

1 An index of floodplain surface complexity

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9 10 Abstract

11 Floodplain surface topography is an important component of floodplain ecosystems. It is the
12 primary physical template upon which ecosystem processes are acted out, and complexity in
13 this template can contribute to the high biodiversity and productivity of floodplain
14 ecosystems. There has been a limited appreciation of floodplain surface complexity because
15 of the traditional focus on temporal variability in floodplains as well as limitations to
16 quantifying spatial complexity. An index of floodplain surface complexity (*FSC*) is developed
17 in this paper and applied to eight floodplains from different geographic settings. The index is
18 based on two key indicators of complexity; variability in surface geometry (*VSG*) and the
19 spatial organization of surface conditions (*SPO*) and was determined at three sampling scales.
20 Relationships between these measures of spatial complexity and environmental drivers,
21 namely; flow variability (mean daily discharge [*Q*], the coefficient of variation of daily
22 discharge [*Q_{CV}*], the coefficient of variation of mean annual discharge [*Q_{CVAnn}*], the
23 coefficient of variation of maximum annual discharge [*Q_{CVMax}*]), sediment yield (*SY*), valley
24 slope (*Vs*), and floodplain width (*Fpw*) were examined. *FSC*, *VSG*, and *SPO* varied between
25 the eight floodplains and these differences depended upon sampling scale. All complexity
26 values declined with increasing *Fpw* in either a power, logarithmic, or exponential function.
27 There was little change in surface complexity with floodplain widths greater than 10 km. *VSG*
28 was significantly related to *SY* and no significant relationships were determined between any
29 of the hydrological variables and floodplain surface complexity.

1 **1 Introduction**

2 The floodplain surface is an important component of floodplain ecosystems. It provides the
3 primary physical template (*sensu* Southwood, 1977) upon which ecosystem and evolutionary
4 processes are acted out (Salo, 1990). Complexity of floodplain surfaces contributes to the
5 relative abundance of physical habitat (Hamilton et al., 2007), high biodiversity (Ward et al.,
6 1999), and elevated levels of productivity (Thoms, 2003), as well as nonlinear ecosystem
7 responses to inundation (Murray et al., 2006; Thapa et al., 2015). The majority of floodplain
8 research has focused on temporal variations and in particular how hydrology drives floodplain
9 structure and function (Junk et al., 1989; Hughes, 1990; Bayley, 1995; Whited et al., 2007).
10 Such a focus has contributed to a limited appreciation of the spatial complexity of floodplain
11 surfaces.

12 There are two main components to the spatial complexity of floodplain surfaces (Scown et al.,
13 2015a). The first component relates to the presence/absence, abundance, and diversity of
14 features or conditions present. This influences the number and range of distinct habitats and
15 potential interactions between those habitats; both of which contribute to complexity (Levin,
16 1998; Phillips, 2003). The second component is concerned with the spatial organization of
17 features or conditions present within a floodplain surface. Spatial organization controls local
18 interactions and feedbacks between features and can emerge in the absence of any global
19 control (Hallet, 1990). It also affects the flux of matter and energy throughout systems
20 (Wiens, 2002). Any measurement of spatial complexity must incorporate both components;
21 something that does not generally occur (Cadenasso et al., 2006). Studies of floodplain
22 surface complexity have been limited because they tend to only measure one of the
23 components of spatial complexity (Scown et al., 2015c). Moreover, many of the measures of
24 spatial complexity that have been proposed are based on categorical ‘patch’ data (e.g.,
25 Papadimitriou, 2002). Such data have limitations because of the qualitative delineation of
26 patch boundaries, loss of information within patches, and subsequent analyses of these data
27 being restricted to the minimum scale at which patches were initially defined (McGarigal et
28 al., 2009). Continuous numerical data have been used in some studies, and single metrics of
29 surface complexity have been developed, such as rugosity or fractal dimension (see review by
30 Kovalenko et al., 2012). These single-metric-based indices do not fully encompass the
31 multivariate nature of spatial complexity; thus, multiple indicators are required to get the full
32 measure of surface complexity (Dorner et al., 2002; Frost et al., 2005; Tokeshi and Arakaki,

1 2012). While frameworks encompassing the multiple dimensions of complexity have also
2 been proposed (e.g., Cadenasso et al., 2006), they have not provided a quantitative measure of
3 spatial complexity (Scown et al., 2015c). Quantitative measures of floodplain spatial
4 complexity are required in order to advance our understanding of the influences of spatial
5 complexity on these ecosystems and how it varies between floodplains.

6 New technologies are available for intensive data capture, such as light detection and ranging
7 (LiDAR), and the analysis of these data using geographic information systems (GIS)
8 overcomes many of the limitations that have inhibited the quantification of spatial
9 complexity. LiDAR provides high resolution, quantitative topographic data over large areas
10 for many landscapes including floodplains. These data are useful for measuring floodplain
11 surface complexity. LiDAR-derived digital elevation models (DEMs) of floodplain surfaces
12 can be used to measure the character and variability of surface features or conditions using a
13 suite of surface metrics (McGarigal et al., 2009) and moving window analyses (Bar Massada
14 and Radeloff, 2010; De Jager and Rohweder, 2012). The spatial organization of these features
15 or conditions can then be measured using spatial correlograms and geostatistical models
16 (Rossi et al., 1992). These quantitative measurements of the two components of spatial
17 complexity can be incorporated into a single multivariate index. The advantages of using
18 single indices that can be decomposed into sub-indices (e.g., for use in assessing ecosystem
19 health [Norris et al., 2007]) have been widely favoured in ecosystem research.

20 A quantitative index of floodplain surface complexity is developed in this study and applied
21 to eight floodplains from different geographic settings. The primary data source is a LiDAR-
22 derived DEM for each floodplain. The character and variability of surface features and
23 conditions and their spatial organization are incorporated into a single quantitative index to
24 enable a comparison of surface complexity between floodplains. The different environmental
25 settings of each floodplain provide an opportunity to determine the influence of
26 environmental controls on floodplain surface complexity. In addition, the index is measured
27 over three sampling scales (moving window sizes) to investigate the effects of scale on
28 floodplain surface complexity. In this study we ask three questions: 1) Does the surface
29 complexity of the eight floodplains differ and is this consistent among sampling scales? 2)
30 Are the two components of spatial complexity related in floodplain surfaces? 3) What
31 environmental factors influence floodplain surface complexity?

1 **2 Study area**

2 Eight floodplain surfaces from different geographic settings were examined in this study (Fig.
3 1, Table 1). The Bidgee, Gwydir, Macquarie, Narran, and Yanga floodplains are all located
4 within the Murray-Darling Basin in S.E. Australia; whereas the floodplain of the Woodforde
5 is located in central Australia approximately 150 km north of the town of Alice Springs. The
6 floodplain of the Shingwedzi is located in N.E. South Africa, in the northern regions of
7 Kruger National Park; and the floodplain of the Upper Mississippi is located within
8 navigation Pool 9 and forms the boundary of the states of Minnesota, Wisconsin and Iowa in
9 the USA. Details of the eight floodplains are provided in Table 1, and in summary, they
10 differed in terms of their degree of valley confinement, climate, and position within the
11 stream network. Four floodplains (the Bidgee, Mississippi, Shingwedzi, and Woodforde) are
12 contained within relatively confined river valley troughs with floodplains width ranging
13 between one and five kilometers. The other four floodplains (the Gwydir, Macquarie, Narran,
14 and Yanga) are all contained within relatively unconfined river valleys with floodplain widths
15 up to 60 kilometers. The eight floodplains also differ in their hydrology and geomorphology,
16 exhibiting a variety of morphological features such as flood channels, oxbows, natural levees,
17 crevasse splays, and back swamps. Detailed descriptions of each of the eight floodplains are
18 provided by Scown et al. (2015a).

19

20 **3 Methods**

21 The Index of Floodplain Surface Complexity (*FSC*) developed for this study was calculated
22 from data extracted from LiDAR-derived digital elevation models (DEMs) for each
23 floodplain. Floodplain extents were delineated using multiple lines of evidence. This
24 delineation was based on examination of breaks of slope in the DEM, contours, changes in
25 vegetation from aerial photography, soil conditions from local soil conservation surveys, and
26 floodwater extents derived from Landsat TM imagery. A buffer within this manually
27 delineated extent was also removed to ensure nothing other than what was deemed to be part
28 of the floodplain was included. Permanently inundated areas were also removed because
29 attaining accurate subsurface land elevations using LiDAR is difficult. Each DEM was then
30 detrended to remove the overall downstream slope. Details of the detrending procedures for
31 each of the floodplains are provided by Scown et al. (2015a; 2015b). Each detrended DEM

1 was subsequently resampled to a $5 \times 5 \text{ m}^2$ grid size using the cubic method in ArcGIS 10.2
2 because this was the finest resolution available for one of the floodplains.

