

# **The Normalized Difference Infrared Index (NDII) as a proxy for root zone moisture content**

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## **Abstract**

With remote sensing we can readily observe the Earth's surface, but direct observation of the sub-surface remains a challenge. In hydrology, but also in related disciplines such as agricultural and atmospheric sciences, knowledge on the dynamics of soil moisture in the root zone of vegetation is essential, as this part of the vadose zone is the core component controlling the partitioning of water into evaporative fluxes, drainage, recharge and runoff. In this paper we tested a novel approach to estimate the catchment-scale soil moisture content in the root zone of vegetation, by using the remotely sensed Normalised Difference Infrared Index (NDII) in the Upper Ping River Basin (UPRB) in Northern Thailand. The NDII is widely used to monitor the Equivalent Water Thickness (EWT) of leaves and canopy. Satellite data from the Moderate Resolution Imaging Spectro-radiometer (MODIS) were used to determine the NDII over an 8-day period, covering the study area from 2001 to 2013. The results show that NDII values decrease sharply at the end of the wet season in October and reach lowest values near the end of the dry season in March. The values then increase abruptly after rains have started, but vary in an insignificant manner from the middle to the late rainy season. This paper hypothesizes that the NDII can be used as a proxy for moisture deficit and hence for the amount of moisture stored in the root zone of vegetation, which is a crucial component of hydrological models. During periods of moisture stress, the 8-day average NDII values were found to correlate well with the 8-day average soil moisture content ( $S_u$ ) simulated by the lumped conceptual hydrological rainfall-runoff model FLEX for 8 sub-catchments in the Upper Ping basin. Even the deseasonalized  $S_u$  and NDII (after subtracting the dominant seasonal signal) showed good correlation during periods of moisture stress. The results clearly show the usefulness of the NDII as a proxy for catchment-scale root zone moisture deficit and therefore as a potentially valuable constraint for the internal dynamics of hydrological models. In dry periods, when plants are exposed to water stress, the EWT (reflecting leaf-water deficit)

decreases steadily, as moisture stress in the leaves is connected to moisture deficits in the root zone. When subsequently the soil moisture is replenished as a result of rainfall, the EWT increases without delay. Once leaf-water is close to saturation - mostly during the heart of the wet season - leaf characteristics and NDII values are not well correlated. However, for both hydrological modelling and water management the stress periods are most important, which is why this product has the potential to becoming a highly efficient model constraint, particularly in ungauged basins.

## 1. Introduction

Estimating the moisture content of the soil from remote sensing is one of the major challenges in the field of hydrology (e.g. [De Jeu et al., 2008](#); [Entekhabi et al., 2010](#)). Soil moisture is generally seen as the key hydrological state variable determining the partitioning of fluxes (into direct runoff, recharge and evaporation) ([Liang et al., 1994](#)), the interaction with the atmosphere ([Legates et al., 2011](#)), and the carbon cycle ([Porporato et al., 2004](#)). The root zone of ecosystems, being the dynamic part of the unsaturated zone, is the key part of the soil related to numerous sub-surface processes ([Shukla and Mintz, 1982](#)). Several remote sensing products have been developed especially for monitoring soil moisture (e.g. SMOS, ERS and AMSR-E), but until now correlations between remote sensing products and observed soil moisture at different depths have been modest at best ([Parajka et al., 2006](#); [Ford et al., 1997](#)). There are a few possible explanations. One is that it is not (yet) possible to look into the soil deep enough to observe soil moisture in the root zone of vegetation ([Shi et al., 1997](#); [Entekhabi et al., 2010](#)), second is that soil moisture observations at certain depths are maybe not the right indicators for the amount of moisture stored in the root zone ([Mahmood and Hubbard, 2007](#)), which is rather determined by the vegetation dependent, spatially variable three-dimensional distribution and density of roots.

These mainstream methods to derive soil moisture from remote sensing have concentrated on direct observation of soil moisture below the surface. The vegetation, through the Vegetation Water Content (VWC), perturbs this picture. As a result, previous studies have tried to determine the VWC from a linear relationship with the Equivalent Water Thickness (EWT) that is measured by the Normalised Difference Infrared Index (NDII) (e.g. [Yilmaz et al., 2008](#)). The NDII was developed by [Hardisky et al. \(1983\)](#) using ratios of different values of near infrared reflectance (NIR) and short wave infrared reflectance (SWIR), defined by:  $(\rho_{\text{NIR}} - \rho_{\text{SWIR}}) / (\rho_{\text{NIR}} + \rho_{\text{SWIR}})$ , similar to the NDVI, which is defined by discrete red and near infrared. Besides for determining the water content of vegetation, the NDII can be effectively used to detect plant water stress according to the property of shortwave infrared reflectance, which is negatively related to leaf water content due to

the large absorption by the leaf (e.g. [Steele-Dunne et al., 2012](#); [Friesen et al., 2012](#); [Van Emmerik et al., 2015](#)). Many studies have found relationships between the equivalent water thickness (EWT) and reflectance at the near-infrared (NIR) and shortwave infrared (SWIR) portion of the spectrum used for deriving NDII ([Hardisky et al., 1983](#); [Hunt and Rock, 1989](#); [Gao, 1996](#); [Ceccato et al., 2002](#); [Fensholt and Sandholt, 2003](#)). [Yilmaz et al. \(2008\)](#) found a significant linear relationship ( $R^2 = 0.85$ ) between equivalent water thickness (EWT) and NDII. Subsequently, they tried to determine a relationship between EWT and vegetation water content (VWC), in order to be able to correct direct moisture observations from space. However, these relationships appeared to be vegetation and crop-type dependent.

Water is one of the determinant environmental variables for vegetation growth, especially in water-limited ecosystems during dry periods. From plant physiology point of view, water absorption from the root zone is driven by osmosis. Subsequently, water transport from the roots to the leaves is driven by water potential differences, caused by diffusion of water out of stomata, called transpiration. This physiological relationship supports the correlation between root zone soil moisture content, moisture tension in the leaves and the water content of plants.

Hence, the root zone moisture deficit is connected to the water content of the canopy/leaves, because soil moisture suction pressure and moisture content in the leaves are directly connected ([Rutter and Sands, 1958](#)). The NDII was developed to monitor leave water content ([Hardisky et al., 1983](#)), so one would expect a direct relation between NDII and root zone moisture deficit. The deficit again is a direct function of the amount of moisture stored in the root zone.

So if leaf water thickness and the suction pressure in the root zone are connected, then the NDII would directly reflect the moisture content of the root zone. It would only reflect the moisture content in the influence zone of roots and not beyond that. Hence the NDII could become a powerful indicator for monitoring root zone moisture content, providing an integrated, depth-independent estimation of how much water is accessible to roots, available for vegetation. In other words, the NDII would allow us to see vegetation as a sort of natural manometer, providing us with information on how much water is available in the sub-surface for use by vegetation. It would be an integrated indicator of soil moisture in the root zone, available directly at the scale of interest.

Thus, the hypothesis is that we can derive the moisture content in the root zone from the observed moisture state of the vegetation by means of the NDII.

