Impacts of future deforestation and climate change on the hydrology of the Amazon basin: a multi-model analysis with a new set of land-cover change scenarios

Matthieu Guimberteau¹, Philippe Ciais¹, Agnès Ducharne², Juan Pablo Boisier³, Ana Paula Dutra Aguiar⁴, Hester Biemans⁵, Hannes De Deurwaerder⁶, David Galbraith⁷, Bart Kruijt⁵, Fanny Langerwisch⁸, German Poveda⁹, Anja Rammig^{8,10}, Daniel Andres Rodriguez¹¹, Graciela Tejada⁴, Kirsten Thonicke⁸, Celso Von Randow⁴, Rita C. S. Von Randow⁴, Ke Zhang¹²12,13,14</sup>, Hans Verbeeck⁶

Correspondence to: M. Guimberteau (matthieu.guimberteau@lsce.ipsl.fr)

Abstract. Neglecting any atmospheric feedback to precipitation, deforestation in Amazon, i.e., replacement of trees by shallow rooted short vegetation, is expected to decrease evapotranspiration (ET). Under energy-limited conditions, this process should lead to higher soil moisture and a consequent increase in river discharge. The magnitude and sign of the response of ET

¹Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-UVSQ, Université Paris-Saclay, F-91191 Gif-sur-Yvette, France

²Sorbonne Universités, UPMC, CNRS, EPHE, UMR 7619 METIS, 75252 Paris, France

³Department of Geophysics, Universidad de Chile, and Center for Climate and Resilience Research (CR2), Santiago, Chile

⁴Centro de Ciência do Sistema Terrestre (CCST), Instituto Nacional de Pesquisas Espaciais (INPE), Av dos Astronautas 1758, 12227-010, São José dos Campos, Brazil

⁵Climate Change and Adaptive Land and Water Management Group, Alterra, Wageningen University & and Research (Alterra), P.O. Box 47, 6700 AA Wageningen, Netherlands

⁶CAVElab – Computational and Applied Vegetation Ecology, Department of Applied Ecology and Environmental Biology, Faculty of Bioscience Engineering, Ghent University, Coupure Links 653, 9000 Ghent, Belgium

⁷School of Geography, University of Leeds, Leeds, United-Kingdom

⁸Earth System Analysis, Potsdam Institute for Climate Impact Research (PIK), P.O. Box 60 12 03, Telegraphenberg A62,14412 Potsdam, Germany

⁹School of Geosciences and Environment, Universidad Nacional de Colombia, Medellín, Colombia

¹⁰TUM School of Life Sciences Weihenstephan, Land Surface-Atmosphere Interactions, Technical University of Munich, Freising, Germany

¹¹Centro de Ciência do Sistema Terrestre (CCST), Instituto Nacional de Pesquisas Espaciais (INPE), Rodovia Presidente Dutra km 39, CP 01, CEP: 12630-000, Cachoeira Paulista, São Paulo, Brazil

¹²Department of Organismic and Evolutionary Biology, Harvard University, Cambridge, MA, USAState Key Laboratory of Hydrology-Water Resources and Hydraulic Engineering, and College of Hydrology and Water Resources, Hohai University, 1 Xikang Road, Nanjing, Jiangsu Province, 210098, China

¹³Cooperative Institute for Mesoscale Meteorological Studies, University of Oklahoma, Norman, OK, USA

¹⁴College of Hydrology and Water Resources, Hohai University, Nanjing, Jiangsu Province, China

to deforestation depends both on land-cover change (LCC), and on climate and CO₂ concentration changes in the future. Using three regional LCC scenarios recently established for the Brazilian and Bolivian Amazon, we investigate the combined impacts of climate change and future deforestation and climate change on the surface hydrology of the Amazon basin for this century at sub-basin scale. For each LCC scenario, three land surface models (LSMs), LPJmL-DGVM, INLAND-DGVM and ORCHIDEE, are forced by bias-corrected climate simulated by three General Circulation Models (GCMs) for different scenarios of the IPCC 4th Assessment Report (AR4). The GCM results indicate that by 2100, without deforestation, the temperature will have increased by a mean of 3.3°C (a range of 1.7 to 4.5°C) over the Amazon basin, intensifing the regional water cycle, whereby precipitation, ET and runoff increase by 8.5, 5.0 and 14%, respectively. However, under this same scenario in south-east Amazonia, precipitation decreases by 10% at the end of the dry season and the three LSMs estimate a 6% decrease of ET, which does not compensate for lower precipitation. Runoff in southeastern Amazonia decreases by 22%, reducing minimum river discharge from the Rio Tapajós catchment by 31% in 2100. The low LCC scenario projects a 7% decline in the area of Amazonian forest by 2100, as compared to 2009; for the high LCC scenario the projection is a 34% decline. In the extreme deforestation scenario, forest loss partly offsets (-2.65%) the positive effect of increase of ET with climate change on increasing ET and slightly amplifies (+3.02.2%) the increase of runoff. Effects of deforestation are more pronounced in the southern part of the Amazon basin, in particular in the Rio Madeira catchment where up to 56% of the 2009 forest area is lost,. The effect of deforestation on water budgets is more severe and at the end of the dry season in the Tapajós and the Xingu catchments, wherewhen deforestation amplifies the decrease of ET due to climate change is amplified by forest area loss. This comes with a Deforestation enhances large increase of runoff during this period (+3527%), offsetting the negative effect of climate change (-22%), and thus balancinges the decrease of low flows in the Rio Tapajós.

20 **Keywords:** Amazon basin, hydrology, land-cover change, land surface models

1 Introduction

The Amazon basin provides a range of ecosystem services. The rivers are used for navigation and hydropower; the forest is an important global sink and store of carbon, and a store of biodiversity; evaporation provides a water vapour source for rainfall downwind. When analysing changes to this ecosystem, it is important to take an integrated approach because each of these services may be affected by, or may affect, the others. Currently, two major changes are taking place simultaneously in Amazonia: deforestation and climate change. From the earlymiddle 1970s, southern Amazonia has experienced widespread deforestation (Moran, 1993) with forest being cleared to create new pasture and cropland (Fearnside, 2005). About 7.3% of the Amazon basin was deforested between 1976 and 2003 (Callède et al., 2008) and a further 2.6% between 2000 and 2010 (Song et al., 2015). At the same time, the background level of CO₂ has been rising and the climate has been changing in response (IPCC, 2013). These changes are expected to continue, to some degree, for the rest of this century.

Here, we focus on future changes to the river hydrology of the Amazon basin. For different deforestation scenarios, we model the changes in river flow from drainage-and-runoff estimated by land surface models (LSMs) driven by forcing data derived from general circulation model (GCM) output. Because of the long transit times of water moving from soil to the

mouth of the Amazon, to simulate discharge requires LSMs to be coupled to a river routing scheme (Biemans et al., 2009; Guimberteau et al., 2012; Langerwisch et al., 2013).

The climate of the Amazon basin is notoriously difficult to model and there is a wide between-GCM variation in the estimated precipitation and its changes (Boisier et al., 2015). This introduces a first level of uncertainty.

Equally, several LSMs exist and, to a greater or lesser extent, they all incorporate existing process knowledge into their parameterizations (Gash et al., 2004; Keller et al., 2009). However, because of their different structures and the values of the parameters used, LSMs also simulate a range of changes in the water and energy balances, produced by changes in vegetation and its structure, phenology and physiology. This introduces a second level of uncertainty.

A third level of uncertainty stems from the development scenarios used. Observed historical deforestation rates in Amazonia are substantially higher different than future from the rates projected in the SSPs (Shared Socio-economic Pathways) of global scenarios (Representative Concentration Pathways, RCPs) used for the last CMIP (Coupled Model Intercomparison Project) assessment. This disparity questions the realism of the globally projected rates at regional scales when considering the regional scales (Soares-Filho et al., 2006; Kay et al., 2013). On the other hand, deforestation can also be reduced by conservation policies, and none of the previous LCC (land-cover change) modelling exercises for Amazonia were able to plausibly anticipate the 84% decrease in deforestation rates in Brazil during the last decade (INPE, 2012; Dalla-Nora et al., 2014). For instance, the previous decade witnessed a drastic change in the deforestation dynamics in the Brazilian Amazon (Dalla-Nora et al., 2014). Until the beginning of the last decade, the aggressive deforestation and illegal land appropriation processes in the region (Becker, 1997; Alves, 2002; Becker, 2004) seemed to be uncontrollable, peaking at 27,772 km² yr⁻¹ in 2004 (INPE, 2016). Clear-cut deforestation rates have been decreasing since then, oscillating around 6000 km² yr⁻¹ during the last 5 years. Previous scenario modelling exercises (Laurance et al., 2001; Soares-Filho et al., 2006; Lapola et al., 2011) which have attempted to project deforestation rates for the Brazilian Amazon highly overestimated the deforestation after 2004. Recent analyses have discussed the role of commodity prices and other economic factors, such as the soy and beef moratoriums, in the slowdown of deforestation rates, although most studies have unveiled the integrated set of actions taken by the Brazilian government to curb deforestation as a decisive factor (Assunção et al., 2012; Macedo et al., 2012; Malingreau et al., 2012; Boucher et al., 2013; Nepstad et al., 2014). These measures included the creation of protected areas, the use of effective monitoring and control systems, and credit restriction mechanisms. The SSPs probably also failed to capture the trajectory of regional LCC because they do not integrate regional land management policies, existing or future road building, or the establishment of conservation areas (Dalla-Nora et al., 2014). At the same time, the future of the region remains highly uncertain, as several factors may contribute to the return of high deforestation rates, including the rapidly expanding global markets for agricultural commodities, large-scale transportation and energy infrastructure projects, and weak institutions (Aguiar et al., 2016). In this complex context, in order to better represent the current situation of the region, in this study we adopted the updated and contrasting scenarios generated in the scope of the AMAZALERT project (Raising the alert about critical feedbacks between climate and long-term land use change in the Amazon, http://www.eu-amazalert.org/home).

Here, we apply three LSMs forced by three different GCM climate projections and introduce more realistic LCC scenarios, combining these in a three dimensional matrix of model simulations to combine the effect of uncertainty in forcing data,

LSMs and development scenarios, This matrix of combinations and allowings us to estimate the magnitude of likely future hydrological changes due to deforestation and climate change and their uncertainty.

2 Materials and methods

2.1 Simulation design and models

The time frame studied includes an historical present period representing current climate conditions (1970-2008) and 21st century projections (20098-2100). Although the domain used in the simulations described below includes the whole Amazon basin (Fig. 1a and Table S1), the analysis focuses on the sub-basinscatchments sensitive to deforestation (Fig. 1b). We selected the southern sub-basinscatchments, which are subjected to a distinct dry season today, and are both sensitive to future precipitation changes (Guimberteau et al., 2013; Boisier et al., 2015) and vulnerable to future deforestation (Coe et al., 2009; Costa and Pires, 2010). These catchments are the Rio Madeira (MAD) and its upstream tributary, the Mamoré (MAM), and the two large south-eastern catchments of the Tapajós (TAP) and Xingu (XIN) (Fig. 1a). We also chose three western catchments, the Purus (PUR), Juruá (JUR) and Upper Solimões (UPSO), and the northern Rio Branco catchment (BRA). These catchments have also experienced deforestation (Nóbrega, 2012; Lima et al., 2014; Barni et al., 2015). The river discharge of the Amazon basin is taken from gauging station at Óbidos. Although this station is the closest to the mouth of the Amazon, it is upstream from the confluence of the Tapajós and Xingu with the main stem of the Amazon (Fig. 1a). The Óbidos data therefore does not contain the contribution of these rivers.

We used three LSMs, namely LPJmL-DGVM-and INLAND-DGVM, which simulates daily water budgets interactively with changes in vegetation physiology, and ORCHIDEE and INLAND-DGVM, which operates with a 30-min time-step (see Table 1 and description in supplementary material). These three models represent the state of the art LSMs inclusive of ecosystem processes controlling runoff and river routing schemes. ORCHIDEE and LPJmL-DGVM are traditionally used for ISIMIP project (The Inter-Sectoral Impact Model Intercomparison Project, https://www.isimip.org/). INLAND-DGVM has been widely tested over South American biomes to represent the biosphere-atmosphere interactions. Thus, the three LSMs are representative of the diversity of approaches to describe the functioning of the coupled system vegetation-hydrology. Moreover, two out of three models integrate different river routing schemes and are thus able to simulate the change of river discharge with climate change and in interaction with the LCC.

