig. 9 B**). Only P, Ge and Cd, exhibiting high affinity
366 to peat (Shotyk et al., 1990, 1992), are not significantly ($p > 0.05$) higher in snow particles compared to the peat

367 column. Finally, the mosses are most depleted by all elements relative to snow PF with only biogenic elements (P, K,
368 Rb, Mn and Cd) known to be concentrated in bryophytes being non-significantly higher in snow particles relative to
369 mosses (Fig. 9 C).

370

371 The PCA of elementary composition of particulate fraction demonstrated the F1 x F2 structure (Fig. S5 A of the
372 Supplement). Here, two groups can be distinguished: highly mobile elements (Na, Ca, V, Ni, Mg, Mn) and low mobile
373 elements (REE, Zr, Pb, Cd, Ga, P). For the particulate fraction, the HCA attributed the elements to 5 formal groups
374 shown in Fig. S5 B and encircled in Fig. S5 A. This distinction, however, is less certain than that of the dissolved
375 fraction and does not allow establishing a clear link between the selected groups and physico-chemical properties of
376 elements or their possible sources in the snow particles. Thus, in the 1st group, among three labile elements (Mg, Na and
377 Ca) we identified V, which may exhibit elevated mobility in the form of anion in carbonate-bearing mineral particles.
378 Divalent metals (Co, Ni, Mn) and Sr constitute the 2nd labile group of elements, yet this group also comprises low-
379 mobile Fe and Cr. The 3rd group of insoluble low mobile elements is marked by the presence of phosphates (REE and
380 P), refractory Zr and volatile Pb. The 4th group of elements revealed by HCA of particles is composed of Sb, Cu and Zn.
381 All these elements are strongly enriched in snow particles over the soil minerals (see Fig. 7, 9 A). The last group of
382 elements in snow particles comprises both labile (Li) and biologically-important Mo, K, Rb, Ba, toxic volatile elements
383 which could bear the signature of anthropogenic pollution (As, Cd) but also low mobile Ti and Ga. We could not
384 identify the link of elements in this group to the degree of snow particles enrichment relative to main “local” substrates
385 of the WSL (moss, peat and clays), shown in Fig. 9.

386

387

388 4. Discussion

389 4.1. Particulate versus dissolved transport of major and trace elements by snow

390 In accord with general knowledge of the Arctic aerosol chemistry (Barrie, 1986; Barrie and Barrie, 1990; Laing
391 et al., 2014, 2015; Nguyen et al., 2013; Pacyna and Ottar, 1989; Shevchenko et al., 2003; Weinbruch et al., 2012), the
392 principal component structure of snow water chemistry identified the combination of lithogenic source (dust and soil
393 particles dissolution, providing low-mobile, insoluble elements such as Al, Fe, Cr, Zr, REEs) and marine aerosols
394 (soluble forms, providing high concentration of mobile elements such as Ca, Mg, Na, Mo, Ni). The latter may also
395 originate from aeolian transport of carbonate-rich soils. The biogenic component may include Mn, Zn, K, Rb, DOC, Si
396 whereas the anthropogenic pollution originates from coal combustion (Sb, Co) and heating systems, gas flaring at the
397 gas oil production site as well as non-ferrous metal-smelter industry (Sb, Zn, Vinogradova et al., 1993) and ground
398 transportation (Pb, Cu, Zn, Cr, Ni, As, Rossini Oliva and Fernández Espinosa, 2007; Sutherland et al., 2000).

399 The soluble highly mobile elements such as alkali and especially alkaline-earth elements, Sb, Mo, W and U
400 demonstrated an increase in their dissolved (< 0.45 µm) concentration with the increase in the total particulate fraction
401 (Fig. 4 B). We interpret this increase in concentration, also correlated with pH_{snow water} increase (Fig. 4 A), as a result of
402 element leaching from soluble minerals such as calcite and dolomite. There was a positive (R² = 0.53, p < 0.05)
403 correlation between % of calcite in the particulate fraction of snow and Ca concentration in snow meltwater (not
404 shown). Therefore, we hypothesize that simultaneous mobilization of carbonate minerals and soluble elements from the
405 soil and rocks to the atmosphere occurs in southern, carbonate-rock bearing provinces where the winter aerosols are
406 generated. The generation of insoluble elements such as trivalent and tetravalent hydrolysates in dissolved fraction of
407 snow occurs independently of snow enrichment in solid particles. Indeed, the decrease, and not increase in insoluble
408 elements dissolved concentration with the increase in particle concentration (Fig. 4 C, D) suggests that these elements

409 are not desorbed or leached from mineral particles, either within **the origin** of aerosol formation or during snow melting
410 and filtration in the laboratory.

411 The majority of measured elements are transported in particulate rather than dissolved fraction in the snow water
412 (**Fig. 8**). This is in general agreement with the results of **other studies in** Scandinavia and Kola Peninsula (Reimann et
413 al., 1996), north-eastern European Russia (Walker et al., 2003) and on drifting ice in the northern Barents Sea (Gordeev
414 and Lisitzin, 2005). An interesting particularity of dissolved fraction of snow in WSL is the increase in soluble fraction
415 of Fe, Cu and **light** REEs north of 63°N (**Fig. 2**). A northward increase in concentration of Fe, La, Ce has been also
416 reported for rivers of the WSL sampled during this period of the year (Pokrovsky et al., **2016a**, see also **Fig. S6 of**
417 **Supplement**). We do not have a straightforward explanation for such a coincidence. **The mobilization** of Fe-rich
418 colloids, occurring in rivers of the northern part of WSL, is not expected to occur in the atmospheric aerosols, since the
419 DOC level in the latter is very low (0.9 ± 0.2 mg/L) and insufficient to stabilize the **dissolved+colloidal** Fe(III); besides,
420 the water surfaces of the north of WSL (thaw ponds and thermokarst lakes rich in **aqueous Fe(III)**) are fully frozen in
421 February and thus cannot generate aerosols. Given that there is no enrichment in particular fraction of Fe and REE in
422 snow collected north of 63°N (not shown), **the** leaching of Fe from snow particles to the soluble fraction of snow in the
423 north **of the** WSL is unlikely. **The atmospheric deposition of anthropogenic or lithogenic origin is also not expected**
424 **since there is no known anthropogenic source of Fe north of 63°N and the lithogenic Fe from underlying clays is not**
425 **possible: the mineral (clay, silt) horizon is permanently frozen in the continuous permafrost zone of the WSL.**

426 The enrichment of snow particulate fraction relative to the earth crust as shown by Al-normalized enrichment
427 coefficient (**Fig. S4**) can be understood via taking into account the particle concentration in snow and microscopic
428 observations. **We suggest that the clays supply most trace elements in the PF.** The atmospheric particles **are known to**
429 **exert** significant impact on soils and ground vegetation (Kabata-Pendias and Pendias, 1984; Rasmussen, 1998; Steinnes
430 and Friedland, 2006). In the case of WSL, the elementary composition of snow particulate fraction was compared with
431 three main reservoirs of elements **within the soil**, sampled over significant latitudinal profile, from 55°N to 68°N
432 (Stepanova et al., 2015). These reservoirs are averaged over full latitudinal range and include *i*) mineral fraction from
433 the bottom of the peat column; *ii*) depth-averaged peat column composition, and *iii*) *Sphagnum* mosses, collected in
434 ombrotrophic bogs, which receive their constituents essentially from the atmosphere (e.g., Santelman and Gorham,
435 1988).

436 The particularity of the northern part of western Siberia lowland is that the active (seasonally unfrozen) soil layer
437 is located within the organic (moss+peat) rather than mineral horizon; the latter is represented by poorly reactive sands
438 and clays (Baulin et al., 1967; Baulin, 1985; Tyrtikov, 1973, 1979). As a result, the surface waters drain essentially
439 organic part of the column which is very poor in lithogenic elements (Pokrovsky et al., 2015, 2016). The supply of
440 mineral particles from the snow therefore may significantly enrich the rivers and lakes in dissolved alkaline earths,
441 metal micronutrients, phosphorus and other elements given high reactivity of incoming silicate and carbonate grains in
442 acidic (pH < 3-4), organic-rich ($10 < \text{DOC} < 50$ mg/L) surface waters of Western Siberia. The degree to which such a
443 supply can lead to overestimation of the calculated chemical weathering export fluxes of cations in the permafrost zone
444 is not possible to quantify. Therefore, in view of the importance of atmospheric input of solid particles for mineral-poor,
445 peat bogs of western Siberia, **the** seasonal, year-round measurements of particulate atmospheric deposition in this region
446 are necessary.

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449

450 **4.2. Metal deposition in the Arctic and the effect of industrial centers and local pollution versus long-term**
451 **transfer on the amount and chemical composition of particulate fraction**

452 Regional background concentrations of dissolved metals in snow of Quebec, Canada are reported to be 1.1, 1.7,
453 and 1.6 mg/L_{meltwater} for Cu, Pb, and Zn, respectively (Telmer et al., 2004). The values for Cu and Pb are comparable
454 with average snow water concentration across the WSL (0.83 and 0.68, respectively) but the concentration of Zn in the
455 WSL snow is significantly higher (10.1±5.0 µg/L, excluding 3 contaminated samples near the Tomsk city). Background
456 concentrations of dissolved Cu, Pb, and Zn in snow of Alaskan Arctic are much lower (0.08, 0.09 and 1.2, respectively,
457 Snyder-Conn et al., 1997). In snow from background areas of north-eastern European Russia, the concentrations of
458 dissolved Cu are near at the same level as in snow from the WSL, whereas the concentrations of dissolved Pb and Zn
459 are 2 times lower (Walker et al., 2003). Concentrations of dissolved Cu and Zn in snow of NW Finland are few times
460 lower than in snow of WSL; concentrations of dissolved Pb are at the same level (Caritat et al., 1998).

461 Significant enrichment in Ni is known for the aerosols of the Arctic Ocean (Shevchenko et al., 2003). It may be
462 linked both to Ni transport from Norilsk and Kola smelters but also with Ni fractionation at the sea surface (Duce et al.,
463 1976). Ni concentration in snow water of the northern part of WSL significantly exceeds that in the thermokarst lakes.
464 The winter snow stock of dissolved Ni is several times higher than the river export of this element during spring flood
465 in the permafrost-bearing zone of the WSL, north of 60°N, and Ni concentration in snow particles exceeds up to 2
466 orders of magnitude its concentration in moss and peat of the territory.

467 The winter-time deposition of dissolved (< 0.45 µm) metals on the surface of northern part of the WSL can be
468 calculated taking into account the mean multi-annual volume of accumulated snow during 8 winter months (in mm of
469 snow water) and the average concentration of elements in February snow collected north of 64°N. The monthly
470 depositions of selected metals (µg m⁻² month⁻¹) on the north of the WSL in the form of snow are equal to 2.8, 12, 15,
471 210 and 0.9 for As, Ni, Pb, Zn, and Cd which is significantly higher than the values for winter deposition of insoluble
472 aerosols into the Russian Arctic (0.22, 0.74, 2.7, 1.3 and 0.056, respectively, Shevchenko et al., 2003). Only V exhibited
473 similar values of Arctic aerosol and snow deposition (0.71 and 0.96 µg m⁻² month, respectively).

