

**Technical note: The use of an interrupted-flow
centrifugation method to characterise preferential flow in
low permeability media**

Richard A. Crane,^{1,2*} Mark O. Cuthbert^{2,3} and Wendy Timms^{2,4}

[1] School of Civil and Environmental Engineering, UNSW, Australia

[2] Connected Waters Initiative Research Centre, UNSW, Australia

[3] School of Geography, Earth and Environmental Sciences, University of Birmingham,
United Kingdom

[4] School of Mining Engineering, UNSW, Australia

Correspondence to: R. A. Crane (r.crane@unsw.edu.au)

23 **Abstract**

24 We present an interrupted-flow centrifugation technique to characterise preferential flow in
25 low permeability media. The method entails a minimum of three phases: centrifuge induced
26 flow, no flow and centrifuge induced flow, which may be repeated several times in order to
27 most effectively characterise multi-rate mass transfer behaviour. In addition, the method
28 enables accurate simulation of relevant in situ total stress conditions during flow by selecting
29 an appropriate centrifugal force. We demonstrate the utility of the technique for characterising
30 the hydraulic properties of smectite clay dominated core samples. All core samples exhibited
31 a non-Fickian tracer breakthrough (early tracer arrival), combined with a decrease in tracer
32 concentration immediately after each period of interrupted-flow. This is indicative of dual (or
33 multi) porosity behaviour, with solute migration predominately via advection during induced
34 flow, and via molecular diffusion (between the preferential flow network(s) and the low
35 hydraulic conductivity domain) during interrupted-flow. Tracer breakthrough curves were
36 simulated using a bespoke dual porosity model with excellent agreement between the data and
37 model output (Nash–Sutcliffe model efficiency coefficient was >0.97 for all samples). In
38 combination interrupted-flow centrifuge experiments and dual porosity transport modelling are
39 shown to be a powerful method to characterise preferential flow in low permeability media.

40

41

42

43

44

45

46 **Key words:** dual porosity, interrupted flow, centrifugation, solute transport, molecular
47 diffusion

48

49

50

Introduction

It is well known that heterogeneities, including biogenic pores/channels, desiccation cracks, fissures, fractures, non-uniform particle size distributions and inter-aggregate pores, are widespread in the subsurface and lead to a range of preferential flow phenomena (Beven and Germann, 1982; Cuthbert et al., 2013; Cuthbert and Tindimugaya, 2010; Flury et al., 1994). The coexistence of relatively high hydraulic conductivity (K) domain(s) and an impermeable one, often termed dual porosity, results in a non-Fickian breakthrough curve. Solute transport in such systems is often characterised by an early arrival of solutes originating from the more mobile domain (macropores) and a slow approach to the final concentration caused by diffusion into the immobile domain (matrix or microporous network). When fitting breakthrough curves, therefore, it is often difficult to differentiate between contributions from the micro- and macropore transport mechanisms. As a consequence, in recent years there has been much research into the development of effective empirical and modelling techniques to characterise solute transport processes for dual porosity systems. One method investigated has been the use of interrupted-flow solute break-through experiments. Amongst the original work on this topic Murali and Aylmore, (1980) discussed the influence of non-constant flow on solute transport in aggregated soil. Brusseau et al., (1989) developed a flow-interruption method for use in measuring rate-controlled sorption processes in soil systems, which was subsequently applied by Koch and Fluhler, (1993) to investigate advection and diffusion phenomena occurring for nonreactive solute transport in aggregated media. The idea proposed was that by interrupting flow during nonreactive tracer breakthrough the degree of non-equilibrium between any fast and slow flow pathways can be determined. Central to this hypothesis is that the magnitude of the change in nonreactive tracer concentration in effluent samples taken immediately after a no-flow period is indicative of such non-equilibrium. Subsequent work within this field has included: determination of physical (e.g., diffusive mass transfer between advective and non-advective water) and chemical (e.g., nonlinear sorption) non-equilibrium processes in soil (Brusseau et al., 1997); determination of nonreactive solute exchange between the matrix porosity and preferential flow paths in fractured shale (Reedy et al., 1996); quantifying the effect of aggregate radius on diffusive timescales in dual porosity media (Cote et al., 1999); numerical modelling of aqueous contaminant release in non-equilibrium flow conditions; (Wehrer and Totsche, 2003) empirical modelling of the release of dissolved organic species (Guimont et al., 2005; Ma and Selim, 1996; Totsche et al., 2006; Wehrer and Totsche, 2005; Wehrer and Totsche, 2009) and heavy metals (Buczko et al., 2004); increasing the efficiency

84 of solute leaching (Cote et al., 2000); empirical modelling of conservative tracer transport in a
85 laminated sandstone core sample (Bashar and Tellam, 2006); and characterising in situ aquifer
86 heterogeneity (Gong et al., 2010). One area where comparatively few studies exist, however,
87 is in characterising the hydraulic properties of aquitards (e.g. clay dominated soils and
88 sediments, shales, mudstones). Such research is of particular interest because preferential flow
89 paths, by their intrinsic nature, can significantly compromise the integrity of aquitard units as
90 local and regional barriers to the movement of groundwater contaminants. There are significant
91 technical difficulties at present, however, in characterising such features at appropriate scales
92 (Cuthbert et al., 2010). For example, it is well known that the K of glacial till is scale dependent,
93 with laboratory permeability measurements often yielding values lower than field based
94 measurements and modelling (Cuthbert et al 2010). As a consequence a key requirement of
95 laboratory scale aquitard characterisation is that the core sample must be of sufficient volume
96 in order to incorporate the key dual porosity features which govern the overall formation. A
97 second technical challenge is that laboratory testing typically requires generation of flow
98 through the sample whilst maintaining relevant in situ hydro-geotechnical conditions. One
99 method which has been demonstrated as effective for this purpose is centrifugation, which is
100 increasingly being used for hydraulic and geotechnical testing of low K materials (Hensley and
101 Schofield., 1991; Nimmo and Mello., 1991; Timms et al., 2009; Timms and Hendry., 2008).
102 Moreover, experiments using geotechnical centrifuges with payload capacities exceeding
103 several kilograms can provide the additional benefit of being able to use core samples of
104 representative scale for the overall formation. Here we present, for the first time, an interrupted-
105 flow methodology using a centrifuge permeameter (CP) to characterise possible dual porosity
106 behaviour of low permeability porous media. A novel dual domain model is also described
107 which has been used to guide physical interpretation of the experimental tracer breakthrough
108 curves.

