



1 **Simulating cold-region hydrology in an intensively drained agricultural watershed in**
2 **Manitoba, Canada, using the Cold Regions Hydrological Model**

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11 **Abstract**

12 Eutrophication and flooding are perennial problems in agricultural watersheds of the
13 northern Great Plains. A high proportion of annual runoff and nutrient transport occurs with
14 snowmelt in this region. Extensive surface drainage modification, frozen soils, and frequent
15 backwater or ice damming impacts on flow measurement represent unique challenges to
16 accurately modeling watershed scale hydrological processes. A physically-based, non-calibrated
17 model created using the Cold Regions Hydrological Modelling platform (CRHM) was
18 parameterized to simulate hydrological processes within a low slope, clay soil, and intensively
19 surface drained agricultural watershed. These characteristics are common to most tributaries of
20 the Red River of the North. Analysis of the observed water level records for the study watershed
21 (La Salle River) indicate that ice cover and backwater issues at time of peak flow may impact the
22 accuracy of both modeled and measured stream flows, highlighting the value of evaluating a
23 non-calibrated model in this environment. Simulations best matched the streamflow record in



24 years when peak and annual discharges were equal to or above the medians of $6.7 \text{ m}^3 \text{ s}^{-1}$ and 1.25
25 $\times 10^7 \text{ m}^3$, respectively, with an average Nash-Sutcliffe efficiency (NSE) of 0.76. Simulation of
26 low-flow years (below the medians) was more challenging (average NSE <0) with simulated
27 discharge overestimated by 90% on average. This result indicates the need for improved
28 understanding of hydrological response in the watershed under drier conditions. Improved
29 simulation of dry years was achieved when infiltration was allowed prior to soil thaw indicating
30 the potential importance of preferential flow. Representation of in-channel dynamics and travel
31 time under the flooded or ice-jam conditions should also receive attention in further model
32 development efforts. Despite the complexities of the watershed being modeled, simulations of
33 flow for average to high flow years and other components of the water balance were robust
34 [snow water equivalency (SWE) and soil moisture]. A sensitivity analysis of the flow routing
35 model suggests a need for improved understanding of watershed functions under both dry and
36 flooded conditions due to dynamic routing conditions, but overall CRHM is appropriate for
37 simulation of hydrological processes in agricultural watersheds of the Red River. Falsifications
38 of snow sublimation, snow transport, and infiltration to frozen soils processes in the validated
39 base model indicate that these processes were very influential to stream discharge generation.

40 **KEY WORDS:** Prairie hydrology; cold-region; snowmelt; physically-based modelling;
41 agricultural watershed; intensively drained.

42 **1. Introduction**

43 Lake Winnipeg, the 10th largest freshwater lake in the world, has experienced rapid
44 eutrophication in the last century due to increased loading of phosphorus (McCullough et al.,
45 2012). The Red River is the largest source of phosphorus (P) to the lake, responsible for around
46 70% of the total annual load (Bourne et al., 2002;Schindler et al., 2012). The majority of the



47 export from the river occurs during hydrological events where P concentration in its tributaries
48 increases with discharge (McCullough et al., 2012). Changes in the discharge of the Red River
49 and its tributaries have been suggested as one of the primary drivers of changes in phosphorus
50 loading to Lake Winnipeg since the 1970s (McCullough et al., 2012;Schindler et al., 2012).
51 Therefore, research to improve understanding of those hydrological processes that control
52 transport is of particular regional importance.

53 Many popular models utilized for the simulation of hydrological processes in agricultural
54 watersheds were initially developed for regions where rainfall driven runoff is the primary
55 contributor to annual water yield. Snowmelt is the main source of stream flow in tributaries of
56 the Red River and accurate representation of hydrological processes in its agricultural tributaries
57 using popular models like the Soil and Water Assessment Tool (SWAT) has proved challenging.
58 While SWAT has been used to estimate water quality targets and beneficial management
59 practices in Canada (Yang et al., 2012), significant modification of the original model is
60 generally required to achieve a good fit with calibration data (Watson et al., 2008;Liu et al.,
61 2012) and improvement of the accuracy of underlying representation of important hydrological
62 processes (e.g. landscape representation, stream routing) is ongoing (Douglas-Mankin et al.,
63 2010).

64 As a result of the problems associated with adapting hydrological models developed for
65 more temperate conditions, a number of models that incorporate cold-region hydrological
66 process have been developed, such as ARHYTHM, VIC, and the Cold Regions Hydrological
67 Modelling (CRHM) platform (Pomeroy et al., 2007). Of these models, CRHM offers the most
68 complete range of physically-based process representation for the Northern Great Plains,
69 including blowing snow, interception and sublimation of snow, energy balance snowmelt,



70 canopy influence on radiation, and infiltration to frozen soils (Pomeroy et al., 2007;Fang et al.,
71 2013). Although CRHM currently does not feature a module to represent nutrient dynamics in
72 either soil or water, such modules are under development (Roste, 2015) and the platform is a
73 powerful tool for assessing watershed nutrient dynamics because of the well-established
74 relationships between phosphorus concentrations and discharge rate in tributaries of the Red
75 River, including the La Salle River (McCullough et al., 2012), which is the focus of the research
76 presented here.

77 The CRHM platform has been successfully used to simulate hydrological processes in a
78 number of contrasting catchments in Canada (Dornes et al., 2008;Fang and Pomeroy, 2008;Ellis
79 et al., 2010),globally in China, Patagonia, and France (Rasouli et al., 2014), and in the
80 intensively studied South Tobacco Creek watershed that drains from the more steeply sloped
81 Pembina/Manitoba Escarpment feature into the Red River in Manitoba (Mahmood et al., 2016).
82 However, CRHM has not yet been used to simulate hydrological processes in the intensively
83 managed lowland agricultural tributaries of the Red River such as the La Salle River.
84 Characteristics in these watersheds include extensive artificial surface drainage, channelization
85 of the stream network, historical drainage of wetlands, clay soils, high fertilizer input, high crop
86 yield, high livestock densities, and highly connected drainage areas with little surface storage.
87 Simulation of cold-region hydrological processes in this landscape is particularly challenging
88 because of the combined effect of climate and land use management on water transport. Large
89 volumes of runoff can be produced when snowmelt is routed over frozen soils (Shook and
90 Pomeroy, 2010;Shook and Pomeroy, 2012) and this process is emphasized where surface
91 drainage enhancements and roadside ditches speed transport (Brunet and Westbrook,
92 2012;Pomeroy et al., 2014). As a result, flow in streams of the Red River watershed tend to



93 exhibit a flashy response to snowmelt with large volumes of runoff from roadside ditches
94 entering into relatively small river channels that are still covered by ice at time of melting. This
95 in-channel ice restricts flow and frequently causes flooding and backwater conditions to develop
96 (Gray and Prowse, 1993). These conditions can lead to error in discharge measurements
97 calculated using stage-rating curves (Mosley and McKerchar, 1993) also invalidate important
98 assumptions of flow routing models (Fread, 1993), thus making accurate prediction of flow more
99 challenging. For this reason, using a non-calibrated, but flexible physically-based model platform
100 is an attractive alternative to models primarily calibrated based on discharge measurements. The
101 application of CRHM will enhance the understanding of the hydrological controls in tributaries
102 of the Red River since concurrent measurements such as SWE and soil moisture can also be
103 utilized to gauge the accuracy of model process representations independent of uncertainties
104 associated with in-stream flow representation or measurement. At the same time, the limitations
105 of existing conceptual models and/or potential problems with discharge measurement should be
106 more apparent in the absence of calibration (Spence et al., 2013).

107 The objectives of this study were to assess the performance of CRHM to simulate the
108 hydrology of an intensively managed agricultural sub-watershed of the Red River Valley and to
109 gain insight about the dominant hydrological controls of stream flow in this unique landscape.
110 Model simulation was completed in a sub-catchment of the La Salle River watershed for the
111 1992-2013 time period. Given global concern regarding the eutrophication of Lake Winnipeg
112 and the importance of discharge from the Red River in determining rates of P loading to the lake,
113 emphasis is placed on evaluating the accuracy of the model in representing snowmelt and high
114 flow events. The research presented here also provides insight into whether there are particular
115 periods of time or weather conditions for which there is poor model performance indicating



116 either limitations of model input data or the need for improved understanding and representation
117 of watershed hydrology in the region.

118 **2. Material and Methods**

119 *2.1 Study site*

120 The study was conducted in a 189 km² sub-catchment of the La Salle River watershed (LS-
121 05OG008; Fig. 1a), a tributary of the larger Red River which is located in the central plains
122 region of Manitoba, Canada (Graveline and Larter, 2006). The study watershed is underlain by
123 lacustrine clay deposited in glacial Lake Agassiz; these deposits consist of a lower, dark grey
124 clay and a thinner upper unit of lighter coloured, calcareous silty clay, with surface texture being
125 predominantly clayey (La Salle Redboine Conservation District, 2007). The study area is
126 located in the Prairie Ecozone, with mean annual temperature around 2.5°C, mean summer
127 temperature of 16°C and mean winter temperature of -13°C; the mean annual precipitation is 560
128 mm, out of which around 25% takes place as snow, while the potential mean annual gross
129 evapotranspiration is about 834 mm (La Salle Redboine Conservation District, 2007).

130 *2.2 Hydrological and meteorological observations*

131 Daily streamflow observations between 1990 and 2013 were obtained from the
132 hydrometric data (HYDAT) database (Environment and Climate Change Canada, 2013) for
133 Water Survey of Canada (WSC) gauging station 05OG008 (La Salle River near Elie; Fig.1b)
134 located at the outlet of the study watershed. Data collection at this location was seasonal from
135 1990 to 1996, and has been continuous from 2002 to present. The annual monitoring period for
136 this station spans from March 1st to October 31st, with no data available during winter months;
137 thus, analyses were carried out using calendar years rather than water years (i.e. October 1st to
138 September 30th) since the hydrometric data does not span over more than a calendar year. A gap



139 in available flow data exists between flooding in 1997 and instrument replacement in 2001.
140 Only flow data is available from HYDAT for the period prior to 1996, while flow and water
141 level were both recorded from 2002 onwards. Notes in the HYDAT metadata pertained to 2004
142 and 2008 indicated equipment malfunctions resulting in loss of data. For this reason, the periods
143 from 1997-2001, 2004, and 2008 were not used for model assessment.

144 The gauging station 05OG008 is located 80 meters downstream from a small stop-log dam,
145 which raises concerns about ice jamming and backwater issues. Ice conditions are flagged in the
146 HYDAT daily records, but backwater and flooding conditions are not indicated for the site.
147 Thus, field notes from site visits were acquired from WSC and used to determine potential
148 backwater conditions. The threshold for flagging these conditions in the present analysis was
149 based on the minimum water level for which backwater was recorded by WSC technicians on the
150 field. This level was 239.3 m, observed on May 19, 2011, which is only 0.2 m below the full
151 supply level (FSL) of the stop-log dam (239.5 m). Potential backwater conditions were assumed
152 whenever the water level was above this threshold. Since water levels in the HYDAT database
153 were only available after 2002, this analysis was carried out between 2002 and 2013. These
154 potential backwater periods are noted in figures displaying measured discharge. However, the
155 presence of backwater conditions, although common, were not consistently documented on the
156 WSC field notes to provide verification of occurrence or impact on measurement accuracy; so
157 even where backwater conditions were suspected, all flow data was assumed accurate and
158 utilized in assessment of model performance.

159 The hourly weather data used to force CRHM was obtained from closest Environment
160 Canada weather station (Environment and Climate Change Canada, 2015) with available data
161 (Fig. 1a). Nearby stations are located at Portage Southport Airport, Winnipeg International



162 Airport, and Marquette (26.6, 47.9, and 9.9 km from the geometric centre of the study area,
163 respectively). Temperature, wind speed, and relative humidity were obtained from the Portage
164 Southport Airport, solar radiation was acquired from the station located at the Winnipeg
165 International Airport, and precipitation was acquired from the weather station in Marquette.
166 Precipitation was available in a daily time-step and was disaggregated to an hourly time-step
167 using HyetosR (Kossieris et al., 2013), which is a package for the temporal stochastic simulation
168 of rainfall process at fine time scales based on Bartlett-Lewis rectangular pulses rainfall model
169 (Koutsoyiannis and Onof, 2001).

