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GIBSI: an integrated modelling system for watershed management – sample applications and current developments

R. Quilbé and A. N. Rousseau

Institut National de la Recherche Scientifique – Centre Eau, Terre et Environnement (INRS-ETE), Université du Québec, 490 rue de la Couronne, Québec (QC), G1K 9A9, Canada

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Correspondence to: A. N. Rousseau (alain.rousseau@ete.inrs.ca)

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Abstract

Hydrological and pollutant fate models have long been developed for research purposes. Today, they find an application in integrated watershed management, as decision support systems (DSS). GIBSI is such a DSS designed to assist stakeholders in watershed management. It includes a watershed database coupled to a GIS and accessible through a user-friendly interface, as well as modelling tools that simulate, on a daily time step, hydrological processes, soil erosion, agricultural pollutant transport and surface water quality. Therefore, GIBSI can be used to assess a priori the effect of management scenarios (reservoirs, land use, waste water effluents, diffuse sources of pollution that is agricultural pollution) on surface hydrology and water quality. For illustration purposes, this paper presents several management-oriented applications using GIBSI on the 6680 km² Chaudière River watershed, located near Quebec City (Canada). They include impact assessments of: (i) timber harvesting; (ii) municipal clean water program; (iii) agricultural nutrient management scenarios; (iv) past land use evolution; (v) possible future agricultural land use evolution under climate change, as well as (vi) determination of achievable performance standards of pesticides management practices. Current and future developments of GIBSI are also presented as these will extend current uses of this tool and make it useable and applicable by stakeholders on other watersheds. Finally, the conclusion emphasizes some of the challenges that remain for a better use of DSS in integrated watershed management.

1 Introduction

Integrated water management at the watershed scale has become a priority in many countries all over the world. For instance, the U.S. Clean Water Act (see Clements et al., 1996; Gariépy et al., 2006) and the European Directive Framework (Official Journal of the European Community, 2000) impose specific objectives in terms of water resources integrity and have led to the creation of numerous watershed organisations

HESSD

4, 1301–1335, 2007

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

that try to apply integrated management principles. These principles basically consist in conciliating all land and water uses while protecting and sustaining water resources. This relies on the involvement of stakeholders and the definition of operating rules. To facilitate the decision making process, there is a need for high quality and accessible data. In this context, scientific research has shifted towards the development of decision support systems (DSS) designed to support the implementation of integrated water management. DSS are practical and user-friendly computer tools that basically rely on a geographical information system (GIS) and a relational database management system (RDBMS) that enable the display of information at any point in space and time, as well as transforming raw data into information relevant to the decision making process (graphs, maps or tables). Most DSS are based on mathematical models and can be used to assess a priori the effect of watershed management scenarios regarding urban, forestry or agricultural issues, on water yield and quality at the watershed scale. Many DSS have been developed all over the world, suited for specific conditions, scales and purposes. Borah and Bera (2004) as well as Rousseau et al. (2005), to name a few, reviewed and compared some DSS. For instance, while some of them are more suited for urban water management, others are specifically developed to assess the effect of agricultural practices on hydrology. Even if most of them are still under development and regularly upgraded, their development is now advanced enough to envision concrete applications for water management purpose (e.g. see Borah and Bera, 2004; He, 2003; Rousseau et al., 2005; Santhi et al., 2001).

As an illustration of the structure, development and possible applications of a DSS, this paper focuses on GIBSI (Gestion Intégrée des Bassins Versants à l'aide d'un Système Informatisé), a Canadian DSS that was originally developed to assist scientists and stakeholders in integrated water management. The development of GIBSI began in 1995 and a first version was released in 1998 (Villeneuve et al., 1998a). A general description is presented in Sect. 2. Since then, several application studies have been conducted and some of them are briefly described in Sect. 3. At the same time, this integrated modelling system (IMS) has been continually upgraded by addition

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

of new models, data post-processing modules or model calibration guidelines. This is presented in Sect. 4.

2 General description of GIBSI

GIBSI is designed to help stakeholders to make decisions in water management at the watershed scale. It can either be used as a data management system or as an impact assessment tool to study the effect of management scenarios on surface water quality using mathematical models. The general structure of GIBSI is depicted on Fig. 1. A detailed presentation of each component may be found in Villeneuve et al. (1998a).

2.1 Data management modules

As most DSS, GIBSI is basically composed of a database, a GIS, a RDBMS and a graphical user interface or GUI (Fig. 2). Attribute data were originally managed using Microsoft Access™ (Simpson, 1994) but this RDBMS has now been replaced by the MySQL® database management system (Pedersen et al., 2005). The database contains spatial data (e.g. location of meteorological stations) and attribute data (i.e., all data associated with spatial data, such as meteorological series). The GIS used is GRASSLAND (L.A.S., 1996). The watershed is discretized into two types of computational elements: (i) river segments that is, one-dimensional elements that support watercourse simulation processes (i.e. streamflow and pollutant transport); and (ii) relatively homogeneous hydrological units (RHHUs) that is, elements corresponding to elementary watersheds that support all the other simulation processes (i.e. runoff generation and sediment and pollutant transport). These computational elements are determined using PHYSITEL (Turcotte et al., 2001), a complementary software program designed specifically to prepare the physiographic database of distributed hydrological models.

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

2.2 Scenarios management modules

Four types of management scenarios can be defined:

- (i) Reservoirs: addition of new reservoirs at any river segment or editing of their characteristics;
- (ii) Agriculture : editing of crop types, nutrient management practices, pesticide treatments, dates of agricultural practices, at any spatial scale (i.e., one or many RHHUs, subwatersheds or administrative units such as municipalities);
- (iii) Wastewater treatment plants: addition of new plants at any river segment or editing of their characteristics (e.g., treatment types, effluent rates);
- (iv) Land use: change of a land use class to another one at any spatial scale (Fig. 2).