109 2. Experimental methods

110 2.1. Core and groundwater sampling methodology

111 The clay core (101.6 mm in diameter, Treifus core barrel, non-standard C size) and
112 groundwater were sourced from a 40 m thick, semi-consolidated, clay-rich alluvium deposit
113 located approximately 100 km south of Gunnedah, New South Wales, Australia (31° 31'9"S,
114 150° 28'7"E). Equipment and procedures for obtaining minimally disturbed cores were
115 compliant with ASTM (2012). See Timms et al., (2014) for a review of the procedure.
116 Groundwater samples were taken from piezometers using standard groundwater quality

sampling techniques (Sundaram et al., 2009). A 240V electric submersible pump (GRUNDFOS MP1) and a surface flow cell were used to obtain representative samples after purging stagnant water to achieve constant field measurements of electrical conductivity, pH, dissolved oxygen (DO) and reduction potential (Eh).

2.2. Centrifuge permeameter theory

During centrifugation increased centrifugal force generates a body force which accelerates both solid and fluid phases within the sample. Centrifugal acceleration at any point within a centrifuge sample is calculated as follows:

$$a = \omega^2 r \quad \text{Eq. (1)}$$

Where a is the centrifugal acceleration (m/s^2), ω is the angular velocity (radian/s), and r is the radius from the axis of rotation (m). The g -level is the scaling factor (a/g) for accelerated gravity, where g is gravity at the Earth's surface.

Vertical hydraulic conductivity, K_v (m/s) is calculated using ASTM (2000) (Eq. 2), where: Q = the steady-state fluid flux (mL/h); A = the sample flow area (cm^2); r_m = the radial distance at the mid-point of the core sample (cm); and RPM = revolutions per minute.

$$K_v = \frac{0.248Q}{Ar_m(RPM)^2} \quad \text{Eq. (2)}$$

The estimated in situ stress applied at the base of the core samples was calculated according to Eq. 3, and assumes that the overlaying formations were fully saturated and of a similar density to the core samples.

$$\sigma_i = \rho_s dg \quad \text{Eq. (3)}$$

Where σ_i = in situ stress (kPa); ρ_s = saturated density of core (kg/m^3); d = depth to the base of the core sample (m BGL); and g = gravitational acceleration (m/s^2). The applied stress at the base of the core (σ_g , kPa) during the centrifuge experiments was calculated according to Eq. 4 (Timms et al. 2014).

$$\sigma_g = [(\rho_b L_c) + \rho_w(L_c + h_w)]a_b \quad \text{Eq. (4)}$$

where ρ_b = core bulk density (kg/m³); L_c = length of CP core specimen (mm); ρ_w = influent density (kg/m³); h_w = height of influent water above CP core specimen (mm); and a_b is the centrifugal acceleration at the base of the CP core specimen.

2.3. Centrifuge permeameter sample preparation

A Broadbent geotechnical centrifuge (GMT GT 18/0.7 F) with a custom built permeameter module (Timms et al., 2014) was used for this study. Prior to mounting into the CP the outer 5 mm of the clay cores were trimmed and the trimmed cores were then inserted into Teflon cylindrical core holders (100 mm internal diameter, 220 mm length) using a custom built mechanical cutting and loading device. The cores were trimmed in order to remove any physical and chemical disturbance associated with the core extraction (drilling) process. A 5 mm thick A14 Geofabrics Bidim geofabric filter (100 micron, $K = 33$ m/s) was placed above and below the sample in order to prevent clogging of the effluent drainage plate with colloid material from the sample. The geofabric filter was held in position above the sample using a plastic clamp.

The core holders (with the core sample held within) were placed into 3000 mL glass beakers containing 1000 mL of groundwater derived from the piezometer at the closest depth to the core sample (see Table 1) and allowed to saturate from the base upwards. In total three core samples were analysed, which were taken from depths of 5.03, 9.52 and 21.75 m BGL. Saturation was performed by immersing the core holder into a reservoir of groundwater with the level of the water 5 cm higher than the top of the core sample. The mass of each core was then monitored every 24 hours until no further increase in mass was recorded, saturation was then assumed to have occurred. The core holders (containing the saturated core samples) were mounted to the CP system via double O-ring seals. An influent head was added to all samples (see Table 1), which was maintained during centrifugation by a custom built automated influent level monitoring and pumping system. The system comprises a carbon fibre EC electrode array which is connected via a fibre optic rotary joint to a peristaltic pump that supplies influent from an external 100 mL burette. Effluent samples were collected in an effluent reservoir and extracted using a 50 mL syringe. All experiments were conducted under steady-state flow,

175 which is defined as a <10% difference between influent and effluent flow rates. The influent
176 volume was determined by manual measurements of the water level in the external burette and
177 effluent volumes were measured by multiplying their mass by their density.

178 **2.4. Interrupted-flow experiment methodology**

179 The idea of interrupting the flow during a breakthrough experiment is to differentiate between
180 advection and diffusion processes. The method comprises a minimum of three phases:

- 181 1. Flow is induced at a constant centrifugal force for a fixed time period with effluent
182 samples collected at multiple periodic intervals. The *g*-level and influent reservoir
183 height are selected so that the maximum total stress on the core approached the
184 estimated in situ stress of the material at the given depth in the formation (Eq. 3 and 4).
185 The time period between each effluent sampling interval is selected in order to gain
186 sufficient effluent volume (namely >1 mL) for accurate volume and nonreactive tracer
187 concentration measurement.
- 188 2. Flow is interrupted (stopped) for a fixed time period during which time the
189 permeameters are disconnected from the centrifuge module and positioned upright, the
190 influent reservoir is also removed to limit any downward migration of solutes. A
191 relatively long interrupted-flow period (>12 hrs) is selected so that slow mass transfer
192 processes can be identified.
- 193 3. Phase 1 is then repeated.