474 The main source of mineral particles in the southern part of latitudinal profile (56–58°N) may be soils of steppe
475 and forest-steppe regions south of WSL, where the land is cultivated and the snow cover is relatively thin. The aeolian
476 transport of soil particles under these conditions may be efficient even in winter (Evseeva et al., 2003). The main source
477 of ash particles in southern part of the profile is the industry and transport of the city of Tomsk. There are numerous
478 studies of snow cover contamination by particles in the vicinity of Tomsk (Boyarkina et al., 1993; Yazikov et al., 2000;
479 Talovskaya et al., 2014). The concentration of particles in snow collected from 58°N to 61°N ranged between 0.85 and
480 5.72 mg/L which is comparable or slightly higher than the values reported for the Arctic snow cover (Darby et al., 1974;
481 Mullen et al., 1972; Nürnberg et al., 1994; Shevchenko et al., 2002, 2010). It is important that in this zone of low PF
482 concentration, combustion spheres, fly ash and black carbon of few µm diameters were dominating. This can explain
483 relatively low concentration of all TE at low PF concentration, as carbon compounds likely contain very low proportion
484 of trace metals. The most important sources of fly ash and black carbon are gas flaring, land transport, heating plants,
485 residential combustion, forest fires (mainly in summer) and industrial plants (Moskovchenko and Babushkin, 2012;
486 Quinn et al., 2008; Stohl et al., 2013). Chemical pollution of atmosphere during gas flaring associated with oil industry
487 is known for the WSL (Raputa, 2013; Yashchenko et al., 2014). The black carbon produced during gas burning is
488 detected not only in western Siberia but in the Russian sector of the Arctic Ocean in high latitudes (Stohl et al., 2013).

489 In the zone 62–64.5°N, where some impact of oil industry is possible, the concentration of insoluble particles in
490 snow were above 10 mg/L, achieving the value of 66.6 mg/L in sample SF36. Backward trajectories to this site using
491 Draxler and Rolf (2003) approach show that last few days before sampling air masses arrived from south-western

492 direction. Accordingly, the particulate fraction in these samples contained mostly mineral particles 1–25 µm size with
493 some fly ash (burning spheres), black carbon. It is possible that mineral particles are supplied here via long-range
494 transport from forest-steppe, steppe and semi-desert regions south and south-west from the study site. Indeed, during
495 winter snow coverage period, the dominant winds in this zone have S, SW and W directions (Moskovchenko and
496 Babushkin, 2012). The events of mineral dust transport over large distances are well known in the boreal zone (Lisitzin,
497 1978, 2011; Shevchenko et al., 2010).

498 Further north of studied latitudinal profile, from 65 to 68°N, the concentration of snow particles ranged from 0.8
499 to 9.2 mg/L. These values are within the background in the Arctic and subarctic (Darby et al., 1974; Mullen et al., 1972;
500 Nürnberg et al., 1994; Shevchenko et al., 2002). The particulate fraction was represented by mineral debris of 1 to 15
501 µm in size, with frequent but not significant presence of spherical ash particles, biogenic strains and porous carbon
502 particles. Because the main source of mineral particles is long-range transport from southern desert and steppe regions,
503 moving to the north decrease the influence of these provinces.

504 We believe that the elevated concentrations of divalent metals, As and Sb should not be interpreted as
505 necessarily pollution from the industrial centers. Rather, volatile Pb, Cd, As may originate from long-range transport of
506 desert material. Therefore, we attempted to distinguish the well-known refractory, non-volatile heavy metals such as
507 Cu, Ni and Co and more volatile elements such as Pb, Cd and As (i.e., Reimann et al., 2000) based on the HCA
508 treatment. For both particulate and dissolved fraction, these elements are located in three or two different groups but
509 never belong to one single group of inter-correlated elements. As such the available data do not evidence similar origin
510 of Cu, Ni and Co, or Pb, Cd, and As in the snapshot of WSL snow sampled in this work.

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512

513 **4.3. Impact of dissolved fraction of snow input on hydrochemistry of lakes and rivers and riverine elementary** 514 **fluxes.**

515 Quantitative comparison of element input to the land surface with winter snow and element export from the
516 watershed of WSL rivers provided the assessment of minimal atmospheric contribution to riverine fluxes. For this
517 comparison, we ignored: *i*) prior melt history of the snow cover, because the freshly fallen snow could be subjected to
518 minimal transformation; *ii*) the efficiency of snow sample to capture the total chemical load transferred from the
519 atmosphere into the snow pack, because only the snow water is postulated to contribute to the riverine flux; and *iii*) to
520 what extent the snowmelt interacts with the topsoil and vegetation litter through which the melt flows into the river.

521 The comparison of dissolved element concentration in lakes and snow demonstrated that the snow loading
522 clearly exceeds the concentrations in lakes of Pb, Zn, Cu, Cd, Sb and Mo (Fig. 5). However, these elements belong to 4
523 various group of elements in dissolved snow fraction, identified by the HCA (Fig. 3 B). In rivers, SO₄, Cr, Co, Ni, Cu,
524 Zn, Mo, Cd, Sb, Cs, W and Pb are dominated by snow input. These elements also belong to 4 various group of the
525 HCA. It thus can be concluded that there is no direct link between the group of elements identified by cluster
526 dendrogram in the snow water and the elements those concentration in rivers or lakes are significantly affected by snow
527 deposition. We believe that a natural cause of this apparent inconsistency is different mechanisms controlling the
528 element distribution in the aerosols (local sources of pollution, remote desert provinces, leaching of soluble elements
529 from particulate fraction) and in surface waters (interaction of melted snow with upper peat and moss/lichen horizons;
530 underground feeding, release of elements from silicate river suspended matter due to abrasion in spring flood).

531 Overall, the impact of the snowmelt on chemical composition of western Siberian thermokarst lakes may be very
532 high. This will be further accentuated by reported increase in the proportion of meltwater that does not reach the main
533 rivers but is stored by the wetlands (i.e., from 20-30% in early 1990s to 50-60% in the mid-2000s, Zakharova et al.,

534 2011). A comparison of snow stock/river water fluxes demonstrates the increase in the influence of atmospheric
535 deposition northward (Figs. 6 and S3 of Supplement). At the same time, the chemical composition of the snow water,
536 although subjected to significant variation, does not exhibit any systematic trend with the latitude (Fig. 2) as follows
537 from the PCA (section 3.1). The reason for this difference may be relatively low fluxes and concentrations in rivers of
538 the northern, permafrost-affected territory of the WSL compared to the southern, permafrost-free zone (Pokrovsky et
539 al., 2015, 2016a). As a result, the impact of atmospheric deposition on the riverine transport is more pronounced in the
540 permafrost zone than in the permafrost-free zone. We expect this effect to be quite general for flat bog tundra areas of
541 northern Eurasia, including, in addition to northern part of western Siberia (~400,000 km²) studied in this work, the
542 Yamal and Gyda Peninsula (122,000 and 160,000 km², respectively), the North-Siberian Lowland (~700,000 km²), the
543 Kolyma Lowland (170,000 km²), and the Yana-Indigirka Lowland (180,000km²) with overall territory close to 1.7
544 million km². The impact of snow deposition on river elementary fluxes should be much lower in permafrost-bearing
545 mountainous terrain such as Central and Eastern Siberia, the Alaskan slopes, north of Scandinavian shield and the
546 Canadian High Arctic. In those territories, two factors may decrease the contribution of snow deposition to river fluxes:
547 1) the impact of local mineral dust for aerosols generation may be well pronounced and 2) the chemical weathering
548 occurs within the mineral seasonally unfrozen layer producing higher fluxes of inorganic components.

549 In contrast, in the lowlands of Northern Eurasia, the rivers drain essentially organic layer (peat bog) terrain, thus
550 mineral feeding of rivers is really low. As it is demonstrated in section 3.2.2 of this study, low chemical (cationic)
551 weathering in the north of the WSL during spring suggests that TDS_c and DIC fluxes in May-June in this and other
552 similar regions are essentially controlled by snowmelt, rather than by weathering. It follows that during the spring
553 period, the intensity of chemical weathering in these latitudes can be a factor of 2 (major cations) to 5 (TE) lower than
554 that deduced from riverine fluxes. However, given that the shares of spring flood period (May-June) in the annual
555 export fluxes are only 5 to 10% for major cations and 10 to 20% for TE (Pokrovsky et al., 2015, 2016), the overall
556 impact of atmospheric deposits on element export fluxes will be strongly pronounced (i.e., ≥ 50% of total measured
557 river flux value) only for elements which have the ratio of the spring-time river export to snow stock less than 0.2, i.e.,
558 SO₄, Cu, Mo, Cd, Sb, Cs, W and Pb. With further increase of winter precipitation in western Siberia (i.e., Bulygina et
559 al., 2009), the impact of snowmelt on element transport to the Arctic Ocean by rivers may increase thus enriching the
560 surface waters in many elements such as Cd, Pb, Sb, Cr, Cu, Ni, As, Mo, Rb, U.

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563 Conclusions

564 The chemical composition of surface layer of snow cover was studied across a 1700-km latitudinal gradient in
565 western Siberia Lowland. The particulate fraction ranged from 0.4 to 66 mg/L_{meltwater} and increased in the regions of
566 enhanced dust deposition from southern steppe and desert provenances, in the proximity of industrial centers and due to
567 fly ash production from gas burning of the oil exploration sites. The lowest concentrations of PF were measured in the
568 NW part of Tomskaya region and north of 66°N (< 10 mg/L, comparable with background values in the Arctic snow). A
569 PCA treatment of the elementary composition of snow water dissolved fraction demonstrated a F1 x F2 structure with the
570 first factor controlling the insoluble, low-mobile lithogenic elements and the 2nd factor acting on alkaline-earth metals,
571 biogenic elements and anions. None of the factors was linked to the latitude.

572 There was an increase in concentration of soluble elements (Ca, Mg, Sr, Mn, Co) and pH in snow water with the
573 increase in mineral (calcite, dolomite) fraction. The concentration of insoluble elements (trivalent and tetravalent
574 hydrolysates, Cr, Pb, Cd, Cu) in snow water did not change or decreased in response to the increase in PF. The elementary
575 composition of PF demonstrated its significant enrichment in most elements relative to mineral soil horizon, peat and moss

576 composition averaged across full latitudinal profile (~1700 km) of WSL. As such, solid atmospheric aerosols may be
577 important factor of insoluble element delivery to the soil surface. The supply of mineral particles from the snow may also
578 significantly enrich the rivers and lakes in dissolved alkaline earths, metal micronutrients, phosphorus and other elements
579 given high reactivity of incoming silicate and carbonate grains in acidic ($\text{pH} < 3\text{-}4$), organic-rich ($10 < \text{DOC} < 50 \text{ mg/L}$)
580 surface waters of Western Siberia.