In this paper, we tested whether there exists a direct and functional relationship between a remote sensing product (the NDII) and the amount of moisture stored in the root zone, as simulated by a

conceptual hydrological model, in which the root zone moisture content is a key state variable in the short and long term dynamics of the rainfall-runoff signal.

The analysis was done using data from the Upper Ping River Basin (UPRB), a tropical seasonal evergreen catchment in northern Thailand. This catchment is adequate for the purpose because there is clear seasonal dynamics, there is a variety of well-gauged sub-catchments with different aridity characteristics, while the phenology is less variable, being an evergreen ecosystem.

## 2. Study site and data

### 2.1 Study site

The Upper Ping River Basin (UPRB) is situated between latitude 17°14'30" to 19°47'52" N, and longitude 98°4'30" to 99°22'30" E in Northern Thailand and can be separated into 14 sub-basins (Fig. 1) (Mapiam, et al., 2014). It has an area of approximately 25,370 km<sup>2</sup> in the provinces of Chiang Mai and Lam Phun. The basin landform ranges from an undulating to a rolling terrain with steep hills at elevations of 1,500 to 2,000 m, and valleys of 330 to 500 m (Mapiam and Sriwongsitanon, 2009; Sriwongsitanon, 2010). The Ping River originates in Chiang Dao district, north of Chiang Mai, and flows downstream to the south to become the inflow for the Bhumiphol dam - a large dam with an active storage capacity of about 9.7 billion m<sup>3</sup> (Sriwongsitanon, 2010). The climate of the region is controlled by tropical monsoons, with distinctive dry and wet seasons and free from snow and ice. The rainy season is influenced by the southwest monsoon and brings about mild to heavy rainfall between May and October. Annual average rainfall and runoff of the UPRB are approximately 1,170 and 270 mm/y, respectively. Avoiding the influence of other factors, these catchments are ideal cases to concentrate on the relationship between NDII and root zone moisture content. The land cover of the UPRB is dominated by forest (Sriwongsitanon and Taesombat, 2011).

## 2.2 Data Collection

### 2.2.1 Satellite data

The satellite data used for calculating the NDII is the MODIS level 3 surface reflectance product (MOD09A1), which is available at 500 m resolution in an 8-day composite of the gridded level 2 surface reflectance products. Each product pixel contains the best possible L2G observation during an 8-day period selected on the basis of high observation coverage, low view angle, absence of

clouds or cloud shadow, and aerosol loading. MOD09 (MODIS Surface Reflectance) is a seven-band product, which provides an estimate of the surface spectral reflectance for each band as it would have been measured at ground level without atmospheric scattering or absorption. This product has been corrected for the effects of atmospheric gases and aerosols (Vermote et al., 2011). The available MODIS data covering the UPRB from 2001 to 2013 were downloaded from <ftp://e4ftl01.cr.usgs.gov/MOLT>. The HDF-EOS Conversion Tool was applied to extract the desired bands (bands 2 (0.841-0.876  $\mu\text{m}$ ) and 6 (1.628-1.652  $\mu\text{m}$ )) and re-projected into Universal Transverse Mercator (Zone 47N, WGS84) from the original ISIN mapping grid.

### 2.2.2 Rainfall data

Data from 65 non-automatic rain-gauge stations covering the period from 2001 to 2013 were used. 42 stations are located within the UPRB while 23 stations are situated in its surroundings. These rain gauges are owned and operated by the Thai Meteorological Department and the Royal Irrigation Department. Quality control of the rainfall data was performed by comparing them to adjacent rainfall data. For each sub-basin, daily spatially averaged rainfall, by inverse distance squared, has been used as the forcing data of the hydrological model.

### 2.2.3 Runoff data

Daily runoff data from 1995 to 2011 at 8 stations located in the UPRB were adequate to be used for FLEX<sup>L</sup> calibration. These 8 stations are operated by the Royal Irrigation Department in Thailand. The locations of these 8 stations and the associated sub-basins are shown in Fig. 1. Runoff data at these stations are not affected by large reservoirs and have been checked for their reliability by comparing them with rainfall data covering their catchment areas at the same periods. Catchment characteristics and available data periods for model calibration of the selected 8 sub-basins are summarized in Table 3.

## 2.3 NDII drought index for the UPRB

The NDII from 2001 to 2013, covering the UPRB, was computed using MODIS bands 2 and 6 reflectance data. The 8-day surface reflectance data of near infrared (band 2: wavelength between 0.841-0.876  $\mu\text{m}$ ) and short wave infrared (band 6: wavelength between 1.628-1.652  $\mu\text{m}$ ) are described by Eq. (1). The 8-day NDII values were averaged over each sub-basin to allow comparison to the 8-day average  $S_u$  (root zone storage reservoir) values extracted from the FLEX<sup>L</sup> model results at each of the 8 runoff stations.

We did not use field observations of soil moisture. One could argue that field observations should be used to link NDII to moisture stress. However, besides not being available, it is doubtful if point observations at fixed depth would provide a correct measure for the moisture content in the root zone. It is more likely that vegetation distributes its roots and adjusts its root density to the specific local conditions and that the root density and distribution is not homogeneous in space and depth.

### 3. Methods

#### 3.1 Estimating vegetation water content using near infrared and short wave infrared

Estimates of vegetation water content (the amount of water in stems and leaves) are of interest to assess the vegetation water status in agriculture and forestry and have been used for drought assessment (Cheng et al., 2006; Gao, 1996; Gao and Goetz, 1995; Ustin et al., 2004; Peñuelas et al., 1993). Evidence from physically-based radiative transfer models and laboratory studies suggests that changes in water content in plant tissues have a large effect on the leaf reflectance in several regions of the 0.7-2.5  $\mu\text{m}$  spectrum (Fensholt and Sandholt, 2003). Tucker (1980) suggested that the spectral interval between 1.55 and 1.75  $\mu\text{m}$  (SWIR) is the most suitable region for remotely sensed leaf water content. It is well known that these wavelengths are negatively related to leaf water content due to a large absorption by leaf water (Tucker, 1980; Ceccato et al., 2002). However, variations in leaf internal structure and leaf dry matter content also influence the SWIR reflectance. Therefore, SWIR reflectance values alone are not suitable for retrieving vegetation water content. To improve the accuracy in estimating the vegetation water content, a combination of SWIR and NIR (0.7 to 0.9  $\mu\text{m}$ ) reflectance information was utilized because NIR is only affected by leaf internal structure and leaf dry matter content but not by water content. A combination of SWIR and NIR reflectance information can remove the effect of leaf internal structure and leaf dry matter content and can improve the accuracy in retrieving the vegetation water content (Ceccato et al., 2001; Yilmaz et al., 2008; Fensholt and Sandholt, 2003).

On the basis of this idea, Hardisky et al. (1983) derived the NDII:

$$NDII = \frac{\rho_{0.85} - \rho_{1.65}}{\rho_{0.85} + \rho_{1.65}} \quad (1)$$

where  $\rho_{0.85}$  and  $\rho_{1.65}$  are the reflectances at 0.85  $\mu\text{m}$  and 1.65  $\mu\text{m}$  wavelengths, respectively. NDII is a normalized index and the values theoretically vary between -1 and 1. A low NDII value and

190 especially below zero means that reflectance from  $\rho_{0.85}$  is lower than the reflectance from  $\rho_{1.65}$   
191 which indicates canopy water stress.