194 All phases can be repeated multiple times in order to record sufficient non-reactive tracer
195 breakthrough which enables the mass transport behaviour to be accurately characterised.
196 Deuterium oxide (D₂O) (Acros Organics, 99.8% concentration) was used as a non-reactive
197 tracer. A concentration of 3.12 mL/L was used, which raised the concentration of D₂O to
198 approximately 200%. This was selected as sufficiently high in order to result in accurately
199 measureable mass transfer changes. Effluent samples were filtered using a 0.2 µm cellulose
200 acetate filter, stored at 4 °C and analysed for δD within 7 days of testing. δD was determined
201 by measuring the ¹H/²H ratio to an accuracy of 0.1% using a Los Gatos DLT100 isotope
202 analyser.

203 **2.5. Dual domain transport modelling**

204 Dual porosity models were created using COMSOL Multiphysics v. 4.4
205 (<http://www.comsol.com>) modified from well-known formulations described, for example, by

Coats and Smith, (1964) and Bear (1987). The purpose of the modelling was to aid physical interpretation of the tracer breakthrough curves and validate the hypothesis that the step changes in tracer concentrations observed during no-flow periods could be explained by the presence of dual porosity in the samples. The models comprised a classical advection-dispersion equation for a mobile zone (subscript m) representing preferential flow pathways with a source/sink term representing exchange of solute with an immobile zone (subscript im). Solute transport in the immobile zone was by diffusion only. The exchanged flux between the immobile and mobile zones was modelled as being proportional to the concentration difference between the zones. The governing equations are as follows:

$$\frac{\partial C_m}{\partial t} = D_m \frac{\partial^2 C_m}{\partial z^2} - \frac{q(t)}{\phi_m} \frac{\partial C_m}{\partial z} - \frac{\gamma}{\phi_m} (C_m - C_{im}) \quad (\text{Eq. 5})$$

$$\frac{\partial C_{im}}{\partial t} = \mu \frac{\partial^2 C_{im}}{\partial z^2} + \frac{\gamma}{\phi_{im}} (C_m - C_{im}) \quad (\text{Eq. 6})$$

$$D_m = \frac{\alpha q(t)}{\phi_m} + \mu \quad (\text{Eq. 7})$$

where C is the δD isotope ratio [1], t is time [T], z is distance along the column [L], q is fluid flux [$L T^{-1}$], α is hydrodynamic dispersivity [L], μ is the coefficient of molecular diffusion [$L^2 T^{-1}$]. The porosity, ϕ , of the mobile and immobile domain is defined as:

$$\phi_m = \frac{V_{p,m}}{V_T} \quad (\text{Eq. 8})$$

$$\phi_{im} = \frac{V_{p,im}}{V_T} \quad (\text{Eq. 9})$$

where $V_{p,m}$ is the pore volume of the mobile domain [L], $V_{p,im}$ is the pore volume of the immobile domain [L] and V_T is the total volume of the saturated core [L]. The mass transfer coefficient, γ [T^{-1}], is defined as:

$$\gamma = \frac{\beta \phi_m \mu}{a^2} \quad (\text{Eq. 10})$$

where β is the dimensionless geometry coefficient, which typically ranges from 3 for rectangular slabs to 15 for spherical aggregates, and a is the characteristic half width of the matrix block [L] (Gerke and van Genuchten, 1993).

The initial concentration conditions were set to zero for both domains for all model runs. During centrifugation periods, a variable solute flux upper boundary condition was used for

the mobile domain varied according to the product of the measured fluid flux & input concentration (C_0) during each experiment as follows:

$$\frac{q(t)}{\phi_m} C_0 = \frac{q(t)}{\phi_m} C_m + D_m \frac{\partial C_m}{\partial z} \quad (\text{Eq. 11})$$

A Dirichlet (constant concentration) upper boundary condition was used for the immobile domain during times of centrifugation. A novel aspect of the models, facilitated by the flexibility of model structure variations possible in COMSOL Multiphysics, was that the upstream transport boundary for both domains was switched to a zero flux condition during the interrupted flow phases. The downstream transport boundary conditions for both domains were given by:

$$\frac{\partial C_{m,im}}{\partial z} = 0 \quad (\text{Eq. 12})$$

At $z = L_c$, where L_c was sufficiently large to ensure the results at the column outlet distance (at $z = L_c$, $L_c \ll L_{DT}$) were not sensitive to the position of the boundary. The total mass flux at the distance from the upstream boundary corresponding with the length of the experimental column was output from the models and integrated over the sampling periods for comparisons to the observed breakthrough curves. μ was calculated as $3.43 \times 10^{-5} \text{ m}^2/\text{d}$ which is the diffusion coefficient of D_2O in H_2O at 25.0°C (Orr and Butler, 1935) multiplied by the average tortuosity of 0.15 reported by (Barnes and Allison, 1988) for clay bearing media. Model output was fitted to the observed data by varying the unconstrained parameters: α and γ . Note that ϕ_m and ϕ_{im} were also considered unconstrained parameters, but their sum was constrained to equal the total ϕ measured for each sample by oven drying at 105°C for 24 hours. In order to quantify the deviation between the recorded data and the dual porosity model the normalised root-mean-square error (NRMSE) and the Nash–Sutcliffe model efficiency coefficient (NSMEC) were calculated (Nash and Sutcliffe, 1970). The mesh size and model tolerance were set sufficiently small so that the results were no longer sensitive to further reduction to ensure the accuracy of the model output. The models runs presented were all executed using an extra fine mesh size and a relative tolerance of 0.00001.