170 *2.3 Watershed delineation and HRU definition*

171 The soils, topography, and land use datasets used as model inputs in this study were
172 derived according to the principles described by Liu et al. (2013) for prediction in ungauged
173 basins. The sub-basins within the La Salle watershed (Fig. 1a) were defined using the watershed
174 delineation tool in SWAT (Yang et al., 2014) and the 90-m digital elevation model (DEM)
175 derived from the NASA Shuttle Radar Topography Mission (SRTM) data. Soil datasets with
176 scales ranging from 1:20,000 to 1:126,720 were obtained from the Manitoba Land Initiative
177 (MLI) database. The soils textures in sub-catchment were 94% clay, 4% silt clay loam, 1% silt
178 clay, and 1% silt loam. The land use datasets were acquired from MLI for non-agricultural land
179 uses and from Agriculture and Agri-Food Canada (AAFC) for cropping systems. Cropping
180 systems, or typical cropping rotations, were derived from nine years (2001-2009) of records from
181 the Manitoba Agricultural Services Corporation (MASC) crop insurance database, as well as
182 land use defined by the Ag-Capture tool between 2009 and 2012 that included each parcel of
183 land in the study area (Yang et al., 2014). The Ag-Capture tool is a geospatial, agricultural land-
184 use inventory and mapping tool developed by Agriculture and Agri-Food Canada (OECD, 2013)



185 that validates remote-sensed land use classification using field surveys. The cropping system
186 assigned to each parcel was based on the dominant or most common crop grown while its
187 representative crop rotation was based on the proportion of the other most common crops grown
188 on the same land parcel over the entire period of record. Cropping systems used in the present
189 model and the crops therein are shown in Table 1.

190 Using SWAT's watershed delineation tool, the entire La Salle watershed was divided into
191 38 sub-basins (Fig. 1a) (Yang et al., 2014). The sub-catchment comprised of sub-basins 2, 3, 5
192 and 33 was used in the present study; they correspond to the gross drainage area draining into the
193 gauging station at the outlet of the overall watershed (05OG008; Fig.1b). The hydrological
194 response units (HRUs) used by CRHM in this study were also based on the HRUs previously
195 defined for the SWAT model (Yang et al., 2014), where HRUs are defined based on unique
196 combinations of land use, soil, and topography. The HRU-definition strategy for applying
197 cropping systems was based on the land-use split method, which divides the land-use type or
198 cropping system into multiple, more specific crop types based on what would be a typical crop
199 rotation for that system. The land-use split method allows for representing crop rotations in the
200 model in a static fashion by distributing the different crops within a cropping system throughout
201 the acreage of the cropping system in a single year (Fig. 2) (Yang et al., 2014). For instance, the
202 "Fall Cereal Cropping System" is comprised of canola, spring wheat, canola, and winter wheat in
203 a four-year rotation (Fig. 2a). Thus, in the four-year rotation, canola is planted 50% of the time
204 (two out of four years), while spring wheat and winter wheat are each planted 25% of the time
205 (one out of four years). The land-use split method distributed this rotation throughout the Fall
206 Cereal cropping system land base within a single year based on the relative percent of each crop
207 (Fig. 2b). Thus, every year, canola, spring wheat, and winter wheat are represented in the model



208 as being planted in 50%, 25%, and 25% of the Fall Cereal acreage, respectively. Once the
209 cropping systems were represented using the land-use split method, the crops were associated to
210 soil types present in the LS-05OG008 sub-catchment. The combination of land use and soil types
211 resulted in the HRUs presented in Table 1. As indicated in this table, agriculture dominates the
212 land use in the study area with urban areas and natural lands occupying only a very small
213 proportion. This is the first time that the land-use split approach has been used in CRHM for
214 simulation of hydrological processes in agricultural watersheds, allowing a more detailed
215 representation of the agricultural land use that included crop types.

216 Due to the highly altered nature of the drainage network in the watershed and its low relief,
217 data collection was undertaken to ensure accurate delineation of routing. Initially the stream
218 network was defined based on digital water feature data from the Manitoba Land Initiative (MLI)
219 (1:20,000 Designated Drains) with supplemental ephemeral streams digitized based on aerial
220 photograph interpretation. The drainage ditch dataset, mainly comprised of roadside ditches,
221 utilized a culvert inventory along with LiDAR DEM (vertical accuracy ± 0.15 m) analysis to
222 determine flow direction to create ditch line segments that passed through culverts and ultimately
223 connected to the stream network. The stream and drainage ditch datasets were merged together
224 to create the final drainage network for the model.

225 *2.4 CRHM module selection*

226 The Cold Region Hydrological Modelling platform (CRHM) was used to develop a custom
227 hydrological model for LS-05OG008. CRHM is a modular objected-oriented platform that
228 allows the creation of customizable models varying from conceptual to physically-based
229 representations of the study area, according to the data availability and uncertainty in process
230 parameters of the basin (Harder and Pomeroy, 2014). A detailed description of CRHM and its



231 modules, which are based on decades of cold-region hydrological research in western and
232 northern Canada, is available in Pomeroy et al. (2007) , with an update for agricultural
233 applications in the Canadian Prairies described by Fang et al. (2010). A set of physically-based
234 modules was defined and arranged sequentially to simulate the dominant hydrological processes
235 in LS-05OG008. Table 2 lists the modules selected, their function, and the sequence in which
236 they were entered into the customized model. Similar model group structures have been
237 successfully used to simulate hydrological processes in the prairie pothole region of the
238 Canadian Prairies (Fang and Pomeroy, 2008;Fang et al., 2010). The same model structure was
239 applied for the four sub-basins of LS-05OG008 (i.e., sub-basins 2, 3, 5 and 33). Each sub-basin
240 was represented in the model by a Group, which is a collection of modules executed in sequence
241 for all HRUs and connected with one another by a Muskingum routing group.

242 *2.5 CRHM module parameterization*

243 The CRHM software does not make provision for calibration utilizing fitted variables;
244 rather, parameters are selected based on the understanding of the hydrological system (Pomeroy
245 et al., 2007) . Since the land-use split approach was used, the HRU distribution was held
246 constant over the simulation period, which allowed for a single set of parameters to be used in
247 the model for each HRU. Sub-basin and HRU physiographic parameters (i.e. area, elevation,
248 slope, aspect, and latitude) were obtained from the DEM and HRU reports generated by SWAT
249 during the watershed delineation and HRU definition for a previous SWAT simulation in the La
250 Salle (Yang et al., 2014) . The parameters of the soil and recharge layers (e.g. soil texture, soil
251 porosity, saturated hydraulic conductivity, and water holding capacity) were obtained or derived
252 from the Soil Survey Reports published by the Province of Manitoba (Michalyna et al.,



253 1972;Ehrlich et al., 1953). The pore size distribution index (λ) was defined based on soil textures
254 associated with clay (Brooks and Corey, 1966;Corey, 1994).

255 Crop seeding dates were defined based on a 10-year average (2000-2009) for each crop
256 based on crop insurance data obtained from MASC. Harvest dates were based on the length of
257 the growing season defined by consultation with farmers in the study area. Leaf area index
258 (LAI), vegetation height, and stalk diameter for the different crops were obtained from the
259 dataset collected by AAFC for the SMAPVEX12 Campaign in 2012 (Kim et al., 2014). The
260 SMAPVEX12 field sites were located between 13 and 45 km south of LS-05OG008 and are
261 located within the Red River watershed and the same or similar ecoregions as the La Salle (Lake
262 Manitoba Plain or Interlake Plain). The Penman–Monteith method was used to calculate
263 evapotranspiration from crop HRUs, while the Priestley-Taylor method was used to calculate
264 evaporation from open-water HRUs (e.g. wetland, river channel).

265 Although located in the Canadian Prairies, the runoff routing in the LS-05OG008 sub-
266 catchment does not follow the typical sequence of land uses further west in this region described
267 by Fang et al. (2010) (i.e. fallow, stubble, and pasture routed to woodland and then to wetland,
268 open water, and river channel). Rather, the flat and intensively managed characteristic of the La
269 Salle watershed result in a lack of any typical routing sequence based on land use. For this
270 reason, uplands were routed directly to drainage channels in LS-05OG008. For each HRU, the
271 routing length was calculated as the median of the distances from each HRU to the drainage
272 network, as obtained using ‘near’ tool in ArcGIS 10.1 (ESRI, Redlands, California). The
273 distances to the drainage network were estimated from the 2009 land use coverage using the Ag-
274 Capture tool.



275 The maximum depression storage of the HRUs was calculated using the methodology
276 described by (Fang et al., 2010) for upland areas, where volume and area were calculated from
277 the same LiDAR DEM used to derive the final drainage network. Briefly, the ‘fill’ tool was used
278 to level off the original DEM and the ‘cut/fill’ tool to calculate the area and volume from the
279 difference between the original and the leveled DEMs. Muskingum routing within and between
280 sub-basins of LS-05OG008 was setup according to (Fang et al., 2013). The storage parameter
281 used in the Muskingum routing module was set to zero, based on the typical pattern observed for
282 regional agricultural runoff monitoring where individual diurnal runoff events at edge-of-field
283 begin as soon as melt starts around noon and stops due to refreezing at night (Tiessen et al.,
284 2010a). In-channel storage was calculated as total reach length (calculated in GIS using the
285 drainage network) divided by average flow velocity, which was estimated using measured
286 hydrographs and channel dimensions (i.e. average flow rate in the simulation period divided by
287 cross sectional area). The dimensionless constant that weights inflow and outflow was set to
288 0.25, which is a common value for natural stream channels (Carter and Godfrey, 1960) .

289 *2.6 Assessment of model simulations*

290 Model simulations were graphically and statistically assessed against streamflow data
291 collected by WSC at station number 05OG008. The statistical metrics used for model assessment
292 in a daily time-step were the Nash-Sutcliffe model efficiency (NSE), model bias (MB), the root
293 mean square difference (RMSD), and the normalized root mean square difference (NMSD)
294 (Fang et al., 2013). Although streamflow simulations were available between 1990 and 2013,
295 some years could not be used for model assessment. Among those are the years of 1990 and
296 1991, which were used as the model warm-up period, and the years having data quality issues as



297 documented in the WSC streamflow database HYDAT (version 10 issued on October 17, 2014).

298 Thus, the years of 1990-1991, 1997-2001, 2004, and 2008 were not used for model assessment.

299 Key aspects of hydrological cycle influencing streamflow generation in cold regions were
300 also used to assess model performance. Simulated snow water equivalent (SWE) was compared
301 to SWE calculated from the depth of snow on ground measured at the Marquette station using
302 the following relationship (Pomeroy and Gray, 1995):

$$303 \quad SWE = 0.01d_s\rho_s, \quad (1)$$

304 where d_s is the depth of snow (cm), ρ_s is the snow density, which was assumed to be 180 kg m^{-3}
305 based on the typical density range between new and settled snow (Paterson, 1994), and SWE is
306 same as defined above (mm). Simulated yearly cumulative evaporation from open water was
307 compared to gross evaporation values published for Portage La Prairie between 1992 and 2000
308 (Martin, 2002), which is the closest location with available data corresponding to the time period
309 of model simulations. Since both values were based on evaporation models and not measured,
310 the evaporation comparison was made only to verify that CRHM simulations were in agreement
311 with values expected for the study area. Simulated volumetric soil water content [VWC,
312 expressed in mm of water in the soil profile, i.e. $\text{VMC} (\text{m}^3 \text{ m}^{-3}) \times \text{profile depth (mm)}$] was
313 compared to VWC simulated across Canada using the National Drought Model (NDM)
314 (Chipanshi et al., 2013). Similarly to evaporation, no direct measurements of soil moisture were
315 available in close proximity to the watershed; thus, the NDM dataset was used to verify only if
316 the VWC simulated by CRHM was within a reasonable range and followed the expected
317 seasonal trend.



318 2.7 Model falsification

319 Taking advantage of the flexible structure of the CRHM platform to simulate physically-
320 based hydrological processes, model falsifications were performed to assess the impact of snow
321 sublimation, blowing snow (which includes both snow sublimation and transport), and frozen-
322 soil infiltration on stream discharge in the study area. These processes are typical of the
323 Canadian Prairies and very influential to runoff generation in the region (Pomeroy et al., 2007).
324 A stepwise model falsification was achieved by sequentially removing the following processes
325 from the model: i) snow sublimation, ii) blowing snow, iii) frozen-soil infiltration, and iv)
326 blowing snow and frozen-soil infiltration, which is the combination of cases (ii) and (iii).
327 Different than the model assessment analysis, which used calendar years, the impact of model
328 falsification was assessed for water years, since a complete time series was available from model
329 simulations. However, statistical metrics used for model assessment (i.e., NSE and MB) were
330 only calculated between March and October, which is the period where observed streamflow is
331 available.