Once scenarios are defined, there are integrated into the database and simulations can be run.

2.3 Simulation modules

GIBSI simulates hydrology, erosion, pollutant transport and surface water quality. It is based on four existing semi-distributed models: (i) HYDROTEL (Fortin et al., 2001a; Fortin et al., 1995), a physically-based, hydrological model compatible with GIS and remote sensing; (ii) RUSLE (Renard et al., 1997; Wischmeier and Smith, 1978) complemented by Yalin's equation (Yalin, 1963) to account for soil erosion and sediment transport capacity; (iii) SWAT (Arnold et al., 1996) and EPIC algorithms (Arnold and Williams, 1995) to simulate the fate of nitrogen, phosphorus and pesticides on cropland; (iv) QUAL2E (Brown and Barnwell, 1987) a water quality model that simulates the biological, physical and chemical processes controlling the fate of pollutants in surface water. The modeling time step is the day. Input data are daily meteorological series (precipitations, minimum temperature and maximum temperature). The user selects

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

the meteorological series, either by defining dates of start and end into the meteorological database, or by selecting characteristic series (e.g., dry summer or wet spring).

2.4 Post-processing tools

Tools have been developed to analyse simulation results. First, data can be visualized using tables, graphs or maps. Regarding pollutant concentrations data, the frequency of exceeding a water quality standard can be calculated or visualised using graphs, which is useful when simulations concern the effect of management practices on specific water uses such as swimming or drinking water. Finally, an environmental benefit-cost analysis can be performed for agricultural scenarios based on the cost of implementation of agricultural beneficial management practices and the environmental benefits from recovering potential water uses due to water quality improvement (see Salvano et al., 2004).

3 Applications of GIBSI

Several applications of GIBSI have been performed since the first version was released in 1998. The aim of this paper is not to describe all of them in detail but rather to provide an overview that illustrates the range of possibilities and limitations of use. All of these applications have been performed on the Chaudière River watershed or sub-watersheds.

3.1 The Chaudière River watershed

The Chaudière River is a tributary of the Saint-Lawrence River, located south of Québec City (Fig. 3). It drains a watershed of 6680 km², mainly forested (64%) and used for agriculture (33%). This watershed was selected for the application and development of GIBSI due to the variety of land uses, agricultural and industrial activities, available data and because it is representative of many watersheds in the Saint-

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Lawrence valley. Soils vary from loamy sand in the downstream portion of the watershed to clay loam in the middle portion and loam in the upstream portion (Duchemin et al., 2001). Agriculture is dominated by animal production, especially pig and dairy farming. That implies that most of the cropland is dedicated to forages and pasture (75% of agricultural land in 1995). It is also important to note that it was identified by the Québec government as the pilot watershed for the implementation of integrated watershed water management in 1993. The population is now around 180 000 inhabitants. Meteorological data come from forty stations distributed all over the watershed.

Calibration of the hydrological model HYDROTEL was performed on the whole watershed (Fortin et al., 2001b) considering measured and simulated streamflows at the outlet. The model efficiency was satisfactory with Nash-Sutcliffe coefficients of 0.88 and 0.83 for 1989–1990 and 1993–1994, respectively. Spatial validation was performed for the Famine and Beaurivage subwatersheds with similar results. A temporal validation was also performed on 1987–1988 and 1990–1991 as well as over a 10-year period. A first calibration of the erosion model was performed in 2002 (unpublished). Regarding the nutrient transport and water quality model, Villeneuve et al. (1998b) compared simulated nitrate and phosphorus concentrations to measured data and found that errors were, in majority, less than 0.1 mgN/L and 0.01 mgP/L respectively. Results were less satisfactory for BOD5. Note that improvements and further calibration of these models are in progress (see Sect. 3.1).

3.2 Effect of timber harvesting on watershed hydrology

3.2.1 Context, objectives and general approach

The objective of this study (Lavigne et al., 2004) was to characterize the ability of GIBSI to simulate the impact of deforestation on the hydrological regime of the Famine River watershed, a subwatershed of the Chaudière River covering an area of 728 km². It is mostly forested, 30% in deciduous forest and 41% in coniferous forest.

3.2.2 Scenarios and simulations

Two scenarios were considered and compared: the basic scenario (BS) consisted in the 1995 land use, while the deforestation scenario (DS) involved a drastical change of forest into bare land. By this way, the bare soil class increased from 1% to 72% of the total area of the watershed. Thirty-one simulations were performed with meteorological daily time series from 1964 to 1995 as input data, each year being considered independently. This was done for each scenario.

