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Impacts of future climate change on urban flood volumes in Hohhot City in Northern China: benefits of climate mitigation and adaptations

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28 **Abstract**

29 As China has become increasingly urbanised, flooding has become a regular occurrence in its
30 major cities. Assessing the effects of future climate change on urban flood volumes is crucial to
31 informing better management of such disasters given the severity of the devastating impacts of
32 flooding (e.g., the 2016 flooding across China). Although recent studies have investigated the
33 impacts of future climate change on urban flooding, the effects of both climate change mitigation
34 and adaptation have rarely been accounted for together in a consistent framework. In this study,
35 we assess the benefits of mitigating climate change by reducing greenhouse gas emissions and
36 locally adapting to climate change by modifying drainage systems to reduce urban flooding
37 under various climate change scenarios through a case study conducted in Northern China. The
38 urban drainage model—Storm Water Management Model—was used to simulate urban flood
39 volumes using current and two adapted drainage systems (i.e., pipe enlargement and low-impact
40 development), driven by bias-corrected meteorological forcing from five general circulation
41 models in the Coupled Model Intercomparison Project Phase 5 archive. Results indicate that
42 urban flood volume is projected to increase by 52% in 2020–2040 compared to the volume in
43 1971–2000 under the business-as-usual scenario (i.e., Representative Concentration Pathway
44 (RCP) 8.5). The magnitudes of urban flood volumes are found to increase nonlinearly with
45 changes in precipitation intensity. On average, the projected flood volume under RCP 2.6 is 13%
46 less than that under RCP 8.5, demonstrating the benefits of global-scale climate change
47 mitigation efforts in reducing local urban flood volumes. Comparison of reduced flood volumes
48 between climate change mitigation and local adaptation (by improving the drainage system)
49 scenarios suggests that local adaptation is more effective than climate change mitigation in
50 reducing future flood volumes. This has broad implications for the research community relative

51 to drainage system design and modelling in a changing environment. This study highlights the
52 importance of accounting for local adaptation when coping with future urban floods.

53

54 **Keywords:** Climate change, urban floods, mitigation, adaptation, drainage systems

55

56 1. Introduction

57 Floods are one of the most hazardous and frequent disasters in urban areas and can cause
58 enormous impacts on the economy, environment, city infrastructure, and human society (Chang
59 et al., 2013; Ashley et al., 2007; Zhou et al., 2012). Urban drainage systems have been
60 constructed to provide carrying and conveyance capacities at a desired frequency to prevent
61 urban flooding. The design of drainage systems is generally based on historical precipitation
62 statistics for a certain period of time, without considering the potential changes in precipitation
63 extremes for the designed return periods (Yazdanfar and Sharma, 2015; Peng et al., 2015;
64 Zahmatkesh et al., 2015). For example, in Danish design guidelines for urban drainage, a 30%
65 and 40% increase in the precipitation intensity is expected for the 10- and 100-year return
66 periods, respectively (Arnbjerg-Nielsen, 2012). The systems are, however, likely to be
67 overwhelmed by additional runoff effects induced by climate change, which may lead to
68 increased flood frequency and magnitude, disruption of transportation systems, and increased
69 human health risk (Chang et al., 2013; Abdellatif et al., 2015). Therefore, it is important to
70 investigate the performance of drainage systems in a changing environment and to assess the
71 potential urban flooding under various scenarios to achieve better adaptations (Mishra, 2015;
72 Karamouz et al., 2013; Yazdanfar and Sharma, 2015; Notaro et al., 2015).

73

74 Impacts of climate change on extreme precipitation and urban flooding have been well
75 documented in a number of case studies. For example, Ashley et al. (2005) showed that flooding
76 risks may increase by almost 30 times in comparison to current situations, and effective
77 adaptation measures are required to cope with the increasing risks in the UK. Larsen et al. (2009)
78 estimated that future extreme one-hour precipitation will increase by 20%~60% throughout
79 Europe. Willems (2013) found that in Belgium the current design storm intensity for the 10-year
80 return period is projected to increase by 50% by the end of this century. Several studies have also
81 investigated the role of climate change mitigation and adaptation in reducing urban flood
82 damages and risks under climate change scenarios (Alfieri et al., 2016; Arnbjerg-Nielsen et al.,
83 2015; Moore et al., 2016; Poussin et al., 2012). To date, however, limited work has been done to
84 investigate the relationship between changes in precipitation intensity and flood volume to
85 provide additional insights into drainage design strategies. More importantly, investigations of
86 the benefits of climate change mitigation (by reducing greenhouse gas emissions [GHG]) and
87 local adaptation (by improving drainage systems) in reducing future urban flood volumes are
88 typically conducted separately, rather than within a consistent framework.

89

90 As China has become increasingly urbanised, flooding has become a regular occurrence in its
91 cities; 62% of Chinese cities surveyed experienced floods and direct economic losses of up to
92 \$100 billion between 2011 and 2014 (China Statistical Yearbook 2015). The 2016 flooding
93 affected more than 60 million people—more than 200 people were killed and \$22 billion in
94 losses were suffered across China. Hence, assessing future changes in urban flooding is very
95 important for managing urban floods by designing new and re-designing existing urban
96 infrastructures to be resilient in response to the impacts of future climate change. While urban

97 floods are speculated to increase in the future (Yang 2000; Ding et al., 2006), their magnitudes
98 are hard to assess because of uncertainties associated with future climate change scenarios, as
99 well as the under-representation of plausible climate change mitigation and adaptation strategies
100 in the models.

101
102 In this study, we chose a drainage system in a typical city in Northern China to illustrate the role
103 of climate change mitigation and local adaptation in coping with future urban flood volumes.
104 Such an investigation of the performance of the present-day drainage system also has important
105 implications for local governments responsible for managing urban flood disasters in the study
106 region. Specifically, we first quantified the effects of future climate change on urban flood
107 volumes as a result of extreme precipitation events for various return periods using the present-
108 day drainage system. We then designed two plausible adaptation strategies for the study region
109 and investigated how much urban flood volume can be reduced by the adapted systems. We also
110 compared the benefits of global-scale climate change mitigation and local adaptation in reducing
111 urban flood volumes to advance our understanding of the effective measures for coping with
112 future urban floods.

113

114 2. Materials and Methods

115 a. Study region

116 The study region (Hohhot City) is located in the south-central portion of Inner Mongolia, China.
117 It lies between the Great Blue Mountains to the north and the Hetao Plateau to the south, which
118 has a north-to-south topographic gradient. The drainage area in year 2010 was about 210.72 km²
119 and it served a residential population of 1.793 million (Figure 1a). The land use types in the

120 region can be classified into five categories: agricultural land (8%), residential areas (38%),
121 industrial land (13%), green spaces (7%), and other facilities (34%, including municipal squares,
122 commercial districts, institutions). The planned drainage area in 2020 is about 307.83 km², which
123 is 50% larger than the current drainage area. The land use categories and distribution are shown
124 in Figure 1b.

125
126 The region is in a cold semi-arid climate zone, characterised by cold and dry winters and hot and
127 humid summers. The regional annual mean precipitation is approximately 396 mm and it
128 exhibits large intra-seasonal variations. Most rain storms fall between June and August, a period
129 that accounts for more than 65% of the annual precipitation. According to local water authorities,
130 the major soil type of the area is a mixture of loam and clay. The current drainage system can be
131 divided into three large sub-basins (Figure 1c) and 326 sub-catchments with a total pipeline
132 length of 249.36 km. The drainage network has a higher pipeline cover rate in the central part,
133 but a rather low design standard for extreme rainfall events with a return period of less than 1
134 year. Historical records of stormwater drainage and flood damage indicate that the region has
135 experienced an increase in flood frequency and magnitude within the context of climate change
136 and urbanisation. During the major flood event on 11 July 2016, the city, especially the western
137 portion of the watershed, was hit by an extreme rainfall event that featured more than 100 mm of
138 rain in 3 hours. The flood event led to the cancellation of at least 8 flights and 17 trains, and
139 delays of several transportation systems. In particular, in the central area, the flood event caused
140 severe traffic jams on major streets and resulted in a number of flooded residential buildings. A
141 new drainage system is therefore required to cope with increasing urban flood volumes and
142 frequencies in the future.

