

1 **Towards a simple representation of chalk hydrology in land**

2 **surface modelling**

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6 **Abstract**

7 Modelling and monitoring of hydrological processes in the unsaturated zone of chalk, a
8 porous medium with fractures, is important to optimize water resources assessment and
9 management practices in the United Kingdom (UK). However, incorporating the processes
10 governing water movement through chalk unsaturated zone in a numerical model is
11 complicated mainly due to the fractured nature of chalk that creates high-velocity preferential
12 flow paths in the subsurface. In general, flow through chalk unsaturated zone is simulated
13 using dual-porosity concept, which often involves calibration of relatively large number of
14 model parameters, potentially undermining applications to large regions. In this study, a
15 simplified parameterization, namely the Bulk Conductivity (BC) model is proposed for
16 simulating hydrology in chalk unsaturated zone. This new parameterization introduces only
17 two additional parameters (namely the macroporosity factor and the soil wetness threshold
18 parameter for fracture flow activation) and uses the saturated hydraulic conductivity from
19 chalk matrix. The BC model is implemented in the Joint UK Land Environment Simulator
20 (JULES) and applied to a study area encompassing the Kennet catchment in the Southern
21 UK. This parameterization is further calibrated at a point-scale using soil moisture profile
22 observations. The performance of calibrated BC model in JULES is assessed and compared
23 against the performance of both the default JULES parameterization and the uncalibrated
24 version of BC model implemented in JULES. Finally, the model performance at the

25 catchment-scale is evaluated against independent data sets (e.g., runoff and latent heat flux).
26 The results demonstrate that the inclusion of the BC model in JULES improves simulated
27 land surface mass and energy fluxes over the chalk-dominated Kennet catchment. Therefore,
28 the simple approach described in this study may be used to incorporate the flow processes
29 through chalk unsaturated zone in large-scale land surface modelling applications.

30 **Keywords:** Chalk hydrology, macroporosity, land surface model, bulk conductivity model.

31 **1. Introduction**

32 Chalk can be described as a fine-grained porous medium traversed by fractures [*Price et al.*,
33 1993]. Previous studies showed that the unsaturated zone of the chalk aquifers plays an
34 important role on groundwater recharge in the UK [e.g., *Lee et al.*, 2006; *Ireson et al.*, 2009].
35 Therefore, both monitoring [e.g., *Bloomfield*, 1997; *Ireson et al.*, 2006] and modelling [e.g.,
36 *Bakopoulou*, 2015; *Brouyère*, 2006; *Ireson and Butler*, 2011, 2013; *Sorensen et al.*, 2014]
37 strategies have been adapted previously to understand the governing hydrological processes
38 in the chalk unsaturated zone.

39 In chalk, the matrix provides porosity and storage capacity, while the fractures greatly
40 enhance permeability [*Van den Daele et al.*, 2007]. Water movement through chalk matrix is
41 slow due to its relatively high porosity (0.3-0.4) and low permeability (10^{-9} - 10^{-8} ms⁻¹). A
42 fractured chalk system, in contrast, conducts water at a considerably higher velocity because
43 of relatively high permeability (10^{-5} - 10^{-3} ms⁻¹) and low porosity (of the order 10^{-4}) of
44 fractures [*Price et al.*, 1993].

45 Simulating water flow through the matrix-fracture system of chalk has been the subject of
46 research for some time. Both conceptual [e.g., *Price et al.*, 2000; *Haria et al.*, 2003] and
47 physics-based [e.g., *Mathias et al.*, 2006; *Ireson et al.*, 2009] models have been proposed
48 previously to describe water flow through chalk unsaturated zone. The physics-based models
49 mentioned above were developed based on dual-continua approach and required relatively

50 large numbers of parameters (i.e., on the order of 20-30 parameters) that were calibrated via
51 inverse modelling using observed soil moisture and matric potential data [e.g., *Ireson et al.*,
52 2009; *Mathias et al.*, 2006].

53 In recent years, representation of chalk has gained attention in land surface modelling. For
54 example, *Gascoign et al.* [2009] applied the Catchment Land Surface Model (CLSM) over the
55 Somme River basin in northern France. A linear reservoir was included in the TOPMODEL
56 based runoff formulation of CLSM to account for the contribution of chalk aquifers to river
57 discharge. *Le Vine et al.* [2016] applied the Joint UK Land Environment Simulator (JULES
58 [*Best et al.*, 2011]) over the Kennet catchment in southern England to evaluate the
59 hydrological limitations of land surface models. In that study, two intersecting Brooks and
60 Corey curves were proposed, which allowed a dual curve soil moisture retention
61 representation for the two distinct flow domains of chalk (i.e., matrix and fracture) in the
62 model. Considering this dual Brooks and Corey curve, a three-dimensional groundwater flow
63 model (ZOOMQ3D [*Jackson and Spink*, 2004]) was coupled to JULES to demonstrate the
64 strong influence of representing chalk hydrology and groundwater dynamics on simulated
65 soil moisture and runoff.

66 The above mentioned studies illustrate the importance of representing chalk in land surface
67 modelling. However, including chalk hydrology in large-scale land surface modelling using
68 the contemporary dual-porosity concept can be complicated due to large number of additional
69 parameters. In this context, we propose a new parameterization, namely the Bulk
70 Conductivity (BC) model as a first step towards a simple chalk representation suitable for
71 land surface modelling. In order to test the proposed parameterization, the BC model is
72 included in JULES (version 4.2), which, by default (i.e., uniform soil column representation
73 using general soil database as typically applied in land surface models), does not represent
74 any chalk feature. In this study, the BC model (included in JULES) is applied at two distinct

75 spatial scales (i.e., point and catchment). At the point-scale, the proposed parameterization is
 76 calibrated using observed soil moisture profile data. This is achieved by randomly sampling
 77 the parameter space and extensively running the model in order to minimize the differences
 78 between observed and simulated soil moisture variability at different depths. Finally, the
 79 proposed model is applied to the Kennet catchment in the Southern England and the fluxes
 80 and states of the hydrological cycle are simulated for multiple years. The simulation results
 81 are evaluated using observed latent heat flux (LE) and runoff data to assess the performance
 82 of the BC model in simulating land surface processes at the catchment scale.

83 **2. A model of flow through chalk unsaturated zone**

84 In this study, the *Bulk Conductivity* (BC) model based on the work by *Zehe et al.* [2001] is
 85 incorporated in JULES to represent the flow of water through the fractured chalk unsaturated
 86 zone. According to this approach, if the relative saturation (S) exceeds a certain threshold (S_0)
 87 at a soil grid, the saturated hydraulic conductivity of chalk matrix (K_s) is increased to a bulk
 88 saturated hydraulic conductivity (K_{sb}) as follows

$$89 \quad K_{sb} = K_s + K_s f_m \frac{S - S_0}{1 - S_0} \quad \text{if } S > S_0 \quad (1)$$

$$90 \quad K_{sb} = K_s \quad \text{if } S \leq S_0 \quad (2)$$

$$91 \quad \text{with } S = \frac{\theta - \theta_r}{\theta_s - \theta_r}$$

92 where f_m is a macroporosity factor (-), θ is soil moisture (m^3m^{-3}), θ_s is soil moisture at
 93 saturation (m^3m^{-3}) and θ_r is the residual soil moisture (m^3m^{-3}). Note that S ranges from zero in
 94 case of completely dry soils to one for fully wet soils.

95 At the first step of evaluation, the K_s , S_0 and f_m parameters are estimated based on existing
 96 literature to assess the performance of the uncalibrated BC model. For the matrix saturated
 97 hydraulic conductivity (K_s), we use $K_s = 1.0 \text{ mmd}^{-1}$ following *Mathias et al.* [2006]. In

98 addition, Equation 1 indicates that the onset of water flow through the fracture system of
99 chalk is controlled by the threshold S_0 . According to *Wellings and Bell* [1980], water flow
100 through fractures dominates over matrix flow in chalk when the pressure head in soil
101 becomes higher than -0.50 mH₂O. We consider a value of $S_0 = 0.80$ for the uncalibrated BC
102 model, which is based on observed soil moisture-matric potential relationship in the study
103 area.

104 Finally, In *Zehe et al.* [2001], f_m was defined as the ratio of the saturated water flow rate in all
105 macropores in a model element to the corresponding value in soil matrix, which can be
106 determined based on the density and length of fractures at small scales. In addition, f_m has
107 also been considered as a calibration parameter previously [e.g., *Blume, 2008; Zehe et al.,*
108 2013]. In this study, we define f_m as a characteristic soil property reflecting the influence of
109 fractures on soil water movement [*Zehe and Blöschl, 2004*] and estimate it from the relative
110 difference of permeability between chalk matrix and fractured chalk system that can be of the
111 order 10^4 - 10^6 according to *Price et al.* [1993]. Consequently, we consider a macroporosity
112 factor of $f_m = 10^5$ for the uncalibrated BC model. In the following step, the BC model is
113 calibrated to minimize the differences between the variability of observed and simulated soil
114 moisture at individual depths. The calibration strategy will be discussed elaborately in section
115 3.5.

116 **3. Methods**

117 **3.1. Study area**

118 The study area encompasses the Kennet catchment located in the Southern England with an
119 area of about 1033 km² (Figure 1a). Generally, Kennet is rural in nature with scattered
120 settlements and has a maximum altitude of approximately 297 m (Above Ordnance Level).
121 The River Kennet discharges into the North Sea through London. The major tributaries of

122 this river are Lambourn, Dun, Enborne, and Foudry Brook. An average annual rainfall of
123 approximately 760 mm was recorded in the catchment over a 40 year period from 1961-1990.
124 Solid geology of the Kennet catchment is dominated by chalk, which is overlain by thin soil
125 layer. While lower chalk outcrops along the northern catchment boundary, progressively
126 younger rocks are found in the southern part. In general, surface runoff production is very
127 limited over the regions of the catchment where chalk outcrops. The flow regime shows a
128 distinct characteristics of slow response to groundwater held within the chalk aquifer [*Le*
129 *Vine et al.*, 2016]. According to *Ireson and Butler* [2013], the unsaturated zone of chalk
130 shows slow drainage over summer and bypass flow during wet periods in this catchment.