3.2.3 Results

The results showed an increase in annual runoff (+57%), spring runoff (+25%) and summer runoff (+138%) after deforestation. For summer low-flow periods, GIBSI simulated an average increase in runoff of 102%. The spring peak flow was also affected, the median value shifting from 201 to 259 m³ s⁻¹ (+26%), as well as the summer peak flow (increase of 74 m³ s⁻¹, +101%). As an illustration, Fig. 4 shows the increase in the summer runoff using the 1995 meteorological series. The runoff response to these precipitation events for DS scenario is larger than for BS scenario, especially between 11 July and 30 August. For the precipitation event of 22 July, runoff response simulated by GIBSI occurred 2 days after the precipitation event (65 mm) with a runoff increase of 135 mm between the two scenarios. These results confirm the fact that deforestation has a great effect on runoff during summer. These changes were consistent with observations from experimental paired watersheds studies found in literature, showing that GIBSI could complement field studies for land management planning on large scale watersheds with respect to surface water resources.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

3.3 Impact of municipal clean water program

3.3.1 Context, objectives and general approach

In 1978, a provincial municipal clean water program (MCWP) was implemented in Québec to control point-source pollution and restore the province's surface water quality. On the Chaudière River watershed, 35 waste water treatment plants (WWTP) were built and the population being connected reached 95% in 1997. The objective of this study (Mailhot et al., 2002) was to assess the effect of this program on water quality. The first step was to characterize WWTP properties (i.e. affluent and effluent water discharge, chemical and physical parameters). Pollutant loads from industrial plants not connected to a municipal sewer network as well as diffuse sources of pollution from urban area or agricultural land were not considered in this study.

3.3.2 Scenarios and simulations

Two scenarios were defined in GIBSI: the first one was associated with the 1983 year corresponding to the pre-MCWP period, and the second one was associated with the 1994 year describing the post-MCWP period. Simulations were performed with both scenarios using meteorological data of years 1983 (dry year) and 1994 (wet year) – two simulations for each scenario.

3.3.3 Results

Simulations results showed that the annual probability of exceeding phosphorus water quality standards (WQSs) at the watershed outlet decreased from 0.53 to 0.40 after MCWP, under 1994 meteorological conditions (wet year). That corresponds to a gain of 49 days under the phosphorus WQS. The results are similar under dry conditions (1983 series), even if phosphorus concentrations remain almost always larger than the WQS during summer, and at the other control points on the watershed. Regarding biological oxygen demand (BOD5), in the year before MCWP the WQS was usually exceeded

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

during several days, however, after MCWP the standard was never exceeded, neither under wet nor dry conditions. Finally, we observed a decrease in median concentration but the concentrations were always smaller than the WQS, even before MCWP. We can conclude from this study that MCWP had a drastic effect on BOD5 and phosphorus but that urban wastewater is still responsible for high probabilities of exceeding phosphorus WQS in the Chaudière River.

3.4 Determination of environmental load allocations from point and diffuse sources

3.4.1 Context, objectives and general approach

This study (Rousseau et al., 2002a, b) consisted in assessing the contribution of pollutant loads associated with wet (non-point source or diffuse) and dry (point source) weather to water pollution. The objective was also to set environmental load allocations (ELAs) or total maximum daily loads (TMDLs) regarding various human activities and water uses, and to determine whether WQSs were achievable. If ELAs from point sources are generally well established based on severe low flows and maximum daily loads, it is not as easy for diffuse sources since they occur during important runoff events and strongly depend on land use and agricultural management practices. Untreated municipal wastewater from a small town (St. Martin) and non-point source pollution were responsible for impairment of the studied river segment located upstream of Sartigan dam which is a place for recreational activities. Pollutants considered were fecal coliforms for bacteriological considerations and phosphorus for aesthetic requirements (eutrophication).

3.4.2 Scenarios and simulations

Based on WQSs, environmental load allocations were calculated for both point and non-point sources, resulting in 7.5 10¹¹ UFC/day and 4.35 kgP/day for low flow conditions. Two base case scenarios were defined considering only point source pollution or

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

both point-source and non-point source pollution. Then, three management scenarios were defined considering: (i) the construction of a WWTP at St. Martin; (ii) the reduction of fertilization rate (from 31 to 13 kgP/ha) by meeting crop requirements for only phosphorus; and (iii) both changes. The simulations were performed using four years (1982–1985) of meteorological data including wet and dry weather years.

3.4.3 Results

As shown on Table 1, the bacteriological WQS was always exceeded more than 84% of the time without the WWTP, but the WWTP reduced this frequency to 0%. Meanwhile, the reduction of the phosphorous fertilization rate resulted in an improvement of water quality as the average probability of exceeding the phosphorus WQS over the five year period dropped from 0.32 to 0.27. This probability was further lowered to 0.19 in the case of the implementation of the St. Martin WWTP. These simulation results also enabled to quantify the contribution of point versus non-point sources for phosphorus loads, which varied dramatically from wet to dry weather conditions: dry weather sources (i.e. urban wastewater) accounted for 63% of the phosphorus loads during the driest summer, and for 12% to 19% for the other summers. This study clearly illustrated how the proposed approach based on probability of exceeding WQs is well suited to characterise ELAs for various human activities.

3.5 Environmental benefit-cost analysis of manure management

3.5.1 Context, objectives and general approach

This study (Salvano et al., 2004, 2006) illustrates the importance of valuing environmental benefits associated with an improvement of water quality when assessing and implementing new agricultural nutrient management plans. In the province of Québec, the Québec Regulation Respecting Agricultural Operations (RRAO) was implemented in 2002 (and modified in 2005) to ensure the protection of water and soil, as well as

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

aquatic life and human health from agricultural pollution. The rationale behind this regulation is to reach a balance between the soil's phosphorus support capacity and the amount of fertilizers. The objective of this study was to evaluate potential benefits generated from water quality improvements within a benefit/cost analysis framework. As most of human services provided by water are not priced in markets, their economic values first have to be estimated using non-market methods. In this case, a benefit transfer procedure was used (see Salvano et al., 2004, 2006 for details). As a first approximation, only the monetary benefits associated with water-based recreational activities were evaluated. First, willingness-to-pay and participation data (number of persons and days per person) were determined based on a survey from Environment Canada (Environment Canada, 2001). Then, the number of days exceeding the phosphorus aesthetic WQS obtained from simulations were used to calculate the total benefit of the management scenario.