332 3. Results

333 3.1 Flow characteristics in the study area between 1990 and 2013

334 As is typical of the cold-region conditions prevailing in the Canadian Prairies, peak stream
335 discharges take place in the spring (Table 3). Thirteen of 15 years exhibited annual peak
336 discharge with snowmelt. The median peak discharge was $6.7 \text{ m}^3 \text{ s}^{-1}$, while the median annual
337 discharge volume was $1.25 \times 10^7 \text{ m}^3$ and the median water yield was 66 mm. Years with peak
338 discharge above the median corresponded to years with annual discharge volume above the
339 median, due to the strong positive correlation between these two variables ($r^2=0.90$), reflecting
340 that most of the annual discharge occurs during spring and is associated with event runoff rather



341 than baseflow. Those years with peak discharge and annual discharge volume equal to or above
342 median values were considered wet years for model assessment purposes and include the years
343 of 1996, 2005, 2006, 2009-2011, and 2013.

344 Seven out of 15 years had ice conditions at time of peak discharge, while two years had
345 peak discharges one day after the end of the ice period (Table 3). Ice conditions at peak
346 discharge were most common in years with cumulative and peak discharges equal to or below
347 the median. Field observations at the site indicate that the absence of ice in wet years was likely
348 due to faster-moving water at high discharge. However, ice moving downstream still may have
349 impacted the flow regime at the gauging station through the formation of ice jams along the river
350 channel and promotion of backwater conditions, which were relatively common (Table 3).

351 Seven out of nine years with both upstream and downstream level monitoring exhibited
352 potential backwater effects (Table 3; Fig. 3). Potential backwater events were generally more
353 associated with years of high discharges. Backwater in this reach could be caused by two factors.
354 The first being that the confluence of the Elm River channel and the La Salle River just
355 downstream of the gauging station 05OG008 (Fig. 1b). Given the gentle topographic gradient of
356 the area and the presence of the small dam, the water backs up in the reach of the La Salle River
357 upstream of the confluence of the two channels. The second factor is the occurrence of ice-
358 damming and potential for build-up from the 4-m dam at Starbuck located 28.6 km downstream
359 following an elevation change of 3 m.

360 *3.2 CRHM simulations of stream discharge*

361 The majority of annual stream discharge occurs in spring and in particular with snowmelt,
362 making this the most important period for hydrological simulations to identify nutrient export



363 potential (Table 3). However, stream discharge measurements during the spring are also
364 associated with the greatest potential for model and measurement uncertainties related to ice
365 conditions and backwater issues. Given the hydrological importance of stream discharges driven
366 by snowmelt, model assessment was performed without removing questionable spring records
367 from the dataset. The only period removed from analysis was the backwater period in July of
368 2005 that occurred following an extreme rainfall runoff event (Table 3; Fig. 3) with a high
369 degree of known uncertainty regarding measurement accuracy caused by backwater and debris
370 buildup (according to WSC field notes), and representing disinformation for model assessment
371 (Beven, 2011).

372 Graphical comparison between observed and simulated stream discharge indicates that the
373 model skill varied for different years (Fig. 4). Records of discharge measurements performed by
374 WSC during site visits were obtained and are also shown in Fig. 4. Simulated time of peak
375 discharge was generally in good agreement with observed values, except for three out of 15 years
376 (i.e. 1993, 2003, and 2007). These three years had peak discharge and annual discharge volume
377 equal to or below the medians for the 1990-2013 period (i.e. $6.7 \text{ m}^3 \text{ s}^{-1}$ for peak discharge and
378 $1.21 \times 10^7 \text{ m}^3$ for annual discharge volume). The magnitude of peak discharges was also
379 reasonably simulated by the model (Fig. 5), although with more variability than timing of peak.
380 Much of this variation was associated with either years of discharge volume below the median
381 (i.e. 1994, 1995, 2002, and 2007) or large peak discharge (i.e. 2006, 2009, and 2011). There was
382 no direct relationship between overestimation of simulated peak discharges (Fig. 5) and
383 overestimation of annual discharge volumes (Fig. 6). However, there was a tendency for
384 simulated annual discharge volumes to be overestimated in low-discharge years (Fig. 6) and to
385 underestimate large peak discharges (Fig. 5).



386 Statistical metrics used to assess model performance are shown in Table 4. The results
387 confirm the trends shown in graphical analysis where the model had very good performance for
388 years with peak discharge equal or larger than the median peak discharge. The Nash-Sutcliffe
389 efficiency (NSE) was consistently above 0.65 and averaged 0.76 for 1992, 1996, 2005, 2006,
390 2009, 2011, and 2013. The year 2010 was an exception, with NSE=0.36. For low discharge
391 years, NSE was negative, indicating that the model was no better than utilizing mean discharge
392 as a predictor of stream discharge. For the years with NSE > 0.65, the model bias (MB) was
393 generally negative, although the average model bias was small (i.e. 5%). This was likely due to
394 the overall underestimation of peak discharges during those years. Conversely, MB was
395 generally positive for years with negative NSE, which correspond to years with annual discharge
396 volumes below the 1990-2013 median. This confirms the difficulty in simulating low discharges
397 as was observed in the graphical analysis. The root mean square difference (RMSD) and the
398 normalized root mean square difference (NRMSD) were small to moderate, given the range of
399 discharges comprised between minimum and peak discharges.

400 3.3 CRHM simulation of SWE, evaporation, and VWC

401 The annual trend in SWE and time of melt was accurately captured by the model (Fig.7).
402 Differences in peak SWE (e.g. 1992 and 1996) were likely due to inter-annual variation snow
403 density, which was not measured and can vary substantially with space and time (Pomeroy and
404 Gray, 1995). However, the assumed snow density value did not introduce a substantial bias into
405 SWE estimation, as most years had good agreement between simulated and observed peak SWE
406 values. CRHM was also able to simulate cumulative open water evaporation that compared well
407 with gross evaporation estimates for the years when data was available (Fig. 8). The average
408 cumulative gross evaporation estimated by AAFC at Portage La Prairie was 686 mm, while that



409 simulated by CRHM for open water was 682 mm. The soil moisture range and variation trend
410 over the year was also well simulated by CRHM in most years (Fig. 9). Despite some
411 differences, CRHM was generally able to match the soil moisture content at the beginning of
412 each growing season and to capture its depletion and recovery over the course of the summer and
413 fall seasons, respectively. The results above are important since they imply that antecedent
414 conditions were reasonably well predicted with accurate representation of watershed
415 hydrological processes and that soil infiltration, soil freezing, and snowmelt runoff patterns can
416 be modelled with greater confidence.

417 *3.4 Sensitivity analysis using the storage parameter in the Muskingum model*

418 A sensitivity analysis was carried out using the storage parameter in the Muskingum model
419 (K storage) for both overland and drainage network flow routing to investigate the adequacy of
420 average values for both wet and dry years (Fig. 10). For wet years, the plot indicates equifinality
421 (Beven, 2011), where different models have similar performance when K storage is in the range
422 between zero and four days for both upland and drainage network (Fig. 10a). The model bias for
423 peak flows aids model selection by defining a narrow band for which the bias is minimized (Fig.
424 10c). However, knowledge of the flow characteristics is still necessary to define the K storage
425 parameter. In this case, K storage of 0 and 3 days for upland and drainage network, respectively,
426 seem reasonable to maximize NSE and minimize peak-flow MB. Conversely, the model
427 performance for dry years tends to improve as K storage increases for both upland and drainage
428 network, although NSE is still in the negative range (Fig. 10b). The contour plot of peak-flow
429 model bias for dry years also suggest large K storage values since low bias is only found for
430 higher K storage values (Fig. 10d).



431 Despite the relative improvement, changing K storage in dry years did not impact model
432 performance appreciably. These results may suggest that the hydrologic controls under dry
433 conditions in this watershed are not strongly topological (i.e. drainage network) but typological
434 (i.e. landscape elements) (Buttle, 2006). In order to assess this possibility, the CRHM option to
435 allow infiltration prior to the major melt event was selected in an attempt to emulate the effect of
436 preferential flow. The NSE and MB for dry years were, respectively, -27.87 and 2.0 (1994), -
437 0.58 and 0.81 (1995), -9.6 and 1.7 (2002), -1.09 and 0.16 (2003), -0.43 and 0.69 (2007), and -
438 0.17 and 0.25 (2012). These results indicate an improvement in model performance in two out of
439 six years (i.e. 2002 and 2012) when compared to the based model (Table 4), which represents
440 better predictions >30% of the time.

441 3.5 Model falsification

442 Comparison of the hydrographs between the base model and the falsified models indicates
443 that sublimation, blowing snow, and infiltration to frozen soils are influential processes in the
444 study area (Fig. 11). Turning off snow sublimation and blowing snow (i.e. snow sublimation and
445 transport combined) in the models resulted in increased peak discharges (Fig. 11b-c), while
446 removing frozen-soil infiltration reduced peak discharge with or without blowing snow (Fig.
447 11d-e). The average change in peak discharge due to sublimation inhibition increased 25.4%,
448 while the change due to blowing snow inhibition was 39.1%; the effect of inhibiting frozen-soil
449 infiltration was the opposite, with reductions around 60% in peak discharge regardless of the
450 blowing snow process (Table 5). Despite the average increase in peak discharge, there were also
451 reductions in peak discharge in particular years when the snow sublimation and blowing snow
452 were removed (i.e. 1993-1994; 1994-1995; 2003-2004; 2006-2007). For frozen-soil infiltration



453 models, removing this process had a very consistent effect on peak discharge with reductions in
454 all but one year (i.e. 2001-2002).

455 One striking effect of the model falsification was that disruption the snow sublimation,
456 blowing snow, and frozen-soil infiltration caused a disconnection between peak discharge and
457 annual discharge, which was a feature of the observed stream discharge (section 3.1). In some
458 years, reductions in peak discharge were accompanied by increases in total discharge (e.g. 2003-
459 2004 for snow sublimation and blowing snow falsifications; 2002-2003 for frozen-soil
460 infiltration models). In other years, the opposite was true (e.g. 2009-2010 for snow sublimation
461 and blowing snow falsifications; no instances for frozen-soil infiltration models; Table 5).

462 Models with snow processes disrupted also presented consistent increases in SWE over the
463 entire simulation period (Table 5). Noteworthy, there was no difference in SWE percent change
464 among the models with snow sublimation and blowing snow falsifications, which indicates no
465 influence of snow transport in SWE. No transport of snow out of the watershed was confirmed
466 by assessing the snow loss variable within CRHM, which indicated total losses smaller than 1
467 mm over the 21 years of simulation (i.e. including the spin-up years) in all cases.

468 The statistical metrics calculated for the falsified models (Table 6) confirmed the loss of
469 model performance when compared to the base model (Table 4), which highlights the
470 importance of snow sublimation, blowing snow, and infiltration to frozen soils processes for
471 accurately simulating stream discharge in the study area. Models with falsified snow processes
472 (i.e. snow sublimation and blowing snow) generally presented a loss of performance, although in
473 specific years, performance remained comparable (e.g. 2010) or improved slightly (e.g. 2005,
474 2006). This improvement was likely due to the increased peak discharge that offset the trend of



475 the model to underestimate large peak discharges in wet years (Fig. 5). Similarly to NSE, model
476 bias tended to become more extreme due to falsification of snow processes.

477 Metrics for falsification of frozen-soil infiltration (including blowing snow falsification or
478 not) were also degraded in relation to the base model. Loss of performance was more severe in
479 wet years due to drastic reduction in peak discharge (Fig. 11). However, performance tended to
480 improve in dry years (e.g. 2005), which was likely due to an offset in the trend of the base model
481 to overestimate discharges in those years. Reduced peak and total discharges due to falsification
482 of frozen-soil infiltration process was due to increased infiltration.