131 **3.2. Field measurements and remotely sensed data**

132 Table 1 summarizes the field measurements and remote sensing data used in this study. We
133 use *in-situ* soil moisture and runoff measurements along with remotely sensed LE data to
134 assess model performance in simulating the mass and energy balance components of the
135 hydrological cycle. Point scale soil moisture measurements at two adjacent sites (~20 m
136 apart) at the Warren Farm (Figure 1) were provided by Centre for Ecology and Hydrology
137 (CEH). A Didcot neutron probe was used at these locations to measure fortnightly soil
138 moisture at different depths below land surface (10 cm apart down to 0.8 m, 20 cm apart
139 between 0.8-2.2 m, and 30 cm apart between 2.2-4.0 m) [*Hewitt et al.*, 2010].

140 The National River Flow Archive (NRFA) coordinates discharge measurements from the
141 gauging station networks across UK. These networks are operated by the Environmental
142 Agency (England), Natural Resources Wales, Scottish Environment Protection Agency, and
143 Rivers Agency (Northern Ireland). We use discharge measurements provided by NRFA to
144 assess model performance in simulating runoff over the Kennet catchment in this study.

145 The MOD16 product of the Moderate Resolution Imaging Spectroradiometer (MODIS) is a
146 part of NASA/EOS project that provides estimation of global terrestrial LE. The LE
147 estimation from MOD16 is based on remotely sensed land surface data [e.g., *Mu et al.*, 2007].
148 In this study, the 8-day and monthly LE data products from MODIS is used to evaluate the
149 model performance in simulating land surface energy fluxes.

150 **3.3. Land surface model**

151 In this study, we use the Joint UK Land Environment Simulator (JULES [e.g., *Best et al.*,
152 2011; *Clark et al.*, 2011]) version 4.2. JULES is a flexible modelling platform with a modular
153 structure aligned to various physical processes developed based on the Met Office Surface
154 Exchange Scheme (MOSES [e.g., *Cox et al.*, 1999; *Essery et al.*, 2003]). Meteorological data
155 including precipitation, incoming short- and longwave radiation, temperature, specific
156 humidity, surface pressure, and wind speed are required to drive JULES. Each grid box in
157 JULES can comprise nine surface types (broadleaf trees, needle leaf trees, C3 grass, C4 grass,
158 shrubs, inland water, bare soil, and ice) represented by respective fractional coverage. Each
159 surface type is represented by a tile and a separate energy balance is calculated for each tile.

160 Subsurface heat and water transport equations are solved based on finite difference
161 approximation in JULES as described in *Cox et al.* [1999]. Moisture transport in the
162 subsurface is described by the finite difference form of Richards' equation. The vertical soil
163 moisture flux is calculated using the Darcy's law. While the top boundary condition to solve
164 the Richards' equation is infiltration at soil surface, the bottom boundary condition in JULES
165 is free drainage that contributes to subsurface runoff.

166 Surface runoff is calculated by combining the equations of throughfall and grid box average
167 infiltration in JULES. In order to direct the generated runoff to a channel network, river
168 routing is implemented based on the discrete approximation of one-dimensional kinematic

169 wave equation [e.g., *Bell et al.*, 2007]. In this approach, river network is derived from the
170 digital elevation model (DEM) of the study area and different wave speeds are applied to
171 surface and subsurface runoff components and channel flows [e.g., *Bell and Moore*, 1998]. A
172 return flow term accounts for the transfer of water between subsurface and land surface [e.g.,
173 *Dadson et al.*, 2010, 2011].

174 **3.4. Model configurations and input data**

175 In this study, simulations are performed at two distinct spatial scales, namely point and
176 catchment. At the point scale, JULES is configured to simulate the mass and energy fluxes at
177 the Warren Farm site (Figure 1a). A total subsurface depth of 5 m is considered in the model
178 with a vertical discretization ranging from 10 cm at the land surface to 50 cm at the bottom of
179 the model domain. Note that this discretization is consistent with the soil moisture
180 measurement depths mentioned in section 3.2. The vegetation type is implemented as C3
181 grass using the default parameters in JULES. Point scale simulations were performed over 2
182 consecutive years from 2003-2005 at an hourly time step. Except for precipitation, hourly
183 atmospheric forcing data to drive JULES was obtained from an automatic weather station
184 operated by the CEH at Warren Farm. In order to estimate hourly precipitation data to run
185 JULES, rain gauge measurements from the Met Office [*Met Office*, 2006] were used. Inverse
186 distance interpolation technique [e.g. *Garcia et al.*, 2008; *Ly et al.*, 2013] was applied on
187 rainfall measurements from 13 gauges closest to Warren Farm (distance varies from 25-60
188 km) to obtain hourly precipitation for the point scale simulations.

189 At the catchment scale, JULES is configured over a study area encompassing the Kennet
190 catchment (Figure 1a) considering a uniform lateral grid resolution of 1 km with 70 x 40 cells
191 in x and y dimensions, respectively. The total subsurface depth and vertical discretization are
192 identical to those of the point scale simulations. Spatially distributed vegetation type

193 information for the study area (Figure 1b) is obtained from the Land Cover Map 2007
194 (LCM2007) dataset [Morton *et al.*, 2011]. Simulations were performed over 5 consecutive
195 years from 2006-2011 at the catchment scale. Note that the simulation periods of catchment
196 and point scale (2003-2005) do not coincide due to the availability of soil moisture
197 measurements described in section 3.2. Spatially distributed meteorological data from the
198 Climate, Hydrology and Ecology research Support System (CHESS) was used to obtain the
199 atmospheric forcing to drive JULES at the catchment scale. The CHESS data includes 1 km
200 resolution gridded daily meteorological variables [Robinson *et al.*, 2015]. This daily data is
201 downscaled using a disaggregation technique described in Williams and Clark [2014] to
202 obtain hourly atmospheric forcing. The flow direction required for river routing is extracted
203 from the USGS HydroSHEDS digital elevation data [Lehner *et al.*, 2008].

204 We estimate the soil hydraulic properties based on texture (Table 2). At the point scale, loam
205 soil is dominant at the Warren Farm site. At the catchment scale, the Harmonized World Soil
206 Database (HWSD) from the Food and Agricultural Organization of UNO (FAO) is used to
207 obtain the texture of different soil types over Kennet (Figure 1c). The saturation-pressure
208 head relationship for different soil types is described using the Van Genuchten [Van
209 Genuchten, 1980] model with parameter values (Table 2) obtained from Schaap and Leij
210 [1998].

211 The hydraulic properties for chalk used in this study are summarized in Table 3. These
212 properties are obtained based on existing literature as a first step when evaluating the
213 uncalibrated BC model. The BC model parameters are subsequently calibrated to minimize
214 the differences between observed and simulated $\Delta\theta$ (section 3.5) at various soil depths.

215 In this study, we consider two different model configurations, namely *default* and *macro*
216 (Figure 2). The *default* configuration corresponds to the standard parameterizations of JULES

217 that does not represent chalk hydrology in the model. In this configuration, each soil column
 218 in JULES is considered to be vertically homogeneous with the soil properties defined in
 219 Table 2, which is motivated by the Met Office JULES Global Land 4.0 configuration
 220 described in *Walters et al.* [2014]. The *macro* configuration, in contrast, explicitly represents
 221 chalk by applying the BC model starting at 30 cm below land surface to the bottom of the
 222 model domain (i.e. 500 cm). Therefore, the soil column in the *macro* configuration can be
 223 divided into topsoil (0-30 cm) and chalk (30-500 cm). The topsoil depth of 30 cm is defined
 224 based on several augured soil samples collected during a field campaign at Warren Farm in
 225 2015 (Figure 2). This depth is corroborated by additional information from the British
 226 Geological Survey (BGS) operated borehole records
 227 (http://www.ukso.org/pmm/soil_depth_samples_points.html), which show that topsoil depths
 228 vary from 10-40 cm over the study area. We apply the *macro* configuration assuming a
 229 spatially homogeneous topsoil depth of 30 cm for both point and catchment scale simulations.
 230 Note that except for this inclusion of chalk, *default* and *macro* configurations are identical in
 231 terms of model set up and input data. It should also be emphasized that *default* represents a
 232 “naïve” configuration deprived of model calibration. Moreover, this configuration does not
 233 represent chalk, which, according to previous studies [e.g., *Le Vine et al.*, 2016], substantially
 234 affects the hydrology of the study area considered here.

235 **3.5. Calibration of the BC model**

236 We calibrate the BC model at the point-scale to minimize the differences between observed
 237 and simulated soil moisture variability ($\Delta\theta$) at different depths. The Root Mean Squared
 238 Error (RMSE) is used as the objective function to optimize the BC model parameters [e.g.,
 239 *Ireson et al.*, 2009]

$$240 \quad RMSE = \frac{1}{nd} \sum_1^{nd} \sqrt{\left(\frac{1}{nt-1} \sum_2^{nt} (\Delta\theta_{d,t}^{obs} - \Delta\theta_{d,t}^{sim})^2 \right)} \quad (3)$$

241 where nd is the number of soil layers, nt is the number of soil moisture observations available
242 for a layer d , $\Delta\theta^{obs}$ is the observed variability of soil moisture and $\Delta\theta^{sim}$ is the simulated
243 variability of soil moisture. Note that we consider $\Delta\theta$ for this optimization because of its
244 relevance to the water flux and recharge through chalk unsaturated zone [e.g., *Ireson and*
245 *Butler*, 2011].