143

144 b. Climate change scenarios

145 Climate projections by five general circulation models (GCMs) from Phase 5 of the Coupled
146 Model Intercomparison Project (CMIP5) archive were obtained from the Inter-Sectoral Impact
147 Model Intercomparison Project (ISI-MIP) (Warszawski et al., 2014). The CMIP5 climate
148 projections were bias-corrected against observed climate for the overlapping period 1950–2000
149 using a quantile mapping method (Piani et al., 2010; Hempel et al., 2013). The bias-corrected
150 CMIP5 climate projections represent a complete climate change picture that includes both the
151 mean property and variation of future climate. Several studies have demonstrated the value of the
152 bias-corrected climate projections in quantifying climate change impacts on global and regional
153 hydrology (e.g., Piontek et al., 2014; Elliott et al., 2014; Haddeland et al., 2014; Leng et al.,
154 2015a,b). In this study, we used the bias-corrected climate from all five GCMs (HadGEM2-ES,
155 GFDL-ESM2M, IPSLCM5A-LR, MIROC-ESM-CHEM, and NorESM1-M) under two
156 Representative Concentration Pathways (RCPs) (i.e., RCP 2.6 and RCP 8.5). The projected
157 urban flood volumes under the business-as-usual scenario RCP 8.5 are compared with those
158 under the climate change mitigation scenario RCP 2.6 to explore the benefits of climate change
159 mitigation in reducing regional urban flood volumes. The possible land-surface-atmosphere
160 interactions that would indirectly affect rainfall and flooding are not considered in this study.

161

162 c. Urban drainage modelling

163 The Storm Water Management Model (SWMM 5.1) developed by the U.S. Environmental
164 Protection Administration is a widely used urban stormwater model that can simulate rainfall-
165 runoff routing and pipe dynamics under single or continuous events (Rossman and Huber, 2016).
166 SWMM can be used to evaluate the variation in hydrological and hydraulic processes and the

167 performance of drainage systems under specific mitigation and adaptation scenarios in the
168 context of global warming. The hydrological component requires inputs of precipitation and
169 subcatchment properties including drainage area, subcatchment width, and imperviousness. The
170 pipe network requires inputs from manholes, pipelines, outfalls, and connections to sub-
171 catchments (Zahmatkesh et al., 2015; Chang et al., 2013). Basic flow-routing models include
172 steady flow, kinematic, and dynamic wave methods. Infiltration can be described by the Horton,
173 Green-Ampt, or Curve Number (SCS-CN) methods. The dynamics of pipe flow are calculated
174 based on the continuity equation and Saint-Venant equations (Rossman and Huber, 2016).
175 Overflow occurs once the surface runoff exceeds the pipe capacity and is expressed as the value
176 of total flood volume (TFV) at each overloaded manhole; i.e., the excess water from manholes
177 after completely filling the pipe system without taking into account the outlet discharges. Other
178 types of model outputs include catchment peak flows, maximum flow rates of pipelines, and
179 flooded hours of manholes. It should be noted that SWMM is not capable of simulating surface
180 inundation dynamics and cannot provide accurate estimation of the inundated zones and depths.
181 The TFV value is thus used to approximately reflect the flood condition and drainage system
182 overloading status. Nevertheless, surface inundation models (e.g., Apel et al., 2009; Horritt and
183 Bates, 2002; Vojinovic and Tutulic, 2009) are applicable if more accurate information about
184 overland flow characteristics is available. In this study, the kinematic wave routing and the
185 Horton infiltration model are used for model simulations. The infiltration capacity parameters for
186 the category of "Dry loam soils with little or no vegetation" are used in the hydrological model to
187 be consistent with the local soil type (Akan, 1993; Rossman and Huber, 2016) (Table 1).

188

189 Rainfall inputs are calculated based on the regional storm intensity formula (SIF) using historical
190 climatic statistics (Zhang and Guan, 2012) (see Equation 1). Application of the SIF is a standard
191 practice for determining design rainfalls in urban drainage modelling in China, and is well
192 documented in the National Guidance for Design of Outdoor Wastewater Engineering
193 (MOHURD, 2011). In fact, the SIF represents an Intensity-Duration-Frequency (IDF)
194 relationship, which is a common approach in literature for estimating design rainfall hydrographs
195 using the Chicago Design Storms (CDS) approach (Berggren et al., 2014; Willems et al., 2012;
196 Zhou et al., 2013).

$$q = \frac{A(1 + D \lg(P))}{(t + b)^c} \quad \text{Eq. (1)}$$

197 where q is the average rainfall intensity, and P and t are the design return period and duration of
198 storm, respectively. The typical temporal resolution considered in SIF for urban drainage
199 modelling is minutes. A , b , c , and D are regional parameters governing the IDF relationship
200 among rainfall intensity, return period, and storm duration. For the study region, the values of A ,
201 b , c , and D were obtained from the local weather bureau and are equal to 635, 0, 0.61, and 0.841,
202 respectively.

203

204 The procedure for applying SIF to obtain CDS is outlined in the National Technical Guidelines
205 for Establishment of Intensity-Duration-Frequency Curve and Design Rainstorm Profile
206 (MOHURD, 2014; Zhang et al., 2008; Zhang et al., 2015). Specifically, for a given return period,
207 the SIF is fitted into the Horner's equation as:

$$i = \frac{a}{(t + b)^c} \quad \text{Eq. (2)}$$

208

209 The synthetic hyetograph based on the Chicago method is computed using Equation 2 and an
 210 additional parameter r (where $0 < r < 1$), which determines the relative time step of the peak
 211 intensity, $t_p = r * t$. The time distribution of rainfall intensity is then described after the peak $t_a =$
 212 $(1-r)*t$ and before the peak $t_b = r*t$ using Equations 3 and 4, respectively, where i_b and i_a are the
 213 instantaneous rainfall intensity before and after the peak:

$$i_a = \frac{a[\frac{(1-c)t_a}{(1-r)} + b]}{(\frac{t_a}{(1-r)} + b)^{1+c}} \quad \text{Eq. (3)}$$

$$i_b = \frac{a[\frac{(1-c)t_b}{r} + b]}{(\frac{t_b}{r} + b)^{1+c}} \quad \text{Eq. (4)}$$

214 In this study, we considered 10 return periods, i.e., the 1-, 2-, 3-, 10-, 20-, 50-, 100-, 200-, 500-,
 215 and 1000-year events. A 4-hour rainfall time series was generated for each return period at 10-
 216 minute intervals based on Equations 1–4. We assumed that the SIF was constant without
 217 considering the non-stationary features in a changing climate. That is, the IDF relationships were
 218 assumed to remain stable in the future and only changes in the daily mean intensity were
 219 considered because of the limited data availability in future sub-hourly climate projections from
 220 which to derive the parameters.

221
 222 As for future climate, the projected changes (i.e., change factors) in precipitation intensity at
 223 various return periods were calculated for each GCM-RCP combination (Table 2). Specifically,
 224 for each year, the annual maximum daily precipitation was determined for both historical and
 225 future periods. The generalised extreme value (GEV) distribution was then fitted separately to
 226 the two sets of daily values (Coles 2001; Katz et al. 2002). The goodness-of-fit was tested by
 227 calculating the Kolmogorov–Smirnov and Anderson–Darling statistics. The value corresponding

228 to each return period was estimated based on the GEV distribution and the changes between
229 future and historical periods were calculated as the change factors. The derived change factor for
230 each return period was then multiplied by the historical design CDS rainfall time series to derive
231 future climate scenarios. We acknowledge that the estimation of changes in extreme precipitation
232 events involves inevitable uncertainties and therefore caution should be exercised when
233 interpreting the relevant results.