2.6. Dual domain model sensitivity testing

Sensitivity analysis of the dual domain model (for the core taken from 5.03 m) was conducted in order to determine how sensitive the model was to changes in the constrained (L_c , ϕ and μ) and unconstrained (ϕ_m , α and γ) parameters. Sensitivity factors for constrained parameters were

Formatted: Subscript

Formatted: Font: Italic

Formatted: Font: Italic, Subscript

Formatted: Subscript

determined according to the estimated percentage error associated with each parameter, whilst $\pm 50\%$ was selected for the unconstrained parameters in order to determine their influence on the NSMEC. The percentage error for L_c was calculated to be $\pm 2.78\%$ due to the core length being 36 mm and the error associated with measurement at each end was ± 0.5 mm. The percentage error for ϕ was calculated to be $\pm 2.79\%$ which comprises the L_c measurement error plus 0.0026% which is the calculated error associated with the two mass measurements. The percentage error for μ was determined to be $\pm 50\%$ due to the range in tortuosity of 0.1 – 0.2 documented by Barnes and Allison, (1988) and references therein.

3. Results and discussion

3.1. D₂O breakthrough

D₂O breakthrough data and best fit dual porosity model output for the interrupted-flow experiments conducted using core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL are displayed in Fig. 1. A close fit was achieved between the dual porosity model output and the original data, with a NSMEC of 0.97, 0.99 and 0.97 and a NRMSE of 5%, 3% and 5% recorded for D₂O breakthrough data from core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively. The D₂O breakthrough curves for all core samples exhibited a relatively elongated shape, with 100% breakthrough not recorded for any of the timescales tested. This was expected given that a ‘long tailing’ is a common feature of dual (or multi) porosity materials, i.e. systems where the mobile domain is coupled to a less mobile, or immobile, domain. In such instances the dominant solute transport mechanism during imposed flow in the mobile domain(s) is typically advection, however, solute exchange also occurs in parallel with the immobile domain(s), typically via molecular diffusion. Following each interrupted-flow (no flow) period a decrease in δD was recorded for all samples, and attributed to the diffusion of D₂O from the preferential flow domain(s) into the low-flow (or immobile) flow domain(s). The shape of the D₂O breakthrough curves and the magnitude of the δD decrease following the interrupted-flow periods are different for all samples, with a 42.6%, 18.5% and 28.4% decrease recorded for the core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively after the first interrupted-flow period. In addition, the K_v of each sample was recorded as different (Fig. 2), with average values of 1.4×10^{-8} m/s, 3.9×10^{-9} m/s and 2.7×10^{-9} m/s for the core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively. The K_v was recorded to decrease during the initial stages of each centrifugation period, and attributed to the partial consolidation of the clay due to the stress applied by the

Formatted: Subscript

Formatted: Subscript

centrifugal force. Following this initial consolidation period a more constant K_v as a function of time was recorded for all cores, indicating that relative equilibrium had been achieved between stress applied by the centrifugal force and the compaction state of the core.

297

Core depth (m BGL)	Estimated in situ total stress, σ_i (kPa)	Influent groundwater depth (m BGL)	Influent EC ($\mu\text{S}/\text{cm}$)	g-level applied	Core length, L_c (mm)	Height of influent water above core, h_c (mm)	K_v (ms^{-1})	Total stress at base of core during centrifugation, σ_g (kPa)
5.03	89	10	18470	20	36	61	1.4×10^{-8}	75
9.52	177	10	18470	20	47	81	3.9×10^{-9}	127
21.75	383	20	13160	80	54	48	5.1×10^{-9}	373

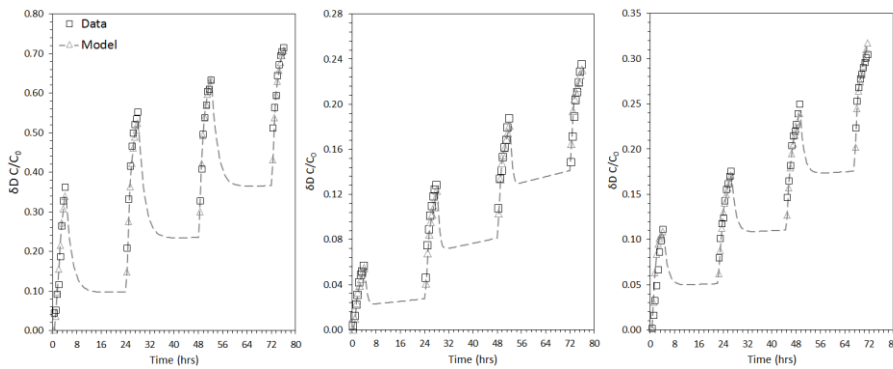
Table 1. Core and influent properties, experimental parameters and K_v results for the interrupted-flow experiments. Calculations are based on Eq. 2 for K_v , Eq. 3 for estimated in situ total stress and Eq. 4 for total stress at the base of core specimen during centrifugation.

3.2. Dual domain model

The close model fits confirm that preferential flow through a dual porosity structure is a plausible hypothesis to explain the shape of the observed breakthrough curves. The unconstrained (ϕ_m , α and γ) parameters that yielded the best dual domain model output fit to the D₂O breakthrough data are displayed in Table 2. It is noted that the pore volume of the mobile domain per total volume of the core, ϕ_m , was modelled to be 0.04, 0.04 and 0.08 for core taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively. With total porosity, ϕ , measured as 0.44, 0.47 and 0.43, this equates to 9.1%, 8.5% and 18.6% of the total pore volume respectively, suggesting that preferential flow features comprise a relatively large proportion of the total pore porosity in each sample. Hydrodynamic dispersivity, α , for best fit model output for all core samples was $L_c/2$, which is larger than typically reported for laboratory scale column experiments (e.g. Shukla et al., 2003). It can be noted that all of the core samples were assumed to have remained saturated throughout the breakthrough experiments because all

Formatted: Subscript

314 influent and effluent flow rates were recorded at steady-state. Whilst dispersion is known to
 315 increase substantially as moisture content decreases from saturation (e.g. Wilson and Gelhar,
 316 1981) it is therefore unlikely that this could have been a factor. The mass transfer coefficient,
 317 γ , was also modelled as different for each core sample with 0.65, 1.50 and 1.20 yielding the
 318 best model fit for the core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL
 319 respectively. Using Eq. 10 the half width of the matrix block (using a β range of 3-15 (3 for
 320 parallel slabs and 15 for spherical aggregates after Gerke and van Genuchten, (1993)), a , is
 321 calculated as within the range of 8.0-17.8 mm, 5.4-12.1 mm and 5.5-12.3 mm for the core
 322 samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively. This suggests that
 323 the preferential flow channels present are likely to be separated by distances in the order of
 324 several mm from each other within the cores. With the dimensions of the cores significantly
 325 greater than these values the model output therefore suggests that several preferential flow
 326 features are present in each core sample.