483 **4. Discussion**

484 The assessment of the hydrological records of the La Salle River at the gauging station
485 05OG008 suggests that ice and backwater pose additional challenges to simulating streamflow in
486 the study area. Peak flows in the river usually took place under ice impacted conditions or
487 directly after clearing of ice (Table 3), which may have impacted both the accuracy of discharge
488 measurements at the site (Environment and Climate Change Canada 2015) and hydrograph
489 pattern. Comments from WSC field notes identified a number of potential sources of uncertainty
490 for discharge measurements. For example, backwater conditions are reported in 2005, 2010 and
491 2011, which correspond to the period of backwater identified in this study. Overbank flow into
492 the riparian vegetation is reported for 2007, which also corresponds to the backwater period
493 identified in the present analysis.

494 Backwater during ice conditions may take place because the bottom of the ice cover causes
495 an increase in flow resistance (Gray and Prowse, 1993) and may affect flow magnitudes and
496 timing. Backing up of water from downstream ice jams after breakup may also impact



497 hydrograph shape promoting high peak flow after release and may increase uncertainty
498 associated with discharge measurements (ASCE, 1996). Increasing flows at the confluence of the
499 La Salle River and the Elm River channel, located downstream of 05OG008 may also create
500 backwater conditions where channel capacity is exceeded. According to the WSC field notes,
501 this seems to be the reason for the backwater in 2011, which was caused by the discharge from a
502 controlled breach in dikes of the Assiniboine River being routed through the Elm River Channel.
503 Overall, backwater conditions seem to be quite frequent in the sub-catchment LS-05OG008
504 according to the present analysis and based on field observations. Due to changes in flow
505 conditions, discharge measurements are recommended on a regular basis after high flow events
506 or ice break-up; if a shift in the rating curve is necessary to accommodate the new flow
507 conditions, a complete new set of discharge measurements must be made (Coulson, 1991). Thus,
508 the peak flows in the observed records could have been exacerbated due to ice conditions and
509 backwater.

510 Model performance may also be reduced during backwater periods. Flow routing was
511 calculated in CRHM using the Muskingum model, which is not accurate for rapidly raising
512 hydrographs routed through flat sloping rivers and neglects backwater effects due to tributary
513 inflows and presence of dams (Fread, 1993). These features are characteristic of the study area
514 and may have influenced the model simulations. Alternative distributed flow models such as the
515 Kinematic wave model and the Muskingum-Cunge model are also impacted by backwater effect,
516 which limits the selection of alternative methods for flood routing. Dynamic routing models that
517 solve the Saint-Venant equations are not limited by backwater conditions; however, no analytical
518 solution of the complete set of equations is available for most practical applications (Fread,
519 1993). Thus, numerical solutions based on finite-element or finite-difference schemes are usually



520 required. Computational inefficiency, numerical instability, and convergence are among the
521 drawbacks of such solutions, which explain the adoption of simpler routing methods in many
522 hydrological models. This remains a major challenge for application of most hydrological
523 models to basins such as the La Salle River basin.

524 Ice covered and backwater conditions create added uncertainty in evaluation of model
525 performance, making evaluation for LS-05OG008 more challenging, but this problem is likely to
526 be characteristics of other rivers in the Red River Valley given the relatively flat topography, the
527 presence of dams along the river channels, presence of ice during peak flows, and frequent
528 backwater at the confluence of different streams. It has been suggested that input errors in
529 precipitation datasets are the dominant source of error in hydrological modelling, while errors in
530 streamflow data are much smaller (Kavetski et al., 2006). While this premise is generally
531 accepted, the results shown in the present study suggest that uncertainty regarding streamflow
532 records of agricultural cold regions can be particularly complex.

533 Despite uncertainty associated with representation of in-stream dynamics during ice and
534 backwater conditions, CRHM was able to capture the overall trend in streamflow with
535 satisfactory simulation of the timing and magnitude of wet years (when peak flows were above
536 the median). However, even in those years, the peak flows were underestimated by the model.
537 Such contrast could be due to different routing conditions taking place during dry and wet years.
538 The contour plots of NSE and peak-flow MB for both wet and dry years suggest that average K
539 storage values are not adequate to represent both hydrological regimes and that dynamical
540 storage might be present in the basin (Fig. 10). Constant values of the Muskingum parameters
541 make them dependent of the hydrological flow data used to derive them, while it has been
542 suggested that a more physically realistic approach is to allow the parameters to vary in time and



543 space according to flow variability (Guang-Te and Singh, 1992). Due to this limitation,
544 methodologies have been devised for application of the Muskingum model with variable
545 parameters (Guang-Te and Singh, 1992; Song et al., 2011). In the present study, the K storage
546 parameter was adequate to represent wet conditions only, since good model performance was
547 achieved by the model for those conditions. However, K storage alone was not enough to explain
548 the poor performance during dry years since the NSE was consistently negative even for large K
549 storage values. Trends towards large K storage (i.e. slow movement of overland runoff and
550 channel routing) and overestimation of discharge for dry years (Figs. 4 and 6) suggest that other
551 physical process such as dynamical macropores affecting infiltration through soil cracks could be
552 underestimated during simulations that were designed to address wet condition flowpaths. In-
553 channel process such as water pooling and natural damming may also explain the large K storage
554 values for drier years.

555 Poor model performance in dry years is not completely unexpected since low flows are
556 generally more difficult to predict than larger flows (Nicolle et al., 2014), particularly for small
557 catchment areas (Stanfield et al., 2009). This difficulty may also have been aggravated if there
558 was flow under or over ice. The general pattern of flow is represented during drier years and
559 overestimation appears not to be caused by overestimation of baseflow; rather, problems are
560 most evident as overestimation of simulated peaks (e.g. years 1994 and 2002 in Fig. 4). While
561 drought conditions and in-stream flow requirements are important management considerations
562 on the Prairies (Fang and Pomeroy, 2007), the smaller magnitude events account for very little of
563 the overall export of water and phosphorus from the La Salle (Corriveau et al., 2013) and CRHM
564 appears to be a promising tool for physically-based simulation of the impact of management



565 change on watershed hydrology in intensively agricultural tributaries of the Red River where
566 elevated nutrient export is of particular concern.

567 Emulation of preferential flow by allowing infiltration prior to the major melt event
568 resulted in improved simulations more than 30% of the time in dry years, which suggests that
569 this mechanism is an important hydrologic control under those conditions. While the method
570 chosen does not capture all the complex nuances of preferential flow, it does allow for enhanced
571 infiltration and mimics the same effects of this mechanism. Interestingly, the two years that
572 resulted in improved simulations (i.e. 2002 and 2012), were dry years preceded by years with
573 wet springs and dry summers. In 2001, well-above-normal precipitation in the spring was
574 followed by dry summer and fall (Wheaton et al., 2008). The same was observed in 2011
575 (Cordeiro et al., 2014). This similar trend highlights the importance of antecedent conditions on
576 preferential flow. In contrast, there was no improvement in dry years followed by dry years (i.e.
577 1995 and 2003). Maybe the simulation of infiltration prior to melt was not enough to capture the
578 magnitude of preferential flow upon prolonged desiccation of the cracking soils in the study area.
579 Regardless, the physically-based nature of the simulations suggests that preferential flow
580 deserves investigation as one of the major hydrological controls driving stream discharge in dry
581 years.

582 This application of CHRM has a very high cultivated proportion of watershed land use,
583 (87%), extremely level topography, and modest depressional storage, which is unique and
584 contrasts to other agricultural applications of CRHM in land use proportions and topographic
585 relief (Mahmood et al., 2016; Fang et al., 2010; Pomeroy et al., 2007; Pomeroy et al., 2014). The
586 crops and cropping systems have been represented with an extremely high level of detail. The
587 land-use split method used in other modelling efforts in this watershed (Yang et al., 2014) to



588 represent crop rotations in a static fashion in a multi-year model exercise, seemed to work well
589 for the application in this study given the good simulations in wet years. While the input datasets
590 had a high level of detail, sensitivity analysis for different datasets (e.g. soils, land uses,
591 topography) would help identify critical inputs where detailed information is not necessary. Such
592 analysis would direct research and monitoring efforts to collect the most relevant information for
593 model setup and parameterization (Ahuja et al., 2002) in other regions of the Canadian Prairies
594 having physiography similar to that of this study.

595 Simulations of SWE, evaporation, and VWC were validated by external datasets,
596 indicating that the model state variables were in good agreement with the major hydrological
597 processes in the agricultural landscape. Small yearly differences in evaporation were likely due
598 to differences in input data, location, and method, where AAFC applies Meyer's revised formula
599 (Martin, 2002) for lakes and CRHM uses the Priestley-Taylor procedure for wetlands and small
600 lakes. Differences in soil moisture values could be due to differences in soil properties used to
601 calculate water holding capacity of the soil. Good simulation of model state variables increase
602 the confidence in the runoff simulations and highlight the value of the physically-based approach
603 used by CRHM to represent these landscapes. These measurements corroborate that the physical
604 description of these hydrological processes are reasonably well understood and represented in the
605 context of the Canadian Prairies, indirectly suggesting that the difficulties found by CRHM in
606 simulating stream flow pertain to runoff routing in the drainage network. The gentle topography,
607 ice-jammed channels at time of snowmelt, and river confluences make it a challenge to simulate
608 channel routing. Thus, research efforts should focus on better understanding river dynamics in
609 these landscapes.



610 Given the good performance in wet years, the present model can also potentially be used
611 for impact assessment of land-use and climate change on nutrient transport dynamics in the study
612 area. Trustworthy representation of the hydrology in the sub-catchment could be used to estimate
613 nutrient transport through fitting of different concentration-discharge (C-Q) models available in
614 the literature (Hall, 1970, 1971; Wilson et al., 2013; Helsel and Hirsch, 2002). Water-quality
615 datasets used for calibration and validation of previous model exercises (Yang et al., 2014) are
616 available to this end in adequate spatial-temporal frequency. This analysis would also allow for a
617 comparison between models in terms of water yield and nutrient export simulations. Such
618 comparisons will be the object of future research in the study area to identify how differing
619 modelling approaches (i.e. calibrated versus non-calibrated simulations) complement each other
620 in simulating flow and nutrient dynamics for future scenarios assessment.

621 The analysis of model falsification indicates that snow sublimation and blowing snow, as
622 well as infiltration to frozen soils, are crucial for accurate simulations of stream discharge in the
623 flat, intensively-managed agricultural watershed of the La Salle River. The prominence of snow
624 sublimation effects over snow transport contrasts with results observed for mountain
625 environments where greater importance of snow transport has been identified (Zhou et al., 2014).
626 However, snow transport within our watershed was still of importance, despite no loss of snow
627 from the watershed. Inclusion of snow transport affected peak and total discharge, with a more
628 pronounced effect on the latter (Table 5). Inhibition of snow sublimation caused a reduction in
629 total discharge in two out of 21 years (9% of the time) when compared to the base model, while
630 inhibition of both snow sublimation and transport caused a reduction of this variable in eight out
631 of 21 years (38% of the time). Accumulation of snow within watersheds tends to occur in
632 association with steep hills and valley slopes (Pomeroy and Goodison, 1997). In the lower slope



633 landscape of the La Salle watershed, snow accumulates in deep drainage ditches and stream
634 channels. If snow transport is inhibited, the simulated accumulation of snow in these topographic
635 features decreases and greater snow melt is simulated for upstream areas. This increases potential
636 for infiltration, which could explain the reduction in total discharge associated with inhibition of
637 snow transport. This pattern stresses the importance of internal snow transport within the
638 watershed to stream discharge generation despite snow transport not impacting peak discharge to
639 the same extent as snow sublimation. The pattern also provides some insight into the potential to
640 site wind barriers such as shelterbelts to retain snow in upstream areas (Steyn et al., 1997) to
641 encourage infiltration and reduce runoff generation with melt.

642 Regarding infiltration into frozen soils, falsifying this process had an overriding effect over
643 falsification of blowing snow since peak discharge was consistently reduced in these models,
644 despite the trend of increased peak and total discharge arising from falsification of snow
645 processes (Table 5). These results emphasize the importance of infiltration to stream discharge
646 generation. This process has been indirectly discussed by van der Kamp et al. (2003), who
647 observed reduced runoff for land uses with enhanced infiltration. Since runoff is the primary
648 source of stream discharge in the Canadian Prairies (Shook and Pomeroy, 2012; Shook and
649 Pomeroy, 2010), it was expected that its relative importance would be more pronounced than that
650 of blowing snow processes. The model falsification indicates that peak discharges would be
651 reduced, on average, by 61% due to inhibition of frozen-soil infiltration, while inhibition of snow
652 processes would increase peak discharges by 39% on average.