234

235 d. Flood volume assessment

236 The TFV values of given rainfall events were simulated by the SWMM. A log-linear relationship
237 is assumed to characterize the changes in flood volume with the increase in precipitation
238 intensity as indicated by return periods (Figure 2a) following Zhou et al. (2012) and Olsen et al.
239 (2015). Generally, more intense rainfall will induce higher TFVs. The TFVs were further linked
240 to their occurrence frequencies to derive the expected flood volume for a flood event at a specific
241 probability (Figure 2b). The total grey area under the curve represents the average total TFVs per
242 year for all floods at various return periods. The contribution of an individual flood event to total
243 TFVs is dependent not only on the flood volume, but also its corresponding probability of
244 occurrence. Intensified precipitation is expected to increase the magnitude of system overflow,
245 resulting in an upward trend in the TFV-return period relationship and increased total TFVs.
246 Mitigation and adaptation are aimed at reducing or preventing the impacts of global warming on
247 flood volumes.

248

249 e. Design of adaptation scenarios

250 In this study, two adaptation scenarios were designed to explore the role of adaptation in
251 reducing urban flood volume within the context of climate change. The first scenario adapted the
252 drainage system as planned by the water authorities to cope with the designed standard of a 3-
253 year design event. It involved two main improvements of the current drainage system—
254 enhancing the pipeline diameter and expanding the pipe network. The design was implemented
255 in the SWMM model as shown in Figure 1c. The number of pipelines of the present-day and
256 adapted systems was 323 and 488, with a total pipe length of 251.6 km and 375.4 km,
257 respectively. In the adapted scenarios, the mean pipeline diameter was about 1.73 m, which
258 increased by 53% compared to that of the present-day system.

259

260 A variety of site-specific factors, such as the imperviousness of land area in the drainage basin,
261 can also influence the performance of a drainage system in managing surface runoff. The second
262 adaptation scenario was to increase the permeable surfaces (e.g., green spaces) and reduce the
263 regional imperviousness in the study region on the basis of pipe capacity enhancement. This
264 scenario is referred to as the Low Impact Development (LID) scenario, and it was used to
265 explore the effectiveness of urban green measures, such as the use of permeable pavements,
266 infiltration trenches, and green roofs. Due to a lack of detailed information about the permeable
267 soil and coverage rates in the study region, the effects of these specific measures cannot be
268 modelled individually. Here, we used a simplified approach by altering the subcatchment
269 imperviousness to reflect the combined effects of infiltration-related measures. We derived such
270 information by comparing the current and planned land use maps using a geographical
271 information system (GIS) and incorporated the changes in land use and imperviousness into the
272 designed LID scenarios. Figure 1d shows the difference in weighted mean imperviousness (WMI)

273 calculated for each subcatchment in the current and planned maps, using the commonly applied
274 impervious factors (Pazwash, 2011; Butler and Davies, 2004) for each land use type. The
275 difference in WMI was used to indicate the area potential for adaptation based on the city plan.
276 For example, a subcatchment with higher positive changes in the WMI indicates that the area is
277 planned to have a land use type with lower imperviousness and therefore is assumed to be more
278 suitable for LID planning, and vice versa.

279

280 3. Results

281 a. Impacts of future climate change on urban flood volumes

282 Figure 3 shows the projected climate change impacts on urban flooding using the present-day
283 drainage system of the near future (i.e., 2020–2040) under the RCP 8.5 scenario. Without climate
284 change mitigation or adaptation, the TFV was projected to increase significantly with the
285 increase of extreme rainfall events for most of investigated return periods (Table 2). Note that the
286 lower bounds for return periods of 1, 3, and 1000 years fall below the current TFV curve due to
287 the decrease in precipitation intensities. Despite the large uncertainty associated with climate
288 projections, in particular with the 1-, 10-, and 1000-year return periods, the poor service
289 performance of the current system in coping with urban flooding was evident. Overall, the urban
290 flood volume was projected to increase by 52% on average by the multi-model ensemble median
291 by 2020–2040; the largest increase (258%) was projected for the 1-year event and the smallest
292 increase (12%) for the 100-year event.

293

294 b. Benefits of climate change mitigation in reducing urban flood volumes

295 Figure 4 shows the comparison of TFVs under the RCP 8.5 scenario (i.e., a business-as-usual
296 scenario) and the RCP 2.6 scenario (i.e., a climate change mitigation scenario). Although large
297 uncertainties exist arising from climate models, it is clear that the simulated TFVs are much
298 smaller under the RCP 2.6 scenario than under the RCP 8.5 scenario, demonstrating the benefits
299 of climate mitigation in reducing local urban flood volumes. Such benefits are especially evident
300 for floods for smaller return periods. For example, an increase of 936 m³ in flood volume is
301 projected with the increase in 1-year extreme rainfall under the business-as-usual climate change
302 scenario (i.e., RCP 8.5), 52% of which would be reduced if climate change mitigation is in place
303 (i.e., under RCP 2.6). Overall, climate change mitigation can reduce future flood volumes by 13%
304 compared to the scenario without mitigation, as indicated by the multi-model ensemble median.
305 Notably, the peak of the total TFV curve was even projected to shift from the 1-year event under
306 the RCP 8.5 scenario to the 3-year event under the RCP 2.6 scenario (Figure 4b). Such a shift in
307 the peak toward smaller return periods combined with a flatter curve demonstrates the important
308 role of climate mitigation in regulating local urban flood volumes.

309

310 c. Benefits of adaptation in reducing urban flood volumes

311 Figure 5 shows the overloaded pipelines (red colour) with and without adaptation. The simulated
312 results under the present 3-year event (recommended service level) and 50-year event (one
313 typical extreme event) were selected to illustrate the role of adaptation in coping with floods in
314 the historical period. As shown in Figure 5a, the simulated locations of overloaded pipelines are
315 in good agreement with historical flood points as recorded by local water authorities. Overall, the
316 percentage of overloaded manholes (POM) and the ratio of flood volume (RFV) are up to 37%
317 and 35% in the current drainage system (Figure 5a), respectively. When experiencing a 50-year
318 extreme rainfall, the POM and RFV increase to 67% and 38%, respectively. This indicates that

319 current pipe capacities are insufficient to cope with extreme rainfall events (Figure 5b). Spatially,
320 the central portion of the city is the most affected region due to the low service level in the area.
321 With proposed adaptations, urban floods can be reduced to zero under a 3-year flood event. Such
322 benefits of local adaptations are also evident when experiencing more intense precipitation
323 events (e.g., 50-year events), for which the POM and RFV reduced from 67% and 50% to 49%
324 and 17%, respectively.

325

326 Figure 6 shows the future changes in urban flood volume (CTFVs) ($CTFV = (TFV_c - TFV_{nc}) / TFV_{nc}$,
327 where c and nc represent the results with and without climate change, respectively) with changes
328 in extreme rainfall for various return periods. The performance of the current drainage system
329 (no adaptation) was found to be less sensitive to future climate change, as indicated by the flatter
330 slope in Figure 6. For example, a similar magnitude of changes in flood volume was projected
331 given changes in extreme rainfall for the return periods of 3, 50, and 500 years; the CTFV is 0.62,
332 0.32 and 0.35 for these periods, respectively. This is because the capacity of the current system is
333 too small to handle extreme rainfall events with return periods larger than 1 year—a condition
334 under which the current drainage system would be flooded completely, not to mention the
335 situations with increased rainfall intensity in the future. Mathematically, the low sensitivity of
336 the current drainage system to changes in extreme rainfall intensity could be attributed to the
337 large value of the denominator in the calculation of $CTFV$.

