

Opinion paper: Modelling catchments as living organisms

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Abstract

Catchment-scale hydrological models frequently miss essential characteristics of what determines the functioning of catchments. The most important active agent in catchments is the ecosystem. It manipulates and partitions moisture in a way that it supports the essential functions of survival and productivity: infiltration of water, retention of moisture, mobilization and retention of nutrients, and drainage. Ecosystems do this in the most efficient way, establishing a continuous, ever-evolving feedback loop with the landscape and climatic drivers. In brief, our hydrological system is alive and has a strong capacity to adjust itself to prevailing and changing environmental conditions. Although most models take Newtonian theory at heart, as best they can, what they generally miss is Darwinian theory on how an ecosystem evolves and adjusts its environment to maintain crucial hydrological functions. Through a Darwinian approach, we can determine the root zone storage capacity of ecosystems, as a crucial component of hydrological models, determining the partitioning of fluxes and the conservation of moisture to bridge periods of drought. Another crucial element of physical systems is the evolution of drainage patterns, both on and below the surface. Models that do not account for these patterns are not physical. The parameters in the equations may be adjusted to compensate for the lack of patterns, but this involves scale-dependent calibration.

In contrast to what is widely believed, relatively simple conceptual models can accommodate these physical processes very efficiently. The parameters of catchment-scale conceptual models, even if they represent physical parameters, such as time scales, thresholds and reservoir sizes, require calibration or estimation on the basis of observations. Fortunately, we see the emergence of new observation systems from space that permit independent estimation of these parameters. As a result, it will become more and more practical to calibrate well-structured conceptual models, even in poorly gauged catchments.

1. Introduction

“The whole is greater than the sum of the parts” and “Everything changes and nothing remains still [...]” are quotes commonly attributed to the Greek philosophers Aristotle (384-322 BC) and Heraclitus (535-475 BC). More recently, but still before Darwin developed his theory on evolution, Alexander von Humboldt (1769-1859) considered nature and its processes as an inseparable entity, where all forces of nature are connected and mutually dependent (Wulf, 2015). Although these concepts were not formulated specifically to describe the movement of water through the natural environment, they very pointedly summarize what controls hydrological functioning at the catchment scale.

Ironically, state-of-the-art catchment-scale hydrological models do, for varying reasons depending on the model under consideration, frequently a poor job in addressing overall system behaviour emerging from the characteristics above. This results in many models being inadequate representations of real-world systems, haunted by large model and/or parameter uncertainties and unreliable predictions.

1 There has now for several decades been an on-going controversy about the individual benefits
2 and flaws of top-down (i.e. conceptual) versus bottom-up (i.e. physically-based) modelling
3 strategies. Beven (1989), for instance, argued that the so-called "physically-based" models fail
4 to use a proper theory of up-scaling, cannot deal adequately with heterogeneity, and suffer
5 from the curse of dimensionality and the sheer impossibility of parameter calibration. These
6 problems have now, almost 3 decades later, not been overcome and still pose limitations to
7 modelling efforts, as recently highlighted by Zehe et al., (2014). Much of the ongoing
8 discussion concentrates on data uncertainty and availability. This is, without doubt, an
9 important and well-justified aspect of the discussion as it helps to improve current modelling
10 practice. Yet, largely not questioning the validity of model concepts themselves, it ignores
11 that a significant proportion of uncertainty in current-generation catchment-scale hydrological
12 models -- both conceptual and physically based -- can be directly linked to the fact that our
13 conceptual understanding of two of the critical aspects of the system, i.e. internal organization
14 and the capacity of the ecosystem to manipulate the system in response to the temporal
15 dynamics of the atmospheric drivers, as encapsulated in the above two quotes, is only
16 insufficiently or often not at all accounted for in these models. One reason for that is the
17 common absence of observations at the modelling scale of interest and our resulting inability
18 to meaningfully characterize natural heterogeneity in the model domain. This leads to the
19 largely indispensable need for model calibration (for both, conceptual and physically based
20 models), which in turn exacerbates our problem to meaningfully parameterize, test and
21 constrain models.

22
23 McDonnell et al (2007), motivated by Dooge's (1986) paper on "Looking for hydrologic
24 laws", concluded that: "In order to make continued progress in watershed hydrology and to
25 bring greater coherence to the science, we need to move beyond the status quo of having to
26 explicitly characterize or prescribe landscape heterogeneity in our (highly calibrated) models
27 and in this way reproduce process complexity but instead explore the set of organizing
28 principles that might underlie the heterogeneity and complexity."; suggesting that we need to
29 find the organising principles underlying the apparent simplicity we can observe in system
30 behaviour.

