

Opinion paper: Catchments as meta-organisms – a new blueprint for hydrological modelling

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Abstract

Catchment-scale hydrological models frequently miss essential characteristics of what determines the functioning of catchments. The most important active agent in catchments is the ecosystem. It manipulates and partitions moisture in a way that it supports the essential functions of survival and productivity: infiltration of water, retention of moisture, mobilization and retention of nutrients, and drainage. Ecosystems do this in the most efficient way, establishing a continuous, ever-evolving feedback loop with the landscape and climatic drivers. In brief, hydrological systems are alive and have a strong capacity to adjust themselves to prevailing and changing environmental conditions. Although most models take Newtonian theory at heart, as best they can, what they generally miss is Darwinian theory on how an ecosystem evolves and adjusts its environment to maintain crucial hydrological functions. In addition, catchments, such as many other natural systems, do not only evolve over time, but develop features of spatial organization, including surface or subsurface drainage patterns, as a by-product of this evolution. Models that fail to account for patterns and the associated feedbacks, miss a critical element of how systems at the interface of atmosphere, biosphere and pedosphere function.

In contrast to what is widely believed, relatively simple, semi-distributed conceptual models have the potential to accommodate organizational features and their temporal evolution in an efficient way. A reason for that being that because their parameters (and their evolution over time) are effective at the modelling scale, and thus integrate natural heterogeneity within the system, they may be directly inferred from observations at the same scale, reducing the need for calibration and related problems. In particular, the emergence of new and more detailed observation systems from space will lead towards a more robust understanding of spatial organization and its evolution. This will further permit the development of relatively simple time-dynamic functional relationships that can meaningfully represent spatial patterns and their evolution over time even in poorly gauged environments.

1. Introduction

“The whole is greater than the sum of the parts” and “Everything changes and nothing remains still [...]” are quotes commonly attributed to the Greek philosophers Aristotle (384-322 BC) and Heraclitus (535-475 BC). More recently, but still before Darwin developed his theory on evolution, Alexander von Humboldt (1769-1859) considered nature and its processes as an inseparable entity, where all forces of nature are connected and mutually dependent (Wulf, 2015). Although these concepts were not formulated specifically to describe the movement of water through the natural environment, they very pointedly summarize what controls hydrological functioning at the catchment scale.

Ironically, state-of-the-art catchment-scale hydrological models do, for varying reasons depending on the model under consideration, frequently a poor job in addressing overall system behaviour emerging from the characteristics above. This results in many models being inadequate representations of real-world systems, haunted by large model and/or parameter uncertainties and unreliable predictions.

1
2 There has now for several decades been an on-going controversy about the individual benefits
3 and flaws of top-down (i.e. conceptual) versus bottom-up (i.e. physically-based) modelling
4 strategies. Beven (1989), for instance, argued that the so-called "physically-based" models fail
5 to use a proper theory of up-scaling, cannot deal adequately with heterogeneity, and suffer
6 from the curse of dimensionality and the sheer impossibility of parameter calibration. These
7 problems have now, almost 3 decades later, not been overcome and still pose limitations to
8 modelling efforts, as recently highlighted by Zehe et al. (2014). Much of the ongoing
9 discussion concentrates on data uncertainty and availability. This is, without doubt, an
10 important and well-justified aspect of the discussion as it helps to improve current modelling
11 practice. Yet, largely not questioning the validity of model concepts themselves, it ignores
12 that a significant proportion of uncertainty in current-generation catchment-scale hydrological
13 models -- both conceptual and physically based -- can be directly linked to the fact that our
14 conceptual understanding of two of the critical aspects of the system, i.e. internal organization
15 and the capacity of the ecosystem to manipulate the system in response to the temporal
16 dynamics of the atmospheric drivers, as encapsulated in the above two quotes, is only
17 insufficiently or often not at all accounted for in these models. One reason for that is the
18 common absence of observations at the modelling scale of interest and our resulting inability
19 to meaningfully characterize natural heterogeneity in the model domain. This leads to the
20 largely indispensable need for model calibration (for both, conceptual and physically based
21 models), which in turn exacerbates our problem to meaningfully parameterize, test and
22 constrain models.

23
24 McDonnell et al. (2007), motivated by Dooge's (1986) paper on "Looking for hydrologic
25 laws", concluded that: "In order to make continued progress in watershed hydrology and to
26 bring greater coherence to the science, we need to move beyond the status quo of having to
27 explicitly characterize or prescribe landscape heterogeneity in our (highly calibrated) models
28 and in this way reproduce process complexity but instead explore the set of organizing
29 principles that might underlie the heterogeneity and complexity.", suggesting that we need to
30 find the organising principles underlying the apparent simplicity we can observe in system
31 behaviour.

