The Normalized Difference Infrared Index (NDII) as a proxy for root 1 zone moisture storage capacity

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9 Abstract

10 With remote sensing we can readily observe the Earth's surface, but looking under the surface into 11 the root zone of vegetation is still a challenge. Yet knowledge on the dynamics of soil moisture in 12 the root zone is essential for agriculture, land-atmosphere interaction and hydrological modelling, 13 alike. In this paper we developed a novel approach to estimate the soil moisture storage deficit in 14 the root zone of vegetation, by using the remotely sensed Normalised Difference Infrared Index 15 (NDII) in the Upper Ping River Basin (UPRB) in Northern Thailand. Satellite data from the 16 Moderate Resolution Imaging Spectro-radiometer (MODIS) was used to evaluate the NDII over 17 an 8-day period, covering the study area from 2001 to 2013. The results show that NDII values 18 decrease sharply at the end of the wet season in October and reach lowest values near the end of 19 the dry season in March. The values then increase abruptly after rains have started, but vary in an 20 insignificant manner from the middle to the late rainy season. The NDII proves to be a very strong 21 proxy for moisture storage deficit in the root zone, which is a crucial component of hydrological 22 models. In addition, the NDII appears to be a reliable indicator for the temporal and spatial 23 distribution of drought conditions in the UPRB. During periods of moisture stress, the 8-day 24 average NDII values were found to correlate very well with the 8-day average soil moisture content (S_u) simulated by the lumped conceptual hydrological rainfall-runoff model FLEX^L for 8 25 26 sub-catchments in the Upper Ping basin. Even the deseasonalized S_u and NDII (after subtracting 27 the dominant seasonal signal) showed good correlation during periods of moisture stress. The 28 results clearly demonstrate the feasibility of NDII as a proxy for root zone moisture stress. In dry 29 periods, when plants are exposed to water stress, the leaf-water deficit increases steadily, and 30 moisture stress in the leaves is connected to moisture deficits in the root zone. Once leaf-water is 31 close to saturation - mostly during the heart of the wet season - leaf characteristics and NDII 32 values are not well correlated. However, to constrain hydrological models or for water 33 management the stress periods are most important, which is why this product can be practical for

34 both hydrological modelling and water management.

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36 1. Introduction

37 Estimating the moisture content of the soil from remote sensing is one of the main challenges in 38 the field of hydrology (e.g. De Jeu et al., 2008; Entekhabi et al., 2010). Soil moisture is generally 39 seen as the key hydrological state variable determining the partitioning of fluxes (into direct 40 runoff, recharge and evaporation) (Liang et. al., 1994), the interaction with the atmosphere 41 (Legates et. al., 2011), and the carbon cycle (Porporato et al., 2004). The root zone of ecosystems, 42 being the dynamic part of the unsaturated zone, is the key part of the soil related to numerous sub-43 surface processes (Shukla and Mintz, 1982). Several remote sensing products have been developed 44 especially for monitoring soil moisture (e.g. SMOS, ERS and AMSR-E), but until now 45 correlations between remote sensing products and observed soil moisture at different depths have 46 been modest at best (Parajka et al., 2006; Ford et al., 1997). There are a few possible explanations. 47 One is that it is not (yet) possible to look into the soil deep enough to observe soil moisture in the 48 root zone of vegetation (Shi et al., 1997; Entekhabi et al., 2010), second is that soil moisture 49 observations at certain depths are maybe not the right indicators for the moisture storage in the 50 root zone (Mahmood and Hubbard, 2007).

51 The hypothesis is that we can derive the soil moisture storage in the root zone from observing the 52 moisture state of the vegetation. We hypothesize that we can link the root zone moisture storage to 53 the water content of the leaves, because soil moisture suction pressure and moisture content in the 54 leaves are directly connected. Water is one of the determinant environmental variables for 55 vegetation growth, especially in water-limited ecosystems during dry periods. From plant 56 physiology point of view, water absorption from the root zone is driven by osmosis. Subsequently, 57 water transport from the roots to the leaves is driven by water potential differences, caused by 58 diffusion of water out of stomata, which we call transpiration. This physiological relationship 59 supports the correlation between root zone soil moisture, moisture tension in the leaves and the 60 water content of plants. In this paper we try to relate a remote sensing product (the NDII, 61 Normalised Difference Infrared Index) to the root zone storage of a conceptual hydrological 62 model, being a key state variable in the short and long term dynamics of the rainfall-runoff signal. 63 In order to do so, we calibrated a conceptual rainfall-runoff model to observed time series in sub-64 basins of the Upper Ping river basin in Thailand and subsequently compared the temporal 65 variability of the root zone storage to the NDII.

66 The NDII was developed by Hunt and Rock (1989) using ratios of different values of near infrared 67 reflectance (NIR) and short wave infrared reflectance (SWIR), defined by: $(\rho_{NIR}-\rho_{SWIR})/(\rho_{NIR}+\rho_{SWIR})$, similar to the NDVI, which is defined by discrete red and near infrared. 68 NDII can be effectively used to detect plant water stress according to the property of shortwave 69 70 infrared reflectance, which is negatively related to leaf water content due to the large absorption 71 by the leaf (e.g. Steele-Dunne et al., 2012; Friesen et al., 2012; Van Emmerik et al., 2015). Many 72 studies have found relationships between the equivalent water thickness (EWT) and reflectance at 73 the near-infrared (NIR) and shortwave infrared (SWIR) portion of the spectrum used for deriving 74 NDII (Hunt and Rock, 1989; Gao, 1996; Ceccato et al., 2002; Fensholt and Sandholt, 2003). Yilmaz et al. (2008) found a significant linear relationship ($R^2 = 0.85$) between equivalent water 75 76 thickness (EWT) and NDII. They also discovered a significant relationship between vegetation 77 water content (VWC), the most successful parameter for retrieval of soil moisture content from 78 microwave data, and EWT; and between EWT and NDII. VWC was therefore linearly related to 79 NDII. Fensholt and Sandholt (2003) derived a shortwave infrared water stress index (SIWSI or 80 NDII) on a daily basis and found a strong correlation with in situ top layer soil moisture 81 measurements from the semiarid Senegal in 2001 and 2002. NDII was therefore selected in this 82 study of the Upper Ping River Basin (UPRB) in northern Thailand because of its potential in 83 detecting equivalent water thickness within the leaves affected by soil moisture storage in the root 84 zone. The relationship between average NDII and root zone moisture storage was evaluated in 85 sub-basins of the UPRB to be used as an indicator to prove the effectiveness of NDII. However, 86 because the NDII is an indicator for water stress, the index is only expected to show a strong link 87 to moisture storage in the root zone when there is a soil moisture deficit. Without water stress 88 occurring within the leaves, particularly during wet periods, NDII would possibly not reflect 89 variation in root zone soil moisture content (Korres et al., 2015).

Instead of comparing the NDII to observed soil moisture by in-field instrumentation, here another approach has been followed. To acquire the information on root zone soil moisture, the lumped (basin average) FLEX^L conceptual hydrological rainfall-runoff model (Fenicia et al., 2011; Gao et al., 2014a; Gao et al., 2014b) was used and calibrated on 8 runoff stations in sub-basins of the UPRB. The simulated root zone storage variation was then compared to the sub-basin average NDII values over the 8 sub-basins.