31 32 **1.1. The whole is greater than the sum of the parts**

33 Observations from a wide range of natural systems strongly suggest that whenever one
34 medium flows through another medium as a result of a gradient, patterns appear (Savenije,
35 2009). On the surface, such patterns facilitate infiltration or drainage with limited soil loss; in
36 the unsaturated zone, patterns facilitate efficient replenishment of moisture deficits and
37 preferential drainage when there is excess moisture; in the groundwater, patterns facilitate the
38 efficient and gradual drainage of groundwater, resulting in linear reservoir recession. In the
39 surface drainage network, patterns facilitate the efficient transport of water and sediments
40 (e.g. Rodriguez-Iturbe and Rinaldo, 2001). A clear analogy with drainage patterns is water
41 flowing through a leaf or blood flowing through a body in a system of veins, providing
42 efficient supply of, for example, water and oxygen, to all parts of the organism (e.g. West et
43 al., 1997). But there are also examples from places afar, such as ice melting on Mars forming
44 similar drainage patterns as in landscapes on Earth. Most conceptual models already
45 implicitly account for such structures by the use of modelling components that represent some
46 sort of preferential flow paths and which are controlled by calibrated parameters, effective at
47 the modelling scale. On the one hand, these parameters integrate the natural heterogeneity of
48 flow resistances, i.e. hydraulic conductivities, of the entire model domain. On the other hand,
49 they also characterize spatial distribution functions that describe connectivity patterns of these
50 flow paths in a spatially heterogeneous domain. In contrast, despite the increasing use of

1 conceptual formulations of preferential flow paths based on dual- or multi-domain flow in
2 newest-generation physically based models (Zehe et al., 2001; Kollet and Maxwell, 2006;
3 Sudicky et al., 2008), many others rely on simple and straightforward aggregation of
4 processes from the lab-scale to the catchment scale, assuming that there is no structure and
5 organization in the system as the modelling scale increases from the grid scale to the full
6 domain of the model application. In both cases a suitable description of the emerging patterns
7 and self-organization which is characteristic for many natural systems (e.g. Bak, 1996) is in
8 addition hindered by the typically elevated number of calibration parameters and the
9 associated equifinality or insufficient description of spatial heterogeneity when using direct
10 observations. Thus, according to these models, the only place in nature where there are no
11 drainage patterns is in the subsurface, i.e. in the root zone, in the unsaturated zone below it,
12 and in the groundwater. This is conceptually wrong, because subsurface drainage patterns,
13 manifest as preferential flow paths and created by diverse biological, physical and chemical
14 processes, do appear at a wide range of spatial and temporal scales. Patterns are created by,
15 for example, animal burrows (e.g. earth worms; Zehe and Flüher, 2001; Schaik et al., 2014),
16 former root channels, soil cracks, rock interfaces, and fissures, which are further reinforced by
17 internal chemical and physical erosion processes. Typically characterized by convergent flow,
18 reduced flow resistance and higher flow velocities, these patterns, as manifestations of
19 organization, provide efficient drainage as well as transport capacity for dissolved or
20 suspended substances. When zooming out to the macroscale, the time dynamic connectivity
21 of these structures frequently emerges as simple functional relationships with system wetness
22 (e.g. Detty and McGuire, 2010; Penna et al., 2011).

23 24 **1.2. Everything changes and nothing remains still**

25 But the problem is not only the absence of patterns. These pattern result from evolution over
26 time. Evolution of climate and landscape have the potential to cause systemic change within
27 catchments. Such a systemic change is unlikely to be picked up at time scales smaller than the
28 calibration period with current model formulations, as the typically constant model
29 parameters define time-invariant functional relationships emerging at the scale of the model
30 domain. Only, and only if the system could be broken down into its smaller, more detailed
31 building blocks and accounting for the relevant physical, chemical and biological processes
32 involved, such a systemic change would emerge from a model. Yet, this is problematic, if not
33 impossible given current-day observation technology and our incomplete understanding of the
34 underlying mechanisms. As an illustrative example, consider the change of the interception
35 pattern over time after the conversion of grassland into forest. If detailed parameterizations of
36 vegetation growth dynamics across the model domain were part of the model, changes in
37 canopy and sub-canopy, and thus in interception pattern over time, would naturally emerge
38 from the model. Given our lack of observations and process knowledge, this is, however, not
39 feasible at scales of actual interest. Rather, functional relationships of the process emerging at
40 larger scales and at lower levels of process detail have to be used. This, however, typically
41 entails that potentially dynamic small-scale processes are lumped into constant parameters,
42 preventing the emergence of a time-variant pattern. It is therefore of critical importance to
43 realize and acknowledge that the hydrological system is not merely a dead configuration of
44 earth material through which water flows. It is the foundation of a living ecosystem that
45 manipulates and adapts the environment so as to facilitate its own survival and reproduction.
46 Ecosystems clearly do not do this in a conscious way with an objective in mind. Rather, the
47 mere fact that they have survived past conditions in competition with other species is proof
48 that they have done so efficiently. The current state of an ecosystem is then the manifestation
49 of its development over the past. This trajectory and not the current structure controls system

1 response. This is Darwinian thinking, alien to the purely mechanistic philosophy on which
2 much of our state-of-the-art modelling concepts are based.

3
4 Hydrological systems, at all spatial scales, from the plot to the catchment scale, therefore may
5 be understood as meta-organisms (e.g. Bosch and Miller, 2016), i.e. systems of living
6 biological entities, that occupy an ecological niche and that interact mutually but also with
7 their inanimate environment. The current appearance and characteristic of these systems is
8 clearly not the endpoint of their trajectories. Ecosystems, and hence hydrological systems,
9 continuously and dynamically evolve over a wide range of temporal and spatial scales. Yet,
10 current generation models are mostly built on the foundations of time-invariant system
11 descriptors. This modelling strategy provides us with system characterizations that are only
12 snapshots in time and that deprive us of developing a better understanding of what drives
13 change and of the systems' future trajectories.