327
 328 Figure 1. Normalised D₂O breakthrough data along with best fit dual porosity model output for
 329 the interrupted-flow experiments conducted using core samples taken from 5.03 m (left), 9.52
 330 m (middle) and 21.75 m (right) depth BGL.

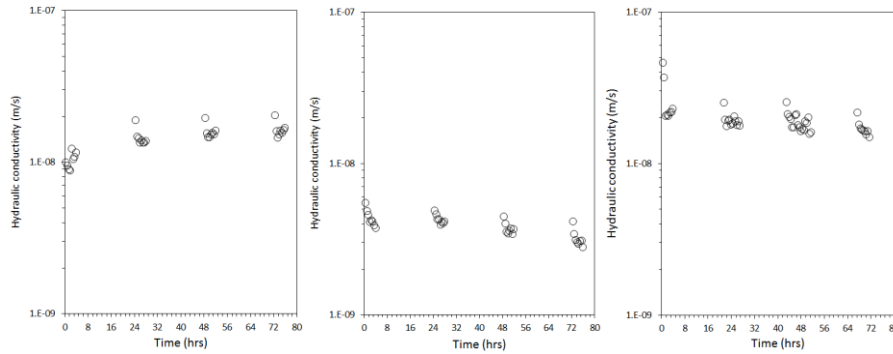


Figure 2. Vertical hydraulic conductivity (m/s), calculated using Eq. 2, for the interrupted-flow experiments conducted using core samples taken from 5.03 m (left), 9.52 m (middle) and 21.75 m (right) depth BGL.

Core depth (m BGL)	Core diameter, D (mm)	Core length, L_c (mm)	Total porosity, ϕ	Pore volume of the mobile domain per total core volume, ϕ_m	Coefficient of molecular diffusion, μ [L^2T^{-1}]	Hydrodynamic dispersivity, α [L]	Mass transfer coefficient, γ [T^{-1}]	Half width of the matrix block, a (mm)
5.03	100	36	0.44	0.06	3.43×10^{-5}	$L_c/2$	0.65	18.3
9.52	100	47	0.47	0.04	3.43×10^{-5}	$L_c/2$	1.50	12.1
21.75	100	55	0.43	0.08	3.43×10^{-5}	$L_c/2$	1.20	12.3

Table 2. Constrained (D , L_c , ϕ , μ) and unconstrained (ϕ_m , α and γ) model parameters. a is calculated using Eq. 10.

Model output for the mobile and immobile domains at the top, middle and base of the core samples is displayed in Fig. 3. It is noted that for all core samples diffusion into the immobile domain during the induced flow periods is relatively significant, with $\delta D_{im}/\delta D_m$ at the end of the first centrifugation (induced flow) period recorded as 0.16, 0.32 and 0.34 for the base of the core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively. With respective average flow rates recorded as 0.017 m/d, 0.007 m/d and 0.015 m/d this behaviour

is not obviously related to the variation in flow rates between the samples, but more likely to the intrinsic properties of the preferential flow domain (namely: ϕ_m , γ and α). It is also noted that for all core samples full equilibration between the mobile and immobile domains occurred ($\delta D_{im} = \delta D_m$) during each no flow period. For example, δD_{im} and δD_m were modelled to be within $\pm 1\%$ of each other after 7.0, 2.6 and 6.1 hrs during the first no flow period for the core samples taken from 5.03 m, 9.52 m and 21.75 m depth BGL respectively.

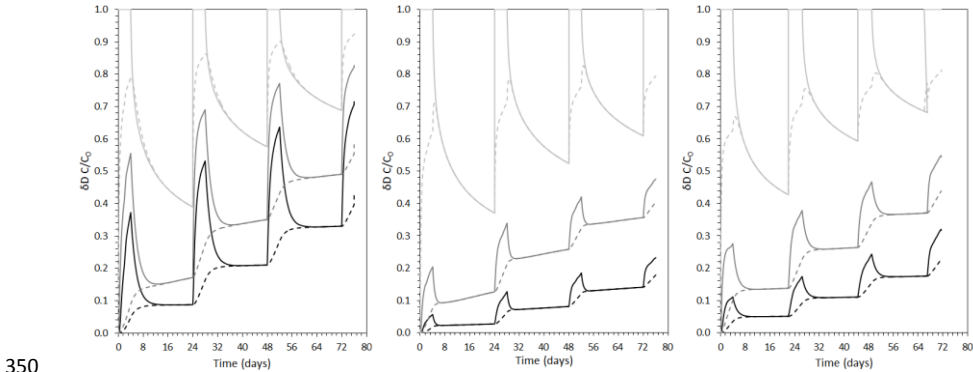


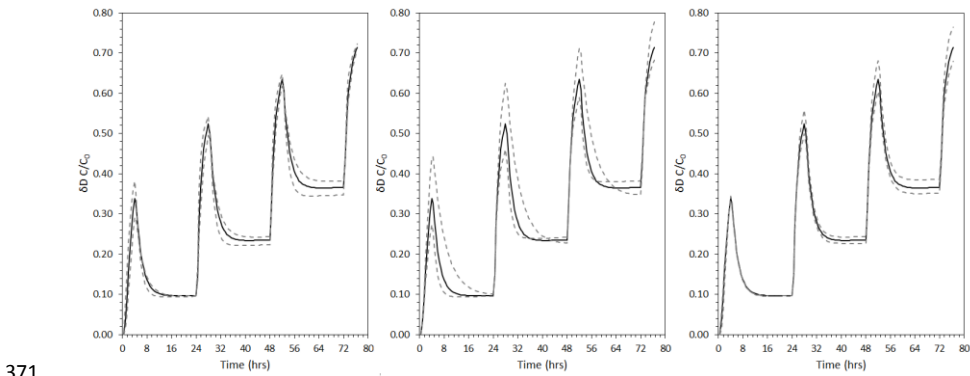
Figure 3. Model output for mobile (solid lines) and immobile (dashed lines) domains for core samples taken from 5.03 m (left), 9.52 m (middle) and 21.75 m (right) depth BGL. The black, dark grey and light grey lines comprise model output for the base, middle and top of the cores respectively.