246 Equation (1) reveals that the calibration of the BC model involves optimizing 3 parameters,
247 namely the saturated hydraulic conductivity of chalk matrix (K_s), saturation threshold (S_0)
248 and macroporosity factor (f_m). *Ireson et al.* [2009] suggested a range of 0.2-2.0 mmd^{-1} for K_s .
249 On the other hand, *Price et al.* [1993] argued that in general, K_s is around 3-5 mmd^{-1} for most
250 chalk soils. Therefore, we consider a range of 0.2-5.0 mmd^{-1} in optimizing K_s . We consider S_0
251 range 0-1, representing the entire physical domain for soil wetness from fully dry to fully
252 wet, respectively. For f_m , a range of 10^4 - 10^6 is considered, which, as discussed earlier, is
253 consistent with the relative difference between the permeability of fractured chalk and chalk
254 matrix according to *Price et al* [1993]. Latin hypercube sampling technique [e.g., *McKay et*
255 *al.*, 2016] is used to generate 2,000 random samples for each BC model parameter within the
256 ranges discussed above. Note that for the K_s parameter, the random sampling was performed
257 from a logarithmic distribution [*Ireson et al.*, 2009]. We perform simulations using these
258 random samples and calculate model performance (Equation 3) to select the optimum
259 parameter values for the BC model for each possible parameter combination as discussed in
260 details in the following section.

261 **4. Results and discussion**

262 **4.1. Point scale simulations**

263 At the point scale, the simulation results are evaluated using soil moisture observations at the
264 Warren Farm site. Figure 3a compares observed and simulated soil moisture (θ) from the

265 *default* and *macro* configurations at 2 m below land surface. Note that the *macro*
266 configuration uses the chalk hydraulic parameters collected from existing literature (Table 3).
267 This figure shows that the *default* configuration considerably underestimates θ throughout the
268 simulation period, which is improved remarkably in case of *macro*. Figure 3b plots observed
269 and simulated soil moisture variability ($\Delta\theta$) from the *default* and *macro* configurations
270 ($\Delta\theta_{default}$ and $\Delta\theta_{macro}$, respectively) at the Warren Farm site. In general, both configurations
271 show discrepancies with observed $\Delta\theta$ with *macro* showing relatively better model
272 performance.

273 The results show that despite the *macro* configuration improves simulated θ , it shows
274 considerable discrepancies with observed $\Delta\theta$, which is consistent throughout the whole chalk
275 profile (results from other model layers are not shown). In order to minimize the differences
276 between observed and modelled $\Delta\theta$ from the *macro* configuration, we calibrate the BC model
277 following the methodology described in section 3.5. The optimization results are summarized
278 in Figure 4. Note that for each combination considered in the optimization, 2,000 model runs
279 were performed using randomly sampled parameters as discussed in section 3.5. In addition
280 to the *default* and *macro* cases, the calibrated cases in Figure 4 correspond to the results from
281 the model runs yielding the lowest RMSE for each parameter combination evaluated.

282 The RMSE between observed and simulated $\Delta\theta$ for the model configurations considered in
283 the optimization is shown in Figure 4a. This figure illustrates that the RMSE of the *default*
284 configuration is larger than that of *macro*, indicating better model performance in
285 reproducing $\Delta\theta$ for the latter (corresponding to a reduction of 15% in RMSE compared to the
286 *default* case). Therefore, the uncalibrated BC model (i.e., *macro* configuration) better
287 reproduces the soil moisture variability compared to *default*. Concerning the calibration of
288 single BC model parameters, Figure 4a shows that S_0 results in a 46% reduction of RMSE

289 compared to the *macro* configuration. Calibrating K_s or f_m individually yields only about 25%
290 reduction of RMSE compared to *macro*.

291 Optimizing both K_s and S_0 simultaneously shows the largest reduction (50%) of RMSE
292 compared to *macro* which coincides with the total RMSE reduction found when all
293 parameters are calibrated. Arguably, the BC model can be implemented in other chalk
294 regions by constraining only S_0 parameter. Such result could potentially be advantageous for
295 transferability to other regions in the UK in order to assess chalk hydrology at large-scale.
296 However since this is the first time the BC model is introduced, we decide to take a
297 conservative approach and select the *macro* configuration with optimized K_s and S_0 (*macro_{opt}*
298 hereafter) to simulate chalk hydrology over the study area that ensures best overall model
299 performance.

300 The lower three panels in Figure 4 presents the BC model parameter values for the *default*
301 and uncalibrated *macro* cases as well as for different combinations of parameters calibrated.
302 The red bars in Figure 4a, b and c highlight the cases in which a given parameter is
303 constrained by optimization. In those cases, the calibrated parameter values are obtained from
304 model runs producing the lowest RMSE. An interesting feature in Figure 4b (calibrating K_s
305 individually) is that the optimization suggests a compensation mechanism in which K_s is
306 increased remarkably in order to physically represent the “effective” flow through the chalk
307 fractures in the BC model. This is not surprising and arguably the simplest way to attempt to
308 improve model performance. For *macro_{opt}*, the values used for K_s is relatively lower than that
309 of uncalibrated *macro* case nevertheless consistent with previous estimates [e.g., Ireson *et al.*,
310 2009]. Figure 4c clearly shows the dominance of S_0 in the BC model as all the relatively low
311 RMSE bars in Figure 4a are associated with S_0 calibration (see red bars in Figure 4c). In
312 addition, calibrated S_0 values for all cases show a consistent constraint around 0.50. Finally,
313 Figure 4d indicates the lack of influence for f_m parameter on model performance.

314 Figure 5 compares $\Delta\theta_{default}$, $\Delta\theta_{macro}$ and $\Delta\theta$ from the *macro_{opt}* configuration ($\Delta\theta_{opt}$) with
315 observed soil moisture variability ($\Delta\theta_{obs}$). As mentioned earlier, $\Delta\theta_{default}$ and $\Delta\theta_{macro}$ show
316 considerable discrepancies with $\Delta\theta_{obs}$ while the *macro* configuration exhibits relatively better
317 performance (Figure 3). Figure 5 illustrates that the overall agreement between observed and
318 simulated $\Delta\theta$ improves substantially in case of *macro_{opt}* compared to *default* and *macro*,
319 which is pronounced especially in the deeper chalk layers. Therefore, this figure indicates
320 that the performance of the BC model in simulating $\Delta\theta$ is further improved by optimizing the
321 K_s and S_0 parameters simultaneously at the Warren Farm site.

322 As mentioned earlier, efficiently reproducing soil moisture variability over the profile is
323 important due to the fact that $\Delta\theta$ significantly affects water flux and recharge through chalk
324 unsaturated zone. The drainage flux through the bottom of soil column (d_b) of a land surface
325 model can be considered as the potential recharge flux to groundwater [e.g., *Sorensen et al.*,
326 2014]. Figure 6 compares the daily sum of d_b from the *default* and *macro_{opt}* configurations at
327 the Warren Farm site. Daily rainfall at this site over the simulation period is shown in Figure
328 6a. In Figure 6b, the *macro_{opt}* configuration shows considerable d_b during the colder months,
329 while relatively slow drainage is observed in summer. In contrast, the *default* configuration
330 shows relatively high d_b in summer compared to the colder months. In general, the recharge
331 rate through chalk unsaturated zone during the warmer periods of the year is lower than that
332 in the winter months [*Wellings and Bell*, 1980; *Ireson et al.*, 2009]. Therefore, the *macro_{opt}*
333 configuration appears to be more consistent with the recharge mechanism in chalk compared
334 to *default*.

335 In this section, the BC model was evaluated at the point scale. The results showed that in
336 general, the *macro* configuration outperforms the *default* case in simulating $\Delta\theta$. In order to
337 improve the model performance even further, model parameter calibration was performed to

338 minimize the differences between observed and simulated $\Delta\theta$ at the point scale. In the next
339 sections, the optimized model (*macro_{opt}*) is evaluated at the catchment scale.

340 **4.2. Catchment scale simulations**

341 At the catchment scale, simulation results from the *default* and *macro_{opt}* configurations are
342 compared with the observations over the Kennet catchment. In order to assess the differences
343 between LE from the *default* and *macro_{opt}* configurations at the catchment scale, Figure 7
344 plots spatially averaged 8-day composites of LE from MODIS (LE_{MOD}) against the LE from
345 these configurations ($LE_{default}$ and LE_{opt} , respectively) over Kennet. The agreement between
346 simulated LE and LE_{MOD} is evaluated using the coefficient of determination (R^2 , see
347 Appendix) and mean bias. Comparison between $LE_{default}$ and LE_{MOD} shows a coefficient of
348 determination of $R^2_{default} = 0.78$ and a mean bias of $bias_{default} = 10.5 \text{ Wm}^{-2}$. The agreement
349 between simulated LE and LE_{MOD} improves in case of the *macro_{opt}* configuration, which is
350 reflected by an increased coefficient of determination of $R^2_{opt} = 0.80$ and a reduced mean bias
351 of $bias_{opt} = 7.1 \text{ Wm}^{-2}$.

352 Figure 7 shows considerable differences between $LE_{default}$ and LE_{opt} for relatively high LE,
353 indicating discrepancies especially during the warmer months of the year. Spatially averaged
354 time series of monthly LE_{MOD} , $LE_{default}$ and LE_{opt} is presented in Figure 8a. This figure shows
355 that the differences between $LE_{default}$ and LE_{opt} increases in summer compared to the colder
356 months of the year, which is consistent with Figure 7. Consequently, the *default* configuration
357 underestimates LE in summer compared to LE_{MOD} , which is improved in case of the *macro_{opt}*
358 configuration. In contrast, the differences between $LE_{default}$ and LE_{opt} are negligible during the
359 colder months of the year.