338

339 With adaptations in place, the flood volume becomes much smaller than that in the current
340 system due to capacity upgrading to hold more water. For example, when experiencing a 10-year
341 extreme rainfall event, the urban flood volumes for the present period (i.e., TFV_{nc}) are 1041,230,

342 274,650 and 180,610 m³ in the current and two adapted systems, respectively, while in the future
343 period, the magnitude of flood volume (i.e., *TFV_c*) is relatively similar among the three drainage
344 systems. Therefore, future CTFVs relative to the historical period are much larger in the adapted
345 systems than in the current system. The larger CTFVs in the adapted systems do not mean a
346 worsened drainage system performance. Rather, they imply that the capacity (i.e., service level)
347 of adapted drainage systems tends to become lower with climate change, while the current
348 drainage system has already reached its peak capacity in handling extreme rainfall events in the
349 historical period and thus shows a low sensitivity to future increases in rainfall intensity under
350 climate change scenarios. Notably, the considerable increases in the CTFVs for return periods of
351 less than 10 years in the adapted systems imply that the designed adaptations can effectively
352 attenuate extreme rainfall events for small return periods. For more extreme rainfall events of
353 return periods ≥ 50 years, more consistent results were found for both adaptation scenarios. This
354 indicates that although the performances of adapted drainage systems are significantly improved
355 compared to that of the current system, the flood volume remains large when experiencing
356 extreme rainfall events with return periods larger than 50 years, because flooding in such cases
357 will push the adapted drainage systems to their upper limits.

358

359 d. Climate mitigation versus drainage adaptation

360 Figure 7 shows the reduced TFVs by climate change mitigation and drainage system adaptation
361 as functions of return period. It is evident that both mitigation and adaptation measures are
362 effective in reducing future urban flood volumes. However, such benefits are projected to
363 weaken gradually with the increase in rainfall intensity (i.e., larger return periods). Importantly,
364 our results show that the two adaptation systems proposed in this study are found to be more

365 effective in reducing urban floods than climate change mitigation. In most cases, the benefits of
366 local adaptation are more than double those of mitigation. In extreme cases, the reduction in TFV
367 achieved by adaptation is five times more than that achieved by climate change mitigation (i.e.,
368 for the return periods of 2–3 years). Such effectiveness of urban flood reduction through
369 drainage system adaptations has profound implications for local governments charged with
370 managing urban flooding in the future. Notably, the second scenario (LID+pipe) exhibited a
371 higher level of flood volume reduction than the pipe scenario in coping with extreme rainfall
372 events for all investigated return periods. This implies that implementation of LID measures to
373 augment drainage system capacity is more effective through reducing upstream loadings
374 compared to updating the pipe system alone.

375

376 It is noted that local soil characteristics could affect the performance of the designed adaptation
377 systems, in particular the LID measures. However, information about soil properties was not
378 available at the subcatchment level in the study region. Here, a set of sensitivity experiments
379 were conducted by adopting different parameters (e.g., infiltration values) associated with
380 possible soil conditions (i.e., dry sand, loam, and clay soils with little or no vegetation in Table 1)
381 for the area. The boundary bars in Figure 7 show the uncertainty range arising from the
382 representation of different soil conditions in the drainage model. The benefits of the designed
383 adaptation measures in reducing urban flood volumes were found to be robust regardless of soil
384 conditions, and such benefits exceeded those of climate change mitigation, confirming our major
385 conclusions found in this study.

386

387 4. Uncertainties and Limitations

388 A number of uncertainties and limitations arise from the model structure, parameter inputs,
389 emission scenarios, GCMs, climate downscaling/bias-correction approaches, etc. Specifically,
390 climate projections by GCMs are subject to large uncertainties, in particular regarding
391 precipitation (Covey et al., 2003) at spatial scales, which are relevant for urban flood modelling.
392 An alternative approach is to simulate future climate using a regional climate model (RCM)
393 nested within a GCM. Such climate projections by RCMs have added value in terms of higher
394 spatial resolution, which can provide more detailed regional climate information. However,
395 various levels of bias would still remain in RCM simulations (Teutschbein and Seibert 2012) and
396 bias corrections of RCM projections would be required; e.g., the European project ENSEMBLES
397 (Hewitt and Griggs 2004; Christensen et al. 2008). To run a RCM was not within the scope of
398 this study; instead, we tended to use publicly available climate projections. Here, we obtained the
399 climate projections from the ISI-MIP (Warszawski et al. 2014), which provides spatially
400 downscaled climate data for impact models. The climate projections were also bias-corrected
401 against observations (Hempel et al. 2013) and have been widely used in climate change impact
402 studies on hydrological extremes such as floods and droughts (e.g., Dankers et al. 2014;
403 Prudhomme et al. 2014; Leng et al. 2015a). It should be noted that we used the delta change
404 factor to derive future climate scenarios as inputs into our drainage model instead of using GCM
405 climate directly. This is because the relative climate change signal simulated by GCMs is argued
406 to be more reliable than the simulated absolute values (Ho et al. 2012). Moreover, we used an
407 ensemble of GCM simulations rather than one single climate model in order to characterise the
408 uncertainty range arising from climate projections. However, disadvantages of this method are
409 that transient climate changes cannot be represented and that changes in intra-seasonal or daily

410 climate variability are not taken into account (Leng and Tang, 2014). Such sources of uncertainty
411 can be explored when improved climate models at finer scales become available (Jaramillo and
412 Nazemi 2017).

413
414 In addition, the SIF parameters were assumed to remain stable in the future and only changes in
415 the daily mean intensity were considered, because future sub-hourly climate projections were not
416 readily available. The full climate variability range would also be under-sampled, although we
417 used five climate models to show the possible range. Given the above limitations, we
418 acknowledge that the modelling results represent the first-order potential climate change impacts
419 on urban floods. Future efforts should be devoted to the representation of dynamic rainfall
420 changes at hourly time steps with consideration of non-stationary climate change.

421
422 Moreover, several assumptions had to be made due to limitations of the current modelling
423 structure and approach. For example, the conveyance capacities of the drainage system and flood
424 volume would largely depend on the state of drainage systems. Hence, a drainage system
425 obstructed by vegetation, waste, or artefacts (cables, pipes, temporary constructions) can make
426 the outcomes of the SWMM calculation significantly different from observations. However,
427 quantifying the impacts of drainage system states on urban flood volumes is not trivial because
428 of the difficulties involved in collecting field data and selecting and using appropriate methods
429 for reasonable assessment of pipe conditions (Ana and Bauwens, 2007; Fenner, 2000), and was
430 not within the scope of this study. With deterioration, such as ageing network, pipe deterioration,
431 blockage, and construction failures, drainage systems were shown to become more vulnerable to
432 extreme rainfalls as demonstrated in previous studies (Dawson et al., 2008; CIRIA, 1997; Davies

433 et al., 2001). It is very likely that our simulated urban flood volumes would be underestimated
434 without considering the changes in drainage conditions (Pollert et al., 2005).

