

1 **Simulating cold-region hydrology in an intensively drained agricultural watershed in**
2 **Manitoba, Canada, using the Cold Regions Hydrological Model**

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12 **Abstract**

13 Eutrophication and flooding are perennial problems in agricultural watersheds of the
14 northern Great Plains. A high proportion of annual runoff and nutrient transport occurs with
15 snowmelt in this region. Extensive surface drainage modification, frozen soils, and frequent
16 backwater or ice damming impacts on flow measurement represent unique challenges to
17 accurately modelling watershed scale hydrological processes. A physically-based, non-calibrated
18 model created using the Cold Regions Hydrological Modelling platform (CRHM) was
19 parameterized to simulate hydrological processes within a low slope, clay soil, and intensively
20 surface drained agricultural watershed. These characteristics are common to most tributaries of
21 the Red River of the North. Analysis of the observed water level records for the study watershed
22 (La Salle River) indicates that ice cover and backwater issues at time of peak flow may impact
23 the accuracy of both modelled and measured stream flows, highlighting the value of evaluating a

24 non-calibrated model in this environment. Simulations best matched the streamflow record in
25 years when peak and annual discharges were equal to or above the medians of $6.7 \text{ m}^3 \text{ s}^{-1}$ and 1.25
26 $\times 10^7 \text{ m}^3$, respectively, with an average Nash-Sutcliff efficiency (NSE) of 0.76. Simulation of
27 low-flow years (below the medians) was more challenging (average NSE <0) with simulated
28 discharge overestimated by 90% on average. This result indicates the need for improved
29 understanding of hydrological response in the watershed under drier conditions. Simulation
30 during dry years was improved when infiltration was allowed prior to soil thaw indicating the
31 potential importance of preferential flow. Representation of in-channel dynamics and travel time
32 under the flooded or ice-jam conditions should also receive attention in further model
33 development efforts. Despite the complexities of the study watershed, simulations of flow for
34 average to high flow years and other components of the water balance were robust [snow water
35 equivalency (SWE) and soil moisture]. A sensitivity analysis of the flow routing model suggests
36 a need for improved understanding of watershed functions under both dry and flooded conditions
37 due to dynamic routing conditions, but overall CRHM is appropriate for simulation of
38 hydrological processes in agricultural watersheds of the Red River. Falsifications of snow
39 sublimation, snow transport, and infiltration to frozen soils processes in the validated base model
40 indicate that these processes were very influential to stream discharge generation.

41 **KEY WORDS:** Prairie hydrology; cold-region; snowmelt; physically-based modelling;
42 agricultural watershed; intensively drained.

43 **1. Introduction**

44 The Red River Basin spans over $122,730 \text{ km}^2$ and encompasses portions of Canada
45 (provinces of Manitoba and Saskatchewan) and United States (North Dakota, South Dakota, and
46 Minnesota), with almost three quarters of its land used for agriculture (Benoy et al., 2016). As a

47 result, most of the hydrology and nutrient transport in that basin is strongly influenced by this
48 land use. The interaction between agricultural management and the hydrology of cold regions
49 remains a topic of ongoing research (Rahman et al., 2014; King et al., 2015) given the need for
50 improved understanding of processes such as prevalence of preferential flow paths, enhanced
51 hydrological connectivity promoted by drainage, and impact of different cropping systems on
52 runoff generation. Hydrological models have been utilized at varying spatial scales to model the
53 hydrology of agricultural areas of cold-climate countries such as Finland (Grizzetti et al., 2003;
54 Knisel and Turtola, 2000), Russia (Schierhorn et al., 2014a; Schierhorn et al., 2014b), and
55 Canada (Yang et al., 2014; Yang et al., 2009). Overall, little research addressing specificities of
56 agriculture in cold region hydrology is available in the literature, although this activity is quite
57 relevant in northern latitude regions such as the northern great plains [North America;
58 (Desaulniers and Gritzner, 2006; Li et al., 2010; Sharp, 1952; Wishart, 2004)], northwest Europe
59 [Scandinavia; (Parry et al., 1988)], and northern Asia (Blanke et al., 2007; Wang et al., 2002).
60 Important challenges remain to modelling the hydrology of northern latitude agricultural
61 watersheds such as integrated modelling of cropping systems and hydrology, representation of
62 processes across spatial scales, and enhanced hydrologic connectivity. For example, crops
63 represented only 30% of the land use in a study in Finland using the Soil and Water Assessment
64 Tool (SWAT) making direct connection to agricultural process more challenging (Grizzetti et al.,
65 2003), while an application of the GLEMS model in that country was done at plot scale (0.11 ha
66 in area) providing detail, but making scaling up a challenge (Knisel and Turtola, 2000). SWAT
67 simulations in Russia have focussed on productivity aspects of wheat only rather than
68 hydrological implications (Schierhorn et al., 2014b; Schierhorn et al., 2014a) and SWAT
69 exercises in Canada have focussed on finer scale simulation of a specific management practices

70 rather than evaluating process representation (i.e. 14.5 km² watershed area; Yang et al., 2009) or
71 were assessed on too coarse (i.e. monthly) time step for physically based modelling of snowmelt
72 (Yang et al., 2014).

73 Many popular models utilized for the simulation of hydrological processes in agricultural
74 watersheds were initially developed for regions where rainfall driven runoff is the primary
75 contributor to annual water yield. Snowmelt is the main source of streamflow in tributaries of
76 the Red River and accurate representation of hydrological processes in its agricultural tributaries
77 using popular models like SWAT has proved challenging. While SWAT has been used to
78 estimate water quality targets and beneficial management practices in Canada (Yang et al.,
79 2012), significant modification of the original model is generally required to achieve a good fit
80 with calibration data (Watson et al., 2008; Liu et al., 2012) and improvement of the accuracy of
81 underlying representation of important hydrological processes (e.g. landscape representation,
82 stream routing) is ongoing (Douglas-Mankin et al., 2010).

83 As a result of the problems associated with adapting hydrological models developed for
84 more temperate conditions, a number of models that incorporate cold-region hydrological
85 processes have been developed, such as ARHYTHM (Zhang et al., 2000), VIC (Cherkauer et al.,
86 2003), the TH-REW model (Mou et al., 2008; Tian et al., 2006), and the Cold Regions
87 Hydrological Modelling (CRHM) platform (Pomeroy et al., 2007). Of these models, CRHM
88 offers the most complete range of physically-based process representation for the Northern Great
89 Plains, including blowing snow, interception and sublimation of snow, energy balance snowmelt,
90 canopy influence on radiation, and infiltration to frozen soils (Pomeroy et al., 2007; Fang et al.,
91 2013; Fang et al., 2010). Although CRHM currently does not feature a module to represent
92 nutrient dynamics in either soil or water, such modules are under development (Roste, 2015) and

93 the platform is a powerful tool for assessing watershed nutrient dynamics because of the well-
94 established relationships between phosphorus concentrations and discharge rate in tributaries of
95 the Red River, including the La Salle River (McCullough et al., 2012), which is the focus of the
96 research presented here.

97 The CRHM platform has been successfully used to simulate hydrological processes in a
98 number of contrasting catchments in Canada (Dornes et al., 2008; Fang and Pomeroy, 2008; Ellis
99 et al., 2010), globally in western China (Zhou et al., 2014), Patagonia (Krogh et al., 2015),
100 German Alps (Weber et al., 2016), and Spanish Pyrenees (Rasouli et al., 2014), and in the
101 intensively studied South Tobacco Creek watershed that drains from the more steeply sloped
102 Pembina/Manitoba Escarpment feature into the Red River in Manitoba (Mahmood et al., 2017).
103 However, CRHM has not yet been used to simulate hydrological processes in the intensively
104 managed lowland agricultural tributaries of the Red River such as the La Salle River.
105 Characteristics in these watersheds include extensive artificial surface drainage, channelization
106 of the stream network, historical drainage of wetlands, clay soils, high fertilizer input, high crop
107 yields, high livestock densities, and highly connected drainage areas with little surface storage.
108 Simulation of cold-region hydrological processes in this landscape is particularly challenging
109 because of the combined effect of climate and land use management on water transport. Large
110 volumes of runoff can be produced when snowmelt is routed over frozen soils (Shook and
111 Pomeroy, 2010; Shook and Pomeroy, 2012) and this process is emphasized where surface
112 drainage enhancements and roadside ditches speed transport (Brunet and Westbrook, 2012;
113 Pomeroy et al., 2014). As a result, flows in streams of the Red River watershed tend to exhibit a
114 flashy response to snowmelt with large volumes of runoff from roadside ditches entering into
115 relatively small river channels that are still covered by ice at time of melting. This in-channel ice

116 restricts flow and frequently causes flooding and backwater conditions to develop (Gray and
117 Prowse, 1993). These conditions can lead to error in discharge measurements calculated using
118 stage-rating curves (Mosley and McKerchar, 1993) and also invalidate important assumptions of
119 flow routing models (Fread, 1993), thus making accurate prediction of flow more challenging.
120 For this reason, using a non-calibrated, but flexible physically-based model platform is an
121 attractive alternative to models primarily calibrated based on discharge measurements. The
122 application of CRHM will enhance the understanding of the hydrological controls in tributaries
123 of the Red River since concurrent measurements such as SWE and soil moisture can also be
124 utilized to gauge the accuracy of model process representations independent of uncertainties
125 associated with in-stream flow representation or measurement. At the same time, the limitations
126 of existing conceptual models and/or potential problems with discharge measurement should be
127 more apparent in the absence of calibration (Spence et al., 2013).

128 While CRHM has performed well in a number of watersheds across Canada, its
129 performance has not been evaluated in intensively managed agricultural watersheds in the
130 Canadian Prairies, which is home to 80% of the country's agriculture (Wheater and Gober,
131 2015). Prairie agro-ecosystems are characterized by modified topography, annual cropping
132 systems, and enhanced hydrologic connectivity that set them apart from other cold regions
133 ecosystems and it has been challenging to accurately model these processes using other
134 platforms. The objective of this study was to perform hydrological simulations in a framework
135 specifically designed to represent agricultural systems typical of the Red River Basin, to identify
136 challenges hindering satisfactory simulation of these systems, and to utilize the process based
137 CRHM model to identify where conceptual hydrological models for cold agricultural regions
138 might be improved. Specifically, CRHM was used to simulate the hydrology of an intensively

139 managed agricultural sub-watershed of the Red River Valley and to gain insight about the
140 dominant hydrological controls of streamflow in this landscape that, although unique, embodies
141 many of the problems faced by other cold regions. For example, phosphorus transport from
142 farmland, which is a problem in the Red River Basin (Rattan et al., 2016; Yates et al., 2012), is
143 also an issue in many other cold-climate regions, including north-west European countries such
144 as Norway, Sweden, United Kingdom and Ireland (Ulén et al., 2007). The research presented
145 here provides insight into those periods of time, weather conditions, and associated process
146 representations with which poor model performance is associated, indicating either limitations of
147 model input data or the need for improved understanding and representation of watershed
148 hydrology in the region.

149 **2. Material and Methods**

150 *2.1 Study site*

151 The study was conducted in a 189 km² sub-catchment of the La Salle River watershed (LS-
152 05OG008; Fig. 1a), a tributary of the larger Red River which is located in the central plains
153 region of Manitoba, Canada (Graveline and Larter, 2006). Only the sub-catchment was used in
154 the present study due to weather data availability in an hourly time-step, which was required to
155 force some physically-based processes in CRHM, such as the Prairie Blowing Snow Model
156 (Pomeroy and Li, 2000). The study watershed is underlain by lacustrine clay deposited in glacial
157 Lake Agassiz; these deposits consist of a lower, dark grey clay and a thinner upper unit of lighter
158 coloured, calcareous silty clay, with surface texture being predominantly clayey (La Salle
159 Redboine Conservation District, 2007). The study area is located in the Prairie Ecozone, with
160 mean annual temperature around 2.5°C, mean summer temperature of 16°C and mean winter
161 temperature of -13°C; the mean annual precipitation is 560 mm, out of which around 25% takes

162 place as snow, while the potential mean annual gross evapotranspiration is about 834 mm (La
163 Salle Redboine Conservation District, 2007).

164 *2.2 Hydrological and meteorological observations*

165 Daily streamflow observations between 1990 and 2013 were obtained from the
166 hydrometric data (HYDAT) database (Environment and Climate Change Canada, 2013) for
167 Water Survey of Canada (WSC) gauging station 05OG008 (La Salle River near Elie; Fig.1b)
168 located at the outlet of the study watershed. Data collection at this location was seasonal from
169 1990 to 1996, and has been continuous from 2002 to present. The annual monitoring period for
170 this station spans from March 1st to October 31st, with no data available during winter months;
171 thus, analyses were carried out using calendar years rather than water years (i.e. October 1st to
172 September 30th) since the hydrometric data does not span over more than a calendar year. A gap
173 in available flow data exists between flooding in 1997 and instrument replacement in 2001.
174 Only flow data is available from HYDAT for the period prior to 1996, while flow and water
175 level were both recorded from 2002 onwards. Notes in the HYDAT metadata pertained to 2004
176 and 2008 indicated equipment malfunctions resulting in loss of data. For this reason, the periods
177 from 1997-2001, 2004, and 2008 were not used for model assessment.