360 In addition, Figure 8b compares the observed and simulated monthly average discharge from
361 the two model configurations at the “Kennet at Theale” gauging station (Figure 1a). This

362 figure depicts that the *default* configuration generally overestimates discharge at this gauging
363 station, which is improved considerably in the case of *macro_{opt}*. We use the Kling-Gupta
364 Efficiency criterion (KGE [Gupta *et al.*, 2009]) to compare the performance of the two model
365 configurations in reproducing observed discharge variability. As mentioned above, the
366 *default* configuration overestimates discharge with $KGE_{\text{default}} = -0.17$. On the other hand, the
367 *macro_{opt}* configuration improves the agreement between observed and simulated discharge,
368 which is reflected by $KGE_{\text{opt}} = 0.51$.

369 In order to summarize the results at catchment scale, Table 4 compares observed and
370 simulated runoff from the two model configurations over the Kennet catchment from 2006-
371 2011. The runoff ratio (*RR*, see Appendix), which is equal to the mean volume of flow
372 divided by the volume of precipitation [e.g., Kelleher *et al.*, 2015], assesses the partitioning
373 of precipitation into runoff over the catchment. The *default* configuration ($RR = 0.82$) shows
374 considerably higher *RR* compared to observation ($RR = 0.40$), indicating overestimation of
375 runoff by the model that is consistent with Figure 8b. Including chalk hydrology in the model
376 remarkably improves the agreement between observed and simulated mean runoff over the
377 Kennet catchment, which is assessed from a runoff ratio of $RR = 0.46$ for the *macro_{opt}*
378 configuration which is much closer to the observed *RR* value than *default*.

379 In Table 4, the relative bias ($\Delta\mu$) of 1.04 between observed and simulated runoff from the
380 *default* configuration again indicates the overestimation by the model. In comparison,
381 *macro_{opt}* shows a smaller relative bias of $\Delta\mu = 0.12$, indicating improved agreement between
382 observed and simulated mean runoff volume compared to *default*. The relative difference in
383 standard deviation ($\Delta\sigma$, see Appendix) compares the variability of observed and simulated
384 flow in Table 4 relating directly to the seasonal change in runoff. This comparison shows that
385 the *default* configuration overestimates the variability of runoff over the Kennet catchment

386 ($\Delta\sigma = 2.04$), which is improved in case of *macro* ($\Delta\sigma = 0.65$). This improvement in
387 reproducing flow variability is also clearly observed in Figure 8b.

388 In this section, the BC model is evaluated using observed mass and energy fluxes over the
389 Kennet catchment. The *default* configuration suggested relatively low summertime LE over
390 the catchment. The agreement between observed and simulated LE was improved in case of
391 the *macro_{opt}* configuration compared to *default*. It was also observed that the overall runoff
392 prediction was considerably improved by *macro_{opt}* compared to *default*. Given its simplicity,
393 our results indicate that the proposed parameterization is suitable for use in land surface
394 modelling applications.

395 **5. Summary and Conclusions**

396 In this study, we proposed a simple parameterization, namely the *Bulk Conductivity* (BC)
397 model to simulate water flow through the matrix-fracture system of chalk in large scale land
398 surface modelling applications. This parameterization was implemented in the Joint UK Land
399 Environment Simulator (JULES) and applied to the Kennet catchment located in the southern
400 UK to simulate the mass and energy fluxes of the hydrological cycle for multiple years. Two
401 model configurations, namely *default* and *macro* were considered with the latter using the BC
402 model to simulate chalk hydrology.

403 The proposed BC model is a single continuum approach of modelling preferential flow [e.g.,
404 *Beven and Germann, 2013*] that involves only 3 parameters, namely the saturated hydraulic
405 conductivity of chalk matrix (K_s), macroporosity factor (f_m) and relative saturation threshold
406 (S_0). Initially, these parameters were estimated from existing literature to assess the
407 performance of the uncalibrated BC model. Finally, the BC model parameters were optimized
408 to minimize the differences between observed and simulated soil moisture variability. Our
409 results indicated that S_0 is by far the most influential parameter in the model when

410 representing water movement through a soil-chalk column. This highlights the simplicity of
411 the proposed BC model for large-scale studies and potential ease in transferability. In
412 comparison, K_s and f_m showed secondary (low) sensitivity on the model performance. Since
413 this study introduces the BC model, we decided however to take a conservative approach.
414 We optimized K_s and S_0 simultaneously for our catchment scale simulations since this
415 combination resulted in the best overall model performance.

416 At the catchment scale, the proposed BC parameterization improved simulated latent heat
417 flux (especially in summer) and the overall runoff compared to *default*. Note that the
418 complexity (i.e., number of parameters) of the BC model for simulating water flow through
419 chalk unsaturated zone is substantially lower compared to more commonly used models for
420 this purpose (e.g., dual-porosity models). Despite its simplicity, the proposed
421 parameterization considerably improves the key hydrological fluxes simulated by JULES at
422 the catchment-scale. Therefore, the BC model can potentially be useful for land surface
423 modelling applications over large-scale chalk-dominated areas.

424 **Acknowledgements**

425 We gratefully acknowledge the support by the “A Multi-scale Soil moisture
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429 providing the data for the point-scale analyses at the Warren Farm. We would also like to
430 thank Miguel Rico-Ramirez (University of Bristol) for helping preparing the precipitation
431 data from the rain gauge network used for the point-scale simulations, Thorsten Wagener
432 (University of Bristol) for his valuable suggestions on model diagnostics, and Joost Iwema
433 (University of Bristol) for helping with the soil samples collected during the 2015 field work

434 campaign. Finally, we would like to thank the reviewers for their comments and suggestions
435 that added to the quality of this manuscript.

436 **Appendix**

437 **Definition of Statistical Metrics**

438 Coefficient of determination (R^2) for observation $y = y_1, \dots, y_n$ and prediction $f = f_1, \dots, f_n$
439 is defined as

$$440 \quad R^2 = 1 - \frac{SS_{res}}{SS_{tot}}$$

441 where SS_{res} is the residual sum of square and SS_{tot} is the total sum of square. SS_{res} and SS_{tot}
442 are defined as

$$443 \quad SS_{res} = \sum_{i=1}^n (y_i - f_i)^2 \quad \text{and}$$

$$444 \quad SS_{tot} = \sum_{i=1}^n (y_i - \bar{y})^2 \quad \text{with } \bar{y} \text{ being the mean of } y.$$

445 Runoff ratio (RR) assesses the portion of precipitation that generates runoff over the
446 catchment. RR is defined as

$$447 \quad RR = \frac{\mu_{runoff}}{\mu_{rain}}$$

448 where μ_{runoff} is mean runoff and μ_{rain} is mean precipitation [e.g., *Kelleher et al.*, 2015].

449 Relative bias ($\Delta\mu$) between observed and simulated time series can be defined as

$$450 \quad \Delta\mu = \frac{\mu_{mod} - \mu_{obs}}{\mu_{obs}}$$

451 where μ_{obs} and μ_{mod} are the mean of observed and simulated time series, respectively. While
452 the optimal value of $\Delta\mu$ is zero, negative (positive) values indicate an underestimation
453 (overestimation) by the model [e.g., *Gudmundsson et al.*, 2012].

454 Relative difference in standard deviation ($\Delta\sigma$) between observed and simulated time series
455 can be defined as

$$456 \quad \Delta\sigma = \frac{\sigma_{mod} - \sigma_{obs}}{\sigma_{obs}}$$

457 where σ_{obs} and σ_{mod} are the standard deviation of observed and simulated time series,
458 respectively [e.g., *Gudmundsson et al.*, 2012].

459

460 **References**

461 Bakopoulou, C. (2015), Critical assessment of structure and parameterization of JULES land
462 surface model at different spatial scales in a UK Chalk catchment, PhD thesis, Imperial
463 College London, UK, available at: <https://spiral.imperial.ac.uk:8443/handle/10044/1/28955>.

464 Bell, V. A. and R. J. Moore (1998), A grid-based flood forecasting model for use with
465 weather radar data: Part 1. Formulation, *Hydrol. Earth Syst. Sc.*, 2, 265-281.

466 Bell, V. A., A. L. Key, R. G. Jones, and R. J. Moore (2007), Development of a high
467 resolution grid-based river flow model for use with regional climate model output, *Hydrol.*
468 *Earth Syst. Sc.*, 11, 532-549.

469 Best, M. J., M. Pryor, D. B. Clark, G. G. Rooney, R. I. H. Essery, C. B. Ménard, J. M.

470 Edwards, M. A. Hendry, A. Porson, N. Gedney, L. M. Mercado, S. Sitch, E. Blyth, O.

471 Boucher, P. M. Cox, C. S. B. Grimmond, and R. J. Harding (2011), The Joint UK Land

472 Environment Simulator (JULES), Model Description – Part 1: Energy and Water

473 Fluxes, *Geosci. Model Dev.*, 4, 677-699.

474 Beven, K., and P. Germann (2013), Macropores and water flow in soils revisited, *Water*

475 *Resour. Res.*, 49, 3071–3092.

476 Bloomfield, J. (1997), The role of diagenesis in the hydrogeological stratification of
477 carbonated aquifers: An example from the chalk at Fair Cross, Berkshire, UK, *Hydrol. Earth*
478 *Syst. Sc.*, 1, 19-33.

