

Author's response

Comments by the Editor

I am happy with your revisions. However, about the title of the paper, I still suggest that the base flow term included in the title. Your study area is a tropical montane watershed. Although the base flow dominates the streamflow in your specific watershed, it is not guaranteed that it is true for all tropic montane watersheds. If you do not want the base flow term appearing in title, it is reasonable to include your specific watershed in the title.

Reply: We have followed the suggestion. We have included the term baseflow in the title of our manuscript.

1 **Manuscript title**

2 Factors influencing stream water-baseflow transit times in tropical montane watersheds

3

4 **Authors**

5 L.E. Muñoz-Villers^{1*}, D.R. Geissert², F. Holwerda¹, J.J. McDonnell^{3,4}

6

7 ¹ Centro de Ciencias de la Atmósfera, Universidad Nacional Autónoma de México, Ciudad

8 Universitaria, Distrito Federal, Mexico

9 ² Instituto de Ecología, A.C., Red de Ecología Funcional, Xalapa, Veracruz, Mexico

10 ³ Global Institute for Water Security, University of Saskatchewan, Saskatoon, Canada

11 ⁴ School of Geosciences, University of Aberdeen, Aberdeen, UK

12

13

14 ^{*)} Correspondence to: Lyssette E. Muñoz-Villers, Centro de Ciencias de la Atmósfera,

15 Universidad Nacional Autónoma de México, Circuito Exterior s/n, Ciudad Universitaria, 04510

16 Ciudad de México, Mexico. Email: lyssette.munoz@atmosfera.unam.mx, Phone: (52) 55-5622-

17 40-89.

18 **Abstract**

19 Stream water mean transit time (MTT) is a fundamental hydrologic parameter that
20 integrates the distribution of sources, flow paths and storages present in catchments. However, in
21 the tropics little MTT work has been carried out, despite its usefulness for providing important
22 information on watershed functioning at different spatial scales in (largely) ungauged basins. In
23 particular, very few studies have quantified stream MTTs and related these to catchment
24 characteristics in tropical montane regions. Here we examined topographic, land use/cover and
25 soil hydraulic controls on baseflow transit times for nested catchments (0.1–34 km²) within a
26 humid mountainous region, underlain by volcanic soil (Andisols) in central Veracruz (eastern
27 Mexico). We used a 2 year record of bi-weekly isotopic composition of precipitation and stream
28 baseflow to estimate MTT. Land use/cover and topographic parameters (catchment area and
29 form, drainage density, slope gradient and length) were derived from GIS analysis. Soil water
30 retention characteristics, and depth and permeability of the soil-bedrock interface were obtained
31 from intensive field measurements and laboratory analysis. Results showed that baseflow MTT
32 ranged between 1.2 and 2.7 years across the 12 study catchments. Overall, MTTs across scales
33 were mainly controlled by catchment slope and the permeability observed at the soil-bedrock
34 interface. In association with topography, catchment form and the depth to the soil-bedrock
35 interface were also identified as important features influencing baseflow MTTs. The greatest
36 differences in MTTs were found both within groups of small (0.1–1.5 km²) and large (14–34
37 km²) catchments. Interestingly, longest stream MTTs were found in the headwater cloud forest
38 catchments.

39
40 *Key words:* stable isotopes, baseflow, catchment transit times, topography, land cover, Andisols,
41 Mexico

43 **1. Introduction**

44 The demand for fresh water is rapidly increasing in the humid tropics due to population
45 growth. Nevertheless, in these regions —particularly the montane tropics— relative little
46 process-based hydrological studies have been performed to quantify the states, stocks, flows and
47 residence times of water. These areas are especially prone to land degradation and deforestation
48 for conversion to agricultural and pasture lands (Asner et al., 2009). Notably, tropical montane

49 cloud forests (TMCF) are unique and hydrologically important ecosystems (Bruijnzeel, 2004),
50 but are among the world's most threatened terrestrial ecosystems (Cayuela et al., 2006; Hamilton
51 et al., 1995; Pope et al., 2015). Yet the hydrological impacts associated with land degradation
52 and forest conversion at different scales remain poorly understood, thus hampering the
53 development of effective local and regional strategies for water resources protection and
54 management.

55 Stream water mean transit time (MTT) is an important hydrologic metric that integrates
56 the variety of flow paths, storages and runoff sources potentially existing in catchments. In
57 humid temperate environments, MTTs estimated from stable isotopes have been used to broadly
58 characterize the hydrological and biogeochemical behavior of catchments (McDonnell et al.,
59 2010), providing important information on catchment resistance and resilience to climate change
60 scenarios (Carey et al., 2010). In these same environments, significant progress has been made in
61 exploring the linkages between baseflow MTTs and catchment characteristics, and the dominant
62 factors controlling stream MTT variability across scales and regions. For example, McGuire et
63 al. (2005) showed for the first time, the dependence of stream water mean residence time on
64 catchment topographic indices (hillslope length and gradient) for multiple nested watersheds in
65 Western Oregon, USA. Further, Broxton et al. (2009) found that stream water isotope variability
66 and estimated MTTs were both related to watershed aspect and slope in the Valles Caldera
67 watershed, New Mexico, USA. In Central Japan, Asano and Uchida (2012) showed that base
68 flow MTT was mainly controlled by the depth of the hydrologically active layer (i.e. depth of the
69 soil-bedrock interface), which was not necessarily related to catchment topography. Perhaps the
70 most extensive work to date has been done in North East Scotland, where several studies have
71 identified soil properties (soil type and permeability) as the main control on stream MTTs (Geris
72 et al., 2015; Rodgers et al., 2005; Soulsby et al., 2006; Tetzlaff et al., 2009a). With the exception
73 of the investigations carried out by McGlynn et al. (2003) in the Maimai watersheds in New
74 Zealand, and more recently by Hale and McDonnell et al. (2016) in the Alsea Watershed Study
75 in the Oregon Coast Range, USA, which both showed strong positive relations between MTTs
76 and catchment area, most studies to date have shown that landscape evolution and organization
77 dictates rainfall-runoff processes in humid temperate environments.

78 In the humid tropics, isotope-inferred stream MTT studies have provided insights into the
79 hydrological functioning of small forested catchments ($< 0.3 \text{ km}^2$; Muñoz-Villers and

80 McDonnell, 2012), and their sensitivity to land use conversion ($< 1.8 \text{ km}^2$; Roa-García and
81 Weiler, 2010). At larger scales (> 2 to 77 km^2), the studies carried out in Ecuador by Crespo et al.
82 (2012) and Timbe et al. (2014) have reported MTT values for various flowing water bodies
83 (springs, creeks, tributaries and rivers), but as yet, the factors controlling the stream water transit
84 times in this and other montane regions of the humid tropics remain to be explored.

85 Here we build upon previous isotope work at our site in central Veracruz, Mexico, where
86 large water storage capacities have been estimated ($\sim 3 \text{ yr}$) for an old-growth TMCF upland
87 catchment based on baseflow MTT (Muñoz-Villers and McDonnell, 2012). The present study is
88 the first in the humid tropics that we are aware of that explores the relationship between stream
89 water MTT and landscape characteristics across catchments ranging different size areas (from
90 0.1 to 34 km^2). Our tropical montane watersheds are underlain by volcanic soil substrates
91 (Andisols). MTT was determined using a 2 year record of rainfall and stream water isotope data.
92 We used metrics such as land cover, topographic parameters and hydrologic properties of the
93 soil-bedrock profile to identify the factors controlling stream MTTs in this environment.
94 Specifically we addressed the following research questions:

- 95
- 96 1. What are the stream mean transit times across catchment scales?
 - 97 2. How do catchment area, topography and subsurface hydrologic properties relate to
98 stream transit times?
 - 99 3. Does land cover have an effect on stream MTT variability?
 - 100 4. Is there a dominant factor controlling stream water transit times in this mesoscale
101 catchment?

102

103 **2. Materials and Methods**

104 **2.1 Study site**

105 The 5th-order Los Gavilanes (LG) river catchment (41 km^2 ; $19^\circ 28' \text{ N} - 97^\circ 01' \text{ W}$) is
106 located on the eastern (windward) slopes of the Cofre de Perote mountain. It is the main stream
107 water supply for the city of Coatepec and surroundings ($\sim 80,000$ inhabitants). The landscape of
108 this region is complex and strongly dissected by perennial streams draining catchments of
109 different sizes. For this study, 12 catchments were selected, ranging in area from 0.1 to 34 km^2

110 and located between 1300 and 3000 m a.s.l. (Figure 1a). Table 1 summarizes the physical
111 characteristics of the study catchments.

112 The mid and upper parts of the LG catchment (1800–3000 m a.s.l.), where the majority
113 of monitored headwaters are located, are characterized by short steep hillslopes covered mostly
114 by pine-oak forest, mature and secondary tropical montane cloud forest, and pasture (Figure
115 1b,c). The lower portions of the LG catchment (1300–1800 m a.s.l.) are characterized by more
116 gentle terrain covered by pasture, fragments of cloud forests on the steeper slopes and shaded
117 coffee plantations below 1400 m.

118 The general climate in the LG catchment is temperate humid with abundant summer rains
119 (Garcia, 1988). About 80 % of the annual rainfall falls as convective storms during the wet
120 season (May–October), when the region is under the influence of the easterly trade wind flow.
121 Maximum ground water recharge and catchment runoff also occur during this season (cf.
122 Muñoz-Villers and McDonnell, 2013). The relatively dry season (November–April) is
123 characterized by light rains and/or fog and drizzle associated with the passage of cold fronts
124 (Holwerda et al., 2010). Fog interception occurs exclusively during this time of year, and
125 accounts for ≤ 2 % of the annual rainfall for the upper part of the LG catchment (Holwerda et al.,
126 2010; Muñoz-Villers et al., 2012, 2015).

127 The local climate varies markedly with elevation. At 1210 m a.s.l. (lower part of the LG
128 catchment), the annual mean daily temperature is 19 °C. Corresponding mean annual rainfall and
129 reference evapotranspiration (ET_0) are 1385 and 1120 mm, respectively (Holwerda et al., 2013).
130 At 2100 m a.s.l. (middle part), the annual mean daily temperature is 15 °C, and mean annual
131 rainfall and ET_0 are 3185 and 855 mm, respectively (Goldsmith et al., 2012; Muñoz-Villers et
132 al., 2012). Finally, at 3000 m a.s.l. (upper part), mean annual temperatures range between 5 and
133 10 °C, and mean annual rainfall is 1900 mm (SMN, 2014).

134 Andisols derived from volcanic ashes are the dominant type of soil across the LG
135 catchment. These soils are characterized by low bulk density, high permeability, high water
136 retention capacity and high organic matter content (Gomez-Tagle et al., 2011). Soil profiles are
137 usually deep, well developed and multilayered (A, A/B, Bw, Bw/C), with silt loam and silty clay
138 loam as the dominant textures (Gómez-Tagle et al., 2011). The parental material is permeable,
139 consisting of moderately weathered andesitic breccias, underlain, in turn, by semi-permeable

140 saprolite that has been weathered from fractured andesitic-basaltic bedrock (cf. Muñoz-Villers
141 and McDonnell, 2012).