### 192 3.2 The semi-distributed FLEX Model

193 The relationship between spatially average NDII and root zone moisture content has been  
194 evaluated in eight sub-basins of the UPRB. Because the NDII is an indicator for water stress, the  
195 index is only expected to show a strong link with the moisture content of the root zone when there  
196 is a soil moisture deficit. Without water stress occurring within the leaves, particularly during wet  
197 periods, NDII would possibly not reflect variation in root zone soil moisture content (Korres et al.,  
198 2015).

199 The remotely sensed NDII values have been compared to the root zone storage as modelled by a  
200 semi-distributed conceptual model; semi-distributed meaning that for each sub-catchment a  
201 separate conceptual model has been used. The different sub-catchments demonstrate a variety of  
202 climatic properties that allow a more rigorous test than a fully lumped model could provide. In this  
203 way, a compromise has been found between the complexity and data requirements of a fully  
204 distributed model and the simplicity of a completely lumped model. One could argue that a fully  
205 distributed conceptual model would have been a better tool to assess the spatial and temporal  
206 pattern obtained by the NDII. This is correct, but this would have required the availability of more  
207 detailed spatially distributed forcing data (particularly rainfall), which was not available.  
208 Moreover, if a semi-distributed lumped model, potentially less accurate than a distributed model,  
209 provides a good correlation with NDVI, then this would be a tougher text than with a fully  
210 distributed model.

211 FLEX (Fig. 2) is a conceptual hydrological model with an HBV-like model structure developed in  
212 a flexible modelling framework (Fenicia et al., 2011; Gao et al., 2014a; Gao et al., 2014b). The  
213 model structure comprises four conceptual reservoirs: the interception reservoir  $S_i$  (mm), the root  
214 zone reservoir representing the moisture storage in the root zone  $S_u$  (mm), the fast response  
215 reservoir  $S_f$  (mm), and the slow response reservoir  $S_s$  (mm). It also includes two lag functions  
216 representing the lag time from storm to peak flow ( $T_{lagF}$ ), and the lag time of recharge from the  
217 root zone to the groundwater ( $T_{lagS}$ ). Besides a water balance equation, each reservoir has process  
218 equations that connect the fluxes entering or leaving the storage compartment to the storage in the  
219 reservoirs (so-called constitutive functions). Table 1 shows 15 mathematical expressions used for  
220 modelling the FLEX. A total of 11 model parameters with their distribution values are shown in  
221 Table 2 and they have to be identified by model calibration. Forcing data include the elevation-  
222 corrected daily average rainfall (Gao et al., 2014a), daily average, minimum and maximum air



223 temperature, and potential evaporation derived by Hargreaves equation (Hargreaves and Samani,  
224 1985).

### 225 3.2.1 Interception reservoir

226 The interception evaporation  $E_i$  ( $\text{mm d}^{-1}$ ) is calculated by potential evaporation  $E_0$  ( $\text{mm d}^{-1}$ ) and  
227 the storage of the interception reservoir  $S_i$  (mm) (Eq. (3)). There is no effective rainfall  $P_e$  ( $\text{mm d}^{-1}$ )  
228 as long as the  $S_i$  is less than its storage capacity  $S_{i,\max}$  (mm) (Eq. (4)) (de Groen and Savenije,  
229 2006).

### 230 3.2.2 Root zone reservoir

231 The moisture content in the root zone is simulated by a 'reservoir' that partitions effective rainfall  
232 into infiltration, and runoff  $R$  ( $\text{mm d}^{-1}$ ), and determines the transpiration by vegetation. Therefore,  
233 it is the core of the FLEX model. For the partitioning between infiltration and runoff we applied  
234 the widely used beta function (Eq. (6)) of the Xinanjiang model (Zhao, 1992; Liang et al., 1992),  
235 developed based on the variable contribution area theory (Hewlett and Hibbert, 1967; Beven,  
236 1979), but which can equally reflect the spatial probability distribution of runoff thresholds. The  
237 moisture storage in the root zone 'reservoir' is represented by  $S_u$  (mm). The beta function defines  
238 the runoff percentage  $C_r$  (-) for each time step as a function of the relative soil moisture content  
239 ( $S_u/S_{u,\max}$ ). In Eq. (6),  $S_{u,\max}$  (mm) is the root zone storage capacity, and  $\beta$  (-) is the shape parameter  
240 describing the spatial distribution of the root zone storage capacity over the catchment. In Eq. (7),  
241 the relative soil moisture and potential evaporation are used to determine the transpiration  $E_t$  ( $\text{mm}$   
242  $\text{d}^{-1}$ );  $C_e$  (-) indicates the fraction of  $S_{u,\max}$  above which the transpiration is no longer limited by soil  
243 moisture stress ( $E_t = E_0 - E_i$ ).

### 244 3.2.3 Response routine

245 In Eq. (8),  $R_f$  ( $\text{mm d}^{-1}$ ) indicates the flow into the fast response routine;  $D$  (-) is a splitter to  
246 separate recharge from preferential flow. In Eq. (9),  $R_s$  ( $\text{mm d}^{-1}$ ) indicates the flow into the  
247 groundwater reservoir. Equation (10) and (11) are used to describe the lag time between storm and  
248 peak flow.  $R_f(t-i+1)$  is the generated fast runoff from the root zone at time  $t-i+1$ ;  $T_{\text{lag}}$  is a  
249 parameter which represents the time lag between storm and fast runoff generation;  $c(i)$  is the  
250 weight of the flow in  $i-1$  days before; and  $R_{\text{fl}}(t)$  is the discharge into the fast response reservoir  
251 after convolution.

252 The linear response reservoirs, representing linear relationships between storages and releases, are  
253 applied to conceptualize the discharge from the surface runoff reservoir, fast response reservoir



and slow response reservoir. In Eq. (12),  $Q_{ff}$  (mm d<sup>-1</sup>) is the surface runoff, with timescale  $K_{ff}$ (d), activated when the storage of fast response reservoir exceeds the threshold  $S_{f,max}$  (mm). In Eq. (14) and (16),  $Q_f$  (mm d<sup>-1</sup>) and  $Q_s$  (mm d<sup>-1</sup>) represent the fast and slow runoff;  $K_f$  (d) and  $K_s$  (d) are the time scales of the fast and slow runoff, respectively.  $Q_m$  (mm d<sup>-1</sup>) is the total amount of runoff simulated from the three individual components, including  $Q_{ff}$ ,  $Q_f$ , and  $Q_s$ .

### 3.2.4 Model calibration

A multi-objective calibration strategy has been adopted in this study to allow for the model to effectively reproduce different aspects of the hydrological response, i.e. high flow, low flow and the flow duration curve. The model was therefore calibrated to three Kling-Gupta efficiencies (Gupta et al., 2009): 1) the K-G efficiency of flows ( $I_{KGE}$ ) measures the performance of hydrograph reproduction especially for high flows; 2) the K-G efficiency of the logarithm of flows emphasizes low flows ( $I_{KGL}$ ), and 3) the K-G efficiency of the flow duration curve ( $I_{KGF}$ ) to represent the flow statistics.