3.3. Sensitivity analysis

Sensitivity analysis plots for a $\pm 50\%$ change in unconstrained parameters (α , γ and ϕ_m) for the core sample taken from 5.03 m depth BGL are displayed in Fig. 4, with corresponding NSMEC data displayed in Table 3. The model fitting efficiency is relatively insensitive to all three unconstrained parameters in the range tested, with a less than 12% change in the NSMEC compared to the NSMEC recorded for the best fit (Table 3). Sensitivity for the estimated % error associated with constrained parameters (ϕ , L_e and μ) are displayed in Fig. 5, with corresponding NSMEC data displayed in Table 3. The model fitting efficiency is also relatively insensitive, with a less than 1% change in the NSMEC compared to the NSMEC recorded for the best fit (Table 3). For the data presented the relatively low sensitivity to the parameters indicates that further testing, such as by dye tracing or geophysical tomography, is necessary to resolve more precisely the nature of the preferential ~~flow paths~~ flow paths. Nevertheless, the

Formatted: Subscript

367 modelling has supported the preferential flow conceptual model we have used to explain the
368 step changes in concentration observed after resting periods. It has also provided a first order
369 approximation of the likely geometry of the ~~flowpaths~~flow paths.
370



371
372 Figure 4. Sensitivity of the dual domain model for the core sample taken from 5.03 m depth
373 BGL due to $\pm 50\%$ change in unconstrained parameters: ϕ_m (LHS), γ (middle) and α (RHS).

Model Parameter	Pore volume of the mobile domain per total pore volume, ϕ_m	Mass transfer coefficient, γ	Hydrodyn amic dispersivity, α	Total porosity, ϕ	Core length, L_c	Coefficien t of molecular diffusion, μ
NSMEC (+ change)	0.925	0.926	0.952	0.974	0.965	0.964
NSMEC (- change)	0.952	0.862	0.964	0.968	0.971	0.975

374
375 Table 3. NSMEC for the core sample taken from 5.03 m depth BGL due to changes in
376 constrained (L_c , ϕ , μ) and unconstrained (ϕ_m , α and γ) model parameters. Changes in
377 constrained parameters comprised the estimated percentage error per each parameter, which
378 was 2.78%, 2.79% and 50% for L_c , ϕ and μ respectively. Changes in unconstrained parameters
379 were $\pm 50\%$. The NSMEC for the best fit was 0.972.

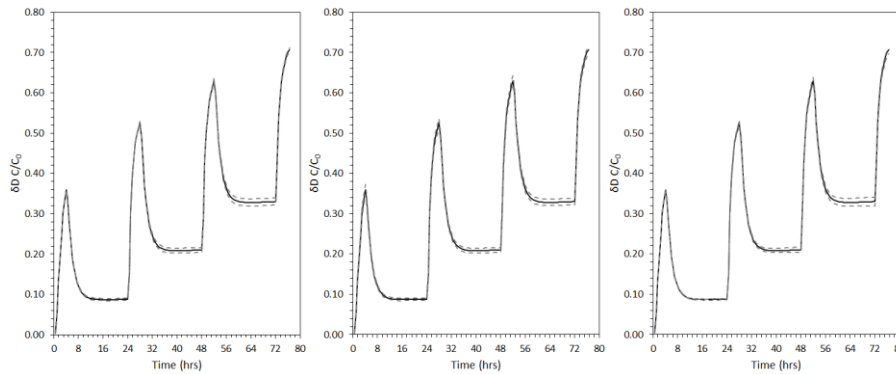


Figure 5. Sensitivity of the dual domain model for the core sample taken from 5.03 m depth BGL for the calculated error associated with the constrained parameters: ϕ (LHS); L_e (middle) and μ (RHS).

3.4 Comparison of dual and single domain modelling

In order to further demonstrate the practicality of the interrupted flow methodology, a numerical experiment was carried out using the dual domain model developed above. Using the best fit parameters from the core from 9.52 m depth BGL, an equivalent simulation to the laboratory experiment described above was run, but *without* interrupted flow phases. The breakthrough curve produced was then fit to the Ogata-Banks equation (Ogata and Banks, 1961) on the assumption that flow was occurring only through a single domain. The resulting fitted was good (NRMSE = 3%) with just one fitting parameter being the dispersion term which yielded a reasonable value of $1.27 \times 10^{-8} \text{ m}^2/\text{s}$. This illustrates that, without the use of interrupted flow phases to reveal the disequilibrium between two or more flow domains, a false assumption could easily be made with regard to the structure and associated transport properties of the core on the basis of a simple 1-D analytical model. This could have very significant consequences for the prediction and management of solute migration through such deposits. This could have very significant consequences for the prediction and management of solute migration through such deposits.

An additional numerical experiment was also undertaken to attempt to match the observed data to a single domain model which included resting phases, since no analytical solution is known for such a simulation. This was accomplished using COMSOL Multiphysics with identical

Formatted: Subscript

settings to the dual domain models described above, but with a disabled immobile domain. Calibrating to the δD breakthrough data recorded for the core from 9.52 m depth BGL by just varying dispersivity, but using the measured porosity, we were unable to achieve a better fit than a NRMSE of 46%, even with an unrealistically high dispersivity. A better fit is possible (NRMSE = 9%, NSMEC = 0.9) if porosity is decreased to 0.1 but, again, only with an unrealistically high value for dispersivity of $1000L_c$, see Figure 6. While such a model may be useful to suggest that the effective porosity of the core through which solute is moving is much less than the total porosity, it is only possible to fit the early time data (e.g. only the first flow stage) very accurately, at the expense of the later time data. Perhaps more importantly than the lower NSMEC (or higher NRMSE) compared to the dual domain models, the single domain model also misses a key feature of the observed breakthrough curves – the decrease in concentration during resting phases. Instead, modelled concentrations increase during resting phases as would be expected in a single domain model due to redistribution of the solute along the core by diffusion. This additional numerical experiment thus strengthens the conclusions of the study, that dual domain behaviour is indicated by our interrupted flow experiment observations, and that single domain models are inappropriate as a means of analysis.