581 The concentrations of Na, Mn, Zn, As, Rb and Sr in winter aerosols are similar (within a factor of 2) to lake
582 water concentrations during spring period. Concentrations of DIC, Cl, SO_4 , Mg, Ca, Cr, Co, Ni, Cu, Mo, Cd, Sb, Cs, Pb
583 and U in filtered snow water are close or higher than those in lakes. In the southern, permafrost-free zone, only Zn, Cd,
584 W, Pb, Cs and Sb fluxes in rivers during May-June period can be provided by dissolved fraction of the snow melt.
585 However, the impact of snow melt on river export fluxes in spring strongly increases northward for DIC, Cl, SO_4 , Na,
586 Mg, Ca, Cd, Pb, Sb, Cr, Cu, Ni, As, Mo, Rb, U. In the permafrost zone, $\geq 50\%$ of riverine fluxes of these elements
587 during spring flood can be provided by the snowmelt. The reason for such high sensitivity of WSL surface reservoirs to
588 atmospheric deposition is feeding of surface waters by essentially organic (moss, peat) soil profiles.

589 The snow deposition of mineral particles on the moss cover developed over the frozen peat in the north of WSL
590 will be mostly pronounced for Sb, Zn, Ni and Cd and in a lesser degree for Mg, Ca, Pb, Mo, and As, since these
591 elements are impoverished in mineral horizons of the WSL. However, quantifying the degree of these changes require a
592 year-round monitoring of liquid and solid atmospheric deposits across the WSL territory. We foresee a possibility to
593 apply the mass balance calculations of atmospheric input to the land surface of other Siberian lowlands of peat bog and
594 thermokarst lake zones, with an overall territory close to 1.7 million km^2 .

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Data availability

598 Full data set of major and trace element concentration in snow water ($< 0.45 \mu\text{m}$) and snow particles sampled
599 across the latitudinal profile of Western Siberia Lowland is available at the Research Gate,
600 <https://www.researchgate.net/publication/309666956>; DOI: 10.13140/RG.2.2.12156.54408.

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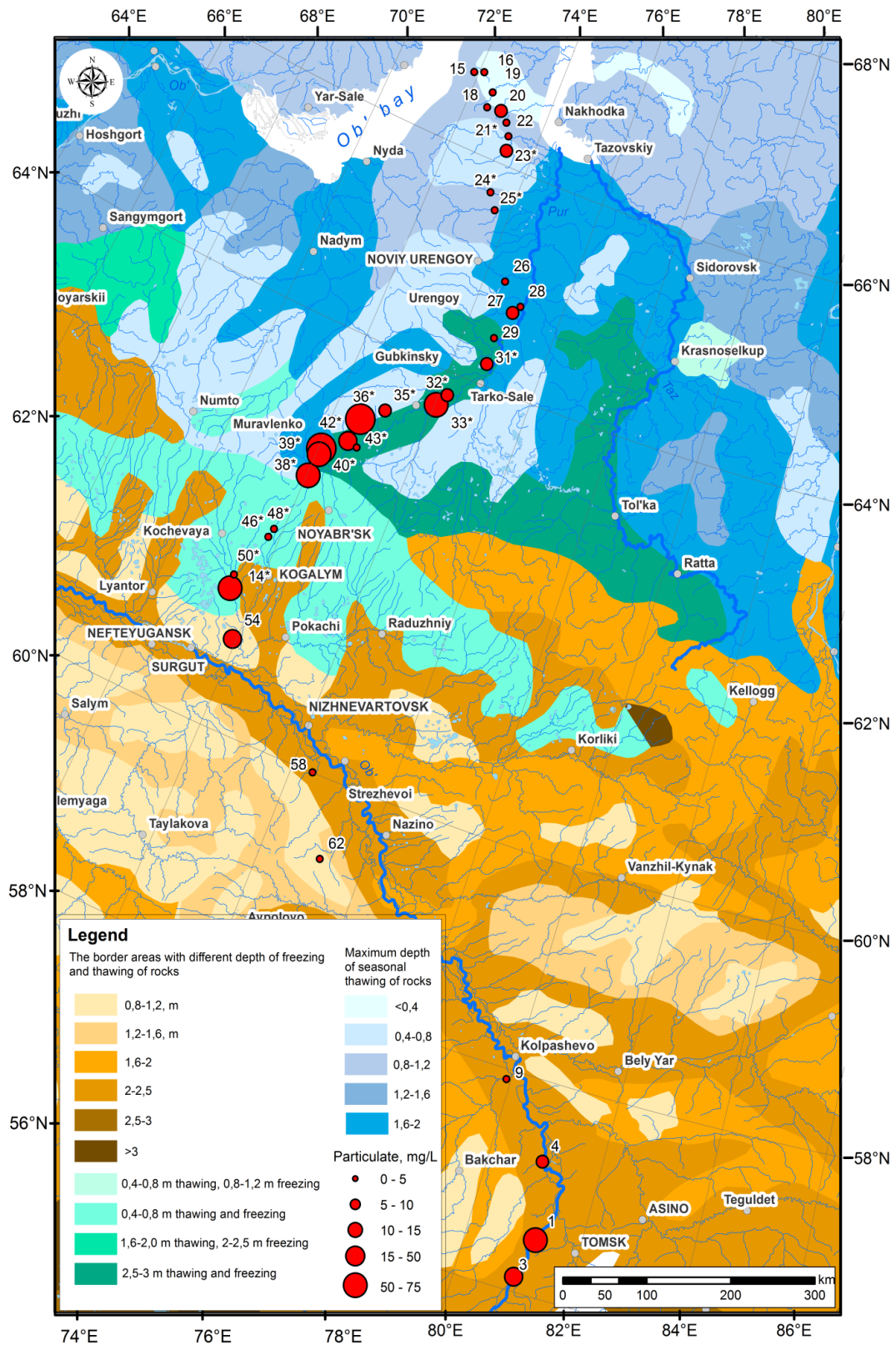
927 **Table 1.** Minimal, maximal, median and geometric mean concentration of dissolved ($\mu\text{g L}^{-1}$ snow water), n=35, and particulate
 928 ($\mu\text{g g}^{-1}$ particles), n=34 snow components. The data for upper continental crust (UPC) are from (Rudnick and Gao, 2003). N.A.
 929 stands for non analyzed.

| Element | Dissolved | | | | Particulate | | | | UPC |
|---|-----------|--------|--------|----------------|-------------|--------|--------|----------------|-------|
| | Min | Max | Median | Geometric mean | Min | Max | Median | Geometric mean | |
| pH | 4.38 | 8.73 | 5.11 | 5.44 | N.A. | N.A. | N.A. | N.A. | N.A. |
| SC, $\mu\text{S cm}^{-1}$ | 9 | 35 | 15.5 | 16.3 | N.A. | N.A. | N.A. | N.A. | N.A. |
| DIC, mg/L | 0.26 | 2.12 | 0.37 | 0.47 | N.A. | N.A. | N.A. | N.A. | N.A. |
| DOC, mg/L | 0.46 | 1.87 | 0.84 | 0.85 | N.A. | N.A. | N.A. | N.A. | N.A. |
| Cl, mg/L | 0.07 | 2.94 | 0.51 | 0.48 | N.A. | N.A. | N.A. | N.A. | N.A. |
| SO₄, mg/L | 0.41 | 2.01 | 0.71 | 0.72 | N.A. | N.A. | N.A. | N.A. | N.A. |
| Li | N.A. | N.A. | N.A. | N.A. | 2.6 | 32.2 | 10.8 | 10.7 | 24 |
| Be | N.A. | N.A. | N.A. | N.A. | 0.12 | 2.11 | 0.59 | 0.59 | 2.1 |
| Na | 47 | 1982 | 295 | 303 | 1452 | 39156 | 6717 | 7314 | 24200 |
| Mg | 19 | 862 | 114 | 114 | 3492 | 156712 | 19089 | 21411 | 14900 |
| Al | 1.6 | 35.2 | 15.5 | 12.3 | 6444 | 138267 | 31079 | 31565 | 81500 |
| P | N.A. | N.A. | N.A. | N.A. | 70 | 1928 | 481 | 503 | 660 |
| Si | 3.5 | 180 | 64.6 | 33.2 | N.A. | N.A. | N.A. | N.A. | N.A. |
| K | 39.2 | 120 | 55.5 | 63.0 | 1682 | 38395 | 5895 | 6023 | 23200 |
| Ca | 57 | 2266 | 267 | 296 | 3944 | 159272 | 17331 | 17775 | 25600 |
| Ti | 0.001 | 0.338 | 0.032 | 0.018 | 194 | 5762 | 674 | 689 | 3800 |
| V | 0.007 | 0.221 | 0.051 | 0.049 | 23.8 | 322 | 67.4 | 69.7 | 97 |
| Cr | 0.027 | 0.340 | 0.111 | 0.117 | 43.8 | 841 | 138 | 156 | 92 |
| Mn | 0.62 | 9.54 | 3.06 | 2.99 | 180 | 1242 | 400 | 404 | 780 |
| Fe | 1.8 | 62.2 | 14.6 | 12.0 | 7206 | 41255 | 15873 | 16488 | 39100 |
| Co | 0.006 | 0.418 | 0.097 | 0.094 | 5.9 | 60.7 | 19.4 | 18.6 | 17.3 |
| Ni | 0.04 | 5.66 | 0.36 | 0.36 | 28.1 | 1067 | 149 | 145 | 47 |
| Cu | 0.16 | 2.51 | 0.57 | 0.63 | 13.1 | 273 | 63.4 | 75.1 | 28 |
| Zn | 1.7 | 31.0 | 8.3 | 8.3 | 70.7 | 3832 | 202 | 255 | 67 |
| Ga | 0.0001 | 0.0185 | 0.0023 | 0.0014 | 1.8 | 26.7 | 8.32 | 7.73 | 17.5 |
| Ge | N.A. | N.A. | N.A. | N.A. | 0.36 | 3.18 | 0.88 | 0.91 | 1.4 |
| As | 0.02 | 0.46 | 0.19 | 0.15 | 3.8 | 67.2 | 16.1 | 16.0 | 4.8 |
| Rb | 0.033 | 0.262 | 0.066 | 0.075 | 6.1 | 124 | 24.1 | 23.4 | 84 |
| Sr | 0.26 | 10.2 | 1.04 | 1.23 | 26.2 | 580 | 117 | 115 | 320 |
| Y | N.A. | N.A. | N.A. | N.A. | 1.52 | 40.2 | 7.1 | 7.5 | 21 |
| Zr | 0.0001 | 0.0403 | 0.0024 | 0.0015 | 7.7 | 383 | 36.8 | 38.5 | 193 |
| Nb | N.A. | N.A. | N.A. | N.A. | 0.80 | 25.5 | 3.54 | 3.72 | 12 |
| Mo | 0.0001 | 0.059 | 0.010 | 0.005 | 0.55 | 10.4 | 2.12 | 2.24 | 1.1 |
| Cd | 0.015 | 0.180 | 0.047 | 0.046 | 0.11 | 3.37 | 0.71 | 0.71 | 0.09 |