3 The *FSC* index is comprised of two sub-indices, which record the two components of spatial
4 complexity; the variability in surface geometry (*VSG*) and the spatial organization of surface
5 conditions (*SPO*). *VSG* is a composite of four surface metrics (Table 2), measured at 50
6 random sample locations throughout each of the floodplains, while *SPO* is calculated from
7 spatial correlogram models of Moran's I over increasing lag distances for each of the four
8 surface metrics from 1000 random sample locations (Table 2). Details of the procedures for
9 calculating each indicator are provided by Scown et al. (2015a). In summary, the surface
10 metrics are used to indicate increasing surface variability, while the spatial correlogram model
11 parameters (range and nugget) are used to indicate increasing 'patchiness' or organization in
12 the surface (Table 2). It is argued here, and elsewhere (Scown, 2015; Scown et al., 2015a),
13 that increasing variability and spatial organization results in increasing spatial complexity. All
14 surface metrics were measured within sampling windows of 50 m, 200 m, and 1000 m radius.
15 These window sizes were chosen based on the identification of scale thresholds between them
16 by Scown et al. (2015b). This enabled us to determine whether any effect of sampling scale
17 occurred.

18 The individual indicators were combined and weighted, using the standardized Euclidean
19 distance procedure, to calculate the overall *FSC* index. This index was used for an overall
20 assessment of floodplain surface complexity and the sub-indices of *VSG* and *SPO* were
21 derived to provide specific interpretations of the two components of spatial complexity for
22 each floodplain surface. An example of *FSC* calculation is given in Equation (1), where *I* is
23 the overall index and *A*, *B*, *C*, ... , *N* are the *n* individual indicators of surface complexity,
24 the details of which are provided in Table 2.

$$I = 1 - \frac{\sqrt{(1 - A)^2 + (1 - B)^2 + (1 - C)^2 + \dots + (1 - N)^2}}{\sqrt{n}} \quad (1)$$

25
26 Calculating the *FSC* index required the *SPO* indicators to have an additional weighting of 0.5,
27 as there were twice as many indicators of *SPO* compared to *VSG*. All indicators were range-
28 standardized and scaled between 0 and 1, hence this index provides a relative measure among
29 those floodplains studied. An index value approaching one indicates the floodplain surface is
30 among the most spatially complex of all floodplains observed, while an index value

1 approaching zero indicates the floodplain surface is among the least spatially complex. The
2 approach used has been applied successfully in developing a large scale index of River
3 Condition (Norris et al., 2007).

4 Relationships between the two components of spatial complexity were also investigated *VSG*
5 and *SPO* at each sampling scale. In addition, relationships between *VSG*, *SPO*, and *FSC* and
6 seven environmental variables were also investigated. The environmental variables were
7 mean daily discharge in ML/day (*Q*), CV daily discharge (*Q_{CV}*), CV mean annual discharge
8 (*Q_{CVAnn}*), CV maximum annual discharge (*Q_{CVMax}*), sediment yield in t/km²/y (*SY*), average
9 valley slope in m/m (*Vs*), and average floodplain width in km (*Fpw*). Detailed calculations of
10 environmental variables are provided by Scown et al. (2015a). Each of these environmental
11 variables reflect an aspect of the flow, sediment, energy, and valley conditions, which have
12 previously been shown to influence floodplain surface morphology (Nanson and Croke, 1992;
13 Warner, 1992). Curve estimation between *VSG*, *SPO*, and *FSC* and each environmental
14 variable at each sampling scale was conducted in SPSS. *Q*, *SY*, and *Vs* were normalized using
15 a logarithmic transformation before analysis.

16

17 **4 Results**

18 **4.1 Floodplain surface complexity (*FSC*)**

19 Floodplain surface complexity, as measured by the *FSC* index, was highly variable among the
20 eight floodplains and across sampling scales. The Gwydir floodplain had the least complex of
21 surfaces across all sampling scales (mean *FSC* of 0.17), while the Shingwedzi floodplain had
22 the most complex surface (mean *FSC* of 0.69) across all scales (Fig. 2). This presumably
23 reflects differences in the geomorphology of these two floodplains. The Shingwedzi
24 floodplain is dissected by numerous channels and gullies, which create highly organized
25 patches of increased topographic relief, whereas the Gwydir floodplain has a relatively flat,
26 featureless surface over larger continuous areas and limited organization around any of the
27 significant surface features. The effect of sampling scale on *FSC* was not consistent across the
28 eight floodplains (Fig. 2), indicating that comparisons among floodplains are scale-dependent.
29 For example, the Gwydir and Narran floodplain surfaces became more complex with
30 increasing window size, whereas the Shingwedzi, Macquarie, and Mississippi floodplains
31 became less complex.

1 **4.2 Variability in surface geometry (VSG)**

2 The *VSG* index was also highly variable among the eight floodplains and across sampling
3 scales (Fig. 3). Again, the Gwydir floodplain consistently had the lowest values for this index
4 over all window sizes (mean *VSG* of 0.06), while the Shingwedzi floodplain consistently had
5 the highest (mean *VSG* of 0.65). This reflects the large differences in topographic relief and
6 variability between these two floodplains. The *VSG* score of 0.00 for the Gwydir floodplain at
7 the 50 m window size indicates that this floodplain had the lowest scores for all four
8 indicators of variability in surface geometry of the eight floodplains studied at this scale.
9 Similar to *FSC*, the effect of sampling scale on *VSG* was not consistent across floodplains
10 (Fig. 3). *VSG* increased with sampling scale for the Narran floodplain, but decreased for the
11 Shingwedzi, Bidgee, Macquarie, and Woodforde floodplains. *VSG* was highest at the 50 m
12 window size and lowest at 200 m for the Mississippi and Yanga floodplains, while it was
13 highest at 200 m and lowest at 50 m for the Gwydir. This indicates that the scale at which
14 surface geometry is most variable depends on the floodplain.

15 **4.3 Spatial organisation of surface conditions (SPO)**

16 The *SPO* index was also highly variable among the eight floodplains and across sampling
17 scales (Fig. 4). Unlike *FSC* and *VSG*, there was no consistency as to which floodplain had the
18 highest and lowest *SPO* across sampling scales. This indicates that no floodplain has
19 consistently the highest or lowest degree of spatial organization of surface conditions among
20 the eight floodplains studied. The effect of sampling scale on *SPO* was also inconsistent
21 across floodplains (Fig. 4). For five of the eight floodplains, *SPO* was lowest at the 200 m
22 window size and highest at 1000 m. For the Mississippi and Woodforde floodplains, the
23 opposite was observed, with *SPO* being highest at 200 m and lowest at 1000 m. The Bidgee
24 floodplain was the only floodplain for which *SPO* increased consistently across all sampling
25 scales. This indicates that the degree of spatial organization of surface conditions is highest at
26 large sampling scales for most floodplains, but at intermediate scales for some. *SPO* was
27 highly variable across window sizes for the Yanga, Woodforde, and Gwydir floodplains. *SPO*
28 was 178 % higher at the 1000 m window size than at 200 m for the Gwydir floodplain and
29 138 % higher for the Yanga floodplain, while for the Woodforde floodplain it was 61 %
30 lower. This indicates a significant change in the spatial organization of these floodplain
31 surfaces between these two sampling scales. The results also showed that floodplain and

1 window size have a greater combined effect on *SPO* among the eight floodplains than on
2 relative *FSC* and *VSG* (Figs. 2, 3, and 4).

3 **4.4 Relationship between *VSG* and *SPO***

4 *SPO* values were, on average, 17 % higher than the *VSG* values. The greatest difference
5 between *SPO* and *VSG* was 0.51 for the Woodforde floodplain, at the 200 m window size,
6 followed by 0.47 for the Bidgee and Yanga floodplains at the 1000 m window size (Figs. 3
7 and 4). The Mississippi floodplain was the only floodplain where *SPO* was lower than the
8 *VSG*, with an average difference of -0.03. This comparison between *SPO* and *VSG* values,
9 suggests surface conditions across the eight floodplains are generally more highly spatially
10 organized than they are geometrically variable.

11 Average variance of *SPO* across sampling scales within a floodplain (0.0212) was almost six
12 times higher than that of *VSG* (0.0037). However, the average *SPO* variance was dominated
13 by a limited number of floodplains; notably the Gwydir, Woodforde, and Yanga floodplains
14 (Fig. 4). Four of the five other floodplains had a less variable *SPO* across sampling scales
15 compared to their *VSG*; with the exception being the Bidgee floodplain. These results of *SPO*
16 variance across sampling scales indicates that, on average, the spatial organization of surface
17 conditions is much more sensitive to sampling scale than the variability of surface geometry.