479 Blume, T. (2008), Hydrological processes in volcanic ash soils: measuring, modelling and
480 understanding runoff generation in an undisturbed catchment, PhD thesis, Institut für
481 Geoökologie, Universität Potsdam, Potsdam, Germany, available at: [https://publishup.uni-](https://publishup.uni-potsdam.de/opus4-ubp/files/1524/blume_diss.pdf)
482 [potsdam.de/opus4-ubp/files/1524/blume_diss.pdf](https://publishup.uni-potsdam.de/opus4-ubp/files/1524/blume_diss.pdf)

483 Brouyère, S. (2006), Modelling the migration of contaminants through variably saturated
484 dual-porosity, dual-permeability chalk, *J. Contam. Hydrol.*, 82, 195-219.

485 Clark, D. B., L. M. Mercado, S. Sitch, C. D. Jones, N. Gedney, M. J. Best, M. Pryor, G. G.
486 Rooney, R. L. H. Essery, E. Blyth, O. Boucher, R. J. Harding, C. Huntingford, and P. M. Cox
487 (2011), The Joint UK Land Environment Simulator (JULES), Model Description – Part 2:
488 Carbon Fluxes and Vegetation Dynamics, *Geosci. Model Dev.* 4, 701-722.

489 Cox, P. M., R. A. Betts, C. B. Bunton, R. L. H. Essery, P. R. Rowntree and J. Smith (1999),
490 The impact of new land surface physics on the GCM simulation of climate and climate
491 sensitivity, *Clim. Dynam.*, 15, 183-203.

492 Dadson, S. J., I. Ashpole, P. Harris, H. N. Davies, D. B. Clark, E. Blyth, and C. M. Taylor
493 (2010), Wetland inundation dynamics in a model of land surface climate: Evaluation in the
494 Niger inland delta region, *J. Geophys. Res.*, 115.

495 Dadson, S. J., V. A. Bell, and R. G. Jones (2011), Evaluation of a grid based river flow model
496 configured for use in a regional climate model, *J. Hydrol.*, 411, 238-250.

497 Dettinger, M. D., and H. F. Diaz (2000), Global characteristics of streamflow seasonality and
498 variability, *J. Hydrometeorol.*, 1, 289-310.

499 Essery, R., M. Best, and P. Cox (2001), MOSES 2.2 technical documentation (Hadley Centre
500 technical note 30), Hadley Centre, Met Office, UK.

501 Garcia, M., C. D. Peters-Lidard, and D. C. Goodrich (2008), Spatial interpolation of
502 precipitation in a dense gauge network for monsoon storm events in the southwestern United
503 States, *Water Resour. Res.*, 44.

504 Gascoin, S., A. Duchare, P. Ribstein, M. Carli, and F. Habtes (2000), Adaptation of a
505 catchment-based land surface model to the hydrological setting of the Somme River basin
506 (France), *J. Hydrol.*, 368, 105-116.

507 Gudmundsson, L., T. Wagener, L. M. Tallaksen, and K. Engeland (2012), Evaluation of nine
508 large-scale hydrological models with respect to the seasonal runoff climatology in Europe,
509 *Water Resour. Res.*, 48.

510 Gupta, H. V., H. Kling, K. K. Yilmaz, and G. F. Martinez (2009), Decomposition of the mean
511 squared error and NSE performance criteria: implications for improving hydrological
512 modelling, *J. Hydrol.*, 377, 80-91.

513 Haria, A. H., M. G. Hodnett, and A. C. Johnson (2003), Mechanisms of groundwater
514 recharge and pesticide penetration to chalk aquifer in southern England, *J. Hydrol.*, 275, 122-
515 137.

516 Hartmann, A., N. Goldscheider, T. Wagener, J. Lange, and M. Weiler (2014), Karst water
517 resources in a changing world: Review of hydrological modeling approaches, *Rev. Geophys.*,
518 52, 218–242, doi:10.1002/2013RG000443.

519 Hewitt, N., M. Robinson, and D. McNeil (2010), Pang and Lambourn hydrometric review
520 2009, Wallingford, NERC/Centre for Ecology and Hydrology, (CEH project number:
521 C04076).

522 Ireson, A. M., S. A. Mathias, H. S. Wheater, A. P. Butler and J. Finch (2009), A model for
523 flow in the chalk unsaturated zone incorporating progressive weathering, *J. Hydrol.*, 365,
524 244-260.

525 Ireson, A. M. and A. P. Butler (2011), Controls on preferential recharge to chalk aquifers, *J.*
526 *Hydrol.*, 398, 109-123.

527 Ireson, A. M. and A. P. Butler (2013), A critical assessment of simple recharge models:
528 application to the UK chalk, *Hydrol. Earth Syst. Sc.*, 17, 2083-2096.

529 Ireson, A. M., H. S. Wheater, A. P. Butler, S. A. Mathias, J. Finch, and J. D. Cooper (2006),
530 Hydrological processes in the chalk unsaturated zone – insight from an intensive field
531 monitoring program, *J. Hydrol.*, 330, 29-43.

532 Ireson, A. M., S. A. Mathias, H. S. Wheater, A. P. Butler, and J. Finch (2009), A model for
533 flow in the chalk unsaturated zone incorporating progressive weathering, *J. Hydrol.*, 365,
534 244-260.

535 Jackson, C. and Spink, A. (2004) User's Manual for the Groundwater Flow Model
536 ZOOMQ3D, IR/04/140, British Geological Survey, Nottingham, UK.

537 Kelleher, C., T. Wagener, and B. McGlynn (2015), Model-based analysis of the influence of
538 catchment properties on hydrologic partitioning across five mountain headwater
539 subcatchments, *Water Resour. Res.*, 51, 4109-4136.

540 Kling, H., M. Fuchs, and M. Paulin (2012), Runoff conditions in the upper Danube basin
541 under an ensemble of climate change scenarios. *Journal of Hydrology*, Volumes 424-425, 6
542 March 2012, 264-277.

543 Lehner, B., K. Verdin, and A. Jarvis (2008), New global hydrography derives from
544 spaceborne elevation data, *EOS, Transactions, AGU*, 89(10), 93-94.

545 Le Vine, N., A. Butler, N. McIntyre, and C. Jackson (2016), Diagnosing hydrological
546 limitations of a land surface model: application of JULES to a deep-groundwater chalk basin,
547 *Hydrol. Earth Syst. Sc.*, 20, 143-159.

548 Lee, L. J. E., D. S. L. Lawrence, and M. Price (2006), Analysis of water-level response to
549 rainfall and implications for recharge pathways in the chalk aquifer, SE England, *J. Hydrol.*,
550 330, 604-620.

551 Ly, S., C. Charles, and A. Degré (2013), Different methods for spatial interpolation of rainfall
552 data for operational hydrology and hydrological modeling at watershed scale. A review,
553 *Biotechnol. Agron. Soc. Environ.* 17, 392-406.

554 Mathias, S. A., A. P. Butler, B. M. Jackson, and H. S. Wheater (2006), Transient simulations
555 of flow and transport in the chalk unsaturated zone, *J. Hydrol.*, 330, 10-28.

556 Met Office (2006), UK hourly rainfall data, Part of the Met Office Integrated Data Archive
557 System (MIDAS), NCAS British Atmospheric Data Centre, 21 March 2016,
558 <http://catalogue.ceda.ac.uk/uuid/bbd6916225e7475514e17fdbf11141c1>.

559 Morton, D., C. Rowland, C. wood, L. Meek, C. Marston, G. Smith, R. Wadsworth, and I. C.
560 Simpson (2011), Final report for LCM2007 – the new UK Land Cover Map (CS technical
561 report no 11/07), Centre for Ecology and Hydrology, UK.

562 Mu, Q., F. A. Heinsch, M. Zhao, and S. W. Running (2007), Development of a global
563 evapotranspiration algorithm based on MODIS and global meteorology data, *Remote Sens.*
564 *Environ.*, 111, 519-536.

565 Price, A., R. A. Downing, and W.M. Edmunds (1993), The chalk as an aquifer. In: Downing,
566 R. A., M. Price, and G. P. Jones *The hydrogeology of the chalk of north-west Europe*. Oxford:
567 Claredon Press. 35-58.

568 Price, M., R. G. Low, and C. McCann (2000), Mechanisms of water storage and flow in the
569 unsaturated zone of chalk aquifer, *J. Hydrol.*, 54-71.

570 Rawls, W. J., D. L. Brankensiek, and K. E. Saxton (1982), Estimation of soil water
571 properties, *Trans. ASAE*, 25(5), 1316–1320.

572 Robinson, E. L., E. Blyth, D. B. Clark, J. Finch, and A. C. Rudd (2015), Climate hydrology
573 and ecology research support system potential evapotranspiration dataset for Great Britain
574 (1961- 2012) [CHESS-PE], NERC-Environmental Information Data Centre.

575 Schär C., D. Lüthi, U. Beyerle, and E. Heise (1999), A soil precipitation feedback: A process
576 study with a regional climate model, *J. Clim.*, 12, 722–741.

577 Schaap, M. G., and F. J. Leij (1998), Database-related accuracy and uncertainty of
578 pedotransfer functions, *Soil Sci.*, 163(10), 765–779.

579 Sorensen, J. P. R., J. W. Finch, A. M. Ireson, and C. R. Jackson (2014), Comparison of varied
580 complexity models simulating recharge at the field scale, *Hydrol. Process.*, 28, 2091-2102.

581 Van den Daele, G. F. A., J. A. Barker, L. D. Connell, T. C. Atkinson, W. G. Darling, and J.
582 D. Cooper (2007), *J. Hydrol.*, 342, 157-172.

583 Van Genuchten, M. Th. (1980), A closed-form equation for predicting the hydraulic
584 conductivity of unsaturated soils, *Soil Sci. Soc. Am. J.*, 44, 892–898.