3.5.2 Scenarios and simulations

This application concerned the Beaurivage River watershed (tributary of the Chaudière River) on which numerous river segments were identified for existing or potential recreational activities (swimming, canoeing, kayaking, hiking). Two scenarios were defined: (i) a base case scenario assuming application of all available manure; and (ii) an on-farm nutrient management scenario based on meeting phosphorus crop requirements with manure and treating any manure surpluses (so called RROA scenario). The cost associated with the implementation of this manure management practice was calculated based on livestock production costs and revenues, manure storage and treatment costs, and fertilization costs for a spatial unit, as compared to the base case scenario. A reduction of benefit is considered as an opportunity cost that is added to other costs. Total cost was then obtained by multiplying the unit cost by cropland area within management units of interest. Two types of spatial management units were considered to evaluate and compare the effect of these scenarios: a group of three contiguous municipalities and two subwatersheds corresponding to two river segments

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

of the Beaurivage River watershed with existing or potential water uses. As the prime period for recreational activities is the summer, simulations were performed using summer meteorological data for years 1977 through 1986 independently.

3.5.3 Results

5 Simulations with the management scenario induced an increase of the number of days of potential activities (i.e. for which phosphorus concentration is lower than phosphorus aesthetic WQS) as compared to the base-case scenario, for the two river segments and the three management units (between 248 and 499 days vs. 213 days for base case scenario). The results of the environmental benefit-cost analysis are given in Table 2.

10 We can see that benefits were similar for all management units but that costs were higher when implementing the scenario at the scale of the municipalities. This results in higher benefit-cost ratios, especially for the upstream river segment, due to the fact that it focuses on a problematic area. Note that all benefit-cost ratios were smaller than one, but a sensitivity analysis showed that this ratio could be larger than one when reducing manure treatment costs, which is technically achievable. However, this methodology does not take into account intrinsic values that would certainly increase the ratios. The development of the proposed environmental benefit-cost analysis methodology represents a first step towards the estimation of all monetary benefits of improving water quality.

20 3.6 Influence of past land use evolution on hydrological regime

3.6.1 Context, objectives and general approach

As illustrated in Sect. 3.2, watershed hydrology is highly sensitive to land use. For instance, clear-cutting induces an increase in water discharge and peak flows. In general, land use all over the world has evolved a lot over the last decades, and it is
25 important to understand and quantify the influence of these changes on hydrology in

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

the past to be able to anticipate the future. This is the objective of this study based on the application of GIBSI in the Chaudière River watershed (Savary et al., 2007¹).

3.6.2 Scenarios and simulations

Seven satellite images (Landsat) were acquired to reconfigure the evolution of land use over the last 30 years (1976, 1981, 1987, 1990, 1995, 1999, 2004). After treatment and classification, the evolution of land use was quantified (Fig. 5). Agriculture and forest (including bush land) followed opposite tendencies, with an increase in agricultural land in the seventies, stagnation in the eighties and a decrease in the nineties. The seven images were integrated into the GIBSI database. For each land use configuration, simulations were run using 30 years of meteorological data (1970–1999), each year being considered independently.

3.6.3 Results

The mean annual water discharge, as well as critical low flow sequences (Q2-7, Q10-7 and Q5-30), were strongly correlated with agricultural land use evolution, with determination coefficients of 0.97, 0.95, 0.92 and 0.93, respectively. This can be explained by the fact that the increase in agricultural land to the detriment of forest induces less evapotranspiration, and thus more available water for overall runoff. This is taken into account by the hydrological model of GIBSI by the mean of the leaf area index and the root depth in the water balance calculation. Moreover, soil surface is more likely to produce faster runoff (translated in the model by a lower Manning coefficient in the calculation of surface water routing). These results show that the water regime of a river is strongly dependent on the evolution of the watershed land use and illustrate how a DSS like GIBSI enables to quantify this influence.

¹Savary, S., Garbouj, M. S., Rousseau, A. N., and Quilbé, R.: Impact of land use changes on the hydrology of the Chaudière River watershed, in preparation, 2007.

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

3.7 Influence of future land use evolution on hydrological regime and water quality in a context of climate change

3.7.1 Context, objectives and general approach

To complement the previous study, Quilbé et al. (2007) have assessed the effect of future land use evolution scenarios under climate change conditions. To do so, the first step was to determine future meteorological sequences that would be used as input data in GIBSI. Three methods were used, based on general circulation models (GCM): delta method, statistical downscaling and a mix of both. Three GCMs were considered for the delta method and only one for the statistical downscaling method. Several gas emission scenarios (GES) were also considered for each case.

3.7.2 Scenarios and simulations

Three land use scenarios were defined. First, the land use of 1995 was used as reference. Secondly, scenario A was driven by economical criteria, by extrapolating the last decade tendency regarding pig production increase. This results in deforestation to create more agricultural fields for feed production and manure application. Finally, scenario B considers the land use distribution as it was in 1976, implying reforestation. It also considers a spatial dispersion of agricultural lands over the watershed. Simulations were run with each scenario over 30 years considered as independent, for reference period (1970–1999) and future period (2010–2039).