First, we performed an historical simulation (19701850-2008, HIST, Table 2) where we forced the LSMs with pre-industrial land cover and the Princeton global climate (Sheffield et al., 2006) at a 1°× 1° spatial resolution and 3-hourly temporal resolution (Fig. 2). This forcing is based on the National Center for Environmental Prediction-National Center for Atmospheric Research (NCEP-NCAR) 6-hourly reanalysis data sets (Kistler et al., 2001) with precipitation, air temperature and radiation biases corrected by hybridization with global monthly gridded observations. The corrected precipitation was disaggregated in space by a statistical downscaling at 1° resolution using relationships developed by the Global Precipitation Climatology Project (GPCP, Huffman et al., 2001), and in time from daily to 3-hourly using the Tropical Rainfall Measuring Mission (TRMM; Huffman et al., 2007) satellite data. A 300-year spin-up was performed by each LSM to ensure equilibrium of carbon

and water pools by recycling the Princeton forcing over the period 1970-2008 with constant pre-industrial atmospheric CO₂ concentration representative of the year 19701850 (278 ppm). Starting from the end of the spin-up state, all the LSMs did the simulation HIST over 1850-2008, forced with increasing CO₂ (from 278 ppm to 385 ppm) and by recycling the Princeton forcing too. We only kept the 39-yr period from 1970 to 2008 as the reference simulation of present conditions (HIST simulation, Table 2). derived from atmospheric observations and with observed climate change. Neither the spin-up nor the HIST runs account for LCC. Each LSM used its own definition of natural land cover and soil parameters (Table 1). The LSMs were not calibrated in present time. Their performance to simulate ET and river discharge in HIST simulation is summarized in Table S2. Other evaluations over the Amazon basin for present time can be found in Langerwisch et al. (2013) for LPJmL-DGVM, in Dias et al. (2015) and Lyra et al. (2016) for INLAND-DGVM and in Guimberteau et al. (2012, 2014) for ORCHIDEE.

Using HIST as initial conditions, multiple future simulations with each LSM forced by three GCMs (see section 2.2), were run from 2009 to 2100 (Table 2, Fig. 2) with increasing CO₂ from the SRES A2 scenario (388 ppm to 856 ppm). For the atmospheric forcing, we used three GCMs from CMIP3 (Table 3), bias-corrected and regridded to 1° as detailed in Zhang et al. (2015) and Morghim et al. (2016) (Fig. 2). First, to define the hydrological response to climate change only, we performed a future simulation with land cover set constant at the value for year 2009 for each LSM (NODEF, Table 2). Then, in order to separate the impacts of future deforestation, we prescribed to each LSM three annual LCC spatial projections (Fig. 2), resulting from the scenarios of Aguiar et al. (2016) and generated in the scope of the AMAZALERT project (Raising the alert about critical feedbacks between climate and long-term land use change in the Amazon, http://www.eu-amazalert.org/home), described in Section 2.3.

2.2 Climate change scenarios

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The projections of future climate (2009–2100) were obtained from simulations of three GCMs from CMIP3 under the SRES A2 scenario for which sub-daily outputs were available for driving LSMs, including the Parallel Climate Model (PCM), the Community Climate System Model (CCSM3), and the Hadley Centre Coupled Model (UKMO-HadCM3) (Table 3). The outputs of the three GCMs were re-gridded to 1.0° and 1-hour resolution using interpolation approaches, and corrected for biases (Zhang et al., 2015; Moghim et al., 2016). Biases in precipitation and temperature fields were corrected by applying the equidistant cumulative distribution function (EDCDF) matching method (Moghim et al., 2016). Specific humidity and downward longwave radiation were then correspondingly adjusted using the bias-corrected temperature data (Moghim et al., 2016). The method for spatial interpolation is the bilinear interpolation. For temporal disaggregation of different models' output variables, two statistical methods were used. CCSM3 generates and stores precipitation data instantaneously every 6 hours, while UKMO-HadCM3 and PCM produces 6-hourly and 3-hourly accumulated precipitation values, respectively. Therefore, two different methods are used for hourly disaggregation of precipitation obtained from these models. Linear temporal interpolation is used for disaggregation of 6-hourly instantaneous CCSM3 precipitation, while a stochastic approach developed for the disaggregation of accumulated precipitation is applied to the UKMO-HadCM3 and PCM precipitation. A method that incorporates the solar zenith angle is used to disaggregate shortwave radiations, while the other meteorological variables are interpolated linearly. Additional details of the downscaling and bias-correction methods are described in Moghim et al. (2016).

2.3 Deforestation scenarios

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The Amazon basin is located in the countries of Bolivia, Brazil, Colombia, Ecuador, Peru and Venezuela. Each country in the basin has its own socioeconomic and institutional context - specific aspects to be taken into consideration when building scenarios in order to avoid oversimplifications. Our methodological choice was to generate new updated scenarios only for Brazil and Bolivia, the most important deforestation hotspots in the basin. The Brazilian portion of the basin covers approximately 50% of the area, being also where most of the deforestation hotspots have been located in the previous decades. Bolivia has also been facing an intensive deforestation process for agricultural expansion around the Santa Cruz area. For the other countries, existing spatial projections were used.

The scenario process followed the "Story and Simulation" (SAS) approach largely adopted in environmental scenarios (Raskin, 2005; Alcamo, 2008). Two stakeholder workshops were held for discussing the whole Brazilian Amazon future, along four axes: natural resources, social development, economic activities and institutional context. The results are multi-dimensional and rich qualitative storylines. Scenario A is a "Sustainability" scenario in terms of socioeconomic, institutional and environmental dimensions, with no deforestation and massive forest restoration. Scenario B stays in the "Middle of the road", maintaining some of the positive trends of the last decade (in the case of Brazil), but not reaching the full potential from an integrated socioeconomic, institutional and environmental perspective. Finally, Scenario C is a pessimistic scenario, named "Fragmentation", consisting of a weakening of the conservation efforts of recent years, including the depletion of natural resources and the return of high deforestation rates.

To feed the spatial model, only some selected elements of the storylines were used - mainly concerning to the natural resources theme: (a) deforestation rates; (b) secondary vegetation dynamics; (c) roads and protected areas network; (d) law enforcement. The quantification process for the Brazilian Amazon is described in Aguiar et al. (2016). For the Bolivian Amazon, expert-driven premises about these same selected elements were adopted - respecting however the Bolivian socioeconomic and political specificities, as explained in Tejada et al. (2015).

These three regional LCC scenarios were generated using a qualitative/quantitative participatory approach (Aguiar et al., 2016). FBased on these elements, future maps of forest area were then simulated using the LuccME (Land use and cover change Modeling Environment, http://www.terrame.org/luccme) model framework, generating annual forest cover maps until 2100 on a grid of 25 km² for the Brazilian Amazon (Aguiar et al., 2016) and the Bolivian Amazon (Tejada et al., 2015). To generate basin-wide LCC projections, the annual spatially explicit results for Brazil and Bolivia were combined with the existing "business-as-usual" projection to the other countries, based on historical deforestation trends spatialized using the CLUE (Conversion of Land Use and its Effects) model (Verburg et al., 2002), as part of the EU-funded ROBIN (Role Of Biodiversity In climate change mitigation) project (Eupen et al., 2014).

Scenario A is a "Sustainability" scenario in terms of socioeconomic, institutional and environmental dimensions, with no deforestation and massive forest restoration. Scenario B stays in the "Middle of the road", maintaining some of the positive trends of the last decade (in the case of Bolivia the deforestation rate increased in this period), but not reaching the full potential from an integrated socioeconomic, institutional and environmental perspective. Finally, Scenario C is a pessimistic scenario,

named "Fragmentation", consisting of a weakening of the conservation efforts of recent years, including the depletion of natural resources and the return of high deforestation rates. Although the participatory process was only carried out in Brazil, similar assumptions were adopted for Bolivia (Tejada et al., 2015). Then, based on a coherent set of premises about the land-use dynamics extracted from the storylines, the scenarios were quantified using the LuceME modelling framework in order to build models for the Brazilian (LuceME/BRAmazon) and Bolivian (LuceME/Bolivia) Amazon.LuccME (Aguiar et al., 2012) is a generic framework to build land-use demand-potential-allocation models. In general terms, ILand change decisions are controlled by an allocation mechanism which uses the suitability of each cell for a given land change transition (potential of change) to distribute a given amount (demand) of change in space. LuccME allows the construction of LUCC models combining existing Demand, Potential and Allocation components according to the needs of a given application and scale of analysis. The modelling components adopted to build the LuccME/BRAmazon and LuccME/Bolivia models are based on the ideas of the Clue model (Verburg et al., 2002)work, in general lines, as follows. In the allocation model, Ccells with positive change potential will receive a percentage of the annual deforestation rate expected to be allocated to the whole area. The amount of change (i.e., new deforestation) in each cell will be proportional to the cell potential, which is recomputed, every year, considering not only the temporal changes in the spatial drivers (for instance, creation/extinction of protected areas, building/paving roads), based on the scenario premises, but also the distance to previously opened areas.

For the Brazilian Amazon, annual spatially-explicit deforestation maps from 2002-2013, provided by the PRODES (Program for the Estimation of Deforestation in the Brazilian Amazon) system (INPE, 2016), were used to calibrate and validate the parameters of the deforestation model, as detailed in Aguiar et al. (2016) (LuccME/BoliviaBRAmazon). Main drivers are related to accessibility (connection to national markets and distance to markets), protected areas and soil fertility. Using these variables, the model correctly captured the different stages of occupation of the new Amazonia frontiers from 2002 to 2013 (Aguiar et al., 2016). The scenario projections generated for the AMAZALERT project cover the period 2014 to 2100. The same premises described by Aguiar et al. (2016) were used to extend the projections from 2050 to 2100. The modelling process for the Bolivian Amazon (LuccME/Bolivia) was similar. Main spatial drivers considered in the model are connectivity to markets, distance to roads, protected areas and slope (Tejada et al., 2015). The observed deforestation data were drawn from the NKMMNH (Noel Kempff Mercado Museum of Natural History, Killeen et al., 2012) from 2001–2008. In this case, the scenario projections run from 2009-2100, adapting the premises described by Tejada et al. (2015) from 2050-2100.

To generate basin-wide LCC projections, the annual spatially explicit results for Brazil and Bolivia were combined with the existing "business-as-usual" projection (defined by the continuation of the current trend) to the other countries, based on historical deforestation trends, as part of the EU-funded ROBIN (Role Of Biodiversity In climate change mitigatioN) project (Eupen et al., 2014).

In this paper, we explore the effects of Scenario A and Scenario C contrasting storylines. Scenario A storyline quantification produced low forest loss (LODEF), whilst Scenario C was quantified into a high (HIDEF) and extreme (EXDEF) forest area loss for the Brazilian Amazon (Table 4). Fig. \$2S1 illustrates the basin-wide maps, combining the LuccME/BRAmazon, the LuccME/Bolivia and the other countries' spatial projections (Eupen et al., 2014). As Fig. \$2S1 illustrates, the results were combined into a 25 x 25 km² grid cell for the whole basin, containing annual information (from 2005 to 2100) about the

percentage of the cell area that was deforested up to that year. These three LCC scenarios (LODEF, HIDEF and EXDEF) were translated into model parameters and prescribed to each LSM (Table 2, Fig. 2) from 2009 to 2100.

2.4 Model results analysis

We selected two 20-year periods, 2040–2059 and 2080–2099, for LSM output analysis. The impact of future climate change alone was estimated for each LSM by the difference between the results of NODEF and HIST in precipitation, ET, runoff and river discharge (Table 2). The impact of future LCC was estimated by taking the difference, for each LSM and GCM-forcing, between the results of future simulations with LCC (LODEF, HIDEF and EXDEF) and without LCC (NODEF). Relative differences are calculated with the same benchmark (HIST simulation).

The spread in the ensemble mean variation (LSMs and GCM-forcings) was measured by the interquartile range (IQR). The consistency of the variations in precipitation, ET and runoff were estimated by the signs of the first (Q_1) and the last (Q_3) quartiles. A decrease (increase) was considered to be consistent if $Q_3 < 0$ ($Q_1 > 0$).

We quantify the relative contribution of GCMs, LSMs and LCC scenarios to uncertainty using an analysis of variance (ANOVA) framework as in e.g. Yip et al. (2011); Sansom et al. (2013); Giuntoli et al. (2015). ANOVA partitions a total sum of squares into portions associated with the various factors. The effect size statistic is usually called eta squared which quantifies the proportion of total variance attributed to each factor.