585 Walters, D. N., K. D. Williams, I. A. Boutle, A. C. Bushell, J. M. Edwards, P. R. Field, A. P.
586 Lock, C. J. Morcrette, R. A. Stratton, J. M. Wilkinson, M. R. Willett, N. Bellouin, A. Bodas-
587 Salcedo, M. E. Brooks, D. Copsey, P. D. Earnshaw, S. C. Hardiman, C. M. Harris, R. C.
588 Levine, C. MacLachlan, J. C. Manners, G. M. Martin, S. F. Milton, M. D. Palmer, M. J.
589 Roberts, J. M. Rodríguez, W. J. Tennant, and P. L. Vidale (2014), The Met Office unified
590 model global atmosphere 4.0 and JULES global land 4.0 configurations, *Geosci. Model Dev.*,
591 7, 361-386.

592 Wellings, S. R., and J. P. Bell (1980), Movement of water and nitrate in the unsaturated zone
593 of upper chalk near Winchester, Hants., England, *J. Hydrol*, 48, 119-136.

594 Williams, K., and D. Clark (2014), Disaggregation of daily data in JULES (Hadley Centre
595 technical note 96), Hadley Centre, Met Office, UK.

596 Zehe, E. and G. Blöschl (2004), Predictability of hydrologic response at the plot and
597 catchment scales: Role of initial conditions, *Water Resour. Res.*, 40.

598 Zehe, E., T. Maurer, J. Ihringer, and E. Plate (2001), Modeling water flow and mass transport
599 in a loess catchment, *Phys. Chem. Earth (B)*, 26, 487-507.

600 Zehe, E., U. Ehret, T. Blume, A. Kleidon, U. Scherer, and M. Westhoff (2013), A
601 thermodynamic approach to link self-organization, preferential flow and rainfall-runoff
602 behaviour, *Hydrol. Earth Syst. Sc.*, 17, 4297-4322.

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605 **Tables**

606 Table 1. Field measurements and remote sensing data.

Data	Spatial scale	Temporal extent	Frequency	Source
Soil moisture	Point ^a	2003-2005	15 day	N. Hewitt (CEH)
Latent heat flux	Global	2006-2011	8 day, 1 month	MODIS
Discharge	Point ^b	2006-2011	1 day	NRFA

607 ^aMeasured at Warren Farm.

608 ^bLocations are shown in Figure 1a.

609 Table 2. Hydraulic properties for different soil types (refer to Figure 1c). Saturated hydraulic
 610 conductivity (K_s) and porosity data are obtained from *Rawls et al.* [1982]. The Van Genuchten
 611 parameters are acquired from *Schaap and Leij* [1998].

Texture	K_s (mmd ⁻¹)	Porosity (-)	α (m ⁻¹)	n (-)
Loam	320	0.463	3.33	1.56
Silt loam	172	0.50	1.2	1.39
Clay	15	0.475	2.12	1.2

612

613 Table 3. Hydraulic properties of chalk

Properties	Uncalibrated		Range for calibration	Calibrated value
	Value	Source		
K_s (mmd ⁻¹)	1.0	Price et al., 1993	0.2 - 5.0	0.31
S_0 (-)	0.8	Observations	0.0 - 1.0	0.46
f_m (-)	10^5	Price et al., 1993	$10^4 - 10^6$	10^{5*}
α (m ⁻¹)	3.0	Le Vine et al., 2016	-	-
n (-)	1.4	Le Vine et al., 2016	-	-

614 * f_m parameter not calibrated

615 Table 4. Comparison between observed and simulated daily average runoff from the two
 616 configurations over the Kennet catchment. Metrics include the Runoff Ratio (RR), relative bias ($\Delta\mu$),
 617 and relative difference in standard deviation ($\Delta\sigma$) (refer to appendix for further information).

Metric	Observed	Simulated (<i>default</i>)	Simulated (<i>macro</i>)
RR	0.40	0.82	0.46
$\Delta\mu$	-	1.04	0.12
$\Delta\sigma$	-	2.04	0.65

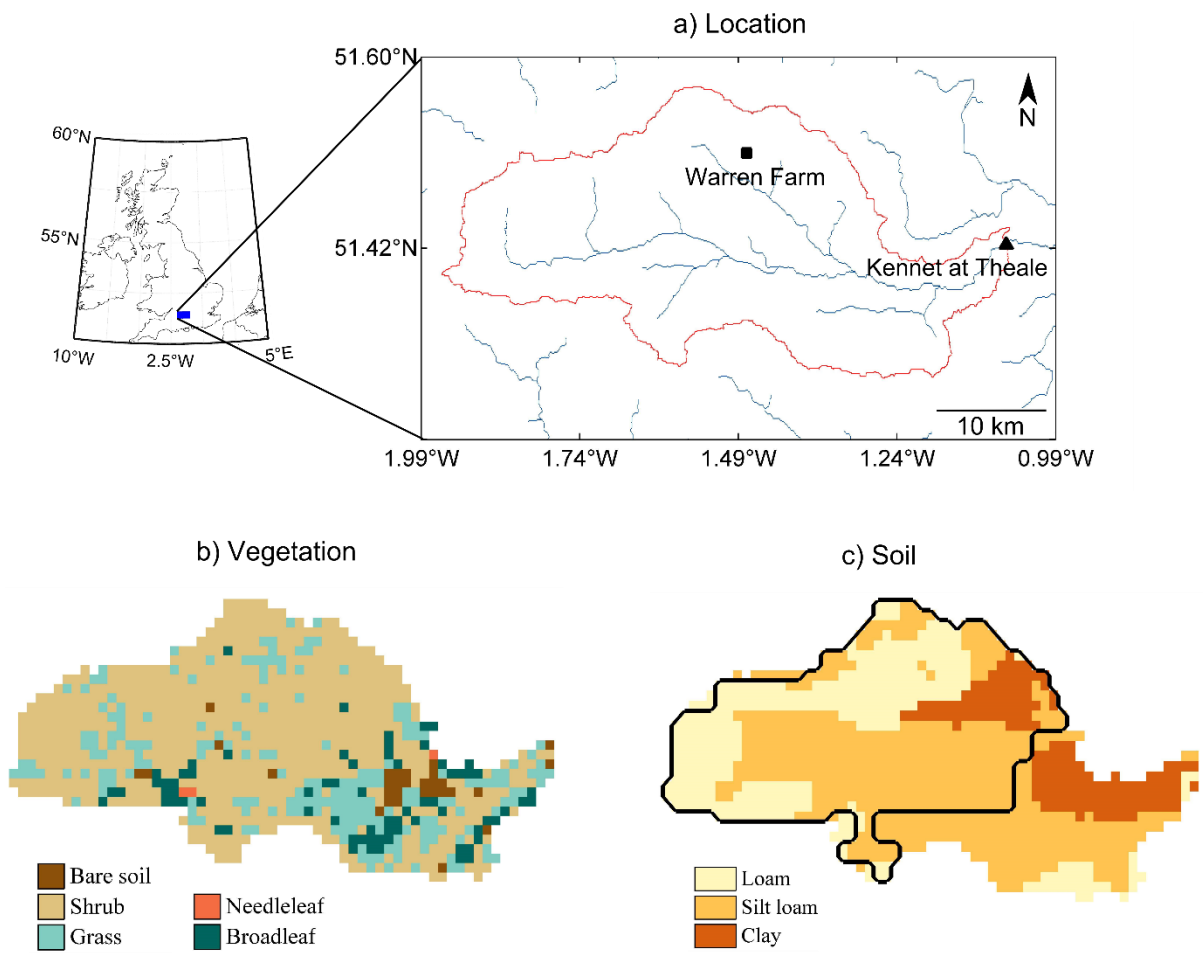
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620 **Figures**

621 Figure 1. (a) Location, (b) vegetation cover and (c) soil texture over the study area. The red
622 line in (a) outlines the Kennet catchment boundary, while the river network is shown in blue.
623 The black triangle in (a) shows the location of the discharge gauging station at the catchment
624 outlet and the black square corresponds to Warren Farm location where point-scale
625 simulations are carried out. The black line in (c) encloses the area of the catchment where
626 chalk is present.

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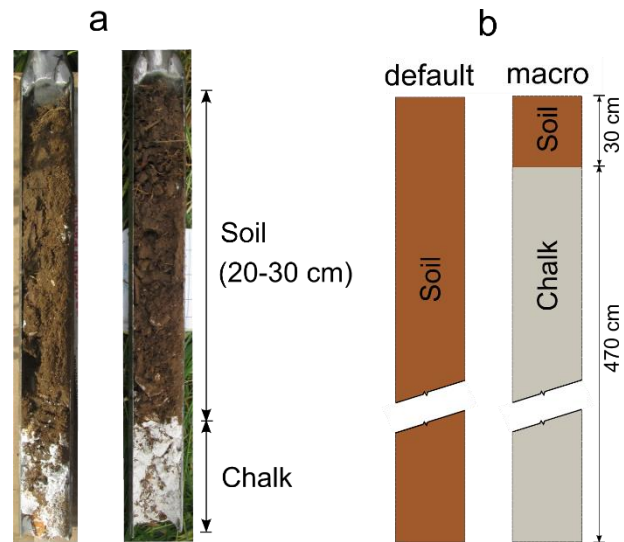
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631 Figure 2. (a) Example of soil profiles collected at Warren Farm during a field campaign in
632 2015 and (b) the two model configurations.