178 The gauging station 05OG008 is located 80 meters downstream from a small water level
179 control structure, which raises concerns about ice jamming and backwater issues. The operation
180 of the control structure is not changed throughout the year (i.e. no stop-logs removed) since the
181 purpose of the structure is to increase the storage of the river channel during the summer months
182 (when discharge is low). Therefore, stream discharge is not significantly affected by human
183 operations of the control structure. Ice conditions are flagged in the HYDAT daily records, but
184 no further detail is provided. It was assumed that ice conditions meant complete or major ice

185 cover at initial ice breakup. While ice conditions are flagged in HYDAT, backwater and flooding
186 conditions are not indicated for the site. Thus, field notes from site visits were acquired from
187 WSC and used to determine potential backwater conditions. The threshold for flagging these
188 conditions in the present analysis was based on the minimum water level for which backwater
189 was recorded by WSC technicians on the field. This level was 239.3 m, observed on May 19,
190 2011, which is only 0.2 m below the full supply level (FSL) of the water level control structure
191 (239.5 m). Potential backwater conditions were assumed whenever the water level was above
192 this threshold. Since water levels in the HYDAT database were only available after 2002, this
193 analysis was carried out between 2002 and 2013. These potential backwater periods are noted in
194 figures displaying measured discharge. However, the presence of backwater conditions, although
195 common, were not consistently documented on the WSC field notes to provide verification of
196 occurrence or impact on measurement accuracy; so even where backwater conditions were
197 suspected, all flow data was assumed accurate and utilized in assessment of model performance.

198 The hourly weather data used to force CRHM was obtained from closest Environment
199 Canada weather station (Environment and Climate Change Canada, 2015) with available data
200 (Fig. 1a). Nearby stations are located at Portage Southport Airport, Winnipeg International
201 Airport, and Marquette (26.6, 47.9, and 9.9 km from the geometric centre of the study area,
202 respectively). Air temperature, wind speed, and relative humidity were obtained from the Portage
203 Southport Airport, solar radiation was acquired from the station located at the Winnipeg
204 International Airport, and precipitation was acquired from the weather station in Marquette.
205 Precipitation was available in a daily time-step and was disaggregated to an hourly time-step
206 using HyetosR (Kossieris et al., 2013), which is a package for the temporal stochastic simulation

207 of rainfall process at fine time scales based on Bartlett-Lewis rectangular pulses rainfall model
208 (Koutsoyiannis and Onof, 2001).

209 *2.3 Watershed delineation and HRU definition*

210 The soils, topography, and land use datasets used as model inputs in this study were
211 derived according to the principles described by Liu et al. (2013) for prediction in ungauged
212 basins. The sub-basins within the La Salle watershed (Fig. 1a) were defined in a previous
213 modelling exercise using the watershed delineation tool in SWAT (Yang et al., 2014) and the 90-
214 m digital elevation model (DEM) derived from the NASA Shuttle Radar Topography Mission
215 data. Soil datasets with scales ranging from 1:20,000 to 1:126,720 were obtained from the
216 Manitoba Land Initiative (MLI) database. The soils textures in sub-catchment were 94% clay,
217 4% silt clay loam, 1% silt clay, and 1% silt loam. The land use datasets were acquired from MLI
218 for non-agricultural land uses and from Agriculture and Agri-Food Canada (AAFC) for cropping
219 systems. Cropping systems, or typical cropping rotations, were derived from nine years (2001-
220 2009) of records from the Manitoba Agricultural Services Corporation (MASC) crop insurance
221 database, as well as land use defined by the Ag-Capture tool between 2009 and 2012 that
222 included each parcel of land in the study area (Yang et al., 2014). The Ag-Capture tool is a
223 geospatial, agricultural land-use inventory and mapping tool developed by Agriculture and Agri-
224 Food Canada (OECD, 2013) that validates remote-sensed land use classification using field
225 surveys. The cropping system assigned to each parcel was based on the dominant or most
226 common crop grown while its representative crop rotation was based on the proportion of the
227 other most common crops grown on the same land parcel over the entire period of record.
228 Cropping systems used in the present model and the crops therein are shown in Table 1.

229 Using SWAT's watershed delineation tool, the entire La Salle watershed was divided into
230 38 sub-basins (Fig. 1a) (Yang et al., 2014). The study area was comprised of sub-basins 2, 3, 5
231 and 33; they correspond to the gross drainage area for the gauging station at the outlet of the
232 overall watershed (05OG008; Fig.1b). The hydrological response units (HRUs) used by CRHM
233 in this study were also based on the HRUs previously defined for the SWAT model (Yang et al.,
234 2014), where HRUs are defined based on unique combinations of land use, soil, and topography.
235 The HRU-definition strategy for applying cropping systems was based on the land-use split
236 method (supplement material).

237 Due to the highly altered nature of the drainage network in the watershed and its low relief,
238 data collection was undertaken to ensure accurate delineation of routing. Initially the stream
239 network was defined based on digital water feature data from the Manitoba Land Initiative (MLI)
240 (1:20,000 Designated Drains) with supplemental ephemeral streams digitized based on aerial
241 photograph interpretation. The drainage ditch dataset, mainly comprised of roadside ditches,
242 utilized a culvert inventory along with LiDAR DEM (vertical accuracy ± 0.15 m) analysis to
243 determine flow direction to create ditch line segments that passed through culverts and ultimately
244 connected to the stream network. The stream and drainage ditch datasets were merged together
245 to create the final drainage network for the model (supplement material).

246 *2.4 CRHM module selection*

247 The Cold Region Hydrological Modelling platform (CRHM) was used to develop a custom
248 hydrological model for LS-05OG008 (a flow chart summarizing the analysis described in the
249 following sections is provided as supplementary material). CRHM is a modular object-oriented
250 platform that allows the creation of customizable models varying from conceptual to physically-
251 based representations of the study area, according to the data availability and uncertainty in

252 process parameters of the basin (Harder and Pomeroy, 2014). A detailed description of CRHM
253 and its modules, which are based on decades of cold-region hydrological research in western and
254 northern Canada, is available in Pomeroy et al. (2007) , with an update for agricultural
255 applications in the Canadian Prairies described by Fang et al. (2010). Physically-based modules
256 were defined and arranged sequentially to simulate the dominant hydrological processes in LS-
257 05OG008. Table 2 lists the modules selected, their function, and the sequence in which they
258 were entered into the customized model (a chart describing the module structure is provided as
259 supplementary material). Similar model group structures have been successfully used to simulate
260 hydrological processes in the prairie pothole region of the Canadian Prairies (Fang and Pomeroy,
261 2008; Fang et al., 2010). The same model structure was applied to the four sub-basins of LS-
262 05OG008 (i.e., sub-basins 2, 3, 5 and 33). Each sub-basin was represented in the model by a
263 Group, which is a collection of modules executed in sequence for all HRUs and connected with
264 one another by a Muskingum routing group.

265 *2.5 CRHM module parameterization*

266 The CRHM software does not make provision for calibration utilizing fitted variables;
267 rather, parameters are selected based on the understanding of the hydrological system (Pomeroy
268 et al., 2007). Since the land-use split approach was used, the HRU distribution was held constant
269 over the simulation period, which allowed for a single set of parameters to be used in the model
270 for each HRU. Sub-basin and HRU physiographic parameters (i.e. area, elevation, slope, aspect,
271 and latitude) were obtained from the DEM and HRU reports generated by SWAT during the
272 watershed delineation and HRU definition for a previous SWAT simulation in the La Salle
273 River watershed (Yang et al., 2014) . The parameters of the soil and recharge layers (e.g. soil
274 texture, soil porosity, saturated hydraulic conductivity, and water holding capacity) were

275 obtained or derived from the Soil Survey Reports published by the Province of Manitoba
276 (Michalyna et al., 1972; Ehrlich et al., 1953). The pore size distribution index (λ) was defined
277 based on soil textures associated with clay (Brooks and Corey, 1966; Corey, 1994).

278 Crop seeding dates were defined based on a 10-year average (2000-2009) for each crop
279 based on crop insurance data obtained from MASC. Harvest dates were based on the length of
280 the growing season defined by consultation with farmers in the study area. Leaf area index
281 (LAI), vegetation height, and stalk diameter for the different crops were obtained from the
282 dataset collected by AAFC for the SMAPVEX12 Campaign in 2012 (Kim et al., 2014). The
283 SMAPVEX12 field sites were located between 13 and 45 km south of LS-05OG008 and are
284 located within the Red River watershed in the same or similar ecoregions as the La Salle River
285 watershed (Lake Manitoba Plain or Interlake Plain). The Penman–Monteith method was used to
286 calculate evapotranspiration from crop HRUs, while the Priestley-Taylor method was used to
287 calculate evaporation from open-water HRUs (e.g. wetland, river channel).

288 Although located in the Canadian Prairies, the runoff routing in the LS-05OG008 sub-
289 catchment does not follow the typical sequence of land uses further west in this region described
290 by Fang et al. (2010) (i.e. fallow, stubble, and pasture routed to woodland and then to wetland,
291 open water, and river channel). Rather, the flat and intensively managed characteristics of the La
292 Salle watershed result in a lack of any typical routing sequence based on land use. For this
293 reason, uplands were routed directly to drainage channels in LS-05OG008. For each HRU, the
294 routing length was calculated as the median of the distances from each HRU to the drainage
295 network, as obtained using ‘near’ tool in ArcGIS 10.1 (ESRI, Redlands, California). The
296 distances to the drainage network were estimated from the 2009 land use coverage using the Ag-
297 Capture tool.

298 The maximum depression storage of the HRUs was calculated using the methodology
299 described by Fang et al. (2010) for upland areas, where volume and area were calculated from
300 the same LiDAR DEM used to derive the final drainage network. Briefly, the ‘fill’ tool was used
301 to level off the original DEM and the ‘cut/fill’ tool used to calculate the area and volume from
302 the difference between the original and the leveled DEMs. Muskingum routing within and
303 between sub-basins of LS-05OG008 was setup according to Fang et al. (2013). The storage
304 parameter used in the Muskingum routing module was set to zero, based on the typical pattern
305 observed for regional agricultural runoff monitoring where individual diurnal runoff events at
306 edge-of-field begin as soon as melt starts around noon and stops due to refreezing at night
307 (Tiessen et al., 2010b). In-channel storage was calculated as total reach length (calculated in GIS
308 using the drainage network) divided by average flow velocity, which was estimated using
309 measured hydrographs and channel dimensions (i.e. average flow rate in the simulation period
310 divided by cross sectional area). The dimensionless constant that weights inflow and outflow was
311 set to 0.25, which is a common value for natural stream channels (Carter and Godfrey, 1960) .

312 *2.6 Assessment of model simulations*

313 Model simulations were graphically and statistically assessed against streamflow data
314 collected by WSC at station number 05OG008. The statistical metrics used for model assessment
315 in a daily time-step were the Nash-Sutcliffe model efficiency (NSE), model bias (MB), the root
316 mean square difference (RMSD), and the normalized root mean square difference (NMSD)
317 (Fang et al., 2013). Although streamflow simulations were available between 1990 and 2013,
318 some years could not be used for model assessment. Among those are the years of 1990 and
319 1991, which were used as the model warm-up period, and the years having data quality issues as

320 documented in the WSC streamflow database HYDAT (version 10 issued on October 17, 2014).
321 Thus, the years of 1990-1991, 1997-2001, 2004, and 2008 were not used for model assessment.

322 Key aspects of the hydrological cycle influencing streamflow generation in cold regions
323 were also used to assess model performance. Simulated snow water equivalent (SWE) was
324 compared to SWE calculated from the depth of snow on ground measured at the Marquette
325 station using the following relationship (Pomeroy and Gray, 1995):

$$326 \quad SWE = 0.01d_s\rho_s, \quad (1)$$

327 where d_s is the depth of snow (cm), ρ_s is the snow density, which was assumed to be 180 kg m^{-3}
328 based on the typical density range between new and settled snow (Paterson, 1994), and SWE is
329 same as defined above (mm). Simulated yearly cumulative evaporation from open water was
330 compared to gross evaporation values published for Portage La Prairie between 1992 and 2000
331 (Martin, 2002), which is the closest location with available data corresponding to the time period
332 of model simulations. Since both values were based on evaporation models and not measured,
333 the evaporation comparison was made only to verify that CRHM simulations were in agreement
334 with values expected for the study area. Simulated volumetric soil water content [VWC,
335 expressed in mm of water in the soil profile, i.e. $\text{VMC} (\text{m}^3 \text{ m}^{-3}) \times \text{profile depth (mm)}$] was
336 compared to VWC simulated across Canada using the National Drought Model (NDM)
337 (Chipanshi et al., 2013). Similarly to evaporation, no direct measurements of soil moisture were
338 available in close proximity to the watershed; thus, the NDM dataset was used to verify only if
339 the VWC simulated by CRHM was within a reasonable range and followed the expected
340 seasonal trend.