3 Results

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3.1 Future scenarios

3.1.1 Climate

By the end of the 21st century, GCM-mean annual temperature increases by 3.3°C in Amazonia. The forcing from the GCM chosen in this study spans the range of climate predictions for Amazonia (Malhi et al., 2009; Zhang et al., 2015). The UKMO-HadCM3 GCM is the driest and warmest model (+4.5°C), predicting Amazon rainfall reductions twice as large as any other CMIP3 GCM (Covey et al., 2003; Cox et al., 2004); PCM simulates a slightly warmer (+1.7°C) but wetter future climate compared to the current climate, and CCSM3 falls in-between (+3.6°C) (Zhang et al., 2015). The strongest warming of 6.1°C by 2100 is found in eastern Amazonia with the UKMO-HadCM3 GCM (not shown). Because of the differences in precipitation changes projected by the three GCMs (between -4.5 and +16.2%) by the end of the century, the average increase of precipitation by 8.5% (190 mm yr⁻¹) across the three GCMs should be considered as very uncertain (Fig. 3a). Precipitation changes are also spatially contrasted. Western and northern Amazonia tend to become wetter in all GCM models, with annual precipitation increases going from 6.5 to 11% (Figs 3a and 4a). In the Upper Solimões and the Branco catchments, the three GCM-forcings give an increase of precipitation. Southern Amazonia also becomes wetter, in particular the Madeira catchment where the three GCM-forcings give a 5% increase in precipitation (Fig. 3a) but with spatial differences. In the Madre de Dios region (see Fig. 1a for location), at least two out of the three GCM-forcings give a decrease in precipitation (Fig. 4a). In south-eastern

Amazonia, there is no change in GCM-mean precipitation over the Tapajós and Xingu catchments but GCM-forcings disagree on the sign and magnitude of the change (Fig. 3a). At least two out of the three GCM-forcings give a decrease of precipitation in the western part of the Tapajós catchment (Fig. 4a).

We focus on the period corresponding to the end of the dry season, from August to October (ASO). Lower precipitation during this period could have critical effects on the vegetation and hydrology (Malhi et al., 2008). In ASO, the spatial patterns of precipitation changes are similar to annual mean changes, but with a larger area of decreased precipitation (Figs. 4a and 4b). As noticed by Guimberteau et al. (2013), south-eastern Amazonia becomes drier in the middle and at the end of the century, with an average ASO precipitation decrease ranging from 10 to 14% in the Tapajós and Xingu catchments (Fig. 3b). At least two out of the three GCM-forcings give a consistent ASO precipitation decrease in most of the grid cells of these two catchments, particularly in the Xingu (Fig. 4b). In other southern regions, GCM-forcings predict wetter ASO conditions with a larger precipitation increase simulated by UKMO-HadCM3 in the southernmost part of the Madeira catchment compared with the two other GCM-forcings (Fig. 4b). In the western and northern catchments, UPSO, PUR and JUR, and the northern BRA, all the GCM-forcings predict a consistent precipitation increase during ASO of between 16 and 30% (Fig. 3b), except in the northernmost part of the Upper Solimões catchment (Fig. 4b).

15 3.1.2 Land-cover change

The total area of Amazonian forest prescribed in 2009 is 5.27 Mkm², i.e., 89% of the total area of the whole basin (Fig. 5). The LODEF scenario projects a 7% decrease in forest area over the Amazon basin by 2099 relative to 2009 (Fig. 6). By comparison, the SSP land-use scenario with the RCP8.5 emission scenario, which broadly corresponds to the SRES A2 storyline of the GCM climate forcing used in this study, gives a forest area loss of 4.6% (Fig. 5), relative to a forest area of 5.03 Mkm² in 2009. By contrast, in both HIDEF and EXDEF scenarios, forest area strongly declines during the next century. By 2100, according to the EXDEF scenario, the area of Amazonian forest is reduced to a value of 3.45 Mkm² (34%) (Figs 5 and 6), i.e., about half of the Amazon basin. LCC scenarios show a high heterogeneity at the resolution of 0.25°, reflecting how fragmentation is simulated in the LuccME model (Fig. S2). The southern catchments experience the highest deforestation rates during the 21st century in all scenarios. The Madeira catchment loses some 14% of its forest area by 2100 in LODEF, and more than 50% in HIDEF and EXDEF (Fig. 6). In the Mamoré southern sub-catchment of the Madeira, forest area loss reaches 60% in HIDEF and EXDEF (Fig. 6) with the deforested area reaching 100% in the upstream part of MAM (Fig. S2f). In the southern Tapajós and Xingu catchments, LODEF and EXDEF give contrasting estimates of forest area: in LODEF, forest area changes by only 1% in 2099 whereas in EXDEF it decreases by approximately 50% (Fig. 6). The western and northern catchments are projected to lose between 2 and 40% of their forest area, depending on the LCC scenario.

3.2 Effects of climate change on evapotranspirationET and runoff

3.2.1 Annual mean changes in ET and runoff

The 8.5% average increase of GCM-estimated annual precipitation (190 mmyr⁻¹) by the end of the century results in a 5% increase in ET (54 mmyr⁻¹) and a 14% increase in runoff (136 mmyr⁻¹) over the entire Amazon basin (Figs 3a, 3c and 3e, respectively). The ensemble spread in annual ET variation is lower (IQR = 110 mmyr⁻¹, Fig. 3c) than the change in runoff (IQR = 420 mmyr⁻¹, Fig. 3e), which is very uncertain (Fig. 3e). Western parts of the basin become wetter, and ensemble-mean ET and runoff consistently increase, by up to 6.5 and 16%, respectively, with a higher spread for runoff variation (IQR > 250 mm yr⁻¹). The largest increase of ensemble-mean runoff (19%) in the Amazon basin, occurs in the northern catchments where the increase in ET is the smallest (<2%). This increase is associated with a large spread between the multiple simulations (IQR close to 600 mmyr⁻¹). In southern parts of the basin, the 5% increase of the ensemble-mean ET is uncertain when considering the Mamoré and the south-eastern sub-catchments (Fig. 3c). Yet, in the foothills of the Andes, the northern Madeira sub-catchment and along the Amazon River (Fig. 7a), the increase of ET is consistent across the multiple simulations. In the northwestern part of the Madeira sub-catchment (MAMMamoré, Beni and Madre de Dios rivers) with an annual precipitation decrease (Fig. 4a), six out of nine simulations show decrease in ET (Fig. 7a). In the Mamoré sub-catchment, mean-ensemble runoff consistently increases by 23% whereas it decreases by 6% in the south-eastern sub-catchments. The runoff changes spread widely across all the simulations in these catchments (IQR \approx 430 mmyr⁻¹, Fig. 3e).

3.2.2 South-eastern catchments: ASO changes in ET and runoff

During ASO, the end of the dry season in the south-eastern catchments, reduced precipitation causes a consistent decrease in ET, e.g., in the Xingu catchment by up to 8% (10 mmmonth⁻¹, Fig. 3d), where at least six out of nine simulations give a consistent ET reduction in many southern grid cells (Fig. 7b). The spread between all the projections simulating ET decrease is lower for the Tapajós than for the Xingu (IQR is 2119 and 27 mm month⁻¹, respectively). Mean-ensemble runoff, already low during ASO in this region, decreases consistently by about 25% and the ensemble spread is low (IQR < 10 mm month⁻¹).

3.3 Effects of deforestation (with background climate change)

3.3.1 Annual mean changes in ET and runoff

Deforestation and climate change lead to a consistent decrease in annual ET in the Amazon basin by the end of the century, of up to 2.52.6% (30 mm yr⁻¹) with the EXDEF scenario (Fig. 8a). The resulting consistent increase of runoff is less than 3%2.2% (Fig. 9a) and both spreads of ET and runoff over the entire basin are small between the multiple forcings and LSMs used (IQR = 30 mm yr⁻¹ for runoff). With the EXDEF scenario, the loss of forest area leads to a continuous ET reduction throughout the 21st century but of different magnitude depending on the simulation type (Fig. 10a). By the end of the 21st century, ORCHIDEE simulates a 58 mm yr⁻¹ ET reduction while LPJmL-DGVM gives a decrease of 12 mm yr⁻¹ (multi GCM-forcing mean). In addition, for a given LSM, the decrease of ET differs according to the GCM-forcing used, notably in the

case of ORCHIDEE, that simulates a decrease twice as large with UKMO-HadCM3 than with PCM. In EXDEF, ET is more strongly affected in the southern and eastern regions where forest area loss is important (compare Fig. S2f and Fig. S2c). A decrease in ET by up to $\sim 150 \,\mathrm{mm\,yr^{-1}}$ is obtained in these regions, where at least six out of nine simulations show a consistent ET decrease in the EXDEF scenario (Fig. 7c). In south-eastern catchments, the strong reduction of forest area in EXDEF leads to a consistent reduction of annual ET by up to 7% in the Tapajós catchment ($\sim 80 \,\mathrm{mm\,yr^{-1}}$, Fig. 8a) and to a consistent increase in runoff by 109% (Fig. 9a). In these two catchments, the spread within the ensemble is higher than in other regions (IQR = $80 \,\mathrm{mm\,yr^{-1}}$ for runoff increase in TAPin the Tapajós catchment). In the western and northern catchments, deforestation and climate change induces a maximum consistent ET (runoff) reduction (increase) of less than 6.5% (6%) in each sub-catchment.

3.3.2 South-eastern catchments: ASO changes in ET and runoff

During ASO in the south-eastern catchments, ensemble-mean ET consistently decreases by up to 1211% and runoff increases by 35up to 27% in the EXDEF scenario (Figs 8b and 9b). ORCHIDEE and INLAND-DGVM simulate the highest ET decreases in August over the Tapajós catchment (by 23 and 12 mm month⁻¹, respectively) while ET decreases most in October in LPJmL-DGVM (-6.0 mm month⁻¹) (Fig. S3c). Deforestation reduces both the evaporation of intercepted rainfall and transpiration, by up to 45% during the wet season in ORCHIDEE (Fig. S24b), and increases soil evaporation by the same order of magnitude. In the Tapajós catchment, in EXDEF scenario, LPJmL-DGVM exhibits strong water limitations and produces nearly no evaporation of intercepted rainfall, while the bare soil evaporation increases during the wet season. Over the Xingu catchment, ET reduction starts one month later in ORCHIDEE (one month earlier in INLAND-DGVM) than in the Tapajós catchment, and seasonal variation of △ET does not change with LPJmL-DGVM (Fig. S3d). Over the Madeira catchment, both ORCHIDEE and INLAND-DGVM simulate a small decrease in ET in the EXDEF simulations during the dry season (by up to -10 mm month⁻¹ in August) while LPJmL-DGVM produces no change of ET (Fig. S3b). Changes in ET component fluxes have the same signs as in the Tapajós but smaller magnitudes (compare Fig. S1S4a and Fig. S1S4b).

3.3.3 South-eastern catchments: uncertainties due to model structure

Deforestation-induced ET variations during the dry season are driven by soil moisture changes which limit ET from dry soils (Juárez et al., 2007; Guimberteau et al., 2014). Thus, ET and runoff variations simulated by the LSMs are strongly linked to their soil hydrology and different soil moisture parameterizations, soil depths (2 m, 3 m and 4 m in the case of ORCHIDEE, LPJmL-DGVM and INLAND-DGVM, respectively) and soil texture maps (Table 1). Looking at specific model behavior, e.g. during the dry season in the Tapajós catchment for the EXDEF scenario and CCSM3 forcing, we found that deforestation decreases soil moisture in the upper layers in ORCHIDEE and INLAND-DGVM (down to 50 cm and 2 m, respectively) while deeper soil moisture increased in these two LSMs (Fig. 11a). These opposing changes of soil moisture in the soil profile are explained by the substitution of the deep-rooted forests by shallow-rooted pasture and crops in the two LSMs (see de Rosnay and Polcher (1998) for ORCHIDEE and Kucharik et al. (2000) for INLAND-DGVM). Short vegetation can only access water for transpiration from the near-surface layers. The resulting deforestation-induced ASO transpiration decrease is

higher with ORCHIDEE ($\sim 30\,\mathrm{mm\,month^{-1}}$ in August) than INLAND-DGVM ($\sim 15\,\mathrm{mm\,month^{-1}}$ in the same month). Yet, in both conditions (with or without deforestation), INLAND-DGVM simulates higher ASO transpiration (Fig. 11a). This can be explained by the higher soil water holding capacity of INLAND-DGVM which enables this LSM to carry over more water from the wet season than ORCHIDEE. This helps to sustain the evaporation during the dry season, as reported by Guimberteau et al. (2014). The simulated LAI being higher in INLAND-DGVM than in ORCHIDEE can also explain the differences between the two LSMs in simulated ET (not shown). In contrast to both ORCHIDEE and INLAND-DGVM, transpiration with LPJmL-DGVM is strongly limited by water availability nearly all year round in southern Amazonia. As a result of this background limitation even without deforestation, under the EXDEF scenario, soil moisture in the deep layers of LPJmL-DGVM decreases only slightly all year long (by $\approx 10\,\mathrm{mm}$, Fig. 11a) and transpiration does not change during the dry season.