435
436 Further, constrained by the one-dimensional modelling approach using SWMM, the
437 performances of LID measures were mainly evaluated according to their effects in reducing
438 water volume from overloaded manholes (Oraei Zare et al., 2012; Lee et al., 2013). That is, the
439 LID adaptation measure was mainly designed to reduce the amount of water rather than slowing
440 down the water speed, which has been demonstrated to be effective in reducing urban floods
441 (Messner et al., 2006; Ashley et al., 2007; Floodsite, 2009). However, it should be noted that
442 most LID measures can reduce runoff volume and flow speed at the same time, although some of
443 the LID measures are primarily designed to slow down the flow speed, i.e., vegetated swales. To
444 examine whether flood retention of a given event is induced by runoff volume or the internal
445 speed control function in the model is difficult and requires detailed data for model validations.
446 Specifically, the required information about surface roughness, soil conductivity, and seepage
447 rate were unavailable at the subcatchment scale in the study region. Therefore, a simplified
448 modelling approach was used to take advantage of existing data, especially for the design of LID
449 measures. With the aid of more detailed field data and planning documents, the design of LID
450 measures could be significantly improved by implementing more advanced approaches (Elliott
451 and Trowsdale, 2007; Zoppou, 2001). Evaluation of other potential adaptation strategies, such as
452 flood retention by rain gardens and green roofs, can be explored in the future to gain additional
453 insights into the performance of LID systems. In particular, the cost-effectiveness of the
454 proposed adaptation measures should be accounted for. Nevertheless, given these limitations,
455 this study stands out from previous climate impact assessment studies of urban flood volumes by

456 having proposed two feasible adaptation strategies and compared their benefits to those from
457 global-scale climate change mitigations through GHG reductions within a consistent framework.
458 Depending on the progress on data collection and the demands of local authorities, more
459 advanced methods for pipe assessment (e.g., considering the changing pipe conditions), LID
460 measures (detailed modelling of LID control), and two-dimensional surface flooding for
461 assessment of flood damage and risk are planned in a future study to provide a more
462 comprehensive analysis of the adaptation measures.

463

464 5. Summary and Conclusions

465 The potential impacts of future climate change on current urban drainage systems have received
466 increasing attention during recent decades because of the devastating impacts of urban flooding
467 on the economy and society (Chang et al., 2013; Zhou et al., 2012; Abdellatif et al., 2015).
468 However, few studies have explored the role of both climate change mitigation and drainage
469 adaptations in coping with urban flooding in a changing climate. This study investigated the
470 performance of a drainage system in a typical city in Northern China in response to various
471 future scenarios. In particular, we assessed the potential changes in urban flood volume and
472 explored the role of both mitigation and adaptation in reducing urban flood volumes in a
473 consistent manner.

474

475 Our results show significant increases in urban flood volumes due to increases in precipitation
476 extremes, especially for return periods of less than 10 years. Overall, urban flood volume in the
477 study region is projected to increase by 52% by the multi-model ensemble median in the period
478 of 2020–2040. Such increases in flood volume can be reduced considerably by climate change

479 mitigation through reduction of GHG emissions. For example, the future TFVs under 1-year
480 extreme rainfall events can be reduced by 50% when climate change mitigation is in place.
481 Besides global-scale climate change mitigation, regional/local adaptation can be implemented to
482 cope with the adverse impacts of future climate change on urban flood volumes. Here, the
483 adaptation measures as designed in this study were demonstrated to be much more effective in
484 reducing future flood volumes than climate change mitigation measures. In general, the reduced
485 flood volumes achieved by adaptation were more than double those achieved by climate change
486 mitigation.

487
488 Through a comprehensive investigation of future urban floods, this study provides much-needed
489 insights into urban flood management for similar urban areas in China, most of which are
490 equipped with highly insufficient drainage capacities. By comparing the reduction of flood
491 volume by climate change mitigation (via reduction of GHG emissions) and local adaptation (via
492 improvement of drainage systems), this study highlights the effectiveness of system adaptations
493 in reducing future flood volumes. This has important implications for the research community
494 and decision-makers involved in urban flood management. We emphasise the importance of
495 accounting for both global-scale climate change mitigation and local-scale adaptation in
496 assessing future climate impacts on urban flood volumes within a consistent framework.

497

498 **Acknowledgements**

499 This research was supported by the Natural Science Foundation of Guangdong Province, China
500 (No. 2014A030310121) and the Scientific Research Foundation for the Returned Overseas
501 Chinese Scholars, State Education Ministry. G. Leng and M. Huang were supported by the

502 Integrated Assessment Research program through the Integrated Multi-sector, Multi-scale
503 Modeling (IM³) Scientific Focus Area (SFA) sponsored by the Biological and Environmental
504 Research Division of Office of Science, U.S. Department of Energy. The Pacific Northwest
505 National Laboratory (PNNL) is operated for the U.S. DOE by Battelle Memorial Institute under
506 contract DE-AC05-76RL01830

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664 **Table 1** Infiltration parameters for three categories of soil in the SWMM simulation

Soil category	Infiltration parameters			
	MaxRate	MinRate	Decay rate	DryTime
	[in/hr]	[in/hr]	[1/hr]	[days]
Dry loam with little or no vegetation	3	0.5	4	7
Dry sand with little or no vegetation	5	0.7	5	5
Dry clay with little or no vegetation	1	0.3	3	9

665

666

667 **Table 2** Projected changes in precipitation intensity under return periods ranging from 1 year to 1000
 668 years by five Global Climate Models under two Representative Concentration Pathways (RCPs)

		1	2	3	10	20	50	100	200	500	1000
GFDL- ESM2M	RCP8.5	2.12	1.23	1.34	1.25	1.27	1.21	1.08	1.12	1.24	1.23
	RCP2.6	1.74	1.08	1.03	1.11	1.07	1.15	1.14	1.15	1.19	1.16
HadGE M2-ES	RCP8.5	0.62	1.08	1.09	1.06	1.01	1.03	1.17	1.26	1.23	1.14
	RCP2.6	0.36	1.2	1.19	1.04	1.02	1.11	1.31	1.26	1.37	1.24
IPSL- CM5A- LR	RCP8.5	1.44	1.17	1.28	1.17	1.08	1.09	1.02	1.1	1.12	1.13
	RCP2.6	0.74	1.04	1.18	1.01	1.06	1.03	1.01	0.99	0.95	1
MIROC -ESM- CHEM	RCP8.5	2.13	1.38	1.3	1.51	1.32	1.23	1.17	1.27	1.16	1.31
	RCP2.6	0.71	1.12	1.14	1.18	1.1	1.07	1.01	1.09	1.01	1.09
NorES M1-M	RCP8.5	2.11	0.96	0.8	1.63	1.35	1.15	1.08	1.01	1.04	0.97
	RCP2.6	0.11	1.09	1.05	1.28	1.17	1.08	1.1	1.18	1.09	1.2

669

670 **List of Figures**

671 **Figure 1** Land use of the study region for the year 2010 (a) and 2020 (b). Pipe network
672 description of current and planned drainage systems (c). Difference in Weighted Mean
673 Imperviousness (WMI) between year 2010 and 2020 (d).

674 **Figure 2** Illustration of flood volume and average total expected total flood volume (TFVs) as a
675 function of return period under a stationary drainage system. The grey area denotes the average
676 total expected TFVs per year considering all kinds of floods.

677 **Figure 3** Projected TFV with changes in precipitation intensity at various return periods under
678 the RCP8.5 scenario for the period of 2020–2040.