The MOSCEM-UA (Multi-Objective Shuffled Complex Evolution Metropolis-University of Arizona) algorithm (Vrugt et al., 2003) was used as the calibration algorithm to find the Pareto-optimal solutions defined by the mentioned three objective functions. This algorithm requires 3 parameters including the maximum number of iterations, the number of complexes, and the number of random samples that is used to initialize each complex. To ensure fair comparison, the parameters of MOSCEM-UA were set based on the number of model parameters. Therefore, the number of complexes is equal to the number of free parameters  $n$ ; the number of random samples is equal to  $n*n*10$ ; and the number of iterations was set to 30000. The model is a widely validated model, which is only used here to derive the magnitude of the root zone moisture storage. Therefore validation is not considered necessary, since the model is merely meant to compare calibrated values of  $S_u$  with NDII.

### 3.3 Deseasonalization

Seasonal signals exist both in NDII and  $S_u$  time series. This can lead to spurious correlation. Therefore we deseasonalized both signals to eliminate this strong signal (Schaeffli & Gupta, 2007) and subsequently compare the deviations from the seasonal signals of both NDII and  $S_u$ . Firstly, the NDII and  $S_u$  were normalized between 0 and 1. Then seasonal patterns of NDII and  $S_u$  were determined as the average seasonal signals, after which they were subtracted from the normalised data.

## 286 4. Results

### 287 4.1 Spatial and seasonal variation of NDII values for the UPRB and its 14 sub-basins

288 To demonstrate the spatial and seasonal behaviour of the NDII over the UPRB, the 8-day NDII  
289 values were aggregated to monthly values for 2001 to 2013. Figure 3 shows examples of monthly  
290 average NDII values for the UPRB in 2004, which is the year with the lowest annual average NDII  
291 value. The figure shows that NDII values are higher during the wet season (May to October) and  
292 lower during the dry season (from November to April). The lower amounts of rainfall between  
293 November and April cause a continuous reduction of NDII values. On the other hand, higher  
294 amounts of rainfall between May and October result in increasing NDII values. However, NDII  
295 values appear to vary little between July and October.

296 The average NDII values during the wet season, the dry season, and the whole year within the 13  
297 years are presented in Table 4. The table also shows the order of the NDII values from the highest  
298 (number 1) to the lowest (number 13). It can be seen that the annual average NDII value for the  
299 whole basin is approximately 0.165, while the average values during the wet and dry season are  
300 about 0.211 and 0.118, respectively. The highest mean annual value (NDII = 0.177) occurred in  
301 2002-2003 and the lowest (NDII = 0.149) in 2004-2005. The highest (NDII = 0.149) and lowest  
302 (NDII = 0.088) dry season values were reported in 2002-2003 and 2004-2005, respectively. On the  
303 other hand, the highest (NDII = 0.224) and lowest (NDII = 0.197) wet season values were  
304 observed in 2006-2007 and 2010-2011, respectively. It can be concluded that a dry season with  
305 relatively low moisture content and a wet season with high moisture content as specified by NDII  
306 values do not normally occur in the same year.

307 The 8-day NDII values were also computed for each of the 14 tributaries within the UPRB from  
308 2001 to 2013. Table 5 shows the monthly averaged NDII values between 2001 and 2013 and the  
309 ranking order for each of the 14 tributaries. The results suggest that Nam Mae Taeng, Nam Mae  
310 Rim, and Upper Mae Chaem sub-basins, which have higher mean annual NDII values, have a  
311 higher moisture content than other sub-basins; while Nam Mae Haad, Nam Mae Li, and Ping  
312 River Section 2 are 3 sub-basins, with lower mean annual NDII values, have lower moisture  
313 content than other sub-basins. Monthly average NDII values for these 6 sub-basins are presented  
314 in Fig. 4. It can be seen that during the dry season, NDII values of the 3 sub-basins with the lowest  
315 values are a lot lower than those of the 3 sub-basins with the highest NDII values. However, NDII  
316 values for these 2 groups are not significantly different during the wet season. The figure also  
317 reveals that NDII values tend to continuously increase from relatively low values in March to

318 higher values in June. The values slightly fluctuate during the wet season before sharply falling  
319 once again when the rainy season ends, and reach their minimum values in February.

## 320 **4.2 FLEX Model results**

321 Calibration of FLEX was done on the 8 sub-catchments that have runoff stations. The results are  
322 summarized in Table 6. The performance of the model was quite good as demonstrated in Table 7.  
323 In Fig. 5, the flow duration curves of runoff stations P.20 and P.21 are presented as examples of  
324 model performance. Table 7 shows the average Kling-Gupta efficiencies values for  $I_{KGE}$ ,  $I_{KGL}$  and  
325  $I_{KGF}$ , which indicate the performance of high flows, low flows, and flow duration curve for the 8  
326 runoff stations. The results for the flow duration curve appear to be better than those of the high  
327 flows and especially the low flows. However, the overall results are acceptable and can be used for  
328 further analysis in this study.

## 329 **4.3 Relation between NDII and root zone moisture storage ( $S_u$ )**

330 The 8-day NDII values were compared to the 8 day average root zone moisture storage values of  
331 the FLEX model. It appears that during moisture stress periods, the relationship can be well  
332 described by an exponential function, for each of the 8 sub-catchments. Table 8 presents the  
333 coefficients of the exponential relationships as well as the coefficients of determination ( $R^2$ ) for  
334 annual, wet season, and dry season values for each sub-catchment. The coefficients are merely  
335 meant for illustration. They should not be seen as functional relationships yet. The corresponding  
336 scatter plots are shown in Fig. 6. It can be clearly seen that the correlation is much better in the dry  
337 season than in the wet season. During the wet season, there may also be short period of moisture  
338 stress, where the exponential pattern can be recognized, but no clear relation is found when the  
339 vegetation does not experience any moisture stress.

340 Examples of deseasonalized and scaled time series of NDII and root zone storage ( $S_u$ ) values for  
341 the sub-catchments P.20 and P.21 are presented in Figure 7. The scaled time series of the NDII  
342 and  $S_u$  values were calculated by dividing their value by the differences between their maximum  
343 and minimum values:  $NDII/(NDII_{max}-NDII_{min})$  and  $S_u/(S_{u,max}-S_{u,min})$ , respectively, while the  
344 maximum and the minimum are the values within the overall considered time series. Figure 7  
345 shows that the scaled NDII and  $S_u$  values are highly correlated during the dry season, but less so  
346 during the wet season. These results confirm the potential of NDII to effectively reflect the  
347 vegetation water content, which, through the suction pressure exercised by the moisture deficit,  
348 relates to the moisture content in the root zone. During dry periods, or during dry spells in the

rainy season, as soon as the leaves of the vegetation experience suction pressure, we see high values of the coefficient of determination.

If the soil moisture in the root zone is above a certain threshold value, then the leaves are not under stress. In the UPRB this situation occurs typically during the middle and late rainy season. The NDII then does not vary significantly while the root zone moisture storage may still vary, albeit above the threshold where moisture stress occurs. This causes a lower correlation between NDII and root zone storage during wet periods. Interestingly, even during the wet season dry spells can occur. We can see in Fig. 6, that during such a dry spell, the NDII and  $S_u$  again follow an exponential relationship.