- 96 2. Study site and data
- 97 **2.1 Study site**

98 The Upper Ping River Basin (UPRB) is situated between latitude 17°14'30" to 19°47'52" N, and longitude 98°4'30" to 99°22'30" E in Northern Thailand and can be separated into 14 sub-99 basins (Fig. 1) (Mapiam, et al., 2014). It has an area of approximately 25,370 km² in the provinces 100 101 of Chiang Mai and Lam Phun. The basin landform ranges from an undulating to a rolling terrain 102 with steep hills at elevations of 1,500 to 2,000 m, and valleys of 330 to 500 m (Mapiam and 103 Sriwongsitanon, 2009; Sriwongsitanon, 2010). The Ping River originates in Chiang Dao district, 104 north of Chiang Mai, and flows downstream to the south to become the inflow for the Bhumiphol dam - a large dam with an active storage capacity of about 9.7 billion m³ (Sriwongsitanon, 2010). 105 106 The climate of the region is controlled by tropical monsoons, with distinctive dry and wet seasons 107 and free from snow and ice. The rainy season is influenced by the southwest monsoon and brings 108 about mild to heavy rainfall between May and October. Annual average rainfall and runoff of the 109 UPRB are approximately 1,170 and 270 mm/y, respectively. Avoiding the influence of other 110 factors, these catchments are ideal cases to concentrate on the relationship between NDII and root 111 zone moisture storage (S_u) . The land cover of the UPRB is dominated by forest (Sriwongsitanon 112 and Taesombat, 2011).

113 **2.2 Data Collection**

114 2.2.1 Satellite data

115 The satellite data used for calculating the NDII is the MODIS level 3 surface reflectance product 116 (MOD09A1), which is at 500 m resolution in an 8-day composite of the gridded level 2 surface 117 reflectance products. Each product pixel contains the best possible L2G observation during an 8-118 day period as selected on the basis of high observation coverage, low view angle, absence of 119 clouds or cloud shadow, and aerosol loading. MOD09 (MODIS Surface Reflectance) is a seven-120 band product, which provides an estimate of the surface spectral reflectance for each band as it 121 would have been measured at ground level without atmospheric scattering or absorption. This 122 product has been corrected for the effects of atmospheric gases and aerosols (Vermote et al., 123 2011). The available MODIS data covering the UPRB from 2001 to 2013 were downloaded from 124 ftp://e4ftl01.cr.usgs.gov/MOLT. The HDF-EOS Conversion Tool was applied to extract the 125 desired bands (bands 2 (0.841-0.876 µm) and 6 (1.628-1.652 µm)) and re-projected into Universal 126 Transverse Mercator (Zone 47N, WGS84) from the original ISIN mapping grid.

127 2.2.2 Rainfall data

- 128 A total of 65 non-automatic rain-gauge stations were selected from 2001 to 2013. 42 stations are
- 129 located within the UPRB while 23 stations are situated in its surroundings. These rain gauges are
- 130 owned and operated by the Thai Meteorological Department and the Royal Irrigation Department.
- 131 Quality control of the rainfall data was performed by comparing them to adjacent rainfall data.
- 132 Rainfall is used as the forcing data of the hydrological model.

133 **2.2.3 Runoff data**

Daily runoff data from 1995 to 2011 at 8 stations located in the UPRB were adequate to be used for FLEX^L calibration. These 8 stations are operated by the Royal Irrigation Department in Thailand. The locations of these 8 stations and the associated sub-basins are shown in Fig. 1. Runoff data at these stations are not affected by large reservoirs and have been checked for their reliability by comparing them with rainfall data covering their catchment areas at the same periods. Catchment characteristics and available data periods for model calibration of the selected 8 sub-basins are summarized in Table 3.

141 **2.3 NDII drought index for the UPRB**

The NDII from 2001 to 2013, covering the UPRB, was computed using MODIS bands 2 and 6 reflectance data. The 8-day surface reflectance data of near infrared (band 2: wavelength between 0.841-0.876 μ m) and short wave infrared (band 6: wavelength between 1.628-1.652 μ m) are described by Eq. (1). The 8-day NDII values were averaged over each sub-basin to allow comparison to the 8-day average S_u (root zone storage reservoir) values extracted from the FLEX^L model results at each of the 8 runoff stations.

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149 **3. Methods**

150 **3.1 Estimating vegetation water content using near infrared and short wave infrared**

Estimates of vegetation water content (the amount of water in stems and leaves) are of interest to assess the vegetation water status in agriculture and forestry and have been used for drought assessment (Cheng et al., 2006; Gao, 1996; Gao and Goetz, 1995; Ustin et al., 2004; Peñuelas et al., 1993). Evidence of physically-based radiative transfer models and laboratory studies have proved that changes in water content in plant tissues have a large effect on the leaf reflectance in several regions of the 0.7-2.5 μ m spectrum (Fensholt and Sandholt, 2003). Tucker (1980) suggested that the spectral interval between 1.55 and 1.75 μ m (SWIR) is the most suitable region 158 for remotely sensed leaf water content. It is well known that these wavelengths are negatively 159 related to leaf water content due to a large absorption by leaf water (Tucker, 1980; Ceccato et al., 160 2002). However, variations in leaf internal structure and leaf dry matter content also influence the 161 SWIR reflectance. Therefore, only SWIR reflectance values are not suitable for retrieving 162 vegetation water content. To improve the accuracy in retrieving the vegetation water content, a 163 combination of SWIR and NIR (0.7 to 0.9 µm) reflectance information was utilized because NIR 164 is only affected by leaf internal structure and leaf dry matter content but not by water content. A 165 combination of SWIR and NIR reflectance information can remove the effect of leaf internal 166 structure and leaf dry matter content and can improve the accuracy in retrieving the vegetation 167 water content (Ceccato et al., 2001; Yilmaz et al., 2008; Fensholt and Sandholt, 2003).

168 On the basis of this idea, Fensholt and Sandholt (2003) derived NDII:

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$$NDII = \frac{\rho_{0.85} - \rho_{1.65}}{\rho_{0.85} + \rho_{1.65}}$$
 (1)

170 where $\rho_{0.85}$ and $\rho_{1.65}$ are the reflectances at 0.85 µm and 1.65 µm wavelengths, respectively. NDII 171 is a normalized index and the values theoretically vary between -1 and 1. A low NDII value and 172 especially below zero means that reflectance from $\rho_{0.85}$ is lower than the reflectance from $\rho_{1.65}$ 173 which indicates canopy water stress.

174 **3.2 FLEX^L Model**

FLEX^L (Fig. 2) is a lumped conceptual hydrological model which has an HBV-like model 175 176 structure developed in a flexible modelling framework (Fenicia et al., 2011; Gao et al., 2014a; Gao 177 et al., 2014b). The model structure comprises four conceptual reservoirs: the interception reservoir S_i (mm), the unsaturated reservoir representing the moisture storage in the root zone S_u (mm), the 178 179 fast response reservoir S_f (mm), and the slow response reservoir S_s (mm). It also includes two lag 180 functions representing the lag time from storm to peak flow (T_{lagF}) , and the lag time of recharge 181 from the root zone to the groundwater (T_{lagS}) . Besides a water balance equation, each reservoir has 182 process equations that connect the fluxes entering or leaving the storage compartment to the 183 storage in the reservoirs (so-called constitutive functions). Table 1 shows 15 mathematical expressions used for modelling the FLEX^L. A total of 11 model parameters with their distribution 184 185 values are shown in Table 2 and they have to be identified by model calibration. Forcing data include the elevation-corrected daily average rainfall (Gao et al., 2014a), daily average, minimum 186 187 and maximum air temperature, and potential evaporation derived by Hargreaves equation 188 (Hargreaves and Samani, 1985).

189 **3.2.1 Interception reservoir**

190 The interception evaporation E_i (mm d⁻¹) is calculated by potential evaporation E_0 (mm d⁻¹) and 191 the storage of the interception reservoir S_i (mm) (Eq. (3)). There is no effective rainfall P_e (mm d⁻¹) 192 as long as the S_i is less than its storage capacity $S_{i,max}$ (mm) (Eq. (4)) (de Groen and Savenije, 193 2006).