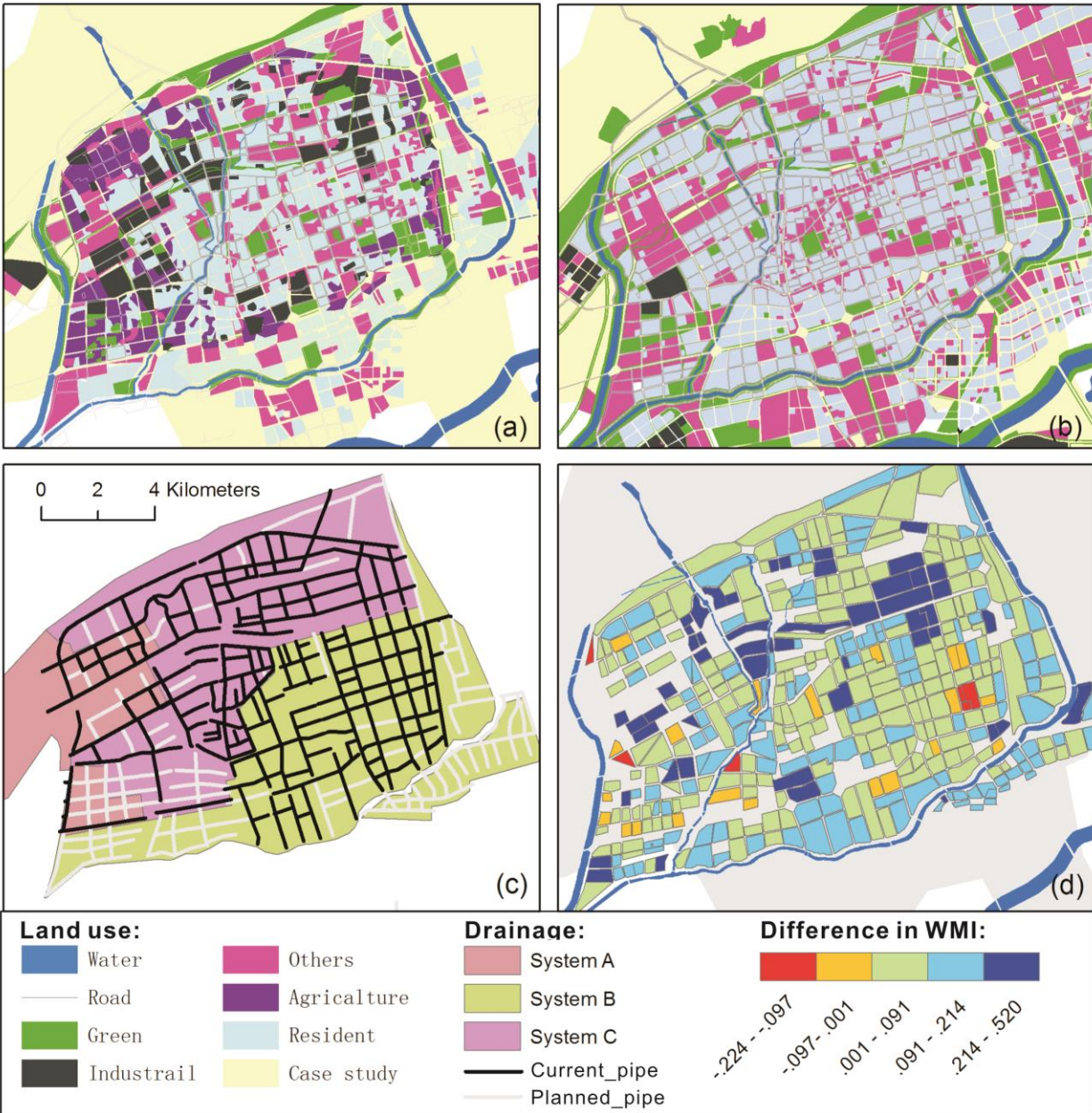
679 **Figure 4** Comparison of (a) flood volume, (b) total TFVs (i.e., the piece-wise integral of flood
680 volume versus the expected frequency with changes in precipitation intensity of various return
681 periods under RCP8.5 (blue) and RCP2.6 (red). (c) is for the reduced TFVs in percentage (i.e.,
682 benefits of climate mitigation) in RCP2.6 relative to RCP8.5 at various return periods.

683 **Figure 5** Spatial distribution of overloaded pipelines (red colour) induced by the 3-year (left
684 column) and 50-year extreme events (right column) without and with adaptations. The total
685 percentage of overloaded manholes (POM) and ratio of flood volume (RFV) are summarised for
686 each scenario. Descriptions of local land use, mainly the traffic network and green spaces, are
687 provided as the background image in (a).

688 **Figure 6** Future changes in flood volumes (CTFVs) relative to historical conditions under the
689 current drainage system (yellow) and two adaptation scenarios (i.e., Pipe in red and Pipe+LID in
690 green) at various return periods.