32 33 **1.1. The whole is greater than the sum of the parts**

34 Observations from a wide range of natural systems strongly suggest that whenever one
35 medium flows through another medium as a result of a gradient, patterns appear (Savenije,
36 2009). On the surface, such patterns facilitate infiltration or drainage with limited soil loss; in
37 the unsaturated zone, patterns facilitate efficient replenishment of moisture deficits and
38 preferential drainage when there is excess moisture; in the groundwater, patterns facilitate the
39 efficient and gradual drainage of groundwater, resulting in linear reservoir recession. In the
40 surface drainage network, patterns facilitate the efficient transport of water and sediments
41 (e.g. Rodriguez-Iturbe and Rinaldo, 2001). A clear analogy with drainage patterns is water
42 flowing through a leaf or blood flowing through a body in a system of vessels, providing
43 efficient supply of, for example, water and oxygen, to all parts of the organism (e.g. West et
44 al., 1997). But there are also examples from places afar, such as ice melting on Mars forming
45 similar drainage patterns as in landscapes on Earth.

46
47 Most conceptual models already implicitly account for such structures by the use of
48 modelling components that represent some sort of preferential flow paths and which are
49 controlled by calibrated parameters, effective at the modelling scale. On the one hand, these
50 parameters integrate the natural heterogeneity of flow resistances, i.e. hydraulic

1 conductivities, of the entire model domain. On the other hand, they also characterize spatial
2 distribution functions that describe connectivity patterns of these flow paths in a spatially
3 heterogeneous domain. In contrast, despite the increasing use of conceptual formulations of
4 preferential flow paths based on dual- or multi-domain flow in newest-generation physically
5 based models (Zehe et al., 2001; Kollet and Maxwell, 2006; Sudicky et al., 2008), many
6 others rely on simple and straightforward aggregation of processes from the lab-scale to the
7 catchment scale, assuming that there is no structure and organization in the system as the
8 modelling scale increases from the grid scale to the full domain of the model application. In
9 both cases a suitable description of the emerging patterns and self-organization, which is
10 characteristic for many natural systems (e.g. Bak, 1996), is in addition hindered by the
11 typically elevated number of calibration parameters and the associated equifinality or
12 insufficient description of spatial heterogeneity when using direct observations.

13
14 Thus, according to these models, the only place in nature where there are no drainage patterns
15 is in the subsurface, i.e. in the root zone, in the unsaturated zone below it, and in the
16 groundwater. This is conceptually wrong, because subsurface drainage patterns, manifest as
17 preferential flow paths and created by diverse biological, physical and chemical processes, do
18 appear at a wide range of spatial and temporal scales. Patterns are created by, for example,
19 animal burrows (e.g. earth worms; Zehe and Flüher, 2001; Schaik et al., 2014), former root
20 channels, soil cracks, rock interfaces, and fissures, which are further reinforced by internal
21 chemical and physical erosion processes. Typically characterized by convergent flow, reduced
22 flow resistance and higher flow velocities, these patterns, as manifestations of organization,
23 provide efficient drainage as well as transport capacity for dissolved or suspended substances.
24 When zooming out to the macroscale, the time dynamic connectivity of these structures
25 frequently emerges as simple functional relationships with system wetness (e.g. Detty and
26 McGuire, 2010; Penna et al., 2011).

27 28 **1.2. Everything changes and nothing remains still**

29 The problem is not only the absence of patterns. These patterns result from evolution over
30 time. Evolution of climate and landscape have the potential to cause systemic change within
31 catchments. Such a systemic change is unlikely to be picked up at time scales smaller than the
32 calibration period with current model formulations, as the typically constant model
33 parameters define time-invariant functional relationships emerging at the scale of the model
34 domain. Only, and only if the system could be broken down into its smaller, more detailed
35 building blocks, accounting for the relevant physical, chemical and biological processes
36 involved, such a systemic change would emerge from a model. Yet, this is problematic, if not
37 impossible given current-day observation technology and our incomplete understanding of the
38 underlying mechanisms. As an illustrative example, consider the change of the interception
39 pattern over time after the conversion of grassland into forest. If detailed parameterizations of
40 vegetation growth dynamics across the model domain were part of the model, changes in
41 canopy and sub-canopy, and thus in interception pattern over time, would naturally emerge
42 from the model. Given the lack of observations and process knowledge, this is, however, not
43 feasible at scales of actual interest. Rather, functional relationships of the process emerging at
44 larger scales and at lower levels of process detail have to be used. This, however, typically
45 entails that potentially dynamic small-scale processes are lumped into constant parameters,
46 preventing the emergence of a time-variant pattern. It is therefore of critical importance to
47 realize and acknowledge that the hydrological system is not merely a dead configuration of
48 earth material through which water flows. It is the foundation of a living ecosystem that
49 manipulates and adapts the environment so as to facilitate its own survival and reproduction
50 (cf. Eagleson, 2005). Ecosystems clearly do not do this in a conscious way with an objective

1 in mind. Rather, the mere fact that they have survived past conditions in competition with
2 other species is proof that they have done so efficiently. The current state of an ecosystem is
3 then the manifestation of its development over the past. The historical evolution and not the
4 current structure or function will help us to understand potential trajectories of the system's
5 response in the future (Harman and Troch, 2014). This is Darwinian thinking, alien to the
6 purely mechanistic philosophy on which much of our state-of-the-art modelling concepts are
7 based.