18 Significant linear relationships between *VSG* and *SPO* were recorded at the 50 m and 200 m
19 window sizes only (Table 3). Overall, *SPO* increased with *VSG* (Fig. 5) and this positive
20 relationship was strongest at the 50 m window size, with more than 61 % of the variance in
21 *SPO* explained by *VSG*, reducing to 56 % at the 200 m window size, and less than 8 % at
22 1000 m window size. The y-intercept of each regression line was greater than 0.1 at all
23 window sizes, while the slope was less than one at 50 m and 1000 m, but greater than one at
24 200 m (Table 3). This provides further indication that *SPO* is generally higher than *VSG* in
25 these eight floodplains.

26 **4.5 Relationships between floodplain surface complexity and environmental** 27 **variables**

28 Floodplain width (Fpw) was the only environmental variable statistically related to any of the
29 three indices of spatial complexity ($p < 0.05$). This variable was significantly related to *FSC*
30 and *VSG* over all window sizes, and to *SPO* over all but the 1000 m window size (Table 4).

1 The decrease in all three complexity indices with increasing Fpw was best explained by either
2 a power, logarithmic, or exponential function (Table 4). In terms of the decrease in *FSC* with
3 increasing Fpw, this was best explained by a power function at all window sizes (Fig. 6a),
4 indicating *FSC* undergoes rapid decline with increases in Fpw, approaching an asymptote at
5 approximately 10 km in Fpw. The modelled change in *FSC* with increasing Fpw was almost
6 identical between the 50 m and 200 m window sizes. At the 1000 m window size, *FSC* was
7 generally lower compared to that at 50 m and 200 m windows sizes in narrow floodplains,
8 before approaching a higher asymptote at larger Fpw. This indicates that broad floodplains
9 generally have higher *FSC* when measured at larger sampling scales, whereas narrow
10 floodplains generally have higher *FSC* when measured at smaller sampling scales.

11 Decreases in *VSG* with increasing Fpw was best explained by a logarithmic function at the 50
12 m window size, a power function at the 200 m window size, and an exponential function at
13 1000 m (Fig. 6b). These models indicate a more rapid initial decline in *VSG* with increasing
14 Fpw at the 200 m window size than at the 50m and 1000 m window sizes. This is followed by
15 approach to a higher asymptote at the 200 m window size above Fpw of approximately 10
16 km, whereas modelled *VSG* continues to decline between Fpw of 10 km and 25 km at the 50
17 m and 1000 m window sizes. This indicates that Fpw has a greater effect on *VSG* in wider
18 floodplains when measured at small and large sampling scales than it does at intermediate
19 scales. The relationship was strongest at the 200 m window size, with more than 80 % of the
20 variance in *VSG* being explained by Fpw.

21 The decrease in *SPO* with increasing Fpw was best explained by a logarithmic function at the
22 50 m and 200 m window sizes (Fig. 6c). The modelled decline in *SPO* was initially more
23 rapid at the 50 m window size than at 200 m, before approaching a higher asymptote at
24 narrower Fpw. This indicates that Fpw has more of an effect on *SPO* in wider floodplains
25 when measured at the 200 m window size than at 50 m. The relationship was strongest at the
26 200 m window size, with more than 77 % of the variance in *SPO* being explained by Fpw.
27 This was reduced to 71 % at the 50 m window size. There was no significant relationship
28 between Fpw and *SPO* at the 1000 m window size (Fig. 6c). This suggests that Fpw exerts
29 little or no control over the spatial organization of surface conditions when measured at large
30 sampling scales.

31 A weak statistical relationship was recorded between *SY* and *VSG*. An increase in *VSG* with
32 increasing *SY* was observed at the 200 m window size ($r^2 = 0.44$; $p = 0.07$). The relatively

1 lower level of significance of this result was attributable to the Gwydir having a high SY but a
2 very low VSG. When the Gwydir floodplain was removed from the analysis, there was a
3 significant and strong linear relationship between log-transformed SY and VSG across all
4 window sizes for the remaining seven floodplains (Table 5, Fig. 7). This relationship was
5 almost identical across all window sizes.

6

7 **5 Discussion**

8 The Euclidean Index of floodplain surface complexity (FSC) used in this study is comprised
9 of the two key components of spatial complexity; the character and variability of features or
10 conditions, and their spatial organization. This index appears to discriminate between
11 floodplains with distinctly different geomorphological features. The multivariate nature of the
12 index, comprised of 12 indicators of surface complexity (Table 2), has advantages over
13 univariate indices that have been applied to measure floodplain surface complexity.
14 Univariate indices fail to incorporate both aspects of surface structure, which contribute to
15 surface complexity (Dorner et al., 2002; Frost et al., 2005; Tokeshi and Arakaki, 2012).
16 Having a single, multivariate-based index is also favorable rather than multiple individual
17 indicators of floodplain surface complexity, as it allows a quantitative measure that can be
18 compared for multiple riverine landscapes. Norris et al. (2007) provide a comparable example
19 of such an application in their assessment of river condition. It is important to note that, the
20 standardization of indicator scores from 0 to 1 is necessary for the Euclidean Index equation
21 (Norris et al., 2007), as the FSC index is a relative index of floodplain surface complexity
22 across a group of floodplains all of which were included in the standardization of the
23 indicators. This is appropriate for examining relationships between floodplain surface
24 complexity and environmental controls, given adequate replication over a range of floodplain
25 settings is achieved. However, it should not be used to compare against indices of other
26 studies, unless all floodplains being compared are included in the calculation of the index.

27 The results of this research demonstrate: floodplain surface complexity to be highly variable
28 among the eight floodplains studied, and floodplain width to exert a significant ‘top-down’
29 control (sensu Thorp et al., 2008) on differences in floodplain surface complexity. These
30 results clearly support geomorphological and ecological thinking that “...*the valley rules the*
31 *stream...*”, as argued first by Hynes (1975) and strongly supported since (e.g., Schumm, 1977;
32 Miller, 1995; Panin et al., 1999; Thoms et al., 2000). In this case, the valley rules the

1 floodplain surface complexity, at least in terms of the ‘top-down’ influences investigated here.
2 The influence of floodplain width on floodplain surface complexity decreases significantly
3 once widths are greater than 10 km. Above 10 km, little change in the index of floodplain
4 surface complexity was recorded. This is likely due to the dissipation of flood energy in wide
5 floodplains, limiting the construction of large topographic features, which contribute to
6 surface complexity. However, subtle topographic features in wide floodplains are also
7 important surface features (Fagan and Nanson, 2004), which may have been overlooked in
8 this index. In narrower, confined settings, where widths are less than 10 km, floodplain
9 construction may be the result primarily of vertical processes (e.g., accretion/incision) leading
10 to more prominent topographic features that exhibit a higher degree of spatial organization
11 and thus increased surface complexity (Nanson and Croke, 1992). Such complexity can lead
12 to the concentration of flood energies in particular areas, promoting episodic catastrophic
13 stripping (Nanson, 1986). The narrowest floodplain examined in this study was, on average,
14 1.5 km in width and the results presented in this study may not be consistent in floodplains
15 narrower than this. In particular, there is a loss of surface complexity when floodplains are
16 contained between artificial levees or embankments (Florsheim and Mount, 2002; Gurnell and
17 Petts, 2002), so floodplain surface complexity should not be considered to increase
18 indefinitely in floodplains approaching a width of 0 km.

19 Valley trough or floodplain width has been identified as a primary controller of floodplain
20 pattern and process in several previous studies. Spatial patterns of flow depth, velocity, and
21 shear stress in overbank flows were found by Miller (1995) to all be influenced by valley
22 width and this influence was particularly noticeable at locations of valley widening or
23 narrowing. Similarly, Thoms et al. (2000) found that valley width had a significant effect on
24 sediment textural character and associated heavy metal concentrations within different
25 morphological units of the Hawkesbury River Valley, New South Wales. In particular, they
26 found higher proportions of silt and clay, and lower proportions of sand and gravel, in wide
27 floodplain sections compared to narrow floodplains. The results of this present research
28 support the findings that floodplain width is an important controller of floodplain pattern and
29 process.

30 The effect of floodplain width was relatively consistent across all three indices examined.
31 This suggests that floodplain width has a similar effect on the variability of floodplain surface
32 geometry, the degree of spatial organization, and overall floodplain surface complexity. This

1 likely explains the significant positive linear relationship between the variability of surface
2 geometry and the spatial organization of surface conditions sub-indices. This relationship
3 likely occurs because environmental conditions, particularly related to floodplain width,
4 which promote higher variability in floodplain surfaces, also cause a high degree of spatial
5 organization. Reinforcing feedbacks between these two components of spatial complexity
6 may also exist. That is, high variability of surface geometry promotes a high degree of spatial
7 organization, and vice versa. Positive feedback is common in complex systems (Levin, 1998;
8 Phillips, 2003), and feedbacks between hydrology, geomorphology, and biology in
9 floodplains may play a part in this (Hughes, 1997).