194 **3.2.2 Unsaturated root zone reservoir**

The unsaturated root zone reservoir partitions effective rainfall into infiltration, and runoff R (mm 195 196 d^{-1}), and determines the transpiration by vegetation. Therefore, it is the core of the FLEX^L model. 197 In this study, we applied the widely used beta function (Eq. (6)) of the Xinanjiang model (Zhao, 198 1992; Liang et al., 1992), developed based on the variable contribution area theory (Hewlett and 199 Hibbert, 1967; Beven, 1979), but which can equally reflect the spatial probability distribution of 200 runoff thresholds. The beta function defines the runoff percentage C_r (-) for each time step as a 201 function of the relative soil moisture content $(S_u/S_{u,max})$. In Eq. (6), $S_{u,max}$ (mm) is the root zone 202 storage capacity, and β (-) is the shape parameter describing the spatial distribution of the root 203 zone storage capacity over the catchment. In Eq. (7), the relative soil moisture and potential evaporation are used to determine the transpiration E_t (mm d⁻¹); C_e (-) indicates the fraction of 204 205 $S_{u,max}$ above which the transpiration is no longer limited by soil moisture stress ($E_t = E_0 - E_i$).

206 **3.2.3 Response routine**

In Eq. (8), R_f (mm d⁻¹) indicates the flow into the fast response routine; D (-) is a splitter to separate recharge from preferential flow. In Eq. (9), R_s (mm d⁻¹) indicates the flow into the groundwater reservoir. Equation (10) and (11) are used to describe the lag time between storm and peak flow. $R_f(t-i+1)$ is the generated fast runoff from the unsaturated zone at time t-i+1; T_{lag} is a parameter which represents the time lag between storm and fast runoff generation; c(i) is the weight of the flow in *i*-1 days before; and $R_{\text{fl}}(t)$ is the discharge into the fast response reservoir after convolution.

The linear response reservoirs, representing linear relationships between storages and releases, are applied to conceptualize the discharge from the surface runoff reservoir, fast response reservoir and slow response reservoir. In Eq. (12), $Q_{\rm ff}$ (mm d⁻¹) is the surface runoff, with timescale $K_{\rm ff}$ (d), activated when the storage of fast response reservoir exceeds the threshold $S_{\rm f,max}$ (mm). In Eq. (14) and (16), $Q_{\rm f}$ (mm d⁻¹) and $Q_{\rm s}$ (mm d⁻¹) represent the fast and slow runoff; $K_{\rm f}$ (d) and $K_{\rm s}$ (d) are the time scales of the fast and slow runoff, respectively. $Q_{\rm m}$ (mm d⁻¹) is the total amount of runoff simulated from the three individual components, including $Q_{\rm ff}$, $Q_{\rm f}$, and $Q_{\rm s}$.

221 **3.2.4 Model calibration**

A multi-objective calibration strategy has been adopted in this study to allow for the model to effectively reproduce different aspects of the hydrological response, i.e. high flow, low flow and the flow duration curve. The model was therefore calibrated to three Kling-Gupta efficiencies (Gupta et al., 2009): 1) the K-G efficiency of flows (I_{KGE}) measures the performance of hydrograph reproduction especially for high flows; 2) the K-G efficiency of the logarithm of flows emphasizes low flows (I_{KGL}), and 3) the K-G efficiency of the flow duration curve (I_{KGF}) to represent the flow statistics.

229 The MOSCEM-UA (Multi-Objective Shuffled Complex Evolution Metropolis-University of 230 Arizona) algorithm (Vrugt et al., 2003) was used as the calibration algorithm to find the Paretooptimal solutions defined by the mentioned three objective functions. This algorithm requires 3 231 232 parameters including the maximum number of iterations, the number of complexes, and the 233 number of random samples that is used to initialize each complex. To ensure fair comparison, the 234 parameters of MOSCEM-UA were set based on the number of model parameters. Therefore, the 235 number of complexes is equal to the number of free parameters n; the number of random samples 236 is equal to n^*n^*10 ; and the number of iterations was set to 30000. The model is a widely validated model, which is only used here to derive the magnitude of the root zone moisture storage. 237 238 Therefore validation is not considered necessary, since the model is merely meant to compare 239 calibrated values of S_u with NDII.

240 **3.3 Deseasonalization**

Seasonal signals exist both in NDII and S_u time series. This can lead to spurious correlation. Therefore we deseasonalized both signals to eliminate this strong signal (Schaefli & Gupta, 2007) and subsequently compare the deviations from the seasonal signals of both NDII and S_u . Firstly, the NDII and S_u were normalized between 0 and 1. Then seasonal patterns of NDII and S_u were determined as the average seasonal signals, after which these were subtracted from the normalised data.

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248 **4. Results**

249 **4.1 Spatial and seasonal variation of NDII values for the UPRB and its 14 sub-basins**

250 To demonstrate the spatial and seasonal behaviour of the NDII over the UPRB, the 8-day NDII 251 values were aggregated to monthly values for 2001 to 2013. Figure 3 shows examples of monthly 252 average NDII values for the UPRB in 2004, which is the year with the lowest annual average NDII 253 value. The figure shows that NDII values are higher during the wet season (May to October) and 254 lower during the dry season (from November to April). The lower amounts of rainfall between 255 November and April cause a continuous reduction of NDII values. On the other hand, higher 256 amounts of rainfall between May and October result in increasing NDII values. However, NDII 257 values appear to vary little between July and October.

258 The average NDII values during the wet season, the dry season, and the whole year within the 13 259 years are presented in Table 4. The table also shows the order of the NDII values from the highest 260 (number 1) to the lowest (number 13). It can be seen that the annual average NDII value for the 261 whole basin is approximately 0.165, while the average values during the wet and dry season are 262 about 0.211 and 0.118, respectively. The highest mean annual value (NDII = 0.177) occurred in 2002-2003 and the lowest (NDII = 0.149) in 2004-2005. The highest (NDII = 0.149) and lowest 263 264 (NDII = 0.088) dry season values were reported in 2002-2003 and 2004-2005, respectively. On the 265 other hand, the highest (NDII = 0.224) and lowest (NDII = 0.197) wet season values were 266 observed in 2006-2007 and 2010-2011, respectively. It can be concluded that a dry season with 267 relatively low moisture content and a wet season with high moisture content as specified by NDII 268 values do not normally occur in the same year.

269 The 8-day NDII values were also computed for each of the 14 tributaries within the UPRB from 270 2001 to 2013. Table 5 shows the monthly averaged NDII values between 2001 and 2013 and the 271 ranking order for each of the 14 tributaries. The results suggest that Nam Mae Taeng, Nam Mae 272 Rim, and Upper Mae Chaem sub-basins, which have higher mean annual NDII values, have a 273 higher moisture content than other sub-basins; while Nam Mae Haad, Nam Mae Li, and Ping 274 River Section 2 are 3 sub-basins, with lower mean annual NDII values, have lower moisture 275 content than other sub-basins. Monthly average NDII values for these 6 sub-basins are presented 276 in Fig. 4. It can be seen that during the dry season, NDII values of the 3 sub-basins with the lowest 277 values are a lot lower than those of the 3 sub-basins with the highest NDII values. However, NDII 278 values for these 2 groups are not significantly different during the wet season. The figure also 279 reveals that NDII values tend to continuously increase from relatively low values in March to 280 higher values in June. The values slightly fluctuate during the wet season before sharply falling 281 once again when the rainy season ends, and reach their minimum values in February.