10 3.3.4 South-eastern catchments: changes in soil moisture explained by GCM precipitation seasonality in presence of deforestation

The amplitudes of the seasonal cycle of precipitation are different between the GCM-forcings. In the UKMO-HadCM3 model, the seasonal amplitude is lower than in CCSM3 and PCM (compare Figs 11a and 11b). In southern Amazonia, the CCSM3 precipitation drops by 79% (-300 mm month⁻¹) between March (wettest month) and May (beginning of the dry season). By contrast, the precipitation drop between these two months is 60% (-180.0 mm month⁻¹) in UKMO-HadCM3. The influence of precipitation from the GCM-forcings on the response of soil moisture variation to deforestation depends on the LSM considered. As a result, soil moisture is lower all year long with LPJmL-DGVM and ORCHIDEE forced by UKMO-HadCM3 due to the dry condition of the soil, even if the deforestation reduces ET. The largest soil moisture decrease occurs with ORCHIDEE from March to June, during the beginning of the dry season. Thus, the change in transpiration simulated by this LSM is highly sensitive to the difference in precipitation changes during the wet-to-dry transition period, between CCSM3 and UKMO-HadCM3.

3.3.5 Changes in runoff and river discharge

The increase of runoff simulated over the catchments translates into an increase of river discharge through the routing schemes of ORCHIDEE and LPJmL-DGVM (INLAND-DGVM does not simulate river discharge). Because of the small effects of deforestation and climate change on the water budget of the entire Amazon basin, changes in river discharge simulated by the LPJmL-DGVM, which is already dry in regions affected by deforestation (see above), are negligible for all the sub-catchments (Fig. 12). The seasonal river discharge simulated by ORCHIDEE at Óbidos, the gauging station closest to the mouth of the Amazon basin, is little affected by deforestation and climate change, with just a slight shift in the discharge increase between December and May. This is explained by an increase in runoff at the end of the dry season in the south (MADMadeira), where river discharge slightly increases between October and April.

The discharge extremes of the southern rivers (Madeira and Tapajós) are affected by deforestation (Fig. 13). For the Madeira, the deforestation in the EXDEF scenario has an opposite effect compared to the effect of climate change on the low flows in both rivers. Namely, for the Madeira, the deforestation induced increase of low flow ranges from 5.0 to 14.53.0 to 10%,

according to different LCC scenarios and compared to a climate change induced reduction by 50% (Fig. 13a). The high-flow increase in the Madeira due to deforestation is very small (up to θ 1.5%) when compared to the climate change effect (+15%). Yet, the spread of the results in high-flow increase due to climate change is significantly reduced when taking the deforestation into account, suggesting some robustness in the simulated impacts of deforestation on high flows. In the Tapajós catchment with the largest future forest area loss, high flows do not change with deforestation (<0.5%), while climate change alone increases them by 12% but with a large spread (Fig. 13b). In contrast, with the deforestation, low flows consistently increase from 156 to 3212.5% in the 2090's, depending on the LCC scenario. These changes are of nearly samehalf-order of magnitude as the effect of climate change (-31%) suggesting that future deforestation is likely to offsetbalance the climate change impact on the hydrology of this catchment.

10 4 Discussion and synthesis

4.1 Does deforestation balance or amplify the impact of climate change on the hydrology of the Amazon basin?

Although with high uncertainties, greenhouse gas-induced climate change will probably enhance the water cycle in Amazonia, increasing annual precipitation, ET and runoff by the end of the century. The three LSMs used in this study simulate an increase of ET, despite the physiological (anti-transpirant) effect of increased CO₂ being accounted for in all of them. However, this behaviour needs to be considered with caution as it is obtained without considering atmospheric feedbacks. Considering the land-atmosphere coupling, deforestation may change precipitation recycling and thus the sign of the water balance over Amazonia (Coe et al., 2009). Consistent with Cook et al. (2012), Langerwisch et al. (2013), Guimberteau et al. (2013) and Sorribas et al. (2016), contrasted precipitation changes are projected between southern and western-northern regions. Comparing FiguresFigs. 4b and 7b, the most pronounced decrease in ET occurs during the end of the dry season, in agreement with Lejeune et al. (2014), in regions where precipitation declines. ET decreases more than precipitation over all the south-eastern catchments (TAP and XIN), i.e., land surface processes incorporated in LSMs reduce the evaporated fraction of precipitation.

It has been suggested that a reduction in the area of Amazonian forest, such as produced by the EXDEF scenario, will push much of Amazonia into a permanently drier climate regime (Malhi et al., 2008). At annual scale, deforestation-reduced ET only partly offsetsting the positive effect of climate change on ET even in EXDEF, so that all the simulations give a net increase of runoff by the end of the century. In south-eastern Amazonia, the ~50% forest area loss in EXDEF combined with climate change leads to a consistent ET decrease which offsets positive changes of ET due to climate change alone. Over the Xingu, our projections of the hydrological budget are consistent with Panday et al. (2015), who also found opposite effects of deforestation and climate during the past 40 years using a combination of long-term observations of rainfall and discharge. Yet, our settings of constant land-use ignore the influence of historic deforestation on ET and may result highly biased estimates of deforestation effects on ET.

Generally, the resulting increase of runoff after deforestation is consistent with other studies such as LCC simulations with LSMs at global scale (Sterling et al., 2013) and LCC experiments on watersheds in Northern Appalachians where water yields increase after intensive cuttings (Rothacher, 1970; Hornbeck et al., 2014). The resulting-increase of annual runoff in the Xingu

catchment (+98%), with increasing deforestation in the future, is of similar order to the results of Stickler et al. (2013), who found a 10 to 12% runoff increase given 40% deforestation in this catchment. Yet, during ASO in the south-eastern catchments, deforestation amplifies the effect of climate change in reducing ET, in particular in the south of the Tapajós catchment and in the north of the Madeira and Xingu catchments where deforested areas are the largest. Thus, deforestation contributes to the increase in runoff (+3527% in TAPthe Tapajós), and thus balances the runoff-reducing effect of climate change (-22% in the Tapajós).

4.2 Consequences on the extreme discharge of the southern rivers

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The ET decrease and runoff increase projected for southern catchments (MADMadeira and the TAPTapajós) by the extreme deforestation scenario applied here (EXDEF) balances the climate change effect on low flows. Climate change alone increases the seasonal amplitude of discharge, and high-flow values. In contrast, deforestation balances this effect by reducing the risk of decrease in low flows in the Madeira and Tapajós in all LCC scenarios; this is related to the decrease of ET during the dry season. The low-flow increase of the Tapajós is consistent with higher future discharge during the dry season, in the Jamanxim sub-catchment (lower Tapajós) (Lamparter et al., 2016). Our result for the Madeira contradicts those of Siqueira Júnior et al. (2015) who found a decrease of low flows with the hydrological model MHD-INPE, combined with the «business-as-usual» scenario (BAU) of Soares-Filho et al. (2006) where deforestation is lower than in the EXDEF scenario. They argue that this behaviour is due to the occurrence of faster flows when deforestation is taken into account; even though this contradicts the fact that LCC scenarios are associated with reduced ET. This comparison highlights the uncertainty in the results of the effect of deforestation on hydrology, depending on whether we use LSMs or hydrological models to simulate river discharge.

4.3 Are the uncertainties of the model responses to deforestation and climate change greater than those from GCM projections? Contributions of GCMs, LSMs and land cover change scenarios to total uncertainty in ET and runoff projections

ET and runoff projections are associated to uncertainties that originate from the GCMs, LSMs, and the LCC scenarios. Using ANOVA, we fould that main uncertainty source is different for ET (Fig. 14a) and runoff (Fig. 14b). LCC provides the largest uncertainty for ET projections over the Amazon basin (38% of total uncertainty in the basin), in particular in the southern catchments (9 and 14% in the Tapajós and Madeira, respectively) and in the Purus and Jurua (10%). The GCM uncertainty is low over the Amazon basin (3%). Yet, significant interactions between GCMs and LCC scenarios (13%) occurs over the Amazon basin, particularly in the western catchments, suggesting that these contributions do not behave linearly. Our results confirm the Large climate change uncertainty, shown in previous studies (Li et al., 2006; Vera et al., 2006; Torres and Marengo, 2013), occurs in the southern water-limited regions of the Madeira, Xingu and Tapajós catchments (between 38 and 83%) with the three GCMs giving different signs of precipitation changes, in particular in the southern water-limited regions of the Xingu and Tapajós catchments. These large uncertainties are also found for runoff projections but in larger proportion in the Madeira (91%) and Mamoré (77%) catchments. These catchments, where the sign of the variation of ET and runoff due to climate change alone is particularly uncertain, due to the large spread in GCM rainfall projections. We showed that

The impact of deforestation combined with climate change on soil moisture and ET also depends on the GCM forcing, Yet, The magnitude of the ET changes due to climate change alone is more uncertain than that induced by the deforestation in the three LSMs assessed. During the wet-to-dry transition period, the strength of the precipitation decrease driven by climate change determines the change of soil water storage due to deforestation which sustains the evaporation during the dry season. Uncertainty in the signal of changes in runoff attributable to LCC only occurs in the Branco catchment (22%) while it is largely attributable to GCMs in southern (up to 91% in the Madeira) and western (73% in the Upper Solimoes) catchments. The overall uncertainty in runoff change over the Amazon basin is dominated by LSMs contribution (53%), particularly in the Xingu and Tapajós catchments (60%), suggesting the difficulty of these models to simulate contrasted variations in runoff that naturally occurs during the year in these catchments. Moreover, uncertainty contribution from LSMs is high for runoff projections in the Purus, Jurua and Branco catchments (between 30 and 45%). Thus, \(\theta\)our study emphasizes the uncertainty associated with the choice of the LSMs to be paired with the GCM forcing, and the effect this choice has on the and their inherent uncertainties the model's parameterizations (energy or water-limited), produces onin the estimation of estimating deforestation impacts on hydrology. Over large river basins like the Amazon, these models have the disadvantage of being rather poorly constrained in their parameterizations of both vegetation functioning (Poulter et al., 2010) and soil hydrology (Christoffersen et al., 2014). In our view, the LSM community needs to strengthen its efforts to cooperate with the soil science community to improve the representation of soil hydrological processes in their models, despite the difference in scale at which they work and the inherent small-scale variability of soil properties. Nevertheless, the uncertainty among the models is lower than among climate projections.

The magnitude of the changes in soil hydrologyET first depends on GCMs-precipitation changes, in particular during the beginning of the dry season and then on the behaviour of each LSM (water versus energy-limited models) in the southern catchments. Uncertainty in runoff changes in the Amazon basin is first attributable to LSMs and then to GCMs, particularly in the south-eastern catchments. The uncertainty attributable to LCC is low in these catchments suggesting some robustness in the response of the hydrology to the deforestation.

5 Conclusions

The construction of new land-cover change scenarios for Amazonia indicates that, by the end of this century, the total forested area of the Amazon basin will have decreased by 7% in the best case, to 34% in the most severe scenario. The most severe forest clearing occurs in southern Amazonia where the Madeira, Xingu and Tapajós catchments experience a 50% decrease in forest area. With a multi-model approach, we found that the replacement of the forests by pasture and crops should only slightly decrease annual evapotranspiration by up to 2.5% and enhance runoff by up to 3.02.2%, for the most severe scenario of the Amazon basin, compared to simulations with climate change only.