341 *2.7 Sensitivity analysis*

342 In order to assess the suitability of a single storage parameter (K storage) in the
343 Muskingum model for wet and dry years, a sensitivity analysis was carried out for both overland
344 and drainage network HRUs. In the analysis, the K storage parameter was varied to 0, 1, 2, 4, 8,
345 and 15 days for the upland HRUs and to 0, 1, 2, 2.5, 3, 4, 8, 15, and 20 days for the river
346 network. The sensitivity of the model performance to K storage was done through a graphical
347 assessment of NSE and MB with variations of that variable.

348 *2.8 Model falsification*

349 Taking advantage of the flexible structure of the CRHM platform to simulate physically-
350 based hydrological processes, model falsifications were performed to assess the impact of snow
351 sublimation, blowing snow (which includes both snow sublimation and transport), and frozen-
352 soil infiltration on stream discharge in the study area. These processes are typical of the
353 Canadian Prairies and very influential to runoff generation in the region (Pomeroy et al., 2007).
354 A stepwise model falsification was achieved by sequentially removing the following processes
355 from the model: i) snow sublimation, ii) blowing snow, iii) frozen-soil infiltration, and iv)
356 blowing snow and frozen-soil infiltration, which is the combination of cases (ii) and (iii). The
357 impact of model falsification was assessed for water years (October 1st – September 30th of the
358 following year) rather than the calendar year, since a complete time series was available from
359 model simulations. However, statistical metrics used for model assessment (i.e., NSE and MB)
360 were only calculated between March and October, which is the period for which observed
361 streamflow was available.

362 **3. Results**

363 *3.1 Flow characteristics in the study area between 1990 and 2013*

364 As typical of the cold-region conditions prevailing in the Canadian Prairies, no major
365 snowmelt events occur in the La Salle River over the winter and peak stream discharge usually
366 takes place in the spring (Table 3). Thirteen out of 15 years exhibited annual peak discharge with
367 snowmelt. The median peak discharge was $6.7 \text{ m}^3 \text{ s}^{-1}$, while the median annual discharge volume
368 was $1.25 \times 10^7 \text{ m}^3$ and the median water yield was 66 mm. Years with peak discharge above the
369 median corresponded to years with annual discharge volume above the median, due to the strong
370 positive correlation between these two variables ($r^2=0.90$), reflecting that most of the annual
371 discharge occurs during spring and is associated with event runoff rather than baseflow. Those
372 years with peak discharge and annual discharge volume equal to or above median values were
373 considered wet years for model assessment purposes and include the years of 1996, 2005, 2006,
374 2009-2011, and 2013.

375 Seven out of 15 years had ice conditions at time of peak discharge, while two years had
376 peak discharges one day after the end of the ice period (Table 3). Ice conditions at peak
377 discharge were most common in years with cumulative and peak discharges equal to or below
378 the median. Field observations at the site indicate that the absence of ice in wet years was likely
379 due to faster-moving water at high discharge. However, ice moving downstream still may have
380 impacted the flow regime at the gauging station through the formation of ice jams along the river
381 channel and promotion of backwater conditions, which were relatively common (Table 3).

382 Seven out of nine years with both upstream and downstream level monitoring exhibited
383 potential backwater effects (Table 3; Fig. 2). Potential backwater events were generally more
384 associated with years of high discharges. Backwater in this reach could be caused by two factors.

385 The first being that the confluence of the Elm River channel and the La Salle River just
386 downstream of the gauging station 05OG008 (Fig. 1b). Given the gentle topographic gradient of
387 the area and the presence of the water level control structure, the water backs up in the reach of
388 the La Salle River upstream of the confluence of the two channels. The second factor is the
389 occurrence of ice-damming and potential for build-up from the 4-m dam at Starbuck located 28.6
390 km downstream following an elevation change of 3 m.

391 *3.2 CRHM simulations of stream discharge*

392 The majority of annual stream discharge occurs in spring and in particular with snowmelt,
393 making this the most important period for hydrological simulations to identify nutrient export
394 potential (Table 3). However, stream discharge measurements during the spring are also
395 associated with the greatest potential for model and measurement uncertainties related to ice
396 conditions and backwater issues. Given the hydrological importance of stream discharges driven
397 by snowmelt, model assessment was performed without removing questionable spring records
398 from the dataset. The only period removed from analysis was the backwater period in July of
399 2005 that occurred following an extreme rainfall runoff event (Table 3; Fig. 2) with a high
400 degree of known uncertainty regarding measurement accuracy caused by backwater and debris
401 buildup (according to WSC field notes), and representing disinformation for model assessment
402 (Beven, 2011).

403 Graphical comparison between observed and simulated stream discharge indicates that the
404 model skill varied for different years (Fig. 3). Records of discharge measurements performed by
405 WSC during site visits were obtained and are also shown in Fig. 3. Simulated time of peak
406 discharge was generally in good agreement with observed values, except for three out of 15 years
407 (i.e. 1993, 2003, and 2007). These three years had peak discharge and annual discharge volume

408 equal to or below the medians for the 1990-2013 period (i.e. $6.7 \text{ m}^3 \text{ s}^{-1}$ for peak discharge and
409 $1.25 \times 10^7 \text{ m}^3$ for annual discharge volume). The magnitude of peak discharges was also
410 reasonably simulated by the model (Fig. 4), although with more variability than timing of peak.
411 Much of this variation was associated with years of discharge volume below the median (i.e.
412 1994, 1995, 2002, and 2007) or large peak discharge (i.e. 2006, 2009, and 2011). There was no
413 direct relationship between overestimation of simulated peak discharges (Fig. 4) and
414 overestimation of annual discharge volumes (Fig. 5). However, there was a tendency for
415 simulated annual discharge volumes to be overestimated in low-discharge years (Fig. 5) and to
416 underestimate large peak discharges (Fig. 4).

417 Statistical metrics used to assess model performance are shown in Table 4. The results
418 confirm the trends shown in graphical analysis where the model had very good performance for
419 years with peak discharge equal or larger than the median peak discharge. The Nash-Sutcliffe
420 efficiency (NSE) was consistently above 0.65 and averaged 0.76 for 1992, 1996, 2005, 2006,
421 2009, 2011, and 2013. The year 2010 was an exception, with NSE=0.36. For low discharge
422 years, NSE was negative, indicating that the model was no better than utilizing mean discharge
423 as a predictor of stream discharge. For the years with NSE > 0.65, the model bias (MB) was
424 generally negative, although the average model bias was small (i.e. 5%). This was likely due to
425 the overall underestimation of peak discharges during those years. Conversely, MB was
426 generally positive for years with negative NSE, which correspond to years with annual discharge
427 volumes below the 1990-2013 median. This confirms the difficulty in simulating low discharges
428 as was observed in the graphical analysis. The root mean square difference (RMSD) and the
429 normalized root mean square difference (NRMSD) were small to moderate, given the range of
430 discharges comprised between minimum and peak discharges.

431 *3.3 CRHM simulation of SWE, evaporation, and VWC*

432 The annual trend in SWE and time of melt was accurately captured by the model (Fig.6).
433 Differences in peak SWE (e.g. 1992 and 1996) were likely due to inter-annual variation in snow
434 density, which was not measured and can vary substantially with space and time (Pomeroy and
435 Gray, 1995). However, the assumed snow density value did not introduce a substantial bias into
436 SWE estimation, as most years had good agreement between simulated and observed peak SWE
437 values. CRHM was also able to simulate cumulative open water evaporation that compared well
438 with gross evaporation estimates for the years when data was available (Fig. 7). The average
439 cumulative gross evaporation estimated by AAFC at Portage La Prairie was 686 mm, while that
440 simulated by CRHM for open water was 682 mm. The soil moisture variation trend over the year
441 was also well simulated by CRHM in most years (Fig. 8). CRHM was generally able to capture
442 the trend in soil moisture content at the beginning of each growing season and to mimic its
443 depletion and recovery over the course of the summer and fall seasons, respectively. Larger
444 departures from the AAFC data in soil moisture occurred in dry years (e.g. 1995, 2003, and
445 2012). The results above are important since they imply that antecedent conditions were
446 reasonably well predicted, especially in wet years, with accurate representation of watershed
447 hydrological processes and that soil infiltration, soil freezing, and snowmelt runoff patterns can
448 be modelled with greater confidence.

449 *3.4 Sensitivity analysis using the storage parameter in the Muskingum model*

450 A sensitivity analysis was carried out using the storage parameter in the Muskingum model
451 (K storage) for both overland and drainage network flow routing to investigate the adequacy of
452 average values for both wet and dry years (Fig. 9). For wet years, the plot indicates equifinality
453 (Beven, 2011), where different models have similar performance when K storage is in the range

454 between zero and four days for both upland and drainage network (Fig. 9a). The model bias for
455 peak flows aids model selection by defining a narrow band for which the bias is minimized (Fig.
456 9c). However, knowledge of the flow characteristics is still necessary to define the K storage
457 parameter. In this case, K storage of 0 and 3 days for upland and drainage network, respectively,
458 seem reasonable to maximize NSE and minimize peak-flow MB. Conversely, the model
459 performance for dry years tends to improve as K storage increases for both upland and drainage
460 network, although NSE is still in the negative range (Fig. 9b). The contour plot of peak-flow
461 model bias for dry years also suggests large K storage values since low bias is only found for
462 higher K storage values (Fig. 9d).

463 Despite the relative improvement, changing K storage in dry years did not impact model
464 performance appreciably. These results may suggest that the hydrological controls under dry
465 conditions in this watershed are not strongly topological (i.e. drainage network) but typological
466 (i.e. landscape elements) (Buttle, 2006). In order to assess this possibility, the CRHM option to
467 allow infiltration prior to the major melt event was selected in an attempt to emulate the effect of
468 preferential flow. The NSE and MB for dry years were, respectively, -27.87 and 2.0 (1994), -
469 0.58 and 0.81 (1995), -9.6 and 1.7 (2002), -1.09 and 0.16 (2003), -0.43 and 0.69 (2007), and -
470 0.17 and 0.25 (2012). These results indicate a potential influence of infiltration prior to the major
471 melt event (used here as a proxy for preferential flow) on model performance in two out of six
472 years (i.e. 2002 and 2012) when compared to the based model (Table 4). Inclusion of this
473 process in the model structure improved predictions more than 30% of the time, although NSE
474 values were still negative in all cases. However, processes other than infiltration could also be
475 influencing model performance, such as an enhanced role of depressional storage.

476 *3.5 Model falsification*

477 Comparison of the hydrographs between the base model and the falsified models indicates
478 that sublimation, blowing snow, and infiltration to frozen soils are influential processes in the
479 study area (Fig. 10). Turning off snow sublimation and blowing snow (i.e. snow sublimation and
480 transport combined) in the models resulted in increased peak discharges (Fig. 10c-d), while
481 removing frozen-soil infiltration reduced peak discharge with or without blowing snow (Fig.
482 10e-f). The average change in peak discharge due to sublimation inhibition increased 25.4%,
483 while the increase due to blowing snow inhibition was 39.1%; the effect of inhibiting frozen-soil
484 infiltration was the opposite, with reductions around 60% in peak discharge regardless of the
485 blowing snow process (Table 5). Despite the average increase in peak discharge, there were also
486 reductions in peak discharge in particular years when the snow sublimation and blowing snow
487 were removed (i.e. 1993-1994; 1994-1995; 2003-2004; 2006-2007). For frozen-soil infiltration
488 models, removing this process had a very consistent effect on peak discharge with reductions in
489 all but one year (i.e. 2001-2002).

490 One striking effect of the model falsification was that disrupting snow sublimation,
491 blowing snow, and frozen-soil infiltration caused a disconnection between peak discharge and
492 annual discharge, which was a feature of the observed stream discharge (section 3.1). In some
493 years, reductions in peak discharge were accompanied by increases in total discharge (e.g. 2003-
494 2004 for snow sublimation and blowing snow falsifications; 2002-2003 for frozen-soil
495 infiltration models). In other years, the opposite was true (e.g. 2009-2010 for snow sublimation
496 and blowing snow falsifications; no instances for frozen-soil infiltration models; Table 5).

497 Models with snow processes disrupted also presented consistent increases in SWE over the
498 entire simulation period (Table 5). Noteworthy, there was no difference in SWE percent change

499 among the models with snow sublimation and blowing snow falsifications, which indicates no
500 influence of snow transport in SWE. No transport of snow out of the watershed was confirmed
501 by assessing the snow loss variable within CRHM, which indicated total losses smaller than 1
502 mm over the 21 years of simulation (i.e. including the spin-up years) in all cases.

503 The statistical metrics calculated for the falsified models (Table 6) confirmed the loss of
504 model performance when compared to the base model (Table 4), which highlights the
505 importance of snow sublimation, blowing snow, and infiltration to frozen soils processes for
506 accurately simulating stream discharge in the study area. Models with falsified snow processes
507 (i.e. snow sublimation and blowing snow) generally presented a loss of performance, although in
508 specific years, performance remained comparable (e.g. 2010) or improved slightly (e.g. 2005,
509 2006). This improvement was likely due to the increased peak discharge that offset the trend of
510 the model to underestimate large peak discharges in wet years (Fig. 4). Similarly to NSE, model
511 bias tended to become more extreme due to falsification of snow processes.