16 **2. The crucial elements of a hydrological model**

17 Any hydrological model that claims to be physical has to properly reflect key elements of
18 hydrological systems. The first key element is the proper reflection of the partitioning points
19 that the ecosystem creates to optimise system functions: infiltration, retention, and drainage.
20 The second key element is that in the landscape patterns emerge, on and below the surface,
21 that facilitate efficient ways of drainage and infiltration.

23 **2.1. Representation of partitioning points**

24 In a hydrological system we can identify two major partitioning zones controlling how and
25 where precipitation is partitioned into different upward, downward or lateral fluxes. The first
26 partitioning zone is located at the (near-) land surface, where precipitation is split into: (1)
27 direct feedback to the atmosphere from canopy interception, ground interception, and open
28 water; (2) infiltration into the root zone; and (3) surface runoff (Hortonian infiltration excess
29 overland flow and Dunne saturation excess overland flow). Water infiltrating into the soil
30 eventually reaches the second partitioning zone, the root zone, which splits the incoming
31 moisture into: (4) transpiration by vegetation; (5) soil evaporation; (6) subsurface saturation
32 and/or infiltration excess flow, e.g. the fill-and-spill theory and/or rapid sub-surface flow
33 through preferential drainage structures within and below the root zone; and (7) percolation to
34 the groundwater.

35
36 If one wants to describe the hydrological functioning of a hydrological unit or catchment, an
37 accurate description critically hinges on a meaningful definition of this partitioning and the
38 residence times of the moisture in the two system partitioning zones. What characterizes and
39 shapes these two partitioning zones and thereby controls their respective functioning, are
40 largely the biotic components of the ecosystem, i.e. vegetation, animals and microorganisms
41 living in a given landscape. In fact, over the past, the ecosystem actively has manipulated (and
42 continues to do so) water fluxes and residence times in a way that the landscape provided the
43 functions that allowed the ecosystem's development to reach its current state. These functions
44 are: (1) facilitating infiltration so as to efficiently recharge root zone soil moisture and to
45 optimise subsurface drainage; (2) retention of sufficient moisture for vegetation to overcome
46 critical periods of drought; (3) efficient drainage of excess water, to ensure sufficient oxygen
47 supply for roots; and (4) maintenance of a healthy substrate with an adequate availability of
48 nutrients. The latter implies the prevention of excessive erosion and leaching of valuable
49 nutrients. If, and only if the current ecosystem manages to modify the substrate so as to
50 satisfy all these functions, it will safeguard long-term survival. It will have to do so

1 efficiently, otherwise, due to an excessive allocation of scarce resources to, for instance, the
2 growth and maintenance of the root system, insufficient resources for surface growth will be
3 available. As a consequence, an inefficient species will experience a disadvantage in the
4 competition with species that are more adapted to the environmental conditions at a given
5 location. They will eventually be replaced by the better adapted species, changing the
6 dynamics and pattern not only of the plant community at that location but also affecting the
7 entire ecosystem around it and thereby its influence on the hydrological functioning. These
8 changes can include for example changes to the root system, the canopy or the animal and
9 microorganism communities in the area. All of which can result in changes to the pathways of
10 water (and nutrients) through the system and eventually affect how the system stores and
11 releases water and nutrients.

12
13 There is increasing experimental (e.g. Brooks et al., 2010; Evaristo et al., 2015) and
14 theoretical evidence (e.g. Hrachowitz et al., 2013; Van der Velde et al., 2015) for such an eco-
15 hydrologically controlled partitioning that regulates these contrasting requirements of storage
16 and drainage of water and nutrients, which has recently been comprehensively summarized in
17 the two-water-worlds hypothesis (McDonnell, 2014; Good et al., 2015). Briefly, root systems
18 extract water and nutrients mainly from the soil matrix, which is characterized by relatively
19 small pore sizes. In contrast, larger pores, having lower specific surfaces and thus less
20 adsorption capacity, only start to fill with increasing moisture content of the soil, when the
21 small pores are increasingly saturated. The lower flow resistances in these larger subsurface
22 features provide less buffer but rather allow for higher flow velocities. They thereby provide
23 an efficient mechanism for water to bypass the soil matrix with little interaction and to drain
24 excess water through a network of preferential channels when the system is in a wet state.
25 Although not independent of each other, water stored in the matrix for transpiration and water
26 in preferential features, generating stream flow, are therefore characterized by distinct age
27 signatures, effectively constituting distinct pools of water (e.g. Hrachowitz et al., 2015). This
28 dual system, satisfying the contrasting hydrological functions of sufficient storage (of water
29 and nutrients) and efficient drainage required by an ecosystem, has developed through co-
30 evolution of climate and hydrology with the ecosystem in a Darwinian process (e.g. Sivapalan
31 et al., 2011; Blöschl et al., 2013). Being in a dynamic equilibrium, the state of such a system
32 at any given time is a manifestation of its past trajectory and reflects the conditions for
33 survival at that time.

34 35 **2.2. The emergence of patterns and their properties**

36 Implicit in relatively simple models with little spatial discretization (i.e. mostly lumped or
37 semi-distributed conceptual models) is that there is an underlying process of maximum
38 efficiency that leads to self-organisation (e.g. Zehe et al, 2013). The Earth system is
39 continuously receiving solar energy. This energy needs to be dissipated in an efficient way to
40 produce entropy (e.g. Michaelian, 2012). According to Kleidon (2016), the process of energy
41 conversion corresponds with maximum power or maximum entropy production, close to the
42 Carnot limit, leading to the evolution of patterns of efficient transport of erosion products.
43 Eventually this self-reinforcing mechanism, i.e. positive feedback loop, creates an organised
44 system of drainage (Kleidon et al., 2013).