8
9 Hydrological systems, at all spatial scales, from the plot to the catchment scale, therefore may
10 be understood as meta-organisms (e.g. Bosch and Miller, 2016), i.e. systems of living
11 biological entities, that occupy an ecological niche and that interact mutually but also with
12 their inanimate environment. The current appearance and characteristic of these systems is
13 clearly not the endpoint of their trajectories. Ecosystems, and hence hydrological systems,
14 continuously and dynamically evolve over a wide range of temporal and spatial scales. Yet,
15 current generation models are mostly built on the foundations of time-invariant system
16 descriptors. This modelling strategy provides us with system characterizations that are only
17 snapshots in time and that deprive us of developing a better understanding of what drives
18 change and of the systems' future trajectories.

2. The crucial elements of a hydrological model

22 Any hydrological model that claims to be physical has to properly reflect key elements of
23 hydrological systems. The first key element is the proper reflection of the partitioning points
24 that the ecosystem creates to optimise system functions: infiltration, retention, and drainage.
25 The second key element is that in the landscape patterns emerge, on and below the surface,
26 that facilitate efficient ways of drainage and infiltration.

2.1. Representation of partitioning points

29 In a hydrological system we can identify two major partitioning zones controlling how and
30 where precipitation is partitioned into different upward, downward or lateral fluxes. The first
31 partitioning zone is located at the (near-) land surface, where precipitation is split into: (1)
32 direct feedback to the atmosphere from canopy interception, ground interception, and open
33 water; (2) infiltration into the root zone; and (3) surface runoff (Hortonian infiltration excess
34 overland flow and Dunne saturation excess overland flow). Water infiltrating into the soil
35 eventually reaches the second partitioning zone, the root zone, which splits the incoming
36 moisture into: (4) transpiration by vegetation; (5) soil evaporation; (6) subsurface saturation
37 and/or infiltration excess flow, e.g. the fill-and-spill theory and/or rapid sub-surface flow
38 through preferential drainage structures within and below the root zone; and (7) percolation to
39 the groundwater.

41 If one wants to describe the hydrological functioning of a hydrological unit or catchment, an
42 accurate description critically hinges on a meaningful definition of this partitioning and the
43 residence times of the moisture in the two system partitioning zones. What characterizes and
44 shapes these two partitioning zones and thereby controls their respective functioning, are
45 largely the biotic components of the ecosystem, i.e. vegetation, animals and microorganisms
46 living in a given landscape. In fact, over the past, the ecosystem actively has manipulated (and
47 continues to do so) water fluxes and residence times in a way that the landscape provided the
48 functions that allowed the ecosystem's development to reach its current state. These functions
49 are: (1) facilitating infiltration so as to efficiently recharge root zone soil moisture and to
50 optimise subsurface drainage; (2) retention of sufficient moisture for vegetation to overcome

1 critical periods of drought; (3) efficient drainage of excess water, to ensure sufficient oxygen
2 supply for roots; and (4) maintenance of a healthy substrate with an adequate availability of
3 nutrients. The latter implies the prevention of excessive erosion and leaching of valuable
4 nutrients. If, and only if the current ecosystem manages to modify the substrate so as to
5 satisfy all these functions, it will safeguard long-term survival. It will have to do so
6 efficiently, otherwise, due to an excessive allocation of scarce resources to, for instance, the
7 growth and maintenance of the root system, insufficient resources for surface growth will be
8 available (e.g. Hildebrandt et al., 2016). As a consequence, an inefficient species will
9 experience a disadvantage in the competition with species that are more adapted to the
10 environmental conditions at a given location. They will eventually be replaced by the better
11 adapted species, changing the dynamics and pattern not only of the plant community at that
12 location but also affecting the entire ecosystem around it and thereby its influence on the
13 hydrological functioning. These changes can include for example changes to the root system,
14 the canopy or the animal and microorganism communities in the area. All of which can result
15 in changes to the pathways of water (and nutrients) through the system and eventually affect
16 how the system stores and releases water and nutrients.

17
18 There is increasing experimental (e.g. Brooks et al., 2010; Evaristo et al., 2015) and
19 theoretical evidence (e.g. Hrachowitz et al., 2013; Van der Velde et al., 2015; Zehe and
20 Jackisch, 2016) for such an eco-hydrologically controlled partitioning that regulates these
21 contrasting requirements of storage and drainage of water and nutrients, which has recently
22 been comprehensively summarized in the two-water-worlds hypothesis (McDonnell, 2014;
23 Good et al., 2015). Briefly, root systems extract water and nutrients mainly from the soil
24 matrix, which is characterized by relatively small pore sizes. In contrast, larger pores, having
25 lower specific surfaces and thus less adsorption capacity, only start to fill with increasing
26 moisture content of the soil, when the small pores are increasingly saturated. The lower flow
27 resistances in these larger subsurface features provide less buffer but rather allow for higher
28 flow velocities. They thereby provide an efficient mechanism for water to bypass the soil
29 matrix with little interaction and to drain excess water through a network of preferential
30 channels when the system is in a wet state. Although not independent of each other, water
31 stored in the matrix for transpiration and water in preferential features, generating stream
32 flow, are therefore characterized by distinct age signatures, effectively constituting distinct
33 pools of water (e.g. Hrachowitz et al., 2015). This dual system, satisfying the contrasting
34 hydrological functions of sufficient storage (of water and nutrients) and efficient drainage
35 required by an ecosystem, has developed through co-evolution of climate and hydrology with
36 the ecosystem in a Darwinian process (e.g. Sivapalan et al., 2011; Blöschl et al., 2013). Being
37 in a dynamic equilibrium, the state of such a system at any given time is a manifestation of its
38 past trajectory and reflects the conditions for survival at that time.