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|-----------|---------|---------|---------|---------|-------|------|------|------|------|
| Sn | N.D. | N.D. | N.D. | N.D. | 1.13 | 29.3 | 7.76 | 7.36 | 2.1 |
| Sb | 0.009 | 0.132 | 0.036 | 0.038 | 1.67 | 27.2 | 5.96 | 6.15 | 0.4 |
| Cs | 0.0015 | 0.0105 | 0.0034 | 0.0036 | 0.32 | 4.78 | 1.35 | 1.24 | 4.9 |
| Ba | 0.74 | 13.6 | 3.35 | 3.32 | 88 | 1664 | 374 | 391 | 628 |
| La | 0.001 | 0.049 | 0.012 | 0.011 | 2.0 | 60.2 | 10.3 | 10.8 | 31 |
| Ce | 0.003 | 0.095 | 0.022 | 0.019 | 4.05 | 128 | 19.0 | 20.6 | 63 |
| Pr | 0.0001 | 0.0084 | 0.0022 | 0.0019 | 0.50 | 15.5 | 2.30 | 2.35 | 7.1 |
| Nd | 0.0013 | 0.0275 | 0.0085 | 0.0067 | 1.86 | 58.6 | 8.32 | 8.70 | 27 |
| Sm | 0.0001 | 0.0072 | 0.0020 | 0.0016 | 0.39 | 11.8 | 1.78 | 1.79 | 4.7 |
| Eu | 0.00010 | 0.00253 | 0.00096 | 0.00083 | 0.11 | 2.56 | 0.45 | 0.47 | 1.0 |
| Gd | 0.0004 | 0.0082 | 0.0022 | 0.0022 | 0.40 | 10.3 | 1.71 | 1.77 | 4.0 |
| Dy | 0.00002 | 0.0041 | 0.0016 | 0.0008 | 0.32 | 7.83 | 1.35 | 1.42 | 3.9 |
| Ho | 0.00006 | 0.00123 | 0.00061 | 0.00054 | 0.06 | 1.51 | 0.26 | 0.27 | 0.83 |
| Er | 0.0002 | 0.0029 | 0.0010 | 0.0010 | 0.18 | 4.71 | 0.77 | 0.80 | 2.3 |
| Tm | 0.00002 | 0.00088 | 0.00011 | 0.00009 | 0.03 | 0.72 | 0.11 | 0.11 | 0.3 |
| Yb | 0.00000 | 0.00289 | 0.00089 | 0.00049 | 0.16 | 4.91 | 0.73 | 0.73 | 1.96 |
| Lu | N.A. | N.A. | N.A. | N.A. | 0.024 | 0.76 | 0.11 | 0.11 | 0.31 |
| Hf | N.A. | N.A. | N.A. | N.A. | 0.25 | 13.2 | 1.10 | 1.18 | 5.3 |
| Ta | N.A. | N.A. | N.A. | N.A. | 0.18 | 4.35 | 0.62 | 0.62 | 0.9 |
| W | 0.002 | 0.108 | 0.020 | 0.017 | 2.0 | 102 | 35.9 | 28.8 | 1.9 |
| Tl | N.A. | N.A. | N.A. | N.A. | 0.04 | 0.73 | 0.23 | 0.23 | 0.90 |
| Pb | 0.02 | 3.67 | 0.51 | 0.38 | 13.2 | 703 | 71.9 | 67.9 | 17 |
| Th | N.A. | N.A. | N.A. | N.A. | 0.43 | 17.1 | 2.22 | 2.33 | 10.5 |
| U | 0.0007 | 0.0063 | 0.0031 | 0.0028 | 0.19 | 4.69 | 0.92 | 0.93 | 2.7 |

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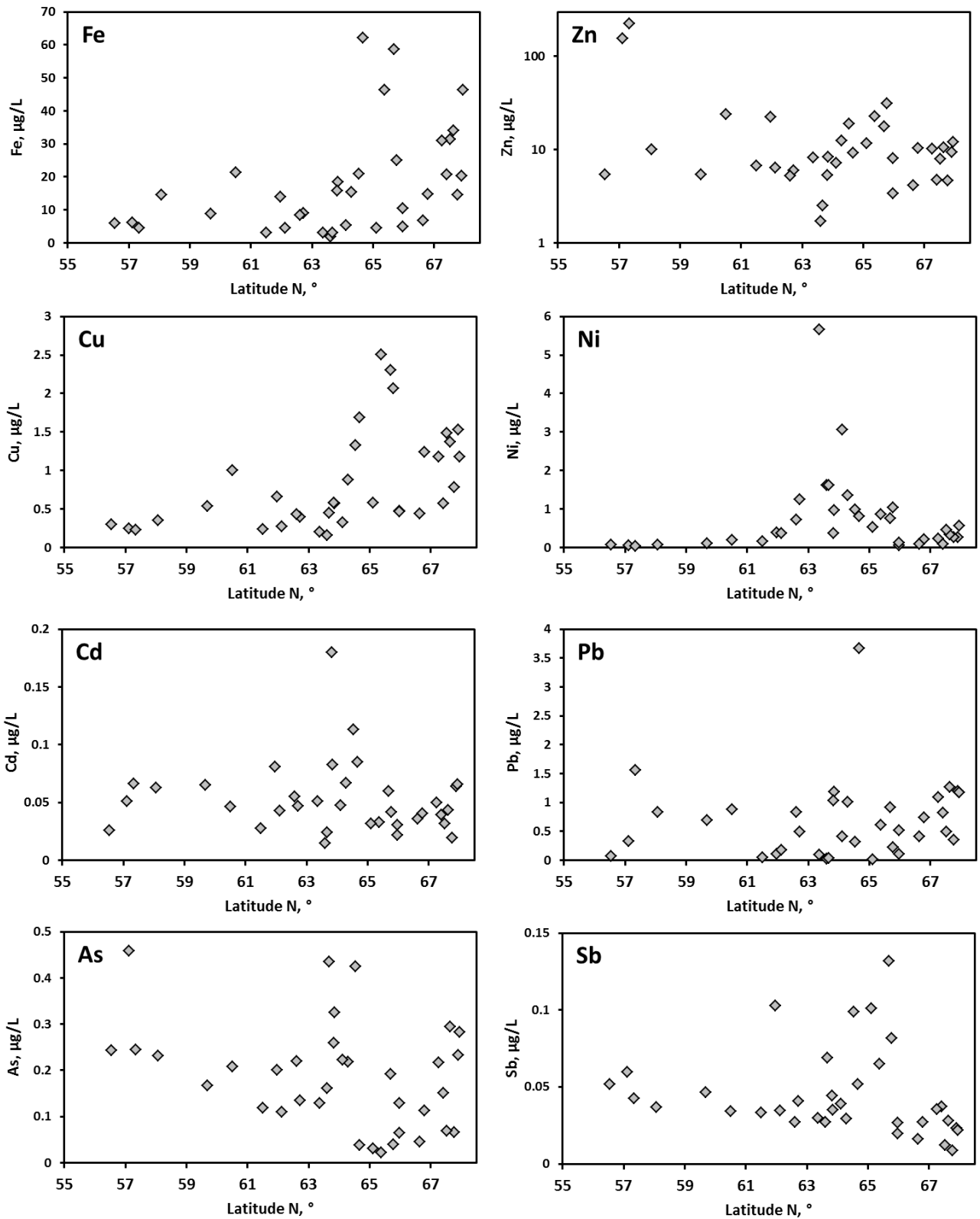
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934 **Figure 1.** Map of the study site. The size of the sampling points reflects the concentration of particulate fraction

935 (mg/L_{snow water})

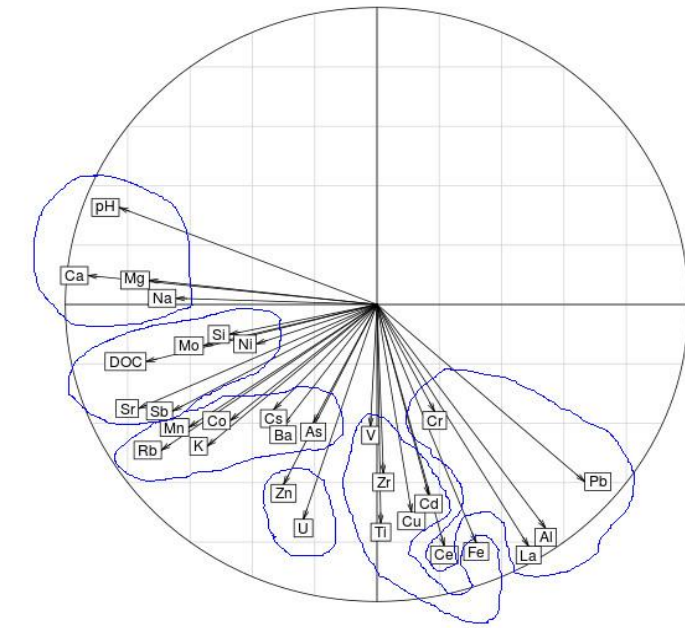
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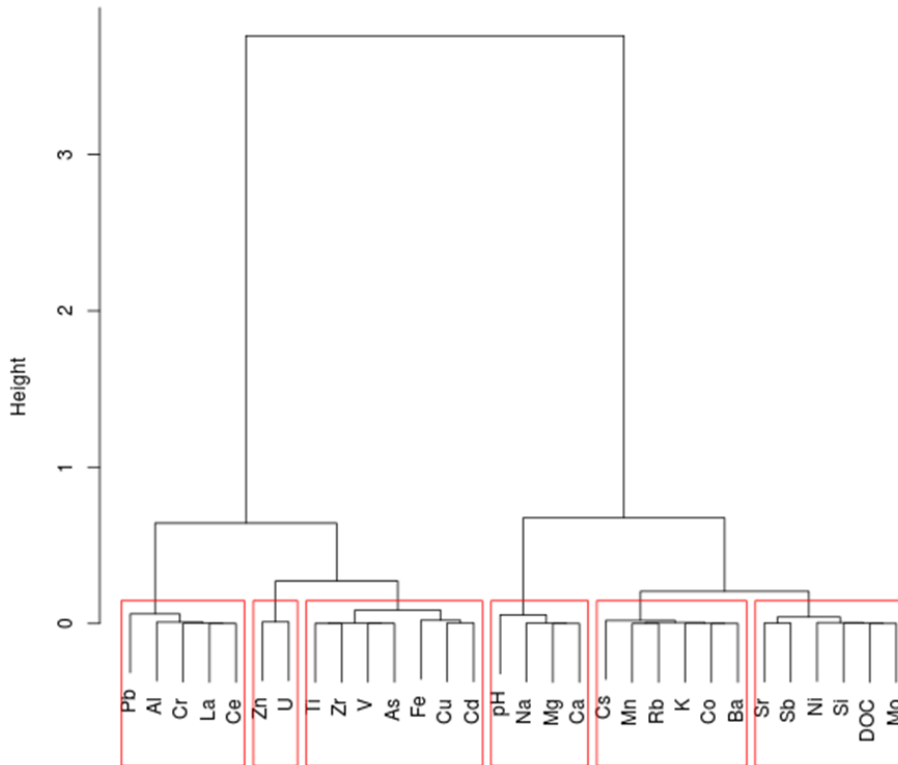
Fig. 2. Examples of dissolved ($< 0.45 \mu\text{m}$) metal concentrations in snow water as a function of latitude. The cause for the elevated concentrations of Ni, Cd and Sb at ca. 64-65°N is most likely industrial impact, but given relatively low number of data points around industrial centers it is hard to prove it unambiguously.