512 Loss of performance was observed with falsification of frozen-soil infiltration (including
513 blowing snow falsification or not) in relation to the base model and was more severe in wet years
514 due to drastic reduction in peak discharge (Fig. 10). However, performance tended to improve in
515 dry years (e.g. 2005), which was likely due to an offset in the trend of the base model to
516 overestimate discharges in those years. Reduced peak and total discharges due to falsification of
517 frozen-soil infiltration process was due to increased infiltration.

518 **4. Discussion**

519 *4.1 Stream discharge uncertainty*

520 The assessment of the hydrological records of the La Salle River at the gauging station
521 05OG008 suggests that ice and backwater pose additional challenges to simulating streamflow in
522 the study area. Peak flows in the river usually took place under ice-impacted conditions or
523 directly after clearing of ice (Table 3), which may have impacted both the accuracy of discharge
524 measurements at the site (Environment and Climate Change Canada 2015) and hydrograph
525 pattern. Comments from WSC field notes identified a number of potential sources of uncertainty
526 for discharge measurements. For example, backwater conditions are reported in 2005, 2010 and
527 2011, which correspond to the period of backwater identified in this study. Overbank flow into
528 the riparian vegetation is reported for 2007, which also corresponds to the backwater period
529 identified in the present analysis.

530 Backwater during ice conditions may take place because the bottom of the ice cover causes
531 an increase in flow resistance (Gray and Prowse, 1993) and may affect flow magnitudes and
532 timing. Backing up of water from downstream ice jams after breakup may also impact
533 hydrograph shape promoting high peak flow after release and may increase uncertainty
534 associated with discharge measurements (ASCE, 1996). Increasing flows at the confluence of the
535 La Salle River and the Elm River channel, located downstream of 05OG008 may also create
536 backwater conditions where channel capacity is exceeded. According to the WSC field notes,
537 this seems to be the reason for the backwater in 2011, which was caused by the discharge from a
538 controlled breach in the dikes of the Assiniboine River being routed through the Elm River
539 Channel. Overall, backwater conditions seem to be quite frequent in the sub-catchment LS-
540 05OG008 according to the present analysis and based on field observations. Due to changes in

541 flow conditions, discharge measurements are recommended on a regular basis after high flow
542 events or ice break-up; if a shift in the rating curve is necessary to accommodate the new flow
543 conditions, a complete new set of discharge measurements must be made (Coulson, 1991). Thus,
544 the peak flows in the observed records could have been exacerbated due to ice conditions and
545 backwater.

546 *4.2 Effect of stream discharge uncertainty on model performance*

547 Model performance may also be reduced during backwater periods. Flow routing was
548 calculated in CRHM using the Muskingum model, which is not accurate for rapidly raising
549 hydrographs routed through flat sloping rivers and neglects backwater effects due to tributary
550 inflows and presence of dams (Fread, 1993). These features are characteristic of the study area
551 and may have influenced the model simulations. Alternative distributed flow models such as the
552 Kinematic wave model and the Muskingum-Cunge model are also impacted by backwater effect,
553 which limits the selection of alternative methods for flood routing. Dynamic routing models that
554 solve the Saint-Venant equations are not limited by backwater conditions; however, no analytical
555 solution of the complete set of equations is available for most practical applications (Fread,
556 1993). Thus, numerical solutions based on finite-element or finite-difference schemes are usually
557 required. Computational inefficiency, numerical instability, and convergence are among the
558 drawbacks of such solutions, which explain the adoption of simpler routing methods in many
559 hydrological models. This remains a major challenge for the application of most hydrological
560 models to catchments such as the La Salle River watershed.

561 Ice covered and backwater conditions create added uncertainty during assessment of model
562 performance, making evaluation for LS-05OG008 more challenging, but this problem is likely to
563 be characteristic of other rivers in the Red River Valley given the relatively flat topography, the

564 presence of water level control structures/dams along the river channels, presence of ice during
565 peak flows, and frequent backwater at the confluence of rivers and streams. It has been
566 suggested that input errors in precipitation datasets are the dominant source of error in
567 hydrological modelling, while errors in streamflow data are much smaller (Kavetski et al., 2006).
568 While this premise is generally accepted, the results shown in the present study suggest that
569 uncertainty regarding streamflow records in agricultural cold regions can be particularly
570 complex.

571 Despite the uncertainty associated with representation of in-stream dynamics during ice
572 and backwater conditions, CRHM was able to capture the overall trend in streamflow with
573 satisfactory simulation of the timing and magnitude of wet years (when peak flows were above
574 the median). However, even in those years, the peak flows were underestimated by the model.

575 *4.3 Sensitivity analysis of the storage parameter in the Muskingum model*

576 Model performance could also have been affected by different routing conditions taking
577 place during dry and wet years. The contour plots of NSE and peak-flow MB for both wet and
578 dry years suggest that constant K storage values are not adequate to represent both hydrological
579 regimes and that dynamical storage might be present in the basin (Fig. 9). Constant values of the
580 Muskingum parameters make them dependent of the hydrological flow data used to derive them,
581 while it has been suggested that a more physically realistic approach is to allow the parameters to
582 vary in time and space according to flow variability (Guang-Te and Singh, 1992). Due to this
583 limitation, methodologies have been devised for application of the Muskingum model with
584 variable parameters (Guang-Te and Singh, 1992; Song et al., 2011). In the present study, the K
585 storage parameter was adequate to represent wet conditions only, since good model performance
586 was achieved by the model for those conditions. However, K storage alone was not enough to

587 explain the poor performance during dry years since the NSE was consistently negative even for
588 large K storage values. Trends towards large K storage (i.e. slow movement of overland runoff
589 and channel routing) and overestimation of discharge for dry years (Figs. 3 and 5) suggest that
590 other physical processes such as dynamical macropores affecting infiltration through soil cracks
591 could be underestimated during simulations that were designed to address wet condition
592 flowpaths. In-channel processes such as water pooling and natural damming may also explain the
593 large K storage values for drier years.

594 *4.4 Model performance in wet years*

595 This application of CHRM has a very high proportion of the watershed used as cropland
596 (87%), extremely level topography, and modest depressional storage, which is unique and
597 contrasts to other agricultural applications of CRHM in land use proportions and topographic
598 relief (Mahmood et al., 2017; Fang et al., 2010; Pomeroy et al., 2007; Pomeroy et al., 2014).
599 The land-use split method used in other modelling efforts in this watershed (Yang et al., 2014) to
600 represent crop rotations in a static fashion in a multi-year model exercise, seemed to work well
601 for the application in this study given the good simulations in wet years. Simulations of SWE,
602 evaporation, and VWC were validated by external datasets, indicating that the model state
603 variables were in good agreement with the major hydrological processes in the agricultural
604 landscape. Small yearly differences in evaporation were likely due to differences in input data,
605 location, and method, where AAFC applies Meyer's revised formula (Martin, 2002) for lakes
606 and CRHM uses the Priestley-Taylor procedure for wetlands and small lakes. Differences in soil
607 moisture values could be due to differences in soil properties used to calculate water holding
608 capacity of the soil. Good simulation of model state variables increases the confidence in the
609 runoff simulations and highlights the value of the physically-based approach used by CRHM to

610 represent agricultural landscapes. Proper simulation of these variables reinforces that the
611 physical description of the relevant hydrological processes are accurate for cold regions,
612 underpinning the use of models that do not require calibration, such as the one developed in the
613 present study for use in cold-climate agricultural areas. Proper simulation of state variables also
614 enables the model to assess other scenarios not simulated here. For example, changes in land use
615 could be simulated by altering model parameters.

616 *4.5 Model performance in dry years*

617 The model did not show satisfactory performance in dry years. Poor model performance in
618 dry years is not completely unexpected since low flows are generally more difficult to predict
619 than larger flows (Nicolle et al., 2014), particularly for small catchment areas (Stanfield et al.,
620 2009). This difficulty may also have been aggravated if there was flow under or over ice. The
621 general pattern of flow is represented during drier years and overestimation appears not to be
622 caused by overestimation of baseflow; rather, problems are most evident as overestimation of
623 peak flows (e.g. years 1994 and 2002 in Fig. 3). While drought conditions and in-stream flow
624 requirements are important considerations on the Prairies from a land use and water management
625 perspective (Fang and Pomeroy, 2007), the smaller magnitude events account for very little of
626 the overall export of water and phosphorus from the La Salle (Corriveau et al., 2013) and CRHM
627 appears to be a promising tool for physically-based simulation of the impact of management
628 change on watershed hydrology in intensively agricultural tributaries of the Red River where
629 elevated nutrient export is of particular concern.

630 Emulation of preferential flow by allowing infiltration prior to the major melt event
631 resulted in improved simulations more than 30% of the time in dry years, which suggests that
632 this mechanism as a hydrological control influencing stream discharge under dry conditions.

633 While the method chosen does not capture all the complex nuances of preferential flow, it does
634 allow for enhanced infiltration and mimics the same effects of this mechanism. Interestingly, the
635 two years that resulted in improved simulations (i.e. 2002 and 2012), were dry years preceded by
636 years with wet springs and dry summers. In 2001, well-above-normal precipitation in the spring
637 was followed by dry summer and fall (Wheaton et al., 2008). The same was observed in 2011
638 (Cordeiro et al., 2014). This similar trend highlights the importance of antecedent conditions on
639 preferential flow. In contrast, there was no improvement in dry years followed by dry years (i.e.
640 1995 and 2003). It is likely that the simulation of infiltration prior to melt was insufficient to
641 account for the magnitude of preferential flow upon prolonged desiccation of the cracking soils
642 in the study area. Granger et al. (1984) showed that dry cracking clays created conditions of
643 “unlimited” infiltration into frozen soils – such high infiltration rates would provide model
644 outputs more similar to streamflow observations. Regardless, the physically-based nature of the
645 simulations suggests that preferential flow deserves investigation as one of the major
646 hydrological controls driving stream discharge in dry years.

647 Preferential flow through soil cracking is inherently linked to soil moisture content, which
648 itself can also affect streamflow in dry years. A sensitivity analysis of prairie snowmelt to
649 drought in Creighton catchment in Saskatchewan using CRHM indicated that large reductions in
650 stream discharge could be driven by reduced winter precipitation, increased winter air
651 temperature, and decreased soil moisture content in the fall (Fang and Pomeroy, 2007). While
652 the area of the Creighton catchment corresponds to less than 10% (i.e. 11.4 km²) of that of the
653 study area in the present study (i.e. 189 km²), it had soils in the low-permeability range (i.e. silt
654 clay and clay loams) and 85% of the land use under cultivated land (stubble or fallow fields),
655 which makes it similar to the characteristics of the La Salle watershed. Simulations of soil

656 moisture (Fig.8) showed a trend to overestimation in dry years (e.g. 1995, 2003, and 2012),
657 which again highlights the importance of antecedent conditions to stream discharge generation
658 and the need to improve the model representation during droughts. In fact, investigations in the
659 Canadian Prairies, although in an internally-drained basin, indicated that snowmelt infiltration is
660 very sensitive to soil moisture in the fall and that hydrologic droughts emerged from low soil
661 moisture conditions (Fang and Pomeroy, 2008). Together these studies indicate the need for
662 further research and improved representation of those processes controlling infiltration and
663 routing of runoff in cold regions with dry antecedent conditions.

664 *4.6 Model falsifications*

665 The analysis of model falsification indicates that snow sublimation and blowing snow, as
666 well as infiltration to frozen soils, are crucial for accurate simulations of stream discharge in the
667 flat, intensively-managed agricultural watershed of the La Salle River. The prominence of snow
668 sublimation effects over snow transport contrasts with results observed for mountain
669 environments where greater importance of snow transport has been identified (Zhou et al., 2014).
670 However, snow transport and redistribution within this watershed was still of importance, despite
671 no loss of snow from the watershed. Inclusion of snow transport affected peak and total
672 discharge, with a more pronounced effect on the latter (Table 5). Inhibition of blowing snow
673 sublimation caused a reduction in total discharge in two out of 21 years (9% of the time) when
674 compared to the base model, while inhibition of both blowing snow sublimation and transport
675 caused a reduction of this variable in eight out of 21 years (38% of the time). Accumulation of
676 snow within watersheds tends to occur in association with steep hills and valley slopes (Pomeroy
677 and Goodison, 1997). In the lower slope landscape of the La Salle watershed, snow accumulates
678 in deep drainage ditches and stream channels. If snow transport is inhibited, the stimulated

679 accumulation of snow in these topographic features decreases and greater snow melt is simulated
680 for upland areas. This increases potential for infiltration, which could explain the reduction in
681 total discharge associated with inhibition of snow transport. This pattern stresses the importance
682 of internal snow redistribution within the watershed to stream discharge generation despite snow
683 transport not impacting peak discharge to the same extent as blowing snow sublimation. The
684 pattern also provides some insight into the potential for wind barrier sites such as shelterbelts to
685 retain snow in upland areas (Steyn et al., 1997) to encourage infiltration and reduce runoff
686 generation with melt.