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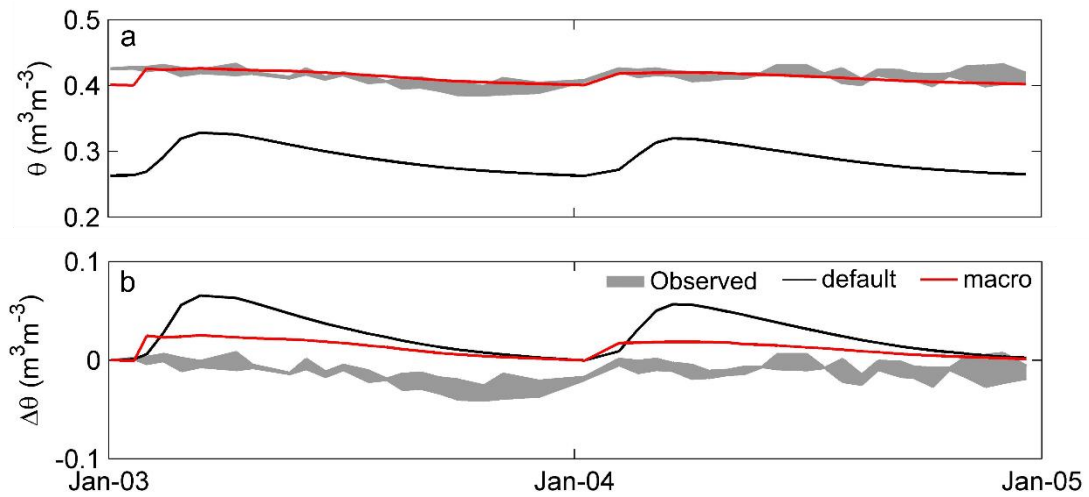
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646 Figure 3. Comparison between observed and simulated (a) soil moisture (θ) and (b) change in
647 soil moisture ($\Delta\theta$) from the *default* and *macro* configurations at a depth of 2 m below land
648 surface at the Warren Farm site. The shaded areas constructed from 2 soil moisture probes at
649 the Warren Farm site denote the range of observed data in these plots.

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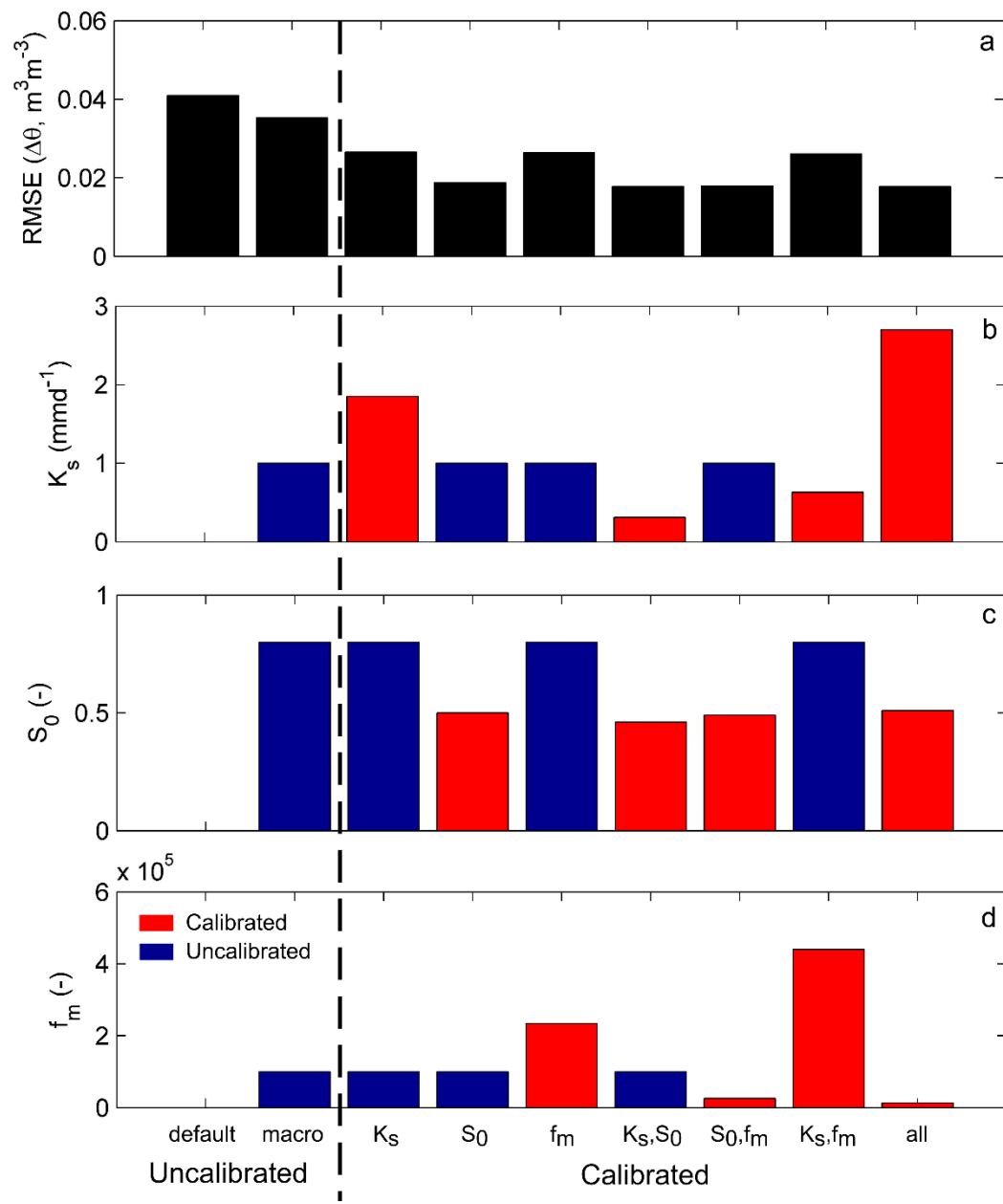
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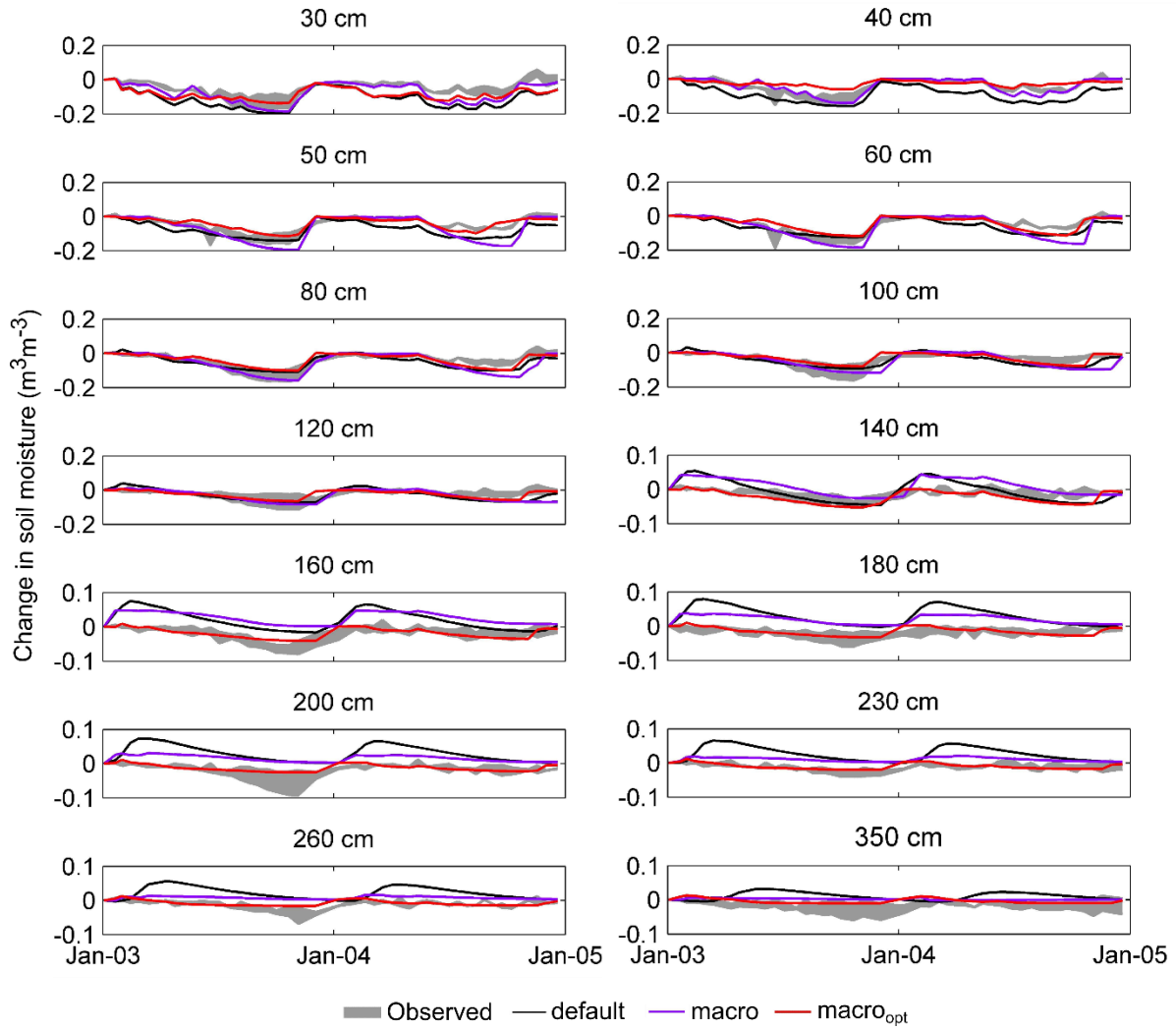
661 Figure 4. (a) Model performance in reproducing observed and simulated $\Delta\theta$, (b) K_s , (c) S_0 and
 662 (d) f_m for different parameter combinations considered in the optimization. For each
 663 parameter (i.e., panels b, c, and d), red bars show cases in which the relevant parameter is
 664 calibrated (either individually or in combination with others); while the blue bars correspond
 665 to cases in which the selected parameter is not calibrated (i.e., fixed value according to
 666 literature as in the *macro* case). Note that except for the *default* and *macro*, the simulation
 667 yielding the lowest RMSE (out of 2,000 model runs) is presented in this plot.