282 **4.2 FLEX^L Model results**

Calibration of FLEX^L was done on the 8 sub-catchments which have runoff stations. The results 283 are summarized in Table 6. The performance of the model was quite good as demonstrated in 284 285 Table 7. In Fig. 5, the duration curves of runoff stations P.20 and P.21 are presented as examples of model performance. Table 7 shows the average Kling-Gupta efficiencies values for I_{KGE} , I_{KGL} 286 and I_{KGF} , which indicate the performance of high flows, low flows, and flow duration curve for the 287 8 runoff stations. The results for the flow duration curve appear to be better than those of the high 288 289 flows and especially the low flows. However, the overall results are acceptable and can be used for 290 further analysis in this study.

4.3 Relation between NDII and root zone moisture storage (*S*_u)

292 The 8-day NDII values were compared to the 8 day average root zone moisture storage values of the FLEX^L model. It appears that during moisture stress periods, the relationship can be well 293 294 described by an exponential function, for each of the 8 sub-catchments. Table 8 presents the coefficients of the exponential relationships as well as the coefficients of determination (\mathbf{R}^2) for 295 296 annual, wet season, and dry season values for each sub-catchment. The corresponding scatter plots 297 are shown in Fig. 6. It can be clearly seen that the correlation is much better in the dry season than 298 in the wet season. During the wet season, there may also be short period of moisture stress, where 299 the exponential pattern can be recognized, but no clear relation is found when the vegetation does 300 not experience any moisture stress.

301 Examples of scaled time series of NDII and root zone storage (S_u) values for the sub-catchments P.20 and P.21 are presented in Figure 7, respectively. The scaled time series of the NDII and S_u 302 values were calculated by dividing their value by the differences between their maximum and 303 minimum values: NDII/(NDII_{max}-NDII_{min}) and S_u /($S_{u,max}$ - S_{umin}), respectively, while the maximum 304 305 and the minimum are the values within the overall considered time series. Figure 7 shows that the 306 scaled NDII and S_u values are highly correlated during the dry season, but less so during the wet 307 season. These results confirm the potential of NDII to effectively reflect the vegetation water 308 content, which, through the suction pressure exercised by the moisture deficit, relates to the 309 moisture content in the root zone. During dry periods, or during dry spells in the rainy season, as 310 soon as the leaves of the vegetation experience suction pressure, we see high values of the 311 coefficient of determination.

If the soil moisture in the root zone is above a certain threshold value, then the leaves are not under stress. In the UPRB this situation occurs typically during the middle and late rainy season. The NDII then does not vary significantly while the root zone moisture storage may still vary, albeit above the threshold where moisture stress occurs. This causes a lower correlation between 316 NDII and root zone storage during wet periods. Interestingly, even during the wet season dry 317 spells can occur. We can see in Fig. 6, that during such a dry spell, the NDII and Su again follow 318 the exponential relationship.

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We can see that the S_u , derived merely from precipitation and energy, is strongly correlated to the vegetation water observed by NDII during condition of moisture stress, without time lag (Figure 6, S1, S2). Introduction of a time lag resulted in reduction of the correlation coefficients (Supplementary material). This confirms the direct response of vegetation to soil moisture stress, which confirms that the NDII can be used as a proxy for root zone moisture storage.

325 The deseasonalized results of dry periods in sub-catchments P.20 and P.21 are shown in Figure 7. We found these variations of deseasonalized NDII and S_u to be similar in these two sub-326 catchments, with the coefficients of determination (R^2) as 0.32 and 0.18 respectively in P.20 and 327 328 P.21. More important than the coefficient of determination is the similarity between the deseasonalized patterns. For P.20, the year 2001 is almost identical, whereas the years 2004 and 329 330 2006 are dissimilar. In general the patterns are well reproduced, especially if we take into account 331 the implicit uncertainties of the lumped hydrological model and the data of precipitation and 332 potential evaporation used in the model. The results of other sub-basins can be found in the 333 supplementary materials.

334

335 **5. Discussion**

5.1 Is vegetation a trouble-maker or a good indicator for moisture storage in the root zone?

337 In bare soil, remote sensors can only detect soil moisture until a few centimeters below the surface 338 (~5cm) (Entekhabi et al., 2010). Unfortunately, for hydrological modelling, the moisture state of 339 the bare surface is of only limited interest. What is of key interest for understanding the dynamics 340 of hydrological systems is the variability of the moisture storage in the unsaturated zone. This 341 variability determines the rainfall-runoff behaviour, the transpiration of vegetation, and the 342 partitioning between different hydrological fluxes. This dynamic part of the unsaturated zone is 343 the root zone of ecosystems. However, observing the soil moisture content in the root zone is still 344 a major challenge (Entekhabi et al., 2010).

What is normally done, is to link the moisture content of the surface layer to the total amount of moisture in the root zone. Knowing the surface soil moisture, the root zone soil moisture can be 347 estimated by an exponential decay filter (Albergel et al., 2008; Ford et al., 2014) or by models 348 (Reichle, 2008) However, the surface soil moisture is only weakly related with root zone soil 349 moisture (Mahmood and Hubbard, 2007); it only works if there is connectivity between the 350 surface and deeper layers and when a certain state of equilibrium has been reached (when the short 351 term dynamics after a rainfall event has leveled out). It is also observed that the presence of 352 vegetation prevents the observation of soil moisture and further deteriorates the results (Jackson 353 and Schmugge, 1991). Avoiding the influence of vegetation in observing soil moisture (e.g. by 354 SMOS or SMAP) is seen as a challenge by some in the remote sensing community (Kerr et al., 355 2001; Entekhabi et al., 2010). Several algorithms have been proposed to filter out the vegetation 356 impact (Jackson and Schmugge, 1991). But is vegetation a trouble-maker, or does it offer an 357 excellent opportunity to directly gauge the state of the soil moisture?

358 In this study, we found that vegetation rather than a problem could become key to sensing the 359 storage of moisture in the root zone. The water content in the leaves is connected to the suction 360 pressure in the root zone (Rutter and Sands, 1958). If the suction pressure is above a certain 361 threshold, then this connection is direct and very sensitive. We found a highly significant 362 correlation between NDII and S_u , particularly during periods of moisture stress. During dry 363 periods, or during dry spells in the rainy season, as soon as the leaves of the vegetation experience 364 suction pressure, we see high values of the coefficient of determination. Observing the moisture 365 content of vegetation provides us with directly information on the soil moisture state in the root 366 zone. We also found that there is almost no lag time between S_u and NDII. This illustrates the fast 367 response of vegetation to soil moisture variation, which makes the NDII a sensitive and direct 368 indicator for root zone moisture storage.

369 **5.2 Implication in hydrological modelling**

370 Simulation of root zone soil moisture is crucial in hydrological modelling (Houser et al., 1998; 371 Western and Blöschl, 1999). Using estimates of soil moisture states could increase model 372 performance and realism, but moreover, it would be powerful information to facilitate prediction 373 in ungauged basins (Hrachowitz et al., 2013). However, until now, it has not been practical (e.g. 374 Parajka et al., 2006; Entekhabi et al., 2010). Assimilating soil moisture in hydrological models, 375 either from top-soil observation by remote sensing, or from the deeper soil column by models 376 (Reichle, 2008), is still a challenge. Several studies showed how difficult it is to assimilate soil 377 moisture data to improve daily runoff simulation (Parajka et al., 2006; Matgen et al., 2012).