10 The textural character of floodplain sediments and local energy conditions during inundation
11 has been postulated as important controls of floodplain morphology (Nanson and Croke,
12 1992). These two drivers would also be expected to influence floodplain surface complexity.
13 In this study, sediment yield was found to have a weak effect on the variability in surface
14 geometry, although relationships were not significant. This may be because estimates of
15 contemporary sediment yield were used in this study, whereas historical sediment yields are
16 relatively more important (Panin et al., 1999). Substantial anthropogenic increases in
17 sediment loads have been reported for the Gwydir floodplain (De Rose et al., 2003). Removal
18 of this floodplain from our analyses, resulted in a significant increase in variability in surface
19 geometry with increasing sediment yield across the seven remaining floodplains. This result
20 suggests that sediment yield may exert ‘top-down’ control on the variability of floodplain
21 surface geometry, although recent anthropogenic changes in sediment yields (Prosser et al.,
22 2001), particularly increased erosion in the catchment due to land use changes, may have
23 delayed ‘lag’ effects on floodplain surfaces which have not yet been observed. Valley slope
24 was used in this study as a surrogate for stream energy, and this was not found to have any
25 effect on overall floodplain surface complexity. More accurate measures of energy conditions
26 such as specific stream power (Nanson and Croke, 1992) may reveal any effects of energy
27 conditions on floodplain surface complexity, if they exist, more clearly. It is also likely that
28 variable flood energy conditions within each floodplain have an effect on localized surface
29 complexity. For example, Fagan and Nanson (2004) found distinct differences in floodplain
30 surface channel patterns among high, intermediate, and low energy areas of the semi-arid
31 Cooper Creek in Australia. They also found the energy of flood flows to be largely controlled
32 by floodplain width.

1 Hydrology has been widely considered the main determinant of floodplain ecosystem pattern
2 and process (Junk et al., 1989; Hughes, 1990; Bayley, 1995; Whited et al., 2007). However,
3 the research presented in this paper indicates that this may not be the case for floodplain
4 surface complexity. None of the four hydrological variables measured here had a significant
5 effect on floodplain surface complexity. This suggests that, although hydrology is largely
6 important in driving floodplain ecosystem processes, floodplain width and sediment
7 conditions appear to exert more control over the complexity of floodplain surfaces. This is
8 important given that floodplain research and restoration is often focused on hydrology,
9 particularly connectivity (e.g., Thoms, 2003; Thoms et al., 2005); whereas valley trough,
10 sediment, and energy conditions may be more important in structuring and maintaining the
11 physical template upon which hydrology acts as an ecosystem driver (Salo, 1990). Loss of
12 floodplain surface complexity due to changes in sediment yield or calibre, or confinement
13 between artificial levees, may be as ecologically important as changes to hydrology and
14 should not be overlooked (Thoms, 2003). It is important to note, however, that some of the
15 eight floodplains studied have experienced anthropogenic alterations to their hydrology. Thus,
16 hydrological parameters based on contemporary data may not reflect the nature of the flow
17 regime that was influential in establishing current surface conditions; lagged effects of altered
18 hydrology on surface complexity may occur in the future (sensu Thoms, 2006).

19 Riverine landscapes and their floodplains are hierarchically organized ecosystems (Dollar et
20 al., 2007; Thorp et al., 2008), being composed of discrete levels of organization distinguished
21 by different process rates (O'Neill et al., 1989). Each level of organization, or holon, has a
22 spatial and temporal scale over which processes occur and patterns emerge (Holling, 1992).
23 The different sampling scales used in this research indicate that the scale at which patterns in
24 floodplain surfaces are most complex depends on the floodplain setting. In particular, wide,
25 unconfined floodplains appear to have higher floodplain surface complexity when measured
26 at larger sampling scales, whereas narrow, confined floodplains have so at smaller sampling
27 scales. Thus, the scale at which floodplain surface complexity is maximized likely relates to
28 the width of the floodplain. Selecting different window sizes tailored to each floodplain
29 individually relative to floodplain width should be the focus of future research. This may
30 reveal consistent effects of scale on floodplain surfaces.

31 These results suggest that the scale of processes that maximize complexity, and potentially
32 biodiversity and productivity (Tockner and Ward, 1999), in floodplains differ between

1 different valley settings. This has implications for understanding and managing the
2 complexity of floodplain ecosystems. Floodplain processes, which operate over certain
3 temporal scales, elicit a response over relative spatial scales (Salo, 1990; Hughes, 1997).
4 Consequently, managing processes at the appropriate scale to achieve desired outcomes is
5 important (Parsons and Thoms, 2007). This has already been recognized for managing
6 floodplain hydrology to maintain biodiversity (Amoros and Bornette, 2002) and these results
7 indicate it is also important for managing the processes that maintain floodplain surface
8 complexity.

9 Recent approaches to examining and understanding ecosystem complexity and the emergent
10 properties that arise from interactions within systems emphasise the importance of
11 heterogeneity, connectivity, and contingency within the landscape (Loreau et al., 2003;
12 Cadenasso et al., 2006). We have presented an index of floodplain surface complexity within
13 such a framework that incorporates measures of variability and spatial organization. These
14 two components of spatial complexity are directly associated with heterogeneity and
15 connectivity (Wiens, 2002), although no direct measure of historical contingency is given in
16 this spatial approach. Metrics and indicators used to measure properties of landscape and
17 ecosystem complexity in the past have largely been based on discrete units and the familiar
18 concept of ‘patches’ (Forman and Godron, 1981). The surface metrics employed in this study
19 are conceptually equivalent to certain patch metrics and a comprehensive comparison of
20 surface and patch metrics is provided by McGarigal et al. (2009). Thus, the approach
21 presented in this study should be considered complimentary to other ecosystem complexity
22 frameworks, such as the meta-ecosystem approach (Loreau et al., 2003), which are based on
23 patches.

24 In terms of the origin and implications of floodplain surface complexity, this research focuses
25 on ‘top-down’ environmental drivers of floodplain surface complexity. ‘Bottom-up’
26 feedbacks from the floodplain ecosystem are also likely to affect surface complexity. For
27 example, vegetation establishment on deposited floodplain sediments is known to produce a
28 positive feedback loop in which more sediment is trapped and semi-permanent morphological
29 features such as islands develop (Nanson and Beach, 1977; Hupp and Osterkamp, 1996). Such
30 feedbacks are likely to influence floodplain surface complexity, particularly in floodplains
31 dominated by such features (Gurnell and Petts, 2002; Stanford et al., 2005). ‘Bottom-up’
32 influences on floodplain surface complexity are difficult to quantify and were not examined in

1 this study. Future research into the influence of vegetation type and density on floodplain
2 surface complexity, particularly in relation to its hydraulic roughness, may provide valuable
3 insights into ‘bottom-up’ controls on floodplain surface complexity. Such data are also
4 available through LiDAR (Straatsma and Baptist, 2008). Effects of floodplain surface
5 complexity on biodiversity and productivity should also be examined in future research. The
6 floodplain surface provides the primary geomorphic template upon which ecosystem and
7 evolutionary processes are acted out (Salo, 1990) and it would be expected that increased
8 surface complexity would promote the range of physical habitats required to maintain
9 floodplain biodiversity (Hamilton et al., 2007).

10 The inclusion of other floodplains, from different regions, in future studies of this nature,
11 would further determine whether the trends observed in this study extend beyond the
12 floodplains investigated here. This study was limited to eight floodplains because of data
13 availability. As high-resolution LiDAR data across many more floodplains are made available
14 to researchers, other analyses such as multiple regression will be possible in studies such as
15 this. Multiple regression would enable the interactive effects of environmental variables to be
16 elucidated, whereas this study was limited to relatively simple linear regression because of the
17 sample size of only eight floodplains.

18

19 **Acknowledgements**

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24

25 **References**

- 26 Amoros, C. and Bornette, G.: Connectivity and biocomplexity in waterbodies of riverine
27 floodplains, *Freshwater Biology*, 47, 761-776, 2002.
- 28 Bar Massada, A. and Radeloff, V. C.: Two multi-scale contextual approaches for mapping
29 spatial pattern, *Landscape Ecology*, 25, 711-725, 2010.