The south-eastern catchments, however, are more vulnerable at the end of the dry season. Compared to forest, crops and pastures fail to sustain their evaporation in a high drought stress context. Given the combination of decreased rainfall due to future climate change and the large forest area loss, evapotranspiration may drop by -109 and -1211% in the Xingu and Tapajós

catchments, respectively, deforestation amplifying the decrease of ET due to climate change alone. The dominant uncertainty associated to Thesethese results strongly depend oncomes from the climate change scenarios. the land surface model used, they vary with the soil hydrology parameterizations; but they also depend on the climate model forcing data. In contrast, by enhancing the runoff, the deforestation balances the negative effect of climate change on runoff in these catchments. As a result, the deforestation in the most intensive scenario balances the risk of decrease in low flows of the Tapajós due to climate change by the end of the century. In these catchments, LSMs are the largest uncertainty source for runoff projections, while the climate change uncertainty dominates in the southern catchments (Madeira and Mamoré). Our results in the Tapajós catchment emphasize the impact of deforestation combined with climate change on hydrological extremes. The deforestation leads to a 3212.5% increase in low flows by the end of the century, which offsetsbalances the opposite impact of climate change.

Biosphere-atmosphere interactions, not accounted for in our study, are also crucial in estimating the progress of forest dieback, whereby forest is replaced by savanna vegetation. During the end of the dry season, we found a strong reduction of ET in south-eastern Amazonia. Evaporation at this time of year provides a critical source of water vapour for precipitation and lower ET can delay the onset of the wet season (Fu and Li, 2004) and reduce the water recycling during this period (Lima et al., 2014). We need to pay careful attention to the intensification and lengthening of droughts during this century; a phenomenon that is commonly predicted by the GCMs for southern Amazonia (Boisier et al., 2015). Whatever its cause, our results emphasize the need to include the deforestation process in climate change simulations. Deforestation has the potential to mask (or unmask) the effects of climate change on surface hydrology.

Competing interests

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The authors declare that they have no conflict of interest.

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Model	Institute	Reference	Model setup	Resolution	
ORCHIDEE ¹	IPSL, Paris, France	Krinner et al. (2005)	River routing including floodplains/swamps,	1.0°, half-hourly	
			FAO soil texture map		
INLAND-DGVM ²	INPE, Sao Jose dos Campos, Brazil	Foley et al. (1996)	Fire emissions, soil texture map No river routing	1.0°, half-hourly	
		Kucharik et al. (2000)			
LPJmL-DGVM ³	PIK, Potsdam, Germany	Sitch et al., 2003	Fire emissions, rRiver routing, FAO soil texture map	1.0°, daily	

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Lund Potsdam Jena managed Land model

Table 1. Models used in this study

Name of the	Scenarios	
simulation	CC	LCC
HIST	no	no
CCSM3 NODEF	yes	no
CCSM3 LODEF	yes	yes
CCSM3 HIDEF	yes	yes
CCSM3 EXDEF	yes	yes
UKMO-HadCM3 NODEF	yes	no
UKMO-HadCM3 LODEF	yes	yes
UKMO-HadCM3 HIDEF	yes	yes
UKMO-HadCM3 EXDEF	yes	yes
PCM NODEF	yes	no
PCM LODEF	yes	yes
PCM HIDEF	yes	yes
PCM EXDEF	yes	yes

Table 2. List of the different simulations performed with the three LSMs (ORCHIDEE, INLAND-DGVM and LPJmL-DGVM) with or without climate change (CC) and land-cover change (LCC).

Institutes, country	Model		Resolution	References
	Name	Acronym	(lat x lon)	
National Control for Atmospheric Descent (NCAD) 110A	Community Climate System Model	CCSM3	~1.4° x 1.4°	Bonan et al. (2002)
National Center for AtmosphericResearch (NCAR), USA	Parallel Climate Model	PCM	~2.8° x 2.8°	Washington et al. (2000)
Hadley Centre for Climate Prediction	W. W. G. & G. 11W.11	UKMO-HadCM3	~2.5° x 3.75°	Gordon et al. (2000)
and Research / Met Office, UK	Hadley Centre Coupled Model			

Table 3. List of the GCMs participating in CMIP3 used in this study with their approximate atmospheric horizontal resolution.

Qualitative scenarios			Quantification of deforestation rates (spatially explicit projections until 2100)				
N	Brief storyline	N	Brazilian Amazon	Bolivian Amazon	Other countries		
Name		Name	(Aguiar et al., 2016)	(Tejada et al., 2015)	(Eupen et al., 2014		
A-Sustainability	"Zero" deforestation scenario.		Annual rate decreasing	Trend of 2005–2008	Same as HIDEF		
	Sustainable land use,		to $3,900 \text{ km}^2 \text{ yr}^{-1}$	until 2013,			
	protected areas,	LODEE	until 2020, and then	then decrease by 50%.			
	indigenous territories,	LODEF	to $1,000 \text{ km}^2 \text{ yr}^{-1}$				
	restrained construction		until 2025, and then				
	of new roads.		stabilizing until 2100.				
			Annual rate increasing	Total deforested area reaches	For each country,		
C-Fragmentation		HIDEF	to 15,000 km ² yr ⁻¹	13 million in 2025 ha,	projected		
			until 2020	then replicates the 2005-2008	according to		
	Return of high		and stabilizing until 2100.	annual rate.	historical trends.		
	deforestation						
	rates	EXDEF	Annual rate increasing	Same as HIDEF	Same as HIDEF		
			to 19,500 km ² yr ⁻¹				
			(1996-2005 historical rate)				
			until 2020 and stabilizing until 2100.				

Table 4. LCC scenarios used in this study

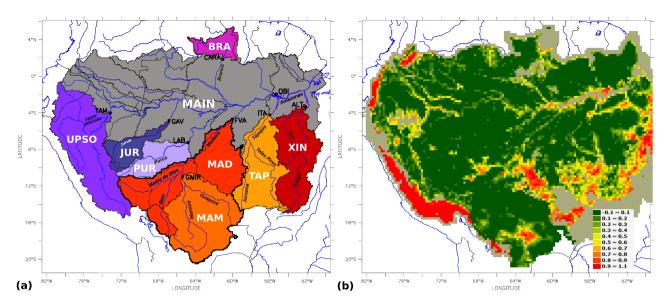


Figure 1. (a) Amazon catchments (name in white, the abbreviations are indicated in Table S1) and the main rivers (adapted from Guimberteau et al., 2012). Localization of the main SO HYBAM gauging stations (the abbreviations are indicated in Table S1). The bold black line delineates the Madeira catchment. Colour is used to distinguish the southern (red), western (purple) and northern sub-catchments (pink). Topographic scale is indicated. (b) Percentage of deforestation in each 25 x 25 km² in 2005 (observed data, Aguiar et al., 2016).

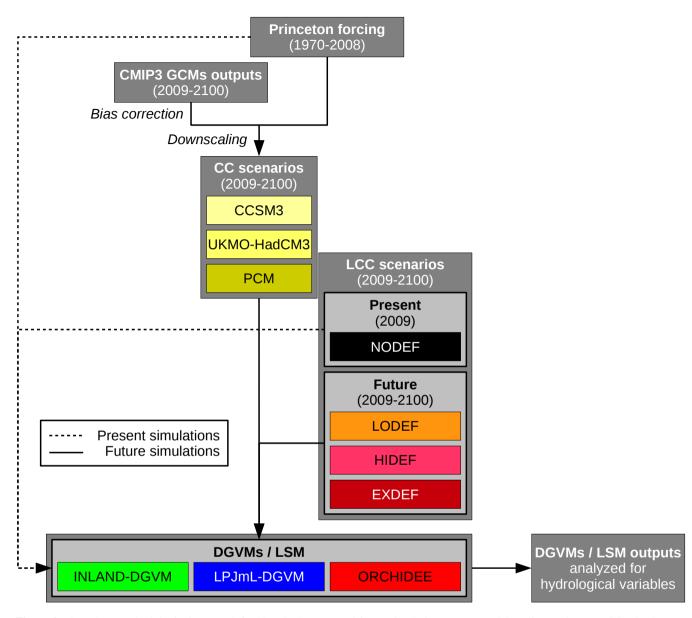


Figure 2. Flow chart methodological approach for historical present and future simulation processes (CC = climate change, LCC = land-cover change). Acronyms of the LCC scenarios are explained in Table 4.

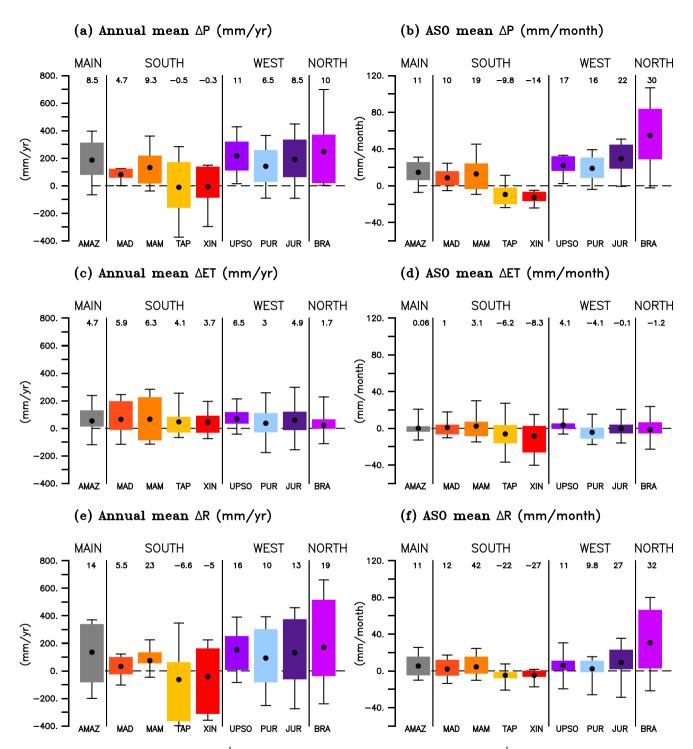


Figure 3. Changes in mean annual $(mmyr^{-1})$ and August to October (ASO)ASO values $(mmmonth^{-1})$ of (a and b) precipitation, (c and d) ET and (e and f) runoff due to climate change only, for the end of the century, over the Amazon basin and eight of its sub-basin catchments (the abbreviations of the sub-basin catchments are indicated in Table S1). Each box plot corresponds to the interquartile range (IQR, distance between the 25^{th} and the 75^{th} percentiles) within each basin catchments indicating the spread of the 3 GCM-forcings (for ΔP) and 3 GCM-forcings x 3 LSMs (for ΔET and ΔR) results (see FigureFig. 1 for colour code). For a given box plot, the black points denote the mean value over the basin catchment, the whiskers extend from the minimum value to the maximum one and the numbers above the box plot indicate the mean relative differences over the basin catchment (%).

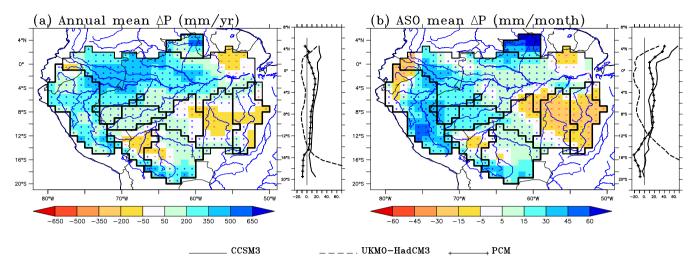


Figure 4. Maps: Spatial change in (a) annual (mmyr⁻¹) and (b) ASO (mmmonth⁻¹) precipitation (mmmonth⁻¹) due to climate change (mean of the three GCM-forcings), for the end of the century. The symbols indicate that more than two GCM-forcings out of three give a precipitation increase (+) or decrease (-) on the grid cell. The black lines delineate the Amazon basin and the sub-basin catchments. **Plots**: Corresponding zonal mean of relative changes in precipitation (%) from the three GCM-forcings.

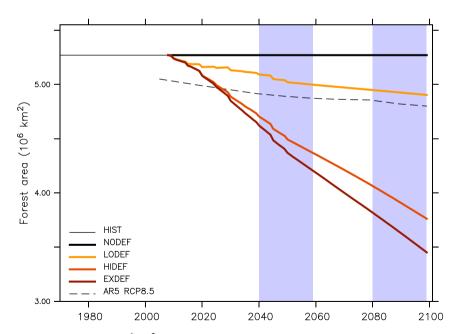


Figure 5. Interannual variation of forest area (10⁶ km²) over the Amazon basin, according to the NODEF scenarios, the three LCC scenarios and the SSP of the AR5 RCP8.5 scenario over 2009-2099. The blue bands indicate the two future periods selected for this study.