3.7.3 Results

The results first show that, without any modification in land use (base case scenario), climate change would induce a decrease of annual water discharge at the outlet of the watershed (mean of -2.7%, with delta method). However, this value can be very different from a GCM/GES combination to another (from -14.1% to +13.8%), pointing out the large uncertainties that are still linked to these methods. Water discharge would

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

strongly increase in winter due to higher temperatures and earlier snow melt (on average +68.5%), and decrease the rest of the year (Fig. 6). This effect is homogeneous within the thirty years of simulation, i.e. both for dry and wet years. Regarding the effect on spring peak flow, results are once again different from one GCM/GES combination to another with, on average, a decrease in peak flow. Finally it is not possible to conclude to any tendency regarding summer low flow modification. When integrating land use evolution scenarios into GIBSI simulations, we obtained opposite effects between the two scenarios, with an increase and a decrease of water discharge in summer and fall, respectively. This is illustrated on Fig. 6 which shows the mean monthly water discharge obtained with GCM ECHAM4, GES A2 and the three land use scenarios (base case, A and B). These results suggest that intensification of agriculture (Scenario A) would mitigate the effect of climate change in summer and fall by generating more runoff and thus higher streamflow in river. The results regarding the effects on sediment follow roughly the same trend as water discharge (see Quilbé et al., 2007). However, this prospective approach has many limitations due to the methodology. First, only tendencies can be pointed out because of the large uncertainty linked to the methodological approach and tools used. Moreover, even if statistical downscaling captures precipitation occurrence, it is weaker at predicting wet days amounts and extremes in our region (Gachon et al., 2005). Finally, the effect on water quality should be considered before concluding about the good or bad effect of land use scenario.

3.8 Definition of achievable agroenvironmental performance standards for pesticides at the watershed scale

3.8.1 Context, objectives and general approach

The Canadian National Agri-Environmental Standards Initiative (NAESI) program aims to develop water quality standards at the watershed scale, in order to lead the development and application of beneficial management practices (BMPs) at the farm scale. This includes: (i) ideal performance standards (IPSSs) that are based on ecotoxico-

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

logical data and specify the desired level of environmental state needed to maintain ecosystem health, and (ii) achievable performance standards (APSs) which represent more realistic standards that could be achieved using recommended available processes, practices and technologies, including BMPs. This means that, to determine APSs, it is necessary to assess a priori the effect of several BMPs on water quality by using a mathematical model that simulates the fate of pollutants at the watershed scale. This approach was first applied for pesticides. After a review and multicriteria comparative analysis of existing pesticide fate models at the watershed scale, GIBSI was selected together with the SWAT and BASINS/HSPF models (Quilbé et al., 2006). Pesticide concentrations data as well as representative agricultural practices were identified with a survey in the Beaurivage River watershed. However, due to the size of the area, it was impossible to identify precisely these practices at the farm scale. Therefore, a large uncertainty remains about the locations, dates and rates of pesticide applications. This uncertainty represents an important limitation for the calibration of the pesticide models. To account for this uncertainty during the simulations, a stochastic process was introduced in GIBSI to determine for each RHHU: (i) the year of crop rotation which determines the crop and thus what kind of pesticides are applied; and (ii) the date of application within the period of time defined. Then, several simulations with different stochastic configurations for crops and application dates provide a range of pesticide concentrations in surface water.

3.8.2 Scenarios and simulations

This study is described by Rousseau et al. (2007). A base case scenario was defined based on current agricultural practices. We considered that pesticides applications were done between 1 and 15 June (16 days). The application rates were 0.65, 0.06 and 1.6 kg/ha for Atrazine (on corn), MCPB (on cereals) and Metolachlor (on corn), respectively. Then three BMP scenarios were created based on this base case scenario: (1) implementation of a 1-m filter strip all along the river network; (2) a 30% reduction in the pesticide application rate; and (3) a combination of both. Simulations were

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

performed over thirty years (1979–1999), each year being considered independently. Three different stochastic configurations for crops spatial distribution and application dates were considered for each year, resulting in 90 simulations for each scenario. The effect of BMPs was examined at the outlet of the Beaurivage River subwatershed, which is the most affected by pesticide pollution.

3.8.3 Results

A cumulative frequency curve (CFC) of pesticide concentration was determined for the month of June since any pesticide level in the stream network beyond this month remains undetectable. These CFCs were based on thirty-year simulations and three stochastic configurations for each scenario. We proposed to define the 9th centile (Q90) of this distribution as a possible value for APS. This means that this value is exceeded only 10% of the time. The results show that this value decreases with the implementation of BMPs (Table 3). For Atrazine, we observed a reduction of 81.7%, 85.8% and 91.9% for scenarios 1, 2 and 3. For MCPB, the effect is weaker with a decrease of 41.4%, 47.5% and 74.7% respectively. Finally, for Metolachlor, Q90 was reduced of 57.6%, 68.6% and 100%, respectively. It is noteworthy that all these concentrations values are all lower than the WQS for aquatic life protection (1.8 $\mu\text{g/L}$, 7.3 $\mu\text{g/L}$ and 8 $\mu\text{g/L}$ respectively for Atrazine, MCPB and Metolachlor; MDDEP, 2006). These results mean that the effect of BMPs on pesticides concentrations may be very different from one pesticide to another. It should be noted that the final APS value can be linked to the ecotoxicological impact by the use of a species sensitivity distribution (the one used for the determination of IPS) as a means to assess a priori the percentage of potentially affected species. Finally, this approach can be easily transposed to other pesticides and other rivers and watersheds. The study clearly shows the utility of tools like GIBSI to determine a priori the efficiency of management plans at the watershed scale and to define APSs.

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

4 Current and future developments of GIBSI

In parallel to these application studies, GIBSI has been continuously upgraded since the first version released in 1998. This concerned database management, interface, scenario management, post-processing tools and models.