142

143 Figure 1

144 Table 1

145

146 **2.2 Field data collection and analysis**

147 **2.2.1 Rainfall measurements**

148 To quantify daily precipitation and its spatial variation along the altitudinal gradient,
149 rainfall was measured at 1560, 2100 and 2400 m a.s.l. (Figure 1a). For the sites at 1560 m
150 (hereafter RA) and 2100 m (SECP), stand-alone tipping bucket rain gauges of the type RG2M
151 (Onset) and Casella CEL, respectively, were used (both with a resolution of 0.2 mm). For the site
152 at 2400 m, rainfall was measured with an ARG100 tipping bucket rain gauge (Environmental
153 Measurements; 0.2 mm) as part of a meteorological station (TG; Figure 1a). The signals from the
154 stand-alone gauges were stored in an HOBO pendant event logger (Onset), whereas for the
155 gauge in the weather station a CR1000 data logger (Campbell Scientific) was used.
156 Measurements at SECP were made continuously from July 2006 to November 2010, whereas
157 measurements at RA and TG covered the period of isotope sampling (see below).

158

159 **2.2.2 Collection and analysis of rain and stream water samples**

160 To establish the records of $\delta^2\text{H}$ and $\delta^{18}\text{O}$ isotope composition of precipitation and
161 streamflow, samplings during non-storm conditions were carried out over two hydrological years
162 (May 2008 – April 2010). Paired with the tipping bucket rain gauges, samples of bulk rainfall
163 were collected using a sampler consisting of a 95 mm diameter funnel assembled to a 40 mm
164 diameter and 400 mm long transparent collection tube. The tube contained a float to minimize
165 evaporation. In addition, the rain water collector was inserted into a 75 mm diameter PVC pipe
166 wrapped with bubble foil insulation to protect the collected water against sunlight and minimize
167 temperature variations. Rainwater sampling intervals ranged between 1 and 25 days, depending
168 on rainfall amount and frequency. For logistical reasons, rainwater collection at the RA site was
169 only possible from March 2009 to April 2010. The missing isotope data (10 months) were
170 completed using a linear regression with data from the SECP site ($r^2 = 0.95$).

171 Previous hydrological work at our research site has shown that baseflow is the dominant
172 streamflow component (~ 90%). Furthermore, isotope and chemical-based hydrograph
173 separation for a series of storms and for catchments dominated by different land covers showed
174 that rainfall-runoff responses are mainly dominated by ground water sources (Muñoz-Villers and
175 McDonnell, 2012, 2013). For these reasons, we focused our stream sampling on baseflow. Grab
176 samples of base flow were collected every 2 weeks at the outlets of the 12 study catchments.
177 These included nine sampling points representing headwaters up to 4th-order streams (numbers
178 1-9; Figure 1a), two main tributaries of the LG river (10 and 11) and the main river (12).

179 All water samples were collected in 30 ml borosilicate glass vials with polycone sealing
180 cap to prevent evaporation. The samples were analyzed for $\delta^2\text{H}$ and $\delta^{18}\text{O}$ on a laser liquid-water
181 isotope spectrometer (Version 2, Los Gatos Research, Inc.) in the Hillslope and Watershed
182 Hydrology Lab at Oregon State University, USA. The isotope values are expressed in permil
183 (‰) relative to Vienna Standard Mean Ocean Water (VSMOW). The precision of $\delta^2\text{H}$ and $\delta^{18}\text{O}$
184 measurements was 0.3 and 0.1‰, respectively.

185

186 **2.3 Transit time model**

187 Biweekly $\delta^2\text{H}$ signatures of stream water and rainfall were used to estimate base flow
188 mean transit time (MTT) and transit time distribution (TTD) for each of the study catchments.
189 First, 2-week volume-weighted means (VWMs) of rainfall isotope composition were calculated
190 for each of the three sampling sites. Second, we followed the McGuire et al. (2005) approach to
191 compare the average isotope signature of baseflow for each study catchment with that of rainfall,
192 and so to identify the elevation at which most recharge occurs. For this, we calculated 2-yr
193 averages, and determined for each catchment which of the rainfall time series had its average
194 $\delta^2\text{H}$ value closest to the average $\delta^2\text{H}$ baseflow value. The overall mean $\delta^2\text{H}$ base flow value was
195 -44.9 ‰ (range: -50.2 to -41.0 ‰ across all catchments), whereas rainfall at TG, SECP and RA
196 had volume-weighted mean $\delta^2\text{H}$ values of -43.0, -37.6 and -33.5 ‰, respectively. Hence, for all
197 of the study catchments, MTT simulations were carried out using the rain isotope data from
198 either TG or SECP. Further, this approach was supported by the fact that both the TG and SECP
199 sites are located at those elevations in the LG catchment where most groundwater recharge
200 occurs, as determined by previous water balance studies (Muñoz-Villers et al., 2012).

201 The $\delta^2\text{H}$ precipitation data collected over 2 hydrological years (May 2008 – April 2010)
202 may be too short to properly constrain stream base flow MTT estimates (cf. Hrachowitz et al.,
203 2009). However, precipitation at our site shows a marked seasonal pattern (Holwerda et al.,
204 2010; Muñoz-Villers et al., 2012). In addition, rainfall isotope signatures shows a strong and
205 consistent variation with rainfall amount (Goldsmith et al., 2012). Therefore, to generate an
206 artificial warm-up period required for the MTT model simulations, we followed the approach of
207 Hrachowitz et al. (2009) and repeated our measured 2-year rainfall time series 15 times (cf.
208 Muñoz-Villers and McDonnell, 2012). We then used a lumped parameter convolution model to
209 predict the $\delta^2\text{H}$ output for the stream water as a weighted sum of its respective past $\delta^2\text{H}$
210 measured input in precipitation (Maloszewski and Zuber, 1993). Mathematically, the stream
211 water outflow composition at any time, $\delta_{out}(t)$, consisted of past inputs lagged $\delta_{in}(t-\tau)$ and
212 weighted by the transfer function $g(\tau)$, representing its lumped TTD (Maloszewski and Zuber,
213 1982):

214

$$215 \delta_{out}(t) = \int_0^{\infty} g(\tau)\delta_{in}(t - \tau)d\tau \quad (1)$$

216 where τ are the lagged times between the input and output tracer composition. The weighting
217 function or TTD describes the travel time of the water from the ground surface to an outflow
218 location (i.e. the catchment outlet) (McGuire and McDonnell, 2010).

219 In this study, we used the most basic TTD models (Exponential, Gamma and Dispersion),
220 which require only one or two distribution parameters to be optimized and have been
221 successfully applied in other studies (e.g. McGuire et al., 2005). The performance of different
222 TTD functions for each of the 12 study catchments was evaluated using the transfer function
223 hydrograph separation model TRANSEP (McGuire and McDonnell, 2010; Weiler et al., 2003).
224 This model utilizes the Generalized Likelihood Uncertainty Estimation (GLUE) methodology
225 (Freer et al., 1996) based on Monte Carlo simulations to determine the identifiability of the
226 individual parameters. Our Monte Carlo analysis of each TTD consisted of 10,000 runs. Model
227 performance was assessed using the Nash–Sutcliffe efficiency E (Nash and Sutcliffe, 1970),
228 based on the best agreement parameter value, where a value of 1 would indicate a perfect fit.
229 Parameter uncertainty was defined as the range between 10th and 90th percentile value for the
230 best 20% performing parameter sets based on E (McGuire and McDonnell, 2010; Seibert and

231 McDonnell, 2010). The overall performance of the TTD models was evaluated using the root
232 mean square error (RMSE).

233

234 **2.4 Terrain analysis**

235 To evaluate whether landscape characteristics had an influence on base flow MTT,
236 several metrics describing catchment topographic and morphometric features were calculated in
237 ILWIS 3.3, a raster and vector GIS system. Catchment area was obtained by delineating and
238 extracting each catchment boundary using a digital contour elevation map (10 x 10 m
239 resolution). Land cover was obtained from a regional land cover/use raster map (20 x 20 m)
240 elaborated by Muñoz-Villers and López-Blanco (2008), using satellite images and ground truth
241 verification data. For vegetation cover, each catchment was classified in one of the following 4
242 categories: (1) > 90 % covered by TMCF; (2) > 60 % covered by any type of forest (pine-oak,
243 TMCF); (3) > 90 % covered by pasture; and (4) even mixture (~ 50-50 %) of pasture and any
244 type of forest.

245 Catchment form factor (a measure of catchment shape), drainage density, slopes and
246 hillslope length were calculated using topographic maps (scale 1: 20,000) and a 10 x 10 m digital
247 elevation model (DEM). Catchment form factor (R_f) and drainage density (D_d) were calculated
248 following Horton (1932). Hillslope length was obtained as the average distance between
249 catchment ridge top and valley bottom. Horizontal and vertical gradients of each pixel in the
250 DEM were used to calculate the mean and the percentage distribution of slopes in each
251 catchment, using for the latter the following six classes: 0-5; 5-10; 10-20; 20-30; 30-45; > 45°.

252

253 **2.5 Soil sampling and analysis**

254 Field surveys, soil samplings and subsequent laboratory analysis were conducted from
255 May 2011 to May 2012. First, hillslope forms (ridge top, mid and valley bottom) were derived in
256 GIS using topographic analysis algorithms (Jenness, 2006) and then overlaid with catchment
257 boundaries. From the intersection of the polygon units thirty-two soil toposequences were
258 selected, distributed mostly in the mid and lower portions of the LG catchment because access to
259 the upper part was very difficult. At each toposequence, soil auger-holes up to 2.2 m deep were
260 performed from ridge top to valley bottom to determine the organization of soil layers along the

261 hillslope. Soil penetration resistance was also measured down to 2 m using a dynamic cone
262 penetrometer, following the design and method of Herrick and Jones (2002).

263 At selected toposequences, soil profile pits of approximately 1.5 m x 1.5 m x 2 m (length,
264 width and depth, respectively; 43 in total) were excavated for detailed soil descriptions following
265 the method of Schoeneberger et al. (2002). In addition, undisturbed soil core samples ($n = 3$) at
266 the soil-bedrock interface in each soil profile pit were taken to determine saturated hydraulic
267 conductivity (K_s) in the laboratory using the constant-head method. Further, a pedotransfer
268 function, correlating the observed K_s and penetration resistance values, was used to extrapolate
269 K_s of the least permeable layer to the catchment scale.

270 In each soil pit, soil samples from the A and B-horizons (solum) were collected for
271 laboratory analysis. Bulk density was determined from samples taken with cylindrical stainless
272 steel cores of 100 cm³ in each horizon ($n = 3$), and dried at 105 °C until constant weight. For soil
273 moisture content at field capacity, undisturbed samples from 5 x 1 cm rings (diameter and height;
274 $n = 3$) were collected, then weighed after reaching saturation and equilibration (normally within
275 48 hours), and placed in a pressure-plate apparatus at 30 kPa. From water retention at field
276 capacity and bulk density values, the amount of water held in the solum (in mm) was calculated.
277 All laboratory analyses were performed in the Soil Laboratory of the Instituto de Ecología A.C.,
278 Xalapa, Veracruz.