691 **Figure 7** Comparison of benefits of climate mitigation and two adaptation strategies in reducing
692 urban flood volumes with changes in precipitation intensities for various return periods, and with
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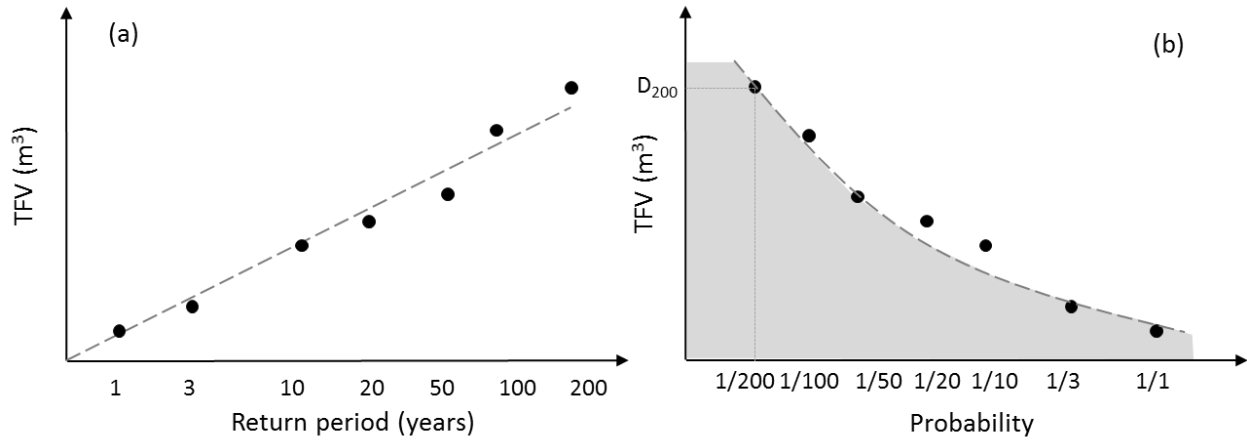
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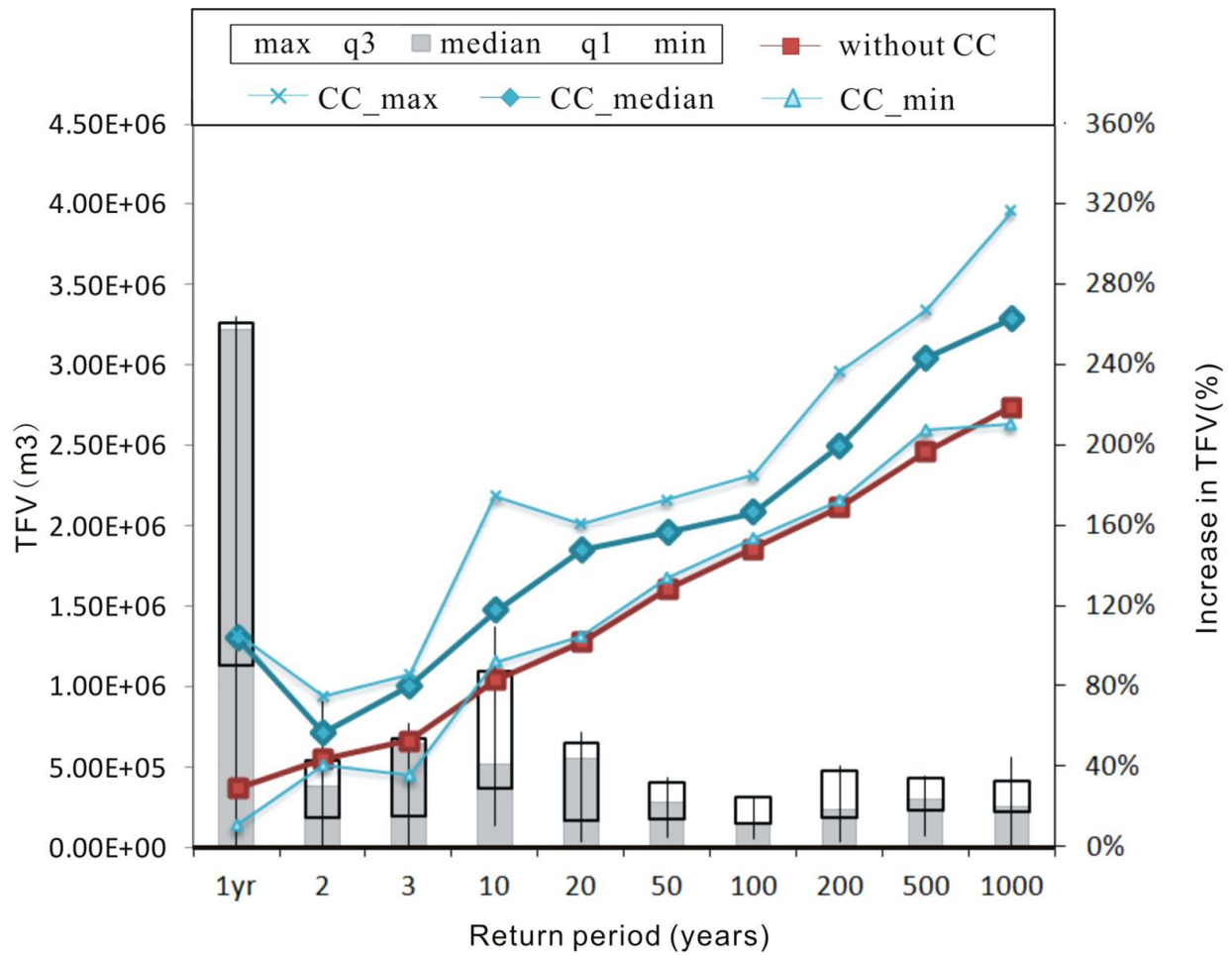