39 40 **2.2. The emergence of patterns and their properties**

41 Implicit in relatively simple models with little spatial discretization (i.e. mostly lumped or
42 semi-distributed conceptual models) is that there is an underlying process of maximum
43 efficiency that leads to self-organisation (e.g. Zehe et al, 2013). The Earth system is
44 continuously receiving solar energy. This energy needs to be dissipated in an efficient way to
45 produce entropy (e.g. Michaelian, 2012). According to Kleidon (2016), the process of energy
46 conversion corresponds with maximum power or maximum entropy production, close to the
47 Carnot limit, leading to the evolution of patterns of efficient transport of erosion products.
48 Eventually this self-reinforcing mechanism, i.e. positive feedback loop, creates an organised
49 drainage system (Kleidon et al., 2013).

50

1 As argued by Dooge (1986), catchments are "complex systems with some degree of
2 organisation"; in other words, it is "organised complexity" (Dooge, 2005). This organisation
3 is dominated by the ecosystem, which is not static but very much alive and continuously
4 evolving. Given the strong evidence for the interactions between hydrological functioning,
5 climate and ecosystem (e.g. Milly, 1994; Rodriguez-Iturbe and Porporato, 2007; Alila et al.,
6 2009; Gao et al., 2014a; Nijzink et al., 2016), it is inconceivable that the hydrological system
7 remains unaltered under climate or land-use change. It is rather adjusting in response to
8 changing environmental conditions and thereby actively and continuously adjusting the
9 partitioning zones at a wide range of spatial and temporal scales. The dominant ecosystem
10 that emerges will, in a Darwinian sense, then tend to maximum efficiency for survival.

11
12 The ecosystem shapes the hydrological system in a way that it converges towards a dynamic
13 equilibrium between infiltration, retention, drainage and limitation of erosion, thereby
14 creating conditions that facilitate its own survival. In a feedback, hydrology on its own term
15 then similarly shapes the ecosystem. If we want to model such systems, we have to realise
16 that our models need to reflect this dynamic and continuous feedback loop. In other words,
17 our models need to be organic and alive, just as natural systems are. Yet, to do this, there is
18 little need to describe the sub-surface partitioning zone, i.e. the unsaturated root zone, in
19 multiple layers with different properties and using root depth estimates. Such data are rarely
20 available at the level of required detail and if they are, they have mostly been obtained from
21 one-time sampling campaigns with no information about their respective temporal
22 trajectories.

23
24 Consider, as a thought experiment, the case of a plant species in a humid climate at a location
25 with a relatively poorly drained soil such as loam. From experiments with individual plants of
26 that species an estimate of average root depth at that location can be obtained. Together with
27 estimates of soil porosity, the water storage capacity in the root zone of that *specific location*
28 can be readily determined. Firstly, this approach ignores that root systems can and do adapt to
29 temporal variability in environmental conditions at time scales relevant for hydrological
30 applications. But moreover, considering that plants of the same species have common limits
31 of operation such as water and nutrient requirements, it is implausible to use the same root
32 depth estimates for the same plant growing in a drier climate and/or at a different location
33 with well-drained, coarser soils, such as sand. The estimated storage capacity of water
34 accessible to plants will be considerably underestimated and will merely reflect the
35 differences in soil properties. However, if the same species survived in a different climate or
36 on that different soil, this implies that it had sufficient access to water and nutrients. In other
37 words, the plant developed a different, i.e. deeper and/or denser, root system that could ensure
38 access to the same volume of water as in the first location (cf. Gao et al., 2014a; DeBoer-
39 Euser et al., 2016; Nijzink et al., 2016). From that we postulate that ecosystems control the
40 hydrological functioning of the root zone in a way that *continuously* optimizes the functions
41 of infiltration, moisture retention, drainage and limitation of erosion.

42
43 The result of such a co-evolution between climate, ecosystem, substrate and hydrological
44 functioning typically exhibits surprisingly simple patterns emerging at larger scales in spite of
45 the complex and highly heterogeneous combination of soils, geology, topography and climate
46 and their mutual interactions at smaller scales. Thus, even relatively simple lumped or semi-
47 distributed conceptual models have in the past shown considerable skill in reproducing
48 hydrological functioning in a wide variety of landscapes and climates. In fact, it is highly
49 likely that these models' relatively simple closure relations, based on simple system
50 descriptions that permit the integration of natural heterogeneity over the model domain, using

1 functional, emergent relationships, are manifestations of energetic optimality, most likely at a
2 state of maximum power (e.g. Kleidon, 2016).