279 Based on the observed range of depths to soil-bedrock interface (DSBI), this variable was
280 divided in 4 classes: very shallow (< 50 cm depth), shallow (> 50-100 cm), moderate deep (100-
281 200 cm) and relatively deep (> 200 cm). Soil K_s and water retention (WR) capacity categories
282 were defined for each site and hillslope sequence. The K_s classes were obtained from the Soil
283 Hydrology Group of the National Engineering Handbook, Part 630 (NRCS-USDA, 2007), and
284 partly modified based on the HOST classification system (Boorman et al., 1995). The WR
285 capacity classes were defined on *ad hoc* ranges. Based on relationships between the soil
286 hydrologic properties and geoforms, the data was extrapolated to the entire LG catchment.

287 Finally, to evaluate differences in the isotopic composition and deuterium excess (d -
288 excess = $\delta^2\text{H} - 8*\delta^{18}\text{O}$; Dansgaard, 1964) values for rainfall and stream water across sites, we
289 used ANOVA t-tests. Statistical relationships between baseflow MTT, depth and permeability at
290 the soil-bedrock interface, soil water retention and landscape characteristics (land cover and
291 topographic variables) were tested using Spearman's rank order correlations. All statistics were

292 evaluated at the 0.05 confidence level and conducted in Sigma plot (version 12, Systat Software
293 Inc.).

294

295 **3. Results**

296 **3.1 Isotopic composition of rainfall and stream water**

297 From May 2008 to April 2010, mean annual precipitation varied from 2670 mm at RA
298 (1560 m), 3476 mm at SECP (2100 m) to 3264 mm at TG (2400 m). The average of 3,476 mm
299 measured at the SECP is somewhat higher (9 %) than the average of $3,185 \pm 305$ (SD) mm
300 measured at that same site between November 2005 and November 2009 (see above).
301 Nevertheless, the rather small difference suggests that the two years of data used in this study are
302 representative of the longer rainfall pattern. For the other sites, no other data than those given
303 here are available.

304 Rainfall showed a clear seasonal pattern, with 80% on average falling during the wet
305 season (May–October). During the same period, a wide range of variation in the biweekly
306 rainfall isotope values was found across elevations. The largest variation (118 ‰ for $\delta^2\text{H}$ and 17
307 ‰ for $\delta^{18}\text{O}$) and most negative (depleted) values (-110 ‰ for $\delta^2\text{H}$ and -16 ‰ for $\delta^{18}\text{O}$) were
308 observed at the highest altitude (2400 m). With decreasing altitude, rainfall isotope values
309 became more positive (enriched) and their range of variation smaller (Figure 2a). However,
310 differences in rainfall isotopic composition among elevations were only suggested for $\delta^{18}\text{O}$ ($p =$
311 0.031). Mean annual deuterium excess (*d*-excess) values of rainfall increased from 15 to 17 ‰
312 with elevation, but differences among sites were not significant ($p \geq 0.05$).

313 Across all sites, $\delta^2\text{H}$ and $\delta^{18}\text{O}$ rainfall signatures in the wet season (-32.3 ± 25.7 ‰ and -
314 5.9 ± 3.2 ‰ SD, respectively) were significantly depleted as compared to the dry season ($-15.6 \pm$
315 13.6 ‰ and -4.3 ± 1.7 ‰; $p \leq 0.05$) (Figure 3). Stream water isotopic composition followed the
316 seasonal pattern as observed for precipitation; however values were damped (range: -49.4 to -
317 41.8 ‰ for $\delta^2\text{H}$ and -8.1 to -7.2 ‰ for $\delta^{18}\text{O}$; Figure 3) as compared to rainfall (range: -106 to 10
318 ‰ for $\delta^2\text{H}$ and -15 to -1 ‰ for $\delta^{18}\text{O}$, respectively). For all stream water sites, differences in
319 isotope composition between the dry and wet seasons were not statistically significant ($p \geq 0.05$).

320 Unlike rainfall, there were no distinct differences in the isotope signatures of the streams,
321 despite the altitude difference of more than 1000 m between the upland headwaters and the
322 downstream LG river tributaries (Figure 2b). During the wet season, *d*-excess values in the

323 streams were rather constant across all sites, whereas during the dry season values were slightly
324 enriched at lower elevations as compared to upper elevations (Table 2).

325 Figure 2c shows that all samples of precipitation and stream water fall along the local
326 meteoric water line (LMWL), with *d*-excess values consistently above 10 ‰ (range of rainfall:
327 10.7 to 24.2 ‰, and streams: 10.4 to 22.5 ‰). The fact that the rainfall and stream baseflow
328 samples had very similar *d*-excess ranges indicates that the catchment water input-output was
329 little affected by evaporation.

330

331 Figure 2

332 Figure 3

333 Table 2

334

335 **3.3 Land cover, topography and soil hydraulic properties**

336 Our GIS analysis showed that approximately 70 % of the LG catchment was covered by
337 some type of forest. Eight out of the 12 study catchments were located in the middle and upper
338 portions of the LG catchment, which are characterized by moderate to steep terrain. These eight
339 catchments ranged in size between 0.1 and 14 km², and were covered predominantly by mature
340 and secondary TMCF (> 50 %; Table 3). Forest was also the dominant land cover in one of the
341 two tributary catchments (Huehueyapan) and in the LG catchment itself. In these two
342 catchments, the upper part was dominated by pine-oak (> 2500 m), the middle part by TMCF and
343 pasture, meanwhile coffee plantations and forest fragments characterized the lower part (< 1400
344 m). Two out of the 12 catchments were dominated by pasture (having areas of 0.1 and 1.5 km²),
345 and only one catchment (1.9 km² of area) was covered by even portions of forest and pasture.

346 Hillslope lengths were shortest (113 m on average) in the smallest catchments (0.1–1.5
347 km²), and longest (273 m) in the largest ones (14–34 km²; Table 3). Slopes of intermediate
348 length (217 m) were found in the 4–9 km² catchments. Mean slope was 32 ± 5° across all
349 catchments. The dominant categories of slopes were 10-20° and 20-30°. Within these groups, the
350 headwater mature and secondary cloud forest catchments (MAT and SEC; ≤ 0.25 km²) showed
351 the highest proportions of the above mentioned slope categories. The pasture headwater
352 catchment (PAS; 0.1 km²) had the highest percentage (46 %) of gentle slopes (0-10°), meanwhile

353 the 20 km² Huehueyapan tributary catchment showed the highest proportion (33 %) of very steep
354 slopes (> 30°).

355 The catchment form factor (R_f) ranged between 0.071 (CATM1 and CATM5) and 0.231
356 (SEC). Drainage density (D_d) ranged from 1.3 to 8.0 km km⁻². Low D_d values (2.4 ± 0.4 km km⁻²)
357 were found at the larger catchments (14-34 km²) whereas higher D_d values (5.3 ± 2.4 km km⁻²)
358 characterized the smaller catchments (0.1-9 km²; Table 3).

359 Soil depth and water retention (WR) capacity of the solum were greatest in hillslopes
360 located in the middle portion of the LG catchment; maximum WR values were observed in the
361 headwater mature (MAT) and secondary (SEC) forest catchments, and in other small catchments
362 < 0.5 km² dominated by TMCF (Category 15; Figure 4a). Catchments with areas of
363 approximately 2 km² were dominated (> 50 %) by soil depths and WR capacities ranging
364 between 1.0 and 1.5 m and 580 and 850 mm, respectively (Category 14). Shallower soil depths
365 (from 0.5 to 1 m) and reduced WR values (from 310 to 510 mm; Category 13) characterized the
366 slope areas (46 % on average) of the larger catchments (9-34 km²). CATM5 showed the highest
367 proportion of area (33 %) covered by very shallow soils and relatively low water retentions
368 (Category 12).

369 Across all sites, the depth to the soil-bedrock interface (DSBI) ranged from 0.5 to more
370 than 2 m, and soil saturated hydraulic conductivity (K_s) at the interface ranged from 1 to nearly
371 40 mm hr⁻¹. However, for the majority of the catchments, the DSBI was between 1.0 and 1.5 m
372 (~ 65 % of the catchment area on average), with corresponding K_s values between 1 and 15 mm
373 hr⁻¹ (Category 2C; Figure 4b). Notably, the SEC was dominated by DSBI between 1 and 2 m
374 (70 % of the catchment area); at some locations in this catchment DSBI was greater than 2 m,
375 with permeabilities at the soil-bedrock interface higher than 36 mm hr⁻¹ (Categories 2A and 1A).
376 In contrast, the Huehueyapan catchment showed the highest percentage of area (30 %) covered
377 by very low DSBI values (0.5 – 1.0 m on average) of all catchments, but K_s ranged from 4 to 36
378 mm hr⁻¹ (Categories 4C and 3B).

379

380 Table 3

381 Figure 4

382

383 **3.2 Stream baseflow MTTs and their relationship with catchments characteristics**

384 Estimated baseflow mean transit time (MTT) ranged between 1.2 and 2.7 years across the
385 12 study catchments. Note the TTD model that we reported for a particular catchment was the
386 one that best fitted the observed baseflow data (Table 4). The root mean square error (RMSE)
387 and Nash-Sutcliffe efficiency value (E) for these model results ranged from 0.8 to 1.5 % (δ^2H)
388 and 0.42 to 0.69, respectively. Table 4 provides further details on the values of the TTD
389 parameters and the uncertainty bounds.

390 Catchment form, slope, land cover and depth to soil-bedrock interface explained each
391 about 50 % of the variance of baseflow MTT across catchments (Figure 5). The positive
392 correlation found between form factor (R_f) and baseflow MTT suggests that catchments with
393 narrow and elongated shapes lead to shorter transit times (Table 5; Figure 5a). Long MTTs were
394 positively correlated with moderately steep catchments (particularly where slopes between 20
395 and 30° predominated; Figure 5c). Conversely, short MTTs were most strongly related to
396 catchments with high proportions of gently slopes (between 5 and 10°). Interestingly, catchments
397 covered by areas with very steep slopes (> 30°) showed very poor correlations with MTTs. Weak
398 correlations were also obtained with catchment drainage density and mean slope length.

399 Soil WR categories determined along the hillslope transects did not explain much of the
400 variation in baseflow MTTs. Instead, a strong and positive relation was observed between MTT
401 and DSBI; specifically longer stream transit times were related to catchment hillslopes where
402 deep (> 2 m) soil-bedrock interfaces dominated (Figure 5f). Conversely, low and negative
403 correlations were obtained with shallower depths (< 1 m; Table 5). Regardless of the DSBI
404 classes, observed K_s values remained generally high across all sites (5–30 mm hr⁻¹; on average).

405 Land cover explained a significant variation of the baseflow MTT (Table 5; Figure 5e);
406 catchments covered by more than 60 % of forests (Categories 1 and 2) had on average longest
407 MTTs (1.9 ± 0.4 (SD) yr) compared to catchments dominated by > 90 % of pasture or evenly
408 mixed covers with pasture and forest (1.5 ± 0.2 yr; Categories 3 and 4).