653 **5. Summary and conclusions**

654 Simulation of streamflow in an intensively managed agricultural tributary of the Red River
655 (the La Salle River) with the CRHM platform revealed a number of topics for which knowledge



656 of hydrological processes in the region and model representation might be improved. 1) Ice and
657 backwater issues are likely to contribute to increased uncertainty in both input hydrometric data
658 and model representations for tributaries of the Red River. Among the drivers of these issues are
659 the low relief of topography of the region, the presence of dams along the river channel, presence
660 of ice during peak flows, and frequent confluences of streams and artificial channels. 2)
661 Simulation of low flow years remains challenging in the La Salle, as is commonly reported for
662 other hydrological models. In low flow years, discharge was overestimated by 90% and a
663 sensitivity analysis of the storage parameter of the Muskingum routing model indicated that
664 averaging this parameter is not adequate for the study area. Also, improved simulations in dry
665 years through emulation of preferential flow by allowing infiltration prior to the major melt event
666 suggests that even where frozen soil predominates, preferential flow may be an important
667 hydrological feature under dry conditions for the high clay content soils of the Red River Valley.
668 As such, dynamical representation of processes such as infiltration through macropores may
669 require revision for drier conditions in the Red River Valley.

670 Despite the potential to improve model representation identified through this research, it is
671 also evident that performance metrics for the CRHM platform indicate very good simulation of
672 peak and yearly cumulative flows in the La Salle River where flows were equal to or above
673 median values (under normal to wet conditions). On average, discharge was underestimated by
674 only 5% in wet years. The good performance of the model in average or above-average flow
675 years indicate that CRHM simulations are likely appropriate for use in regional nutrient-transport
676 assessments where export is largely determined by hydrological drivers in the study area and for
677 assessment of land-use and climate change impacts on streamflow.



678 Falsifications of snow sublimation, snow transport, and infiltration to frozen soils
679 processes in the validated base model indicate that these processes were very influential to
680 stream discharge generation. Inhibition of snow sublimation would represented an average
681 increase in peak discharge around 25%, while inhibition of blowing snow, which includes both
682 snow sublimation and transport, would cause an increase in peak discharge around 39%.
683 Simulation of infiltration without changes to model structure to account for frozen-soils would
684 cause a reduction in peak discharge around 61%.

685 **7. Team list**

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687 Agri-Food Canada); Patsy Michiels (Land resource analyst; Science and Technology Branch,
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690 Technology Branch, Agriculture and Agri-Food Canada).

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696 **9. Code availability**

697 CRHM codes are available through model developers. Details can be found at
698 <http://www.usask.ca/hydrology/CRHM.php>.



699 **10. Data availability**

700 The weather and hydrometric datasets used in this research are publicly accessible through
701 the Environment and Climate Change Canada.

702 Weather Data: Environment and Climate Change Canada. 2015. Historical Climate Data.
703 Available at: <http://climate.weather.gc.ca/>. Access: February 22, 2016.

704 Hydrometric data: Environment and Climate Change Canada. 2015. HYDAT Database-
705 National Water Data Archive. Available at: [https://ec.gc.ca/rhc-
706 wsc/default.asp?lang=En&n=9018B5EC-1](https://ec.gc.ca/rhc-wsc/default.asp?lang=En&n=9018B5EC-1). Access: February 22, 2016.

707 **11. Appendices**

708 The manuscript has no appendices

709 **12. Supplement link**

710 **Plot data to be uploaded and link to be generated.**

711 **13. Author contribution**

712 M.R.C. Cordeiro, H.F. Wilson, and J. Vanrobaeys conceived the modelling objectives,
713 scope, and strategy; M.R.C. Cordeiro and J. Vanrobaeys acquired the input data; M.R.C.
714 Cordeiro, J.P. Pomeroy, and X. Fang developed the custom model for analysis; M.R.C. Cordeiro,
715 H.F. Wilson performed data analysis; M.R.C. prepared manuscript with contribution from all co-
716 authors.

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943



944 Table 1. List of hydrologic response units (HRUs) used in the LS-05OG008 watershed.

HRU ID	HRU acronym [†]	Cropping system/Land use	Crop	Soil texture	Occurrence			
					SB 2	SB 3	SB 5	SB 33
1	IVPO/SICL	Irrigated Vegetable	Potato	Silty clay loam	Yes	Yes	No	Yes
2	IVPO/C	Irrigated Vegetable	Potato	Clay	No	Yes	No	No
3	IVCR/SICL	Irrigated Vegetable	Carrot	Silty clay loam	Yes	Yes	No	Yes
4	IVCR/C	Irrigated Vegetable	Carrot	Clay	No	Yes	No	No
5	PRSB/SIC	Pulse Non-Row	Soybean	Silty clay	No	Yes	No	No
6	PRSB/C	Pulse Non-Row	Soybean	Clay	No	Yes	No	No
7	PFAF/C	Perennial Forage	Alfalfa	Clay	Yes	Yes	No	Yes
8	PRSW/SIC	Pulse Non-Row	Spring wheat	Silty clay	No	Yes	No	No
9	PRSW/C	Pulse Non-Row	Spring wheat	Clay	No	Yes	No	No
10	PFSW/C	Perennial Forage	Spring wheat	Clay	Yes	Yes	No	Yes
11	OSSW/C	Oilseed - Spring Cereal	Spring wheat	Clay	Yes	Yes	Yes	Yes
12	FCSW/C	Fall Cereal	Spring wheat	Clay	Yes	Yes	Yes	No
13	FCWW/C	Fall Cereal	Winter wheat	Clay	Yes	Yes	Yes	No
14	PRCA/SIC	Pulse Non-Row	Canola	Silty clay	No	Yes	No	No
15	PRCA/C	Pulse Non-Row	Canola	Clay	No	Yes	No	No
16	OSCA/C	Oilseed - Spring Cereal	Canola	Clay	Yes	Yes	Yes	Yes
17	FCCA/C	Fall Cereal	Canola	Clay	Yes	Yes	Yes	No
18	FYDL/SICL	Feedlot	–	Silty clay loam	No	No	No	Yes
19	FYDL/C	Feedlot	–	Clay	Yes	No	No	No
20	URLD/SICL	Urban (low density)	–	Silty clay loam	Yes	Yes	No	Yes
21	URLD/C	Urban (low density)	–	Clay	Yes	Yes	Yes	Yes
22	URMD/SIL	Urban (medium density)	–	Silt loam	No	No	Yes	No
23	URMD/SICL	Urban (medium density)	–	Silty clay loam	Yes	No	No	No
24	URMD/SIC	Urban (medium density)	–	Silty clay	No	Yes	No	No
25	URMD/C	Urban (medium density)	–	Clay	Yes	Yes	Yes	No
26	WETL/WA	Wetland/water	–	–	Yes	No	No	No
27	River Channel	River	–	–	Yes	Yes	Yes	Yes

945 [†] First two letters indicate cropping system/land use; third and fourth letters indicate crop; letter(s) after the slash indicate soil texture.



946

Table 2. Modules used in the customized CRHM model to simulate hydrological process in the LS-05OG008 watershed.

Sequence [†]	Module	Description
1	Basin	Holds commonly used physical and control parameters
2	Global radiation	Calculates theoretical global radiation, direct and diffuse solar radiation, as well as maximum sunshine hours based on latitude, elevation, ground slope, and azimuth (Garnier and Ohmura, 1970). This module provides radiation input to modules 3, 8, and 13
3	Observation	Adjusts the temperature and precipitation variables in the input weather data to variations in environmental lapse rate, elevation, and wind-induced under-catch
4	Crop growth	Simulates a linear crop development over the growing season
5	Sunshine hours	Estimates sunshine hours from incoming shortwave radiation and provide input to modules 8 and 13
6	Walmsley's windflow	A parametric version of the wind flow model (Walmsley et al., 1989) that adjusts the wind speed change due to local topographic features and provides input to module 11
7	Incoming long-wave radiation	Estimates incoming long-wave radiation from the air temperature and the atmospheric transmittance, which is estimated from measured short-wave radiation and theoretical global radiation and provides input to module 13 (Sicart et al., 2006)
8	All-wave radiation	Estimates the snow-free net all-wave radiation from the calculated short-wave radiation by Garnier and Ohmura (1970) and the calculated net long-wave radiation (Brun, 1932) using sunshine hours, air temperature and humidity (Granger and Gray, 1990) and provides inputs to module 9
9	Evaporation	Estimates actual evapotranspiration from unsaturated surfaces (Monteith, 1965) and evaporation from saturated surfaces such as stream channels (Priestley and Taylor, 1972). These algorithms modify the moisture content in the interception, ponded surface water, and soil column stores, as well as in the stream channel
10	Canopy	Estimates the snowfall and rainfall intercepted by the forest canopy and updates the under-canopy snowfall and rainfall and calculates short-wave and long-wave sub-canopy radiation (Ellis et al., 2010) with options for open environment (no canopy adjustment of snow mass and energy) and forest environment (adjustment of snow mass and energy from forest canopy)
11	Prairie Blowing Snow Model	Simulates snow sublimation and transport between HRUs (Pomeroy and Li, 2000)
12	Albedo	Estimates the snow albedo during the winter and the melt period (Gray and Landine, 1987). This module indicates the beginning of melt for module 13
13	Energy-Budget Snowmelt Model	Estimates snowmelt for snowpack in the Canadian Prairies (Gray and Landine, 1988) by calculating the energy balance of radiation, sensible heat, latent heat, ground heat, advection from rainfall, and change in internal energy
14	Infiltration	Estimates snowmelt infiltration into frozen soils (Gray et al., 1985) and rainfall into unfrozen soils based on texture and ground cover (Ayers, 1959)
15	Hydraulic conductivity estimator	Darcy's law for unsaturated flow is used to calculate the drainage factors of module 16 utilizing the unsaturated hydraulic conductivity calculated using Brooks and Corey relationship (Fang et al., 2013)
16	Soil	Estimates the soil moisture, groundwater flow, and interactions between ground- and surface-water throughout the year (Leavesley et al., 1983; Dornes et al., 2008; Fang et al., 2010)



17	Volumetric soil moisture	Converts soil moisture to volumetric equivalent using the variables from module 16 and determines fall status for module 14
18	Muskingum routing	Routes runoff between HRU and sub-basins using the Muskingum method (Chow, 1964)

947 † Sequence in which the modules were entered into each CRHM Group using the Macro window.



948 Table 3. Streamflow characteristics of the study area for the years used in the analysis.

Year	Water Yield (mm)	Discharge				Peak Date	Ice conditions	Backwater
		Annual [†] (m ³)	Snowmelt (m ³)	Snowmelt Proportion (%)	Peak (m ³ s ⁻¹)			
1992	64	1.21 × 10 ⁷	7.27 × 10 ⁶	60	6.7	April 10 th	March 1 st – April 12 th	No data for analysis
1993	66	1.25 × 10 ⁷	4.51 × 10 ⁶	36	5.6	April 07 th	March 1 st – April 10 th	No data for analysis
1994	18	3.33 × 10 ⁶	9.32 × 10 ⁵	28	0.7	April 10 th	March 1 st – April 16 th	No data for analysis
1995	61	1.15 × 10 ⁷	6.56 × 10 ⁶	57	5.0	March 31 st	March 1 st – April 15 th	No data for analysis
1996	99	1.87 × 10 ⁷	1.35 × 10 ⁷	72	13.5	April 29 th	March 1 st – April 28 th	No data for analysis
2002	10	1.94 × 10 ⁶	1.94 × 10 ⁶	100	1.6	April 16 th	March 1 st – April 16 th	No backwater
2003	18	3.49 × 10 ⁶	3.49 × 10 ⁶	100	2.1	April 2 nd	March 1 st – April 09 th	No backwater
2005 [‡]	83	1.57 × 10 ⁷	4.54 × 10 ⁶	29	10.0	April 08 th	March 1 st – April 04 th	April 3 rd – April 14 th June 29 th – July 24 th
2006	115	2.18 × 10 ⁷	2.18 × 10 ⁷	100	16.5	April 10 th	March 1 st – April 07 th	April 3 rd – April 19 th
2007	38	7.26 × 10 ⁶	7.26 × 10 ⁶	100	4.6	April 12 th	March 1 st – April 05 th	March 31 st – April 13 th
2009	89	1.69 × 10 ⁷	1.69 × 10 ⁷	100	13.3	April 17 th	March 1 st – April 16 th	April 11 th – April 26 th
2010	110	2.09 × 10 ⁷	1.04 × 10 ⁷	50	10.7	June 1 st	March 1 st – April 05 th	May 30 th – June 06 th September 2 nd – September 3 rd
2011	150	2.84 × 10 ⁷	1.59 × 10 ⁷	56	15.7	April 13 th	March 1 st – April 13 th	April 04 th – April 20 th May 20 th – May 21 st
2012	26	4.89 × 10 ⁶	1.71 × 10 ⁶	35	2.5	May 29 th	March 1 st – March 24 th	No backwater
2013	76	1.44 × 10 ⁷	9.38 × 10 ⁶	65	9.4	May 04 th	March 1 st – May 1 st	April 28 th – May 08 th

949 [‡] Backwater issues late in the summer; [†] Total flow from March 1st to October 31st.