668

669 Figure 5. Comparison between observed and simulated $\Delta\theta$ from *default*, *macro* and *macro_{opt}*
 670 configurations at various depths below land surface. The shaded areas, which are constructed
 671 from 2 soil moisture probes at the Warren Farm site, denote the range of $\Delta\theta$.

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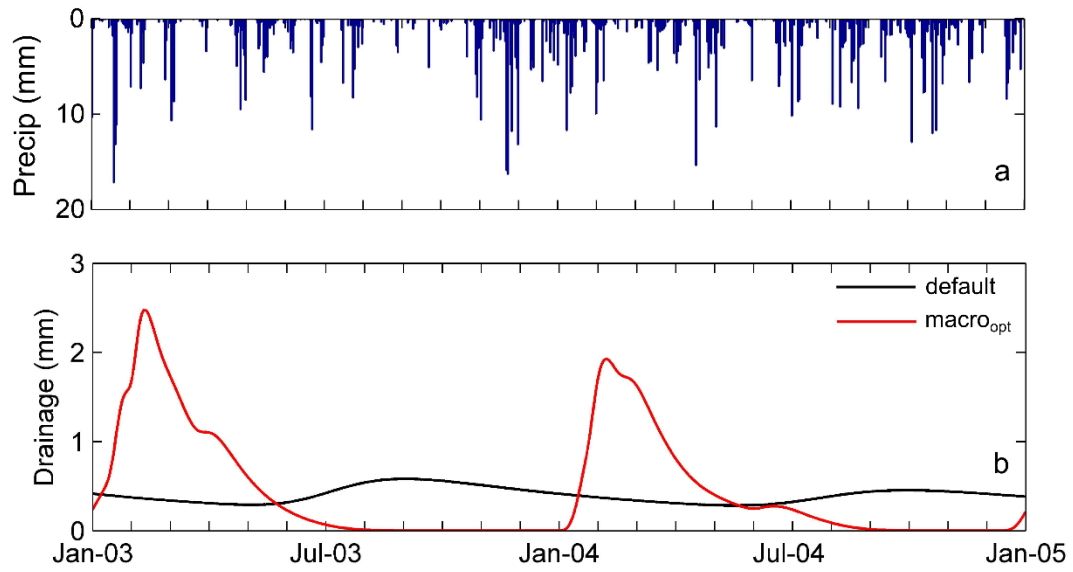
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679 Figure 6. (a) Daily precipitation and (b) daily drainage through the bottom of the soil column
680 at Warren Farm over the two simulated years (2003-2005).

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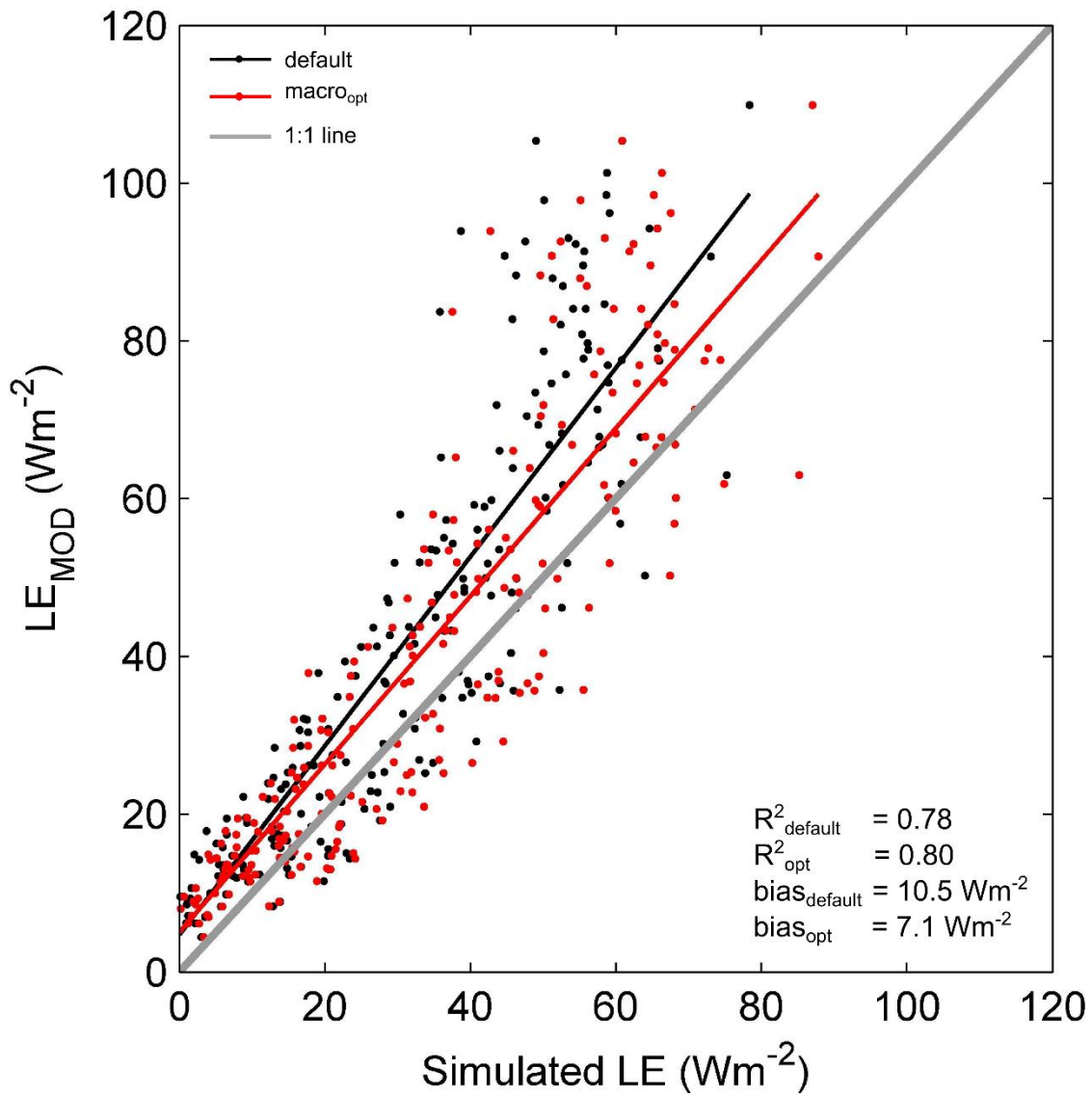
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693 Figure 7. Catchment average 8 day composites of MODIS estimated LE (LE_{MOD}) against
694 simulated LE from *default* and *macro* configurations ($LE_{default}$ and LE_{macro} , respectively) along
695 with the linear models fitted for $LE_{default}$ (black line) and LE_{macro} (red line). The 1:1 line is
696 shown in grey, which represents the perfect fit between LE_{MOD} and simulated LE .

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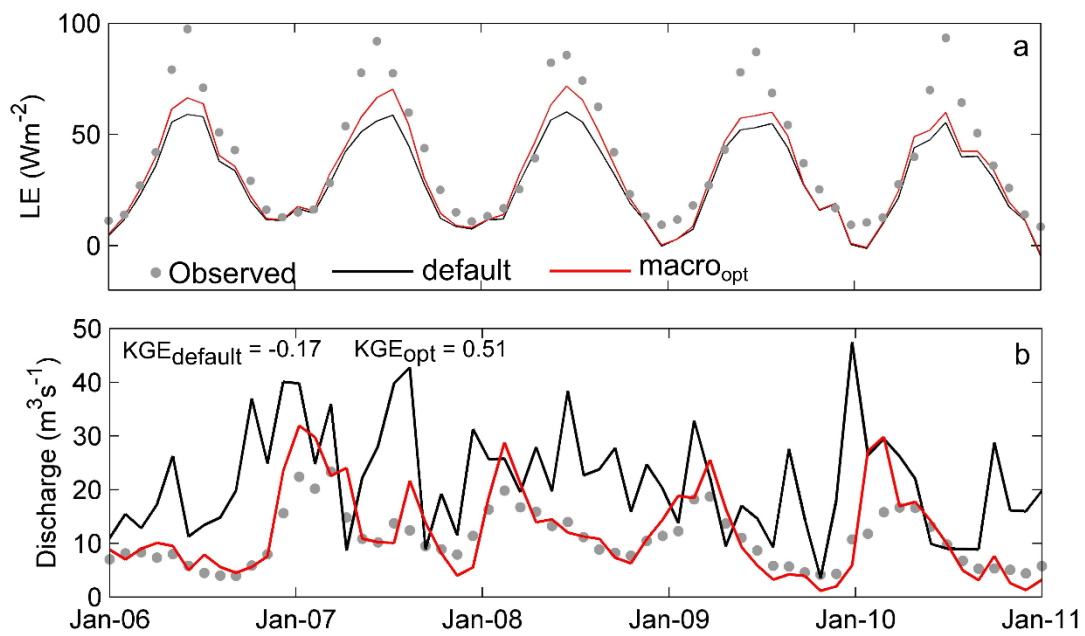
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701 Figure 8. (a) Spatially averaged monthly latent heat flux (LE) from MODIS, *default* and
702 *macro_{opt}* configurations over the Kennet catchment and (b) monthly average observed and
703 simulated discharge from the *default* and *macro_{opt}* configurations at the “Kennet at Theale”
704 gauging station.

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