409 Baseflow MTT showed no relation to catchment area (Table 5; Figure 5b). However, at
410 the smallest scale (< 0.3 km²), major differences in the MTT were found (1.5 to 2.7 yr). At the
411 intermediate scale (4–9 km²), differences in MTTs (1.4 to 1.9 yr) were small among catchments.
412 At the larger scale (> 14 km²), some more variation in the baseflow MTTs was observed (1.2 to 2
413 yr). The 20 km² Huehueyapan showed the shortest baseflow transit times (1.2 yr) compared to

414 other large catchments examined—this was also the lowest MTT estimated across all the study
415 catchments.

416

417 Table 4

418 Table 5

419 Figure 5

420

421 **4. Discussion**

422 **4.1 How do our baseflow MTTs compare to those found in other tropical montane streams?**

423 Our stable isotope data showed that wet season rainfall is the main catchment stream
424 water source in this tropical montane region. This is similar to findings in other humid tropical
425 environments (Crespo et al., 2012; Roa García and Weiler, 2010; Scholl and Murphy, 2014), but
426 contrasts with results for temperate regions where seasonality in flow regime is usually much
427 more pronounced and, consequently, stream water tends to reflect input sources from different
428 seasons (Brooks et al., 2012; Mueller et al., 2013; Peralta-Tapia et al., 2015). Our estimates of
429 base flow transit times ranged between 1.2 and 2.7 years across the 12 study catchments. These
430 rather long transit times suggest deep, and presumably long subsurface flow paths, contributing
431 to sustain catchment baseflows across scales (0.1 to 34 km²) and seasons.

432 Comparing our results with those obtained by Roa-García and Weiler (2010) for three
433 adjacent headwater catchments differing in size (0.6–1.8 km²) and land cover (forest versus
434 pasture) in central-western Colombia, our baseflow MTTs for the cloud forest catchments (~ 2.7
435 yr; 0.1–0.3 km²) are almost twice the value obtained for their forest-dominated catchment (1.4
436 yr). Further, for two pasture-dominated catchments, these authors obtained MTTs that differed
437 considerably (0.1 and 1.4 yr), which they attributed to differences in soil permeability.
438 Furthermore, the relatively short stream MTTs in the Andean catchments were attributed to the
439 relatively low hydraulic conductivities that characterize the volcanic soils (Acrodoxic
440 Hapludans) of that region, limiting rain water percolation and promoting near-surface flow (Roa-
441 García and Weiler, 2010). This contrasts with our sites, where deep subsurface flow rather than
442 shallow lateral flow is the dominant flowpath for runoff generation (Muñoz-Villers and
443 McDonnell, 2012, 2013).

444 In southern Ecuador, Crespo et al. (2012) used a simple sine-wave approach to estimate
445 the baseflow MTTs for a 74 km² nested mesoscale watershed (the San Francisco river basin),
446 underlain mostly by Histosols. They found MTTs on the order of 0.7–0.9 yr for nine cloud forest
447 catchments (1.3–74 km²). Further, for a 0.8 km² pasture catchment, they reported a MTT of 0.8
448 yr. Shallow lateral subsurface flow and high catchment runoff ratios (76–81 %) due to relatively
449 low topsoil and subsurface permeabilities (14–166 mm hr⁻¹) characterized the hydrology of that
450 montane area (Crespo et al., 2012). In contrast, soil hydraulic conductivities at our site were
451 higher (~ 400 mm hr⁻¹ on average across land covers; Muñoz-Villers et al., 2015), leading to
452 lower (annual) rainfall-runoff ratios (35–50 %), and hydrological responses mainly driven by
453 groundwater sources, which likely explain the much larger catchment water storage capacities of
454 our systems.

455 For eight of the catchments in the San Francisco river basin previously investigated by
456 Crespo et al. (2012), Timbe et al. (2014) obtained much higher MTTs values by fitting several
457 TTD models. For seven cloud forest dominated catchments (1.3–77 km²), they reported an
458 average MTT value of 2.1 yr, while for a pasture catchment they obtained a MTT value (3.9 yr)
459 that was twice the average value for the forests. However, the authors did not provide an
460 explanation of why they found longer MTTs and contradictory results (i.e. higher MTT in the
461 pasture than in the forests) compared to the earlier work by Crespo et al. (2012).

462

463 **4.2 Factors determining baseflow MTTs in this tropical montane watershed**

464 It is well known that topography plays an important role in the transit time of water
465 through catchments (Tetzlaff et al., 2009a), particularly in montane environments (cf. McGuire
466 et al., 2005). Our findings are consistent with previous work and show that longest baseflow
467 MTTs are related to rounded shapes of catchment (0.19–0.23), where moderate slope gradients
468 (20–30°) predominate. In contrast, catchments with elongated forms—regardless of their internal
469 slope assemblages—produced the shortest MTT estimates. Our interpretation is that in narrow
470 forms, the hydrological connectivity between hillslopes and the stream is higher than in
471 catchments with more rounded shapes. This in turn would increase the frequency of water table
472 formation and response to precipitation leading to shorter water travel times. Related work on
473 this was carried out by Hrachowitz et al. (2009) in the Scottish Highlands, who evaluated the
474 influence of topography on stream MTT. In their study, form factor ratios and drainage densities

475 were computed for 20 different catchments (< 1 to 35 km²). Their work showed that elongated
476 forms of catchments were roughly distinguished from rounded shapes. Drainage density,
477 however, characterized much better the catchments topography of that region showing a strong
478 and inverse relationship with stream MTTs. They found high drainage density values associated
479 to high percentages of responsive soil cover (peat soils) as rapid water routed via overland flow
480 enhances connectivity between hillslopes and stream channel network. In our study site,
481 drainage density was inversely related to slope length (data not shown) and showed no
482 correlation with soil type as Andisols dominate entirely the hillslopes of our catchments.

483 We also explored the influences of land cover on baseflow MTT. Our findings showed
484 that catchments covered predominantly by forests had longer MTT estimates compared to
485 catchments dominated by pasture. We attributed this to topographic differences among sites
486 more than land cover effects since most forested areas are themselves located on steep terrain.
487 This is supported by results obtained by Muñoz-Villers and McDonnell (2013), who investigated
488 the streamflow dynamics at the mature and secondary TMCF and pasture headwater catchments.
489 They found on average 50 % lower baseflow in the pasture at the end of the dry season compared
490 to forests, explained by lower recharge of subsurface water storages due to smaller catchment
491 gradients (cf. Sayama et al., 2011; Tetzlaff et al., 2009b) in the pasture, and lower surface soil
492 infiltration capacities caused by animal grazing compaction ($30 \pm 14 \text{ mm hr}^{-1}$ versus 696 ± 810
493 mm hr^{-1}). Thus, the fact that forested catchments have steeper slopes and higher topsoil
494 infiltration capacities might be a more likely explanation for their higher subsurface water
495 storage capacities.

496 In general, very few studies have investigated the effect of land cover on catchment
497 stream MTTs. Mueller et al. (2013) studied the influence of shrub cover area on MTTs in four
498 micro catchments in the Swiss Alps. They found that soil and bedrock hydraulic characteristics
499 had a stronger control on stream transit times rather than land cover. High subsurface flow
500 promoted by fast soil water percolation through fractured bedrock, which can contain karstic
501 rock in deeper layers, dominated the catchment water storage, mixing and release in this alpine
502 environment. More recently, Geris et al. (2015) investigated the relative influence of soil type
503 and vegetation cover on storage and transmission processes in a headwater catchment (3.2 km²)
504 in northeast Scotland. Forested and non-forested sites were compared on poorly drained
505 Histosols in riparian zones and freely draining Podzols on steeper hillslopes. Their results

506 showed that soil permeability properties linked to soil type rather than vegetation influences
507 were dominant features on water storage dynamics at the plot and catchment scales.

508 Our study determined the depth and permeability of the soil-bedrock interface through
509 intensive and extensive measurements in the subsurface over numerous hillslope transects across
510 the LG catchment. This is rather unlike most studies that have derived flow path depths and
511 source contributing areas to stream discharge from surface topography based on digital terrain
512 models (Hrachowitz et al., 2010; McGuire et al., 2005; Tetzlaff et al., 2009b) or from
513 geochemical tracers such as SiO₂ (Asano and Uchida, 2012). Our approach showed that
514 hillslopes with deeper soils along with high hydraulic conductivities at the soil-bedrock interface
515 allowed more subsurface water transmission and storage, leading to longer catchment baseflow
516 transit times. In this case, longest stream MTTs (ca. 3 yr) were obtained in the mature and
517 secondary TMCF headwater catchments, associated to their highest percentage of area covered
518 by deep soil-bedrock profiles related in turn to their moderate steep relief, and greatest
519 subsurface permeabilities. Previous work at these sites showed that the very high permeability of
520 the Andisols (1000 mm hr⁻¹ at 0.1 m to 4 mm hr⁻¹ at 1.5 m depth; Karlsen, 2010) and underlying
521 volcanic substrate promote vertical and fast soil water percolation and recharge of deeper
522 sources, as the preferred flow path mechanism controlling catchment water storage and storm
523 runoff responses (Muñoz-Villers and McDonnell, 2013).

524 Across all catchments, the observed range of saturated hydraulic conductivities at the
525 soil-bedrock interface was from 5 to 30 mm hr⁻¹, suggesting little impedance for water to
526 continue percolating vertically below the soil profile and to recharge ground water reservoirs.
527 This could explain the generally long MTTs found across sites (1.8 yr on average). Further, we
528 observed greatest depths to bedrock at mid and ridge top hillslope positions (data not shown).
529 Thus these topographic features seem to be the main contributing areas to subsurface water
530 recharge. While soil water retention capacities were also greatest at mid and ridge top slope
531 positions, they did not explain much of the variation in baseflow MTTs.

532 These findings are partly consistent with those obtained by Asano and Uchida (2012) in
533 central Japan, who examined the baseflow MTT spatial variation for a 4.3 km² forested montane
534 watershed underlain by granitic soils. They used the dampening of the isotopic signal as a proxy
535 for the relative difference in MTTs among locations. They also used dissolved silica as a tracer
536 to identify the contributing depth of the flow path to stream discharge. Their work showed that

537 the depth of hydrologically active soil-bedrock layer was the main factor determining catchment
538 water storage. Longer baseflow MTT were associated to increased flow path contributions
539 related in turn to hillslope length and topography. McGuire et al. (2005) also showed strong
540 correlations between catchment terrain indices (flow path length) and mean stream residence
541 times for seven catchments (0.085–62.4 km²) in the western Cascade Mountains of Oregon,
542 USA, showing that landscape organization was the main factor controlling catchment-scale water
543 transport.

544 While some investigations have reported that catchment area controls the variation in
545 stream MTT (i.e. Hale and McDonnell, 2016; McGlynn et al., 2003), the majority of the work
546 published to date has shown no relation between MTT and catchment size for catchments
547 ranging between 0.1 and 200 km² (Crespo et al., 2012; McGuire et al., 2005; Mueller et al.,
548 2013; Rodgers et al., 2005; Soulsby et al., 2006). Our findings support these latter studies and
549 show that increasing catchment area does not lead to longer mean stream travel times.