There are several reasons why we have not compared our results with soil moisture observations in the field. Firstly, observations of soil moisture are not widely available. Moreover, it is not 380 straightforward to link classical soil moisture observations to the actual moisture available in the 381 root zone. Most observations are conducted at fixed depths and at certain locations within a highly 382 heterogeneous environment. Without knowing the details of the root distribution both in the 383 horizontal and vertical plain, it is hard, if not impossible, to estimate root zone soil moisture. We 384 should realize that it is difficult to observe root zone soil moisture even at a local scale. But 385 measuring root zone soil moisture at a catchment scale is even more challenging. State-of-the-art 386 remote sensing techniques can observe spatially distributed soil moisture, but what they can see is 387 only the top layer if not blocked by vegetation. The top layer moisture may be correlated with the 388 root zone storage, but it is definitely not the same.

By observing the moisture content of the leaves, the NDII represents the soil moisture storage condition of the entire root zone, which is precisely the information that hydrological models require. This study clearly shows the strong temporal correlation between S_u and NDII. From the relationship between NDII and S_u , we can directly derive a proxy for the soil moisture state, which can potentially be assimilated in hydrological models. This method would be extremely useful for prediction of discharge in ungauged basins.

We should, of course, be aware of regional limitations. This study considered a tropical seasonal evergreen ecosystem, where periods of moisture stress regularly occur. In ecosystems which shed their leaves, or go dormant, other conditions may apply. We need further investigations into the usefulness of this approach in catchments with different climates. In addition, the phenology of the ecosystem is of importance, which should be taken into consideration in follow-up research.

400

401 **6. Conclusions**

The NDII was used to investigate drought for the UPRB from 2001 to 2013. Monthly average NDII values appear to be spatially distributed over the UPRB, in agreement with seasonal variability and landscape characteristics. NDII values appear to be lower during the dry season and higher during the wet season as a result of seasonal differences between precipitation and evaporation. The NDII appears to correlate well with the moisture storage in the root zone, offering an interesting proxy variable for calibration of hydrological models in ungauged basins.

To illustrate the importance of NDII as a proxy for root zone moisture storage in hydrological models, we applied the $FLEX^{L}$ model to assess the root zone soil moisture content (S_u) of 8 subcatchments of the UPRB controlled by 8 runoff stations. The results show that the 8-day average NDII values over the study sub-basin correlate well with the 8-day average S_u for all subcatchments during dry periods (average R² equals 0.87), and less so during wet spells (average R² 413 equals 0.61). The NDII appears to be a good proxy for root zone moisture content during dry
414 spells when leaves are under moisture stress. The natural interaction between rainfall, soil
415 moisture, and leave water content can be visualised by the NDII, making it an important indicator
416 both for hydrological modelling and drought assessment.

417

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Reservoirs	Water balance equations	Equation	Constitutive equations	Equation
- · · ·	dS		$E_{i} = \begin{cases} E_{0}; S_{i} > 0\\ 0; S_{i} = 0 \end{cases}$	(3)
Interception	$\frac{\mathrm{d}S_{\mathrm{i}}}{\mathrm{d}t} = P - E_{\mathrm{i}} - P_{\mathrm{e}}$	(2)	$P_{e} = \begin{cases} 0; S_{i} < S_{i,max} \\ P; S_{i} = S_{i,max} \end{cases}$	(4)
Unsaturated reservoir	$\frac{\mathrm{d}S_{\mathrm{u}}}{\mathrm{d}t} = P_{\mathrm{e}} - R - E_{\mathrm{t}}$	(5)	$\frac{R}{P_{\rm e}} = 1 - (1 - \frac{S_{\rm u}}{(1 + \beta)S_{\rm u,max}})^{\beta}$	(6)
	$\frac{dt}{dt} = T_e = K - E_t$	(3)	$E_t = (E_0 - E_i) \cdot \min(1, \frac{S_u}{C_e S_{u,max}(1 + \beta)})$	(7)
			$R_{\rm f} = R \cdot D$	(8)
			$R_{\rm s} = R \cdot (1 - D)$	(9)
Splitter and Lag function			$R_{\rm fl}(t) = \sum_{i=1}^{T_{\rm lag}} c(i) \cdot R_{\rm f}(t-i+1)$	(10)
			$c(i) = i / \sum_{u=1}^{T_{\text{tag}}} u$	(11)
Fast records:-	$dS_{f} = R Q Q$	(12)	$Q_{\rm ff} = \max(0, S_{\rm f} - S_{\rm f,max}) / K_{\rm ff}$	(13)
Fast reservoir	$\frac{\mathrm{d}S_{\mathrm{f}}}{\mathrm{d}t} = R_{\mathrm{fl}} - Q_{\mathrm{ff}} - Q_{\mathrm{f}}$	(12)	$Q_{ m f}=S_{ m f}$ / $K_{ m f}$	(14)
Slow reservoir	$\frac{\mathrm{d}S_{\mathrm{s}}}{\mathrm{d}t} = R_{\mathrm{s}} - Q_{\mathrm{s}}$	(15)	$Q_{\rm s} = S_{\rm s} / K_{\rm s}$	(16)

572 Table 1. Water balance and constitutive equations used in FLEX^L.

Parameter	Range	Parameters	Range
$S_{i,max}$ (mm)	(0.1, 6)	$K_{ff}(d)$	(1, 9)
$S_{u,max}$ (mm)	(10, 1000)	$T_{lagF}(d)$	(0, 5)
β(-)	(0, 2)	T_{lagS} (d)	(0, 5)
<i>C</i> _e (-)	(0.1, 0.9)	$K_f(d)$	(1, 40)
D (-)	(0, 1)	$K_{s}\left(\mathrm{d} ight)$	(10, 500)
$S_{f,max}$ (mm)	(10, 200)		

574 Table 2. Parameter range of the FLEX^L Model.

Sub-basin	Mae Taeng at Ban Mae Taeng (P.4A)	Nam Mae Chaem at Kaeng Ob Luang (P.14)	Ping River at Chiang Dao (P.20)	Nam Mae Rim at Ban Rim Tai (P.21)	Nam Mae Klang at Pracha Uthit Bridge (P.24A)	Nam Mae Khan at Ban Klang (P.71)	Nam Mae Li at Ban Mae E Hai (P.76)	Nam Mae Tha at Bar Sop Mae Sapuad (P.77)
Area (km ²)	1902	3853	1355	515	460	1771	1541	547
Altitude range (m)	1020	991	790	731	888	828	618	641
Average channel slope (%)	0.78	0.81	0.80	0.72	0.98	0.69	0.41	0.63
Average forest and agricultural areas (%)	81.9, 16.5	91.8, 7.4	80.9, 12.8	86.1, 11.6	79.7, 14.2	86.1, 10.1	69.7, 20.1	80.4, 12.7
Average rainfall depth	953 (88%)	883 (92%)	1076 (88%)	1019 (90%)	860 (88%)	1090 (89%)	1092 (91%)	757 (88%)
(wet season/ dry season) (mm)	130 (12%)	75 (8%)	150 (12%)	115 (10%)	121(12%)	132 (11%)	106 (9%)	88 (10%)
Number of years data is coincident with NDII	11	7	12	11	12	9	12	12
Data period	1995-2011	1995-2007	1995-2012	1995-2011	1995-2012	1996-2009	1996-2012	1996-2012

576 Table 3. Catchment characteristics and data period for selected 8 sub-basins in the UPRB.

Table 4. Average NDII values during the wet season, the dry season, and the whole year from
2001 to 2013, and their order of moisture content (Range from 1 to 13. Less value indicates less
NDII) for the entire Upper Ping River Basin.