950 Table 4. Metrics used for model assessment.

Year	NSE	MB	RMSD	NRMSD
1992	0.69	-0.10	0.51	0.89
1993	-0.79	0.31	1.13	1.92
1994	-27.91	2.00	0.68	4.31
1995	-0.58	0.81	1.10	2.02
1996	0.81	0.50	1.00	1.14
2002	-14.04	2.01	0.96	3.92
2003	-1.08	0.16	0.69	1.58
2005	0.85	-0.12	0.93	0.77
2006	0.76	-0.38	2.31	0.84
2007	-0.43	0.69	1.65	1.81
2009	0.65	-0.31	2.18	1.02
2010	0.36	-0.26	1.50	1.52
2011	0.84	-0.29	0.98	0.73
2012	-0.60	0.30	0.27	1.16
2013	0.72	0.36	0.91	1.34

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952



953 Table 5. Percent change in peak discharge, total discharge, and snow water equivalent (SWE) between falsified models and base models.

Water Year	Snow sublimation			Blowing snow			Frozen-soil infiltration			Frozen-soil infiltration and blowing snow		
	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE
1992-1993	39.4	12.3	23.0	69.5	7.1	23.0	-65.7	-2.8	0.0	-65.0	-5.3	23.0
1993-1994	-3.8	2.1	24.8	-2.0	-2.4	24.8	-34.6	18.4	0.0	-34.4	24.4	24.8
1994-1995	-11.3	-1.9	41.8	-13.1	-1.5	41.8	-77.9	-24.3	0.0	-78.1	-25.3	41.8
1995-1996	24.3	8.7	39.5	47.4	5.3	39.5	-78.5	-29.2	0.0	-82.1	-38.0	39.5
1996-1997	180.1	45.7	39.7	350.5	47.3	39.7	-32.7	8.4	0.0	-43.1	-12.5	39.7
1997-1998	29.5	11.4	19.9	33.6	-1.3	19.9	-75.5	5.4	0.0	-75.4	5.3	19.9
1998-1999	7.3	4.9	62.8	3.1	-2.4	62.8	-42.4	43.4	0.0	-41.8	46.0	62.8
1999-2000	14.3	4.8	30.4	14.6	1.6	30.4	-41.4	8.1	0.0	-41.3	7.2	30.4
2000-2001	19.0	17.0	13.9	16.0	18.4	13.9	-79.7	-12.8	0.0	-79.7	-17.9	13.9
2001-2002	25.1	10.3	22.0	30.5	3.4	22.0	21.9	13.7	0.0	24.0	12.6	22.0
2002-2003	17.0	6.1	151.3	23.8	-1.9	151.3	-72.6	42.7	0.0	-72.2	44.1	151.3
2003-2004	-8.2	9.2	17.3	-1.3	16.3	17.3	-80.1	-21.3	0.0	-79.8	-26.0	17.3
2004-2005	30.4	9.1	19.4	27.5	2.2	19.4	-60.0	-7.1	0.0	-60.0	-12.4	19.4
2005-2006	16.2	8.9	22.3	10.5	1.0	22.3	-89.8	-27.0	0.0	-92.4	-30.8	22.3
2006-2007	-4.0	18.9	28.7	-4.0	9.6	28.7	-84.7	-11.1	0.0	-84.5	-15.0	28.7
2007-2008	99.3	15.4	53.0	89.3	-4.5	53.0	-52.0	14.5	0.0	-62.3	10.0	53.0
2008-2009	36.5	10.7	16.2	49.1	5.2	16.2	-80.4	-4.1	0.0	-76.2	-2.7	16.2
2009-2010	1.7	-0.6	23.6	22.3	-5.8	23.6	-36.1	6.6	0.0	-37.0	5.8	23.6
2010-2011	2.9	6.5	19.6	21.8	5.6	19.6	-80.7	-18.8	0.0	-83.8	-24.7	19.6
2011-2012	16.1	5.0	54.2	4.0	-2.2	54.2	-53.8	36.9	0.0	-53.2	38.9	54.2
2012-2013	2.5	2.3	13.1	28.0	5.6	13.1	-79.5	-29.4	0.0	-82.8	-34.3	13.1
Average	25.4	9.8	35.1	39.1	5.1	35.1	-60.8	0.5	0.0	-62.0	-2.4	35.1

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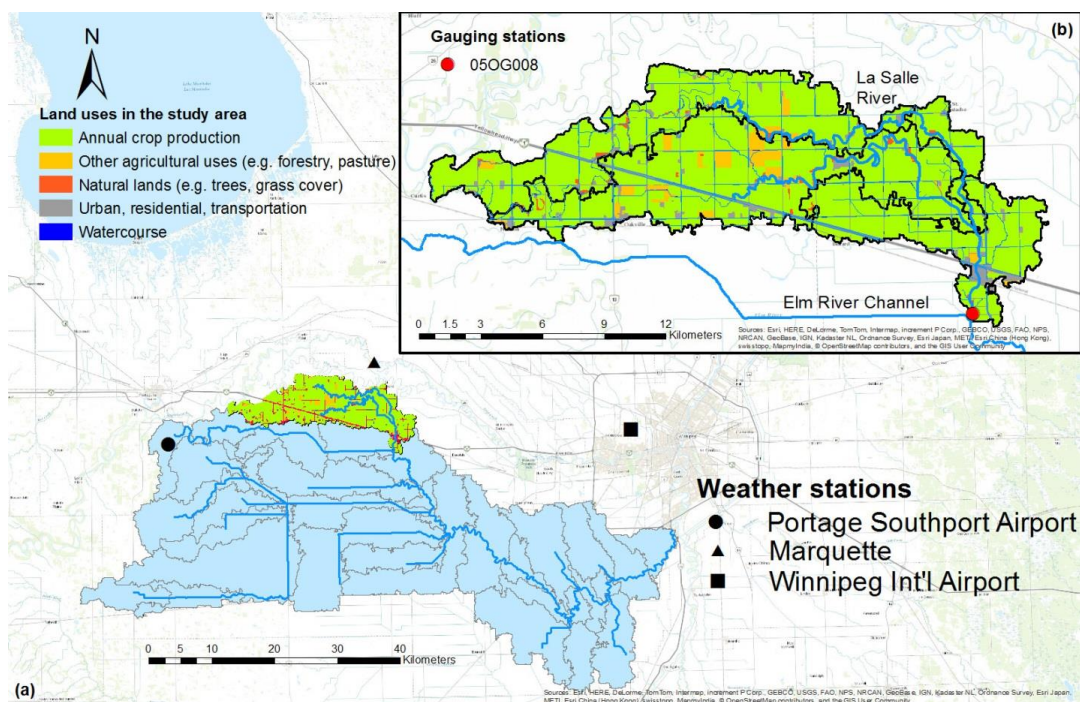
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Table 6. Selected statistical metrics (i.e. NSE and Model Bias) for falsified models.

Year	Snow sublimation		Blowing snow		Frozen-soil infiltration		Frozen-soil infiltration and blowing snow	
	NSE	MB	NSE	MB	NSE	MB	NSE	MB
1992	-0.43	0.17	-0.62	0.10	-0.21	-10.80	-0.15	-0.10
1993	-0.80	0.47	-0.59	0.40	-0.02	13.60	0.02	0.09
1994	-29.18	2.09	-28.89	1.91	-11.43	216.00	-11.54	2.22
1995	-0.23	0.77	-0.59	0.79	0.31	26.30	0.34	0.23
1996	0.48	0.64	0.12	0.59	0.22	-6.90	0.22	-0.21
2002	-39.64	2.80	-41.84	2.46	-1.37	85.00	-0.73	0.67
2003	-1.52	0.30	-1.74	0.13	-0.14	9.80	-0.01	0.05
2005	0.89	0.03	0.90	-0.08	-0.05	-40.90	-0.05	-0.50
2006	0.83	-0.29	0.81	-0.35	-0.18	-77.20	-0.19	-0.81
2007	-0.86	1.13	-0.88	0.96	-0.10	-30.10	0.00	-0.44
2009	0.68	-0.16	0.66	-0.16	-0.07	-64.50	0.02	-0.63
2010	0.35	-0.26	0.33	-0.30	0.30	-28.80	0.31	-0.31
2011	0.89	-0.23	0.86	-0.24	0.10	-47.30	0.10	-0.53
2012	-1.16	0.38	-0.64	0.26	-0.14	63.40	-0.29	0.68
2013	0.67	0.39	0.09	0.44	0.01	-12.00	0.05	-0.19
Average	-4.60	0.55	-4.80	0.46	-0.85	6.37	-0.79	0.02



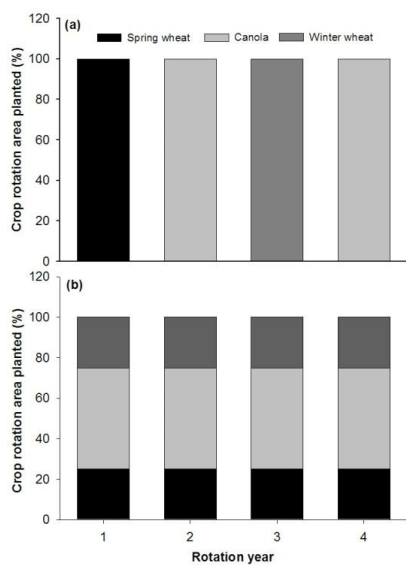
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958 Figure 1. The La Salle watershed (a) and the sub-watershed used in the study, which drains into Water
959 Survey Canada gauging station 05OG008 (b).

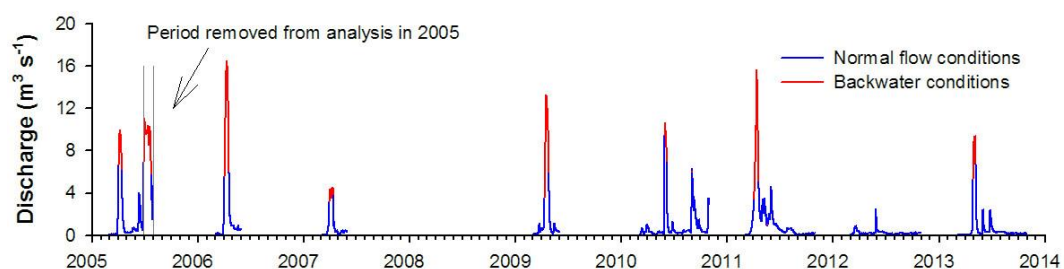
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962 Figure 2. Example of how the Fall Cereal Cropping System rotation 4-year rotation (a) was represented every
963 year in the model using the land-use split method (b).