45
46 As argued by Dooge (1986), catchments are "complex systems with some degree of
47 organisation"; in other words, it is "organised complexity" (Dooge, 2005). This organisation
48 is dominated by the ecosystem, which is not static but very much alive and continuously
49 evolving. Given the strong evidence for the interactions between hydrological functioning,
50 climate and ecosystem (e.g. Milly, 1994; Rodriguez-Iturbe and Porporato, 2007; Alila et al.,

1 2009; Gao et al., 2014a; Nijzink et al., 2016), it is inconceivable that the hydrological system
2 remains unaltered under climate or land-use change. It is rather adjusting in response to
3 changing environmental conditions and thereby actively and continuously adjusting the
4 partitioning zones at a wide range of spatial and temporal scales. The dominant ecosystem
5 that emerges will, in a Darwinian sense, then tend to maximum efficiency for survival.

6
7 The ecosystem shapes the hydrological system in a way that it converges towards a dynamic
8 equilibrium between infiltration, retention, drainage and limitation of erosion, thereby
9 creating conditions that facilitate its own survival. In a feedback, hydrology on its own term
10 then similarly shapes the ecosystem. If we want to model such systems, we have to realise
11 that our models need to reflect this dynamic and continuous feedback loop. In other words,
12 our models need to be organic and alive, just as natural systems are. Yet, to do this, there is
13 little need to describe the sub-surface partitioning zone, i.e. the unsaturated root zone, in
14 multiple layers with different properties and using root depth estimates. Such data is rarely
15 available at the level of required detail and if it is, it has mostly been obtained from one-time
16 sampling campaigns with no information about their respective temporal trajectories.

17
18 Consider, as a thought experiment, the case of a plant species in a humid climate at a location
19 with a relatively poorly drained soil such as loam. From experiments with individual plants of
20 that species an estimate of average root depth at that location can be obtained. Together with
21 estimates of soil porosity, the water storage capacity in the root zone of that *specific location*
22 can be readily determined. Firstly, this approach ignores that root systems can and do adapt to
23 temporal variability in environmental conditions at time scales relevant for hydrological
24 applications. But moreover, considering that plants of the same species have common limits
25 of operation such as water and nutrient requirements, it is implausible to use the same root
26 depth estimates for the same plant growing in a drier climate and/or at a different location
27 with well-drained, coarser soils, such as sand. The estimated storage capacity of water
28 accessible to plants will be considerably underestimated and will merely reflect the
29 differences in soil properties. However, if the same species survived in a different climate or
30 on that different soil, this implies that it had sufficient access to water and nutrients. In other
31 words, the plant developed a different, i.e. deeper and/or denser, root system that could ensure
32 access to the same volume of water as in the first location (cf. Gao et al., 2014a; DeBoer-
33 Euser et al., 2016; Nijzink et al., 2016). Thus, ecosystems control the hydrological
34 functioning of the root zone in a way that *continuously* optimizes the functions of infiltration,
35 moisture retention, drainage and limitation of erosion.

36
37 The result of such a co-evolution between climate, ecosystem, substrate and hydrological
38 functioning typically exhibits surprisingly simple patterns emerging at larger scales in spite of
39 the complex and highly heterogeneous combination of soils, geology, topography and climate
40 and their mutual interactions at smaller scales. Thus, even relatively simple lumped or semi-
41 distributed conceptual models have in the past shown considerable skill in reproducing
42 hydrological functioning in a wide variety of landscapes and climates. In fact, it is highly
43 likely that these models' relatively simple closure relations, based on simple system
44 descriptions that permit the integration of natural heterogeneity over the model domain, using
45 functional, emergent relationships, are manifestations of energetic optimality, most likely at a
46 state of maximum power (e.g. Kleidon, 2016).

47
48 Apparently, ecosystems are capable of creating resilience against variability and, in that
49 process, create predictable behaviour within an otherwise complex environment. Hence, mere

1 upscaling from the lab-scale to the landscape scale is insufficient if the ecosystem is not
2 included as an active agent creating resilience against the variability of nature.

3 **3. Why can simple conceptual models meaningfully represent these system properties?**

4 Several hydrologists have remarked on the paradox that instead of more complexity,
5 simplicity emerges in catchment behaviour as more processes come into play (e.g. Sivapalan
6 2003a). This happens at a scale where the hydrological unit has sufficient size to achieve a
7 certain level of organisation. Self-organisation leads to less complexity (Dooge, 2005).
8 Conceptual models, being a configuration of relatively simple relationships, seem therefore
9 adequate to deal with systems that have reached some degree of organisation. But it is not
10 merely the simplicity.

11
12
13 Let us consider a conceptual model that consists of three main stores: the surface reservoir,
14 the root zone reservoir and the groundwater reservoir. The surface reservoir represents the
15 retention of moisture by canopy and ground interception, which has a relatively small storage
16 capacity from which the moisture can evaporate directly back to the atmosphere. Above the
17 capacity threshold the moisture is split into infiltration and surface runoff, depending on a
18 threshold defined by the infiltration capacity. There is nothing non-physical about this. The
19 key lies in the infiltration function, but this is not particular for conceptual models.