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dissolved fraction

Cluster Dendrogram



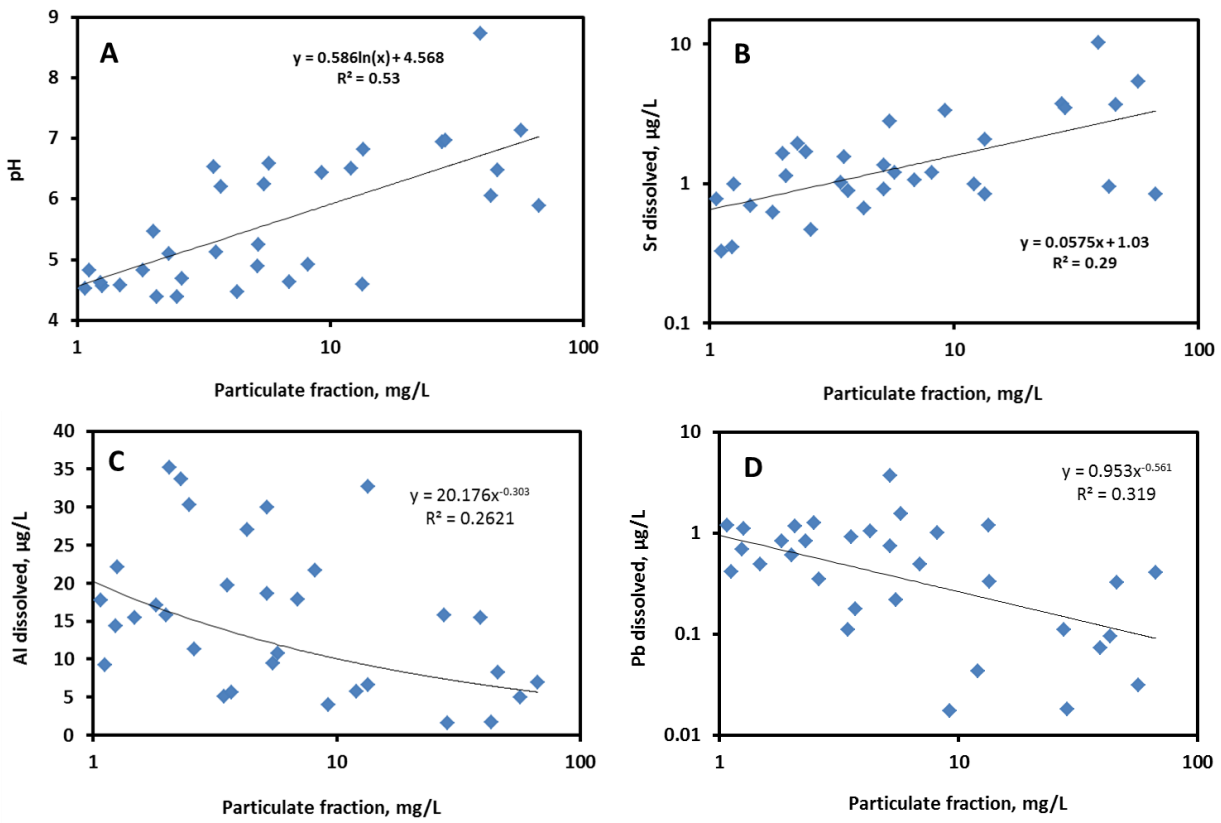
dissolved fraction

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Figure 3 A: PCA Factorial map F1x2 of elements of a reconstructed table for the dissolved fraction. Partition of elements into 6 groups revealed by a CAH is shown by a contour line. **B:** Dendrogram of a hierarchical cluster performed on variables of a reconstructed table for the dissolved fraction using the Pearson correlation as a distance measure and Ward's method for the linkage rule.

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970 **Figure 4.** pH value (A) and Sr (B), Al (C) and Pb (D) concentration in dissolved fraction of snow as a function of
971 concentration of particles. Note log X scale for Sr and Pb.

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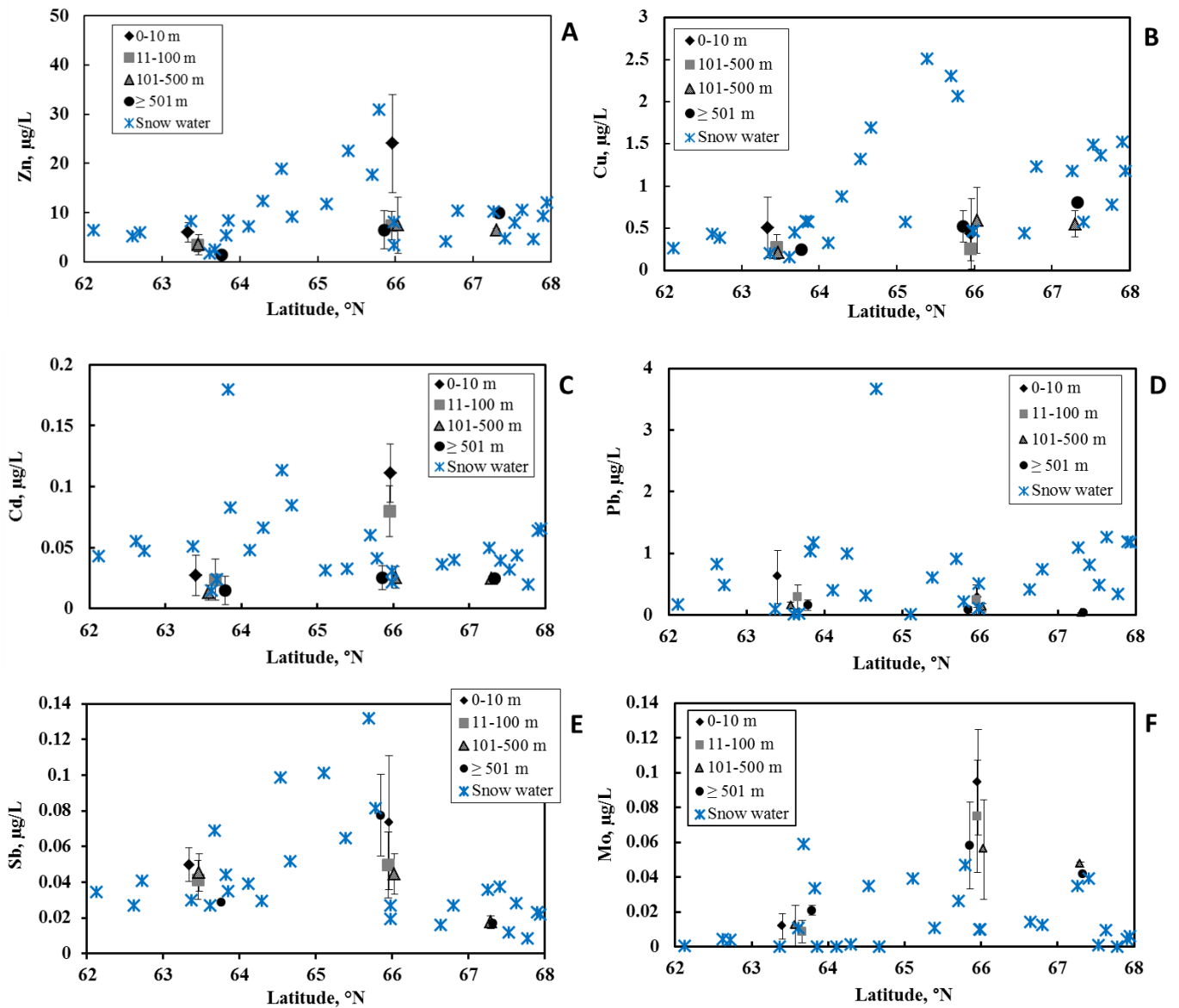
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988 **Figure 5.** Snow water soluble ($< 0.45 \mu\text{m}$) concentration of Zn (A), Cu (B), Cd (C), Pb (D), Sb (E) and Mo (F) (blue
989 asterisk) compared with average concentrations in thermokarst lakes of different size in western Siberia (black symbols)
990 along the latitudinal gradient. Diamonds, squares, triangles and circles represent the lakes of four diameters: 0-10, 10-
991 100, 100 to 500, and > 501 m, respectively. The error bars represent the 2 s.d. of mean concentration for at least 10
992 lakes.