5 4.1 Model development and calibration

Regarding models, only the calibration of the hydrological model has been reported in the literature (Fortin et al., 2001b). This explains why most of the applications presented in this paper concern the effect of watershed management on water discharge. Regarding the pollutant transport models (erosion, nutrients, pesticides), an important work was performed to improve them over the last three years, and their thorough calibration is currently in the process and results will be submitted for publication soon. Moreover, a pathogen transport model has been developed (Rogel, 2007). It simulates the fate of fecal coliforms resulting from manure application on crops as well as pasture, accounting for bacterial mortality, partitioning and transport in erosion and runoff.

15 Finally, an indicator of ecological integrity is currently under development (Grenier et al., 2006b). Indeed, physical and chemical data are not sufficient to determine the quality of an aquatic medium. For instance, high nitrogen and phosphorus concentrations in a stream or a lake indicate a risk for eutrophication but do not give information on the real state of the aquatic environment, which depends on many other factors
20 such as pH or dissolved oxygen. This indicator will be based on benthic diatoms and macro-invertebrate communities. In a previous study (Grenier et al., 2006a), reference conditions of each stream type sampled in Southern Quebec (Canada) were defined using benthic diatoms and environmental variables characterizing streams and watersheds (regional reference site approach). Two diatom reference communities were
25 sufficient to define reference conditions, one for neutral conditions and one for alkaline conditions. Based on these results, classification tree models (prediction models) were created using reference sites to identify watershed characteristics responsible for

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

the discrimination between the two reference communities (neutral and alkaline). The model was then used to predict which diatom reference community should be present in an impacted stream under potential natural conditions. The results from this study were used to develop a diatom-based index that can be used to evaluate the degradation status of a site by comparing the actual ecological conditions with appropriate reference conditions. The same approach is now being applied for benthic macro-invertebrate communities. Finally, the developed models will be integrated into GIBSI to predict the structure of diatoms and macro-invertebrate communities in rivers based on simulated physical and chemical variables.

4.2 Development of an application protocol of GIBSI and implementation on other watersheds

Until now, GIBSI has been developed and applied on only one watershed, i.e. the Chaudière River. A user's guide already exists (Villeneuve et al., 2003) but it only explains how to use GIBSI once it is implemented. However, the implementation and application procedure of GIBSI, like many DSS, is still complex and follows several steps that have to be done carefully, for example: identification of user's needs, model selection, data acquisition, database construction, model adaptation if needed, model calibration, scenario definition, simulations and results analysis. Moreover, it should be always kept in mind that all models are based on simplification hypotheses that have to be considered in the end of the process, i.e. when interpreting the results and conclusions. Therefore, in order to allow stakeholders to apply it on other watersheds, an application protocol is being developed, explaining the different steps to follow and giving tools to clarify the procedure. It includes an inventory of needed data, a dictionary of the database and a guide on how to construct the database, discretize the watershed, apply and calibrate the models. Moreover, these two documents (user's guide and application protocol) as well as the GUI will eventually be translated in English and Spanish.

Meanwhile, applications on other watersheds have already begun: GIBSI is now be-

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

ing implemented on a Mexican watershed (Arcediano watershed, Santiago River) by the Instituto Mexicano de Tecnología del Agua (IMTA). The objective is to improve water management on this watershed, in collaboration with the watershed management committee. Moreover, in the next two years GIBSI will be implemented on four watersheds throughout Canada to determine APSs for pesticides (see Sect. 3.8): Yamaska river (QC), Wilmot-Dunk (PEI), South Nations (ON) and Salmon Arm (CB).

5 Conclusions

The practical applications of GIBSI presented in this paper illustrate the wide range of possibilities offered by such a tool to assess the effect of land use management, wastewater treatment or agricultural practices on water quantity and quality at the watershed scale. Current and future work involves addition and calibration of models as well as development of a water ecological integrity indicator. More generally, this demonstrates how a DSS like GIBSI may be used to incorporate sound science into legislative and political decisions regarding water management. However, the integration and the use of a DSS like GIBSI for operational integrated watershed management issues still face major challenges. One of them is that DSS are often too complex for operational purposes. Indeed, if stakeholders do need tools that are sound science based, they also need them to be easy to use, apply and understand. Thus, this requires a simplification as well as the development of tools and guides that facilitate the technical transfer from research to management, as it is being done for GIBSI. Also, new concepts based on user needs and receptivity have to be investigated (see Jake-
man and Letcher, 2003; McIntosh et al., 2007). Moreover, the uncertainty in model outputs has to be quantified to make the DSS more reliable and the decision making process easier (Mannina et al., 2006; Wu et al., 2006). Even so, it is clear that DSS application will always need a close dialogue between users and developers, i.e. between stakeholders and scientists, within an interdisciplinary framework, and this is certainly one of the greatest interests of such tools.

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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HESSD

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Table 1. Probability of exceeding bacteriological WQS ($WQS_{CF}=200$ CFU/100 mL) and aesthetic WQS ($WQS_P=0.03$ mg/L) during the summer period from 1982 to 1985 for the three scenarios : base case scenario (Sc A), fertilisation based on crop requirements (Sc B), fertilisation based on crop requirement and treatment of Saint-Martin wastewater (Sc C).

	P([CF]>WQS _{CF})					P([P]>WQS _P)				
	1982	1983	1984	1985	1982–1985	1982	1983	1984	1985	1982–1985
Sc A	1	1	0.84	0.94	0.94	0.36	0.52	0.28	0.20	0.32
Sc B	1	1	0.84	0.94	0.94	0.33	0.46	0.22	0.11	0.27
Sc C	0	0	0	0	0	0.22	0.33	0.17	0.09	0.19

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Table 2. Results of the environmental benefit-cost analysis.