20
21 The unsaturated root zone storage in a conceptual model can be brought in tune with the
22 storage requirement of the vegetation. This can be derived in a Darwinian sense and can lead
23 to scale-independent estimates of root zone storage capacity for given ecosystems (Gao et al.,
24 2014a; Nijzink et al., 2016; Wang-Erlandsson, 2016). This is a fully physical storage
25 capacity. When the store is full, sub-surface runoff and recharge is generated. At aggregate
26 scale there is spatial heterogeneity in the landscape, which leads to a distribution of thresholds
27 above which runoff is generated, describing the connectivity pattern of that system. This can
28 be done by using any suitable distribution function, such as in the Xinangjiang (Zhao & Liu,
29 1995) or VIC model (e.g. Liang et al., 1994). If the runoff mechanism is sub-surface flow, the
30 threshold is sub-surface saturation above a less permeable layer (e.g. McDonnell, 2009); if the
31 mechanism is saturation excess overland flow, it describes the increasing saturated area of a
32 catchment (Dunne & Black, 1970). Again, this is purely physical, as long as the right runoff
33 mechanism is applied to the appropriate landscape: sub-surface (McDonnell) flow on
34 hillslopes and Dunne overland flow on landscapes where groundwater can reach the surface
35 (wetlands and riparian zones). The routing of the flow toward the stream network can be done
36 by simple transfer functions, linear reservoirs or cascades. This is just a matter of routing and
37 does not affect the partitioning or the water balance.

38
39 Finally, data from catchments worldwide suggest that groundwater systems at the catchment-
40 scale function in many cases as linear reservoirs in natural catchments, manifest in the
41 frequently observed exponential recession of the hydrograph during rainless periods, in
42 particular in lower order, upland streams where time lags introduced by channel routing are
43 limited compared to the modelling time scale. Why the dynamic part of the groundwater is
44 organised in this simple way is still one of the fundamental questions in hydrology, but the
45 answer is likely to be found in the theory of maximum power or maximum entropy
46 production. Whether or not the answer to this question will be found sooner or later does not
47 affect the viewpoint that the exponential depletion of groundwater is physical and real. The
48 linear reservoir is not more or less physical than Darcy's equation.

1 In spite of their rather low level of detail, conceptual models are quite capable of representing
2 these processes in a simple and adequate way, provided we account for differences in
3 landscape, ecosystem and land cover. If there is considerable heterogeneity in the climatic
4 drivers (precipitation and energy) and if these drivers are available at grid-scale, then the
5 stocks of the conceptual models can be distributed spatially, so as to account for the spatial
6 heterogeneity of the moisture states. Conceptual models do not have to be lumped, as long as
7 the system descriptors reflect the processes at the hydrological unit scale at which they
8 emerge.

9 10 **4. What are the practical consequences?**

11 Ironically, the above implies that bringing in more physics -- i.e. the right kind of physics --
12 into our models makes them simpler. Apparently, simplicity -- that is to say the right kind of
13 simplicity -- enhances the physics of our hydrological models. If a model is complex, yet fails
14 to reproduce patterns emerging at the macroscale that characterize real-world systems as a
15 result of the evolution of the system over the past, that may be an indication of a lack of
16 physics, or of the wrong application of physics. In other words, zooming out to the
17 macroscale allows to focus on the pattern and processes emerging at that scale, which are, due
18 to the ever improving remote sensing technology, increasingly observable at the actual
19 modelling scale (see Figure 1). This offers opportunities for prediction in ungauged basins. As
20 emphasized by Sivapalan et al. (2003b), our limited ability to predict hydrological behaviour
21 is an indication of our lack of understanding of *essential* physical processes at the *macroscale*.
22 This is of particular importance the scarcer detailed observations at suitable spatial resolutions
23 and scales are. In fact, it was this inability that was the main trigger for the PUB science
24 decade (2003-2012).

25
26 There is already a wide range of remotely sensed data available that allow modellers to
27 directly exploit spatial patterns emerging at the macroscale for use in models. For example, as
28 different parts of the landscape can be associated with different dominant hydrological
29 processes, topographical indicators extracted from globally available digital elevation models,
30 such as the topographic wetness index (TWI; Beven and Kirkby, 1979; Ambroise et al., 1996;
31 Freer et al., 2004) or more recently the height above the nearest drainage (HAND; Rennó et
32 al., 2008; Savenije, 2010; Nobre et al., 2011; Gharari et al., 2011) have proven highly
33 valuable for model development, as illustrated by the example of a landscape-informed semi-
34 distributed formulation of a conceptual model in Figure 2. Similarly, increased detail in land
35 cover maps, including also products such as leaf area index, allows to account for the effect
36 spatial pattern of different vegetation types (e.g. Cuo et al., 2009; Li et al., 2009; Samaniego
37 et al., 2010), while the higher temporal resolution of snow cover maps permits an improved
38 representation of spatial patterns of snow accumulation and depletion (e.g. Rodell and
39 Houser, 2004; Andreadis and Lettenmaier, 2006; Nester et al., 2012). As shown by a range of
40 recent studies, these information sources can serve as efficient tools to constrain spatially
41 explicit or semi-distributed models (both conceptual and physically based) while ensuring a
42 meaningful representation of spatial patterns (e.g. Gao et al., 2014b,2016).