- 1 Bayley, P. B.: Understanding large river: floodplain ecosystems, *BioScience*, 45, 153-158,
2 1995.
- 3 Cadenasso, M. L., Pickett, S. T. A., and Grove, J. M.: Dimensions of ecosystem complexity:
4 heterogeneity, connectivity, and history, *Ecological Complexity*, 3, 1-12, 2006.
- 5 Carey, W. C.: Formation of floodplain lands, *J. Hydr. Eng. Div.-ASCE*, 95, 981-994, 1969.
- 6 De Jager, N. R. and Rohweder, J. J.: Spatial patterns of aquatic habitat richness in the Upper
7 Mississippi River floodplain, USA, *Ecological Indicators*, 13, 275-283, 2012.
- 8 De Rose, R. C., Prosser, I. P., Wiese, M., and Hughes, A. O.: Patterns of erosion and
9 sediment and nutrient transport in the Murray-Darling Basin, Technical Report 32/03, CSIRO
10 Land and Water, Canberra, Australia, 2003.
- 11 Dollar, E. S. J., James, C. S., Rogers, K. H., and Thoms, M. C.: A framework for
12 interdisciplinary understanding of rivers as ecosystems, *Geomorphology*, 89, 147-162, 2007.
- 13 Fagan, S. D. and Nanson, G. C.: The morphology and formation of floodplain-surface
14 channels, Cooper Creek, Australia, *Geomorphology*, 60, 107-126, 2004.
- 15 Florsheim, J. L. and Mount, J. F.: Restoration of floodplain topography by sand-splay
16 complex formation in response to intentional levee breaches, Lower Cosumnes River,
17 California, *Geomorphology*, 44, 67-94, 2002.
- 18 Forman, R. T. T. and Godron, M.: Patches and structural components for a landscape ecology,
19 *BioScience*, 31, 733-740, 1981.
- 20 Gurnell, A. M. and Petts, G. E.: Island-dominated landscapes of large floodplain rivers, a
21 European perspective, *Freshwater Biology*, 47, 581-600, 2002.
- 22 Hallet, B.: Spatial self-organization in geomorphology: from periodic bedforms and patterned
23 ground to scale-invariant topography, *Earth-Science Reviews*, 29, 57-75, 1990.
- 24 Hamilton, S. K., Kellndorfer, J., Lehner, B., and Tobler, M.: Remote sensing of floodplain
25 geomorphology as a surrogate for biodiversity in a tropical river system (Madre de Dios,
26 Peru), *Geomorphology*, 89, 23-38, 2007.
- 27 Holling, C. S.: Cross-scale morphology, geometry, and dynamics of ecosystems, *Ecological*
28 *Monographs*, 62, 447-502, 1992.

- 1 Hughes, F. M. R.: The influence of flooding regimes on forest distribution and composition in
2 the Tana River floodplain, Kenya, *Journal of Applied Ecology*, 27, 475-491, 1990.
- 3 Hughes, F. M. R.: Floodplain biogeomorphology, *Progress in Physical Geography*, 21, 501-
4 529, 1997.
- 5 Hupp, C. R. and Osterkamp, W. R.: Riparian vegetation and fluvial geomorphic processes,
6 *Geomorphology*, 14, 277-295, 1996.
- 7 Hynes, H.: The stream and its valley, *Verhandlungen des Internationalen Verein Limnologie*,
8 19, 1-15, 1975.
- 9 Junk, W. J., Bayley, P. B., and Sparks, R. E.: The flood pulse concept in river-floodplain
10 systems, in *Proceedings of the International Large River Symposium*, edited by D. P. Dodge,
11 pp. 110-127, Canadian Special Publication of Fisheries and Aquatic Sciences No. 106,
12 Canadian Government Publishing Centre, Ottawa, Ontario, 1989.
- 13 Kovalenko, K., Thomaz, S., and Warfe, D.: Habitat complexity: approaches and future
14 directions, *Hydrobiologia*, 685, 1-17, 2012.
- 15 Lecce, S. A.: Spatial patterns of historical overbank sedimentation and floodplain evolution,
16 Blue river, Wisconsin, *Geomorphology*, 18, 265-277, 1997.
- 17 Levin, S. A.: Ecosystems and the biosphere as complex adaptive systems, *Ecosystems*, 1,
18 431-436, 1998.
- 19 Loreau, M., Mouquet, N., and Hold, R. D.: Meta-ecosystems: a theoretical framework for a
20 spatial ecosystem ecology, *Ecology Letters*, 6, 673-679, 2003.
- 21 McGarigal, K., Tagil, S., and Cushman, S.: Surface metrics: an alternative to patch metrics for
22 the quantification of landscape structure, *Landscape Ecology*, 24, 433-450, 2009.
- 23 Miller, A. J.: Valley morphology and boundary conditions influencing spatial patterns of
24 flood flow, in *Natural and Anthropogenic Influences in Fluvial Geomorphology: The Wolman*
25 *Volume*, edited by: Costa, J. E., Miller, A. J., Potter, K. W., and Wilcock, P. R., pp. 57-82,
26 American Geophysical Union, Washington, D. C., 1995.
- 27 Murray, O., Thoms, M., and Rayburg, S.: The diversity of inundated areas in semiarid flood
28 plain ecosystems, in *Sediment Dynamics and the Hydromorphology of Fluvial Systems*,
29 edited by J. S. Rowan et al., pp. 277-286, IAHS Press, Wallingford, U. K., 2006.

- 1 Nanson, G. C.: Episodes of vertical accretion and catastrophic stripping: a model of
2 disequilibrium flood-plain development, *Geological Society of America Bulletin*, 97, 1467-
3 1475, 1986.
- 4 Nanson, G. C. and Beach Forest, H. F.: succession and sedimentation on a meandering-river
5 floodplain, northeast British Columbia, Canada, *Journal of Biogeography*, 4, 229-251, 1977.
- 6 Nanson, G. C. and Croke, J. C.: A genetic classification of floodplains, *Geomorphology*, 4,
7 459-486, 1992.
- 8 Norris, R. H., Linke, S., Prosser, I. P., Young, W. J., Liston, P., Bauer, N., Sloane, N., Dyer,
9 F., and Thoms, M.: Very-broad-scale assessment of human impacts on river condition,
10 *Freshwater Biology*, 52, 959-976, 2007.
- 11 O'Neill, R. V., Johnson, A. R., and King, A. W.: A hierarchical framework for the analysis of
12 scale, *Landscape Ecology*, 3, 193-205, 1989.
- 13 Panin, A. V., Sidorchuk, A. Y., and Chernov, A. V.: Historical background to floodplain
14 morphology: examples from the East European Plain, in *Floodplains: Interdisciplinary*
15 *Approaches*, edited by S. B. Marriott and J. Alexander, pp. 217-229, Geological Society
16 Special Publication No. 163, Geological Society, London, U. K, 1999.
- 17 Papadimitriou, F.: Modelling indicators and indices of landscape complexity: an approach
18 using G.I.S., *Ecological Indicators*, 2, 17-25, 2002.
- 19 Parsons, M. and Thoms, M. C.: Hierarchical patterns of physical–biological associations in
20 river ecosystems, *Geomorphology*, 89, 127-146, 2007.
- 21 Phillips, J. D.: Sources of nonlinearity and complexity in geomorphic systems, *Progress in*
22 *Physical Geography*, 27, 1-23, 2003.
- 23 Prosser, I. P., Rutherford, I. D., Olley, J. M., Young, W. J., Wallbrink, P. J., and Moran, C. J.:
24 Large-scale patterns of erosion and sediment transport in river networks, with examples from
25 Australia, *Marine and Freshwater Research*, 52, 81-99, 2001.
- 26 Rossi, R. E., Mulla, D. J., Journel, A. G., and Eldon, H. F.: Geostatistical tools for modeling
27 and interpreting ecological spatial dependence, *Ecological Monographs*, 62, 277-314, 1992.
- 28 Salo, J.: External processes influencing origin and maintenance of inland water-land ecotones,
29 in *The Ecology and Management of Aquatic-Terrestrial Ecotones*, edited by R. J. Naiman and
30 H. Décamps, pp. 37-64, Unesco, Paris, France, 1990.