Scenario	Management unit	Total Benefits ¹ (\$)	Net Benefits ² (\$)	Net Costs ³ (\$)	Benefit-Cost Ratio
Reference	Watershed	34,932,318	–	–	–
RRAO	Municipalities	39 879 554	4 947 236	13 358 754	0.37
	Subwatershed 1	39 343 865	4 411 547	6 017 262	0.73
	Subwatershed 2	40 168 706	5 236 388	9 282 867	0.56

¹ The benefits are for recreational activities associated with water use for both river segments.

² That is the difference between the benefits of the scenario RRAO for the management unit with the benefits of the base-case scenario.

³ The costs were calculated for a management period of one-year.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Table 3. Q90 values (in $\mu\text{g/L}$) of Atrazine, MCPB and Metolachlor cumulative frequency curves obtained from GIBSI simulations.

	Base case	Scenario		
		1 (1-m filter strip)	2 (application rate)	1+2
Atrazine	0.454	0.080	0.060	0.026
MCPB	1.913	1.086	0.833	0.484
Metolachlor	0.423	0.179	0.133	0.000

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

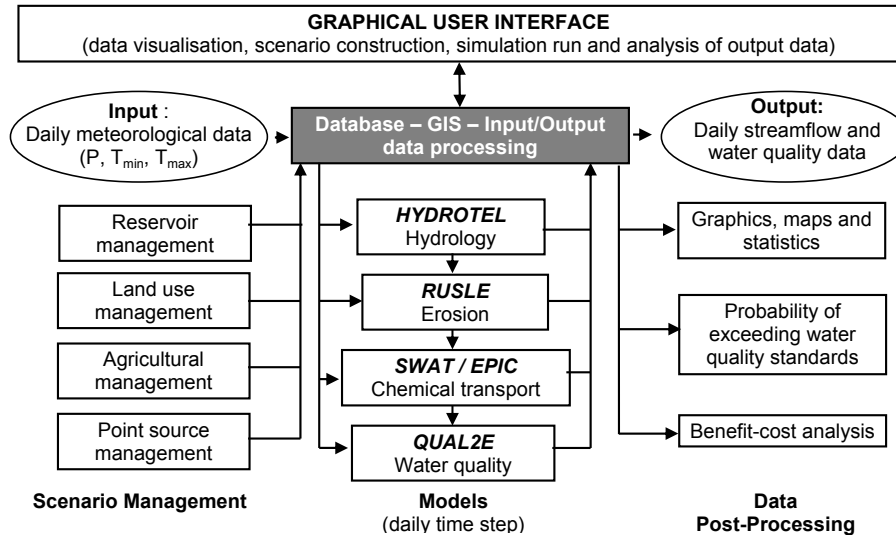


Fig. 1. Representation of GIBSI components and structure.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

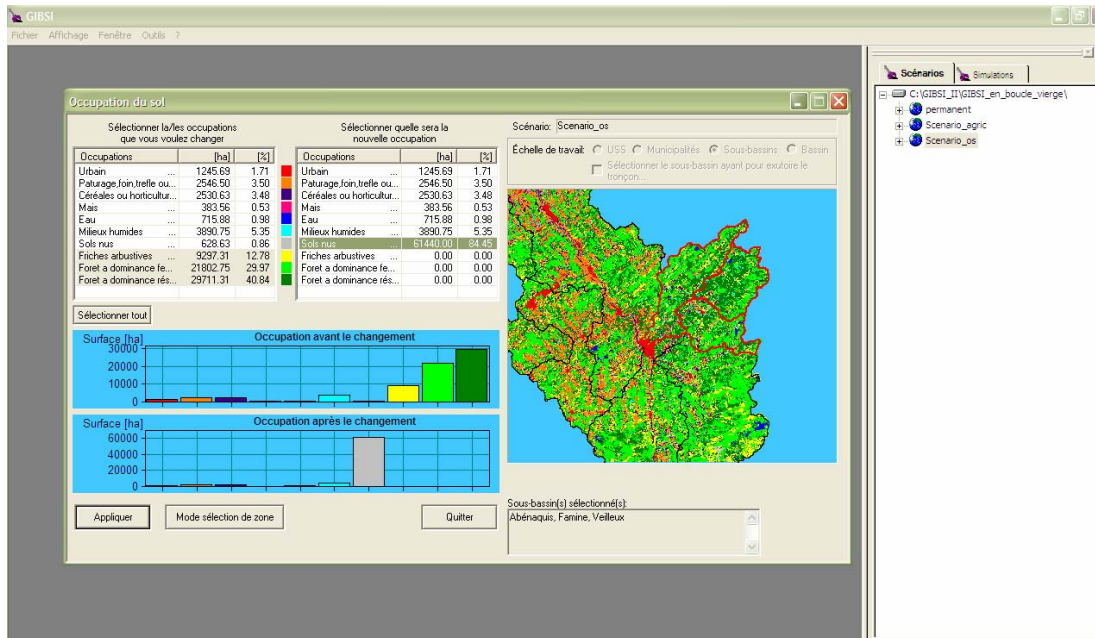


Fig. 2. Example of GIBSI window for the definition of land use management scenario.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Fig. 3. Chaudière River watershed.