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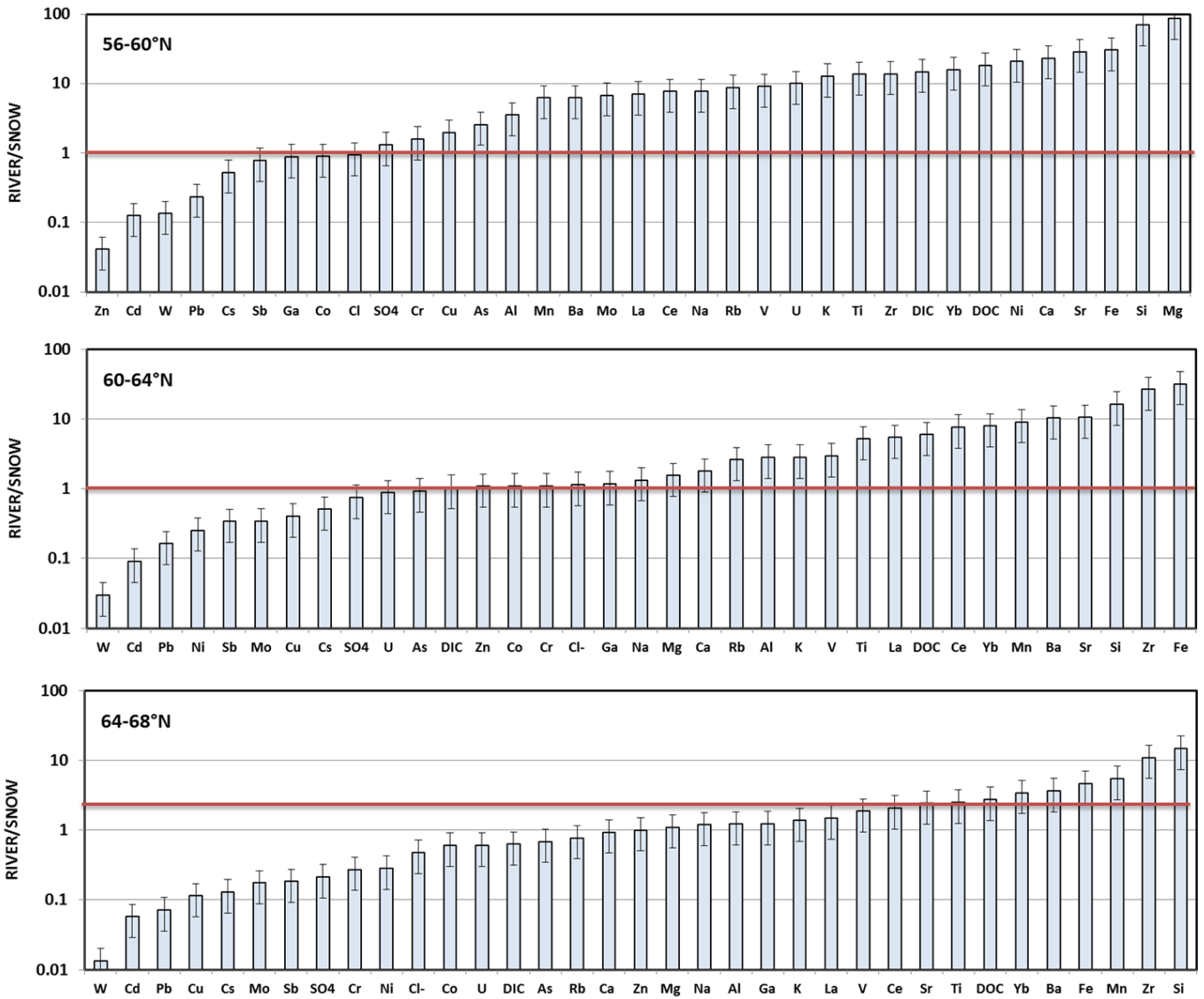
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1002 **Figure 6.** The ratio of mean dissolved flux of rivers in three latitudinal zones (56-60°N, 60-64°N, and 64-68°N) of

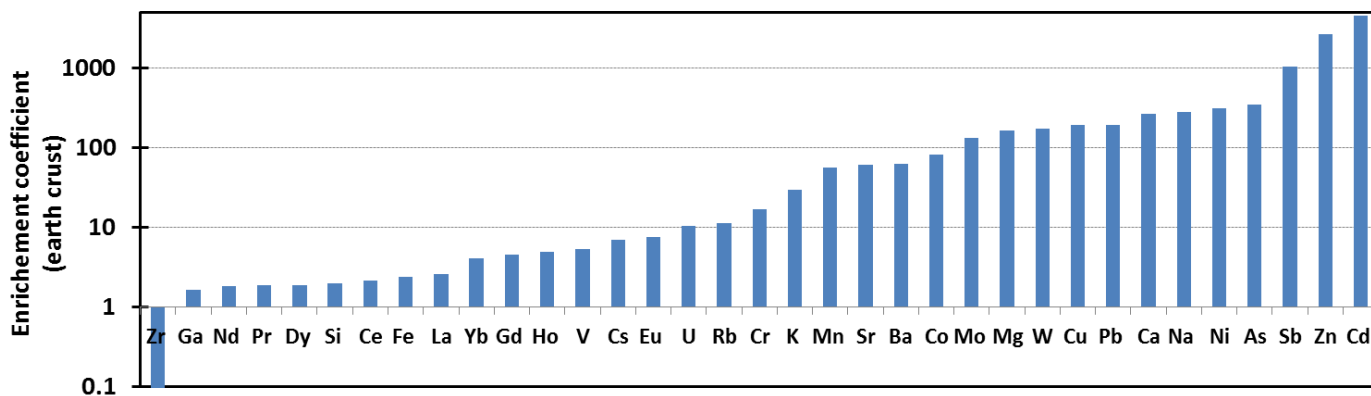
1003 WSL to the stock of dissolved fraction of snow. For this calculation, the snow volume (in mm of water) accumulated

1004 over full winter and mean river runoff over May and June were used.

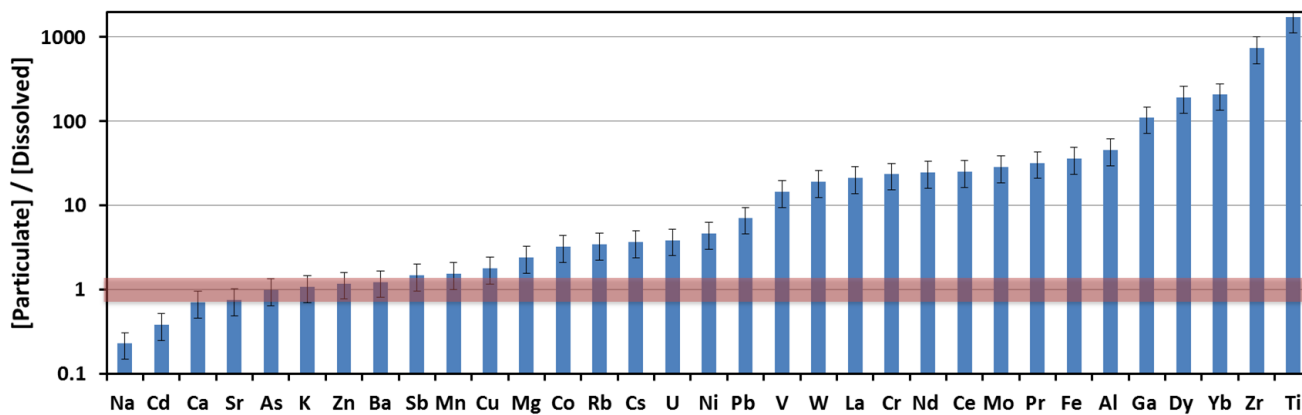
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5 **Figure 7.** The latitude-averaged Al-normalized enrichment coefficient of snow particles with respect to the earth crust.



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Figure 8. The average values (56 to 68°N) of the ratio of particulate to dissolved element concentration in snow water of western Siberia. Bold red line indicates statistically non-significant deviation from 1.

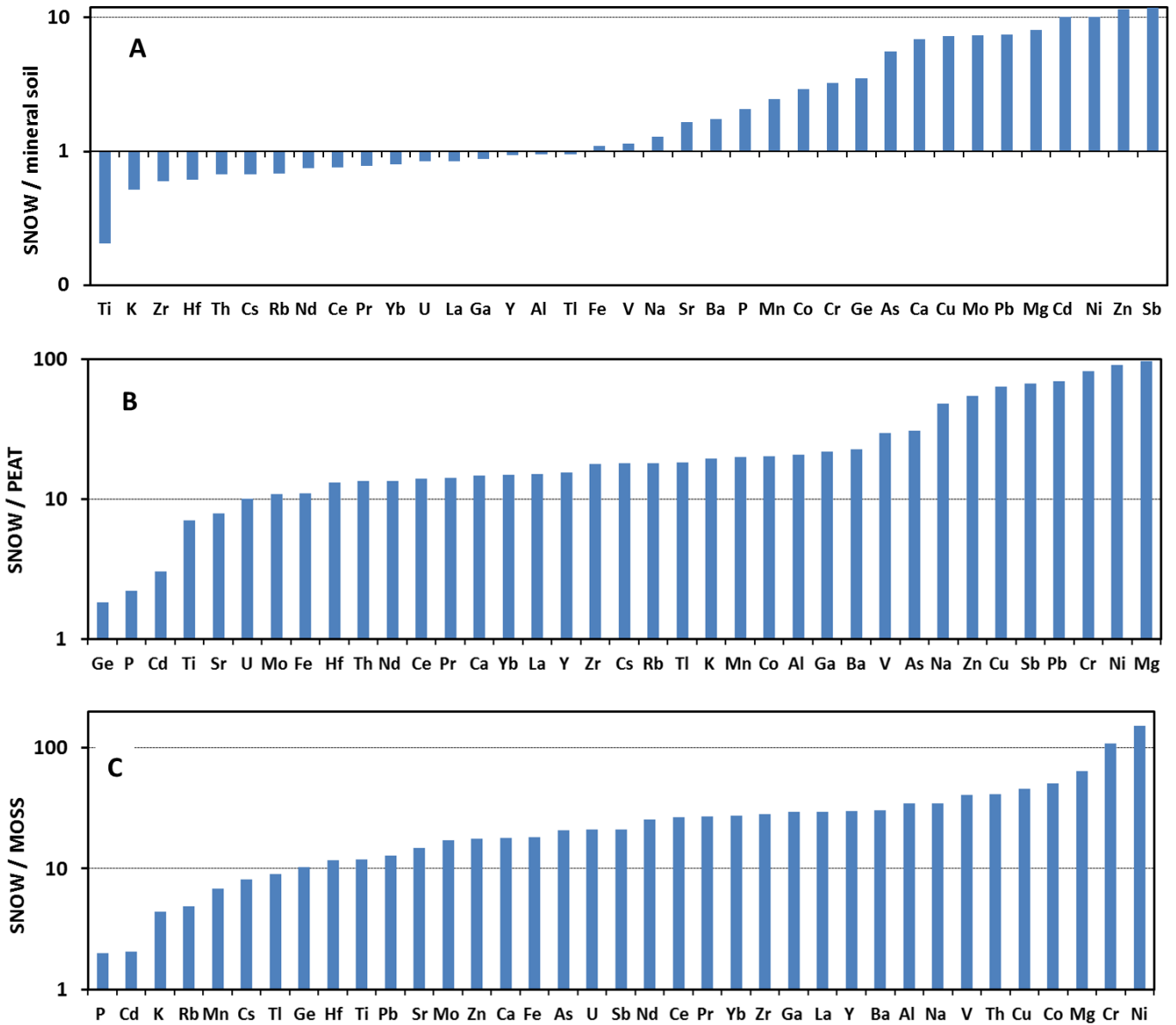
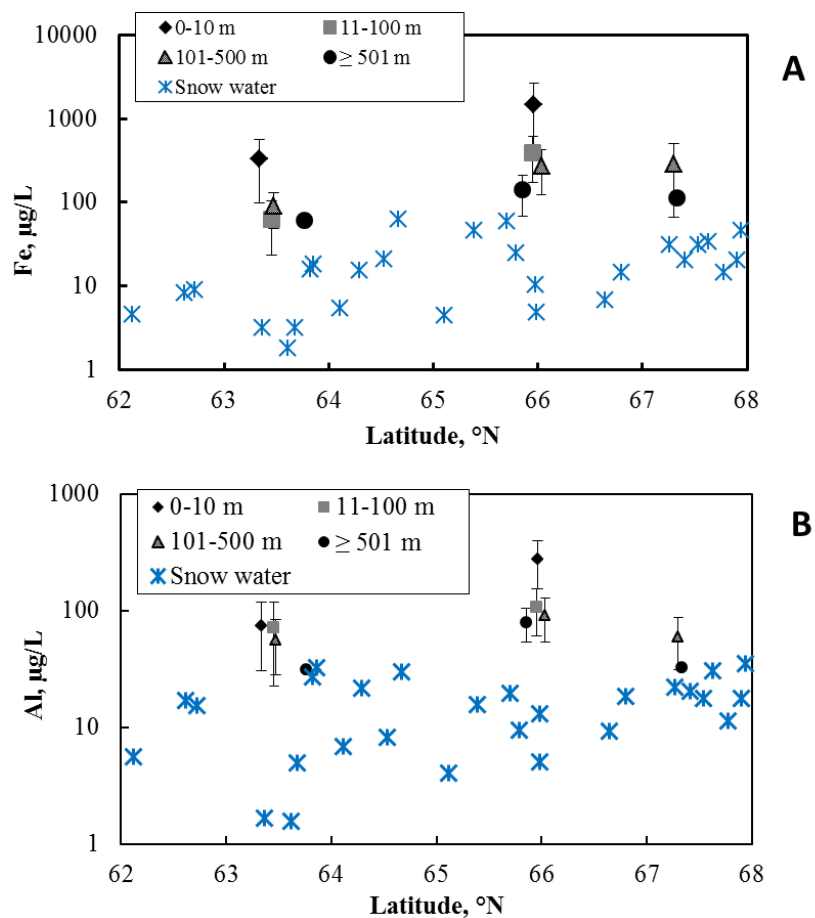