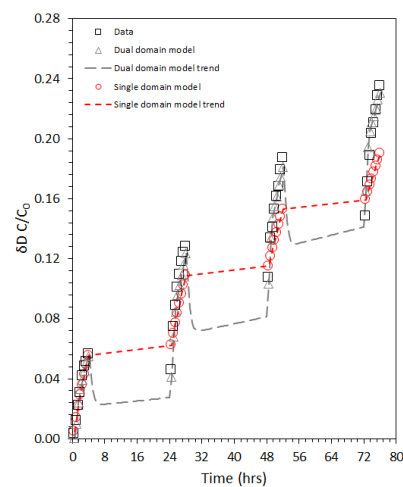


Figure 6. Comparison of single and dual domain interrupted flow transport model output best-fit simulations for the core taken from 9.52 m BGL.

4. Conclusions and outlook

Formatted: Font: Italic

Formatted: Font: Italic, Subscript

Formatted: Font: Times New Roman, 12 pt

Solute transport in the subsurface can be influenced by multiple nonlinear, rate-limited processes, and it is often difficult to determine which processes predominate for any given system. In this work we demonstrate the utility of interrupted-flow solute transport experiments using a [CP-centrifuge permeameter](#) to quantify the relative contributions of preferential flow pathways and surrounding matrix porosity to mass transfer processes in low permeability dual porosity materials. Dual domain transport modelling was used to validate the hypothesis that the step changes in tracer concentrations observed during no-flow periods could be explained by the presence of dual porosity in the samples. The modelling also enabled a first order approximation of the physical properties of the two domains to be inferred. Smectite clay core samples were used (101.6 mm in diameter) as an example low K dual porosity media, however, it is anticipated that the methodology would also be suitable for the characterisation of any dual porosity material where mass transfer occurs via both advection and diffusion (e.g. fractured rock, heterogeneous soils, mine tailings). The methodology entails a minimum of three phases: induced flow, no flow, and induced flow, however, this may be repeated several times in order to most effectively characterise the multi-rate mass transfer behaviour. In addition, it is necessary to tailor the induced flow rate, interrupted-flow timescales and non-reactive tracer concentrations in order to most effectively identify different mass transfer processes whilst also simulating realistic total stress conditions. Future work will seek to further investigate the structure of the clay samples studied using quantitative tomography techniques (e.g. X-ray computed tomography and magnetic resonance imaging) and how these physical features can be integrated into site scale numerical flow modelling.

Acknowledgements

The authors would like to thank Dayna McGeeney (School of Civil and Environmental Engineering) and Mark Whelan (School of Mining Engineering) from the Connected Waters Initiative Research Centre, UNSW, Australia, for their technical support. The work was financially supported by Program 1B of the National Centre for Groundwater Research and Training, supported by the Australian Research Council and the National Water Commission, and the Gary Johnson Trust. Dr Mark Cuthbert was financially supported by the European Community's Seventh Framework Programme [FP7/2007–2013] under grant agreement No.299091.

References

- ASTM. Standard practice for thin-walled tube sampling of soils for geotechnical purposes. ASTM D1587-08(2012)e1, West Conshohocken, PA, 2012.
- ASTM. Standard test method for determining unsaturated and saturated hydraulic conductivity in porous media by steady-state centrifugation. ASTM D6527, West Conshohocken, PA, 2000.
- Barnes, C. J., Allison, G. B.: Tracing of water movement in the unsaturated zone using stable isotopes of hydrogen and oxygen. *Journal of Hydrology*, 100, 143-176, 1988.
- Bashar, K., Tellam, J. H.: Non-reactive solute movement through saturated laboratory samples of undisturbed stratified sandstone. The Geological Society of London Special Publications., 263, 233-251, 2006.
- Bear, J. Bachmat, Y. Introduction to *Modeling of Transport Phenomena in Porous Media*. Springer Science & Business Media. ISBN: 0792305574, 1990.
- Beven, K., Germann, P.: Macropores and water flow in soils. *Water Resources Research*, 18, 1311-1325, 1982.
- Brusseau, M. L., Hu, Q., Srivastava, R.: Using flow interruption to identify factors causing nonideal contaminant transport. *Journal of Contaminant Hydrology*, 24, 205-219, 1997.
- Brusseau, M. L., Rao, P. S. C., Jessup, R. E. and Davidson, J. M.: Flow interruption: a method for investigating sorption nonequilibrium. *J. Contam. Hydrol.*, 4, 223-240, 1989.
- Buczko, U., Hopp, L., Berger, W., Durner, W., Peiffer, S., Scheithauer, M.: Simulation of chromium transport in the unsaturated zone for predicting contaminant entries into the groundwater. *Journal of Plant Nutrition and Soil Science*, 167, 284-292, 2004.
- Coats, K. H., Smith, B. D.: Dead-end pore volume and dispersion in porous media. *Soc. Pet. Eng. J.*, 4, 73-84, 1964.
- Cote, C. M., Bristow, K. L., Ross, P. J.: Increasing the efficiency of solute leaching: impacts of flow interruption with drainage of the "preferential flow paths." *Journal of Contaminant Hydrology*, 43, 191-209, 2000.
- Cote, C. M., Bristow, K. L., Ross, P. J.: Quantifying the influence of intra-aggregate concentration gradients on solute transport. *Soil Science Society of America Journal*, 63, 757-767, 1999.
- Cuthbert, M. O., Mackay, R., Tellam, J. H. Thatcher, K. E.: Combining unsaturated and saturated hydraulic observations to understand and estimate groundwater recharge through glacial till. *Journal of Hydrology*, 391(3), 263-276, 2010.
- Cuthbert, M.O, and C Tindimugaya.: The importance of preferential flow in controlling groundwater recharge in tropical Africa and implications for modelling the impact of climate

change on groundwater resources. *Journal of Water and Climate Change*, 1(4), 234-245, 2010.