550 We also found that baseflow MTTs were more variable in smaller catchments (0.1-1.5
551 km² sizes) where topography imposed its strongest effect (cf. Hrachowitz et al., 2010; Tetzlaff et
552 al., 2009b). Further, longer MTT were found at the forest-dominated headwater catchments (\leq
553 0.25 km²; ~ 3 yr). This is similar to the findings obtained by Timbe et al. (2014) in a tropical
554 montane cloud forest watershed underlain by Histosols in southern Ecuador, who reported longer
555 and larger variation of MTTs in small streams (0.1 – 5 km²; $\sim 3 \pm 1.09$ yr) in comparison to
556 downstream tributaries and main river channels (10 – 77 km²; $\sim 2 \pm 0.08$ yr). At the intermediate
557 scale (4 – 9 km²), our differences in MTTs were small and associated probably to catchment
558 topography. Unexpectedly, MTTs showed a slight convergent pattern at this scale (Figure 5b)
559 (cf. Hrachowitz et al., 2010; Timbe et al., 2014). Beyond this scale (> 9 km²), catchment
560 processes determining streamflow behavior seems to be different. For instance, the Huehueyapan
561 watershed (20 km²) showed the shortest MTT (1.2 yr) across all catchments investigated. We
562 attributed its low water storage capacity to its narrow form, combination of gentle and very steep
563 slope areas (the latter limiting the development of soil on its hillslopes), and shallow depths to
564 bedrock. Alternatively, the main outlet of the LG catchment (35 km²) has an MTT of 2.2 yr that
565 nearly doubled the MTT value of its main tributary (Huehueyapan catchment). This might
566 suggest that runoff processes of smaller catchments does not necessary combine to define MTT

567 at larger scales ($> 14 \text{ km}^2$)(cf. Shaman et al., 2004), probably due to changes in geomorphology,
568 related in turn to past landscape formation of this volcanic area.

569

570 **5. Conclusions**

571 This study provides an important first step towards a better understanding of the
572 hydrology of tropical montane regions and the factors influencing stream water transit times in
573 these environments. Our estimates of baseflow MTT ranged between 1.2 and 2.7 years across 12
574 catchments (0.1 to 34 km^2) in central Veracruz, Mexico, suggesting deep and presumably long
575 subsurface flow paths contributing to sustain baseflows, particularly during dry periods. Our
576 findings showed that catchment slope and the permeability observed at the soil-bedrock interface
577 are the key factors controlling baseflow MTT in this tropical montane region. Longest stream
578 MTTs were found in the cloud forest headwater catchments, related to their moderate steep
579 slopes, and greater transmissivity at the soil-bedrock interface. Conversely, the MTT was
580 shortest in one tributary of the main river outlet, which was mainly attributed to its high
581 proportions of both gentle and very steep slopes. In association with topography, catchment form
582 and the depth to the soil-bedrock interface were also identified as important features influencing
583 baseflow MTT variability across scales. More specifically, longer baseflow MTTs appeared to be
584 related to rounded shapes of catchments and deeper soil-bedrock interfaces. Greatest depths to
585 bedrock were particularly observed in the mid and ridge top hillslope positions, thus these
586 topographic locations seemed to be the main contributing areas for catchment subsurface water
587 recharge. Major differences in MTTs were found both within groups of small (0.1 – 1.5 km^2) and
588 large (14 – 34 km^2) catchments, related mostly to catchment slope and morphology, and to much
589 lesser extent to land cover.

590

591 **Author contribution**

592 L.E.M. and J.J.M. developed the idea of this research. L.E.M., D.R.G. and F.H. collected data.
593 L.E.M. and D.R.G analyzed and interpreted data. L.E.M. wrote the first draft of the manuscript.
594 D.R.G. and F.H. edited and commented on this first draft. D.R.G. and J.J.M. edited and
595 commented on the second draft and the final version.

596

597 **Acknowledgements**

598 We thank the Municipality of Coatepec (Veracruz, Mexico) and the residents of Plan de
599 San Antonio, Loma Alta and Tierra Grande for permitting us to work on their land. Tina Garland
600 and Caroline Patrick are thanked for their assistance in analyzing the water samples for isotopes.
601 Adán Hernández, Sergio Cruz, Luiz Martínez and Enrique Meza are thanked for their great help
602 in the field. We appreciate the valuable comments of two anonymous reviewers. Finally, this
603 research was founded by the US National Science Foundation (NSF/DEB 0746179) grant to H.
604 Asbjornsen, T.E. Dawson and J.J. McDonnell, and by Mexico-CONACyT (CB-106788) grant to
605 D.R. Geissert.

606

607 **References**

608 Asano, Y. and Uchida, T.: Flow path depth is the main controller of mean baseflow transit times
609 in a mountainous catchment, *Water Resour. Res.*, 48, W03512, doi: 10.1029/2011WR010906,
610 2012.

611

612 Asner, G.P., Rudel, T.K., Aide, T.M., Defries, R., and Emerson, R.: A contemporary assessment
613 of change in humid tropical forests, *Cons. Biol.*, 26, 1386-1395, 2009.

614

615 Asbjornsen, H., Manson, R., Scullion, J., Holwerda, F., Muñoz-Villers, L., Alvarado-Barrientos,
616 M.S., Geissert, D., Gomez-Tagle, A., Dawson, T., McDonnell, J., and Bruijnzeel, L.A.:
617 Interactions between payments for hydrologic services, landowner decisions, and
618 ecohydrological consequences: Synergies and disconnection in the cloud forest zone of central
619 Veracruz, Mexico, *Ecol. Soc.*, In review.

620

621 Boorman, D.B., Hollis, J.M., Lilly, A.: Hydrology of soil types: a hydrologically-based
622 classification of the soils of the United Kingdom. *Inst. Hydrol.*, Wallingford, 1995.

623

624 Brooks, J.R., Wigington, P.J., Phillips, D.L., Comeleo, R., and Coulombe R.: Willamette River
625 Basin surface water isoscape ($\delta^{18}\text{O}$ and $\delta^2\text{H}$): temporal changes of source water within the river,
626 *Ecosphere*, 3, 39, 2012.

627

628 Broxton, P.D., Troch, P.A., and Lyon, S.W.: On the role of aspect to quantify water transit times
629 in small mountainous catchments, *Water Resour. Res.*, 45, W08427,
630 doi:10.1029/2008WR007438, 2009.

631

632 Bruijnzeel, L.A.: Hydrological functions of tropical forests: Not seeing the soil for the trees?,
633 *Agr. Ecosyst. Environ.*, 104, 185–228, 2004.

634

635 Carey, S.K., Tetzlaff, D., Seibert, J., Soulsby, C., Buttle, J., Laudon, H., McDonnell, J.,
636 McGuire, K., Caissie, D., Shanley, J., Kennedy, M., Devito, K., and Pomeroy, J.W.: Inter-
637 comparison of hydro-climatic regimes across northern catchments: synchronicity, resistance and
638 resilience, *Hydrol. Process.*, 24, 3591-3602, 2010.

639

640 Cayuela, L., Golicher, D.J., and Rey-Benayas, J.M.: The extent, distribution, and fragmentation
641 of vanishing montane cloud forest in the Highlands of Chiapas, Mexico, *Biotropica*, 38, 544-554,
642 2006.

643

644 Crespo, P., Bücker, A., Feyen, J., Vaché, K.B., Frede, H-G., and Breuer, L.: Preliminary
645 evaluation of the runoff processes in a remote montane cloud forest basin using Mixing Model
646 Analysis and Mean Transit Time, *Hydrol. Process.*, 26, 3896-3910, 2012.

647

648 Dansgaard, W.: Stable Isotopes in Precipitation, *Tellus*, 16, 436-68, 1964.

649

650 Freer, J., Beven, K., and Ambrose, B.: Bayesian estimation of uncertainty in runoff prediction
651 and the value of data: An application of the GLUE approach, *Water Resour. Res.*, 32, 2161-
652 2173, 1996.

653

654 Garcia, E.: Modificaciones al sistema de clasificación climática de Köppen, Offset Larios,
655 México, D. F., México, 217 pp., 1988.

656

657 Geris, J., Tetzlaff, D., McDonnell, J., and Soulsby, C.: The relative role of soil type and tree
658 cover on water storage and transmission in northern headwater catchments, *Hydrol. Process.*, 29,
659 1844-1860, 2015.

660

661 Goldsmith, G.R., L.E. Muñoz-Villers, F. Holwerda, J.J. McDonnell, H. Asbjornsen, T.E.
662 Dawson: Stable isotopes reveal linkages among ecohydrological processes in a seasonally dry
663 tropical montane cloud forest, *Ecohydrology*, 5, 779–790, 2012.

664

665 Gómez-Tagle Jr., C. A., Geissert, D., Pérez-Maqueo, O. M., Marin-Castro, B. E., and Rendon-
666 Lopez, M. B.: Saturated hydraulic conductivity and land use change, new insights to the
667 payments for ecosystem services programs: a case study from a tropical montane cloud forest
668 watershed in Eastern Central Mexico, in: *Developments in Hydraulic Conductivity Research*,
669 edited by: Dikinya, O., InTech, 225–247, 2011.

670

671 Hale, V.C. and McDonnell, J.J.: Effect of bedrock permeability on stream baseflow mean transit
672 time scaling relations: (1) A multi-scale catchment intercomparison, *Water Resour. Res.*, doi:
673 10.1002/2014WR016124, 2016.

674

675 Hamilton, L.S., Juvik, J.O., and Scatena, F.N.: The Puerto Rico tropical montane cloud forest
676 symposium: introduction and workshop synthesis, in: *Tropical Montane Cloud Forests*.
677 *Ecological Studies*, edited by: Hamilton, L.S., Juvik, J.O., Scatena, F.N., Springer, Verlag, New
678 York , 110, 1–16, 1995.

679

680 Herrick, J.E. and Jones, T.L.: A dynamic cone penetrometer for measuring soil penetration
681 resistance, *Soil Sci. Soc. Am. J.*, 66, 1320-1324, 2002.

682

683 Holwerda, F., Bruijnzeel, L.A., Muñoz-Villers, L.E., Equihua, M., and Asbjornsen, H.: 2010.
684 Rainfall and cloud water interception in mature and secondary lower montane cloud forests of
685 central Veracruz, Mexico, *J. of Hydrol.*, 384,84-96, 2010.

686

687 Holwerda, F., Bruijnzeel, L.A., Barradas, V.L., and Cervantes, J.: The water and energy
688 exchange of a shaded coffee plantation in the lower montane cloud forest zone of central
689 Veracruz, Mexico, *Agr. Forest Meteorol.*, 173, 1-13, 2013.