687 Regarding infiltration into frozen soils, falsifying this process had an overriding effect over
688 falsification of blowing snow since peak discharge was consistently reduced in these models,
689 despite the trend of increased peak and total discharge arising from falsification of snow
690 processes (Table 5). These results emphasize the importance of infiltration to stream discharge
691 generation. This process has been indirectly discussed by Gray et al. (1986) and van der Kamp et
692 al. (2003), who observed reduced runoff for land uses with enhanced infiltration. Since runoff is
693 the primary source of stream discharge in the Canadian Prairies (Shook and Pomeroy, 2012;
694 Shook and Pomeroy, 2010), the importance of frozen soil infiltration could be more pronounced
695 than that of the blowing snow process to streamflow generation. The model falsification
696 indicates that peak discharges would be reduced, on average, by 61% where infiltration into
697 frozen-soil is modelled at non-frozen rates, while inhibition of snow processes would increase
698 peak discharges by 39% on average.

699 **5. Summary and conclusions**

700 Simulation of streamflow in an intensively managed agricultural tributary of the Red River
701 (the La Salle River) with the CRHM platform revealed a number of topics for which knowledge

702 of hydrological processes in the region and model representation might be improved. 1) Ice and
703 backwater issues are likely to contribute to increased uncertainty in both measured hydrometric
704 data and model representations for tributaries of the Red River. Among the drivers of these
705 issues are the low relief of topography of the region, the presence of water level control
706 structures/dams along the river channel, presence of ice during peak flows, and frequent
707 confluences of streams and artificial channels. Future modelling efforts in the region should
708 focus on estimating the model uncertainty arising from ice and dam effects on hydrometric data.
709 2) Simulation of low flow years remains challenging in the La Salle, as is commonly reported for
710 other hydrological models. In low flow years, discharge was overestimated by 90% and a
711 sensitivity analysis of the storage parameter of the Muskingum routing model indicated that
712 averaging this parameter is not adequate for the study area. Also, improved simulations in dry
713 years through emulation of preferential flow by allowing infiltration prior to the major melt event
714 suggest that even where frozen soil predominates, preferential flow may be an important
715 hydrological feature under dry conditions for the high clay content soils of the Red River Valley.
716 As such, dynamic representation of processes such as infiltration through macropores may
717 require revision for drier conditions in the Red River Valley. Research efforts using modelling
718 frameworks should try to implement simulation of preferential flow pathways, especially under
719 dry conditions.

720 Despite the potential to improve model representation identified through this research, it is
721 also evident that performance metrics for the CRHM platform indicate very good simulation of
722 peak and annual cumulative flows in the La Salle River where flows were equal to or above
723 median values (under normal to wet conditions). On average, annual discharge was
724 underestimated by only 5% in wet years. The good performance of the model in average or

725 above-average flow years indicates that CRHM simulations are likely appropriate for use in
726 regional nutrient-transport assessments where export is largely determined by hydrological
727 drivers in the study area and for assessment of land-use and climate change impacts on
728 streamflow.

729 Falsifications of blowing snow transport and sublimation, and infiltration to frozen soils
730 processes in the validated base model indicate that these processes were very influential to
731 stream discharge generation. Inhibition of snow sublimation would represent an average increase
732 in peak discharge around 25%, while inhibition of blowing snow, which includes both snow
733 sublimation and transport, would cause an increase in peak discharge around 39%. Simulation of
734 infiltration without changes to model structure to account for frozen-soils would cause a
735 reduction in peak discharge around 61%.

736 **7. Team list**

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747 the original authors are given credit and the appropriate citation details are mentioned.

748 **9. Code availability**

749 CRHM codes are available through model developers. Details can be found at
750 <http://www.usask.ca/hydrology/CRHM.php>.

751 **10. Data availability**

752 The weather and hydrometric datasets used in this research are publicly accessible through
753 the Government of Canada's Open Data portal (<http://open.canada.ca>) and Environment and
754 Climate Change Canada websites:

755 Weather Data: Environment and Climate Change Canada. 2015. Historical Climate Data.
756 Available at: <http://climate.weather.gc.ca/>. Access: February 22, 2016.

757 Hydrometric data: Environment and Climate Change Canada. 2015. HYDAT Database-
758 National Water Data Archive. Available at: [https://ec.gc.ca/rhc-](https://ec.gc.ca/rhc-wsc/default.asp?lang=En&n=9018B5EC-1)
759 [wsc/default.asp?lang=En&n=9018B5EC-1](https://ec.gc.ca/rhc-wsc/default.asp?lang=En&n=9018B5EC-1). Access: February 22, 2016.

760 **11. Appendices**

761 The manuscript has no appendices

762 **12. Supplement link**

763 **Plot data to be uploaded and link to be generated.**

764 **13. Author contribution**

765 M.R.C. Cordeiro, H.F. Wilson, and J. Vanrobaeys conceived the modelling objectives,
766 scope, and strategy; M.R.C. Cordeiro and J. Vanrobaeys acquired the input data; M.R.C.

767 Cordeiro, J.P. Pomeroy, and X. Fang developed the custom model for analysis; M.R.C. Cordeiro,
768 H.F. Wilson performed data analysis; M.R.C. prepared manuscript with contribution from all co-
769 authors.

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776 **15. Disclaimer**

777 **aaa**

778 **16. References**

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Table 1. List of hydrologic response units (HRUs) used in the LS-05OG008 watershed.

HRU ID	HRU acronym [†]	Cropping system/Land use	Crop	Soil texture	Occurrence			
					SB 2	SB 3	SB 5	SB 33
1	IVPO/SICL	Irrigated Vegetable	Potato	Silty clay loam	Yes	Yes	No	Yes
2	IVPO/C	Irrigated Vegetable	Potato	Clay	No	Yes	No	No
3	IVCR/SICL	Irrigated Vegetable	Carrot	Silty clay loam	Yes	Yes	No	Yes
4	IVCR/C	Irrigated Vegetable	Carrot	Clay	No	Yes	No	No
5	PRSB/SIC	Pulse Non-Row	Soybean	Silty clay	No	Yes	No	No
6	PRSB/C	Pulse Non-Row	Soybean	Clay	No	Yes	No	No
7	PFAF/C	Perennial Forage	Alfalfa	Clay	Yes	Yes	No	Yes
8	PRSW/SIC	Pulse Non-Row	Spring wheat	Silty clay	No	Yes	No	No
9	PRSW/C	Pulse Non-Row	Spring wheat	Clay	No	Yes	No	No
10	PFSW/C	Perennial Forage	Spring wheat	Clay	Yes	Yes	No	Yes
11	OSSW/C	Oilseed - Spring Cereal	Spring wheat	Clay	Yes	Yes	Yes	Yes
12	FCSW/C	Fall Cereal	Spring wheat	Clay	Yes	Yes	Yes	No
13	FCWW/C	Fall Cereal	Winter wheat	Clay	Yes	Yes	Yes	No
14	PRCA/SIC	Pulse Non-Row	Canola	Silty clay	No	Yes	No	No
15	PRCA/C	Pulse Non-Row	Canola	Clay	No	Yes	No	No
16	OSCA/C	Oilseed - Spring Cereal	Canola	Clay	Yes	Yes	Yes	Yes
17	FCCA/C	Fall Cereal	Canola	Clay	Yes	Yes	Yes	No
18	FYDL/SICL	Feedlot	–	Silty clay loam	No	No	No	Yes
19	FYDL/C	Feedlot	–	Clay	Yes	No	No	No
20	URLD/SICL	Urban (low density)	–	Silty clay loam	Yes	Yes	No	Yes
21	URLD/C	Urban (low density)	–	Clay	Yes	Yes	Yes	Yes
22	URMD/SIL	Urban (medium density)	–	Silt loam	No	No	Yes	No
23	URMD/SICL	Urban (medium density)	–	Silty clay loam	Yes	No	No	No
24	URMD/SIC	Urban (medium density)	–	Silty clay	No	Yes	No	No
25	URMD/C	Urban (medium density)	–	Clay	Yes	Yes	Yes	No
26	WETL/WA	Wetland/water	–	–	Yes	No	No	No
27	River Channel	River	–	–	Yes	Yes	Yes	Yes

[†] First two letters indicate cropping system/land use; third and fourth letters indicate crop; letter(s) after the slash indicate soil texture.

Table 2. Modules used in the customized CRHM model to simulate hydrological process in the LS-05OG008 watershed.

Sequence [†]	Module	Description
1	Basin	Holds commonly used physical and control parameters
2	Solar radiation	Calculates theoretical global radiation, direct and diffuse solar radiation, as well as maximum sunshine hours based on latitude, elevation, ground slope, and azimuth (Garnier and Ohmura, 1970). This module provides radiation input to modules 3, 8, and 13
3	Observation	Adjusts the temperature and precipitation variables in the input weather data to variations in environmental lapse rate, elevation, and wind-induced under-catch
4	Crop growth	Simulates a linear crop development over the growing season
5	Sunshine hours	Estimates sunshine hours from incoming shortwave radiation and provide input to modules 8 and 13
6	Walmsley's windflow	A parametric version of the wind flow model (Walmsley et al., 1989) that adjusts the wind speed change due to local topographic features and provides input to module 11
7	Long-wave radiation	Estimates incoming long-wave radiation from the air temperature and the atmospheric transmittance, which is estimated from measured short-wave radiation and theoretical global radiation and provides input to module 13 (Sicart et al., 2006)
8	Summer net radiation	Estimates the snow-free net all-wave radiation from the calculated short-wave radiation by Garnier and Ohmura (1970) and the calculated net long-wave radiation (Brunt, 1932) using sunshine hours, air temperature and humidity (Granger and Gray, 1990) and provides inputs to module 9
9	Evaporation	Estimates actual evapotranspiration from unsaturated surfaces (Monteith, 1965) and evaporation from saturated surfaces such as stream channels (Priestley and Taylor, 1972). These algorithms modify the moisture content in the interception, ponded surface water, and soil column stores, as well as in the stream channel
10	Canopy	Estimates the snowfall and rainfall intercepted by the forest canopy and updates the under-canopy snowfall and rainfall and calculates short-wave and long-wave sub-canopy radiation (Ellis et al., 2010) with options for open environment (no canopy adjustment of snow mass and energy) and forest environment (adjustment of snow mass and energy from forest canopy)
11	Prairie Blowing Snow Model (PBSM)	Simulates snow sublimation and transport between HRUs (Pomeroy and Li, 2000)
12	Albedo	Estimates the snow albedo during the winter and the melt period (Gray and Landine, 1987). This module indicates the beginning of melt for module 13
13	Energy-Budget Snowmelt Model (EBSM)	Estimates snowmelt for snowpack in the Canadian Prairies (Gray and Landine, 1988) by calculating the energy balance of radiation, sensible heat, latent heat, ground heat, advection from rainfall, and change in internal energy
14	Infiltration	Estimates snowmelt infiltration into frozen soils (Gray et al., 1985) and rainfall into unfrozen soils based on texture and ground cover (Ayers, 1959)
15	Hydraulic conductivity estimator	Darcy's law for unsaturated flow is used to calculate the drainage factors of module 16 utilizing the unsaturated hydraulic conductivity calculated using Brooks and Corey relationship (Fang et al., 2013)
16	Soil	Estimates the soil moisture, groundwater flow, and interactions between ground- and surface-water throughout the

year (Leavesley et al., 1983; Dornes et al., 2008; Fang et al., 2010)

17 Volumetric soil moisture Converts soil moisture to volumetric equivalent using the variables from module 16 and determines fall status for module 14

18 Muskingum routing Routes runoff between HRU and sub-basins using the Muskingum method (Chow, 1964)

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† Sequence in which the modules were entered into each CRHM Group using the Macro window.

Table 3. Streamflow characteristics of the study area for the years used in the analysis.

Year	Water Yield (mm)	Discharge					Ice conditions	Backwater
		Annual [†] (m ³)	Snowmelt (m ³)	Snowmelt Proportion (%)	Peak (m ³ s ⁻¹)	Peak Date		
1992	64	1.21×10^7	7.27×10^6	60	6.7	April 10 th	March 1 st – April 12 th	No data for analysis
1993	66	1.25×10^7	4.51×10^6	36	5.6	April 07 th	March 1 st – April 10 th	No data for analysis
1994	18	3.33×10^6	9.32×10^5	28	0.7	April 10 th	March 1 st – April 16 th	No data for analysis
1995	61	1.15×10^7	6.56×10^6	57	5.0	March 31 st	March 1 st – April 15 th	No data for analysis
1996	99	1.87×10^7	1.35×10^7	72	13.5	April 29 th	March 1 st – April 28 th	No data for analysis
2002	10	1.94×10^6	1.94×10^6	100	1.6	April 16 th	March 1 st – April 16 th	No backwater
2003	18	3.49×10^6	3.49×10^6	100	2.1	April 2 nd	March 1 st – April 09 th	No backwater
2005 [‡]	83	1.57×10^7	4.54×10^6	29	10.0	April 08 th	March 1 st – April 04 th	April 3 rd – April 14 th June 29 th – July 24 th
2006	115	2.18×10^7	2.18×10^7	100	16.5	April 10 th	March 1 st – April 07 th	April 3 rd – April 19 th
2007	38	7.26×10^6	7.26×10^6	100	4.6	April 12 th	March 1 st – April 05 th	March 31 st – April 13 th
2009	89	1.69×10^7	1.69×10^7	100	13.3	April 17 th	March 1 st – April 16 th	April 11 th – April 26 th
2010	110	2.09×10^7	1.04×10^7	50	10.7	June 1 st	March 1 st – April 05 th	May 30 th – June 06 th September 2 nd – September 3 rd
2011	150	2.84×10^7	1.59×10^7	56	15.7	April 13 th	March 1 st – April 13 th	April 04 th – April 20 th May 20 th – May 21 st
2012	26	4.89×10^6	1.71×10^6	35	2.5	May 29 th	March 1 st – March 24 th	No backwater
2013	76	1.44×10^7	9.38×10^6	65	9.4	May 04 th	March 1 st – May 1 st	April 28 th – May 08 th

[‡] Backwater issues late in the summer; [†] Total flow from March 1st to October 31st.