	Wet season	Dry season	
Year	(May-October)	(November-April)	Annual
2001-2002	0.223 (2)	0.119 (7)	0.171 (4)
2002-2003	0.205 (9)	0.149 (1)	0.177 (1)
2003-2004	0.218 (5)	0.091 (12)	0.155 (12)
2004-2005	0.210 (8)	0.088 (13)	0.149 (13)
2005-2006	0.200 (11)	0.128 (3)	0.164 (7)
2006-2007	0.224 (1)	0.111 (10)	0.168 (5)
2007-2008	0.222 (3)	0.130 (2)	0.176 (2)
2008-2009	0.221 (4)	0.123 (5)	0.172 (3)
2009-2010	0.213 (7)	0.101 (11)	0.157 (11)
2010-2011	0.197 (13)	0.128 (4)	0.163 (8)
2011-2012	0.216 (6)	0.116 (9)	0.166 (6)
2012-2013	0.201 (10)	0.118 (8)	0.159 (10)
2013-2014	0.199 (12)	0.123 (6)	0.161 (9)
Average	0.211	0.118	0.165
Maximum	0.224	0.149	0.177
Minimum	0.197	0.088	0.149

Sub-basin	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Average
Ping River Section 1	0.14 (7.5)	0.06 (7.4)	0.02 (8.8)	0.07 (8.9)	0.17 (8.4)	0.21 (6.2)	0.22 (4.5)	0.22 (6.1)	0.24 (7.5)	0.23 (8.3)	0.22 (7.8)	0.18 (7.2)	0.16 (8)
Nam Mae Ngad	0.17 (5.2)	0.11 (5.9)	0.07 (6.2)	0.10 (6.3)	0.18 (6.9)	0.21 (7.1)	0.21 (7.5)	0.22 (8.0)	0.23 (9.2)	0.23 (7.9)	0.23 (6.4)	0.20 (5.7)	0.18 (6)
Nam Mae Taeng	0.21 (1.3)	0.16 (1.0)	0.13 (1.2)	0.14 (2.1)	0.19 (3.9)	0.21 (6.1)	0.22 (6.0)	0.23 (4.5)	0.25 (3.1)	0.25 (2.6)	0.26 (1.2)	0.24 (1.7)	0.21 (1)
Ping River Section 2	0.07 (11.5)	0.02 (9.8)	0.01 (9.2)	0.04 (11.6)	0.13 (13.1)	0.18 (13.0)	0.18 (13.5)	0.19 (13.3)	0.21 (13.6)	0.21 (12.7)	0.17 (13.4)	0.12 (13.5)	0.13 (12)
Nam Mae Rim	0.17 (5.3)	0.13 (4.3)	0.10 (3.9)	0.13 (3.3)	0.20 (2.6)	0.22 (3.7)	0.22 (4.0)	0.24 (2.5)	0.26 (1.3)	0.26 (1.2)	0.24 (3.7)	0.20 (5.6)	0.20 (2)
Nam Mae Kuang	0.09 (9.4)	0.03 (9.5)	0.02 (9.3)	0.05 (10.1)	0.15 (10.0)	0.20 (8.1)	0.21 (8.1)	0.22 (8.2)	0.24 (7.0)	0.23 (7.5)	0.20 (10.4)	0.14 (10.7)	0.15 (9)
Nam Mae Ngan	0.18 (4.0)	0.13 (4.4)	0.10 (4.9)	0.13 (4.1)	0.19 (3.9)	0.21 (5.3)	0.22 (5.5)	0.23 (5.2)	0.25 (3.9)	0.24 (4.5)	0.24 (4.5)	0.22 (4.0)	0.19 (5)
Nam Mae Li	0.05 (12.5)	-0.04 (12.5)	-0.04 (12.7)	0.02 (12.1)	0.14 (11.9)	0.19 (11.8)	0.20 (9.7)	0.23 (8.3)	0.23 (9.9)	0.21 (13.0)	0.18 (13.2)	0.13 (12.5)	0.12 (13)
Nam Mae Klang	0.19 (3.3)	0.13 (3.5)	0.12 (2.8)	0.14 (2.3)	0.20 (2.9)	0.22 (4.8)	0.22 (7.2)	0.23 (7.6)	0.23 (8.6)	0.24 (7.2)	0.24 (4.5)	0.22 (3.3)	0.20 (4)
Ping River Section 3	0.06 (11.7)	-0.03 (12.5)	-0.04 (12.3)	0.03 (11.2)	0.15 (9.3)	0.21 (7.2)	0.21 (8.7)	0.21 (9.9)	0.22 (11.4)	0.21 (11.9)	0.19 (11.2)	0.15 (10.3)	0.13 (11)
Upper Nam Mae Chaem	0.20 (1.9)	0.15 (2.0)	0.12 (2.3)	0.13 (4.2)	0.18 (6.7)	0.20 (9.5)	0.21 (9.2)	0.21 (9.1)	0.24 (6.2)	0.25 (3.9)	0.26 (2.1)	0.24 (1.6)	0.20 (3)
Lower Nam Mae Chaem	0.09 (9.8)	0.006 (10.7)	-0.007 (10.8)	0.05 (10.2)	0.15 (10.2)	0.20 (10.2)	0.20 (9.9)	0.21 (8.9)	0.23 (9.5)	0.23 (8.3)	0.21 (8.9)	0.16 (9.2)	0.14 (10)
Nam Mae Haad	0.03 (14.0)	-0.07 (14.0)	-0.06 (13.8)	0.003 (12.9)	0.15 (10.0)	0.21 (5.8)	0.22 (6.4)	0.23 (6.2)	0.24 (5.2)	0.22 (9.7)	0.19 (11.2)	0.12 (12.4)	0.12 (14)
Nam Mae Tuen	0.13 (7.6)	0.05 (7.7)	0.05 (7.0)	0.10 (5.9)	0.19 (5.2)	0.21 (6.2)	0.22 (4.9)	0.222 (7.2)	0.23 (8.7)	0.24 (6.2)	0.23 (6.5)	0.20 (6.5)	0.17 (7)
Average	0.13	0.06	0.04	0.08	0.17	0.20	0.21	0.22	0.24	0.23	0.22	0.18	0.16
Maximum	0.21	0.16	0.13	0.14	0.20	0.22	0.22	0.24	0.26	0.26	0.26	0.24	0.21
Minimum	0.03	-0.07	-0.06	0.003	0.13	0.18	0.18	0.19	0.21	0.21	0.17	0.12	0.12

582 Table 5. Monthly average NDII values between 2001 and 2013 and the order of basin moisture content for each of 14 sub-basins within the UPRB.

Runoff station	S _{i,max}	S _{u,max}	Ce	Beta	D	K_{f}	Ks	T_{lagF}	T_{lagS}	S _{f,max}	K_{ff}
Kunon station	(mm)	(mm)	(-)	(-)	(-)	(days)	(days)	(days)	(days)	(mm)	(days)
P.4A	2.0	463	0.30	0.66	0.77	2.9	42	1.1	49	93	9.1
P.14	2.3	269	0.55	1.16	0.65	4.0	63	1.5	39	155	7.6
P.21	2.3	388	0.31	0.90	0.64	2.1	66	2.4	48	33	2.5
P.20	2.0	324	0.47	0.50	0.79	7.7	103	1.0	25	69	1.7
P.24A	3.2	209	0.77	1.53	0.89	3.2	267	1.5	44	24	4.2
P.76	2.3	486	0.62	0.32	0.89	2.4	191	2.7	3	130	7.4
P.77	4.5	344	0.48	0.27	0.75	1.5	65	1.2	30	164	5.6
P.71	4.3	532	0.34	0.46	0.90	3.5	80	1.8	15	179	6.5

585 Table 6. FLEX^L parameters calibrated at 8 runoff stations located in the UPRB.

Station	Data period	I_{KGE}	I_{KGL}	I _{KGF}
P.4A	1995-2009	0.822	0.667	0.963
P.14	1995-2007	0.796	0.442	0.966
P.21	1995-2009	0.814	0.718	0.985
P.20	1995-2011	0.792	0.685	0.964
P.24A	1995-2011	0.623	0.598	0.945
P76	2000-2011	0.539	0.665	0.916
P.77	1999-2011	0.775	0.612	0.970
P.71	1996-2009	0.823	0.714	0.975
Average		0.748	0.638	0.961

587 Table 7. FLEX^L model performance at 8 runoff stations.

Runoff station		Annual	_	W	et Seas	on	Dry Season Relationship			
	Re	lations	hip	Re	lations	hip				
	a	b	R^2	a	b	R^2	a	b	\mathbf{R}^2	
P.4A	11.2	12.4	0.66	11.1	12.9	0.53	12.6	11.2	0.90	
P.14	21.9	9.8	0.81	19.2	10.8	0.71	24.6	8.5	0.92	
P.20	52.3	7.4	0.79	36.2	9.1	0.72	59.7	6.7	0.91	
P.21	30.8	9.0	0.68	27.8	9.3	0.53	30.6	9.22	0.86	
P.24A	22.1	8.5	0.60	24.2	8.3	0.41	22.4	8.1	0.81	
P.71	2.1	19.9	0.77	1.9	20.5	0.65	2.3	19.0	0.87	
P.76	10.1	13.6	0.85	8.1	14.4	0.74	10.8	14.6	0.87	
P.77	35.4	8.0	0.70	20.7	10.2	0.61	40.6	7.7	0.83	
Averag	-	-	0.73	-	-	0.61	-	-	0.87	
e										

Table 8. Exponential relationships between the average NDII values and simulated root zone moisture storage (S_u) in the 8 sub-basins controlled by the 8 runoff stations.

591 Note: $S_u = ae^{bNDII}$

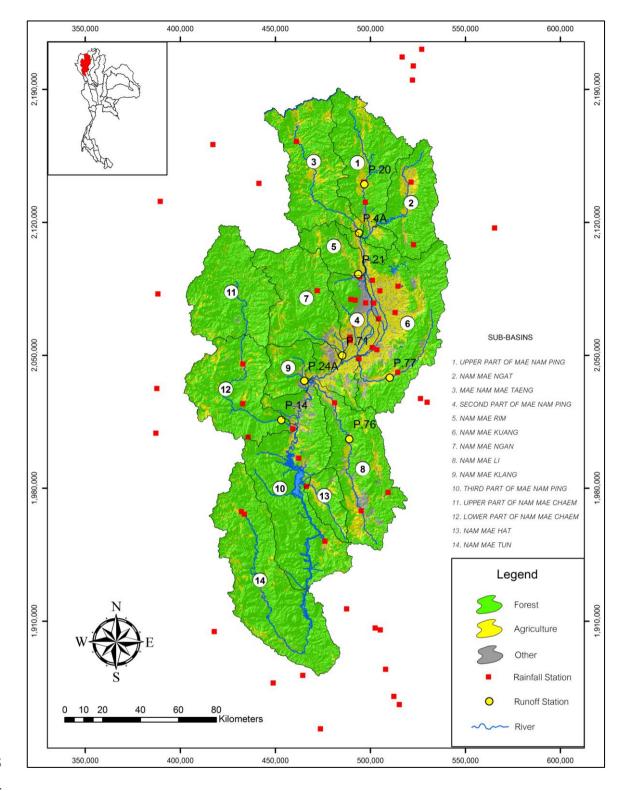


Figure 1. The Upper Ping River Basin (UPRB) and the locations of the rain-gauge and runoffstations. The numbers indicate the 14 sub-basins of the UPRB.

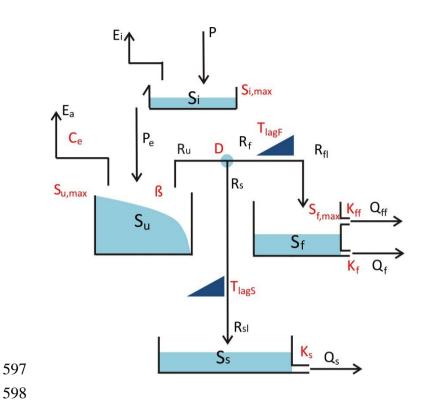


Figure 2. Model structure of the FLEX^L.

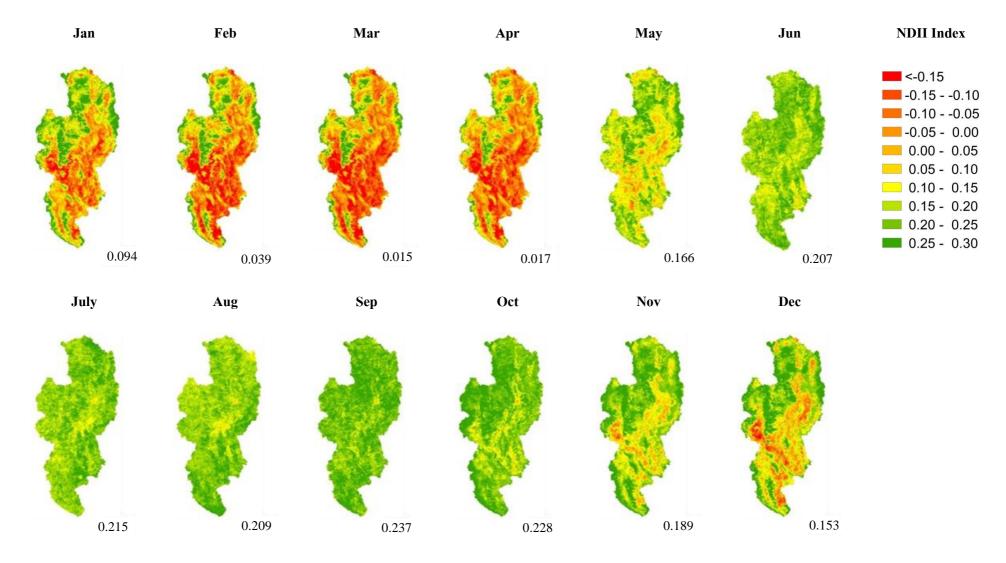


Figure 3. Monthly average NDII values for the UPRB in 2004. The green color indicates an NDII between 0.15 and 0.30, yellow between 0 and 0.15, orange between -0.15 and 0 and red an NDII<-0.15) representing relatively high-, medium-, low-, and very low- root zone moisture content.

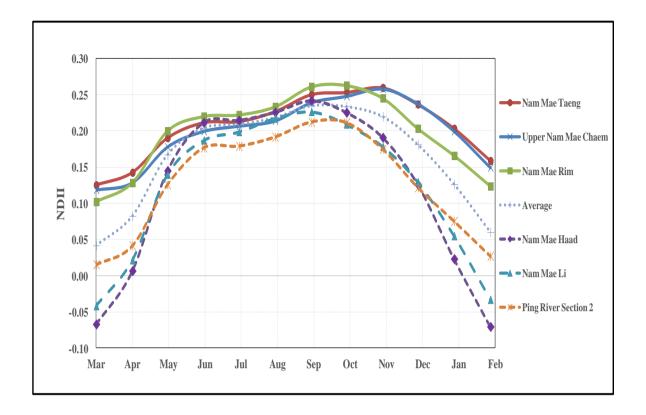


Figure 4. Monthly average NDII values for 6 sub-basins compared to the basin average in the UPRB. Note that three wettest and three driest basins are presented in this graph.

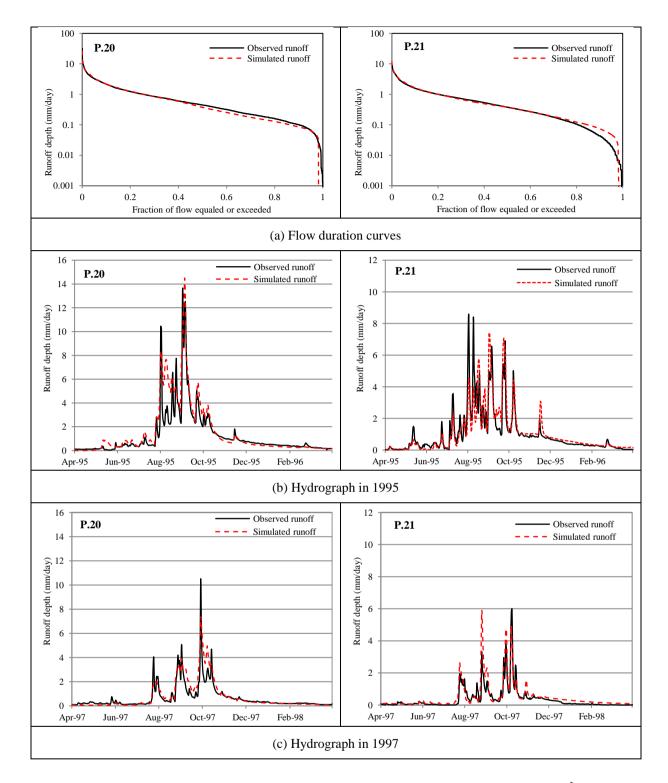
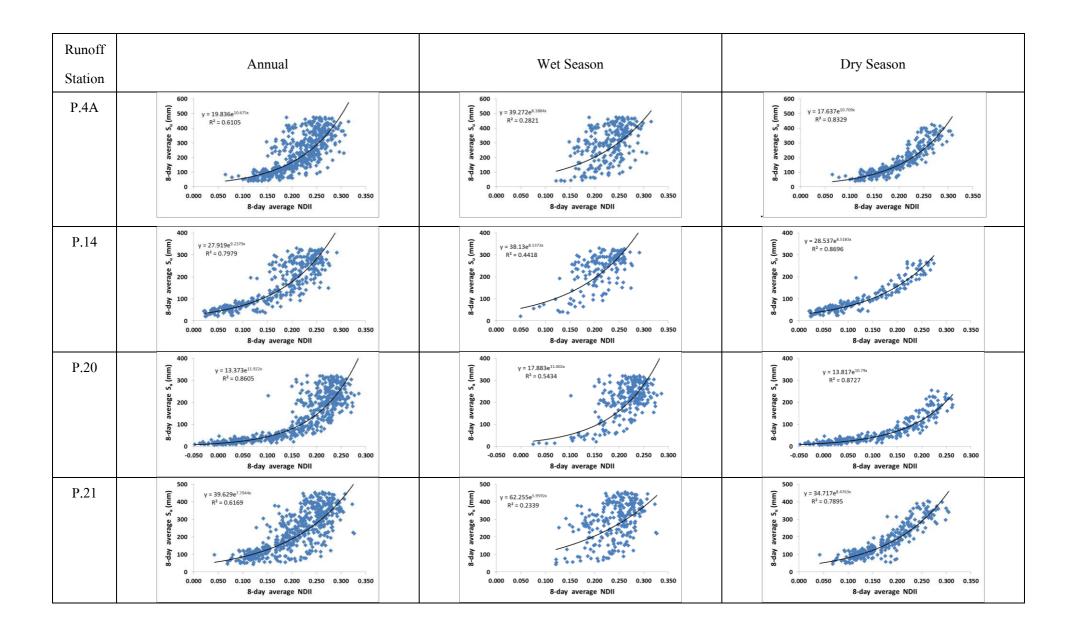


Figure 5. Examples of flow duration curves and simulated hydrographs using FLEX^L at runoff stations P.20 and P.21.



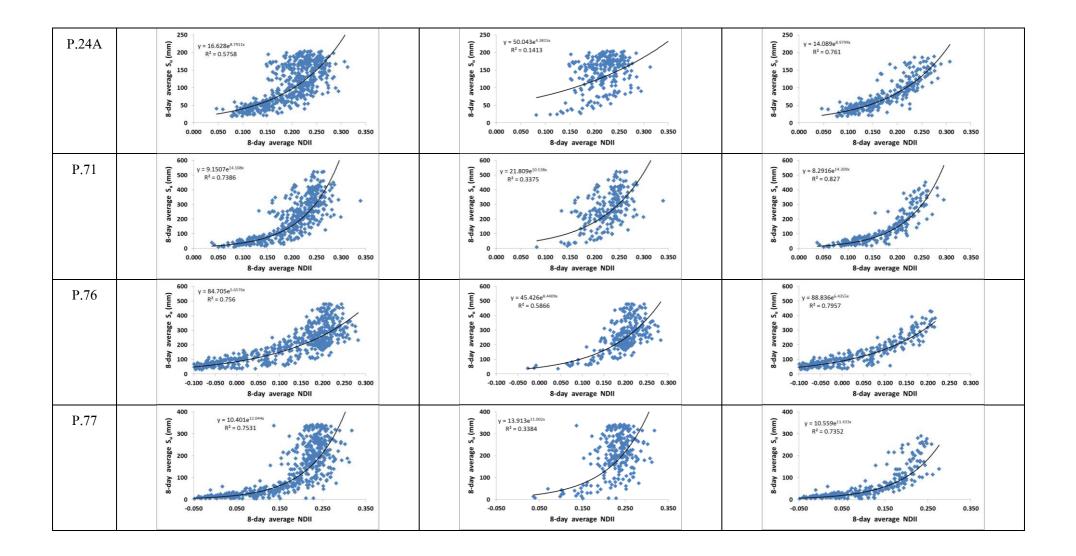


Figure 6. Scatter plots between the average NDII and the average root zone moisture storage (S_u) for 8 sub-basins controlled by runoff stations.

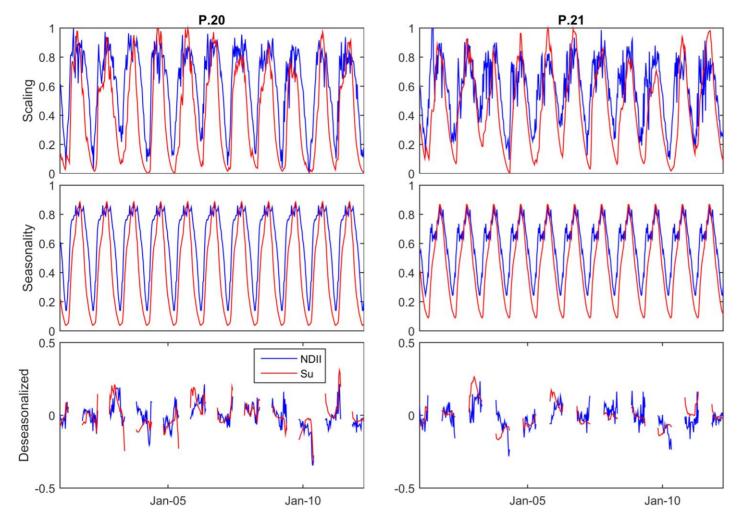


Figure 7. Scaled time series, seasonality and de-seasonalized (dry seasons) time series of the 8-days-averaged NDII values compared to the 8-daysaveraged simulated root zone moisture storage (S_u) in Nam Mae Rim sub-basin at P.20 (Chiang Dao) and P.21 (Ban Rim Tai) runoff stations. The coefficients of determination (R^2) of the de-seasonalized NDII and S_u are 0.32 and 0.18 respectively for P.20 and P.21. For the results of all the 8 sub-basins, please refer to the supplementary material.