3
4 Apparently, ecosystems are capable of creating resilience against variability and, in that
5 process, create predictable behaviour within an otherwise complex environment. Hence, mere
6 upscaling from the lab-scale to the landscape scale is insufficient if the ecosystem is not
7 included as an active agent creating resilience against the variability of nature.

8 9 **3. Why can simple conceptual models meaningfully represent these system properties?**

10 Several hydrologists have remarked on the paradox that instead of more complexity,
11 simplicity emerges in catchment behaviour as more processes come into play (e.g. Sivapalan
12 2003a). This happens at a scale where the hydrological unit has sufficient size to achieve a
13 certain level of organisation. Self-organisation leads to less complexity (Dooge, 2005).
14 Conceptual models, being a configuration of relatively simple relationships, seem therefore
15 adequate to deal with systems that have reached some degree of organisation. But it is not
16 merely the simplicity.

17
18 Let us consider a conceptual model that consists of three main stores: the surface reservoir,
19 the root zone reservoir and the groundwater reservoir. The surface reservoir represents the
20 retention of moisture by canopy and ground interception, which has a relatively small storage
21 capacity from which the moisture can evaporate directly back to the atmosphere. Above the
22 capacity threshold the moisture is split into infiltration and surface runoff, depending on a
23 threshold defined by the infiltration capacity. There is nothing non-physical about this. The
24 key lies in the infiltration function, but this is not particular for conceptual models.

25
26 The unsaturated root zone storage in a conceptual model can be brought in tune with the
27 storage requirement of the vegetation. This can be derived in a Darwinian sense and can lead
28 to scale-independent estimates of root zone storage capacity for given ecosystems (Gao et al.,
29 2014a; Nijzink et al., 2016; Wang-Erlandsson, 2016). This is a fully physical storage
30 capacity. When the store is full, sub-surface runoff and recharge is generated. At aggregate
31 scale there is spatial heterogeneity in the landscape, which leads to a distribution of thresholds
32 above which runoff is generated, describing the connectivity pattern of that system. This can
33 be done by using any suitable distribution function, such as in the Xinangjiang (Zhao & Liu,
34 1995) or VIC model (e.g. Liang et al., 1994). If the runoff mechanism is sub-surface flow, the
35 threshold is sub-surface saturation above a less permeable layer (e.g. McDonnell, 2009); if the
36 mechanism is saturation excess overland flow, it describes the increasing saturated area of a
37 catchment (Dunne & Black, 1970). Again, this is purely physical, as long as the right runoff
38 mechanism is applied to the appropriate landscape: sub-surface flow on hillslopes and Dunne
39 overland flow on landscapes where groundwater can reach the surface (wetlands and riparian
40 zones). The routing of the flow toward the stream network can be done by simple transfer
41 functions, linear reservoirs or cascades. This is just a matter of routing and does not affect the
42 partitioning or the water balance.

43
44 Finally, data from catchments worldwide suggest that groundwater systems at the catchment-
45 scale function in many cases as linear reservoirs in natural catchments, manifest in the
46 frequently observed exponential recession of the hydrograph during rainless periods, in
47 particular in lower order, upland streams where time lags introduced by channel routing are
48 limited compared to the modelling time scale. Why the dynamic part of the groundwater is
49 organised in this simple way is still one of the fundamental questions in hydrology, but the
50 answer is likely to be found in the theory of maximum power or maximum entropy

1 production. Whether or not the answer to this question will be found sooner or later does not
2 affect the viewpoint that the exponential depletion of groundwater is physical and real. The
3 linear reservoir is not more or less physical than Darcy's equation.

4
5 In spite of their rather low level of detail, conceptual models are quite capable of representing
6 these processes in a simple and adequate way, provided we account for differences in
7 landscape, ecosystem and land cover. If there is considerable heterogeneity in the climatic
8 drivers (precipitation and energy) and if these drivers are available at grid-scale, then the
9 stocks of the conceptual models can be distributed spatially, so as to account for the spatial
10 heterogeneity of the moisture states. Conceptual models do not have to be lumped, as long as
11 the system descriptors reflect the processes at the hydrological unit scale at which they
12 emerge.

13 14 **4. What are the practical consequences?**

15 Ironically, the above implies that bringing in more physics - i.e. the right kind of physics -
16 into our models makes them simpler. Apparently, simplicity - that is to say the right kind of
17 simplicity - enhances the physics of our hydrological models. If a model is complex, yet fails
18 to reproduce patterns emerging at the macroscale that characterize real-world systems as a
19 result of the evolution of the system over the past, that may be an indication of a lack of
20 physics, or of the wrong application of physics. In other words, zooming out to the
21 macroscale allows to focus on the pattern and processes emerging at that scale, which are, due
22 to the ever improving remote sensing technology, increasingly observable at the actual
23 modelling scale (see Figure 1). This offers opportunities for prediction in ungauged basins. As
24 emphasized by Sivapalan et al. (2003b), our limited ability to predict hydrological behaviour
25 is an indication of our lack of understanding of *essential* physical processes at the *macroscale*.
26 This is of particular importance the scarcer detailed observations at suitable spatial resolutions
27 and scales are. In fact, it was this inability that was the main trigger for the PUB science
28 decade (2003-2012).