We can see that the  $S_u$ , derived merely from precipitation and energy, is strongly correlated to the vegetation water observed by NDII during condition of moisture stress, without time lag (Figure 6, S1, S2). Introduction of a time lag resulted in reduction of the correlation coefficients (Supplementary material). This confirms the direct response of vegetation to soil moisture stress, which confirms that the NDII can be used as a proxy for root zone moisture content.

The deseasonalized results of dry periods in sub-catchments P.20 and P.21 are shown in Figure 7. We found these variations of deseasonalized NDII and  $S_u$  to be similar in these two sub-catchments, with the coefficients of determination ( $R^2$ ) as 0.32 and 0.18 respectively in P.20 and P.21. More important than the coefficient of determination is the similarity between the deseasonalized patterns. For P.20, the year 2001 is almost identical, whereas the years 2004 and 2006 are dissimilar. In general the patterns are well reproduced, especially if we take into account the implicit uncertainties of the lumped hydrological model, the uncertainties in the 8-day derived NDII, and the data of precipitation and potential evaporation used in the model. The results of other sub-basins can be found in the supplementary materials.

## 5. Discussion

### 5.1 Is vegetation a trouble-maker or a good indicator for the moisture content of the root zone?

In bare soil, remote sensors can only detect soil moisture until a few centimetres below the surface (~5cm) (Entekhabi et al., 2010). Unfortunately, for hydrological modelling, the moisture state of the bare surface is of only limited interest. What is of key interest for understanding the dynamics

380 of hydrological systems is the variability of the moisture content of the root zone, in which the  
381 main dynamics take place. This variability determines the rainfall-runoff behaviour, the  
382 transpiration of vegetation, and the partitioning between different hydrological fluxes. However,  
383 observing the soil moisture content in the root zone is still a major challenge (Entekhabi et al.,  
384 2010).

385 What is normally done, is to link the moisture content of the surface layer to the total amount of  
386 moisture in the root zone. Knowing the surface soil moisture, the root zone soil moisture can be  
387 estimated by an exponential decay filter (Albergel et al., 2008; Ford et al., 2014) or by models  
388 (Reichle, 2008). However, the surface soil moisture is only weakly related with root zone soil  
389 moisture (Mahmood and Hubbard, 2007); it only works if there is connectivity between the  
390 surface and deeper layers and when a certain state of equilibrium has been reached (when the short  
391 term dynamics after a rainfall event has levelled out). It is also observed that the presence of  
392 vegetation prevents the observation of soil moisture and further deteriorates the results (Jackson  
393 and Schmugge, 1991). Avoiding the influence of vegetation in observing soil moisture (e.g. by  
394 SMOS or SMAP) is seen as a challenge by some in the remote sensing community (Kerr et al.,  
395 2001; Entekhabi et al., 2010). Several algorithms have been proposed to filter out the vegetation  
396 impact (Jackson and Schmugge, 1991), also based on NDII (e.g. Yilmaz et al., 2008). But is  
397 vegetation a trouble-maker, or does it offer an excellent opportunity to directly gauge the state of  
398 the soil moisture?

399 In this study, we found that vegetation rather than a problem could become key to sensing the  
400 storage dynamics of moisture in the root zone. The water content in the leaves is connected to the  
401 suction pressure in the root zone (Rutter and Sands, 1958). If the suction pressure is above a  
402 certain threshold, then this connection is direct and very sensitive. We found a highly significant  
403 correlation between NDII and  $S_u$ , particularly during periods of moisture stress. During dry  
404 periods, or during dry spells in the rainy season, as soon as the leaves of the vegetation experience  
405 suction pressure, we see high values of the coefficient of determination. Observing the moisture  
406 content of vegetation provides us with directly information on the soil moisture state in the root  
407 zone. We also found that there is almost no lag time between  $S_u$  and NDII. This illustrates the fast  
408 response of vegetation to soil moisture variation, which makes the NDII a sensitive and direct  
409 indicator for root zone moisture content. In fact, the canopy acts as a kind of manometer for the  
410 root zone moisture content.

## 411 5.2 The validity of the hypothesis

412 In natural catchments, it is not possible to prove a hypothesis by using a calibrated model. There  
413 are too many factors contributing to the uncertainty of results: the processes are too heterogeneous,  
414 the observations are not without error, the climatic drivers are too uncertain and heterogeneous and  
415 finally there is substantial model uncertainty, both in the semi-distributed hydrological model and  
416 in the remote sensing model used to determine the 8-day NDII product. In this case we have  
417 selected a lumped conceptual model, which is good at mimicking the main runoff processes, but  
418 which lacks the detail of distributed models. Distributed models, however, require detailed and  
419 spatially explicit information (which is missing) and are generally over-parameterized, turning  
420 them highly unreliable in data-scarce environments. On top of this there is considerable doubt if  
421 they provide the right answers for the right reasons.

422 This paper is not a modelling study, but a test of the hypothesis whether the observed NDII  
423 correlates with the modelled root zone storage. We have seen in Figure 6 that the correlation is  
424 strong during periods of moisture stress, but that when the root zone is near saturation the  
425 correlation is weak. But we also saw that even in the wet season, during short dry spells, the  
426 correlation is strong. Even when the seasonality is removed, the patterns between NDII and  $S_u$  in  
427 Figure 7 are similar, although there are two dry seasons when this is less the case (in 2004 and  
428 2006). So given the implicit uncertainty of the hydrological model, the uncertainty of the  
429 meteorological drivers, as well as the river discharges to which the models have been calibrated,  
430 and the uncertainty associated with the relationship between NDII and EWT, the good  
431 correspondence between the NDII and the root zone storage of the model during periods of  
432 moisture stress gives strong support to the hypothesis, which therefore cannot be rejected. It is in  
433 our view even likely that the differences between the signals of the NDII and the  $S_u$  are rather  
434 related to model uncertainty, the uncertainty of the climatic drivers, the uncertainty in the  
435 relationship between NDII and EWT, and the problems of determining accurate NDII estimates  
436 over 8-days periods, than due to a weak correlation between the root zone storage and the NDII.

### 437 **5.3 Implication in hydrological modelling**

438 Simulation of root zone soil moisture is crucial in hydrological modelling (Houser et al., 1998;  
439 Western and Blöschl, 1999). Using estimates of soil moisture states could increase model  
440 performance and realism, but moreover, it would be powerful information to facilitate prediction  
441 in ungauged basins (Hrachowitz et al., 2013). However, until now, it has not been practical (e.g.  
442 Parajka et al., 2006; Entekhabi et al., 2010). Assimilating soil moisture in hydrological models,  
443 either from top-soil observation by remote sensing, or from the deeper soil column by models

444 (Reichle, 2008), is still a challenge. Several studies showed how difficult it is to assimilate soil  
445 moisture data to improve daily runoff simulation (Parajka et al., 2006; Matgen et al., 2012).