Cuthbert, M.O., R Mackay, JR Nimmo.: Linking soil moisture balance and source-responsive models to estimate diffuse and preferential components of groundwater recharge. *Hydrology and Earth System Sciences*, 17(3), 1003-1019, 2013.

Flury, M., Fluhler, H., Jury, W. A., Leuenberger, J.: Susceptibility of soils to preferential flow of water: a field study. *Water Resources Research*, 30, 1945-1954, 1994.

Gerke, H. H., van Genuchten, M. T.: Evaluation of a first-order water transfer term for variably saturated dual-porosity flow models. *Water Resources Research*, 29(4), 1225-1238, 1993.

Gong, R., Lu, C., Wu, W.-M. Cheng, H., Gu, B., Watson, D.B., Criddle, C.S., Kitanidis, P.K., Brooks, S.C., Jardine, P.M. Luo, J.: Estimating kinetic mass transfer by resting-period measurements in flow-interruption tracer tests. *Journal of Contaminant Hydrology*, 117, 37-45, 2010.

Guimont, S., Perrin-ganier, C., Real, B., Schiavon, M.: Effects of soil moisture and treatment volume on bentazon mobility in soil. *Agronomy for Sustainable Development*, 25, 323-329, 2005.

Hensley, P. J., and Schofield, A. N.: Accelerated physical modelling of hazardous-waste transport. *Geotechnique*, 41(3), 447-465, 1991.

Nash, J. E., Sutcliffe, J. V.: River flow forecasting through conceptual models part I — A discussion of principles., *Journal of Hydrology*, 10(3), 282-290, 1970.

Koch, S., Fluhler, H.: Non-reactive solute transport with micropore diffusion in aggregated porous media determined by a flow-interruption method. *Journal of Contaminant Hydrology*, 14, 39-54, 1993.

Ma, L., Selim, H. M.: Solute transport in soils under conditions of variable flow velocities. *Water Resources Research*, 32, 3277-3283, 1996.

Murali, V. and Aylmore, L. A. G.: No-flow equilibrium and adsorption dynamics during ionic transport in soils. *Nature (London)*, 283, 467-452, 1980.

Nimmo, J. R., and K. A. Mello.: Centrifugal techniques for measuring saturated hydraulic conductivity. *Water Resources Research*, 27(6), 1263-1269, 1991.

Ogata, A., Banks, R. B.: A solution of the differential equation of longitudinal dispersion in porous media. *Fluid Movement in Earth Materials*, Geological Survey Professional Paper., 411-A. <http://pubs.usgs.gov/pp/0411a/report.pdf>

Orr, W. J. C., Butler, J. A. V.: The rate of diffusion of deuterium hydroxide in water. *J. Chem. Soc.*, 303, 1273-1277, 1935.

526 Reedy, O. C., Jardine, P. M., Wilson, G. V. and Selim, H. M.: Quantifying the diffusive mass
527 transfer of nonreactive solutes in columns of fractured saprolite using flow interruption. *Soil*
528 *Sci. Soc. Am.*, 60(5), 1376–1384, 1996.

529 Shukla, M. K., Lal, R., Owens, L. B. and Unkefer, P.: Land use and management impacts on
530 structure and infiltration characteristics of soils in the North Appalachian region of Ohio. *Soil*
531 *Science*, 168(3), 167-177, 2003.

532 Sundaram, B., Feitz, A., Caritat, P. de, Plazinska, A., Brodie, R., Coram, J. and Ransley, T.:
533 Groundwater Sampling and Analysis – A Field Guide. Geoscience Australia, Record 2009/27
534 95, 2009.

535 Timms, W. A., Crane, R. A., Anderson, D. J., Bouzalakos, S., Whelan, M., McGeeney, D.,
536 Rahman, P. F., Guinea, A., and Acworth, R. I.: Vertical hydraulic conductivity of a clayey-silt
537 aquitard: accelerated fluid flow in a centrifuge permeameter compared with in situ conditions.
538 *Hydrology and Earth System Sciences Discussion*, 11, 3155-3212, 2014.

539 Timms, W., Hendry, M.: Long-term reactive solute transport in an aquitard using a centrifuge
540 model. *Ground Water*, 46, 616-628, 2008.

541 Timms, W., Hendry, M.J., Muise, J., Kerrich, R.: Coupling centrifuge modeling and laser
542 ablation inductively coupled plasma mass spectrometry to determine contaminant retardation
543 in clays. *Environ. Sci. Technol.*, 43, 1153–1159, 2009.

544 Totsche, K. U., Jann, S., Kogel-Knabner, I.: Release of polycyclic aromatic hydrocarbons,
545 dissolved organic carbon, and suspended matter from disturbed NAPL-contaminated gravelly
546 soil material. *Vadose Zone Journal*, 5, 469-479, 2006.

547 Wehrer, M., Totsche, K. U.: Detection of non-equilibrium contaminant release in soil columns:
548 Delineation of experimental conditions by numerical simulations. *Journal of Plant Nutrition*
549 *and Soil Science*, 116, 475-483, 2003.

550 Wehrer, M., Totsche, K. U.: Determination of effective release rates of polycyclic aromatic
551 hydrocarbons and dissolved organic carbon by column outflow experiments. *European*
552 *Journal of Soil Science*, 56, 803-813, 2005.

553 Wehrer, M., Totsche, K. U.: Difference in PAH release processes from tar-oil contaminated
554 soil materials with similar contamination history. *Chemie de Erde – Geochemistry*, 69, 109-
555 124, 2009.

556 Wilson, J. L., and L. W. Gelhar.: Analysis of longitudinal dispersion in unsaturated flow 1. The
557 analytical method, *Water Resour. Res.*, 17(1), 122– 130, 1981.

558

559

560

561
562
563
564
565
566
567
568
569
570
571
572
573
574