690

691 Horton, R.E.: Drainage-basin characteristics, *EOS T. Am. Geophys. Un.*, 13, 350-361, 1932.

692

693 Hrachowitz, M., Soulsby, C., Tetzlaff, D., Dawson, J.J.C., Dunn, S.M., and Malcolm, I.A.: Using
694 long-term data sets to understand transit times in contrasting headwater catchments, *J. of*
695 *Hydrol.*, 367, 237–248, 2009.

696

697 Hrachowitz, M., Soulsby, C., Tetzlaff, D., and Speed, M.: Catchment transit times and landscape
698 controls—does scale matter?, *Hydrol. Process.*, 24, 117-125, 2010.

699

700 Kirchner, J.W., X. Feng, C. Neal: Catchment-scale advection and dispersion as a mechanism for
701 fractal scaling in stream tracer concentrations, *J. Hydrol.*, 254, 82– 101, 2001.

702

703 Jenness, J.: Topographic Position Index (tpi_jen.avx) extension for ArcView 3.x, v. 1.3a.,
704 Jenness Enterprises, 2006.

705

706 Karlsen, R.: Stormflow processes in a mature tropical montane cloud forest catchment, Coatepec,
707 Veracruz, Mexico, MSc. thesis, VU Univ., Amsterdam, Netherlands, 110 pp., 2010.

708

709 Maloszewski, P. and Zuber, A.: Determining the turnover time of groundwater systems with the
710 aid of environmental tracers. 1. Models and their applicability, *J. of Hydrol.*, 57, 207-231, 1982.

711

712 Maloszewski, P. and Zuber, A.: Tracer experiments in fractured rocks: Matrix diffusion and
713 validity of models, *Water Resour. Res.*, 8, 2723–2735, 1993.

714

715 McDonnell, J.J., McGuire, K., Aggarwal, P., Beven, K.J., Biondi, D., Destouni, G., Dunn, S.,
716 James, A., Kirchner, J., Kraft, P., Lyon, S., Maloszewski, P., Newman, B., Pfister, L., Rinaldo,
717 A., Rodhe, A., Sayama, T., Seibert, J., Solomon, K., Soulsby, C., Stewart, M., Tetzlaff, D.,

718 Tobin, C., Troch, P., Weiler, M., Western, A., Wörman, A., and Wrede, S.: How old is
719 streamwater? Open questions in catchment transit time conceptualization, modelling and
720 analysis, *Hydrol. Process.*, 24, 1745-1754, 2010.
721

722 McGuire, K.J., McDonnell, J.J., Weiler, M., Kendall, C., McGlynn, B.L., Welker, J.M., and
723 Seibert, J.: The role of topography on catchment scale water residence time, *Water Resour. Res.*,
724 41, W05002, doi: 10.1029/2004WR003657, 2005.
725

726 McGuire, K.J. and McDonnell, J.J.: Hydrological connectivity of hillslopes and streams:
727 Characteristic time scales and nonlinearities, *Water Resour. Res.*, 46, W10543,
728 doi:10.1029/2010WR009341, 2010.
729

730 Mueller, M.H., Weingartner, R., and Alewell, C.: Importance of vegetation, topography and flow
731 paths for water transit times of base flow in alpine headwater catchments, *Hydrol. Earth Syst.*
732 *Sc.*, 17, 1661–1679, 2013.
733

734 Muñoz-Villers, L.E., López-Blanco, J.: Land use/cover changes using Landsat TM/ETM images
735 in a tropical and biodiverse mountainous area of central eastern Mexico, *Int. J. Remote Sens.*, 29,
736 71–93, 2008.
737

738 Muñoz-Villers, L.E. and McDonnell, J.J.: Runoff generation in a steep, tropical montane cloud
739 forest catchment on permeable volcanic substrate, *Water Resour. Res.*, 48, W09528, 2012.
740

741 Muñoz-Villers, L.E. and McDonnell, J.J.: Land use change effects on runoff generation in a
742 humid tropical montane cloud forest region, *Hydrol. Earth Syst. Sc.*, 17, 3543–3560, 2013.
743

744 Muñoz-Villers, L.E., Holwerda, F., Alvarado-Barrientos, M.S., Geissert, D., Marín-Castro, B.E.,
745 Gómez-Tagle, A., McDonnell, J.J., Asbjornsen, H., Dawson, T.E., and Bruijnzeel, L.A.: Efectos
746 hidrológicos de la conversión del bosque de niebla en el centro de Veracruz, México, *Bosque*,
747 36, 395–407, 2015.
748

749 Muñoz-Villers, L.E., Holwerda, F., Gómez-Cárdenas, M., Equihua, M., Asbjornsen, H.,
750 Bruijnzeel, L.A., Marín-Castro, B.E., and Tobón, C.: Water balances of old-growth and
751 regenerating montane cloud forests in central Veracruz, Mexico, *J. of Hydrol.*, 462-463, 53–66,
752 2012.

753

754 Nash, J.E. and Sutcliffe, J.V.: River flow forecasting through conceptual models, I, A discussion
755 of principles, *J. of Hydrol.*, 10, 282–290, 1970.

756

757 NRCS-USDA: National Engineering Handbook, Part 630, chap. 7, Hydrologic Soil Groups,
758 2007.

759

760 Peralta-Tapia, A., Sponseller, R.A., Tetzlaff, D., Soulsby, C., and Laudon, H.: Connecting
761 precipitation inputs and soil flow pathways to stream water in contrasting boreal catchments,
762 *Hydrol. Process.*, 29, 3546–3555, 2015.

763

764 Pope, I., Bowen, D., Harbor, J., Shao, G., Zanotti, L., and Burniske, G.: Deforestation of
765 montane cloud forest in the Central Highlands of Guatemala: contributing factors and
766 implications for sustainability in Q’eqchi’ communities, *Int. J. Sust. Dev. World* 22, 201-212,
767 2015.

768

769 Roa-García, M.C. and Weiler, M.: Integrated response and transit time distributions of
770 watersheds by combining hydrograph separation and long-term transit time modeling, *Hydrol.*
771 *Earth Syst. Sc.*, 14, 1537–1549, 2010.

772

773 Rodgers, P., Soulsby, C., Waldron, S., and Tetzlaff, D.: Using stable isotope tracers to assess
774 hydrological flow paths, residence times and landscape influences in a nested mesoscale
775 catchment, *Hydrol. Earth Syst. Sc.*, 9, 139-155, 2005.

776

777 Sayama, T., McDonnell, J.J., Dhakal, A., and Sullivan, K.: How much water can a watershed
778 store?, *Hydrol. Process.*, 25, 3899-3908, 2011.

779

780 Schoeneberger, P.J., Wysocki, D.A., Benham, E.C., and Broderson, W.D.: Field book for
781 describing and sampling soils. Version 2.0. Lincoln, USA, National Soil Survey Center, Natural
782 Resources Conservation Service, USDA, 2002.

783

784 Scholl, M.A. and Murphy, S.F.: Precipitation isotopes link regional climate patterns to water
785 supply in a tropical mountain forest, eastern Puerto Rico, *Water Resour. Res.*, 50, 4305-4322,
786 2014.

787

788 Seibert, J. and McDonnell, J.J.: Land-cover impacts on streamflow: A change- detection
789 modelling approach that incorporates parameter uncertainty, *Hydrolog. Sci. J.*, 55, 316-332,
790 2010.

791

792 Shaman, J., Stieglitz, M., and Burns, D.: Are big basins just the sum of small catchments?,
793 *Hydrol. Process.*, 18, 3195-3206, 2004.

794

795 SMN: <http://smn.cna.gob.mx/>, 2014.

796

797 Soulsby, C., Tetzlaff, D., Rodgers, P., Dunn, S., and Waldron, S.: Runoff processes, stream
798 water residence times and controlling landscape characteristics in a mesoscale catchment: An
799 initial evaluation, *J. of Hydrol.*, 325,197-221, 2006.

800

801 Tetzlaff, D., Seibert, J., McGuire, K.J., Laudon, H., Burn, D.A., Dunn, S.M., and Soulsby, C.:
802 How does landscape structure influence catchment transit time across different
803 geomorphic provinces?, *Hydrol. Process.*, 23, 945-953, 2009a.

804

805 Tetzlaff, D., Seibert, J., and Soulsby, C.: Inter-catchment comparison to assess the influence of
806 topography and soils on catchment transit times in a geomorphic province; the Cairngorm
807 mountains, Scotland, *Hydrol. Process.*, 23,1874-1886, 2009b.

808

809 Timbe, E., Windhorst, D., Crespo, P., Frede, H-G., Feyen, J., and Breuer, L.: Understanding
810 uncertainties when inferring mean transit times of water trough tracer-based lumped-parameter

811 models in Andean tropical montane cloud forest catchments, *Hydrol. Earth Syst. Sc.*, 18, 1503-
812 1523, 2014.

813

814 Weiler, M., McGlynn, B.L., McGuire, K.J., and McDonnell, J.J.: How does rainfall become
815 runoff? A combined tracer and runoff transfer function approach, *Water Resour. Res.*, 39, 1315,
816 2003.

Table 1. Topographic characteristics of the 12 catchments (0.1–34 km²) investigated

#ID	Catchment	Area (km ²)	Stream order	Mean elevation (m a.s.l.)	Elevation range (m a.s.l.)
1	MAT	0.25	1	2160	2020-2300
2	SEC	0.12	2	2130	2040-2220
3	PAS	0.10	1	2400	2320-2480
4	CATM1	0.46	2	2230	1980-2480
5	CATM2	0.62	2	2230	1980-2480
6	CATM3	1.9	3	2380	2000-2760
7	CATM4	1.5	2	2240	1860-2620
8	CATM5	4.1	2	2050	1340-2760
9	CATM6	8.9	4	1980	1340-2620
10	PUENTE ZARAGOZA	13.5	4	2030	1300-2760
11	HUEHUEYAPAN	19.7	4	2120	1300-2940
12	LOS GAVILANES	33.5	5	2120	1300-2940

Table 2. Annual, and wet and dry season means of the isotopic composition of rainfall (3 sites) and stream water (12 sampling locations) plus corresponding values of *d*-excess, as calculated from 2 years of data (April 2008 – May 2010)

Rainfall	VWM Annual			VWM Wet season			VWM Dry season		
	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess
TG (2400 m)	-43.0	-7.5	17.0	-48.2	-8.0	15.8	-23.7	-5.5	20.3
SECP (2100 m)	-37.6	-6.7	16.0	-43.6	-7.4	15.6	-18.9	-4.6	17.9
RA (1560 m)	-33.4	-6.1	15.4	-44.0	-7.4	15.2	-12.2	-3.7	17.4
Catchments	Mean annual			Mean wet season			Mean dry season		
	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess	$\delta^2\text{H}$	$\delta^{18}\text{O}$	<i>d</i> -excess
MAT	-42.5	-7.3	15.9	-43.1	-7.4	16.1	-41.8	-7.2	15.8
SEC	-41.8	-7.2	15.8	-42.5	-7.3	15.9	-40.9	-7.0	15.1
PAS	-47.7	-7.9	15.5	-47.7	-7.9	15.5	-47.6	-7.8	14.8
CATM1	-49.4	-8.1	15.4	-48.9	-8.1	15.9	-50.1	-8.1	14.7
CATM2	-47.2	-7.8	15.2	-47.0	-7.8	15.4	-47.4	-7.8	15.0
CATM3	-46.8	-7.8	15.6	-46.4	-7.7	15.2	-47.4	-7.9	15.8
CATM4	-44.8	-7.5	15.2	-42.5	-7.6	18.3	-44.3	-7.4	14.9
CATM5	-42.5	-7.3	15.9	-42.8	-7.3	15.6	-42.1	-7.3	16.3
CATM6	-41.8	-7.2	15.8	-42.2	-7.2	15.4	-41.3	-7.2	16.3
PUENTE ZARAGOZA	-42.1	-7.3	16.3	-42.5	-7.3	15.9	-41.7	-7.2	15.9
HUEHUEYAPAN	-46.1	-7.7	15.5	-46.6	-7.8	15.8	-45.4	-7.7	16.2
LOS GAVILANES	-43.3	-7.4	15.9	-43.7	-7.4	15.5	-43.0	-7.4	16.2