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Table 4. Metrics used for model assessment.

Year	NSE	MB	RMSD	NRMSD
1992	0.69	-0.10	0.51	0.89
1993	-0.79	0.31	1.13	1.92
1994	-27.91	2.00	0.68	4.31
1995	-0.58	0.81	1.10	2.02
1996	0.81	0.50	1.00	1.14
2002	-14.04	2.01	0.96	3.92
2003	-1.08	0.16	0.69	1.58
2005	0.85	-0.12	0.93	0.77
2006	0.76	-0.38	2.31	0.84
2007	-0.43	0.69	1.65	1.81
2009	0.65	-0.31	2.18	1.02
2010	0.36	-0.26	1.50	1.52
2011	0.84	-0.29	0.98	0.73
2012	-0.60	0.30	0.27	1.16
2013	0.72	0.36	0.91	1.34

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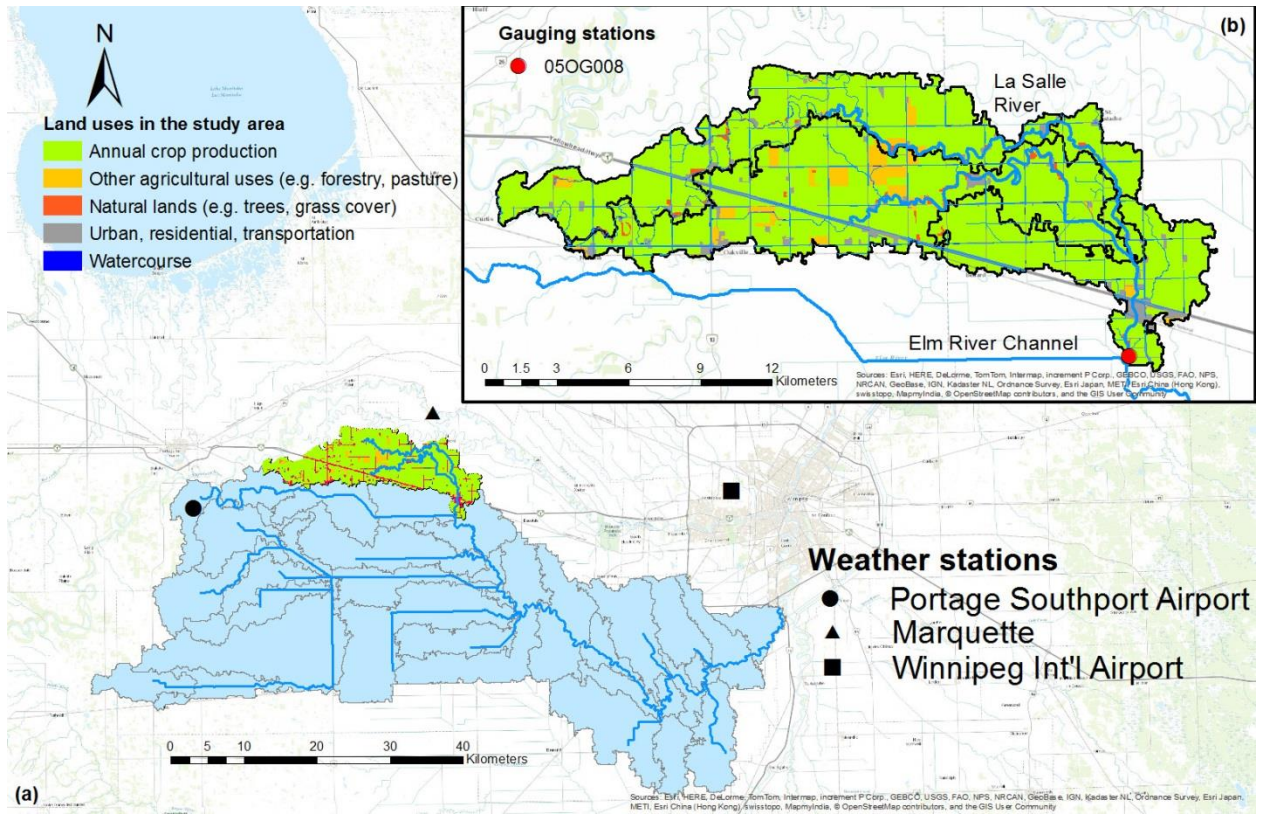
Table 5. Percent change in peak discharge, total discharge, and snow water equivalent (SWE) between falsified models and base models.

Water Year	Snow sublimation			Blowing snow			Frozen-soil infiltration			Frozen-soil infiltration and blowing snow		
	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE	Peak discharge	Total discharge	SWE
1992-1993	39.4	12.3	23.0	69.5	7.1	23.0	-65.7	-2.8	0.0	-65.0	-5.3	23.0
1993-1994	-3.8	2.1	24.8	-2.0	-2.4	24.8	-34.6	18.4	0.0	-34.4	24.4	24.8
1994-1995	-11.3	-1.9	41.8	-13.1	-1.5	41.8	-77.9	-24.3	0.0	-78.1	-25.3	41.8
1995-1996	24.3	8.7	39.5	47.4	5.3	39.5	-78.5	-29.2	0.0	-82.1	-38.0	39.5
1996-1997	180.1	45.7	39.7	350.5	47.3	39.7	-32.7	8.4	0.0	-43.1	-12.5	39.7
1997-1998	29.5	11.4	19.9	33.6	-1.3	19.9	-75.5	5.4	0.0	-75.4	5.3	19.9
1998-1999	7.3	4.9	62.8	3.1	-2.4	62.8	-42.4	43.4	0.0	-41.8	46.0	62.8
1999-2000	14.3	4.8	30.4	14.6	1.6	30.4	-41.4	8.1	0.0	-41.3	7.2	30.4
2000-2001	19.0	17.0	13.9	16.0	18.4	13.9	-79.7	-12.8	0.0	-79.7	-17.9	13.9
2001-2002	25.1	10.3	22.0	30.5	3.4	22.0	21.9	13.7	0.0	24.0	12.6	22.0
2002-2003	17.0	6.1	151.3	23.8	-1.9	151.3	-72.6	42.7	0.0	-72.2	44.1	151.3
2003-2004	-8.2	9.2	17.3	-1.3	16.3	17.3	-80.1	-21.3	0.0	-79.8	-26.0	17.3
2004-2005	30.4	9.1	19.4	27.5	2.2	19.4	-60.0	-7.1	0.0	-60.0	-12.4	19.4
2005-2006	16.2	8.9	22.3	10.5	1.0	22.3	-89.8	-27.0	0.0	-92.4	-30.8	22.3
2006-2007	-4.0	18.9	28.7	-4.0	9.6	28.7	-84.7	-11.1	0.0	-84.5	-15.0	28.7
2007-2008	99.3	15.4	53.0	89.3	-4.5	53.0	-52.0	14.5	0.0	-62.3	10.0	53.0
2008-2009	36.5	10.7	16.2	49.1	5.2	16.2	-80.4	-4.1	0.0	-76.2	-2.7	16.2
2009-2010	1.7	-0.6	23.6	22.3	-5.8	23.6	-36.1	6.6	0.0	-37.0	5.8	23.6
2010-2011	2.9	6.5	19.6	21.8	5.6	19.6	-80.7	-18.8	0.0	-83.8	-24.7	19.6
2011-2012	16.1	5.0	54.2	4.0	-2.2	54.2	-53.8	36.9	0.0	-53.2	38.9	54.2
2012-2013	2.5	2.3	13.1	28.0	5.6	13.1	-79.5	-29.4	0.0	-82.8	-34.3	13.1
Average	25.4	9.8	35.1	39.1	5.1	35.1	-60.8	0.5	0.0	-62.0	-2.4	35.1

Table 6. Selected statistical metrics (i.e. NSE and Model Bias) for falsified models.

Year	Snow sublimation		Blowing snow		Frozen-soil infiltration		Frozen-soil infiltration and blowing snow	
	NSE	MB	NSE	MB	NSE	MB	NSE	MB
1992	-0.43	0.17	-0.62	0.10	-0.21	-10.80	-0.15	-0.10
1993	-0.80	0.47	-0.59	0.40	-0.02	13.60	0.02	0.09
1994	-29.18	2.09	-28.89	1.91	-11.43	216.00	-11.54	2.22
1995	-0.23	0.77	-0.59	0.79	0.31	26.30	0.34	0.23
1996	0.48	0.64	0.12	0.59	0.22	-6.90	0.22	-0.21
2002	-39.64	2.80	-41.84	2.46	-1.37	85.00	-0.73	0.67
2003	-1.52	0.30	-1.74	0.13	-0.14	9.80	-0.01	0.05
2005	0.89	0.03	0.90	-0.08	-0.05	-40.90	-0.05	-0.50
2006	0.83	-0.29	0.81	-0.35	-0.18	-77.20	-0.19	-0.81
2007	-0.86	1.13	-0.88	0.96	-0.10	-30.10	0.00	-0.44
2009	0.68	-0.16	0.66	-0.16	-0.07	-64.50	0.02	-0.63
2010	0.35	-0.26	0.33	-0.30	0.30	-28.80	0.31	-0.31
2011	0.89	-0.23	0.86	-0.24	0.10	-47.30	0.10	-0.53
2012	-1.16	0.38	-0.64	0.26	-0.14	63.40	-0.29	0.68
2013	0.67	0.39	0.09	0.44	0.01	-12.00	0.05	-0.19
Average	-4.60	0.55	-4.80	0.46	-0.85	6.37	-0.79	0.02