- 1 Schumm, S. A.: *The Fluvial System*, John Wiley and Sons, New York, N. Y, 1977.
- 2 Scown, M. W., Thoms, M. C., and De Jager, N. R.: Floodplain complexity and surface
3 metrics: Influences of scale and geomorphology, *Geomorphology*, 245, 102-116, 2015a.
- 4 Scown, M. W., Thoms, M. C., and De Jager, N. R.: Measuring floodplain spatial patterns
5 using continuous surface metrics at multiple scales, *Geomorphology*, 245, 87-101, 2015b.
- 6 Scown, M. W., Thoms, M. C., and De Jager, N. R.: Measuring spatial pattern in floodplains: a
7 step towards understanding the complexity of floodplain ecosystems, in *River Science:
8 Research and Management for the 21st Century*, edited by: Gilvear, D. J., Thoms, M. C.,
9 Wood, P., and Greenwood, M. G., John Wiley and Sons, London, UK, 2015c.
- 10 Southwood, T. R. E.: Habitat, the templet for ecological strategies?, *Journal of Animal
11 Ecology*, 46, 337-365, 1977.
- 12 Stanford, J. A., Lorang, M. S., and Hauer, F. R.: The shifting habitat mosaic of river
13 ecosystems, *Verhandlungen des Internationalen Verein Limnologie*, 29, 123-136, 2005.
- 14 Straatsma, M. W. and Baptist, M. J.: Floodplain roughness parameterization using airborne
15 laser scanning and spectral remote sensing, *Remote Sensing of Environment*, 112, 1062-1080,
16 2008.
- 17 Thapa, R., Thoms, M. C., and Parsons, M.: An adaptive cycle hypothesis of semi-arid
18 floodplain vegetation productivity in dry and wet resource states, *Ecohydrology*,
19 doi:10.1002/eco.1609, in press, 2015.
- 20 Thoms, M. C.: Floodplain–river ecosystems: lateral connections and the implications of
21 human interference, *Geomorphology*, 56, 335-349, 2003.
- 22 Thoms, M. C.: Variability in riverine ecosystems, *River Research and Applications*, 22, 115-
23 151, 2006.
- 24 Thoms, M. C., Parker, C. R., and Simons, M.: The dispersal and storage of trace metals in the
25 Hawkesbury River valley, in *River Management: The Australasian Experience*, edited by S.
26 Brizga and B. Finlayson, pp. 197-219, John Wiley and Sons, Chichester, U. K, 2000.
- 27 Thoms, M. C., Southwell, M., and McGinness, H. M.: Floodplain–river ecosystems:
28 fragmentation and water resources development, *Geomorphology*, 71, 126-138, 2005.

- 1 Thorp, J. H., Thoms, M. C., and Delong, M. D.: *The Riverine Ecosystem Synthesis: Towards*
2 *Conceptual Cohesiveness in River Science*, Elsevier, Amsterdam, 2008.
- 3 Tockner, K. and Ward, J. V.: Biodiversity along riparian corridors, *Archiv für Hydrobiologie,*
4 *Supplementband, Large Rivers*, 11, 293-310, 1999.
- 5 Ward, J. V., Tockner, K., and Schiemer, F.: Biodiversity of floodplain river ecosystems:
6 ecotones and connectivity, *Regulated Rivers: Research & Management*, 15, 125-139, 1999.
- 7 Warner, R. F.: Floodplain evolution in a New South Wales coastal valley, Australia: spatial
8 process variations, *Geomorphology*, 4, 447-458, 1992.
- 9 Warner, R. F.: Floodplain stripping: another form of adjustment to secular hydrologic regime
10 change in Southeast Australia, *Catena*, 30, 263-282, 1997.
- 11 Whited, D. C., Lorang, M. S., Harner, M. J., Hauer, F. R., Kimball, J. S., and Stanford, J. A.:
12 Climate, hydrologic disturbance, and succession: drivers of floodplain pattern, *Ecology*, 88,
13 940-953, 2007.
- 14 Wiens, J. A.: Riverine landscapes: taking landscape ecology into the water, *Freshwater*
15 *Biology*, 47, 501-515, 2002.
- 16

1 Table 1. Summary of the geographical and climatic settings of the eight study floodplains.

Floodplain name	Valley setting	Climate	Stream network setting
Bidgee	Confined	Semi-arid/temperate	Lowland continuous
Gwydir	Unconfined	Semi-arid/temperate	Lowland terminal
Macquarie	Unconfined	Semi-arid/temperate	Lowland continuous
Mississippi	Confined	Continental	Upland continuous
Narran	Unconfined	Semi-arid	Lowland terminal
Shingwedzi	Confined	Sub-tropical	Upland continuous
Woodforde	Confined	Arid	Headwaters continuous
Yanga	Unconfined	Semi-arid/temperate	Lowland continuous

2

1 Table 2. Summary of the indicators used to calculate the index of Floodplain Surface
 2 Complexity (*FSC*). Averages and standard deviations of the surface metrics (left columns) are
 3 calculated from 50 random sample locations throughout each floodplain. The nugget and
 4 range from the Moran's I spatial correlograms (right columns) are extracted from the
 5 exponential isotropic models fit to these. See Scown et al. (2015a) for detailed calculation
 6 procedures.

Indicators of variability in surface geometry		Indicators of spatial organisation of surface conditions	
Average standard deviation of surface heights	Indicates variability in surface elevation within an area	Spatial correlogram exponential isotropic model nugget ($\times 4$ metrics)	Indicates strength of spatial organisation
Average coefficient of variation of surface heights	Indicates variability in surface elevation relative to the mean elevation within an area	Inverse of the spatial correlogram exponential isotropic model range ($\times 4$ metrics)	Indicates patchiness or fragmentation in spatial organisation
Standard deviation of skewness of surface heights	Indicates variability in erosional and depositional features within an area		
Average standard deviation of surface curvature	Indicates how convoluted the surface is		

7

1 Table 3. Results from regression analyses of *SPO* against *VSG* at each of the three window
2 sizes.

	Best model	F	<i>d.f.</i>	p	r^2
50 m	$y = 0.703x + 0.223$	9.676	1, 7	0.02	0.61
200 m	$y = 1.135x + 0.120$	7.627	1, 7	0.03	0.56
1000 m	$y = 0.329x + 0.429$	0.472	1, 7	0.52	0.07

3

1 Table 4. Results from regression analyses of *FSC*, *VSG*, and *SPO* against *Fpw* at each of the
 2 three window sizes.

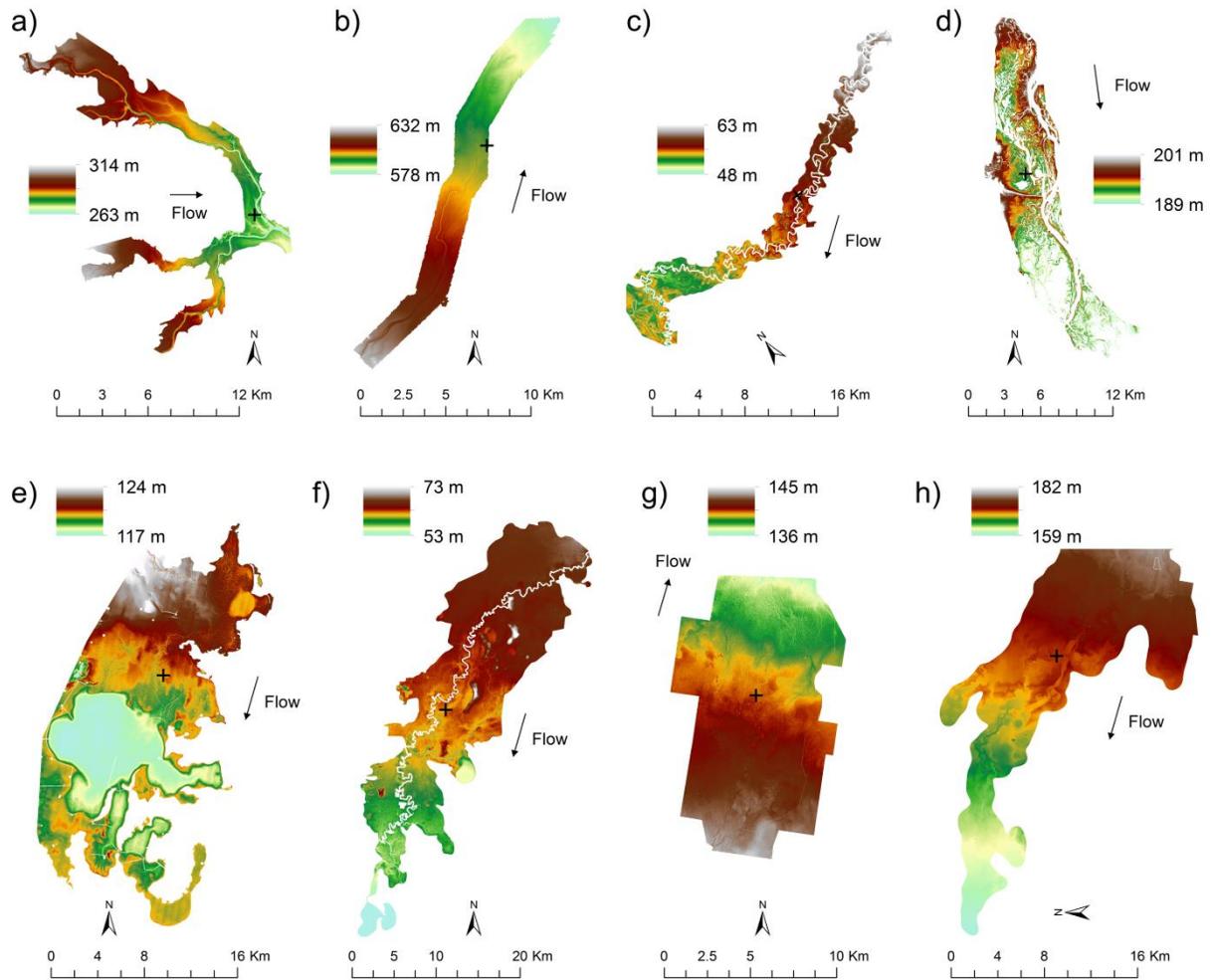
	Best model	F	<i>d.f.</i>	p	r ²
<i>FSC</i>	50 m $y = 0.765x^{-0.414}$	10.344	1, 7	0.02	0.63
	200 m $y = 0.762x^{-0.420}$	25.523	1, 7	0.00	0.81
	1000 m $y = 0.549x^{-0.213}$	5.871	1, 7	0.05	0.50
<i>VSG</i>	50 m $y = -0.151 \ln x + 0.630$	9.642	1, 7	0.02	0.62
	200 m $y = 0.627x^{-0.418}$	26.319	1, 7	0.00	0.81
	1000 m $y = 0.472e^{-0.064x}$	13.574	1, 7	0.01	0.69
<i>SPO</i>	50 m $y = -0.145 \ln x + 0.737$	14.515	1, 7	0.01	0.71
	200 m $y = -0.204 \ln x + 0.866$	20.586	1, 7	0.00	0.77
	1000 m	0.570	1, 7	0.48*	0.09