Table 3. Slope length and gradient, drainage density, form factor and land cover of the study catchments

Catchment	Mean slope length (m)	Drainage density (km km ⁻²)	Form factor (-)	% of cover per slope class						Land cover category and % of vegetation coverage
				0-5°	5-10°	10-20°	20-30°	30-45°	>45°	
MAT	123	4.8	0.222	6	14	35	33	11	1	[1]: 100% mature TMCF
SEC	105	5.8	0.231	6	14	33	31	15	1	[1]: 100% secondary TMCF
PAS	68	7.0	0.164	7	39	36	13	5	0	[3]: 90% pasture; 10% shrubs
CATM1	80	7.9	0.071	6	21	28	26	18	1	[2]: 67% TMCF; 31% pasture; 2% pine-oak forest
CATM2	77	8.0	0.093	6	21	29	26	17	1	[2]: 66% TMCF; 33% pasture; 1% pine-oak forest
CATM3	190	2.6	0.122	7	30	34	17	11	1	[4]: 49% pasture; 36% TMCF; 15% pine-oak forest
CATM4	150	3.2	0.087	6	30	36	17	10	1	[4]: 55% pasture; 34% TMCF
CATM5	208	6.8	0.071	9	23	29	20	17	2	[2]: 55% TMCF; 35% pasture; 6% pine-oak forest
CATM6	225	1.3	0.131	5	15	25	25	25	5	[2]: 65% TMCF; 27% pasture; 5% pine-oak forest
PUENTE ZARAGOZA	235	2.9	0.187	7	17	26	24	22	4	[2]: 62% TMCF; 29% pasture; 5% pine-oak forest
HUEHUEYAPAN	300	2.0	0.134	13	16	18	21	25	8	[2]: 43% TMCF; 29% pine-oak forest; 21% pasture
LOS GAVILANES	285	2.4	0.220	9	16	21	22	20	7	[2]: 51% TMCF; 24% pasture; 20% pine-oak forest

Table 4. Stream baseflow MTTs, and corresponding model parameters and model efficiencies

Catchments	MTT (days)	Model ^a	Model parameters	Model efficiency	
MAT	958	Gamma (α, β)	α, β 0.74 (0.70, 0.85), 1299 (524, 1137)	E 0.53	RMSE ($\delta^2\text{H}$, ‰) 1.5
SEC	975	Gamma (α, β)	0.74 (0.59, 0.93), 1326 (484, 2329)	0.68	1.4
PAS	548	Exponential (τ_m)	τ_m 548 (493, 609)	0.57	1.0
CATM1	531	Exponential (τ_m)	τ_m 531(514, 550)	0.58	1.0
CATM2	636	Dispersion (τ_m, D_p)	τ_m, D_p 636 (463,824) 0.66 (0.44, 0.89)	0.66	1.1
CATM3	624	Dispersion (τ_m, D_p)	τ_m, D_p 624 (536,734) 0.85(0.68, 0.96)	0.45	1.0
CATM4	522	Dispersion (τ_m, D_p)	τ_m, D_p 522 (451,571) 2.2 (1.4, 3.0)	0.53	1.4
CATM5	710	Exponential (τ_m)	τ_m 710 (555, 859)	0.63	0.8
CATM6	702	Exponential (τ_m)	τ_m 702 (550, 856)	0.64	0.9

PUENTE ZARAGOZA	633	Exponential (τ_m)	τ_m 633 (520, 751)	0.64	0.9
HUEHUEYAPAN	424	Exponential (τ_m)	τ_m 424 (371, 482)	0.63	1.2
LOS GAVILANES	788	Exponential (τ_m)	τ_m 788 (646, 935)	0.42	1.0

MTT is the mean transit time, E is the Nash-Sutcliffe efficiency and RMSE is the root mean square error. Numbers in parenthesis are the 10th and 90th percentile values of the MTT estimates and the model parameters.
^a Refer to the supplementary information for the corresponding formulas of the TTD models.

Table 5. Spearman's rank correlation coefficients (r_s) between baseflow MTT and land cover, catchment area, topographic characteristics and subsurface hydrologic properties for the study catchments.

	r_s
Land cover	-0.74
Area (km ²)	-0.09
Form factor (-)	0.56
Drainage density (km km ⁻²)	0.05
Mean slope length (m)	-0.13
Slope 0-5°	-0.22
Slope 5-10°	-0.63
Slope 10-20°	-0.01
Slope 20-30°	0.57
Slope 30-45°	0.04
Slope > 45°	0.06
DSBI > 200 cm	0.48
100 < DSBI ≤ 200 cm	-0.28
50 < DSBI ≤ 100 cm	-0.15
DSBI ≤ 50 cm	-0.08
Soil WR per category	
11	-0.08
12	0.24
13	-0.18
14	0.30
15	-0.25

Figures and captions

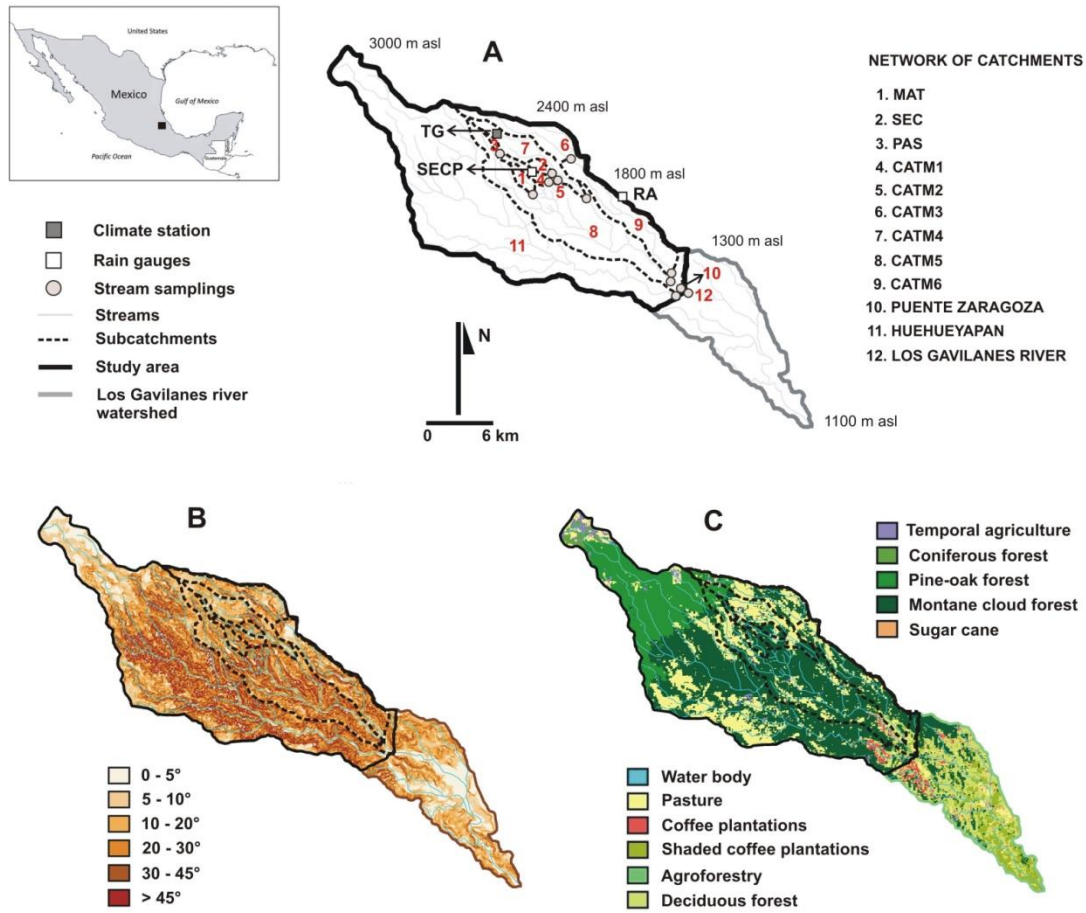


Figure 1. Location of the study site in central Veracruz, eastern Mexico, and maps of the Los Gavilanes catchment showing (A) the stream and rain water collection points; (B) slopes; and (C) land covers (see text for further explanation).