446 There are several reasons why we have not compared our results with soil moisture observations in  
447 the field. Firstly, observations of soil moisture are not widely available. Moreover, it is not  
448 straightforward to link classical soil moisture observations to the actual moisture available in the  
449 root zone. Most observations are conducted at fixed depths and at certain locations within a highly  
450 heterogeneous environment. Without knowing the details of the root distribution, both horizontally  
451 and vertically, it is hard, if not impossible, to estimate the water volume accessible to plants  
452 through their root systems. We should realize that it is difficult to observe root zone soil moisture  
453 even at a local scale. But measuring root zone soil moisture at a catchment scale is even more  
454 challenging. State-of-the-art remote sensing techniques can observe spatially distributed soil  
455 moisture, but what they can see is only the near-surface layers if not blocked by vegetation. The  
456 top layer moisture may or not be correlated with the root zone storage, amongst others depending  
457 on the vegetation type, but it is definitely not the same.

458 By observing the moisture content of the leaves, the NDII represents the soil moisture content of  
459 the entire root zone, which is precisely the information that hydrological models require as this is  
460 the component that controls the occurrence and magnitude of storage deficits and thereby the  
461 moisture dynamics of a system. This study clearly shows the strong temporal correlation between  
462  $S_u$  and NDII. From the relationship between NDII and  $S_u$ , we can directly derive a proxy for the  
463 soil moisture state at the actual scale of interest, which can potentially be assimilated in  
464 hydrological models. Being such a key state variable, the NDII-derived  $S_u$  could become a  
465 potentially powerful and otherwise unavailable constraint for the soil moisture component of  
466 hydrological models. This would mean a breakthrough in hydrological modelling as it would  
467 allow a robust parameterization of water partitioning into evaporative fluxes and drainage even in  
468 data scarce environments. Given the implicit uncertainties in hydrological modelling, this new and  
469 readily available proxy could potentially enhance our implicitly uncertain modelling practice.  
470 More importantly it would open completely new venues for modelling ungauged parts of the  
471 world and could become extremely useful for discharge prediction in ungauged basins  
472 (Hrachowitz et al., 2013).

473 We should, of course, be aware of regional limitations. This study considered a tropical seasonal  
474 evergreen ecosystem, where periods of moisture stress regularly occur. In ecosystems which shed  
475 their leaves, or go dormant, other conditions may apply. We need further investigations into the  
476 usefulness of this approach in catchments with different climates. In addition, the phenology of the



ecosystem is of importance, which should be taken into consideration in follow-up research. Finally, a comparison with distributed or semi-distributed models would be a further test of the value of the NDII as proxy for the root zone moisture content.

480

## 481 **6. Conclusions**

The NDII was used to investigate drought for the UPRB from 2001 to 2013. Monthly average NDII values appear to be spatially distributed over the UPRB, in agreement with seasonal variability and landscape characteristics. NDII values appear to be lower during the dry season and higher during the wet season as a result of seasonal differences between precipitation and evaporation. The NDII appears to correlate well with the moisture content in the root zone, offering an interesting proxy variable for calibration of hydrological models in ungauged basins.

To illustrate the importance of NDII as a proxy for root zone moisture content in hydrological models, we applied the FLEX model to assess the root zone soil moisture storage ( $S_u$ ) of 8 sub-catchments of the UPRB controlled by 8 runoff stations. The results show that the 8-day average NDII values over the study sub-basin correlate well with the 8-day average  $S_u$  for all sub-catchments during dry periods (average  $R^2$  equals 0.87), and less so during wet spells (average  $R^2$  equals 0.61). The NDII appears to be a good proxy for root zone moisture content during dry spells when leaves are under moisture stress. The natural interaction between rainfall, soil moisture, and leave water content can be visualised by the NDII, making it an important indicator both for hydrological modelling and drought assessment.

The potential of using the NDII to constrain model parameters (such as the power of the beta function  $\beta$ , recharge splitter  $D$  and  $C_e$  in the transpiration function) in ungauged basins is an important new venue, which could potentially facilitate the major question of prediction in ungauged basins.

501

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508 **References**

- 509 Albergel, C., Rüdiger, C., Pellarin, T., Calvet, J. C., Fritz, N., Froissard, F., Suquia, D., Petitpa, A.,  
510 Piguet, B., and Martin, E.: From near-surface to root-zone soil moisture using an exponential  
511 filter: an assessment of the method based on in-situ observations and model simulations, *Hydrol.*  
512 *Earth Syst. Sci.*, 12(6), 1323-1337, 2008.
- 513 Beven, K. J. and M. J. Kirkby. 1979. A physically based, variable contributing area model of basin  
514 hydrology. *Hydrological Sciences Bulletin*. 24(1). 43-69.
- 515 Ceccato, P., Flasse, S., and Grégoire, J. M.: Designing a spectral index to estimate vegetation  
516 water content from remote sensing data: Part 2, Validations and applications, *Remote Sens.*  
517 *Environ.*, 82, 198–207, doi:10.1016/S0034-4257(02)00036-6, 2002.
- 518 Ceccato, P., Flasse, S., Tarantola, S., Jacquemoud, S., and Grégoire, J. M.: Detecting vegetation  
519 leaf water content using reflectance in the optical domain, *Remote Sens. Environ.*, 77, 22– 33,  
520 doi:10.1016/S0034-4257(01)00191-2, 2001.
- 521 Cheng, Y. B., Zarco-Tejada, P. J., Riaño, D., Rueda, C. A., and Ustin, S. L.: Estimating vegetation  
522 water content with hyperspectral data for different canopy scenarios: Relationships between  
523 AVIRIS and MODIS indexes, *Remote Sens. Environ.*, 105, 354–366,  
524 doi:10.1016/j.rse.2006.07.005, 2006.
- 525 de Groen, M. M., and Savenije, H. H. G.: A monthly interception equation based on the statistical  
526 characteristics of daily rainfall, *Water Resour. Res.*, 42, W12417, doi:10.1029/2006WR005013,  
527 2006.
- 528 De Jeu, R. A. M., Wagner, W., Holmes, T. R. H., Dolman, A. J., van de Giesen, N. C., and  
529 Friesen, J.: Global Soil Moisture Patterns Observed by Space Borne Microwave Radiometers and  
530 Scatterometers, *Surv. Geophys.*, 28, 399–420, doi 10.1007/s10712-008-9044-0, 2008.
- 531 Entekhabi, D., Nioku, E. G., O'Neill, P. E., Kellogg, K. H., Crow, W. T., Edelstein, W. N., Entin,  
532 J. K., Goodman, S. D., Jackson, T. J., Johnson, J., Kimball, J., Piepmeier, J. R., Koster, R. D.,  
533 Martin, N., McDonald, K. C., Moghaddam, M., Moran, S., Reichle, R., Shi, J.-C., Spencer, M. W.,  
534 Thurman, S. W., Leung, T., and Van Zyl, J.: The Soil Moisture Active Passive (SMAP) Mission,  
535 *Proc. IEEE.*, 98, 704–716, 2010.
- 536 Fenicia, F., Kavetski, D., and Savenije, H. H. G.: Elements of a flexible approach for conceptual  
537 hydrological modeling: 1. Motivation and theoretical development, *Water Resour. Res.*, 47,  
538 W11510, doi:10.1029/2010WR010174, 2011.

539 Fensholt, R., and Sandholt, I.: Derivation of a shortwave infrared stress index from MODIS near-  
 540 and shortwave infrared data in a semiarid environment, *Remote Sens. Environ.*, 87, 111–121,  
 541 doi:10.1016/j.rse.2003.07.002, 2003.