Figure 9. The ratios of the average concentrations of elements in snow particles (neglecting sample SF22) to those in mineral soil (A), peat (B) and mosses (C) of WSL. The peat, moss, and underlying mineral horizons data are averaged over the latitude of 55 to 68°N as described in Stepanova et al (2015). Note normal Y scale for mineral soil (A) and log Y scale for peat and moss (B, C).

SUPPLEMENTARY INFORMATION



5

Figure S1. Snow water soluble ($< 0.45 \mu\text{m}$) concentration of Fe and Al (blue asterisk) compared with average concentrations in thermokarst lakes of different size in western Siberia (closed symbols) along the latitudinal gradient.

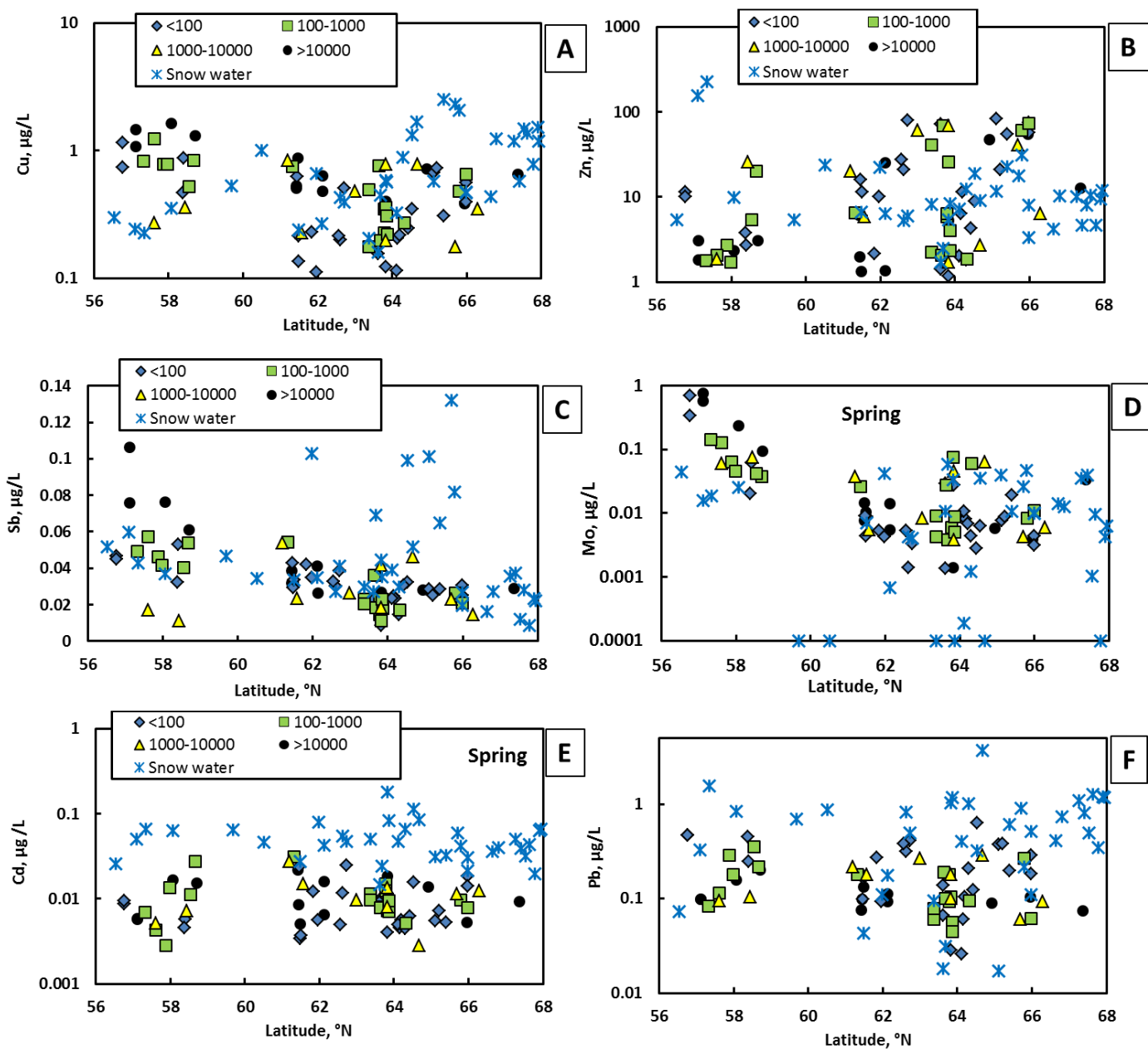
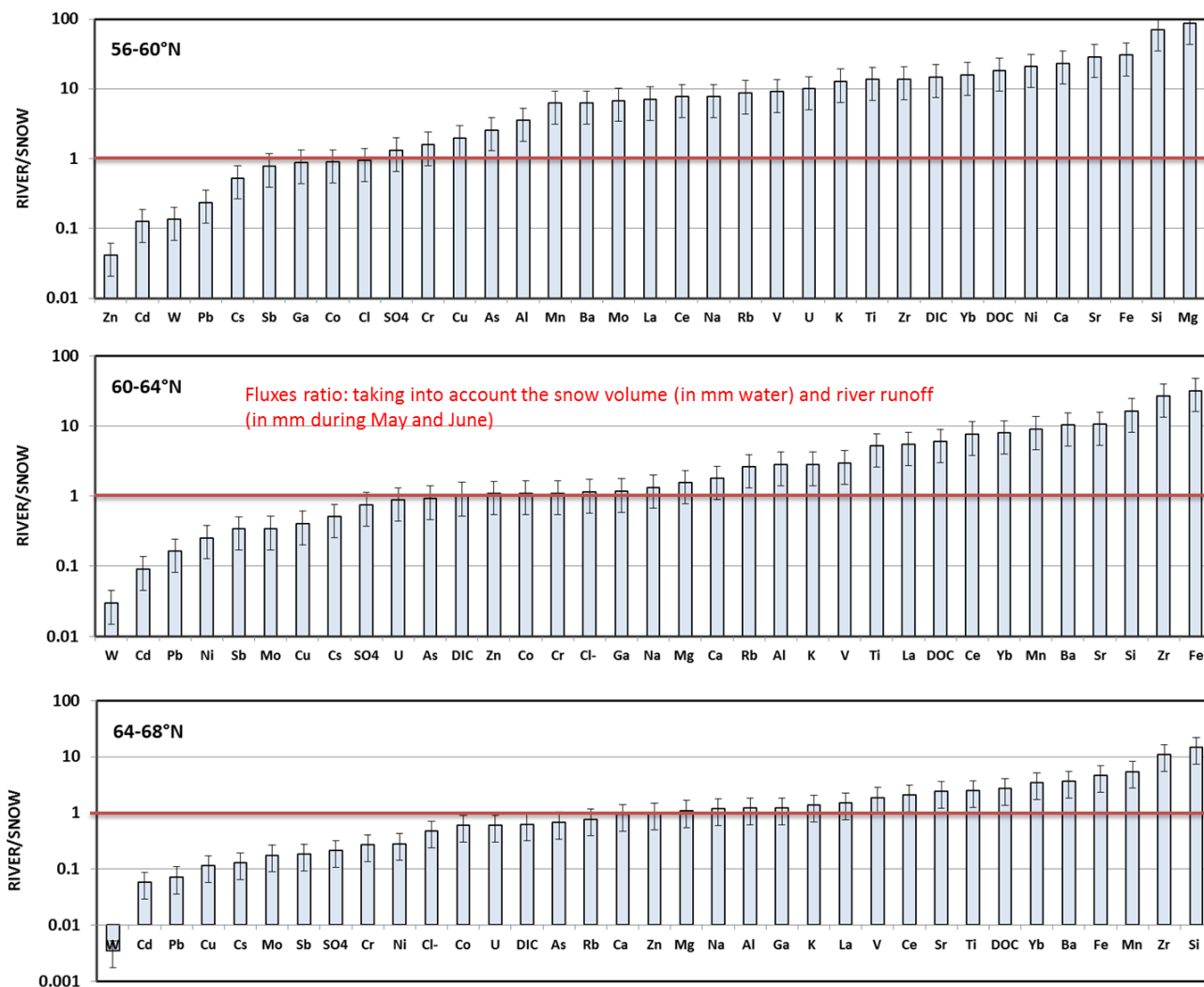


Figure S2. Snow water soluble ($< 0.45 \mu\text{m}$) of Zn (A), Cu (B), Cd (C), Pb (D), Sb (E) and Mo (F) (blue asterisk) compared with actual concentrations in rivers during spring flood (May-June) of different size of the watershed (closed diamonds, squares, triangles and circles correspond to < 100 , $100\text{-}1000$, $1000\text{-}10,000$ and $> 10,000$ km² surface area, respectively) in western Siberia along the latitudinal gradient.

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5 **Figure S3.** The ratio of mean dissolved flux of rivers in three latitudinal zones (56-60°N, 60-64°N, and 64-68°N) of WSL to the stock of dissolved fraction of snow. For this calculation, the snow volume (in mm of water) accumulated over full winter and mean river runoff in May and June were used.