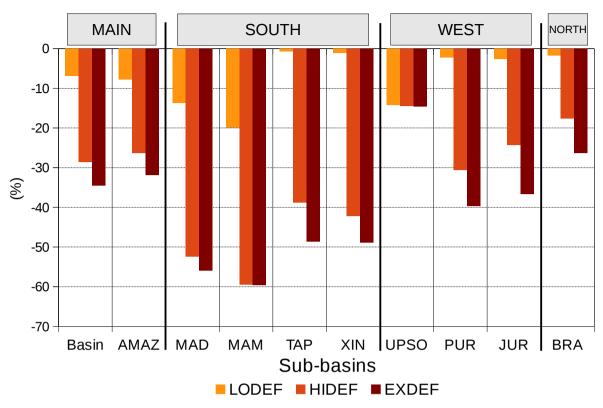


Figure 6. Forest area decrease (%) over the different sub-basin catchments of the Amazon basin between each of the three LCC scenarios in 2099 and the NODEF scenario in 2009 (the abbreviations of the sub-basin catchments are indicated in Sect. 2.1 and Table S1).

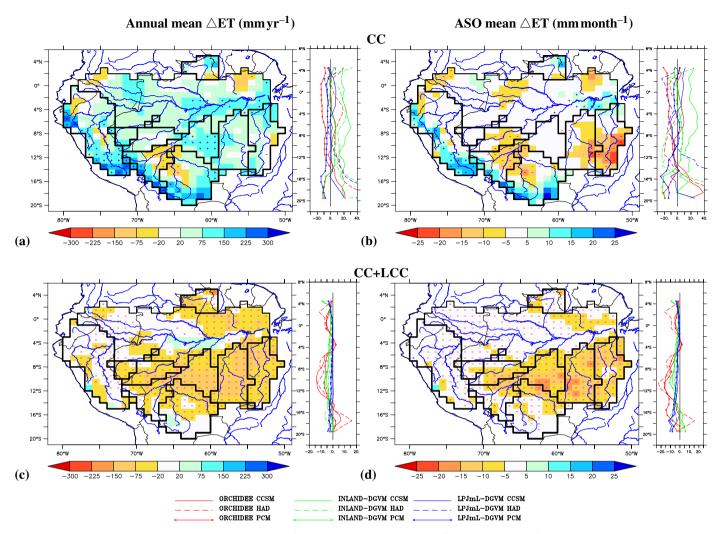


Figure 7. Maps: Spatial change in (**a,c**) annual (mmyr⁻¹) and (**b,d**) ASO (mm month⁻¹) ET (mmmonth⁻¹) due to (**a,b**) climate change (mean of the three GCM-forcings) and (**c,d**) deforestation combined with climate change (EXDEF), for the end of the century. The symbols indicate that more than six simulations out of nine (3 GCM-forcings x 3 LSMs) give an increase (+) or a decrease (-) of ET on the grid cell. The black lines delineate the Amazon basin and the sub-basin catchments. **Plots**: Corresponding zonal mean of relative changes in ET (%) from each of the nine simulations.

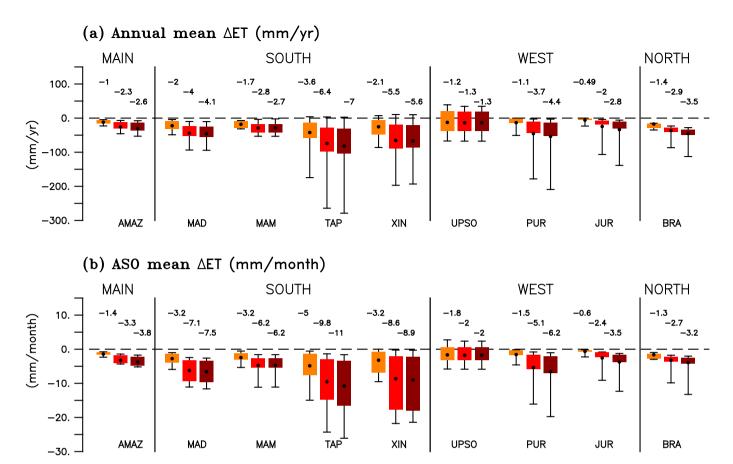


Figure 8. (a) Mean annual (mmyr⁻¹) and (b) ASO values (mmmonth⁻¹) of ET changes due to deforestation and assuming future climate change, for the end of the century, over the Amazon basin and eight of its sub-basin-catchments (the abbreviations of the basin-catchments are indicated in Table S1). Each box plot corresponds to the interquartile range (IQR, distance between the 25th and the 75th percentiles) within each basin-catchment indicating the spread of the 3 GCM-forcings x 3 LSMs results for one LCC scenario (see FigureFig. 6 for colour code). For a given box plot, the black points denote the mean value over the basin-catchment, the whiskers extend from the minimum value to the maximum one and the numbers above the box plot indicate the mean relative differences over the basin-catchment (%).

(a) Annual mean $\triangle R$ (mm/yr) MAIN SOUTH WEST NORTH 300. 0.97 0.61 2.5 7.7 7.9 1.7 1.6 5.1 3.6 2.2 200. (mm/yr) 100. AMAZ UPS0 PUR MAD MAM TAP JUR BRA XIN

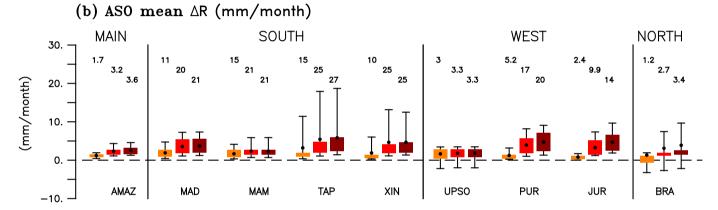


Figure 9. As in Figure Fig. 8 but for runoff.

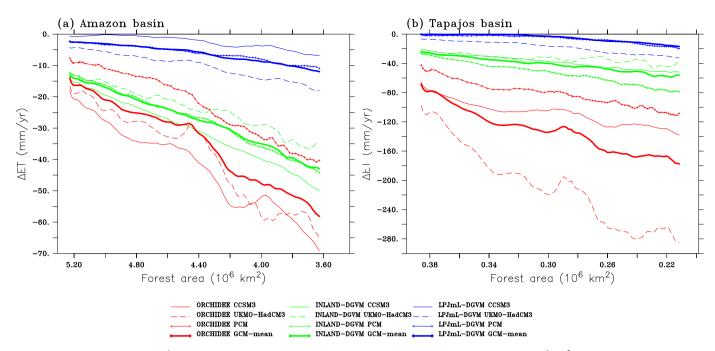


Figure 10. ET changes (mm yr⁻¹) over 2009-2100 (1-year running mean) as a function of tree area decrease (10^6 km^2) from scenario EXDEF within the (a) Amazon and (b) Tapajós basins catchments.

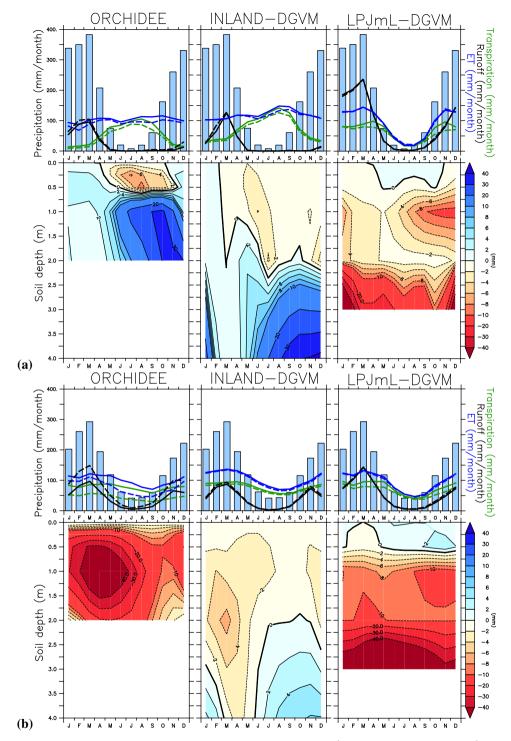


Figure 11. Impact of deforestation combined with climate change on ET (mm month⁻¹), transpiration (mm month⁻¹), runoff (mm month⁻¹) and soil moisture (mm) over the Tapajós catchment for the three LSMs, for two different end of century climates: (a) CCSM3 and (b) UKMO-HadCM3. **Top panels**: seasonal cycle of precipitation (mm month⁻¹) with colour bars and seasonal cycles of ET, transpiration and runoff with plain/dashed lines for NODEF/EXDEF LCC scenarios. **Bottom panels**: Corresponding Asoil moisture (mm) due to deforestation combined with climate change. Results for PCM forcing are similar to those for CCSM3 forcing (not shown).

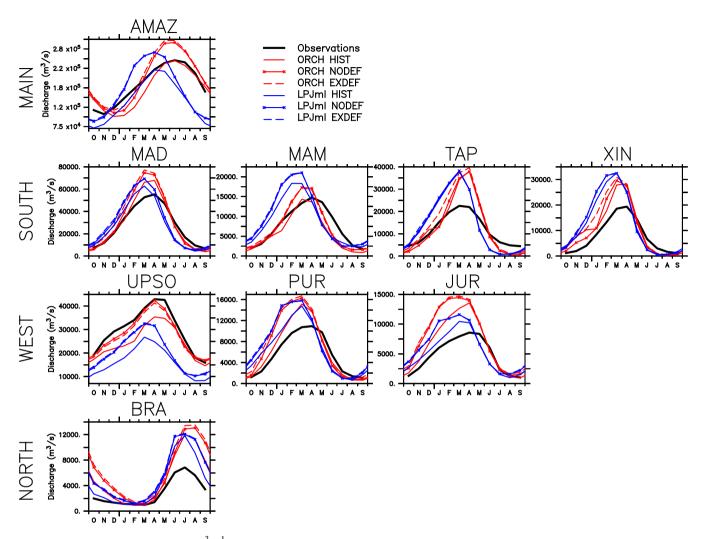


Figure 12. Seasonal river discharge (m³ s⁻¹) simulated by ORCHIDEE and LPJmL-DGVM from HIST (averaged over 1970-2008) and from NODEF and EXDEF LCC scenarios (mean of the three GCM-forcings for each scenario, averaged over 2080-2099) at the gauging stations over the Amazon sub-basin catchments (the abbreviations of the catchments are indicated in Table S1). The results from HIST simulations are compared with the observations from the SO HYBAM (averaged over 1970-2008). River discharge is not represented for INLAND-DGVM because it does not include a routing scheme.

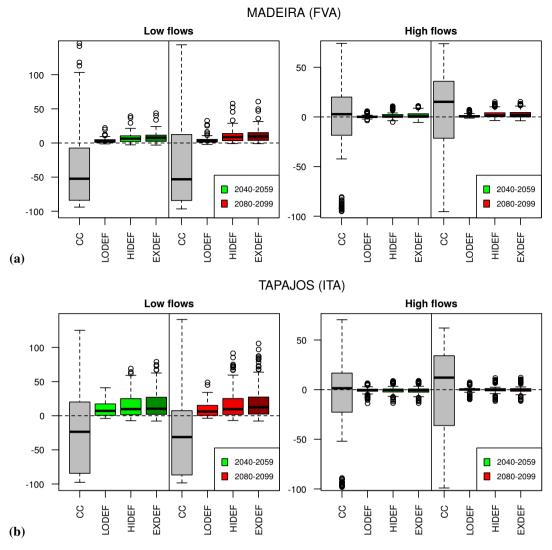


Figure 13. Relative change (%) of the first deciles (i.e., low flow, left panels) and the last deciles (i.e., high flow, right panels) of river discharge due to climate change (grey) and deforestation combined to climate change (three LCC scenarios) of (a) the Madeira (at FVA) and (b) the Tapajós (at ITA), for the middle (green) and the end (red) of the century. The changes are simulated by the ensemble of six simulations (2 LSMs x 3 GCM-forcings). The change of low flows and high flows cannot be represented with INLAND-DGVM because it does not include a routing scheme. The boxes correspond to the interquartile range (IQR, the distance between the 25th and the 75th percentiles), the bold horizontal line in each box is the median and the whiskers extend from the minimum value to the maximum value unless the distance from the minimum (maximum) value to the first (third) quartile is more than 1.5 times the IQR. Circles indicate the outliers that are 1.5*IQR below (above) the 25th percentile (75th percentile).