3

- 1 Table 5. Results from regression analyses of *VSG* against $\log_{10}(SY) + 1$ at each of the three
 2 window sizes with Gwydir removed.

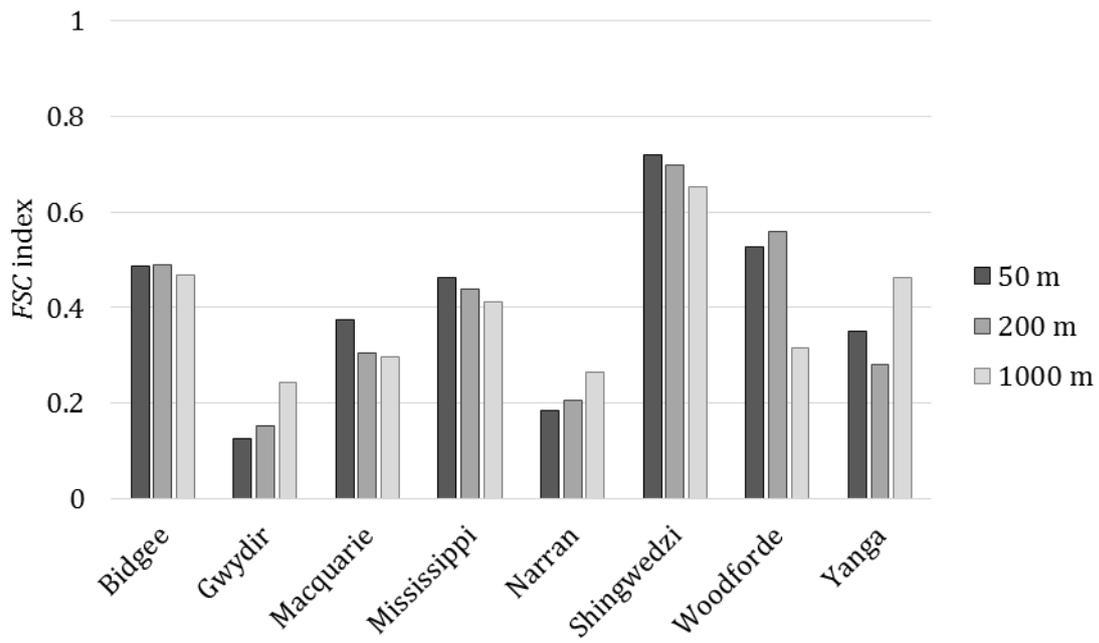
	Best model	F	<i>d.f.</i>	p	r^2
50 m	$y = 0.183x + 0.088$	50.497	1, 6	0.00	0.91
200 m	$y = 0.158x + 0.084$	18.179	1, 6	0.00	0.78
1000 m	$y = 0.142x + 0.088$	36.076	1, 6	0.00	0.88

3



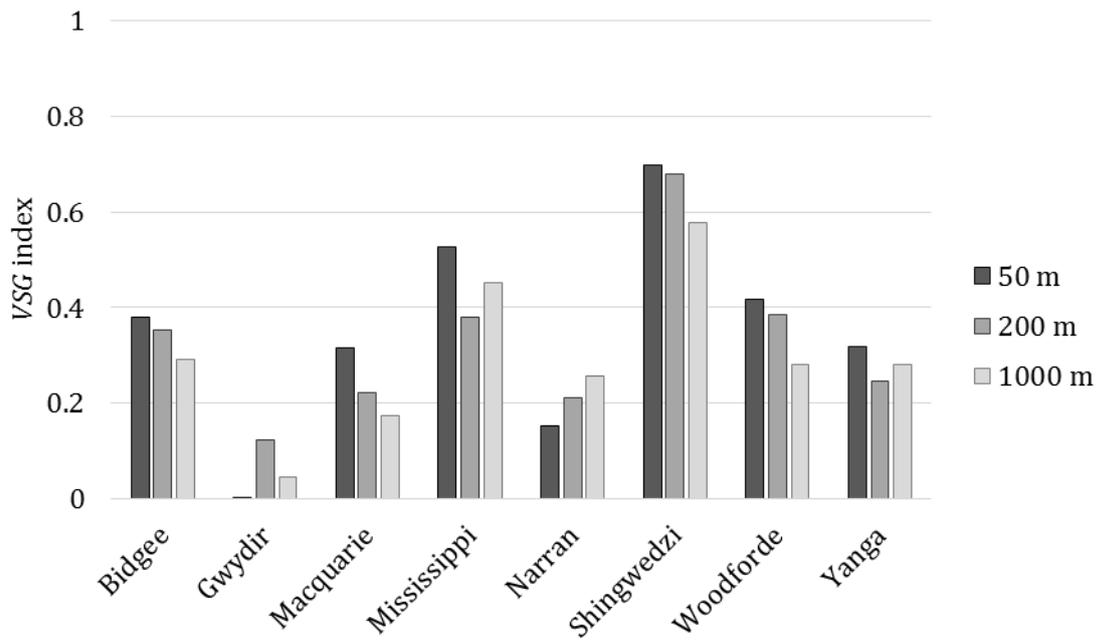
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Figure 1. Digital elevation models displaying the floodplain surface in meters above sea level for each study site (crosses indicate coordinates listed): a) Shingwedzi ($31^{\circ}24'E$, $23^{\circ}05'S$); b) Woodforde ($133^{\circ}20'E$, $22^{\circ}21'S$); c) Bidgee ($143^{\circ}24'E$, $34^{\circ}42'S$); d) Mississippi ($91^{\circ}15'W$, $43^{\circ}29'N$); e) Narran ($147^{\circ}23'E$, $29^{\circ}48'S$); f) Yanga ($143^{\circ}42'E$, $34^{\circ}30'S$); g) Macquarie ($147^{\circ}33'E$, $30^{\circ}41'S$); h) Gwydir ($149^{\circ}20'E$, $29^{\circ}16'S$).