43
44 A further example that illustrates the value of remote sensing data to identify and quantify
45 pattern emerging at the macroscale are spatially distributed estimates of precipitation and
46 evaporation. Recent work suggests that the catchment-scale moisture retention capacity in the
47 unsaturated root zone, one of the most important parameters in terrestrial hydrological
48 systems, can be estimated based on a Darwinian theory. If an ecosystem has been able to
49 survive critical periods of drought, where the evaporation E was larger than the precipitation
50 P , then apparently it had sufficient storage to overcome this drought. By simulating the

1 storage variation resulting from P and E time series, the root zone storage capacity that the
 2 ecosystem designed can be estimated (e.g. Gao et al., 2014a; De Boer-Euser et al., 2016;
 3 Nijzink et al., 2016; Wang-Erlandsson et al., 2016; Figure 3). With this method, the root zone
 4 storage capacity of each landscape element can in principle be determined at any scale where
 5 information on E and P is available.

6
 7 Similarly, time series of remotely sensed gravity anomalies can be related to spatio-
 8 temporally varying water storage patterns i.e. GRACE (e.g. Wahr et al., 2004). This
 9 information was in the past already successfully used to evaluate or constrain hydrological
 10 models (e.g. Winsemius et al., 2006; Krakauer and Temimi, 2011; Milzow et al. 2011; Xie et
 11 al., 2012; Reager et al., 2014). However, spatial organization allows to take this even a step
 12 further. The stream flow recession during dry periods, when the root zone is disconnected
 13 from groundwater (e.g. McDonnell, 2014) and stream flow is sustained exclusively by the
 14 groundwater, is characterized by an exponential decrease (i.e. linear reservoir) emerging at
 15 the macroscale in many catchments worldwide. During such periods, the water balance
 16 reduces to a relation between the groundwater storage S_g and the groundwater dominated
 17 outflow Q_g , which are assumed to be linearly related by a recession time scale k_g :

$$18 \quad \frac{dS_g}{dt} = -Q_g = -\frac{S_g}{k_g} \quad (1)$$

19
 20 GRACE provides estimates of changes of the total water equivalent storage W , which is the
 21 sum of all water stores (surface, unsaturated and saturated zones). During the dry season,
 22 when there is a disconnect between the (sub-)surface and the groundwater, the temporal
 23 gradient of the surface and sub-surface stores can be replaced by $(P-E)$. If we subtract $(P-E)$
 24 from the temporal gradient of W (dW/dt), we thus obtain the recession of the groundwater
 25 storage (dS_g/dt , see Figure 4):

$$26 \quad \frac{dS_g}{dt} = \frac{dW}{dt} + (E - P) = -\frac{S_g}{k_g} \quad (2)$$

27
 28 The temporal recession of S_g obeys the same exponential function as the recession in the
 29 drainage network during dry periods, acting as a linear reservoir, implying that the time scale
 30 of the recession k_g reflects the recession parameter at the scale of the model application.

31 **4.4 Can we predict runoff without ground stations?**¹

32 Thus already with the present remote sensing-based tools, we can derive crucial hydrological
 33 parameters from pattern and organization identified through independent data sources (see
 34 Figure 2): the root zone storage capacity $S_{u,max}$ for different vegetation classes from E and P
 35 products; and the recession time scale k from gravity observations. If subsequently we
 36 estimate interception capacities S_i from land cover information, which can be done with
 37 reasonable accuracy (e.g. Samaniego et al., 2010), then there are, when using a conceptual
 38 model, only few parameters left to calibrate, such as the exponent β of the spatial distribution
 39 function describing the connectivity of fast flow paths (e.g. preferential flow), a splitter D
 40 describing the connectivity of preferential recharge, and the fast recession time scales k_f .
 41 Because in the above we have not yet simulated the entire time series, what one could do next
 42 is to drive a simple conceptual model with P and calibrate on the time series of E and W (e.g.
 43 Winsemius et al., 2009). This would allow estimation of the remaining three parameters.

¹ based on the poster presented at the symposium in honour of Eric Wood (Princeton 3 June 2016)

1
2 At the present level of technology there is still considerable uncertainty in the estimation of E ,
3 P and W time series. But the quality of these products is gradually improving. In addition, we
4 have more and more access to accurate altimetry, which could in the future allow meaningful
5 calibration on water levels, making use of hydraulic equations. Already now, calibration on
6 lake levels is possible, and a few studies even already ventured in using altimetry for the
7 determination of accurate river geometry, river levels and, using hydraulic equations,
8 calibration of runoff on water levels (e.g. Sun et al., 2012, 2015).

9 10 **5. Conclusions**

11
12 As hydrological scientists, we would like all our models to be based on solid physics. On this
13 issue we do not disagree. What we sometimes disagree on, is what type of physics we need to
14 include. It is clear that both model concepts, whether "top down" conceptual, or "bottom-up"
15 physically-based, have an important role to play in discovering the physics of underlying
16 pattern formation. But for both concepts applies that if a model does not contain the pattern
17 and characteristics of an active organising agent, i.e. the ecosystem, then the model cannot
18 claim to be physical as this active agent organises moisture retention, infiltration and
19 preferential drainage.