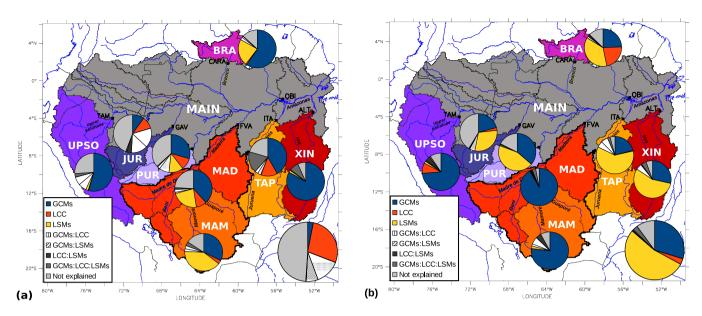


Figure 14. Contributions of GCMs, LCC scenarios, LSMs and interactions between each to total uncertainty in (a) ET and (b) runoff changes, for the end of the century, for the Amazon basin (bottom right pie chart) and eight of its catchments.

Impacts of future deforestation and climate change on the hydrology of the Amazon basin: a multi-model analysis with a new set of land-cover change scenarios

Matthieu Guimberteau¹, Philippe Ciais¹, Agnès Ducharne², Juan Pablo Boisier³, Ana Paula Dutra Aguiar⁴, Hester Biemans⁵, Hannes De Deurwaerder⁶, David Galbraith⁷, Bart Kruijt⁵, Fanny Langerwisch⁸, German Poveda⁹, Anja Rammig^{8,10}, Daniel Andres Rodriguez¹¹, Graciela Tejada⁴, Kirsten Thonicke⁸, Celso Von Randow⁴, Rita C. S. Von Randow⁴, Ke Zhang¹²12,13,14, Hans Verbeeck⁶

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SUPPLEMENTARY MATERIAL

Models

LPJmL-DGVM (Lund Potsdam Jena managed Land model)

The process-based dynamic global vegetation and hydrology model LPJmL-DGVM calculates carbon and the corresponding water fluxes with a daily time step and a spatial resolution of 0.5 x 0.5 (lat/lon) (Sitch et al., 2003; Gerten et al., 2004b; Bondeau et al., 2007; Rost et al., 2008; Schaphoff et al., 2013). Potential natural vegetation

¹Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-UVSQ, Université Paris-Saclay, F-91191 Gif-sur-Yvette, France

²Sorbonne Universités, UPMC, CNRS, EPHE, UMR 7619 METIS, 75252 Paris, France

³Department of Geophysics, Universidad de Chile, and Center for Climate and Resilience Research (CR2), Santiago, Chile ⁴Centro de Ciência do Sistema Terrestre (CCST), Instituto Nacional de Pesquisas Espaciais (INPE), Av dos Astronautas 1758, 12227-010, São José dos Campos, Brazil

⁵Climate Change and Adaptive Land and Water Management Group, Alterra, Wageningen University & and Research (Alterra), P.O. Box 47, 6700 AA-Wageningen, Netherlands

⁶CAVElab – Computational and Applied Vegetation Ecology, Department of Applied Ecology and Environmental Biology, Faculty of Bioscience Engineering, Ghent University, Coupure Links 653, 9000 Ghent, Belgium

⁷School of Geography, University of Leeds, Leeds, United-Kingdom

⁸Earth System Analysis, Potsdam Institute for Climate Impact Research (PIK), P.O. Box 60 12 03, Telegraphenberg A62.14412 Potsdam. Germany

⁹School of Geosciences and Environment, Universidad Nacional de Colombia, Medellín, Colombia

¹⁰TUM School of Life Sciences Weihenstephan, Land Surface-Atmosphere Interactions, Technical University of Munich, Freising, Germany

¹¹Centro de Ciência do Sistema Terrestre (CCST), Instituto Nacional de Pesquisas Espaciais (INPE), Rodovia Presidente Dutra km 39, CP 01.CEP: 12630-000, Cachoeira Paulista, São Paulo, Brazil

¹² Department of Organismic and Evolutionary Biology, Harvard University, Cambridge, MA, USAS tate Key Laboratory of Hydrology-Water Resources and Hydraulic Engineering, and College of Hydrology and Water Resources, Hohai University, 1 Xikang Road, Nanjing, Jiangsu Province, 210098, China

¹³Cooperative Institute for Mesoscale Meteorological Studies, University of Oklahoma, Norman, OK, USA

¹⁴College of Hydrology and Water Resources, Hohai University, Nanjing, Jiangsu Province, China

and the main processes controlling its dynamics are calculated from inputs of climate data (temperature, precipitation and cloud cover), atmospheric CO2, and soil texture. The main processes included in LPJmL-DGVM are photosynthesisthe water balance, carbon balance, vegetation (based on Farquhar et al., 1980 and Collatz et al., 1992), auto- and heterotrophic respiration, establishment, phenology, mortality and phenology fire disturbance, These processes lead to the dynamics of carbon stored in the vegetation, litter and soil. The daily water balance of the soil is calculated by a simple bucket model, consisting of 5 soil layers of 20 cm, 30 cm, 50 cm, 1 m and 1 m depth, resulting in a cumulative depth of 3m. Water from precipitation that is not intercepted by vegetation enters the first soil layer depending on the amount of rainfall and the water saturation of the soil layer. The water that enters the first soil layer either evaporates, transpires or percolates to deeper soil layers. Evaporation from the canopy depends on the intercept water and the leaf area index of the vegetation. Evaporation from soil only occurs on bare soil and depends on the energy available for vaporization (potential evapotranspiration, PET) Plant transpiration is closely coupled to stomatal activity and photosynthesis and is calculated as a function of soil water supply and atmospheric demand (Sitch et al., 2003). All excess water above field capacity runs off as surface or subsurface runoff. The water is simulated to percolate from the first layer through the deeper soil layers based on a storage routine technique (Schaphoff et al., 2013) and is added to the runoff as baseflow component (Gerten et al., 2004b). The runoff is routed through a gridded river network (Vörösmarty et al., 2000), with a constant flow velocity of 1 m s⁻¹ (Rost et al., 2008). Human processes like irrigation extraction and the operation of large reservoirs is explicitly accounted for (Rost et al., 2008; Biemans et al., 2011). Simulated water fluxes include evaporation, soil moisture, snowmelt, runoff, discharge, interception and transpiration. The carbon balance includes a detailed simulation of photosynthesis (based on Farquhar et al. (1980) and Collatz et al. (1992)), autotrophic and heterotrophic respiration, allocation of carbon to the plant compartments, establishment, mortality and phenology (Sitch et al., 2003). These processes are in LPJmL-DGVM calculateds for the performance of nine plant functional types (PFTs) representing natural vegetation for in each grid cell. Eeach of these PFT representing an assortment of species classified as being functionally similar. However, in the Amazon basin, LPJmL primarily simulates three of these plant functional types, representing tropical evergreen and deciduous forest and C4 grasses. In this study for the Amazon basin, LPJmL-DGVM primarily simulates three of these plant functional types, representing tropical evergreen and deciduous forests and C4 grasses. LPJmL-DGVM also includes crop growth and harvest of so-called crop functional types on managed land as well as managed grassland (Bondeau et al., 2007). LPJmL-DGVM has been provened to reproduce observed patterns of biomass production, the global water balance, river discharge, : it also includes managed landtropical vegetation dynamics and fire (Cramer et al., 2001; Sitch et al., 2003; Wagner et al., 2003; Gerten et al., 2004a, 2008; Rost et al., 2008; Biemans et al., 2009; Poulter et al., 2009; Fader et al., 2010; Thonicke et al., 2010). It has been shown that the observed patterns in water fluxes (including soil moisture, evapotranspiration and runoff) are comparable to stand-alone global hydrological models (Wagner et al., 2003; Gerten et al., 2004a; Gordon et al., 2004; Gerten et al., 2008; Biemans et al., 2009; Haddeland et al., 2011). Several studies on Amazonia have been conducted showing the effect of climate change on NPP (Poulter et al., 2009), on carbon stocks (Gumpenberger et al., 2010), on the risk for forest dieback (Rammig et al., 2010) and also on riverine-related changes such as inundation patterns on patterns of inundation duration and inundated area (Langerwisch et al., 2013).

INLAND-DGVM (INtegrated model of LAND surface processes)

INLAND (Foley et al., 1996; Kucharik et al., 2000) is the land surface module of the Brazilian Earth System Model (BESM), and represents virtually all relevant aspects of the land surface to the climate system. BESM is a world-class global coupled model of the climate system currently being developed within Brazil's Climate Change Program that includes modules of atmospheric and ocean general circulation, the terrestrial and marine biosphere, cryosphere, carbon cycles, and aerosols. INLAND simulates 12 different PFTs competing for available resources

within the grid cell and the relative success of each PFT determines its fractional coverage. The model allows trees and herbaceous plants or grasses to experience different light and water availability: while trees in the upper canopy have priority in capturing available light (thus shading the shrubs and grasses in the lower part of the canopy), the herbaceous plants are able to capture soil water first when it infiltrates the ground (Foley et al., 1996). INLAND uses the mechanistic treatment of canopy photosynthesis proposed by Farquhar et al. (1980) and the semi-mechanistic Ball-Berry approach to estimate stomatal conductance (Ball et al., 1987; Collatz et al., 1992), computing gross photosynthesis, maintenance respiration and growth respiration to yield the annual carbon balance for each PFT, and the vegetation dynamics module simulates biomass changes for each PFT on a yearly time step. The model uses specific soil water-stress functions to down-regulate the gross primary productivity of vegetation as soils dry.

INLAND-DGVM is premised to be single, physically consistent model that solves the energy, water, carbon, and momentum balance of the soil-vegetation-atmosphere system and can be directly incorporated within Atmospheric Climate models. Based on the LSX package of Thompson and Pollard (1995), it represents canopy and soil physics processes by explicitly diagnosing the temperature of the vegetation in two canopy layers (e.g. trees versus shrubs and grasses) and of its soil layers, as well as air temperature and specific humidity within canopy air spaces, driven by the radiation balance of the vegetation and the ground, and the diffusive and turbulent fluxes of sensible heat and water vapor. In order to resolve the diurnal cycle, the model solves the canopy physics at its shortest time step (depending on the user choice, usually 30-60 min). The total amount of evapotranspiration is treated as the sum of three water vapor fluxes: evaporation from the soil, evaporation of water intercepted by the vegetation and canopy transpiration.

The model state description includes 6 soil layers with varying thicknesses (to simulate the diurnal and seasonal variations of heat and moisture in the total soil depth) that are parameterized with biome-specific root biomass distributions of Jackson et al. (1996). This permits a different root length density for each layer in the profile.

The dynamics of soil volumetric water content are simulated for each layer. Soil moisture is based on Richards' flow equation, where the soil moisture change in time and space is a function of soil hydraulic conductivity, soil water retention curve, plant water uptake, and upper and lower boundary conditions. The water budget is controlled by the rate of infiltration (Green and Ampt, 1911), evaporation of water from the soil surface, the transpiration stream originating from plants, and redistribution of water in the profile. The modeling of water flow in unsaturated soils requires the description of water uptake by plant roots. Water uptake by roots is represented by a sink term in the macroscopic Richards equation and only considers stress due to dry conditions through a simple heuristic approach that represents the influence of soil water stress on gross photosynthesis rates (Foley et al., 1996). The drainage from the bottom soil layer is modeled assuming gravity drainage and neglects interactions with groundwater aquifers. Foley et al. (1996); Kucharik et al. (2000) give additional descriptions of the IBIS model land surface physics, which is essentially transferred unaltered to INLAND-DGVM.

ORCHIDEE (ORganising Carbon and Hydrology In Dynamic EcosystEms)

ORCHIDEE (Krinner et al., 2005) is the land component of the IPSL (Institut Pierre Simon Laplace) coupled climate model. It simulates the energy and water fluxes between the soil, the vegetation, and the atmosphere through the SECHIBA (Schématisation des Echanges Hydriques à l'Interface entre la Biosphère et l'Atmosphère, Ducoudré et al., 1993; de Rosnay and Polcher, 1998) module, while —andthe CO₂ fluxes and ecosystem carbon cycling throughare described by the STOMATE (Saclay Toulouse Orsay Model for the Analysis of Terrestrial Ecosystems, Viovy, 1996) module. When coupled with SECHIBA, STOMATE links the fast hydrological and biophysical processes with the carbon dynamics. STOMATE also contains a dynamic vegetation model, but this module was not activated for this study. In each grid cell, up to 12 plant functional types (PFTs) can be represented simultaneously, in addition to bare soil. LAI dynamics (from carbohydrate allocation) is simulated by STOMATE

which models the allocation of assimilates, autotrophic respiration components, foliar development, mortality and litter and soil organic matter decomposition. A factor representing drought stress (McMurtrie et al., 1990) linearly computes the rate of ribulose bisphosphate (RuBP) regeneration and the carboxylation rate.