29
30 There is already a wide range of remotely sensed data available that allow modellers to
31 directly exploit spatial patterns emerging at the macroscale for use in models. For example, as
32 different parts of the landscape can be associated with different dominant hydrological
33 processes, topographical indicators extracted from globally available digital elevation models,
34 such as the topographic wetness index (TWI; Beven and Kirkby, 1979; Ambroise et al., 1996;
35 Freer et al., 2004) or more recently the height above the nearest drainage (HAND; Rennó et
36 al., 2008; Savenije, 2010; Nobre et al., 2011; Gharari et al., 2011) have proven highly
37 valuable for model development, as illustrated by the example of a landscape-informed semi-
38 distributed formulation of a conceptual model in Figure 2. Similarly, increased detail in land
39 cover maps, including also products such as leaf area index, allows to account for the spatial
40 patterns of different vegetation types (e.g. Cuo et al., 2009; Li et al., 2009; Samaniego et al.,
41 2010), while the higher temporal resolution of snow cover maps permits an improved
42 representation of spatial patterns of snow accumulation and depletion (e.g. Rodell and
43 Houser, 2004; Andreadis and Lettenmaier, 2006; Nester et al., 2012). As shown by a range of
44 recent studies, these information sources can serve as efficient tools to constrain spatially
45 explicit or semi-distributed models (both conceptual and physically based) while ensuring a
46 meaningful representation of spatial patterns (e.g. Gao et al., 2014b,2016).

47
48 A further example that illustrates the value of remote sensing data to identify and quantify
49 patterns emerging at the macroscale are spatially distributed estimates of precipitation and
50 evaporation. Recent work suggests that the catchment-scale moisture retention capacity in the

1 unsaturated root zone, one of the most important parameters in terrestrial hydrological
 2 systems, can be estimated based on a Darwinian theory. If an ecosystem has been able to
 3 survive critical periods of drought, where the evaporation E was larger than the precipitation
 4 P , then apparently it had sufficient storage to overcome this drought. By simulating the
 5 storage variation resulting from P and E time series, the root zone storage capacity that the
 6 ecosystem designed can be estimated (e.g. Gao et al., 2014a; De Boer-Euser et al., 2016;
 7 Wang-Erlandsson et al., 2016; Figure 3). With this method, the root zone storage capacity of
 8 each landscape element can in principle be determined at any scale where information on E
 9 and P is available. Such observations can also be used to simulate the evolution of the root
 10 zone storage capacity as a result of land use or climate change (Nijzink et al., 2016).

11
 12 Similarly, time series of remotely sensed gravity anomalies can be related to spatio-
 13 temporally varying water storage patterns i.e. GRACE (e.g. Wahr et al., 2004). This
 14 information was in the past already successfully used to evaluate or constrain hydrological
 15 models (e.g. Winsemius et al., 2006; Krakauer and Temimi, 2011; Milzow et al. 2011; Xie et
 16 al., 2012; Reager et al., 2014). However, spatial organization allows to take this even a step
 17 further. The stream flow recession during dry periods, when the root zone is disconnected
 18 from groundwater (e.g. McDonnell, 2014) and stream flow is sustained exclusively by the
 19 groundwater, is characterized by an exponential decrease (i.e. linear reservoir) emerging at
 20 the macroscale in many catchments worldwide. During such periods, the water balance
 21 reduces to a relation between the groundwater storage S_g and the groundwater dominated
 22 outflow Q_g , which are assumed to be linearly related by a recession time scale k_g :

$$\frac{dS_g}{dt} = -Q_g = -\frac{S_g}{k_g} \quad (1)$$

23
 24
 25
 26 GRACE provides estimates of changes of the total water equivalent storage W , which is the
 27 sum of all water stores (surface, unsaturated and saturated zones). During the dry season,
 28 when there is a disconnect between the (sub-)surface and the groundwater, the temporal
 29 gradient of the surface and sub-surface stores can be replaced by $(P-E)$. If we subtract $(P-E)$
 30 from the temporal gradient of W (dW/dt), we thus obtain the recession of the groundwater
 31 storage (dS_g/dt):

$$\frac{dS_g}{dt} = \frac{dW}{dt} + (E - P) = -\frac{S_g}{k_g} \quad (2)$$

32
 33
 34
 35 The temporal recession of S_g obeys the same exponential function as the recession in the
 36 drainage network during dry periods, acting as a linear reservoir, implying that the time scale
 37 of the recession k_g reflects the recession parameter at the scale of the model application.