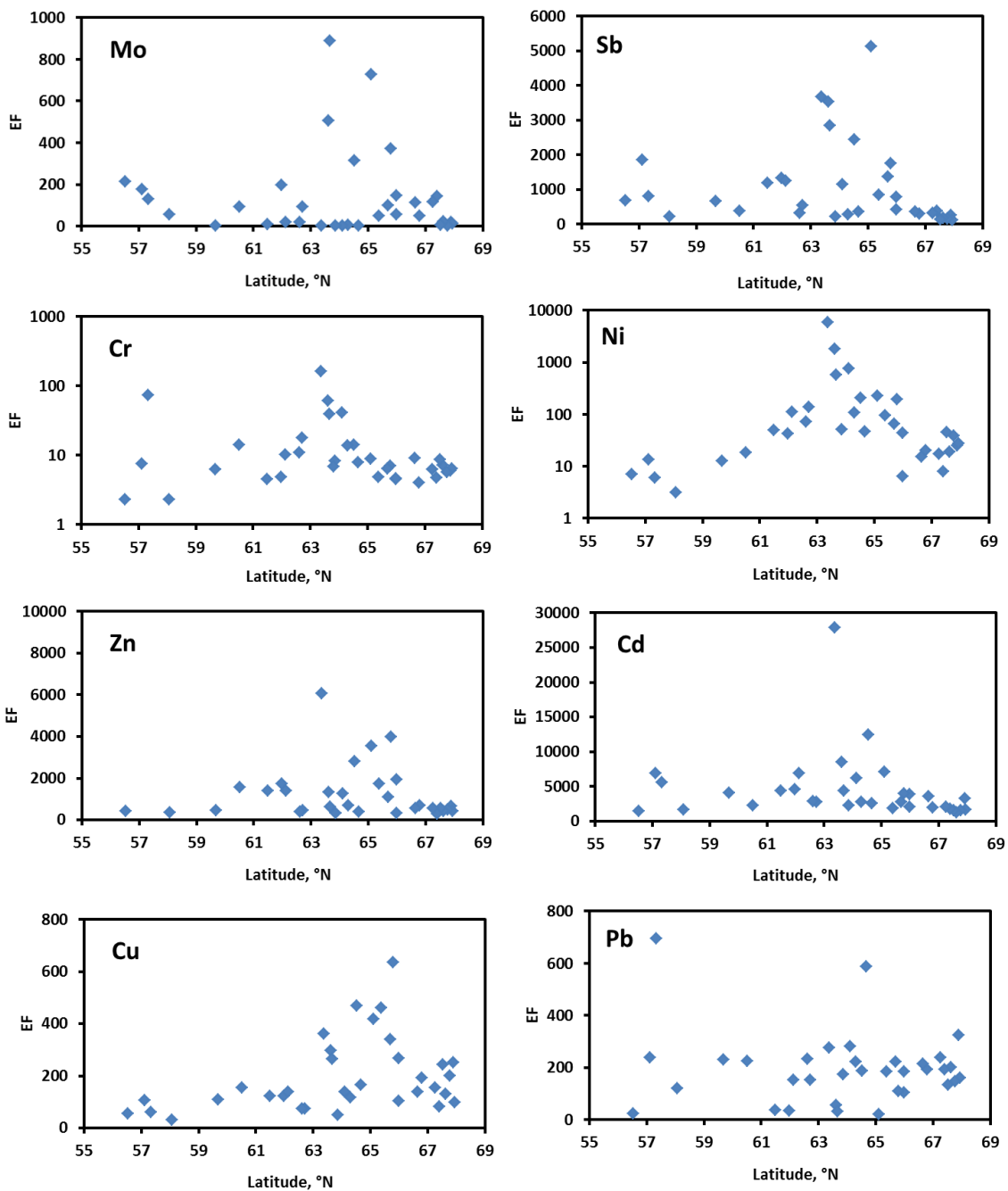


Figure S4. The enrichment factor (relative to the upper part of continental earth crust) of particulate fraction of snow for Mo, Sb, Cr, Ni, Zn, Cd, Cu and Pb as a function of latitude.

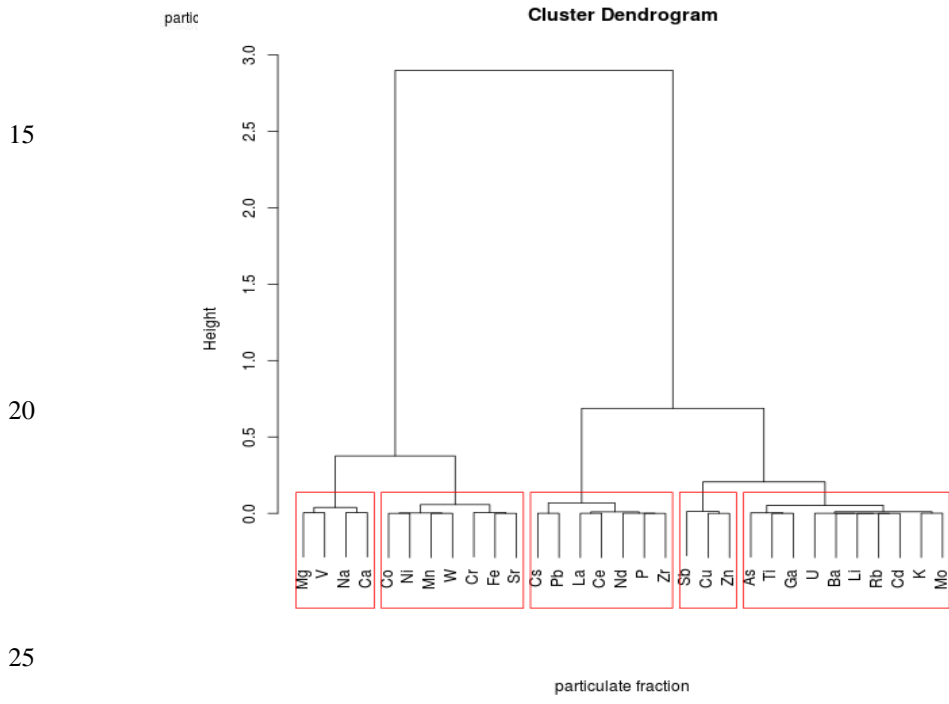
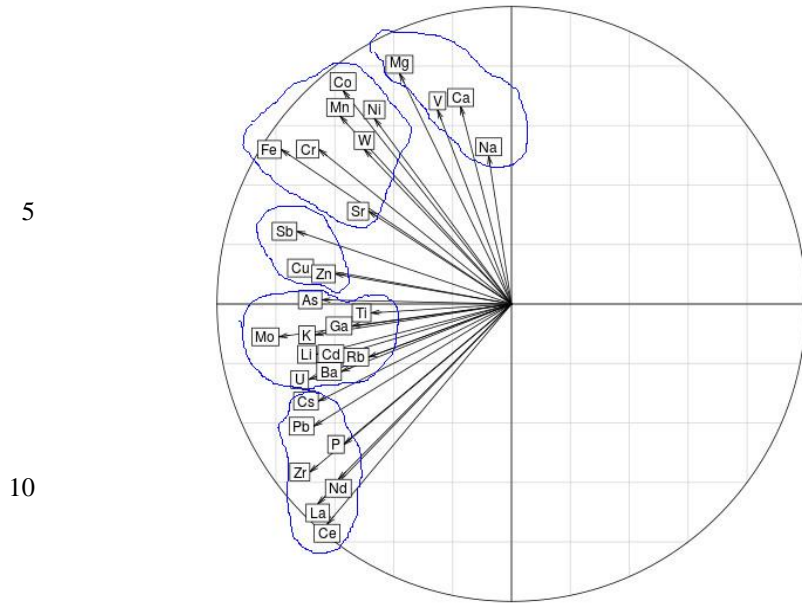
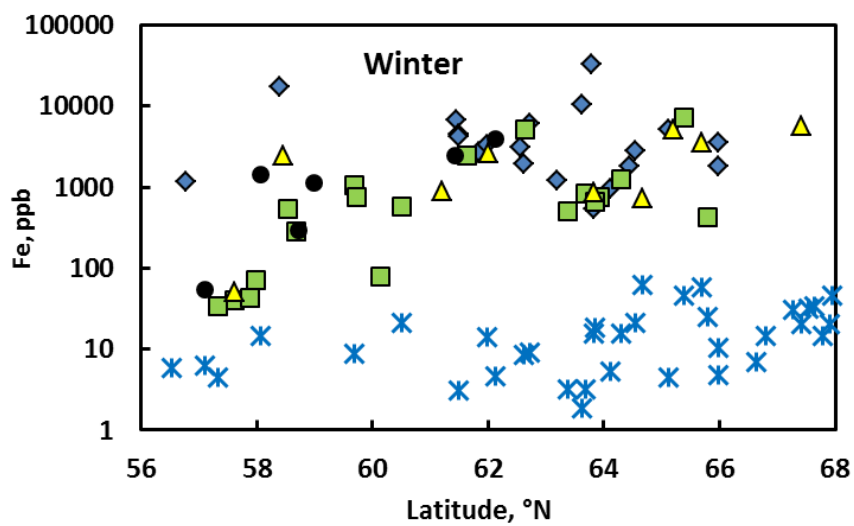


Figure S5 A. PCA Factorial map F1xF2 of variables (elements) of a reconstructed table for the particulate fraction. Partition of elements into 5 groups revealed by a CAH is reported by a contour line. **B:** Dendrogram of a hierarchical cluster performed on variables of a reconstructed table for the particulate fraction using Pearson correlation distance as distance measure and Ward's method for the linkage rule.



- 10 **Figure S6.** Comparison of Fe concentrations in rivers of different size watershed in snow water ($< 0.45 \mu\text{m}$) sampled in February- March 2014 across the WSL. See the legend for symbols in Fig S3. Note an increase in concentration north of $63\text{-}64^\circ\text{N}$, for both rivers and snow water.

5 **Table S1.** Mineralogical composition of selected snow particles.

| No | Quartz | Albite | K-Fs | Calcite | Dolomite | Chlorite | Illite | Phlogopite | Amphibole | Pyroxene | Chrysotile | Magnesite | Forsterite | Talc | Magnetite |
|-------|--------|--------|------|---------|----------|----------|--------|------------|-----------|----------|------------|-----------|------------|------|-----------|
| SF-1 | 37 | 20 | 11 | 19 | < 1 | 4 | 7 | | 2 | | | | | | |
| SF-3 | 38 | 27 | 9 | 3 | | 10 | 9 | | 3 | | | | | | |
| SF-14 | 20 | 8 | 6 | 1 | 48 | 5 | 4 | | 3 | 1 | 2 | | | 1 | |
| SF-31 | 30 | 20 | 12 | | 8 | 8 | | 12 | 6 | | 3 | | | | |
| SF-33 | 35 | 16 | 9 | | 10 | 3 | | 16 | 4 | | 3 | 4 | | | |
| SF-36 | 47 | 7 | 11 | | 1 | 4 | | 5 | 3 | 3 | 12 | | 4 | 3 | |
| SF-38 | 48 | 3 | 6 | | 1 | 6 | | 4 | 3 | | 16 | | 6 | 4 | 3 |
| SF-39 | 41 | 8 | 6 | 4 | 8 | 5 | | 12 | 2 | 2 | 8 | | 2 | 2 | |
| SF-40 | 35 | 12 | 9 | 6 | 3 | 4 | | 8 | 4 | 3 | 10 | | 4 | 2 | |