-The drought stress and the leaf age of the vegetation directly influence the photosynthetic capacity (Farquhar et al., 1980; Collatz et al., 1992; Verbeeck et al., 2011; de Weirdt et al., 2012), and indirectly the stomatal conductance (Ball et al., 1987), which. The latter controls the transpiration as and is a function of two profiles: water availability in the soil column and a fixed root density profile for each PFT, and the soil moisture profile (de Rosnay and Polcher, 1998). Canopy interception is proportional to LAI and the corresponding evaporation proceeds at potential rate, like the soil evaporation. In the latter case, however, soil moisture can become limiting if the upward diffusion to the top soil layer cannot supply enough water to sustain the required potential rate.

The soil hydrology is represented by an 11-layer scheme which simulates vertical unsaturated flows based on the Richards equation, which allows capillary riseSoil moisture redistribution is described by a multi-layer scheme to solve the Richards equation for vertical unsaturated flow under the effect of root uptake (de Rosnay et al., 2002; Campoy et al., 2013). The hydraulic conductivity and diffusivity depend on soil moisture following the Van Genuchten (1980) model: the required parameters are taken from (Carsel and Parrish, 1988), and depend on the dominant soil texture in each grid-cell, based on the 1° × 1° texture map by Zobler (1986). For numerical integration, tThe 2-m soil column is divided into 11 discrete layers, with thickness increasing geometrically with depth, while the saturated hydraulic conductivity exponentially decreases with depth, to account for increased compaction and reduced bioturbation (Beyen and Kirkby, 1979) Variation in soil texture among grid cells is taken into account by means of three different soil types (coarse, medium and fine textured). Their spatial distribution is diagnosed by interpolating the 1° × 1° Food and Agriculture Organization texture map (FAO, 1978) by Zobler (1986), upscaled to the working resolution of ORCHIDEE by only keeping the dominant texture in each grid cell. The precipitation rate and the soil hydraulic conductivity govern the partitioning between soil infiltration and surface runoff-production and soil infiltration, which involves a time splitting procedure inspired from Green and Ampt (1911) to describe the propagation of the wetting front. The second contribution to total runoff is free gravitational drainage at the bottom of the soil.

The routing module (Polcher, 2003; Ngo-Duc et al., 2005; Guimberteau et al., 2012) calculates the daily discharge in each grid-cell and to the ocean. ItStreamflow routing is based relies on a series of linear reservoirs along the drainage networks, derived atfrom a 0.5° resolution data set (Vörösmarty et al., 2000). The routing scheme also includes a floodplain/swamp parameterization (d'Orgeval et al., 2008), recently improved by Guimberteau et al. (2012) for the Amazon basin, by introducing a new floodplain/swamp map. The simulation of the hydrology by the model ORCHIDEE has been widely tested over the Amazon basin and its sub-catchments (Guimberteau et al., 2012; Getirana et al., 2014; Guimberteau et al., 2014).

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Location	Station		River		Latitude	Longitude	Area (km ²)
	Name	Abbreviation	$_{ m Name}$	Abbreviation			
MAIN	Óbidos	OBI	Amazonas	AMAZ	-1.95	-55.30	4,680,000
SOUTH	Fazenda Vista Alegre	FVA	Madeira	MAD	-4.68	-60.03	1,293,600
	Guajará-Mirim	GMIR	$\operatorname{Mamor\acute{e}}$	MAM	-10.99	-65.55	532,800
	Itaituba	ITA	Tapajós	TAP	-4.24	-56.00	461,100
	Altamira	ALT	Xingu	XIN	-3.38	-52.14	$469,\!100$
WEST	Tamshiyacu	TAM	Upper Solimões	UPSO	-4.00	-73.16	726,400
	Lábrea	LAB	Purus	PUR	-7.25	-64.80	230,000
	Gavião	GAV	Juruá	JUR	-4.84	-66.85	170,400
NORTH	Caracaraí	CARA	Branco	BRA	+1.83	-61.08	130,600

Table S1. List of the gauging stations for the studied sub-catchments. Sources: SO HYBAM (Observation Service of the Geodynamical, hydrological and biogeochemical control of erosion/alteration and material transport in the Amazon, Orinoco and Congo basins, Cochonneau et al., 2006).

	Di.	M - J -1	Relative bias (%)		Correlation coefficient		NRMSE (%)	
	Basin	Model	Q	\mathbf{ET}	Q	ET	Q	ET
MAIN	AMAZ	INLAND-DGVM	-22.4	-1.8	-	0.60	-	14.1
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	-21.9	+1.9	0.77	0.55	36.6	25.0
		ORCHIDEE	-5.9	-4.6	0.91	0.58	14.1	17.2
	MAD	INLAND-DGVM	-28.3	+0.2	-	0.89	-	13.6
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	-2.2	-9.5	0.89	0.83	33.5	28.9
		ORCHIDEE	-5.5	-1.7	0.99	0.88	20.2	15.2
•	MAM	INLAND-DGVM	-60.9	+0.92	-	0.99	-	15.0
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+14.0	-14.8	0.73	0.91	47.4	30.6
SOUTH -		ORCHIDEE	-22.2	-3.0	0.91	0.98	43.4	18.0
SOUTH.	TAP	INLAND-DGVM	+10.5	-3.0	-	0.02	-	13.4
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+25.1	-6.8	0.90	0.45	53.9	34.0
		ORCHIDEE	+16.6	-3.3	0.96	0.11	47.5	11.5
	XIN	INLAND-DGVM	+47	-1.9	-	0.17	-	14.9
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+59.1	-5.9	0.83	-0.01	66.6	34.5
		ORCHIDEE	+34.1	-4.4	0.94	0.31	46.1	12.6
	UPSO	INLAND-DGVM	-57.4	+2.2	-	0.32	-	22.9
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	-45.2	-0.9	0.93	0.87	86.0	18.5
		ORCHIDEE	-17.2	-10.0	0.96	0.31	23.9	25.5
•	PUR	INLAND-DGVM	+9.3	+2.6	-	0.83	-	9.8
WEST		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+18.6	+1.7	0.86	0.27	39.3	24.0
		ORCHIDEE	+15.8	-0.9	0.96	0.79	31.6	10.1
	JUR	INLAND-DGVM	+9.3	-0.05	-	0.86	-	9.3
		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+10.2	+7.3	0.89	0.02	29.7	17.0
		ORCHIDEE	+39.4	-4.1	0.96	0.82	40.1	10.4
	BRA	INLAND-DGVM	+47.1	+17.1	-	0.74	-	21.0
NORTH		$\mathrm{LPJmL} ext{-}\mathrm{DGVM}$	+53.3	+12.5	0.99	0.06	51.3	33.1
		ORCHIDEE	+69.3	+10.9	0.96	0.61	58.8	15.0

Table S2. Bias (%), correlation and NRMSE (Normalized Root Mean Square Error) (%) against the observations, of discharge and ET, for each catchment, for HIST period. Observed discharge comes from SO HYBAM and ET is estimated by the product of Jung et al. (2010).

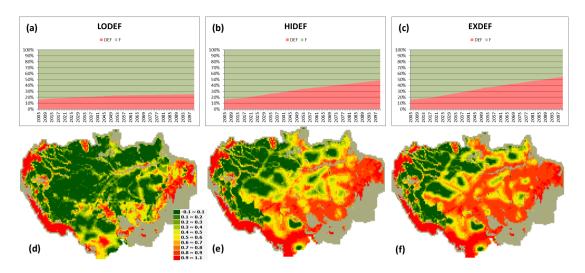


Figure S1. Percentage of dDeforested area (%) in each 25 x 25 km² for the LCC scenarios LODEF (a and d), HIDEF (b and e) and EXDEF (c and f).

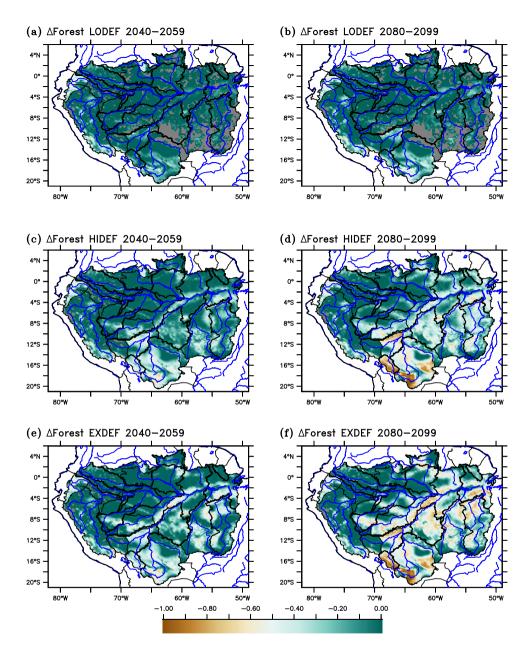


Figure S2. Decrease of forest fraction for the three LCC scenarios (for the two time periods) compared with the NODEF scenario in 2009 over the Amazon basin. Grey colour indicates no change of forest fraction.

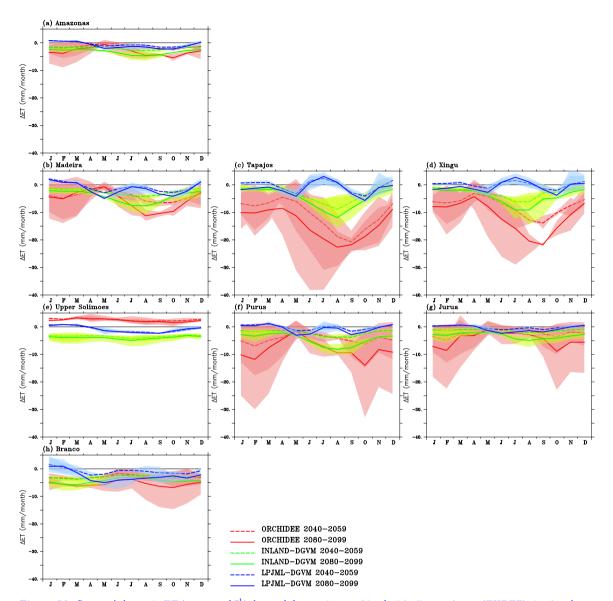


Figure S3. Seasonal change in ET (mm month⁻¹) due to deforestation combined with climate change (EXDEF) simulated by the three LSMs over the Amazon basin and its catchments, averaged over the two future periods. For a given LSM and period, the shaded area defines the envelope enclosing the range with plausible climate futures.

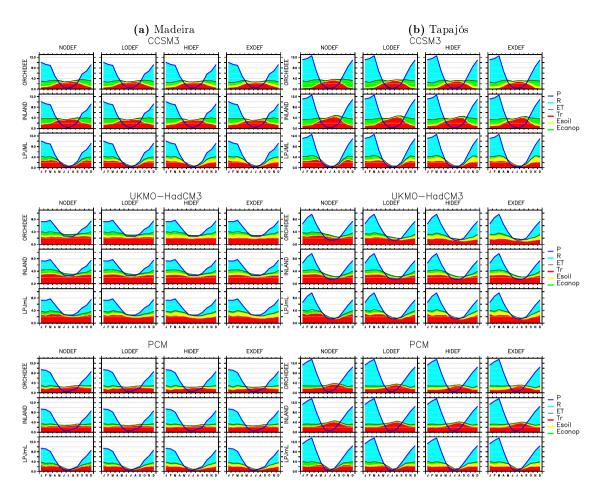


Figure S24. For each GCM-forcing, monthly mean seasonalities of the water budget components (including the ET components) (mm d^{-1}) from the three LSMs (linesrows) and for each NODEF and LCC scenarios (columns) over (a) the Madeira and (b) the Tapajós catchments. The variables of the water budget are: precipitation (P), runoff (R) and evaportanspiration (ET). The variables of the ET components are: transpiration (Tr), soil evaporation (Esoil) and evaporation of canopy interception (Ecanop).