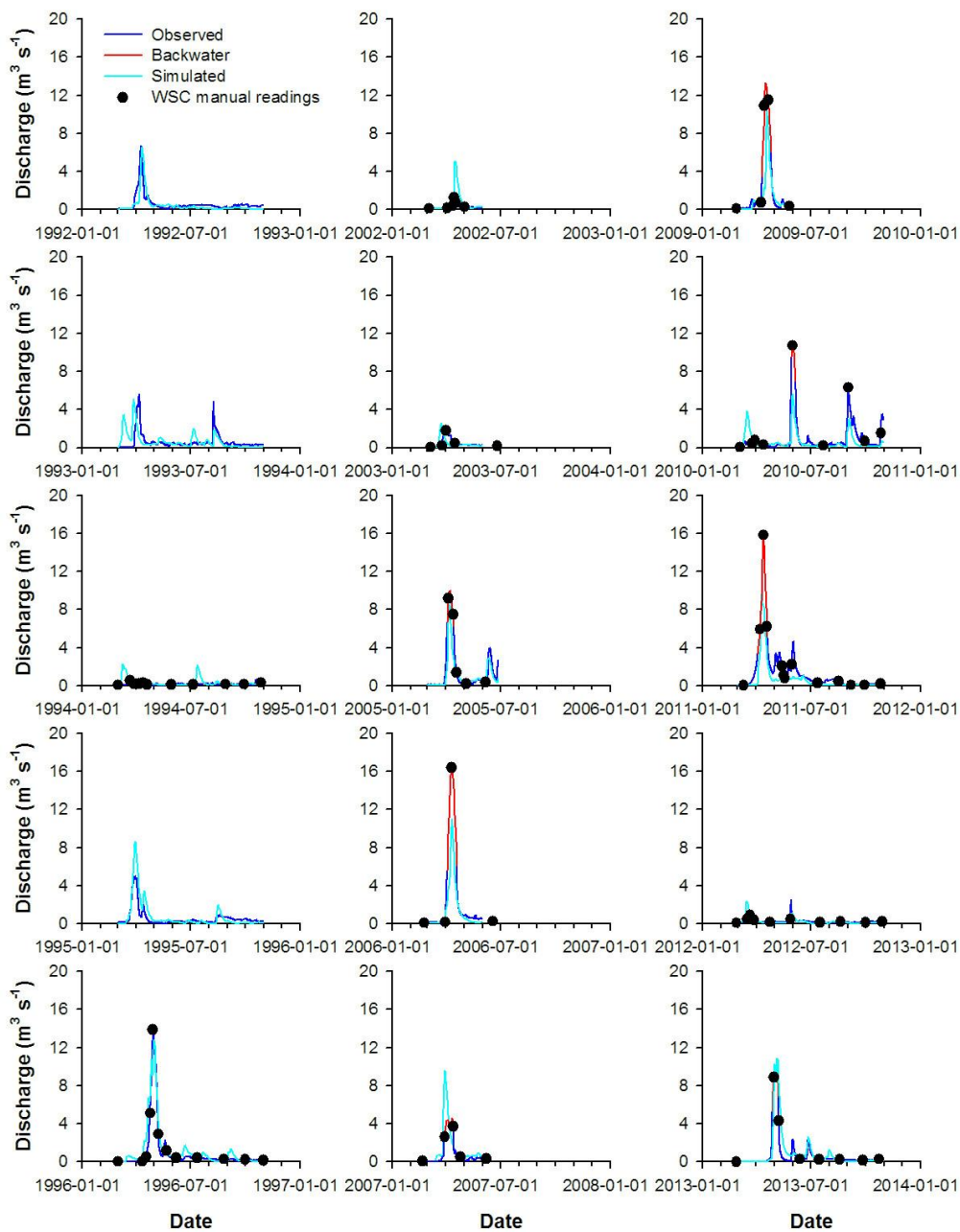
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966 Figure 3. Yearly hydrographs indicating periods of potential backwater issues. Years of 2002, and 2003 not
967 presented since these years did not have backwater issues. Years of 2004 and 2008 were not included in the analysis
968 since quality issues were identified in the metadata of the records. The period between lines in 2005 indicate the
969 records removed from the dataset during model assessment.

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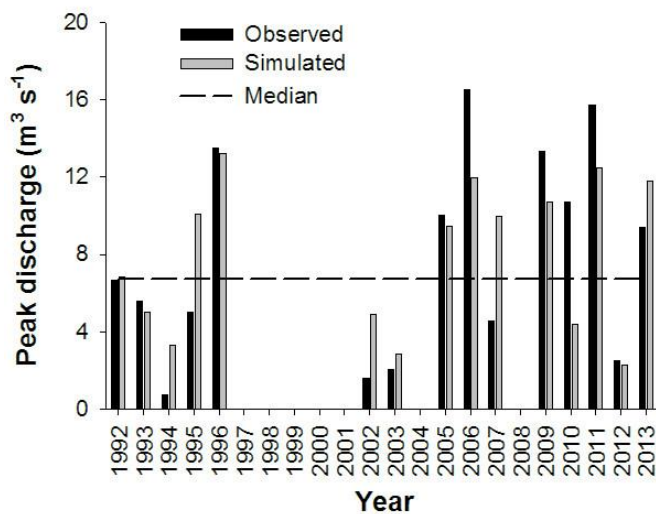


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Figure 4. Comparison of observed and simulated stream discharge between 1992 and 2013 for years with good records in the HYDAT database.

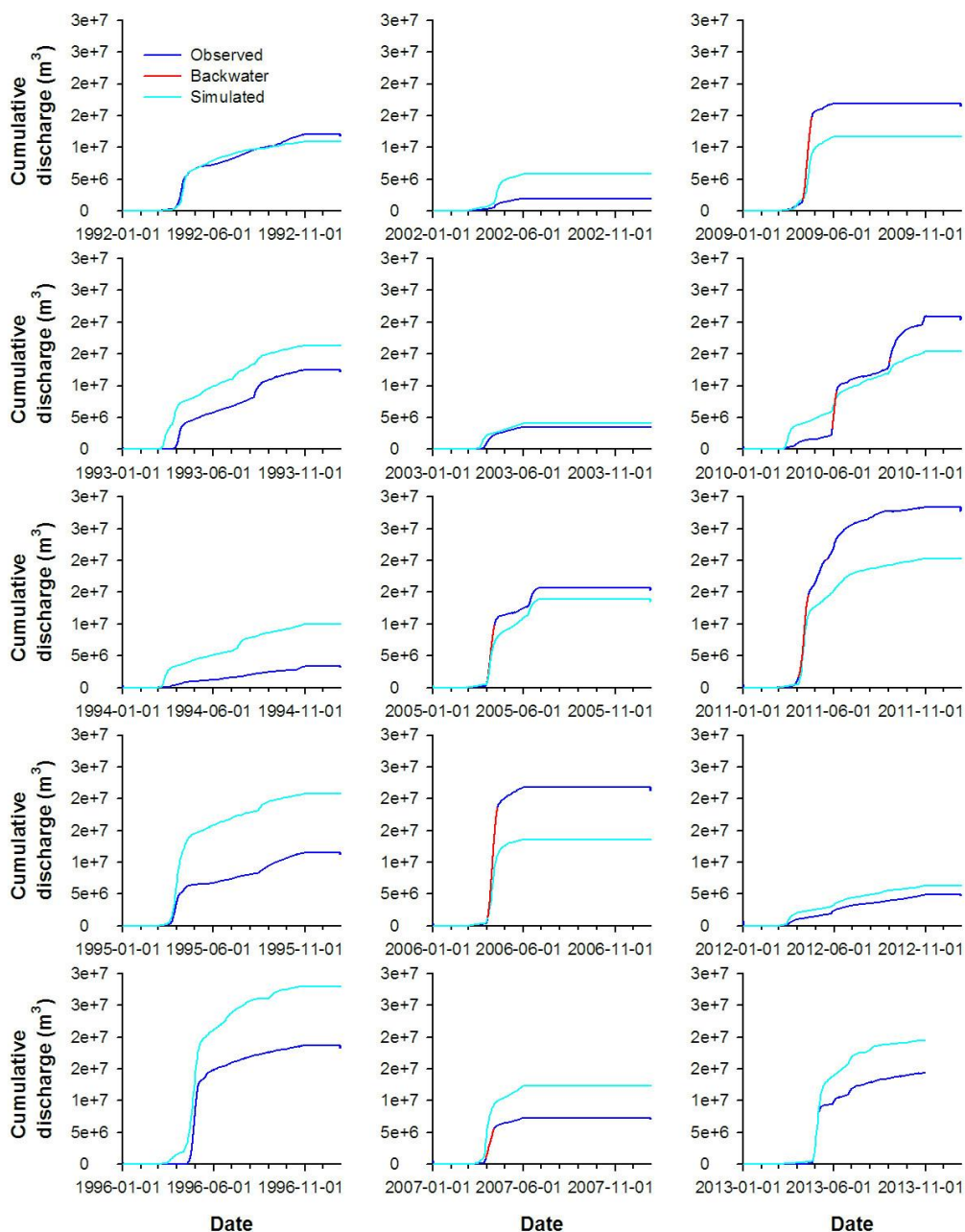
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976 Figure 5. Comparison of observed and simulated peak discharge between 1992 and 2013 for years with good
977 records in the HYDAT database. No data available in the HYDAT database between 1997 and 2001. Years of 2004
978 and 2008 were not included in the analysis since quality issues were identified in the metadata of the records.

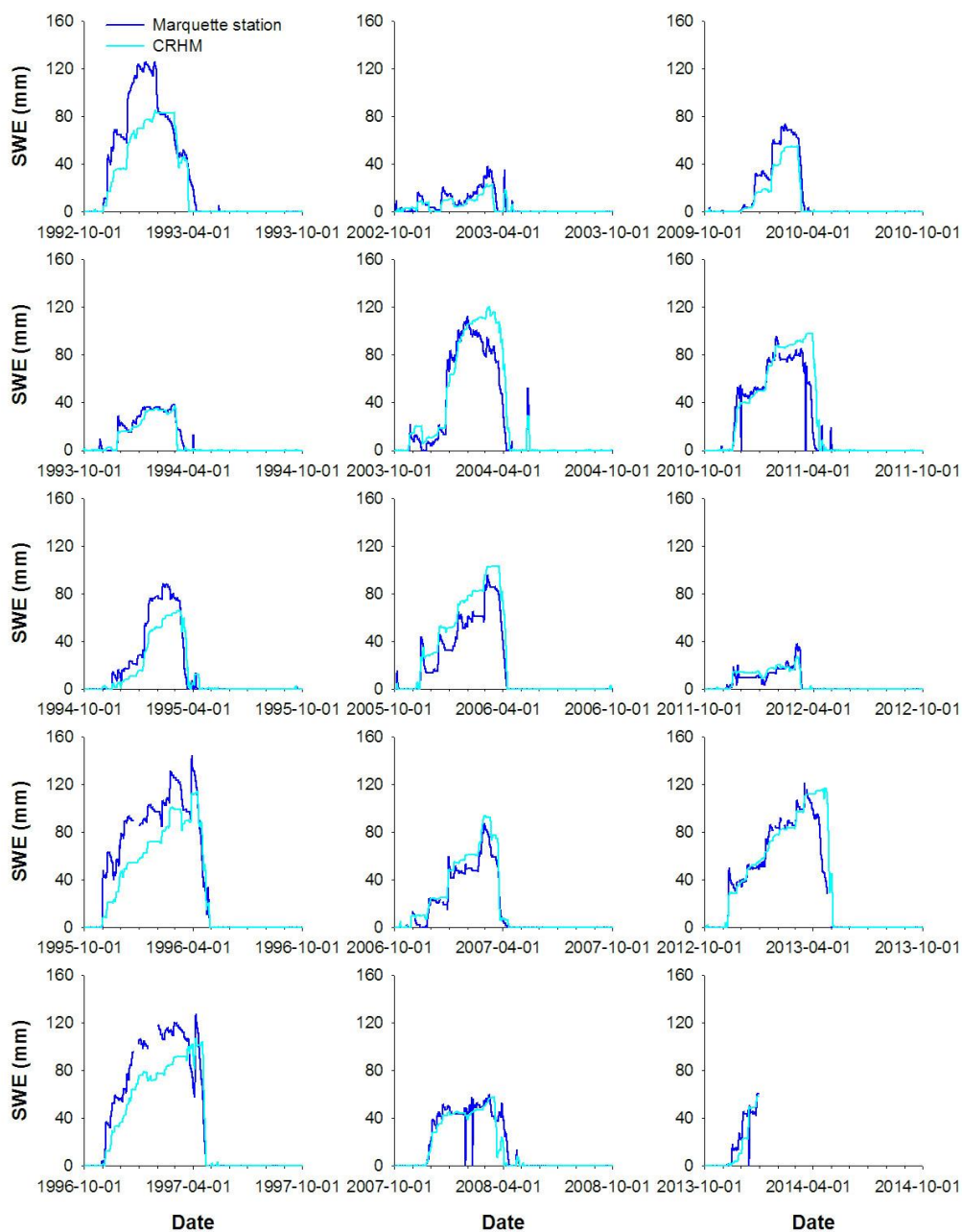
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981 Figure 6. Comparison of observed and simulated annual cumulative discharge between 1992 and 2013 for
 982 years with good records in the HYDAT database.