542 Ford, T. W., Harris, E., and Quiring, S. M.: Estimating root zone soil moisture using near-surface  
 543 observations from SMOS, *Hydrol. Earth Syst. Sci.*, 18(1), 139-154, 2014.

544 Friesen, J., Steele-Dunne, S. C., and van de Giesen, N.: Diurnal Differences in Global ERS  
 545 Scatterometer Backscatter Observations of the Land Surface, *IEEE Transactions on Geoscience*  
 546 and Remote Sensing, vol. 50, issue 7, pp. 2595-2602, 2012.

547 Gao, B. C.: NDWI - A normalized difference water index for remote sensing of vegetation liquid  
 548 water from space, *Remote Sens. Environ.*, 58, 257–266, doi:10.1016/S0034-4257(96)00067-3,  
 549 1996.

550 Gao, B. C., and Goetz, A. F. H.: Retrieval of equivalent water thickness and information related to  
 551 biochemical components of vegetation canopies from AVIRIS data, *Remote Sens. Environ.*, 52,  
 552 155–162, doi:10.1016/0034-4257(95)00039-4, 1995.

553 Gao, H., Hrachowitz, M., Fenicia, F., Gharari, S., and Savenije, H. H. G.: Testing the realism of a  
 554 topography driven model (Flex-Topo) in the nested catchments of the Upper Heihe,  
 555 China, *Hydrol. Earth Syst. Sci.*, 18, 1895–1915, doi:10.5194/hess-18-1895-2014, 2014(a).

556 Gao, H., Hrachowitz, M., Schymanski, S. J., Fenicia, F., Sriwongsitanon, N., and Savenije, H. H.  
 557 G.: Climate controls how ecosystems size the root zone storage capacity at catchment scale,  
 558 *Geophys. Res. Lett.*, 41, 7916–7923, doi:10.1002/2014GL061668, 2014(b).

559 Gupta, H. V., Kling, H., Yilmaz, K. K., and Martinez, G. F.: Decomposition of the mean squared  
 560 error and NSE performance criteria: implications for improving hydrological modeling, *J. Hydrol.*,  
 561 377, 80–91, <http://dx.doi.org/10.1016/j.jhydrol.2009.08.003>, 2009.

562 Hardisky, M. A., V. Klemas, and R. Michael Smart. "The influence of soil salinity, growth form,  
 563 and leaf moisture on the spectral radiance of *Spartina alterniflora* canopies." *Photogrammetric*  
 564 *Engineering and Remote Sensing* 49 (1983): 77-83.

565 Hargreaves GH, Samani ZA: Reference crop evapotranspiration from temperature. *Appl Engine*  
 566 *Agric.* 1(2):96–99, 1985.

567 Houser, P. R., Shuttleworth, W. J., Famiglietti, J. S., Gupta, H. V., Syed, K. H., and Goodrich, D.  
 568 C.: Integration of soil moisture remote sensing and hydrologic modeling using data assimilation,  
 569 edited, 1998.

570 Hunt, E. R. Jr. and Rock, B. N.: Detection of changes in leaf water content using near- and middle-  
 571 infrared reflectances, *Remote Sens. Environ.*, 30, 43–54, doi:10.1016/0034-4257(89)90046-1,  
 572 1989.

573 Hrachowitz, M., Savenije, H.H.G., Blöschl, G., McDonnell, J.J., Sivapalan, M., Pomeroy, J.W.,  
 574 Arheimer, B., Blume, T., Clark, M.P., Ehret, U., Fenicia, F., Freer, J.E., Gelfan, A., Gupta, H.V.,  
 575 Hughes, D.A., Hut, R.W., Montanari, A., Pande, S., Tetzlaff, D., Troch, P.A., Uhlenbrook, S.,  
 576 Wagener, T., Winsemius, H.C., Woods, R.A., Zehe, E., and Cudennec, C.: A decade of  
 577 Predictions in Ungauged Basins (PUB); a review. *Hydrol. Sci. J.*, 58 (6), 1–58, doi:  
 578 10.1080/02626667.2013.803183, 2013.

579 Hewlett, J.D. and Hibbert, A.R. 1967. Factors affecting the response of small watersheds to  
 580 precipitation in humid regions. IN *Forest Hydrology* (eds. W.E. Sopper and H.W. Lull). Pergamon  
 581 Press, Oxford. pp. 275-290.

582 Korres, W., Reichenau, T. G., Fiener, P., Koyama, C. N., Bogen, H. R., Cornelissen, T., Baatz,  
 583 R., Herbst, M., Dieckkrüger, B., Vereecken, H., and Schneider, K.: Spatio-temporal soil moisture  
 584 patterns – A meta-analysis using plot to catchment scale data, *J. Hydrol.*, 520, 326–341,  
 585 doi:10.1016/j.jhydrol.2014.11.042, 2015.

586 Jackson, T. J., and Schmugge, T. J.: Vegetation effects on the microwave emission of soils,  
 587 *Remote Sens. Environ.*, 36(3), 203-212, 1991.

588 Kerr, Y. H., Waldteufel, P., Wigneron, J.-P., Martinuzzi, J.-M., Font, J., and Berger, M.: Soil  
 589 Moisture Retrieval from Space: The Soil Moisture and Ocean Salinity (SMOS) Mission, *IEEE*  
 590 *Transactions on Geoscience and Remote Sensing*, 39(8), 1729-1735, 2001.

591 Legates, D. R., Mahmood, R., Levina, D. F., DeLiberty, T. L., Quiring, S. M., Houser, C., and  
 592 Nelson, F. E.: Soil moisture: A central and unifying theme in physical geography, *Prog. Phys.*  
 593 *Geogr.*, 35, 65–86, 2011.

594 Liang, X., Lettenmaier, D. P., Wood, E. F., and Burges, S. J.: A simple hydrologically based  
 595 model of land surface water and energy fluxes for general circulation models, *Journal of*  
 596 *Geophysical Research: Atmospheres*, 99, 14 415-14 428, 1994.

597 Mahmood, R., and Hubbard, K. G.: Relationship between soil moisture of near surface and  
 598 multiple depths of the root zone under heterogeneous land uses and varying hydroclimatic  
 599 conditions, *Hydrological Processes*, 25, 3449-3462, doi: 10.1002/hyp.6578, 2004.

600 Mapiam, P. P., Sharma, A., and Sriwongsitanon, N.: Defining the Z~R relationship using gauge  
601 rainfall with coarse temporal resolution: Implications for flood forecast. *J. Hydrol. Eng.*, 19,  
602 04014004, doi: 10.1061/(ASCE)HE.1943-5584.0000616, 2014.

603 Mapiam, P. P., and Sriwongsitanon, N.: Estimation of the URBS model parameters for flood  
604 estimation of ungauged catchments in the upper Ping river basin, Thailand. *ScienceAsia*, 35, 49–  
605 56, 2009.

606 Matgen P, Fenicia F, Heitz S, Plaza D, de Keyser R, Pauwels VRN, et al. Can ASCAT-derived  
607 soil wetness indices reduce predictive uncertainty in well-gauged areas? A comparison with in situ  
608 observed soil moisture in an assimilation application. *Adv Water Resour* 2012;44:49–65.  
609 doi:http://dx.doi.org/10.1016/j.advwatres.2012.03.022.