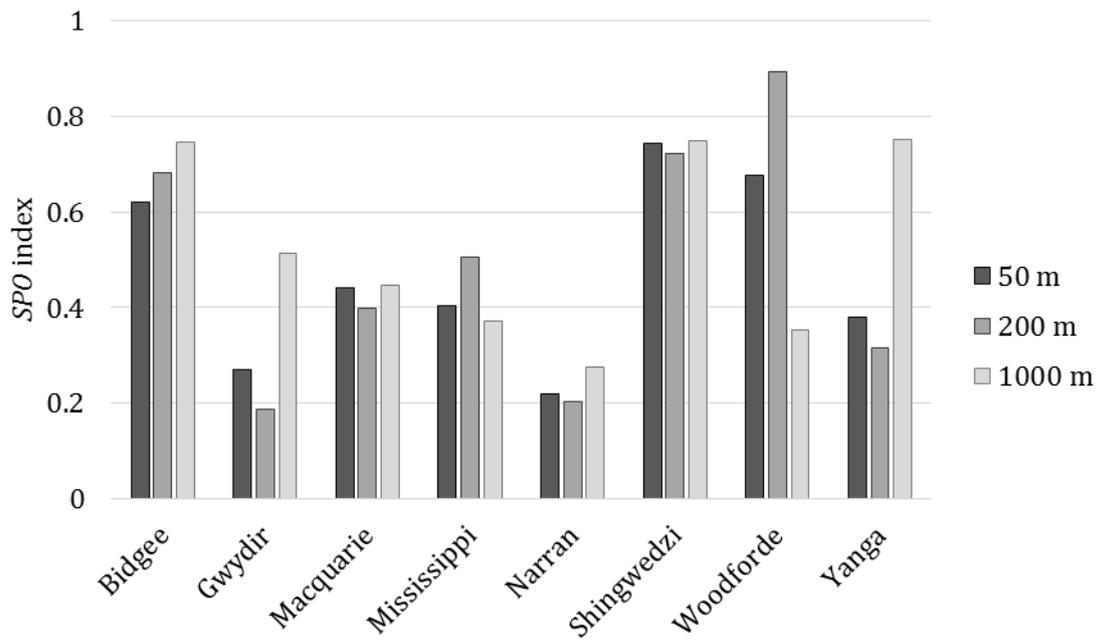
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Figure 2. Index of floodplain surface complexity (*FSC*) for the eight floodplains at each of the three window sizes.



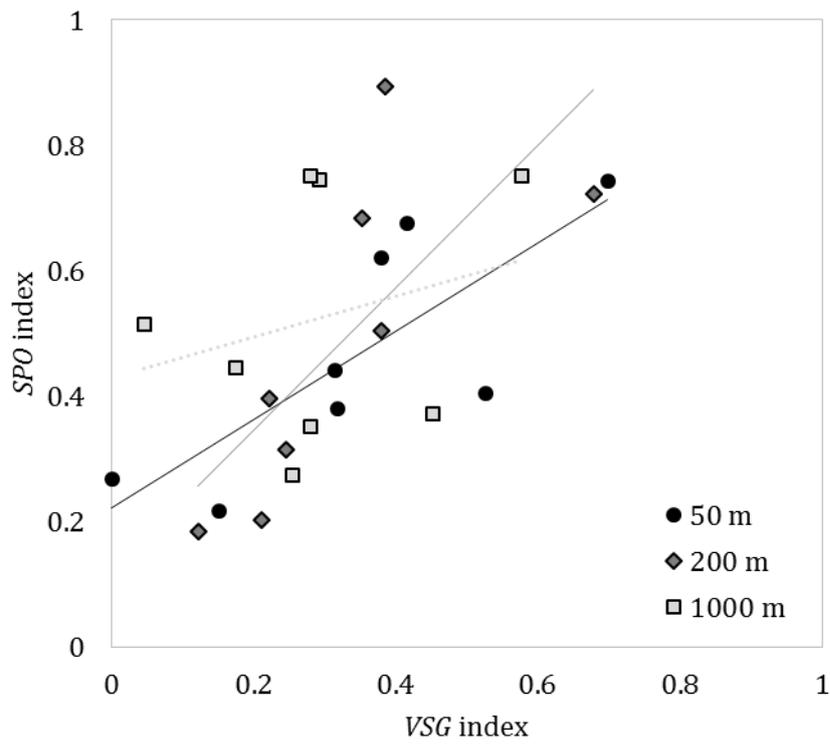
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Figure 3. Index of variability in surface geometry (VSG) for the eight floodplains at each of the three window sizes.



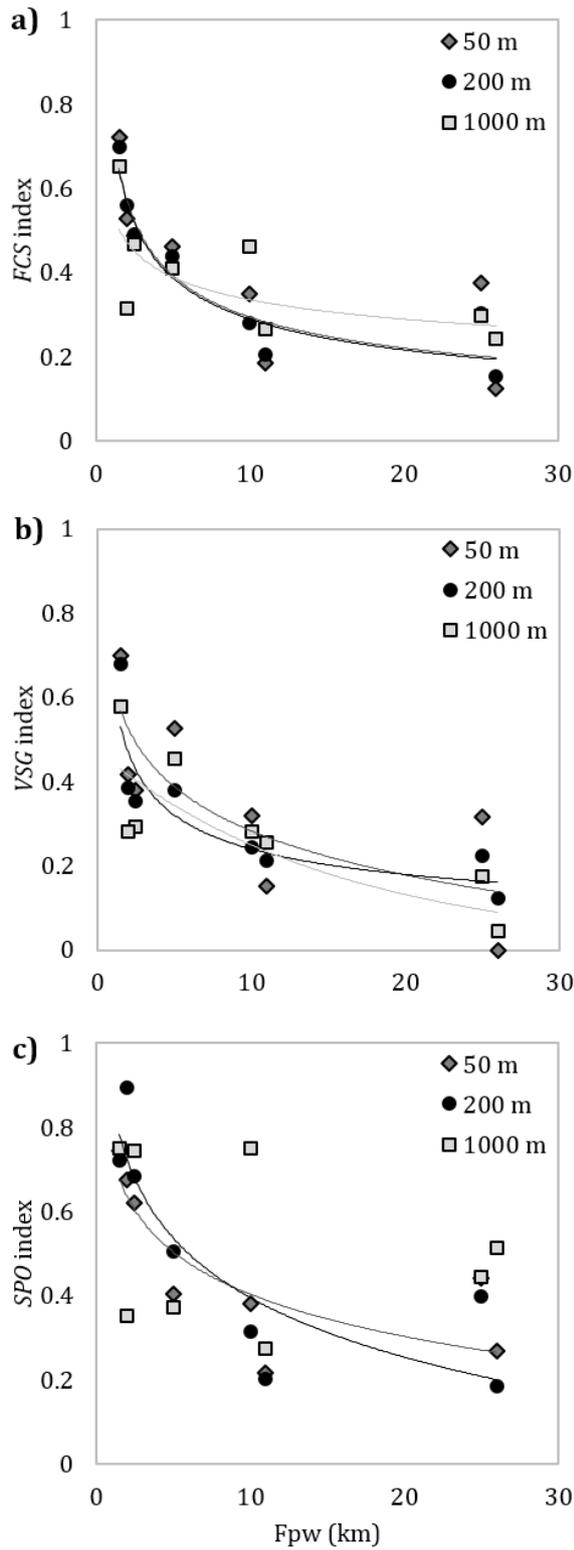
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Figure 4. Index of spatial organisation of surface conditions (*SPO*) for the eight floodplains at each of the three window sizes.



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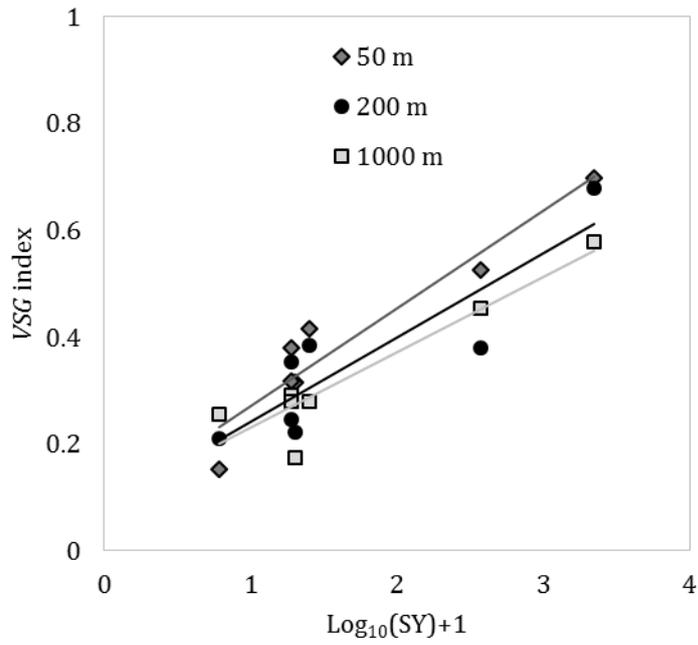
Figure 5. Linear relationships between variability in surface geometry (*VSG*) and spatial organisation of surface conditions (*SPO*) at each of the three window sizes.



1

2 Figure 6. Power relationships between floodplain width (Fpw) and a) floodplain surface
 3 complexity (*FCS*), b) variability of surface geometry (*VSG*), and c) spatial organisation of
 4 surface conditions (*SPO*) at each of the three window sizes.

5



1

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3 Figure 7. Linear relationships between log-transformed SY and variability of surface

4 geometry (VSG) at each of the three window sizes with Gwydir removed.