38 39 **4.4 Can we predict runoff without ground stations?**¹

40 Thus already with the present remote sensing-based tools, we can derive crucial hydrological
 41 parameters from pattern and organization identified through independent data sources (see
 42 Figure 2): the root zone storage capacity $S_{u,max}$ for different vegetation classes from E and P
 43 products; and the recession time scale k from gravity observations. If subsequently we
 44 estimate interception capacities S_i from land cover information, which can be done with
 45 reasonable accuracy (e.g. Samaniego et al., 2010), then there are, when using a conceptual
 46 model, only few parameters left to calibrate, such as the exponent β of the spatial distribution

¹ based on the poster presented at the symposium in honour of Eric Wood (Princeton 3 June 2016)

1 function describing the connectivity of fast flow paths (e.g. preferential flow), a splitter D
2 describing the connectivity of preferential recharge, and the fast recession time scales k_f .
3 Because in the above we have not yet simulated the entire time series, what one could do next
4 is to drive a simple conceptual model with P and calibrate on the time series of E and W (e.g.
5 Winsemius et al., 2009). This would allow estimation of the remaining three parameters.

6
7 At the present level of technology there is still considerable uncertainty in the estimation of E ,
8 P and W time series. But the quality of these products is gradually improving. In addition, we
9 have more and more access to accurate altimetry, which could in the future allow meaningful
10 calibration on water levels, making use of hydraulic equations. Already now, calibration on
11 lake levels is possible, and a few studies even already ventured in using altimetry for the
12 determination of accurate river geometry, river levels and, using hydraulic equations,
13 calibration of runoff on water levels (e.g. Sun et al., 2012, 2015).

14 15 **5. Conclusions**

16
17 As hydrological scientists, we would like all our models to be based on solid physics. On this
18 issue we do not disagree. What we sometimes disagree on, is what type of physics we need to
19 include. It is clear that both model concepts, whether "top down" conceptual, or "bottom-up"
20 physically-based, have an important role to play in discovering the physics of underlying
21 pattern formation. But for both concepts applies that if a model does not contain the pattern
22 and characteristics of an active organising agent, i.e. the ecosystem, then the model cannot
23 claim to be physical as this active agent organises moisture retention, infiltration and
24 preferential drainage.

25
26 If we realise that our physical system is organised, following some form of optimality,
27 whether we call it maximum entropy production or maximum power, then our hydrological
28 world becomes simpler and even more predictable. In recent years, the focus on small-scale
29 physics and the believe in the ever-increasing computer power, have prevented us from
30 developing holistic modelling strategies that provide plausible descriptions of how nature
31 really works at the macro scale (e.g. Savenije, 2001) and which can be encapsulated in
32 already relatively simple formulations of conceptual models.

33
34 The good news is that these holistic approaches match very well with the newly arising
35 remote sensing-based tools that are increasingly getting better. The chances are not remote
36 that the global ambition of the PUB decade to predict runoff in ungauged basins at acceptable
37 levels of certainty will be reached in the not too distant future. This is of course, provided we
38 use the right physics.