610 Parajka, J., Naeimi, V., Blöschl, G., Wagner, W., Merz, R., and Scipal, K.: Assimilating  
611 scatterometer soil moisture data into conceptual hydrologic models at the regional scale, *Hydrol.*  
612 *Earth Syst. Sci.*, 10(3), 353-368, 2006.

613 Peñuelas, J., Filella, I., Biel, C., Serrano, L., and Savé, R.: The reflectance at the 950–970 nm  
614 region as an indicator of plant water status, *Int. J. Remote Sens.*, 14, 1887–1905,  
615 doi:10.1080/01431169308954010, 1993.

616 Porporato, A., Daly, E., and Rodriguez-Iturbe, I.: Soil water balance and ecosystem response to  
617 climate change, *The American Naturalist*, 164, 625-623, 2004.

618 Reichle, R. H.: Data assimilation methods in the Earth sciences, *Adv. Water Resour.*, 31(11),  
619 1411-1418, 2008.

620 Rutter, A. J., and Sands, K. (1958). The relation of leaf water deficit to soil moisture tension in  
621 *Pinus sylvestris*, L. 1. The effect of soil moisture on diurnal change in water balance. *New*  
622 *Phytol.* 57: 50--65.

623 Schaefli, B., & Gupta, H. V. Do Nash values have value?. *Hydrological Processes*,  
624 21(15), 2075-2080, 2007.

625 Shi, J., Wang, J., Hsu, A. Y., O'Neill, P. E., and Engman, E. T.: Estimation of bare surface soil  
626 moisture and surface roughness parameter using L-band SAR image data, *Geoscience and Remote*  
627 *Sensing, IEEE Transactions on*, 35(5), 1254-1266, 1997.

628 Shukla, J., & Mintz, Y. (1982). Influence of land-surface evapotranspiration on the earth's climate.  
629 *Science*, 215(4539), 1498-1501.

630 Sriwongsitanon, N.: Flood forecasting system development for the Upper Ping River Basin.  
631 Kasetsart Journal (Natural Science), 44(4), 2010.

632 Sriwongsitanon, N. and Taesombat, W.: Effects of land cover on runoff coefficient, J. Hydrol.,  
633 410, 226–238, doi:10.1016/j.jhydrol.2011.09.021, 2011.

634 Steele-Dunne, S. C., Friesen, J., and van de Giesen, N.: Using Diurnal Variation in Backscatter to  
635 Detect Vegetation Water Stress, IEEE Transactions on Geoscience and Remote Sensing, vol. 50,  
636 issue 7, pp. 2618-2629, 2012.

637 Tucker, C. J.: Remote sensing of leaf water content in the near infrared, Remote Sens. Environ.,  
638 10, 23–32, doi:10.1016/0034-4257(80)90096-6, 1980.

639 Ustin, S. L., Roberts, D. A., Gamon, J. A., Asner, G. P. and Green, R. O.: Using Imaging  
640 Spectroscopy to Study Ecosystem Processes and Properties. BioScience, 54, 523–534,  
641 doi:10.1641/0006-3568(2004)054[0523:UISTSE]2.0.CO;2, 2004.

642 Van Emmerik, T., Steele-Dunne, S.C., Judge, J. and van de Giesen, N.C.: Impact of Diurnal  
643 Variation in Vegetation Water Content on Radar Backscatter From Maize During Water Stress,  
644 IEEE Transactions on Geoscience and Remote Sensing, vol. 53, issue 7,  
645 doi:10.1109/TGRS.2014.2386142, 2015.

646 Vermote, E. F., Kotchenova, S. Y., and Ray, J. P.: MODIS Surface Reflectance User's Guide.  
647 Web site: <http://modis-sr.ltdri.org>, 2011.

648 Vrugt, J. A., Gupta, H. V., Bastidas, L. A., Bouten, W. and Sorooshian, S.: Effective and efficient  
649 algorithm for multiobjective optimization of hydrologic models, Water Resour. Res., 39, 1214.  
650 doi:10.1029/2002WR001746, 2003.

651 Western, A. W., and Blöschl, G.: On the spatial scaling of soil moisture, J. Hydrol., 217(3–4), 203–  
652 224, 1999.

653 Yilmaz, M. T., Hunt, E. R. Jr., and Jackson, T. J.: Remote sensing of vegetation water content  
654 from equivalent water thickness using satellite imagery, Remote Sens. Environ., 112, 2514–2522,  
655 [doi:10.1016/j.rse.2007.11.014](https://doi.org/10.1016/j.rse.2007.11.014), 2008.

656 Zhao, R. J.: The Xinanjiang model applied in China, J. Hydrol., 135, 371–381, doi:10.1016/0022-  
657 1694(92)90096-E, 1992.

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659 Table 1. Water balance and constitutive equations used in FLEX<sup>L</sup>.

Reservoirs	Water balance equations	Equation	Constitutive equations	Equation
Interception	$\frac{dS_i}{dt} = P - E_i - P_e$	(2)	$E_i = \begin{cases} E_0; S_i > 0 \\ 0; S_i = 0 \end{cases}$	(3)
			$P_e = \begin{cases} 0; S_i < S_{i,\max} \\ P; S_i = S_{i,\max} \end{cases}$	(4)
Root zone reservoir	$\frac{dS_u}{dt} = P_e - R - E_t$	(5)	$\frac{R}{P_e} = 1 - (1 - \frac{S_u}{(1+\beta)S_{u,\max}})^\beta$	(6)
			$E_t = (E_0 - E_i) \cdot \min(1, \frac{S_u}{C_e S_{u,\max} (1+\beta)})$	(7)
Splitter and Lag function			$R_f = R \cdot D$	(8)
			$R_s = R \cdot (1 - D)$	(9)
			$R_{fl}(t) = \sum_{i=1}^{T_{lag}} c(i) \cdot R_f(t-i+1)$	(10)
			$c(i) = i / \sum_{u=1}^{T_{lag}} u$	(11)
Fast reservoir	$\frac{dS_f}{dt} = R_{fl} - Q_{ff} - Q_f$	(12)	$Q_{ff} = \max(0, S_f - S_{f,\max}) / K_{ff}$	(13)
			$Q_f = S_f / K_f$	(14)
Slow reservoir	$\frac{dS_s}{dt} = R_s - Q_s$	(15)	$Q_s = S_s / K_s$	(16)

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661     Table 2. Parameter range of the FLEX<sup>L</sup> Model.

Parameter	Range	Parameters	Range
$S_{i,max}$ (mm)	(0.1, 6)	$K_{ff}$ (d)	(1, 9)
$S_{u,max}$ (mm)	(10, 1000)	$T_{lagF}$ (d)	(0, 5)
$\beta$ (-)	(0, 2)	$T_{lagS}$ (d)	(0, 5)
$C_e$ (-)	(0.1, 0.9)	$K_f$ (d)	(1, 40)
$D$ (-)	(0, 1)	$K_s$ (d)	(10, 500)
$S_{f,max}$ (mm)	(10, 200)		

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