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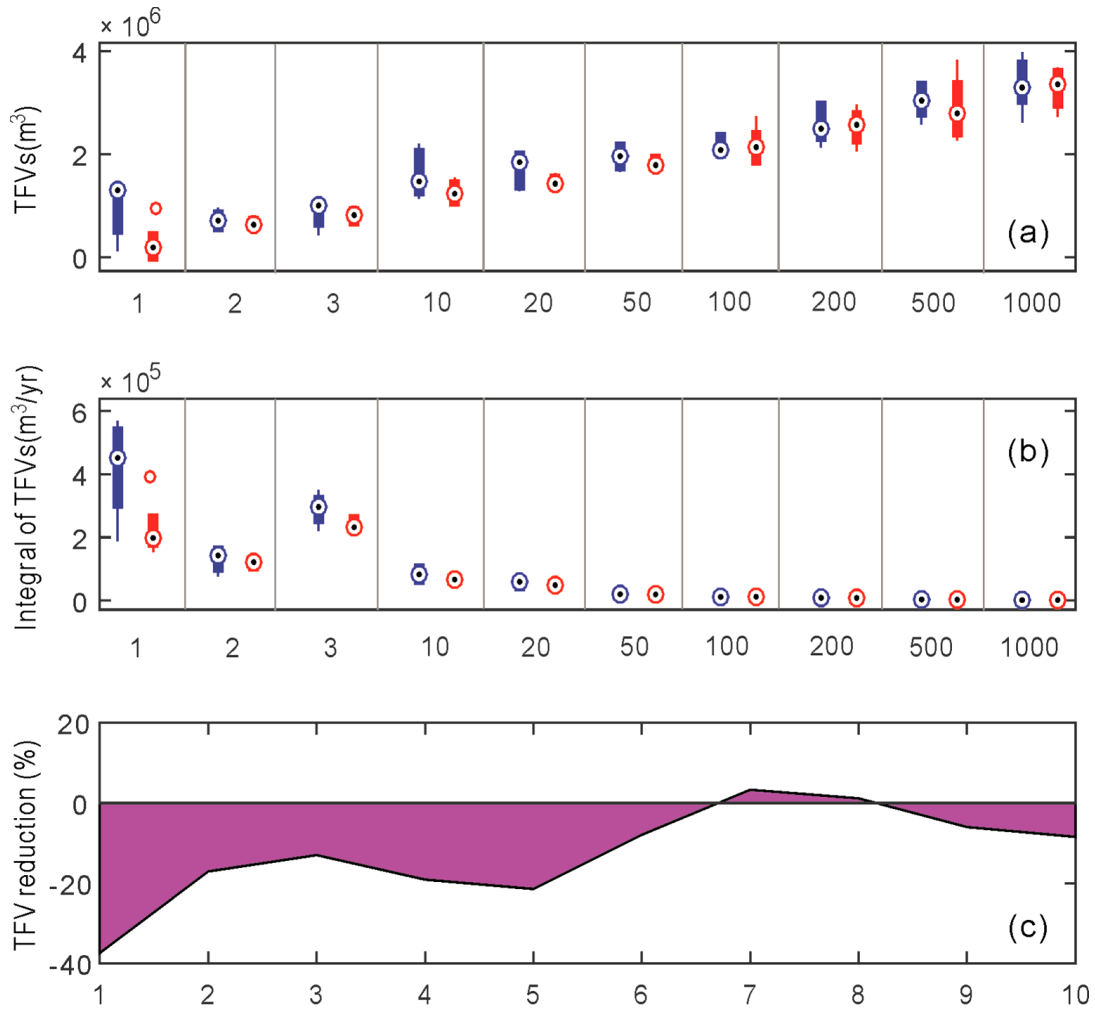
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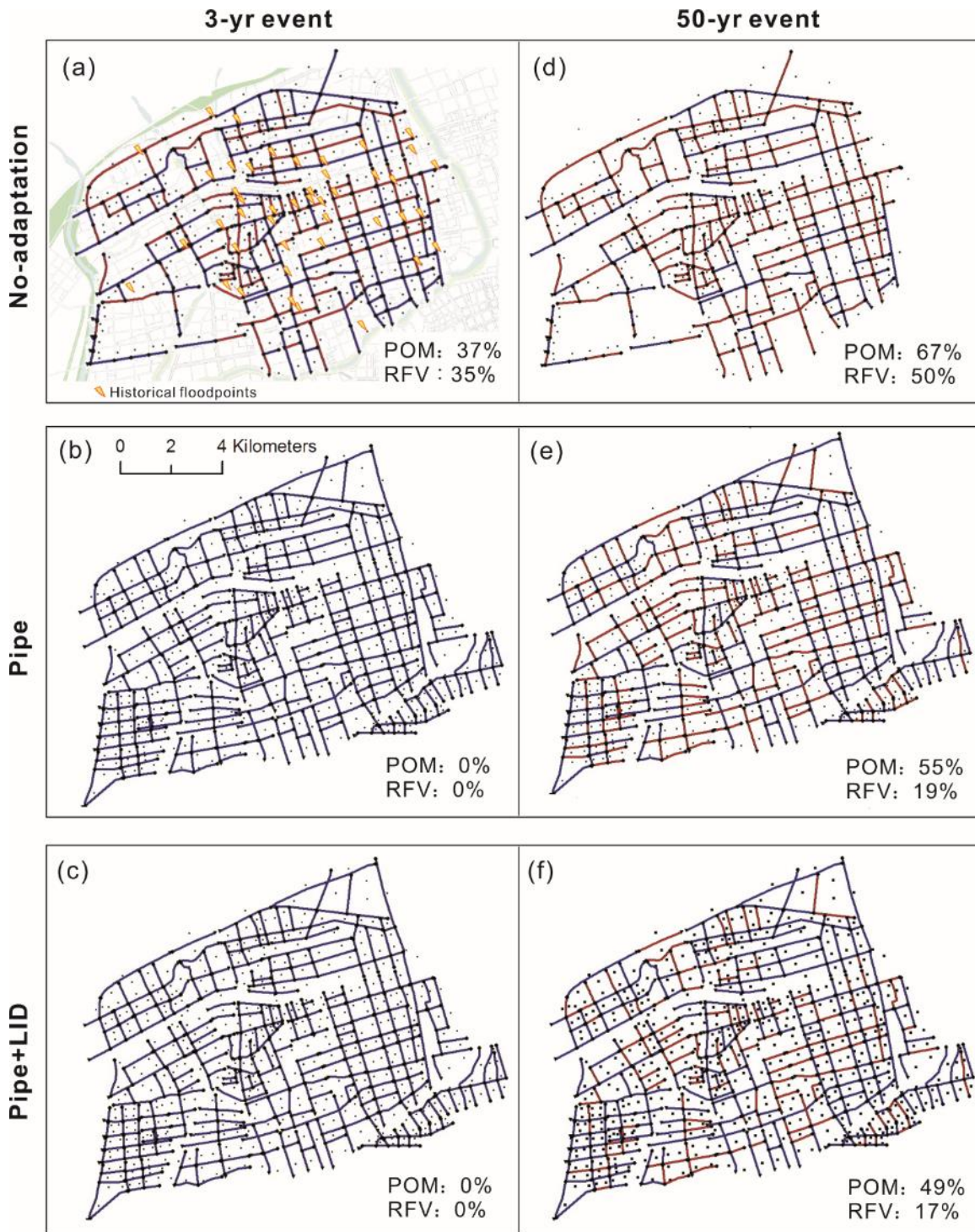
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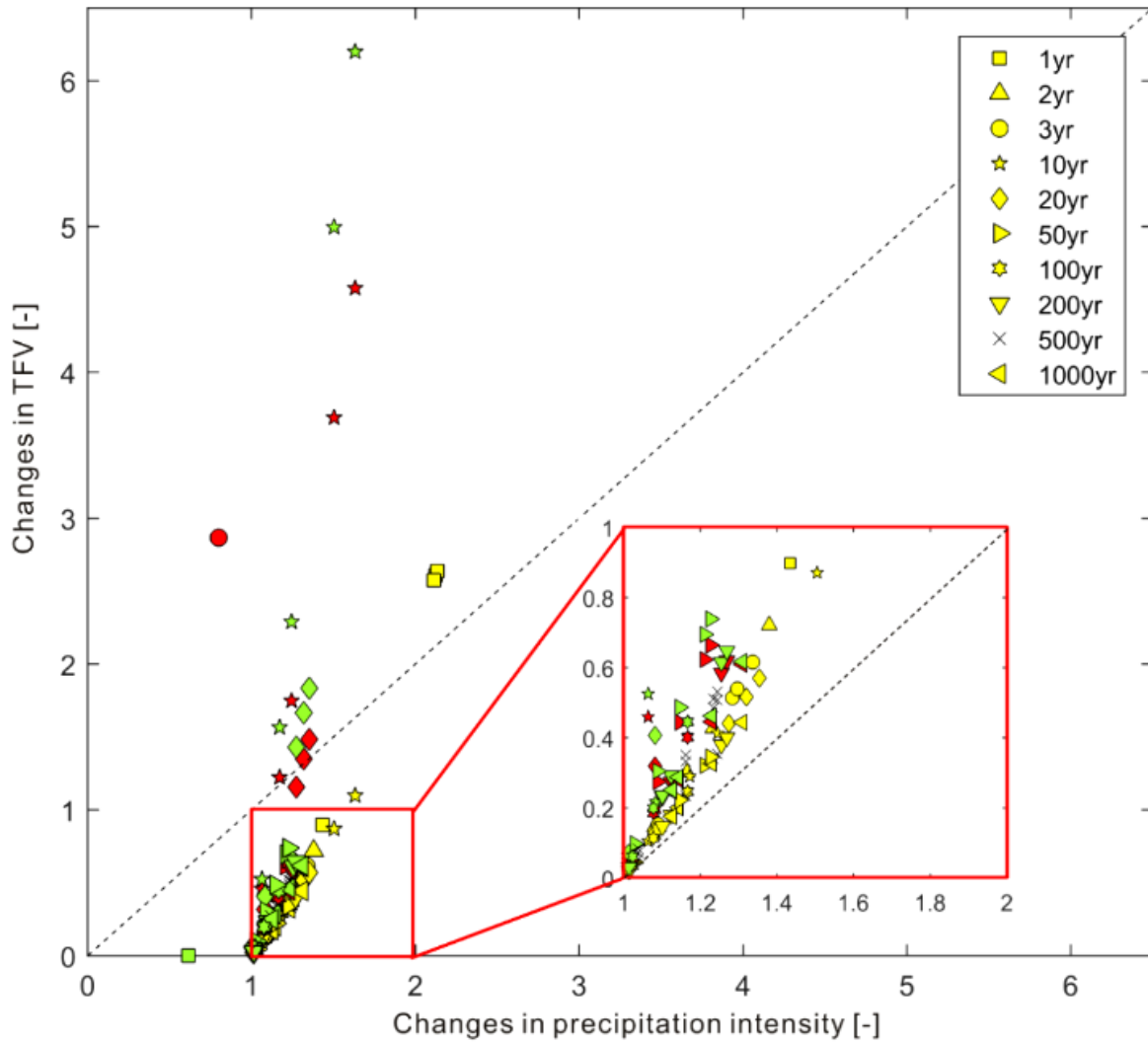
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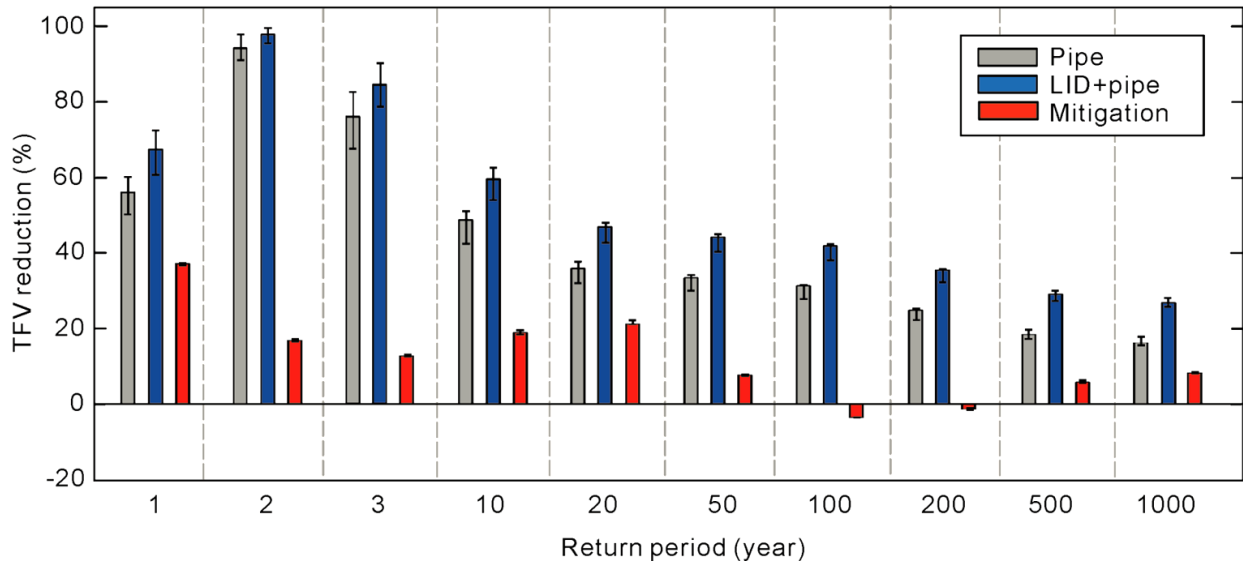
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