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and
A. N. Rousseau

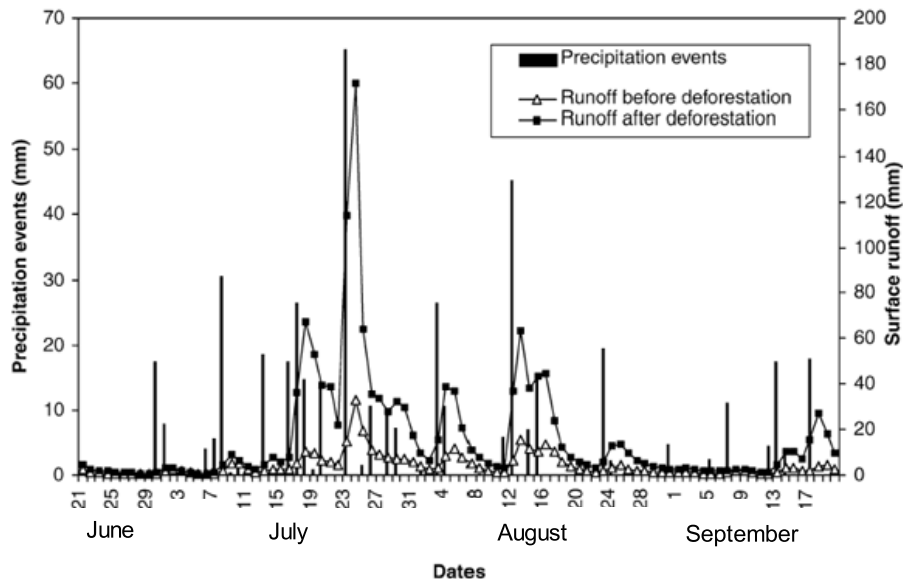


Fig. 4. Peak runoff increase after deforestation for the 1995 meteorological series (from Lavigne et al., 2004).

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Applications and developments of GIBSI

R. Quilbé and A. N. Rousseau

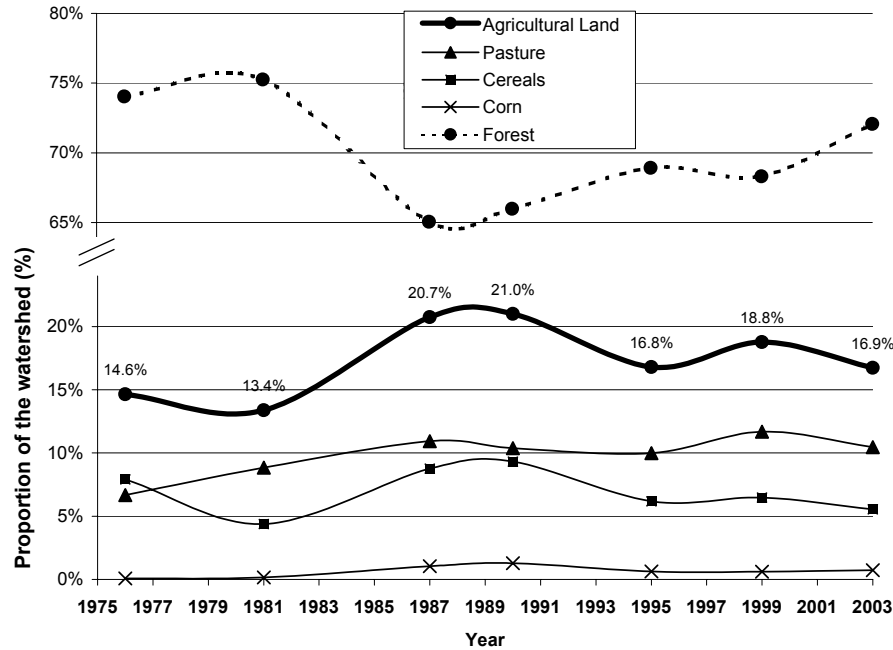


Fig. 5. Evolution of land use on the Chaudière River watershed from 1976 to 2004.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

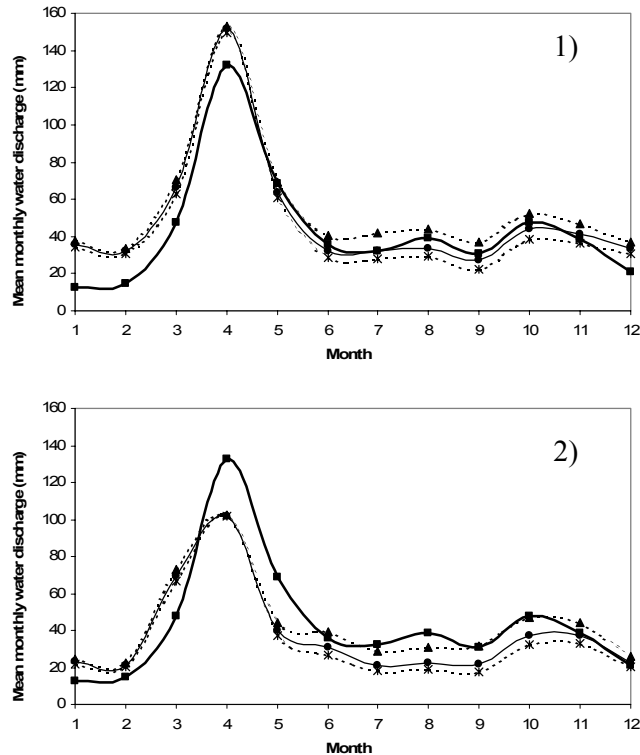


Fig. 6. Effect of climate change and land use evolution on mean monthly water discharge. Only the two GCM/GES combinations that give the extreme effects among all GCM/GES are considered here: HadCM3-A2b (1) and ECHAM4-A2 (2). In both cases, the bold line with squares represents the present conditions as simulated by the model. Other lines represent future water discharge considering base case land use evolution scenario (thin line, circles), scenario A (dotted line, triangles) and scenario B (dotted line, stars).

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion