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2 **Nonlinear effects of locally heterogeneous hydraulic**

3 **conductivity fields on regional stream-aquifer exchanges**

4 Jinfeng Zhu¹, C Larrabee Winter^{2,*}, Zhongjing Wang^{1,3,*}

5 ¹ Department of Hydraulic Engineering, Tsinghua University, Beijing, 100084, China

6 ² Department of Hydrology and Water Resources and Program in Applied

7 Mathematics, University of Arizona, Tucson, AZ, 85721, USA

8 ³ State Key Laboratory of Hydro-Science and Engineering, Tsinghua University, Beijing,

9 100084, China

10

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12 *Corresponding author information:

13 Email: winter@email.arizona.edu

14 Tele: +1 520 621 7120

15 Email: zj.wang@mail.tsinghua.edu.cn

16 Tele: +86 10 62782021

17

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1 **Abstract**

2 Computational experiments are performed to evaluate the effects of locally
3 heterogeneous conductivity fields on regional exchanges of water between stream and
4 aquifer systems in the Middle Heihe River Basin (MHRB) of northwestern China.
5 The effects are found to be nonlinear in the sense that simulated discharges from
6 aquifers to streams are systematically lower than discharges produced by a base
7 model parameterized with relatively coarse effective conductivity. A similar, but
8 weaker, effect is observed for stream leakage. The study is organized around three
9 hypotheses: (H1) small-scale spatial variations of conductivity significantly affect
10 regional exchanges of water between streams and aquifers in river basins, (H2)
11 aggregating small-scale heterogeneities into regional effective parameters
12 systematically biases estimates of stream-aquifer exchanges, and (H3) the biases
13 result from slow-paths in groundwater flow that emerge due to small-scale
14 heterogeneities. The hypotheses are evaluated by comparing stream-aquifer fluxes
15 produced by the base model to fluxes simulated using realizations of the MHRB
16 characterized by local (grid-scale) heterogeneity. Levels of local heterogeneity are
17 manipulated as control variables by adjusting coefficients of variation. All models are
18 implemented using the MODFLOW simulation environment, and the PEST tool is
19 used to calibrate effective conductivities defined over 16 zones within the MHRB.
20 The effective parameters are also used as expected values to develop log-normally
21 distributed conductivity (K) fields on local grid scales. Stream-aquifer exchanges
22 are simulated with K fields at both scales and then compared. Results show that the
23 effects of small-scale heterogeneities significantly influence exchanges with
24 simulations based on local-scale heterogeneities always producing discharges that are
25 less than those produced by the base model. Although aquifer heterogeneities are
26 uncorrelated at local scales, they appear to induce coherent slow-paths in groundwater
27 fluxes that in turn reduce aquifer-stream exchanges. Since surface water-groundwater
28 exchanges are critical hydrologic processes in basin-scale water budgets, these results
29 also have implications for water resources management.

30

1. Introduction

Exchanges of water between streams and aquifers are critical elements in the coupled dynamics of watersheds. Groundwater discharge to streams maintains natural hydrologic systems through base flow during periods of low stream flow, while leakage from streams is an important source of groundwater recharge (Sophocleous et al., 1995; Hantush, 2005; Newman et al., 2006). The magnitudes of such fluxes vary on scales ranging from sub-regional (or zonal) to local scales. Local variations in the spatial distribution of aquifer characteristics are known to affect groundwater flow and stream-aquifer exchanges on sub-regional scales (Schmidt et al. 2006; Kalbus et al., 2009; Mendoza et al., 2015), but it is not usually possible to measure system parameters and states with sufficient accuracy and level of detail to specify local variations throughout a watershed (Wroblecky et al., 1998; Kalbus et al., 2006). Furthermore, hydrologic system parameters can vary on multiple spatial and temporal scales and are subject to experimental error (Molz, 2000; Genereux et al., 2008).

Computational experiments, on the other hand, allow the effects of heterogeneity to be investigated by consistently varying the values of parameters, such as hydraulic conductivity, as control variables and observing the resulting effects on simulated system states (Winter et al., 2004; Bruen and Osman, 2004; Hantush, 2005). In this study the sensitivity of regional hydrologic systems to locally heterogeneous aquifer hydraulic conductivity is explored by simulating stream-aquifer exchanges in the Middle Heihe River Basin (MHRB) of northwestern China, a typical semi-arid basin. Fluxes in alluvial aquifers of the MHRB are usually represented as two-dimensional processes (Huang, 2012).

In addition to their importance for water resources management, stream-aquifer exchanges are a convenient measure of overall watershed performance because they summarize the states of system fluxes at well-defined locations where they are relatively easy to quantify. Two scales of heterogeneity are represented in the simulations, zonal and local (or grid-scale). Local-scale conductivity is manipulated as the only control variable in the computational experiments. All other system

1 parameters, including streambed conductance, are the same in every simulation.

2 Multi-physics models of the coupled hydrologic fluxes of the MHRB are
3 implemented using the MODFLOW simulation environment ([McDonald and](#)
4 [Harbaugh, 2003](#)). The PEST tool ([Doherty et al., 1994](#)) is used to calibrate an
5 effective aquifer hydraulic conductivity for each of sixteen zones within the MHRB.
6 The zones were defined in previous hydrogeological studies of the MHRB ([Hu et al.,](#)
7 [2007](#)). The effective zonal parameters are also used as expected values to develop
8 realizations of log-normally distributed conductivity fields on local grid scales.
9 Standard deviations are defined by means of coefficients of variation (CV) ranging
10 from 0.5 to 2.0. Ten realizations of a random conductivity field are produced for each
11 CV. Random conductivity values are uncorrelated in space. Monthly stream-aquifer
12 exchanges are calculated for each realization and for the zonally parameterized
13 model.

14 The study is organized in terms of three explicit hypotheses about the effects of
15 local-scale heterogeneity:

16 **H1.** Local-scale heterogeneities significantly affect regional
17 stream-aquifer water exchanges in the MHRB.

18 **H2.** Systematic biases arise in estimates of stream-aquifer exchanges if
19 local-scale heterogeneities are smoothed by aggregating conductivity into
20 effective values for the set of 16 sub-regional zones.

21 **H3.** The biases are the result of slow-paths in groundwater flow that
22 emerge due to small-scale heterogeneities in conductivity fields.

23 The hypotheses H1-3 are tested by comparing simulated stream-aquifer exchanges
24 produced by the local realizations to the zonal simulation.

25 Hypotheses related to H1 and H2 have been investigated in a few recent studies.
26 [Kalbus et al. \(2009\)](#) evaluated the relative effects of aquifer conductivity and
27 streambed conductance by simulating stream-aquifer interactions in a 220m reach of a
28 small stream in Germany. They found that heterogeneity in the aquifer influenced

1 discharge to the stream more strongly than did variations in the streambed itself.
2 [Schmidt et al. \(2006\)](#) also studied the spatial distribution and magnitude of
3 groundwater discharge to a stream with a simple analytical model that was mainly
4 focused on groundwater discharge at the reach scale. [Kurtz et al. \(2013\)](#) generated
5 multiple realizations of stream-aquifer interactions in the Limmat aquifer system in
6 Zurich Switzerland. They allowed riverbed hydraulic conductivities to take one of
7 four different levels of heterogeneity ranging from local variability at each grid-point
8 to effective conductances of only 5, 3, and 2 values. They found that effective
9 conductance did not always reproduce fluxes obtained from base simulations where
10 system parameters were perfectly known, and furthermore, simulations based on
11 effective parameters gave biased estimates of net exchanges between aquifer and
12 stream. [Mendoza et al. \(2015\)](#) used a land surface model with strong a priori
13 constraints on parameters to argue that such models can perform poorly when spatial
14 variability and hydrologic connectivity of a region are represented coarsely. [Lackey et](#)
15 [al. \(2015\)](#) used a synthetic stream-aquifer system to show that modeling streambed
16 conductance as a homogeneous property can lead to errors in estimated stream
17 depletion.

18 The current study contributes to the understanding of these topics in two ways.
19 First, results indicate that locally heterogeneous hydraulic conductivity fields lead to
20 systematic reductions of regional stream-aquifer exchanges, especially in the direction
21 of aquifer-stream discharges. The causal mechanism appears to be slow paths that
22 emerge in groundwater flows in response to local heterogeneities. These local-scale
23 fluxes are spatially coherent and are not averaged out by aggregating conductivity
24 into a few effective zonal parameters. The nonlinear effects of these coherent
25 structures are propagated to regional scales. Second, hypotheses H1-3 have practical
26 implications for allocations of water resources since many resource management
27 decisions depend on the results of computational models like the base model used
28 here ([Perkins and Sophocleous, 1999](#); [Fleckenstein et al., 2006](#)). A regional basin is
29 partitioned into a few zones, on the basis of physical, geographical and geological

1 criteria (Eaton, 2006), and effective zonal parameters are set by calibration
2 (Christensen et al., 1998). Since effectively parameterized zonal models cannot
3 produce the nonlinearities observed here, the base model is found to systematically
4 overestimate regional stream-aquifer exchanges. This holds true for all levels of
5 heterogeneity investigated here as controlled by CV.

6 The paper is organized as follows. In Section 2, the experimental framework used
7 to model the MHRB and the stochastic methods used to produce locally
8 heterogeneous hydraulic conductivity fields are described. Section 3 reports methods
9 and results of investigations of H1 and H2. It presents results comparing simulations
10 of stream-aquifer exchanges derived from (1) a base model whose effective
11 conductivity parameters are specified by zone and (2) stochastic realizations of
12 heterogeneous local-scale conductivity. Section 4 focuses on H3. The emergence of
13 coherent slow paths in the flow fields is investigated as the source of the bias in base
14 model estimates. Results are summarized and discussed in Section 5, and a few
15 directions for future experiments are sketched.

16

17 **2. Experiment setting and model development**

18 The experimental approach is to compare alternative numerical models of
19 hydrologic system dynamics that produce stream-aquifer exchanges in the Middle
20 Heihe River Basin (MHRB), a closed basin in northwestern China of area of 8778
21 km². Stream-aquifer exchanges are estimated by (1) a so-called base model that
22 specifies hydraulic conductivity on zonal scales and (2) a set of related models that
23 incorporate local-scale heterogeneity in their parameterizations. Zonal conductivity in
24 the base model is calibrated using the PEST parameter estimation tool (Doherty et al.,
25 1994; Sophocleous et al., 1999), and the base model assigns the same calibrated value
26 of hydraulic conductivity to every grid point in a given zone.

27 The level of heterogeneity of hydraulic conductivity is the experimental control
28 variable in local-scale simulations. It is set by adding varying levels of randomness to
29 the zonal values of the base model. The result is a random field of conductivities with

1 expected values that are the same in a given zone as the base model's effective
2 conductivity. Otherwise, system parameters including specific yield and streambed
3 conductance and boundary conditions are the same for all simulations. All
4 computations are conducted using the MODFLOW simulation environment ([Leake](#)
5 and [Prudic, 1991](#); [McDonald and Harbaugh, 2003](#); [Rodríguez et al., 2006](#)).

6 **2.1. Study area**

7 The MHRB is contained in the Heihe River Basin, the second largest closed river
8 basin in China, which drains a total area of 142.9 thousand km². The MHRB is in a
9 semi-arid region with average annual rainfall of approximately 50-200mm/y, most of
10 which occurs from May to October. The river originates in the Qilian Mountains of
11 Qinghai Province and then flows through Gansu Province to western Inner Mongolia.
12 The average annual inflow from mountain areas upstream is 1.58 billion m³ as
13 measured at the Yingluoxia Gauge (YLX), and the discharge through the Zhengyixia
14 Gauge (ZYX) is 0.95 billion m³. The 184 km long reach from YLX to ZYX defines
15 the MHRB study area (Fig. 1), which is known to be influenced by temporal patterns
16 and spatial distributions of stream-aquifer exchanges ([Zhou et al., 2011](#)).

17 Intensive stream seepage and groundwater discharge occur in the study area along
18 the river in fluvial and alluvial fans. The exchanges proceed in two directions:
19 groundwater flows through the streambed into the stream (discharging gaining
20 stream), and stream water infiltrates through the streambed into the groundwater
21 system (leaking losing stream) which is the same as general states presented in
22 ([Winter, 1995](#); [Kalbus et al., 2006](#)). Groundwater discharge to streams maintains
23 natural hydrologic systems during periods of low stream flow through base flow
24 ([Sphocleous et al., 1995](#); [Hantush, 2005](#)). Streams are an important source of
25 groundwater recharge at higher elevations. Conjunctive use of surface and subsurface
26 waters is a potentially important resource management tool in semi-arid regions like
27 the Heihe River Basin whose potential also depends on system state estimates.

28 The Middle Heihe River leaks into a sandy fluvial area immediately after YLX
29 and then enters an alluvial plain composed of fine soil. At that point substantial stream

1 seepage, diversion, and groundwater discharge (base flow) occur along the stream.
2 The Middle Heihe River is a wide and shallow stream, and the area of the stream
3 channel is much greater than the lateral areas of both banks. Thus the exchange of
4 stream water and groundwater occurs primarily in the vertical direction. The plain of
5 fine-grained soil along the river is the main area where springs and groundwater
6 emerge. Thirty irrigation areas are distributed throughout the study area. Groundwater
7 tables range from 0m to 300m beneath the land surface, and aquifer thickness varies
8 from tens to a couple of hundred meters.

9 **2.2. Numerical model of stream-aquifer exchanges**

10 All stream-aquifer interactions are simulated using the numerical modeling tool
11 MODFLOW with the stream package (STR) for one-dimensional stream flow and
12 two-dimensional groundwater flow in this study. The stream package ([Leake and](#)
13 [Prudic, 1991](#)) simulates stream flow with Manning's equation and interactions with
14 groundwater flow using Richard's equation, which assumes vertical, gravity-driven
15 flow and neglects capillarity. This is an acceptable assumption for the typical alluvial
16 sediments of the kind found in the MHRB ([Spanoudaki et al., 2009; Huang, 2012](#)).

17 The stream and aquifer systems are coupled in the model through an iterative
18 process in which convergence of state variables linking the two domains is used as the
19 criterion for accepting a solution. The stream stage and groundwater table are iterated
20 at each time step until the differences between two iterations are within a small
21 tolerance. Seepage is calculated from the product of the head difference times a
22 streambed conductance. In regional-scale groundwater models it is reasonable to
23 assume a homogeneous low-conductivity streambed within a heterogeneous aquifer
24 ([Kalbus et al., 2009](#)). Conductance is assigned based on an existing calibration and
25 measured data values between 1-3 m/d ([He and Zhao, 2007; Zhou et al., 2011](#)).

26 Since aquifer thickness is small compared to its horizontal dimensions, one layer
27 is sufficient for the vertical discretization. [Workman et al. \(1997\)](#) pointed out that the
28 discrete cell size can be 1-2km in regional groundwater flow models, as a river can
29 affect water table elevations as far as 1525 m from the middle of the stream. Thus the

1 study domain is discretized into 155 rows and 172 columns, with each cell size 1km x
2 1km. The aquifer is simulated as a free-surface boundary able to fluctuate in response
3 to recharge from irrigation fields, evapotranspiration, flow to drains, and interaction
4 with streams. The recharge package (RCH) is used to input rainfall and irrigation
5 seepages as vertical boundaries, and the recharge distribution is set according to the
6 data from [Hu et al. \(2007\)](#). The ETS package is employed to simulate phreatic
7 evaporation considering the multiple-linear relationship with groundwater depth.
8 Based on previous studies ([Zhou et al., 2011](#); [Tian et al., 2015](#)), spring flow is also
9 calculated in irrigation areas by representing the drainage area as drain cells.

10 The groundwater table observed on January 1 of the year 2000 is input and
11 interpolated as initial head for all simulations. The time step is 5 days, and the
12 simulation period is one year. Stream inflow at the YLX Gauge and groundwater
13 lateral recharges from mountain areas are used as an upper boundary, and outflow at
14 the ZYX Gauge is taken as a lower boundary ([Zhou et al., 2011](#)). In the mountain
15 front region, the aquifer has a high proportion of pebbles, sand and gravel, while the
16 plain of the lower area is composed of thick unconsolidated alluvial fills with alluvial
17 sand, silt, and clay ([Hu et al., 2007](#)).

18 **2.3. Base model**

19 Previous research has identified 16 major hydrologic zones within the MHRB
20 based on variations of hydraulic conductivity and specific yield ([Hu et al., 2007](#); [Jia et](#)
21 [al., 2009](#)). In this study, the PEST parameter estimation system is used to calibrate
22 aquifer parameters (zonal horizontal conductivity K_z and specific yield S_z) for the
23 different zones, $z = 1, \dots, 16$. The parameterization of conductivity by constant
24 effective values, one each for the 16 coarse zones, is the foundation of the base model.
25 The 16 zonal conductivities are also used as expected values for the stochastic
26 generation of locally refined conductivity fields as described in the next sub-section.

27 The objective function for PEST estimates is the sum of squared differences
28 between observations and simulations of groundwater tables at 34 wells distributed
29 through the basin. Effective parameters K_z and S_z are optimized and estimated

1 separately for the 16 zones of the domain due to computational limitations. The
2 lower and upper levels of parameter settings for conductivity are 0.5~50 m/d and are
3 0.01~0.25 for specific yield without considering anisotropy. The procedure of PEST
4 includes an initial MODFLOW simulation, calculation of the Jacobian matrix for each
5 parameter, and different Lambda upgrade. After the calibration of K_z values the
6 specific yields of the system, S_z , are optimized with PEST for transient conditions by
7 minimizing the groundwater table errors at observation wells.

8 **2.4. Stochastic realization of conductivity fields**

9 Stochastic simulations are the basis for local-scale realizations of hydraulic
10 conductivity fields, $K(x)$, used in comparisons. The variable x corresponds to a point
11 in the 155 x 172 computational grid used in all simulations of hydrologic system
12 dynamics in the MHRB. Local parameters $K(x)$ are defined by log-normally
13 distributed perturbations on the zonal conductivity field of the base model,

14
$$\ln K(x) = Y(x) = Y_z + Y'(x), \quad Y'(x) \sim N(0, V_z) \text{ for all } x \in \text{zone } z. \quad (1)$$

15 The zonal mean, Y_z , and variance, V_z , are defined in terms of the base model
16 parameters, K_z ,

17
$$Y_z = \ln \frac{K_z}{\sqrt{1 + C}}, \quad V_z = \text{Var}[Y(x)] = \ln(1 + C), \quad x \in \text{zone } z \quad (2)$$

18 where C is the Coefficient of Variation (CV).

19 In order to compare the effects of different levels of total heterogeneity,
20 realizations of conductivity fields are generated corresponding to six values $C = 0.1,$
21 $0.2, 0.5, 0.8, 1.0, 2.0$. Grid-scale analyses at different levels of heterogeneity are based
22 on a sample of 10 realizations of log-normally distributed random conductivity fields
23 for each value of C . Local scale simulations of groundwater flow and stream-aquifer
24 exchanges are produced using realizations of K fields based on Eqns. (1) and (2) in
25 the general model.

26 Grid point values of conductivity are sampled independently, so no correlation
27 structure is imposed on local conductivities. It is generally accepted that hydraulic

1 conductivity is a random field that is typically correlated over a continuous range of
2 small scales, becoming uncorrelated once a characteristic correlation length is reached
3 (Zhang, 2002; Rubin, 2003). The effect of small-scale correlations on regional
4 velocities and mass transport is understood through second-order for spatially
5 stationary conductivity fields (Gelhar and Axness, 1983; Winter et al., 1984; Neuman
6 and Orr, 1993) and for a wider range of fields that are heterogeneous at local and
7 regional scales (Winter and Tartakovsky, 2002). The small scale of our study is set
8 by the 1 km x 1 km size of our grid cells. Since many alluvial systems have
9 correlation lengths that are much less than that, e.g., Rehfeldt et al. (1992) and Riva et
10 al. (2006), it is informative, and in some cases realistic, to investigate the effect that a
11 field of locally independent, identically distributed conductivities has on the state
12 variables of regional models. Indeed, our experiments show that stream-aquifer
13 exchanges estimated by a typical regional simulation like the base case can be biased
14 even when local conductivities are uncorrelated.

15 **2.5. Tests of Hypotheses**

16 Comparisons between simulations produced by the base model and local-scale
17 models depend on either normalized squared departures between fluxes produced by
18 the base model and the locally heterogeneous models or ratios between the same
19 fluxes. The results of comparisons are said to support H1 and H2 if simulated values
20 produced by the base model exhibit systematic bias with respect to local-scale
21 simulations of stream-aquifer exchanges. H3 is evaluated by comparing normalized
22 squared departures of groundwater fluxes produced by the base model with samples
23 produced by locally resolved simulations.

24 **3. Simulated stream-aquifer exchanges**

25 This section focuses on H1 and H2, the hypotheses that simulations of exchanges
26 of water between the stream and aquifer systems produced using locally
27 heterogeneous conductivity fields are systematically less than estimates produced by
28 the base model. Results are given separately for the base model and for models with

1 parameterizations affected by local scales of heterogeneity and then compared.
2 Levels of CV are manipulated as control parameters in the experiments. Simulated
3 groundwater fluxes are used later (Section 4) to evaluate H3, the hypothesis that
4 locally heterogeneous conductivity affects the emergence of preferential paths in the
5 aquifer.

6 **3.1. Base model calibration and simulation results**

7 The lower fluvial plain of the MHRB is mainly composed of relatively low
8 permeability silts and clay, while the upland aquifer, containing mostly sand mixed
9 with gravel, is more permeable (Hu et al., 2007). This distribution is reflected in
10 estimates of zonal conductivity and specific storage, K_z and S_z , derived from a PEST
11 calibration based on a record of 5-day groundwater table variations in the year 2000
12 (Fig. 1). The differences between well heights simulated using the PEST calibration
13 and observations yield a residual mean of -0.19 m, standard error of 0.021 m, root
14 mean square (RMS) of 1.4 m, and normalized RMS of 0.77%. The PEST calibration
15 results specify higher conductivity for upstream areas near mountain fronts than for
16 the lower fluvial plain in accord with the characteristic geology of the basin.

17 Simulated river stages and groundwater tables for June and December along the
18 183km river distance are displayed in Fig. 2a and Fig. 2b. The groundwater table is
19 generally much lower than the stream stage for about 25 km below Yingluoxia Gauge
20 (YLX), indicating a losing reach of stream, as can be seen from these two figures. The
21 groundwater table becomes quite shallow in the plain downstream of YLX, and the
22 stream may be losing or gaining at different seasons and locations until the outlet at
23 the Zhengyixia Gauge (ZYX).

24 River stage in June is relatively higher than that in December (Fig. 2c), as the
25 stream flow is higher in flood season. Nevertheless, the groundwater table along the
26 river in December is a bit higher because there is almost no pumping for irrigation in
27 that month. Stream inflow from YLX exhibits clear seasonal variations within the
28 year with a large pulse of water in Spring-Summer due to snowmelt (Fig. 2d). This
29 effect is clearly reflected by monthly stream leakage to groundwater ($m^2/km/day$). On

1 the other hand, groundwater discharge to the river (or base flow) is relatively stable
2 and less controlled by the seasonal inflow from station YLX.

3 The pattern of groundwater flow is not only controlled by the distribution of
4 hydraulic conductivity, but also by the configuration of the water table. Magnitudes of
5 average groundwater flow velocity are higher near the mountain fronts than in the
6 lower plains (Fig. 3a). Average monthly stream leakage rates (\bar{Q}_b), where the average
7 is taken over all monthly values, exhibit critical transition points between upwelling
8 and downwelling zones (Fig 3b). The stream is losing in deep groundwater areas near
9 mountain fronts and becomes a gaining stream in the lower plain where groundwater
10 tables are shallow.

11 The intensity of stream-aquifer interactions also varies seasonally. Interactions in
12 summer and autumn show similar patterns with water exchanges of losing streams
13 and gaining streams stronger than the monthly average. Stream inflow is higher and
14 recharge to groundwater (precipitation and irrigation seepage) is relatively larger in
15 these seasons. Thus, leakage from losing streams and groundwater discharge along
16 gaining streams both become larger than the corresponding seasonal average. The
17 magnitudes of water exchange are more spatially variable in the high-flow seasons of
18 winter and spring: the magnitude augmentation (compared with seasonal average) of
19 water exchanges for losing streams and gaining streams can be as large as 1041 and
20 $317 \text{ m}^2/\text{km/day}$ respectively. However, the water exchanges of the losing streams and
21 gaining streams become much weaker in spring and winter, as the stream inflow from
22 YLX is relatively low and the groundwater recharge (precipitation and irrigation
23 seepage) is quite small.

24 **3.2. Influence of local heterogeneity on simulated results**

25 Groundwater simulations based on local (grid-scale) simulations of hydraulic
26 conductivity $K(x)$ display more spatially heterogeneous system states than the base
27 model. Realizations of highly heterogeneous $K(x)$ fields are produced by applying
28 the method of Eqns. (1) and (2) in Section 2. Fig. 4 displays grid-scale perturbations
29 of $K(x)$ from the zonal mean K_z for coefficients of variation (CV) of magnitude $C =$

1 0.1 and $C = 2$, as indicated by $K(x)/K_z$. Higher perturbations from zonal mean
 2 conductivity occur for higher CV, which follows from the definition of zonal standard
 3 deviation (Eqn. 2).

4 Grid-scale simulations, $W_G(x)$, of groundwater tables at the end of the simulation
 5 period (Dec) are more variable than the base result, $W_B(x)$. Point-wise differences
 6 between the two, $\Delta W(x) = W_B(x) - W_G(x)$, are about equally distributed between grid
 7 cells where the grid-scale water table is higher than the base case ($\Delta W(x) < 0$) and
 8 lower ($\Delta W(x) > 0$), although there is a slightly greater chance of a rise. The number
 9 of cells where the absolute change in the water table is large ($|\Delta W(x)| > 1$ m) increases
 10 systematically with CV, and the number of cells where $|\Delta W(x)| < 1$ m decreases
 11 correspondingly. The point-wise differences exhibit a pattern of compact spatially
 12 coherent areas where water tables consistently increase or decrease together. This is
 13 despite the lack of any correlation in the values of grid-scale conductivity.

14 It is convenient to use an indicator,

15 $I_m = Q_m / Q_m^{(B)}$ (3)

16 to compare monthly stream-aquifer exchanges produced by local grid-scale models,
 17 Q_m , with monthly exchanges produced by the base model. Here $m = 1, \dots, 12$
 18 indicates the month.

19 The base model simulation, $Q_m^{(B)}$, of total water exchanged between the stream
 20 and aquifer during month m summed over all $x = 1, \dots, n$ grid cells along the stream is,

21 $Q_m^{(B)} = \sum_{x=1}^n Q_m^{(B)}(x)$ (4)

22 Similarly, Q_m , the mean total exchange for month m averaged across all $r = 1, \dots, N_r$
 23 (=10) locally refined models is

24 $Q_m^{(r)} = \sum_{x=1}^n Q_m^{(r)}(x)$ and $Q_m = \frac{1}{N_r} \sum_{r=1}^{N_r} Q_m^{(r)}$. (5)

25 The mean monthly value of the aquifer-stream discharge indicator is always less
 26 than 1 for all values of CV: the base model produces discharge values that are
 27 systematically larger than the locally heterogeneous models (Fig. 5). When two

1 standard deviation confidence intervals are calculated for I_m , a few intervals
2 corresponding to the smallest CVs include 1 for some months. In those cases
3 small-scale heterogeneities increase discharge in some local areas and the overall
4 effect of heterogeneity is reduced. When CV equals 0.1, the monthly aquifer-stream
5 discharge is quite close to the base model value (i.e., the expected value) since
6 perturbations around the expected values are small in that case. On the other hand,
7 mean indicator values decrease to nearly 0.8 as CV increases from 0.1 to 2, and
8 standard deviations continue to increase. As heterogeneity in local models increases,
9 the monthly aquifer-stream discharge indicator decreases and uncertainty as measured
10 by standard deviation increases. The discharge ratio also shows some seasonal
11 patterns, with smaller effects of heterogeneity in high-flow months, such as August
12 and September. Compared with other months, the aquifer-stream discharge quantity is
13 larger and the uncertainty is less in these months.

14 The mean monthly value of the corresponding stream-aquifer leakage indicator
15 is also less than 1 in most months, and the standard deviations show increasing trends
16 with increasing CVs, but the relative magnitudes of leakage are less than those of
17 discharge and are closer to the base model (Fig. 6). Taking CV equaling to 2 for
18 example, the ranges of mean values for stream-aquifer leakage and discharge are
19 respectively 0.97-0.99 and 0.81-0.93, and the respective ranges of standard deviations
20 for stream-aquifer leakage and discharge are 0.004-0.013 and 0.020-0.044.

21 The entire Middle Heihe River includes upstream segments where the dominant
22 exchange mechanism is leakage and downstream segments where discharge is
23 dominant. To separate these two confounding classes of fluxes, spatial effects of
24 heterogeneous conductivity are further investigated by calculating indicator values
25 ($Q_m/Q_m^{(B)}$) for four major sub-streams in the MHRB. Results for discharge and leakage
26 are shown for all 10 realizations of conductivity for each value of CV (Figs. 7 and 8).
27 Stream 1 is an upstream segment while Streams 2-4 comprise the main stem of the
28 river in the alluvial plain. The indicator values are almost all less than 1 for average
29 monthly discharge from these four sub-streams (Fig. 7), corresponding to the

1 relatively larger impact of conductivity heterogeneity on groundwater processes
2 leading to discharge. Stream-aquifer leakage is larger than the corresponding base
3 model in every sub-stream during at least some months (Fig. 8).

4 The stream-aquifer water exchanges of these four sub-streams respond
5 differently to local heterogeneity, reflecting the different groundwater flow regimes of
6 the different sub-regions. For example, Stream 1 is mainly a losing stream located in a
7 mountain front area with a deep groundwater table, thus aquifer heterogeneity has a
8 relatively small effect on stream-aquifer exchanges. Streams 2-4 go through the plain
9 where water exchanges are affected by pumping and recharge schemes, and the
10 differences between Q_m and $Q_m^{(B)}$ are more obvious. Since groundwater depths are
11 shallow and extensive stream-aquifer interactions occur in these areas, the effects of
12 aquifer heterogeneity on water exchanges of these streams are relatively strong. The
13 values of $Q_m/Q_m^{(B)}$ are smaller during summer and autumn, when the stream inflow
14 from YLX and precipitation are relatively large.

15

16 **4. Slow paths**

17 In this section evidence is presented for the hypothesis (H3) that the systematic
18 bias observed in Section 3 is due to the response of groundwater flow to local
19 heterogeneities in conductivity fields. Spatially coherent slow paths are seen to
20 emerge in groundwater flows simulated in realizations of heterogeneous conductivity
21 fields, and their density increases as the CV of experiments increases. This occurs
22 despite the lack of any correlation in the spatial distribution of local-scale
23 conductivity itself.

24 **4.1. Local scale conductivity field heterogeneity**

25 It is well-known that groundwater system parameters exhibit heterogeneity on
26 multiple scales ranging down to the smallest (Neuman et al., 1987; Neuman, 1990,
27 1994). Here the statistics of locally heterogeneous conductivity fields are analyzed
28 and compared to the mean fields corresponding to the calibrated conductivity

1 parameters of the base model. A single realization is chosen at random from the 10
2 available per level of CV, and its statistics are analyzed. Results for other
3 realizations are essentially the same.

4 The point x is a grid cell in the z^{th} zone. The ratio

5 $\rho_z(x) = K(x)/K_z$ (6)

6 compares $K(x)$, the local conductivity at x to K_z , the expected value of $K(x)$ defined by
7 the calibrated zonal conductivity parameterizing the base model. The summary
8 statistics of $\rho_z(x)$ describe the overall behavior of normalized local conductivity when
9 adjusted for zone (Table 1). The arithmetic means of the $\rho_z(x)$ equal 1 for all values of
10 CV, implying that the conductivity field is indeed a realization drawn from an
11 ensemble of fields whose zonal expected values are the base model values derived
12 from the initial calibration. The standard deviations of the $\rho_z(x)$ increase with CV.
13 Both these behaviors are expected. Since hydraulic conductivity is bounded below by
14 0, the ranges of minimum values are also limited; nonetheless, minimum values
15 decrease by two orders of magnitude, which is consistent with the increase in CV.
16 Maximum values increase with increasing CV. Calculations of Moran's I (not shown)
17 indicate that values of K are not spatially correlated for any CV.

18 The geometric mean of $\rho_z(x)$ declines from 0.96 to 0.76 as the heterogeneity of
19 the conductivity field (CV) increases. Since the effective conductivity of a
20 heterogeneous 2D field is equal to its geometric mean (Matheron, 1967; Gutjahr et al.,
21 1978), groundwater flow is expected to be slower in more heterogeneous fields where
22 CV is higher. A related effect is observed in the simulations of aquifer-stream
23 discharge reported in Section 3.2 where both overall discharge and discharge in
24 individual stream segments (Fig. 5 and Fig. 7) decline as CV increases. This is further
25 suggested by the trend in the percentage of grid cells with $K(x)$ values less than K_z (i.e.
26 $\rho_z(x) < 1$) which increase from 54.8% to 64.2% with CV increasing from 0.1 to 2. It
27 should be noted that the median value also shows a decreasing trend from 0.99 to

1 0.84.

2 These trends are evident in grid-scale $K(x)$ field comparison maps based on the
3 realization used to construct Table 1 (Fig. 9). Maps for other realizations are similar.
4 Black grid cells are locations where $K(x) < K_z$, while white grid cells have values
5 $K(x) > K_z$. The number of black grids clearly increases with heterogeneity (CV),
6 illustrating the increasing percentage of area occupied by locations with smaller
7 conductivity than expected.

8 **4.2. Emergence of slow paths**

9 To evaluate hypothesis H3, we compare groundwater fluxes $q(x)$ and $q_b(x)$ obtained
10 respectively from (i) the flow simulation based on the same realization of a locally
11 heterogeneous conductivity field, $K(x)$, that was selected in Section 4.1 and (ii) the
12 simulation based on the calibrated base model field, K_z . We adopt the normalized
13 difference

$$14 \quad \Delta q(x) = (q(x) - q_b(x)) / \sqrt{q(x) \bullet q_b(x)} \quad (7)$$

15
16 to indicate differences between flow simulations. Here $\Delta q(x) < 0$ indicates a point
17 where flow due to local variations of conductivity, $q(x)$, is less than flow due to the
18 base model, $q_b(x)$. The normalization removes large-scale effects due to zonal
19 averages from the analysis.

20 Summary statistics of flow (Table 2) support the hypothesis that locally
21 heterogeneous conductivity fields induce reduced overall flows even though the
22 heterogeneous conductivities are spatially independent. The arithmetic mean
23 difference drops from -0.05 to -0.51 as CV increases from 0.1 to 2, and the median
24 and minimum values exhibit a declining trend as well, implying that groundwater
25 flow estimated by the local-scale model is slower at most points than the groundwater
26 flow estimates from the base model. Meanwhile, standard deviations steadily
27 increase from 0.175 to 0.833, indicating that groundwater flow becomes more
28 variable as CV increases. Additionally the percent of area that is occupied by
29 relatively lower groundwater flow ($q(x) < q_b(x)$) increases with increasing

1 heterogeneity.

2 This fact is amplified by examining the spatial distribution of normalized
3 differences $\Delta q(x)$ (Fig. 10). Other realizations exhibit similar patterns. Black grid
4 cells correspond to locations where $\Delta q(x) < 0$, while white grid cells are points where
5 $\Delta q(x) > 0$. As aquifer heterogeneity increases, the area covered by black grid cells
6 expands, consistent with the last line of Table 2. This is somewhat more pronounced
7 in low-lying regions where zonal mean hydraulic conductivities are low (cf. Fig. 1).

8 The distribution of locations where $\Delta q(x) > 0$ ($\Delta q(x) < 0$) is fairly uniform in the
9 simulation with relatively low heterogeneity ($CV = 0.1$). Areas in zone 1 where there
10 seems to be a slightly higher concentration of $\Delta q(x) < 0$ may be exceptional. It should
11 also be noted that relatively strong local anisotropies in the direction of the stream
12 appear throughout the basin, except in zone 1. The relatively uniform spatial
13 distribution of the $\Delta q(x)$ index begins to disappear as CV increases. When
14 heterogeneity is high ($CV = 1.0, 2.0$), areas where the $\Delta q(x)$ index is less than zero
15 generally coincide with zones of low conductivity (like zones 1, 2, 3, etc. in Fig. 1).
16 Zones of high conductivity are somewhat harder to pick out, but there are obvious
17 areas where $\Delta q(x) > 0$ that correspond to areas of high zonal conductivity (like zones
18 10, 8, 9, etc. in Fig. 1).

19 Coherent areas of relatively high and low flow build up in the groundwater
20 simulations (Fig. 10), especially at high levels of heterogeneity (high CV). This seems
21 related to the increase of the area occupied by points of low conductivity with
22 increasing heterogeneity (Table 1). Yet points of relatively low conductivity do not
23 exhibit spatial coherence (cf. Fig. 9), nor should they since the random perturbation to
24 conductivity that is added at any given point is chosen independently of all other
25 points. The spatial coherence of low flux areas, as measured by $\Delta q(x) < 0$, results
26 from the cumulative effect on continuous flows of blockages arising from spatially
27 independent low conductivities. The flow paths of other simulations are statistically
28 the same as this one, although their actual locations and geometry are different.

29 **5. Summary and Concluding Remarks**

1 This study focuses on three related hypotheses about the effects of locally
2 variable hydraulic conductivity fields on regional exchanges of water between
3 streams and aquifers: (H1) small-scale heterogeneities of hydraulic conductivity
4 significantly affect simulated stream-aquifer water exchanges in river basins, and
5 hence, computational projections of them; (H2) systematic biases arise in estimates of
6 exchanges if small-scale heterogeneities are smoothed by aggregation into a few
7 sub-regions; and (H3) the biases result from slow-paths in groundwater flow that
8 emerge due to small-scale heterogeneities.

9 The study addresses these hypotheses through computational experiments by
10 simulating system states of the hydrologic systems of the Middle Heihe River Basin
11 (MHRB) of northern China. The study compares (i) estimates of stream-aquifer
12 exchanges and groundwater fluxes produced by a base model that reflects current
13 practice of parameterizing hydraulic conductivity on sub-regional scales to (ii)
14 estimates produced by a related set of models whose conductivity parameters are
15 spatially heterogeneous on local (grid-point) scales. The computational experiments
16 provide evidence that regional system states, specifically exchanges of water between
17 the MHRB's streams and aquifers, respond significantly to local variations of
18 conductivity.

19 Local-scale heterogeneities in conductivity fields cannot be resolved by effective
20 zonal parameterizations used in the base model. The cumulative effects of local
21 heterogeneities combine to produce spatially coherent "slow paths" in the resulting
22 simulations of groundwater flows. Flow paths in locally heterogeneous systems
23 appear to intersect enough relatively low conductivity areas compared to the base
24 model to reduce groundwater flow regardless of zone. The systematic behavior
25 observed in the local-scale experiments is also consistent with theoretical results of
26 [Matheron \(1967\)](#) and [Gutjahr et al. \(1978\)](#).

27 The aquifer-stream discharge response is stronger than the response of
28 stream-aquifer leakage. That may result from discharge's greater sensitivity to
29 variations of hydraulic conductivity in the MHRB. Stream segments that are further
30 downstream are also more affected by heterogeneity than upstream segments. This

1 seems consistent with the overall pattern of leakage in the Heihe River Basin where
2 recharge occurs primarily in upland reaches that are not directly connected to the
3 water table and average conductivities are relatively high, and thus are relatively
4 unaffected by variations in conductivity.

5 The results presented here strongly suggest that local-scale variations in the
6 conductivity system parameter are not averaged out as scales of representation
7 increase, but instead significantly affect regional exchanges of water between streams
8 and aquifers (H1): in these experiments local heterogeneity has non-linear effects on
9 states of the regional system. That is so even when the level of heterogeneity in
10 conductivity is fairly small, as measured by the coefficient of variation of the
11 simulated conductivity fields (CV). These effects increase with CV. This is especially
12 true of discharge, which the sub-regionally scaled base model systematically
13 over-estimates (H2). The emergence of slow paths in the groundwater flow field (H3)
14 appears to be the source of the non-linearities.

15 These non-linear effects have implications for water resources management in
16 addition to their intrinsic scientific interest. Since the introduction of computational
17 tools capable of approximating the dynamics of basin-scale hydrologic systems, it has
18 become common to use effectively (zonally) parameterized coupled models to
19 manage regional water resources (Chen and Shu, 2006; Werner et al., 2006; Dragoni
20 et al., 2013). Although estimates of system states produced by such aggregated
21 models have been assumed to approximate watershed dynamics up to reasonable
22 amounts of unsystematic random error, this assumption is generally not tested. The
23 base model used here, and the methods used to produce it, is typical of such zonally
24 parameterized models. It has been an open question whether zonal parameterizations
25 can produce estimates of system states that account for the effects of local-scale
26 heterogeneity that are known to exist in parameters (Schmidt et al., 2006; Rodríguez
27 et al., 2013). The results of this study reveal a systematic bias in approximations of
28 stream-aquifer exchanges in the MHRB produced by the base model. This is
29 especially true of discharge. Similar effects have been observed in other studies
30 (Kalbus et al., 2009; Kurtz et al., 2013; Mendoza et al., 2015; Lackey et al., 2015).

1 Most of the evidence for (or against) hypotheses like H1-3 will continue to be
2 computational. Several additional lines of investigation suggest themselves
3 immediately. First is the extension of this kind of study to other basins, both natural
4 and synthetic. Extensions to three dimensions may be required in some cases. The
5 respective roles of zonal and local heterogeneity can be investigated explicitly by
6 applying Analysis of Variance ([Winter et al., 2006](#)), or other multivariate statistical
7 techniques.

8 Local-scale heterogeneities are independent in this study. This is a weak
9 assumption in the sense that it places a minimal number of restrictions on the spatial
10 structure of conductivity fields, but its influence on conclusions about H1-3 is unclear.
11 Hydraulic conductivity fields are generally correlated, so experiments that evaluate
12 the effects of different levels and kinds of correlation should be pursued. Synthetic
13 conductivity fields that reproduce specified correlation structures can provide large
14 numbers of realizations of parameter fields with given values of control variables like
15 correlation length and spatial anisotropy.

16 Finally, the existence of slow paths in groundwater flows, is inferred indirectly
17 here from summary statistics (cf. the last line in Table 2) or by qualitative
18 interpretations of results (Fig. 10). A Langrangian analysis based on particle
19 tracking would allow the interaction between variations in conductivity and
20 groundwater fluxes to be quantified more directly.

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1

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38

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40

41

1 **Tables**

2
3

Table 1 Grid-scale K field statistics

$\rho_z = K(x)/K_z$	CV=0.1	CV=0.5	CV=1	CV=2
Arithmetic mean	1	1	1	1
Standard deviation	0.3	0.65	0.98	1.12
Min	0.2	0.04	0.024	0.006
Max	4.1	14.62	37.47	21.64
Median	0.99	0.942	0.89	0.84
Geometric mean	0.96	0.85	0.78	0.76
Percent of $K(x)/K_z < 1$	54.8%	59.0%	62.0%	64.2%

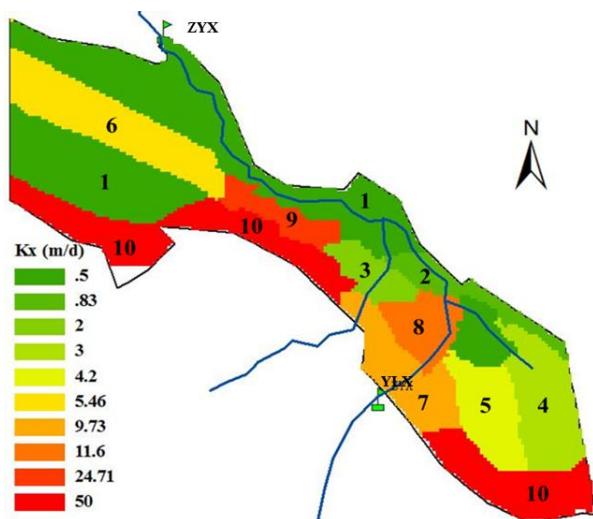
4
5

Table 2 Groundwater flow field statistics with local heterogeneity

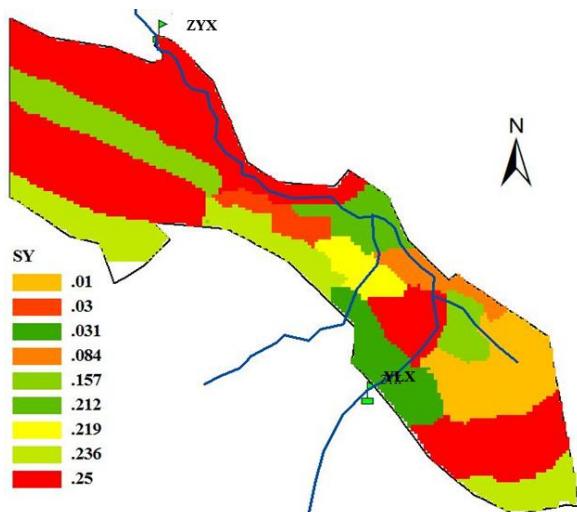
$\Delta q(x)$	CV=0.1	CV=0.5	CV=1	CV=2
Arithmetic mean	-0.05	-0.21	-0.34	-0.51
Standard deviation	0.175	0.398	0.569	0.833
Min	-1.833	-3.338	-4.901	-12.4
Max	1.54	1.429	3.458	2.721
Median	-0.017	-0.075	-0.14	-0.234
Percent of $q(x) < q_b(x)$	61.9%	71.0%	75.5%	79.5%

6
7

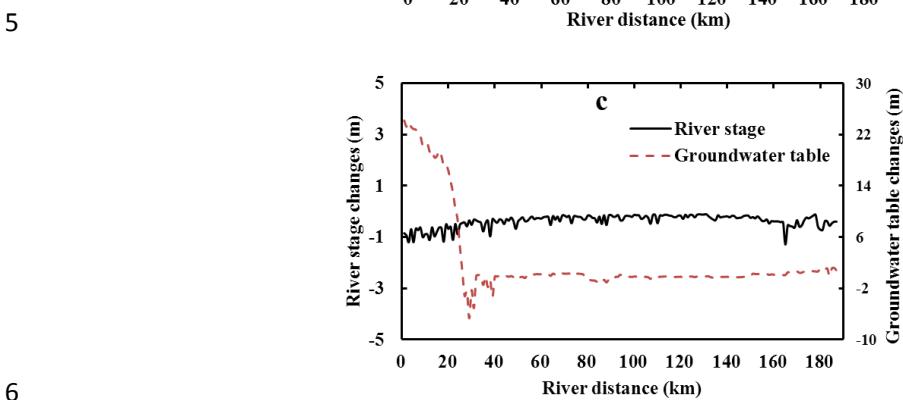
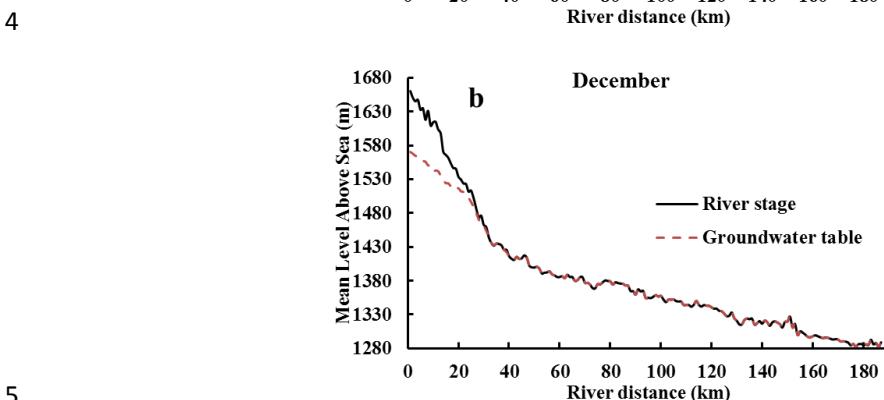
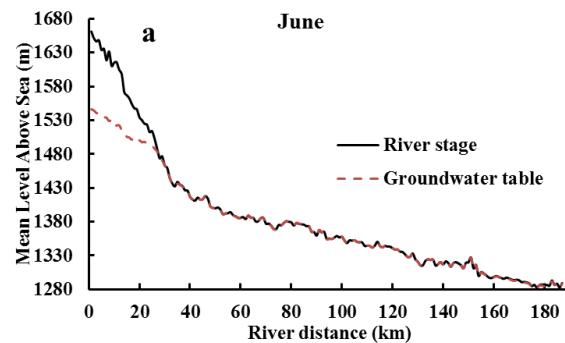
8 **Figures**



9



1
2 Fig. 1. PEST estimated aquifer conductivity K_z and Specific yield S_z
3



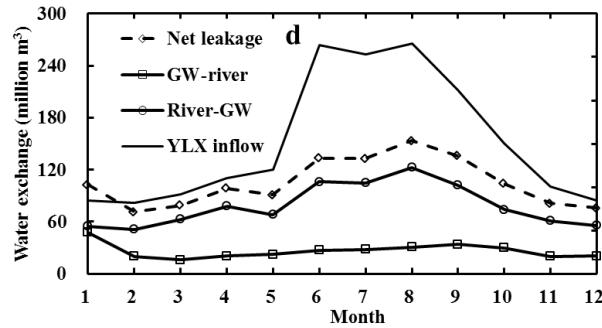


Fig. 2. Simulated groundwater table and river stage along river distance in (a) Jun and (b) Dec; (c) difference between Dec and Jun (Dec-Jun); (d) temporal stream-aquifer exchanges within year

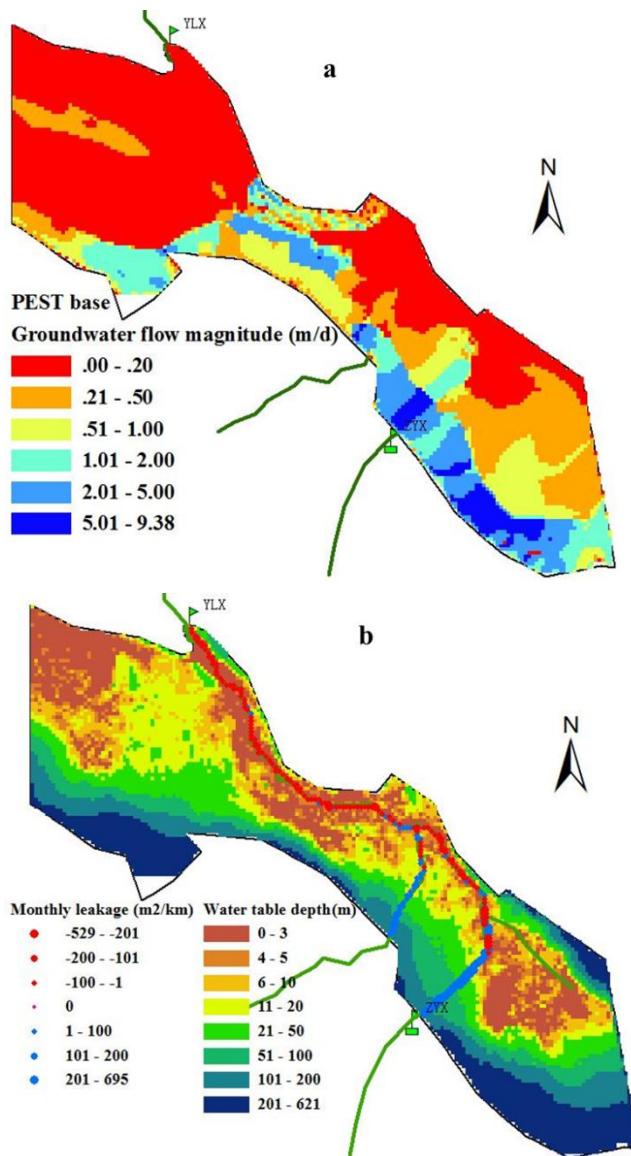
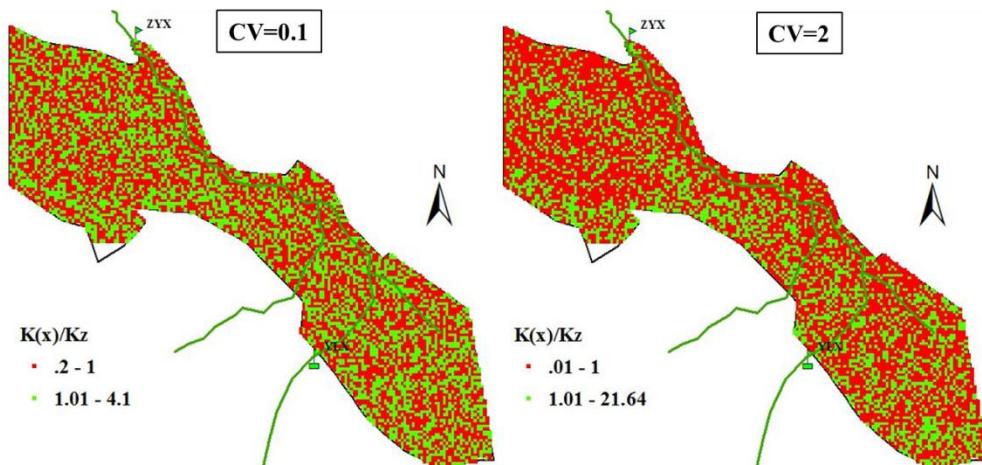


Fig. 3. Groundwater flow rate q_0 distribution (a); Average monthly leakage rate \bar{Q}_b and groundwater table depth (b)

1

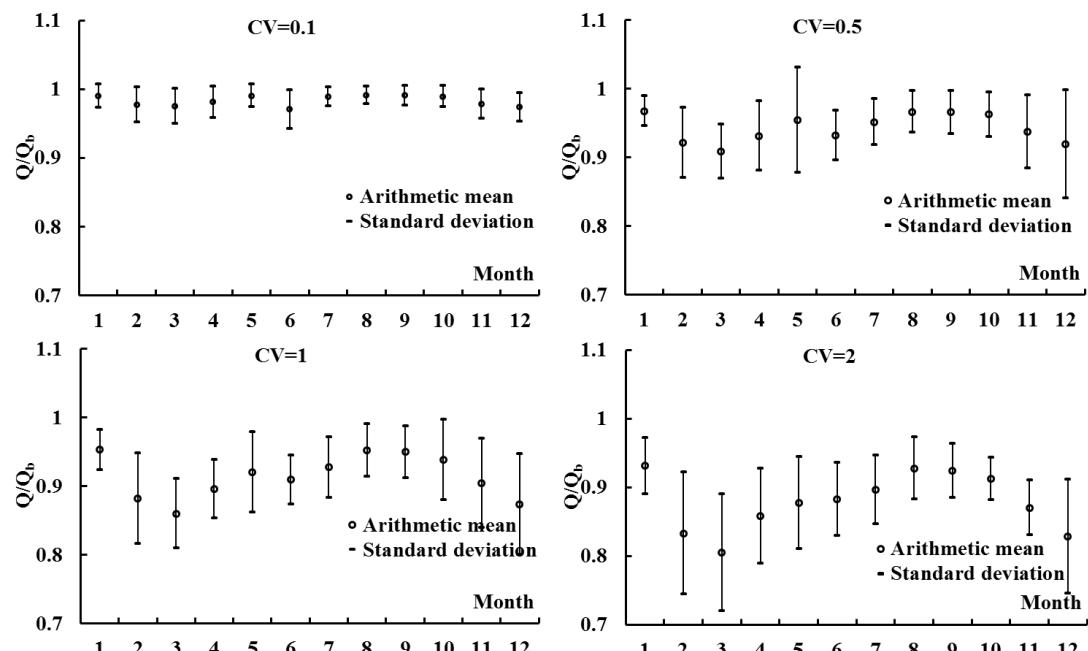


2

3 Fig. 4. Grid-scale realization of K field perturbation when CV equals 0.1 and 2

4

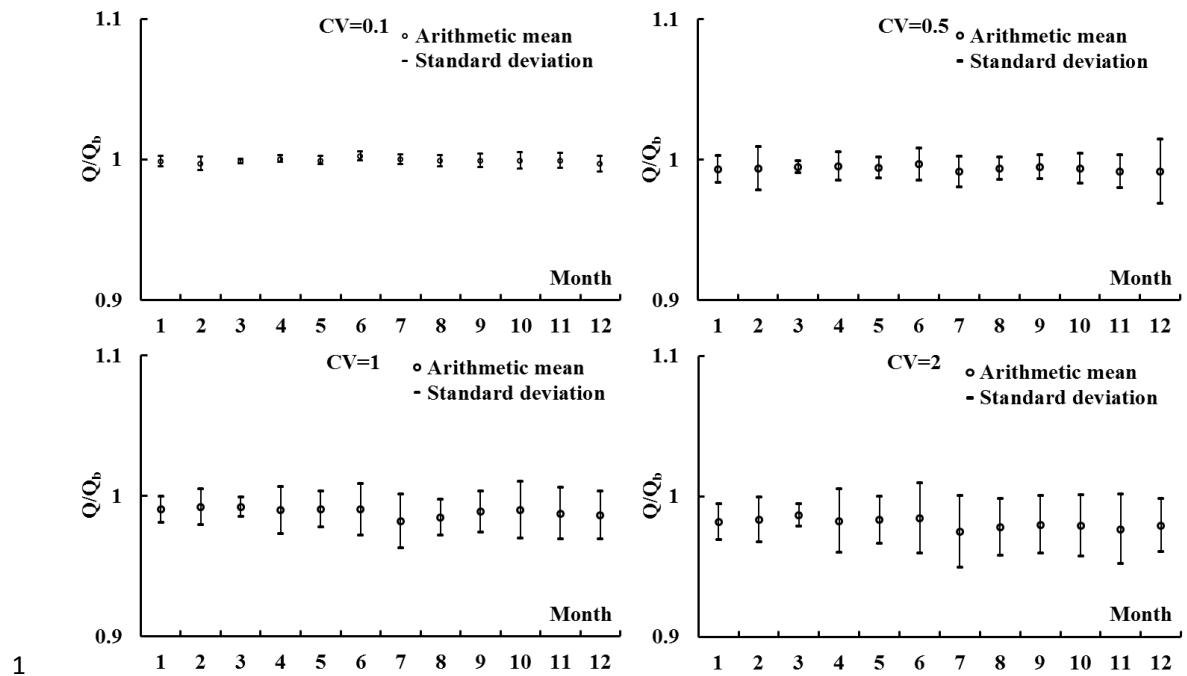
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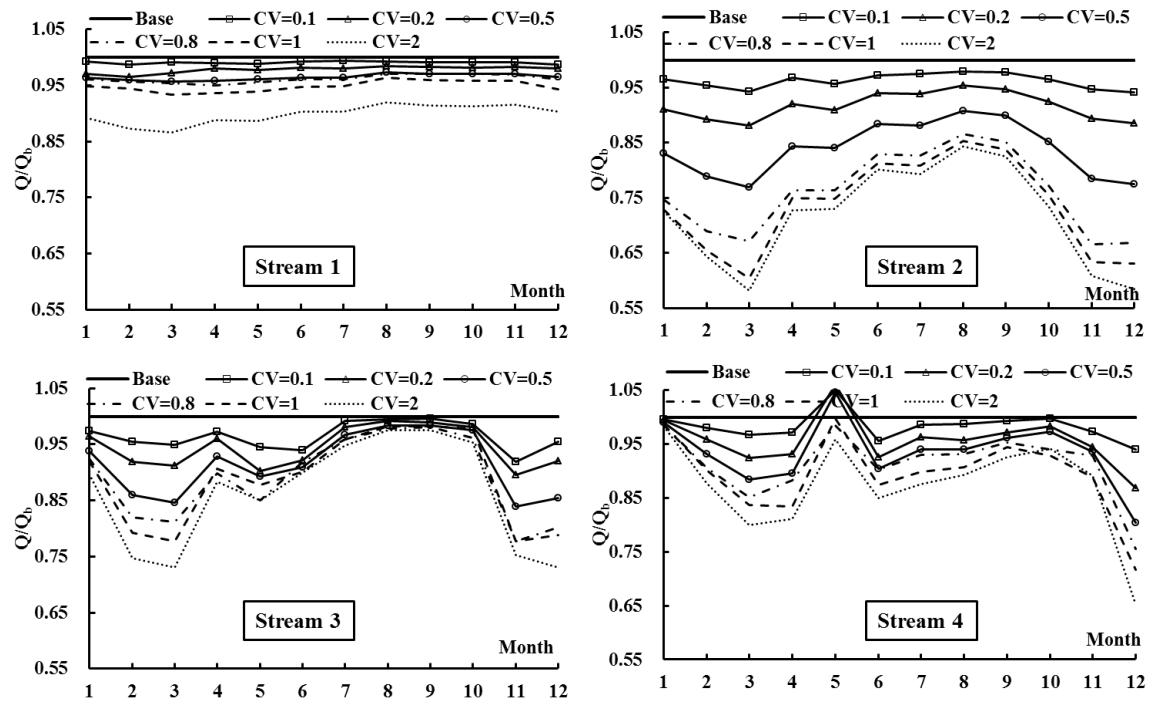
7 Fig. 5. Means and confidence intervals of aquifer-stream discharge from local model

8



2 Fig. 6. Means and confidence intervals of stream-aquifer leakage from local model

3



4 Fig. 7. Aquifer-stream discharges of different streams from refined model

5

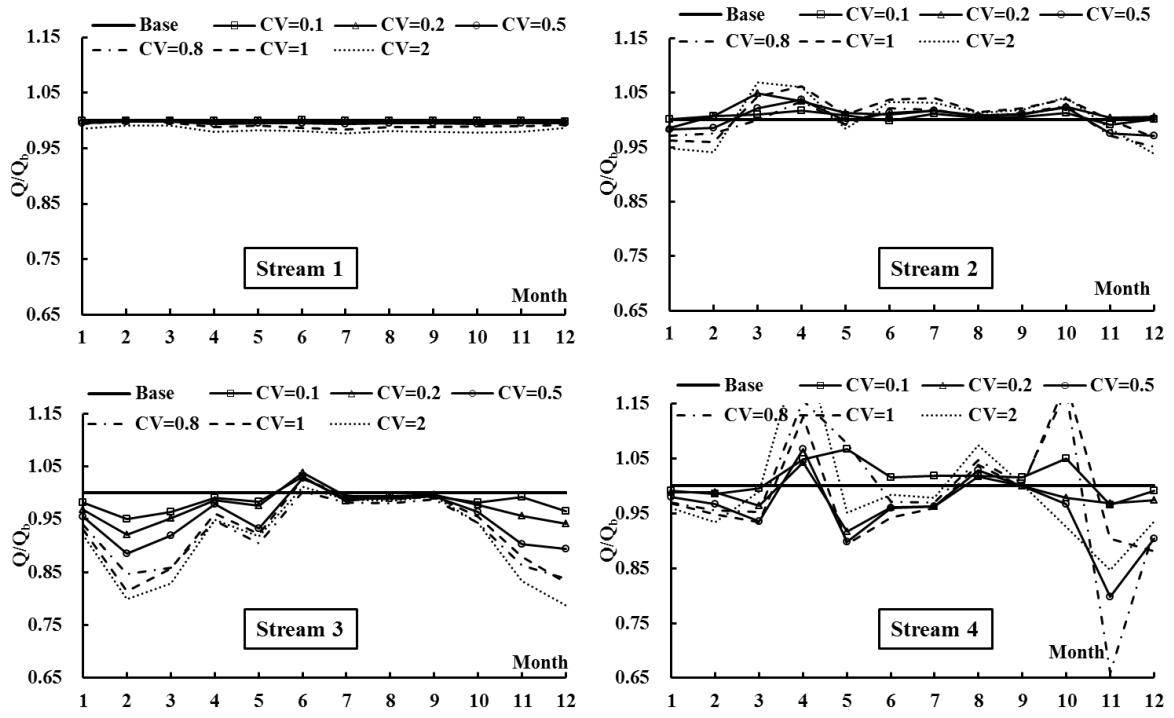
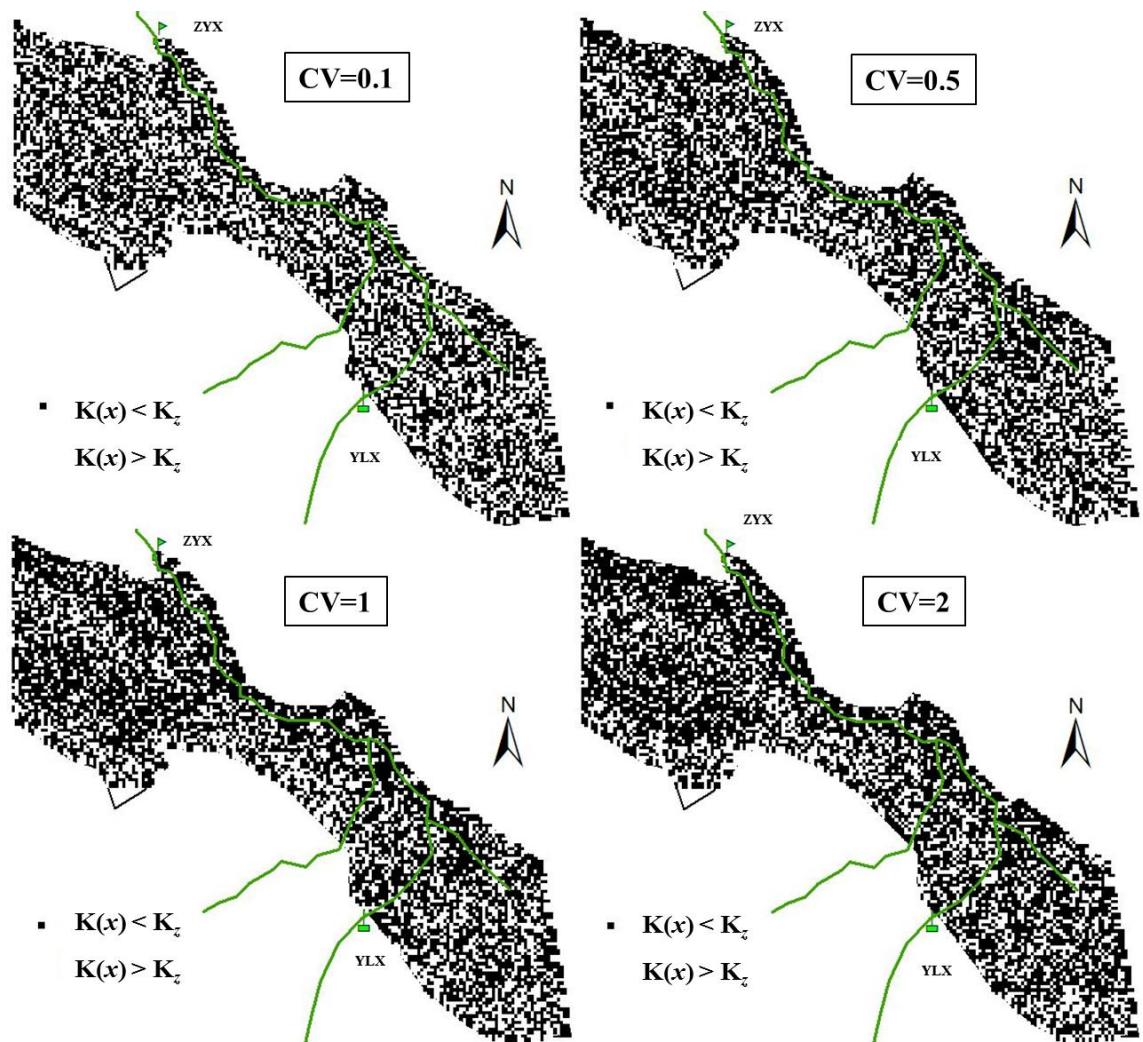


Fig. 8. Stream-aquifer leakages of different streams from refined model



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2 Fig. 9. Grid-scale $K(x)$ field map for CV equaling to 0.1, 0.5, 1 and 2
3

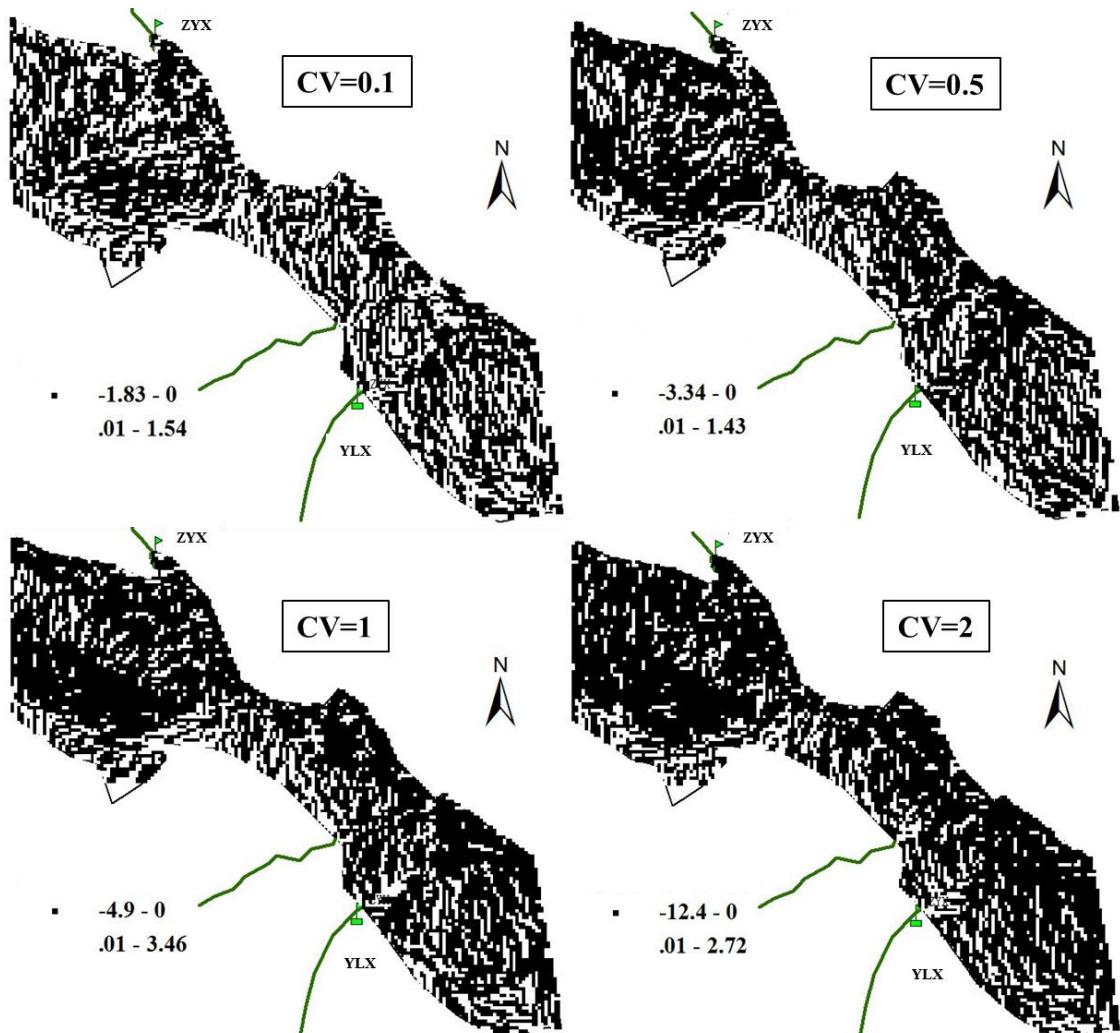


Fig. 10. The field of normalized groundwater flow differences $\Delta q(x)$ with grid-scale $K(x)$

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