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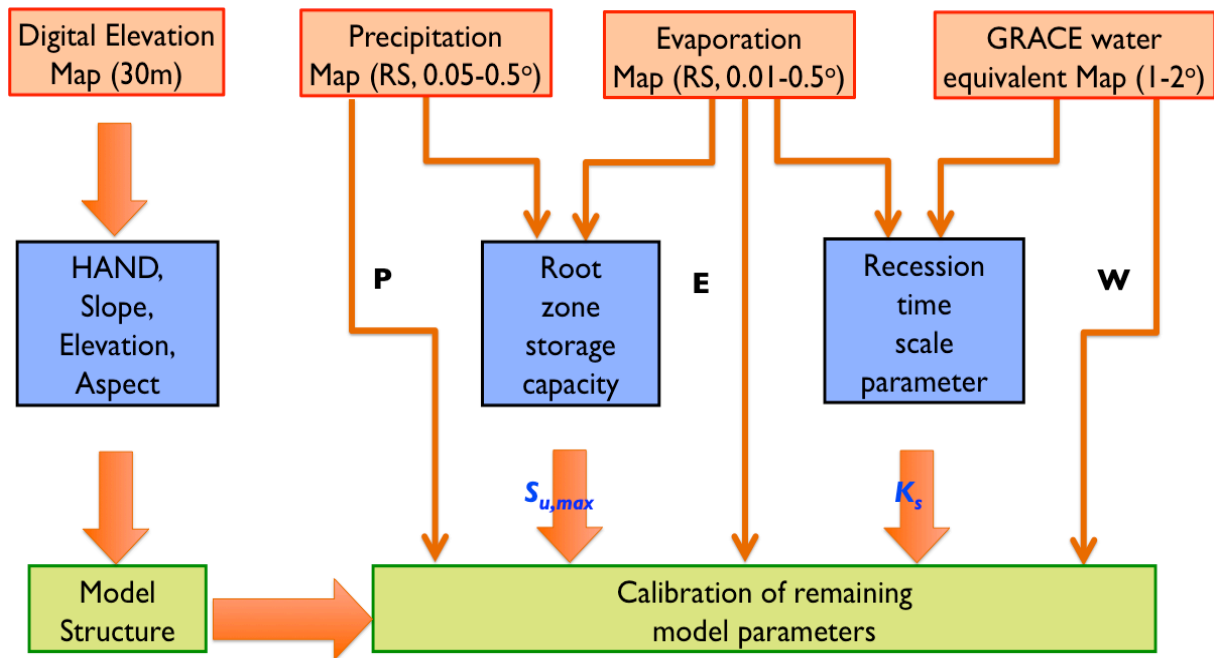
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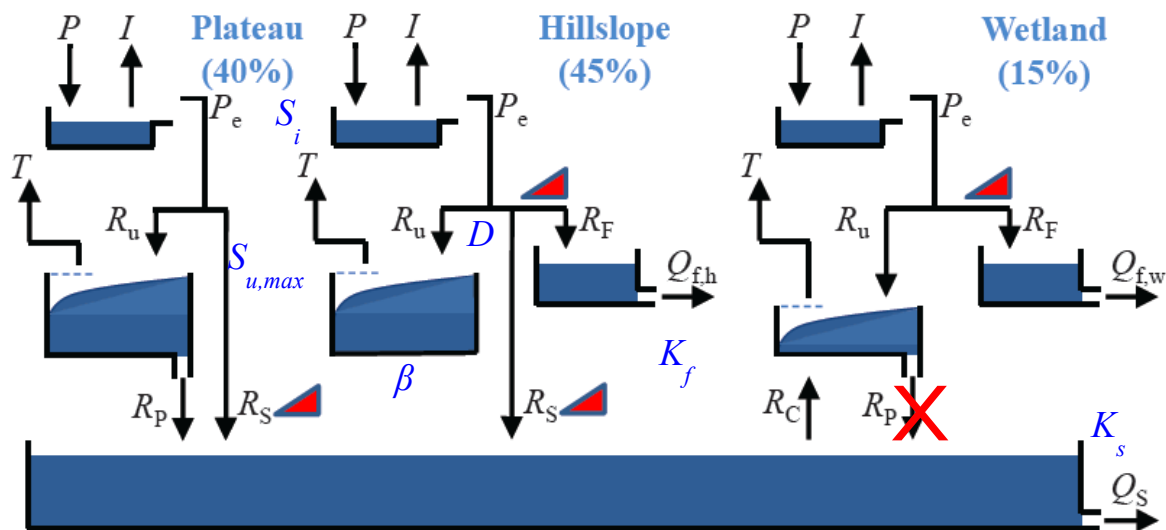
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1
2 **Figure 1.** Flow diagram for Prediction in Ungauged Basins

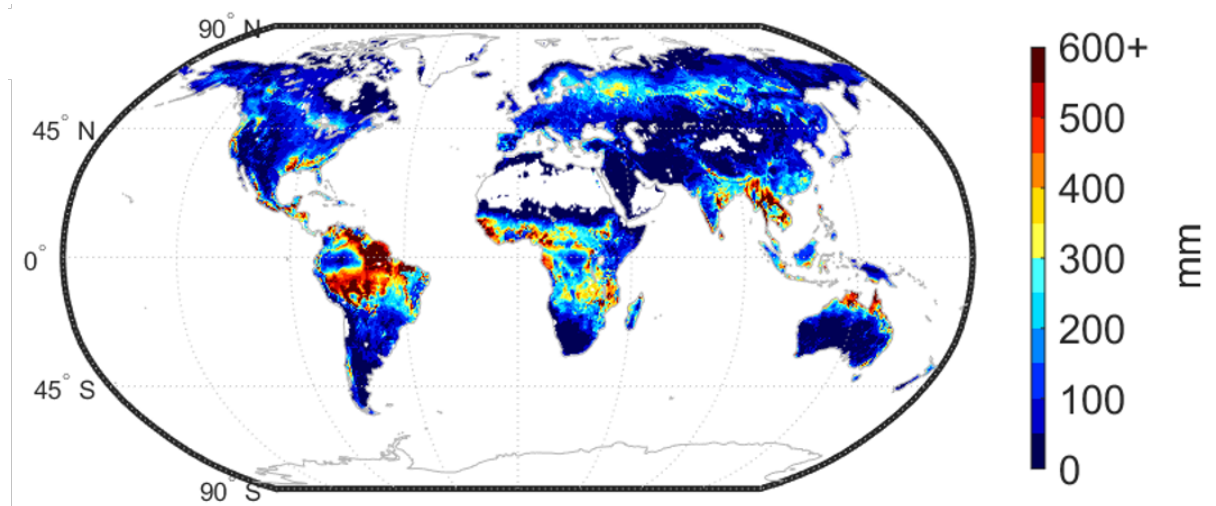
3



4
5 **Figure 2.** Example of a structure for a semi-distributed model consisting of three
6 hydrologically distinct functional units based on the respective areal proportions of three
7 landscape classes as derived from a digital elevation model, connected by a common
8 groundwater system (after Gharari et al., 2014).

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Figure 3. Root zone storage capacity as determined by Wang Erlandsson et al. (2016)

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