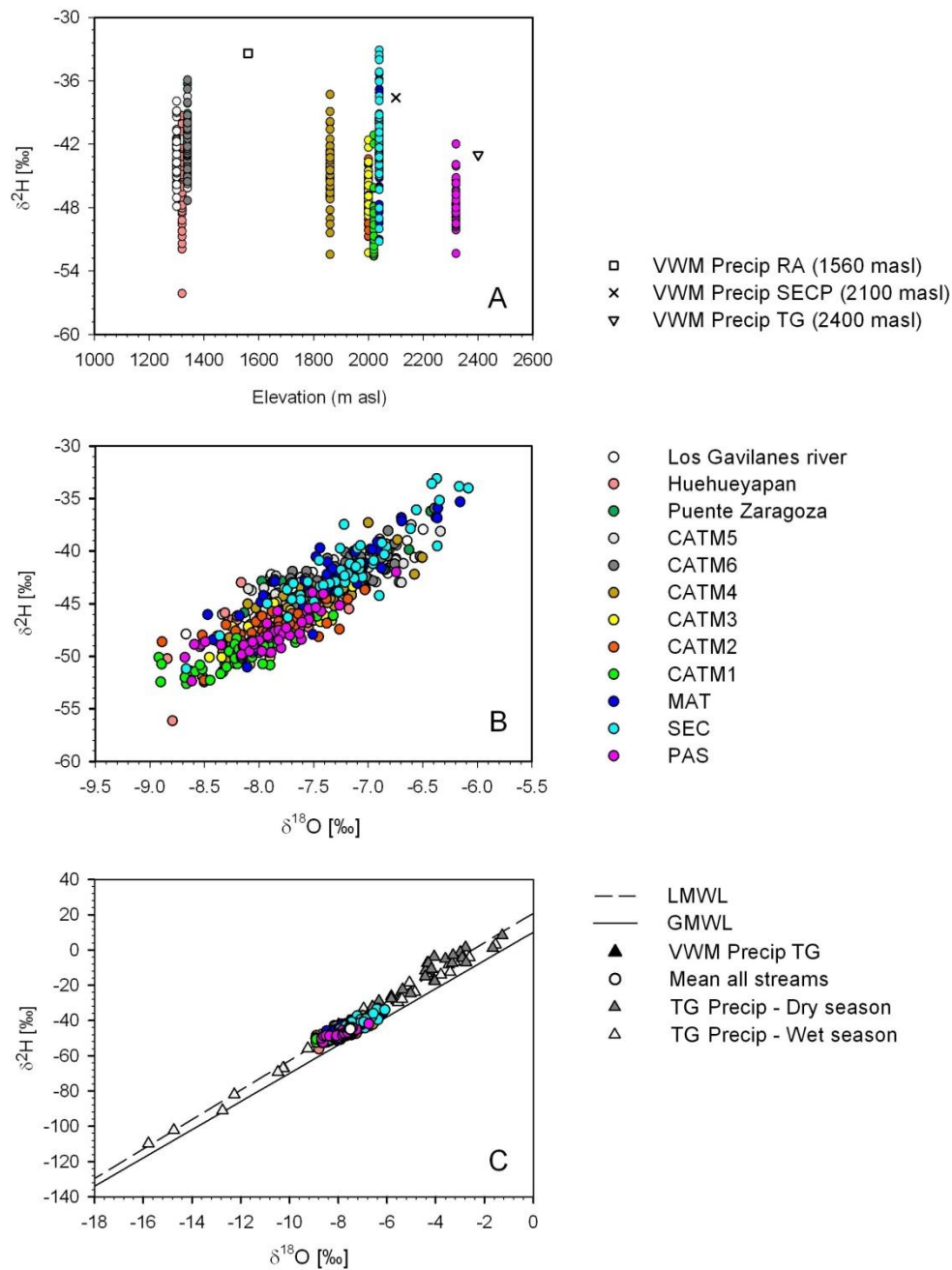


Figure 2. (A) Relationship between $\delta^2\text{H}$ signatures of stream base flow and elevation of the catchment outlets (i.e. the sampling locations), and volume-weighted means (VWMs) of the deuterium composition of rainfall at the three elevations within the Los Gavilanes river catchment; (B) $\delta^{18}\text{O}$ versus $\delta^2\text{H}$ signatures of baseflow from the 12 catchments investigated; and (C) Isotope ($\delta^{18}\text{O}$ versus $\delta^2\text{H}$) signatures of rainfall and stream baseflow. The local meteoric water line (LMWL; dashed line) is based on the 2008-2010 precipitation at TG, and reads: $\delta^2\text{H}=8.36* \delta^{18}\text{O} + 20.37$; the solid line represents the global meteoric water line (GMWL): $\delta^2\text{H}=8* \delta^{18}\text{O} + 10$.

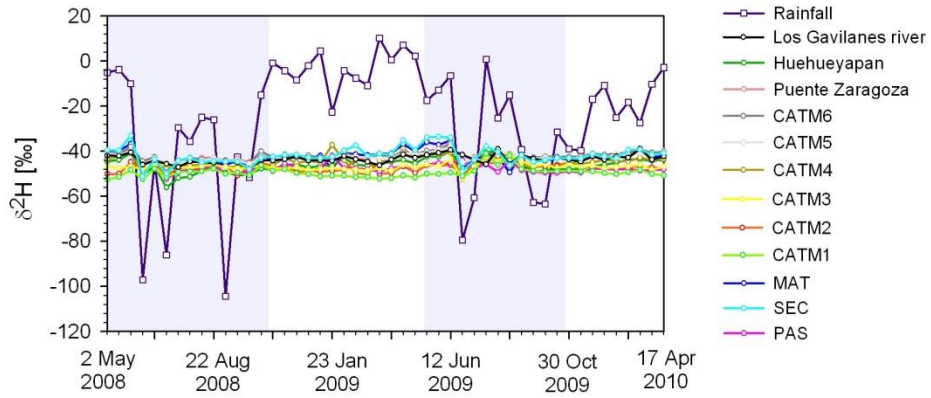


Figure 3. Biweekly values of $\delta^2\text{H}$ composition of stream baseflow for each of the 12 study catchments, and corresponding values of deuterium composition of rainfall at 2400 m (TG) for the period between May 2008 and April 2010. The shaded areas indicate the wet seasons.

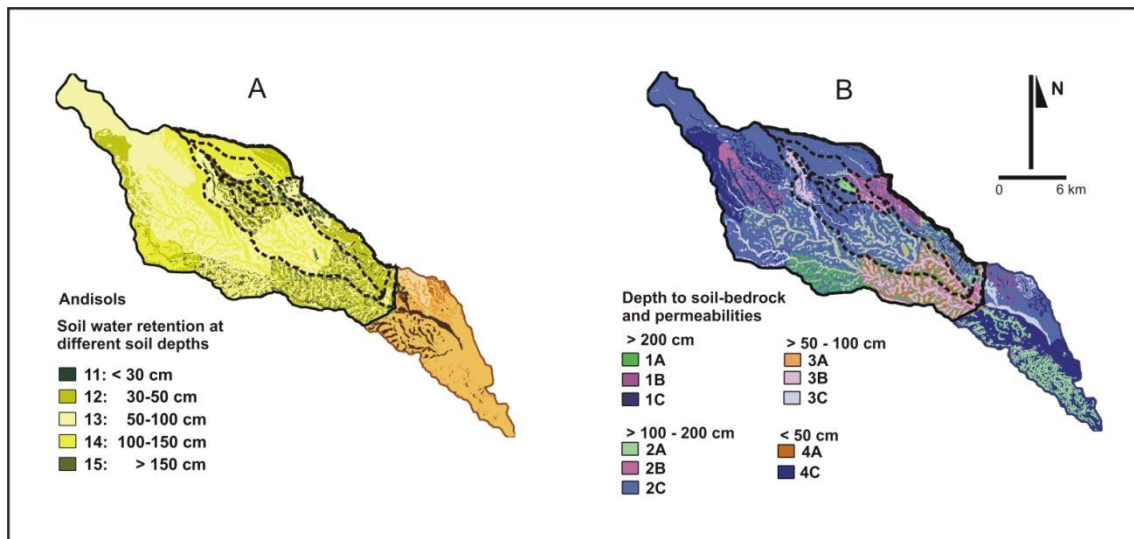


Figure 4. Map of hydro-pedologic properties of the Los Gavilanes river catchment. (A) Soil water retention at field capacity in the solum. Category 11: < 180 mm; Category 12: $\geq 180 \leq 310$ mm; Category 13: $\geq 310 \leq 580$ mm; Category 14: $\geq 580 \leq 850$ mm; and Category 15: ≥ 850 mm. (B) Depth to soil-bedrock interface and corresponding saturated hydraulic conductivities (K_s). For depth < 100 cm, K_s categories A, B and C correspond to: $K_s > 36$; $14 < K_s \leq 36$ and $1 < K_s \leq 14$ mm hr⁻¹, respectively. For depth > 100 cm, A, B and C correspond to $K_s > 144$; $36 < K_s \leq 144$ and $4 < K_s \leq 36$ mm hr⁻¹, respectively.

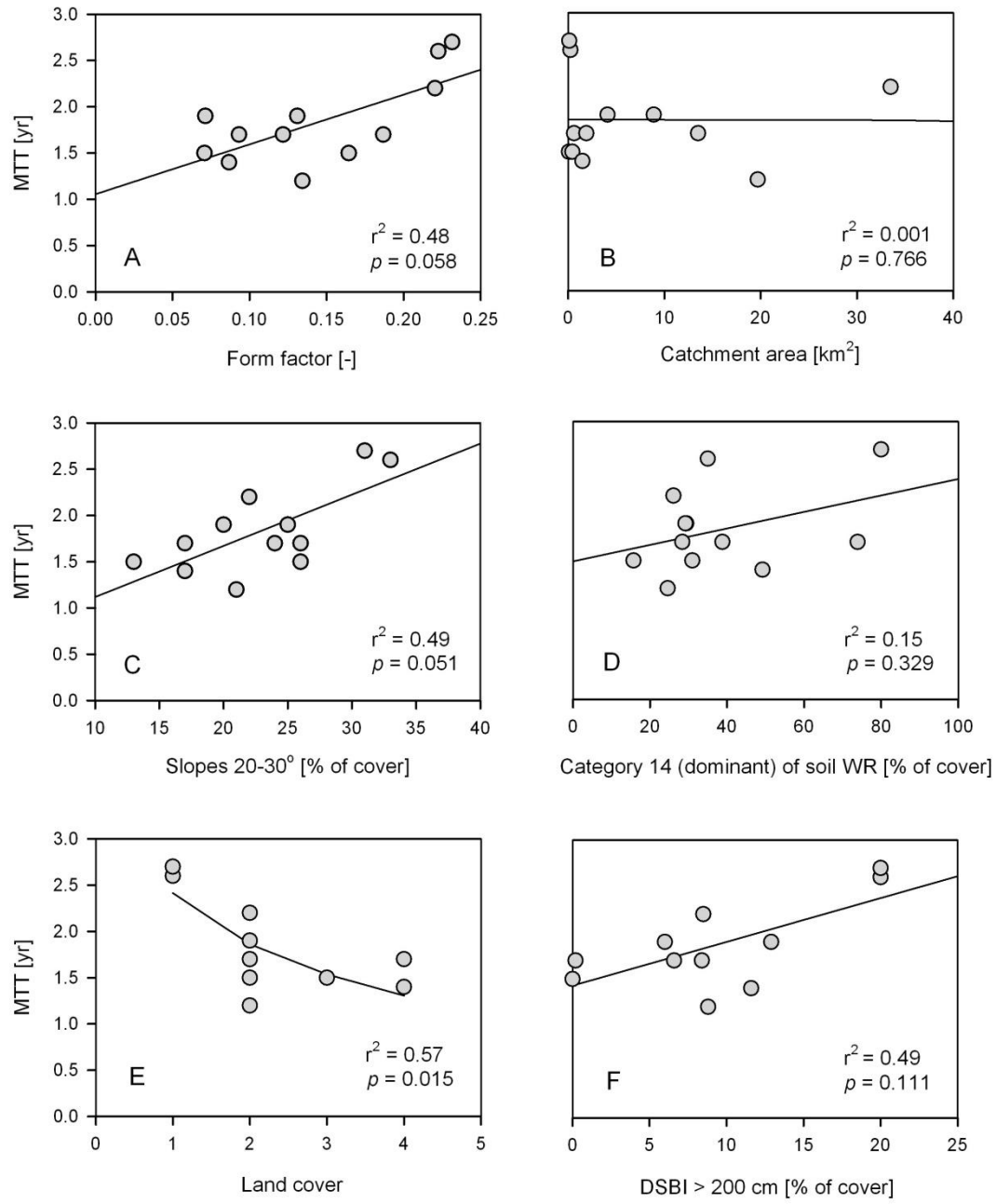


Figure 5. Regressions between stream baseflow MTTs and topographic features, subsurface properties, land cover and catchment area for the study catchments.