20
21 If we realise that our physical system is organised, following some form of optimality,
22 whether we call it maximum entropy production or maximum power, then our hydrological
23 world becomes simpler and even more predictable. In recent years, the focus on small-scale
24 physics and the believe in the ever-increasing computer power, have prevented us from
25 developing holistic modelling strategies that provide plausible descriptions of how nature
26 really works at the macro scale (e.g. Savenije, 2001) and which can be encapsulated in
27 already relatively simple formulations of conceptual models.

28
29 The good news is that these holistic approaches match very well with the newly arising
30 remote sensing-based tools that are increasingly getting better. The chances are not remote
31 that the global ambition of the PUB decade to predict runoff in ungauged basins at acceptable
32 levels of certainty will be reached in the not too distant future. This is of course, provided we
33 use the right physics.

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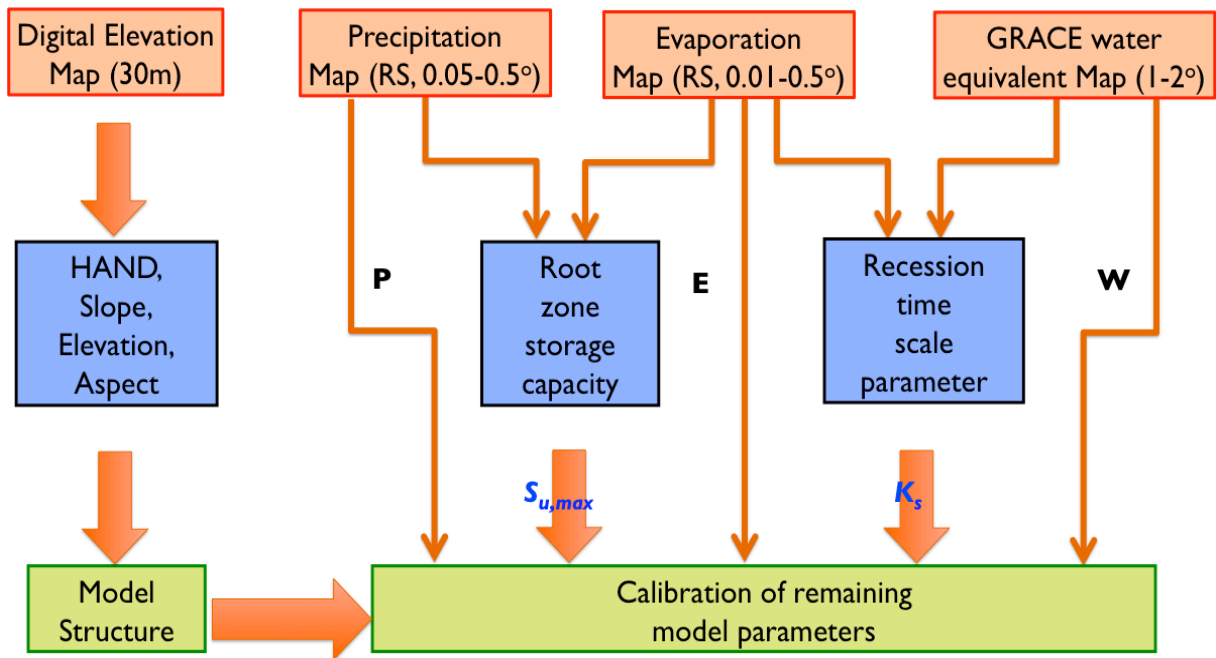
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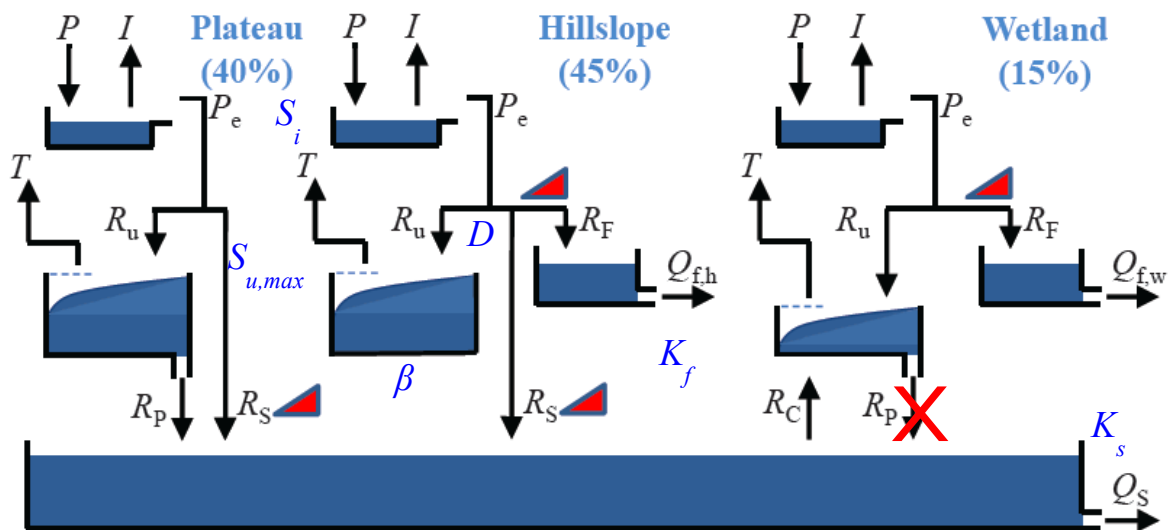
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1
2 **Figure 1.** Flow diagram for Prediction in Ungauged Basins