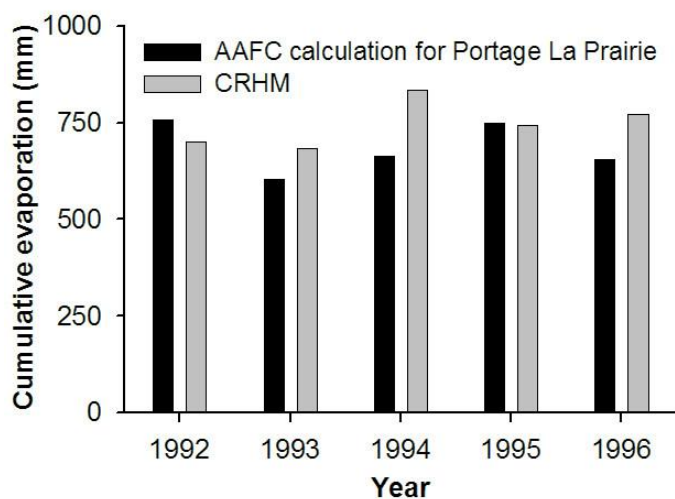
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985 Figure 7. Comparison of observed and simulated snow water equivalent (SWE) between 1992 and 2013 for
986 years with good records in the HYDAT database. SWE was calculated assuming a snow density of 180 kg m^{-3} .

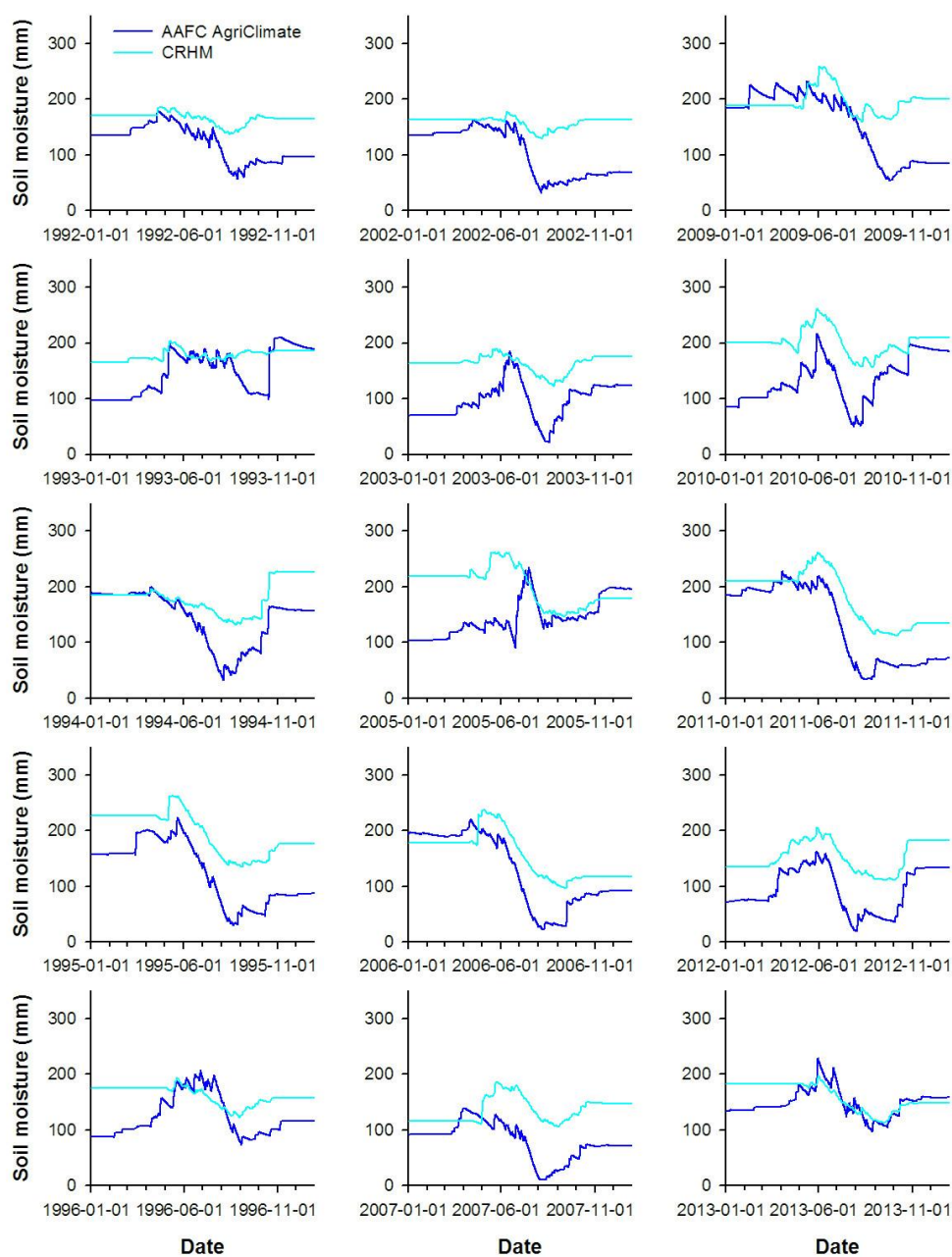
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989 Figure 8. Comparison of evaporation estimates produced by the Agriculture and Agri-Food Canada (AAFC)
990 and CRHM between 1992 and 1996.

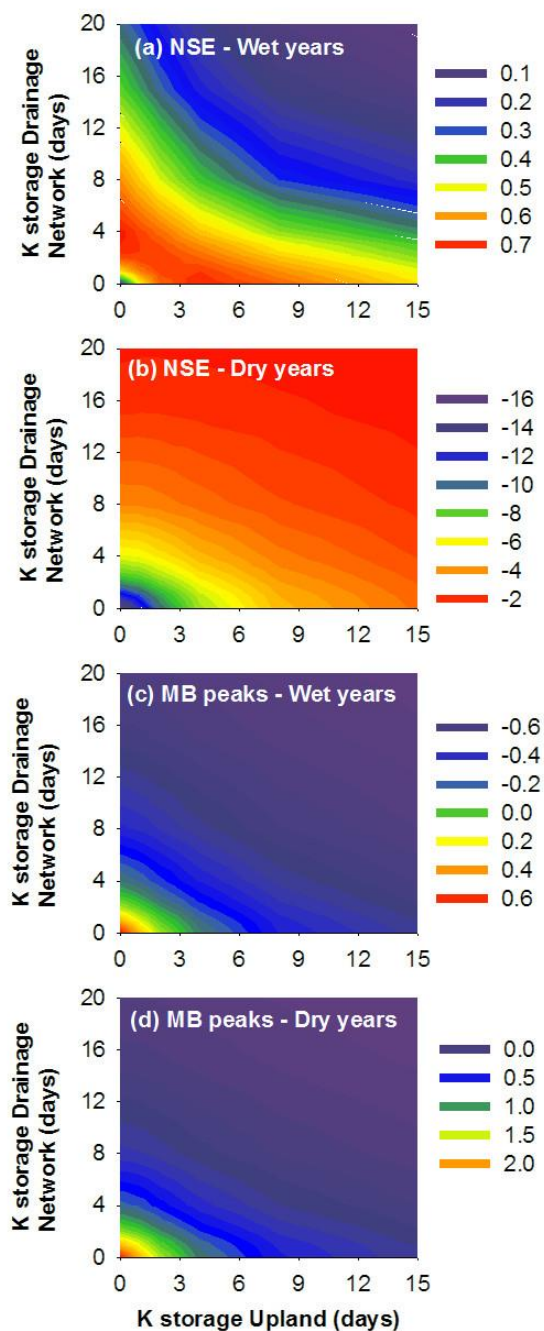
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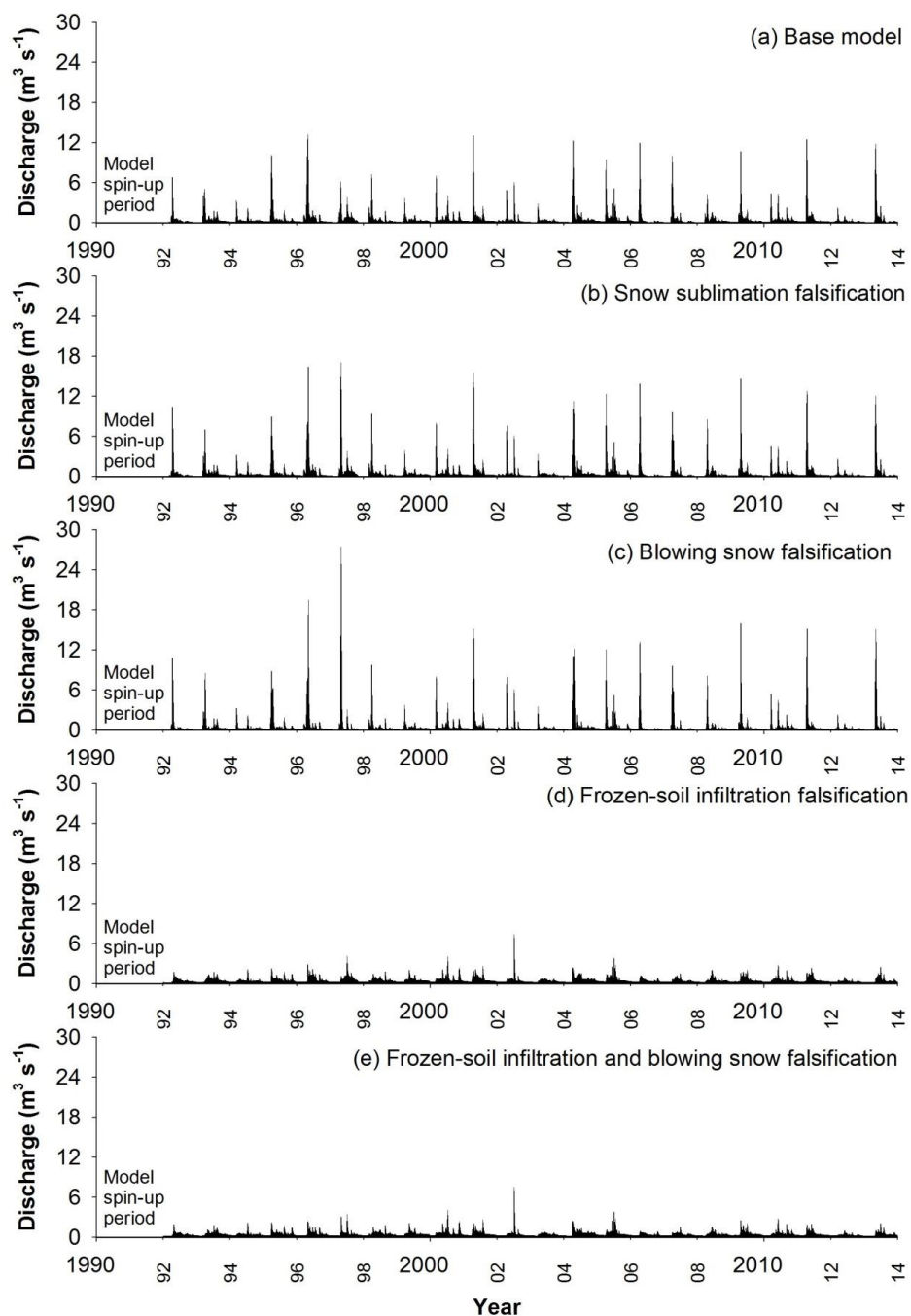
993 Figure 9. Comparison between simulated volumetric soil water content (VWC; expressed in mm of water in
994 the soil profile) produced by the National Drought Model (NDM) and CRHM between 1992 and 2013 for years with
995 good records in the HYDAT database.

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998 Figure 10. Sensitivity analysis of the K storage parameter in the Muskingum model for up land and drainage
999 network using Nash-Sutcliff efficiency (NSE) and peak-flow model bias (MB) as objective functions.



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Figure 11. Hydrographs of the entire simulation period (1992-2013) for the base model (a) and the different model falsifications, which include inhibition of snow sublimation (b), blowing snow (c), frozen-soil infiltration (d) and frozen-soil infiltration combined with blowing snow (e).