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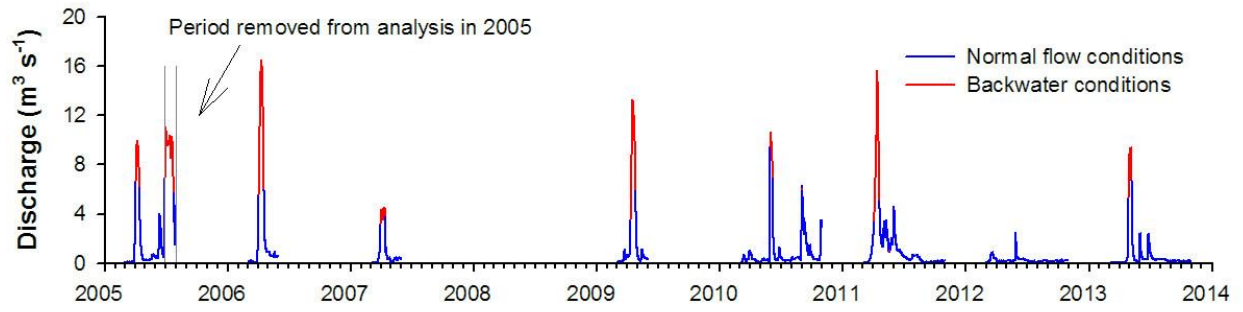
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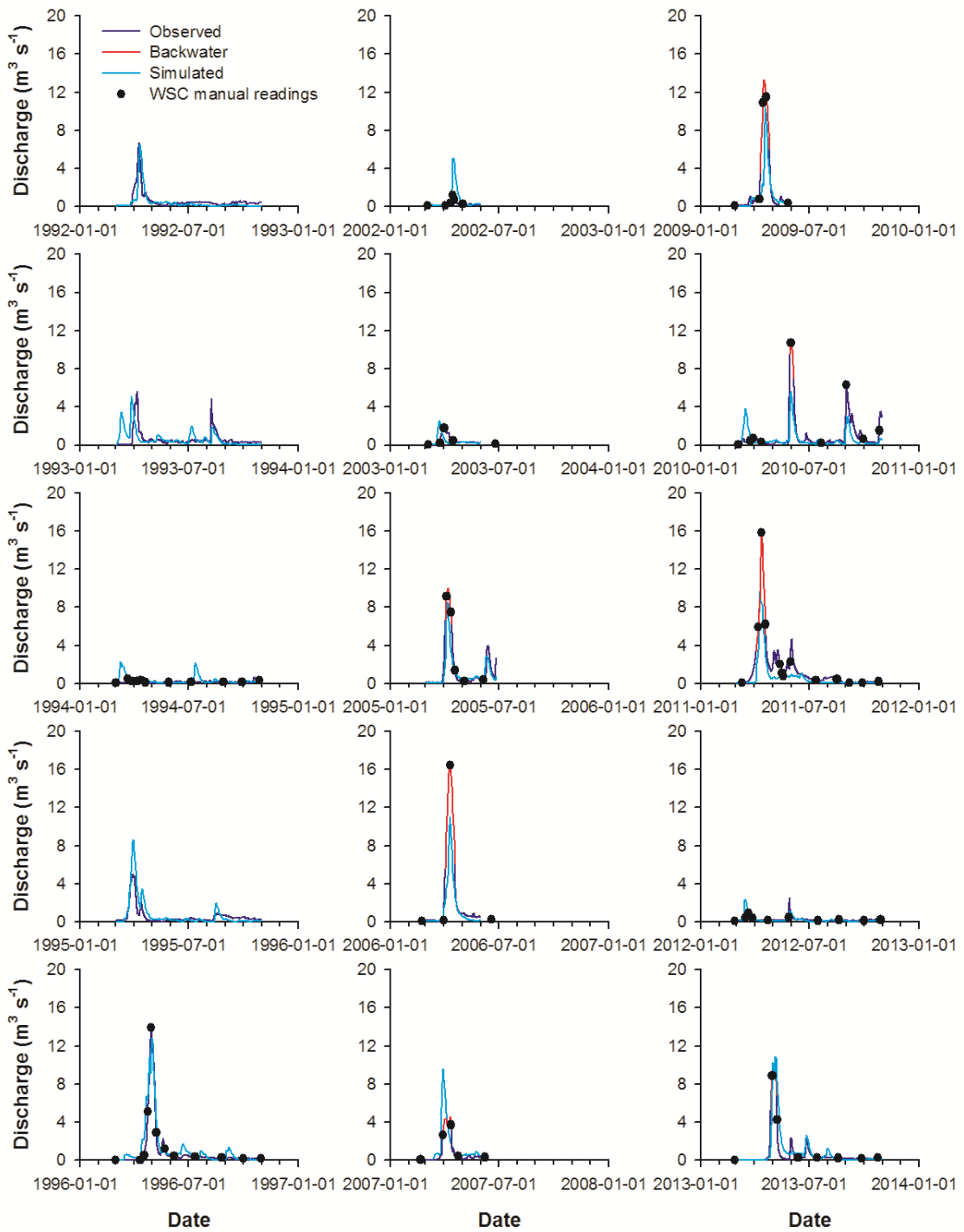
Figure 1. The La Salle watershed (a) and the sub-watershed used in the study, which drains into Water Survey Canada gauging station 05OG008 (b).



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1077 Figure 2. Yearly hydrographs indicating periods of potential backwater issues. Years of 2002, and 2003 not
 1078 presented since these years did not have backwater issues. Years of 2004 and 2008 were not included in the analysis
 1079 since quality issues were identified in the metadata of the records. The period between lines in 2005 indicate the
 1080 records removed from the dataset during model assessment.

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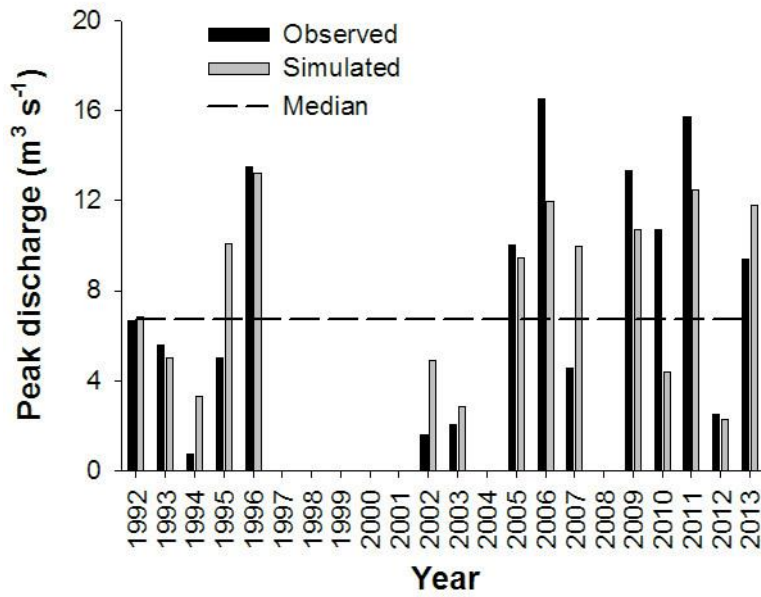
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Figure 3. Comparison of observed and simulated stream discharge between 1992 and 2013 for years with good records in the HYDAT database.

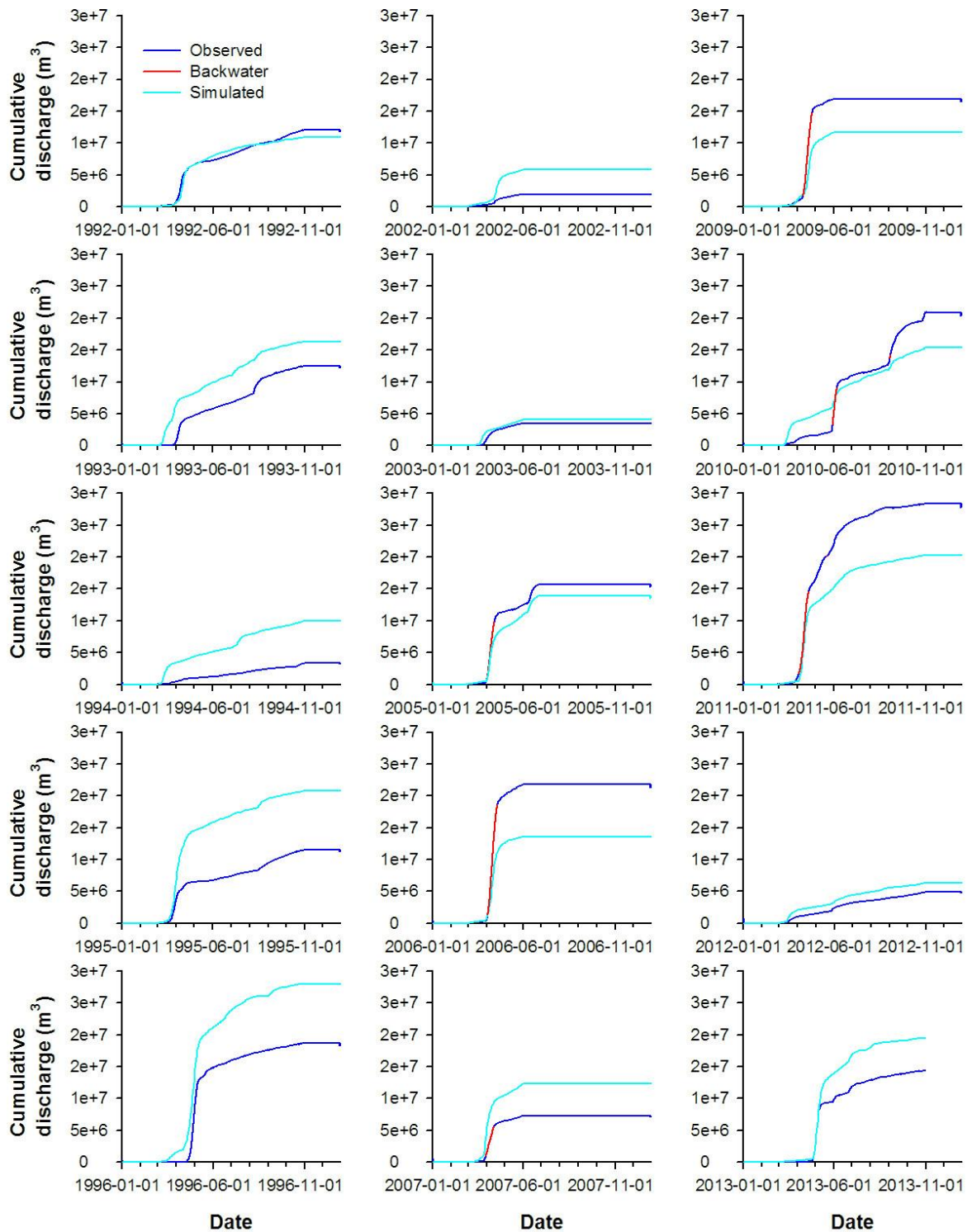
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1087 Figure 4. Comparison of observed and simulated peak discharge between 1992 and 2013 for years with good
 1088 records in the HYDAT database. No data available in the HYDAT database between 1997 and 2001. Years of 2004
 1089 and 2008 were not included in the analysis since quality issues were identified in the metadata of the records.

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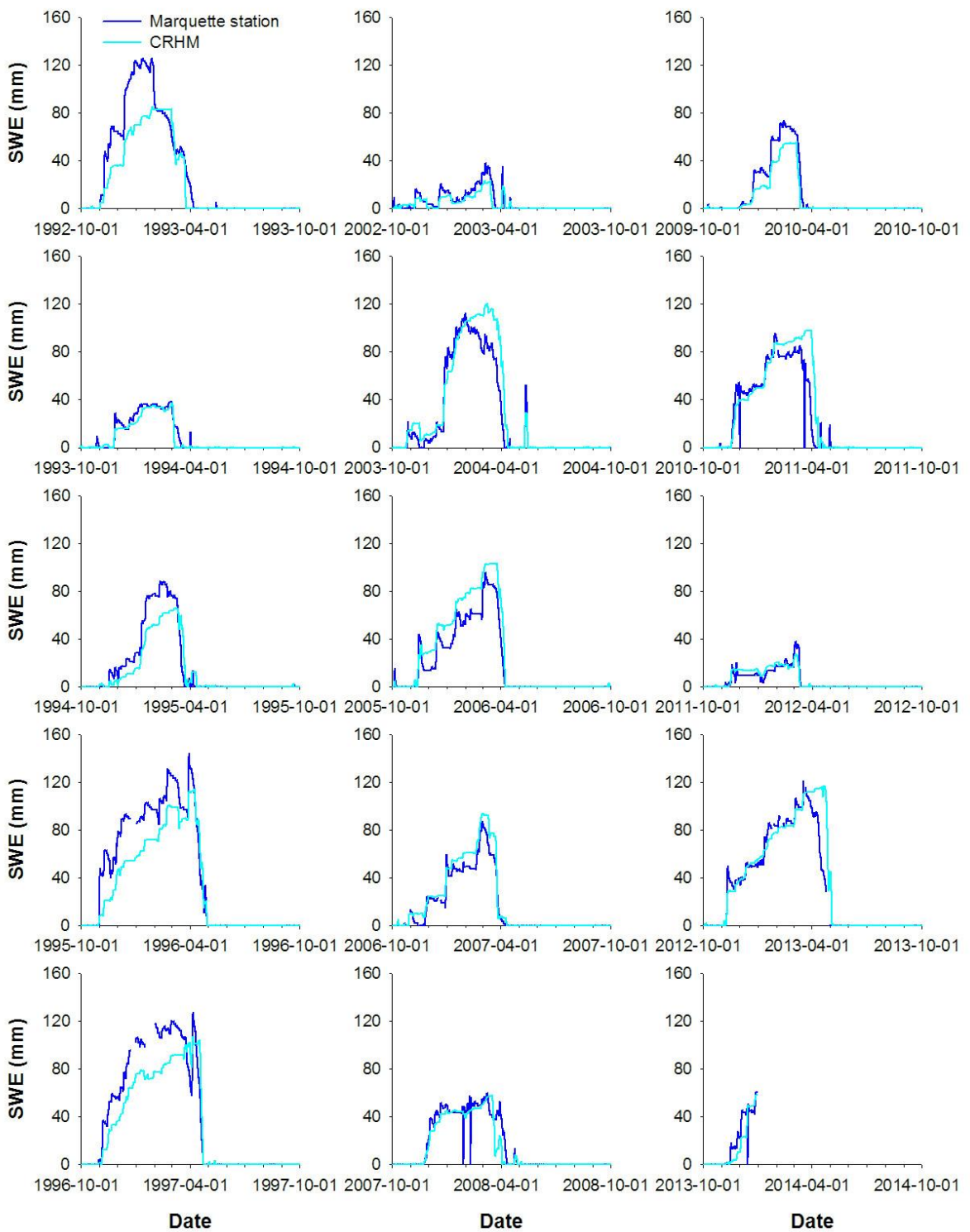
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Figure 5. Comparison of observed and simulated annual cumulative discharge between 1992 and 2013 for years with good records in the HYDAT database.