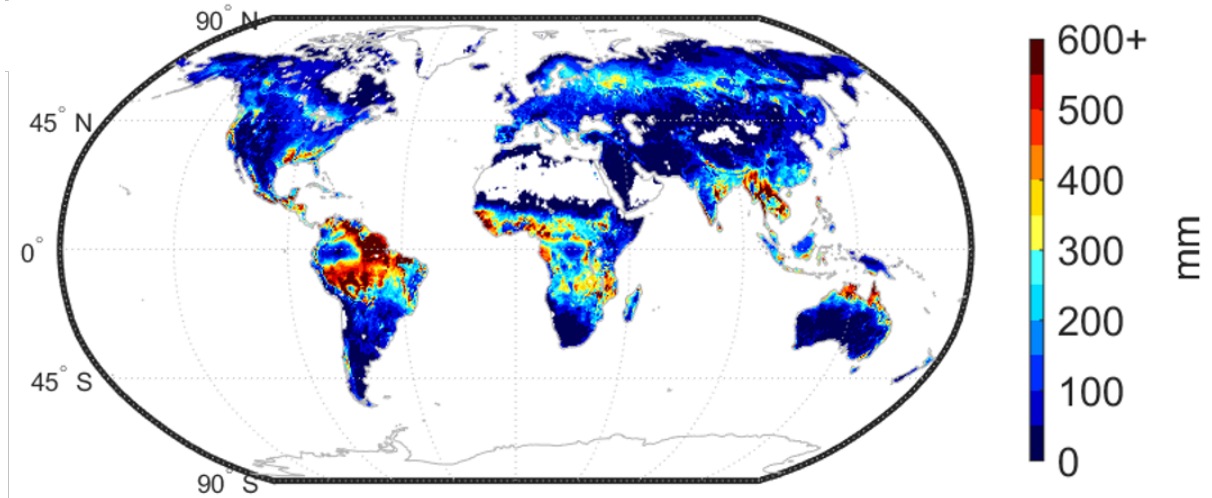
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4
5 **Figure 2.** Model structure derived from DEM, showing three landscape classes and the
6 groundwater system connecting them.

7

1

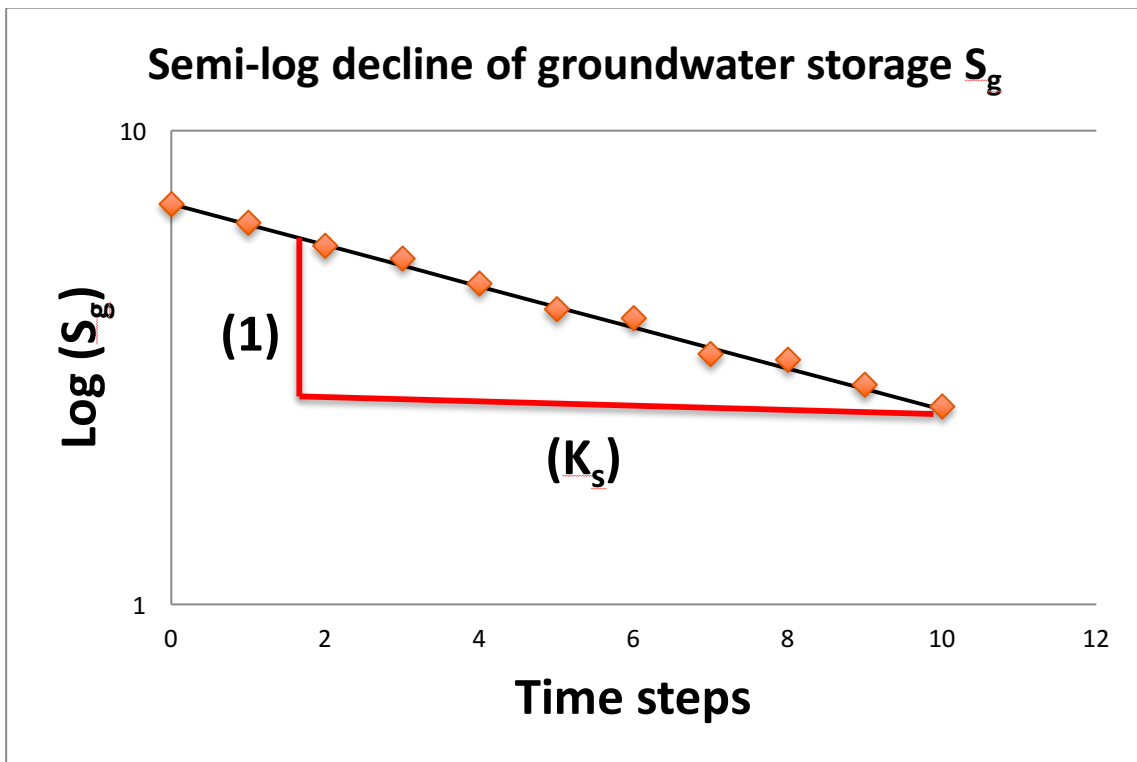


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Figure 3. Root zone storage capacity as determined by Wang Erlandsson et al. (2016)

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Figure 4. Indication of how groundwater recession can be determined by gravity observations from space.