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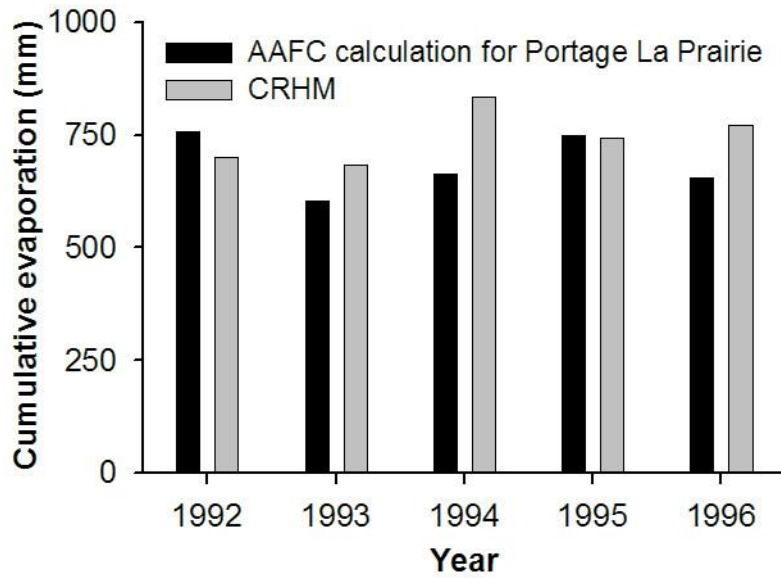
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Figure 6. Comparison of observed and simulated snow water equivalent (SWE) between 1992 and 2013 for years with good records in the HYDAT database. SWE was calculated assuming a snow density of 180 kg m^{-3} .

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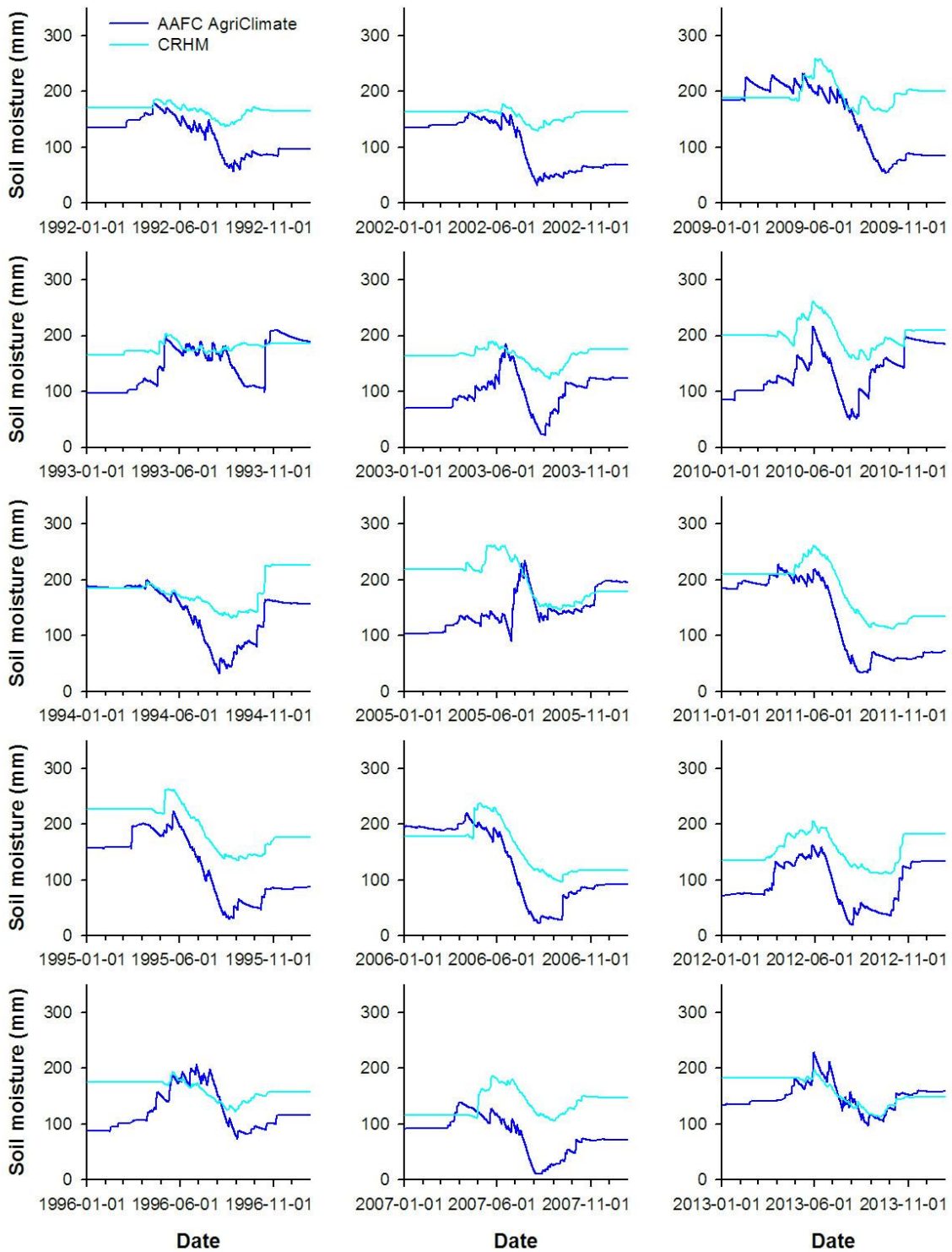
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1100 Figure 7. Comparison of evaporation estimates produced by the Agriculture and Agri-Food Canada (AAFC)
 1101 and CRHM between 1992 and 1996.

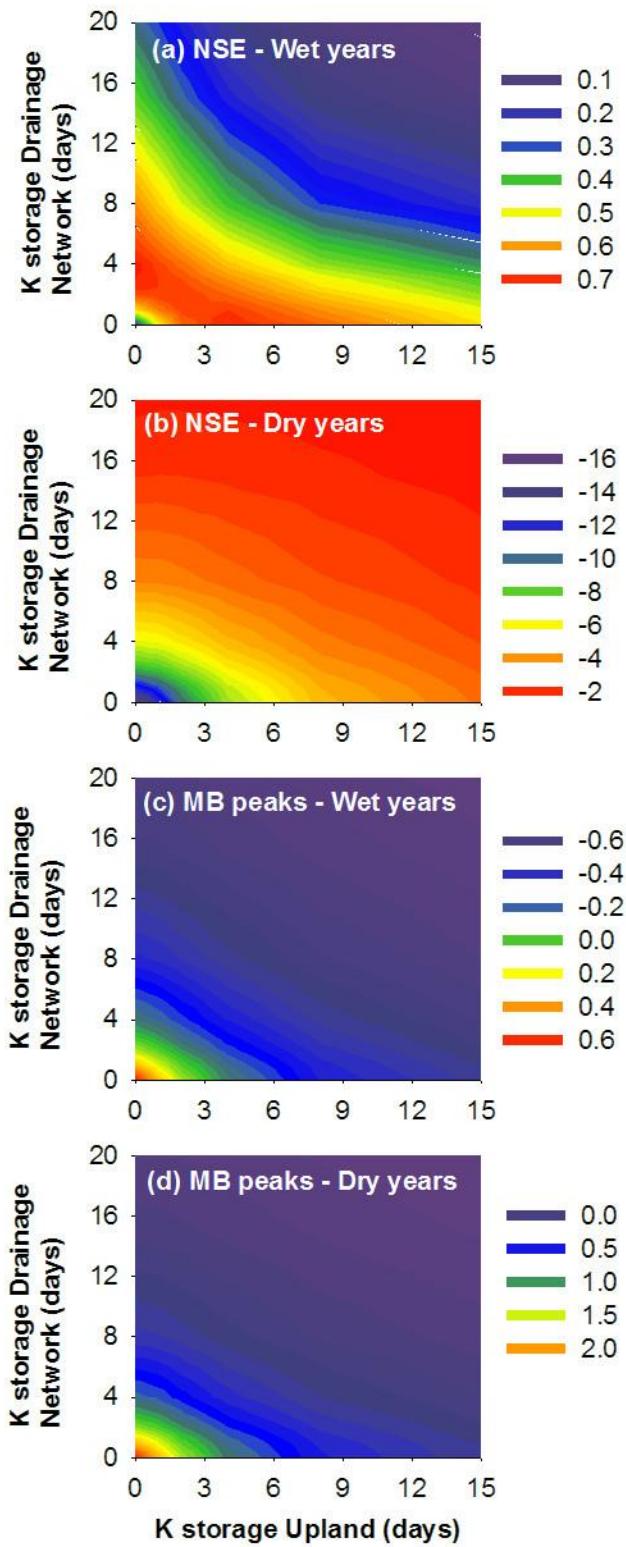
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1104 Figure 8. Comparison between simulated volumetric soil water content (VWC; expressed in mm of water in
 1105 the soil profile) produced by the National Drought Model (NDM) and CRHM between 1992 and 2013 for years with
 1106 good records in the HYDAT database.

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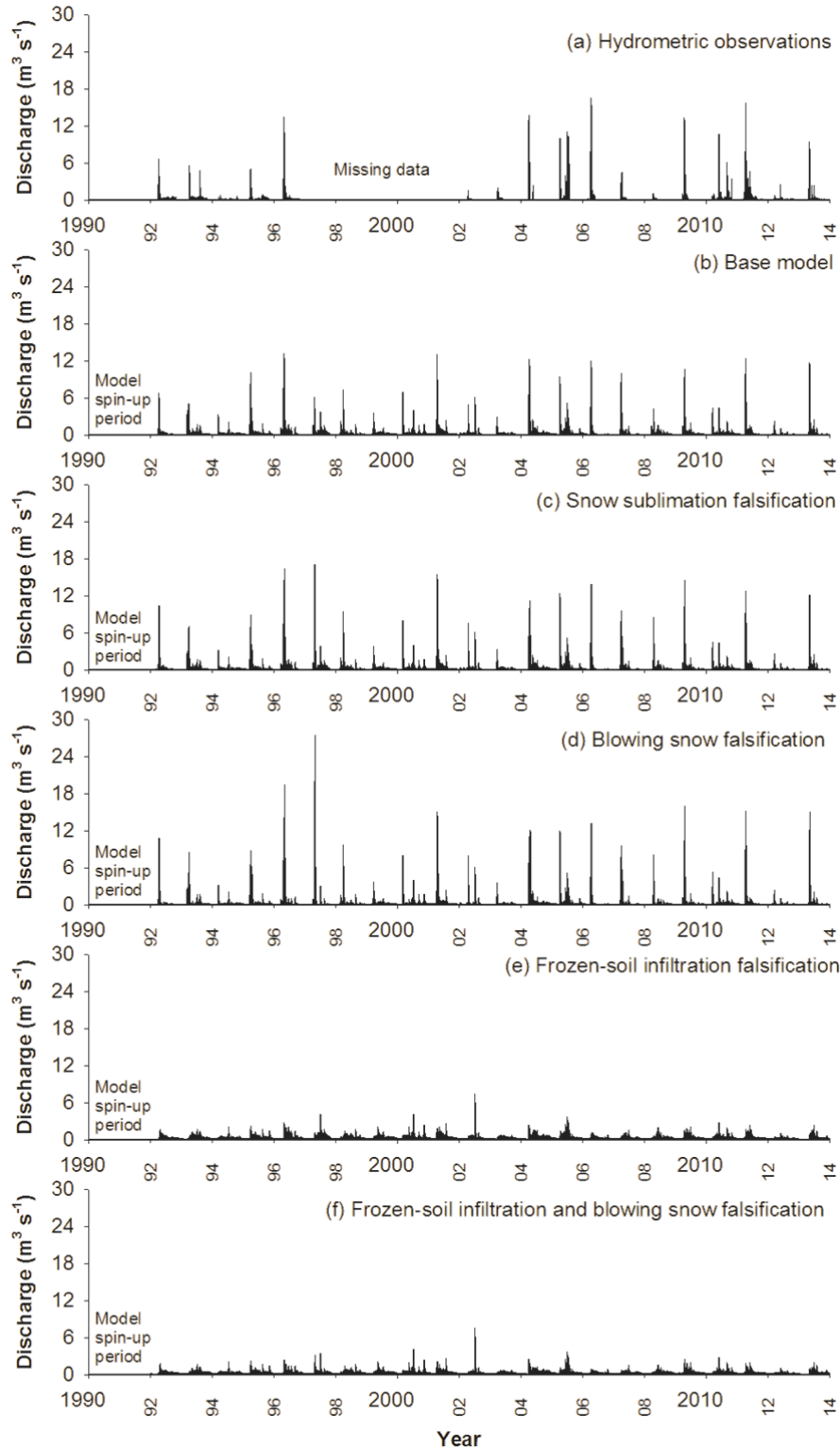


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Figure 9. Sensitivity analysis of the K storage parameter in the Muskingum model for upland and drainage network using Nash-Sutcliffe efficiency (NSE) and peak-flow model bias (MB) as objective functions.



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1112 Figure 10. Observed hydrograph (a) and model simulations (1992-2013) for the base model (b) and the
 1113 different model falsifications, which include inhibition of snow sublimation (c), blowing snow (d), frozen-soil
 1114 infiltration (e) and frozen-soil infiltration combined with blowing snow (f). Model spin-up period shown for
 1115 reference only.