

1 Modelling hyporheic processes for regulated rivers under transient hydrological and
2 hydrogeological conditions

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1 Abstract

2 Understanding the effects of major hydrogeological controls on hyporheic exchange and bank
3 storage is essential for river water management, groundwater abstraction, restoration and
4 ecosystem sustainability. Analytical models cannot adequately represent complex settings
5 with, for example, transient boundary conditions, varying geometry of surface water-
6 groundwater interface, unsaturated and overland flow, etc. To understand the influence of
7 parameters such as (1) sloping river banks, (2) varying hydraulic conductivity of the riverbed
8 and (3) different river discharge wave scenarios on hyporheic exchange characteristics such as
9 (a) bank storage, (b) return flows and (c) residence time, a 2-D hydrogeological conceptual
10 model and, subsequently, an adequate numerical model were developed. The numerical model
11 was calibrated against observations in the aquifer adjacent to the hydropower regulated Lule
12 River, Northern Sweden, which has predominantly diurnal discharge fluctuations during
13 summer and long-lasting discharge peaks during autumn and winter. Modelling results
14 revealed that bank storage increased with river wave amplitude, wave duration and smaller
15 slope of the river bank, while maximum exchange flux decreased with wave duration. When a
16 homogeneous clogging layer covered the entire river-aquifer interface, hydraulic conductivity
17 positively affected bank storage. The presence of a clogging layer with hydraulic conductivity
18 $<0.001 \text{ m d}^{-1}$ significantly reduced the exchange flows and virtually eliminated bank storage.
19 The bank storage return/fill time ratio was positively related to wave amplitude and the
20 hydraulic conductivity of the interface and negatively to wave duration and bank slope.
21 Discharge oscillations with short duration and small amplitude decreased bank storage and,
22 therefore, the hyporheic exchange, which has implications for solute fluxes, redox conditions
23 and the [potential of riverbeds as fish spawning locations](#)~~spawning potential of riverbeds~~.
24 Based on these results, river regulation strategies can be improved by considering the effect of
25 certain wave event configurations on hyporheic exchange to ensure harmonious
26 hydrogeochemical functioning of the river-aquifer interfaces and related ecosystems.

27 Keywords: river-aquifer interaction, hyporheic zone, bank storage, regulated rivers,
28 groundwater dependent ecosystem, modelling

29 1 Introduction

30 Surface water-groundwater interfaces have recently received growing research interest
31 (Sophocleous, 2002) and have become the focus of multiple water resources management
32 policies (Klöve et al., 2011). The hyporheic zone that harbours river-aquifer interactions plays
33 a key role in riverine and riparian ecosystem functioning (e.g. Krause et al., 2011). The
34 reactive nature of this zone maintains exchange and transformation of solutes along the
35 pathways between surface water and groundwater. The hyporheic zone is an appreciated
36 habitat for hyporheos, microorganisms and bacteria occupying the space below and along the
37 river channel (Boulton et al., 1998). Besides having negative impacts on the ecosystem of the
38 zone itself, with changes in hyporheic water composition (Calles et al., 2007; [Siergieiev et al.,
39 2014c](#)), alteration of the hyporheic functionality due to surface water-aquifer disconnection
40 can also severely modify neighbouring ecosystems. [As a result, riparian zones and adjacent
41 wetlands may experience changes in groundwater table, water and nutrient fluxes.](#)

1 | ~~Furthermore, R~~restricted hyporheic exchange limits mobilisation of solutes from the riparian
2 | zone and their fluxes into the river during key hydrological events as a result of river-aquifer
3 | disconnection (Burt and Pinay, 2005), which can ~~further-in turn~~ affect surface water quality
4 | (Valett et al., 1996).

5 | A large number of rivers worldwide are obstructed by dams (Nilsson et al., 2005) and are
6 | therefore subject to artificial discharge fluctuations. These fluctuations stress hyporheic
7 | exchange flows, which often results in degradation of the river-aquifer continuum. A major
8 | impact is the alteration of river sediment transport and, subsequently, increased colmation of
9 | the river-bed (Brunke and Gonser, 1997; Blaschke et al., 2003), which deteriorates river-
10 | aquifer hydraulic connectivity (Burt and Pinay, 2005) and controls functional changes in the
11 | hyporheic zone (Siergieiev et al., 2014c). In addition, inundation of the river banks by
12 | construction of ~~run-of-river~~-reservoirs changes the shape of the river-aquifer interface,
13 | successively affecting the hyporheic exchange (Doble et al., 2012a).

14 | Therefore, it is necessary to understand the hydrogeological functioning of this interface
15 | under artificial conditions, such as hydropower regulated rivers, in order to incorporate this
16 | knowledge into water resource management and thereby improve the functional behaviour of
17 | the hyporheic zone by optimising river discharge strategies (e.g. Hanrahan, 2008).

18 | The hyporheic zone size, bank storage volume and bank fluxes vary with river stage
19 | fluctuations (amplitude, duration) and river bank conditions (slope, hydraulic conductivity).
20 | Todd (1955) provided a first theoretical analysis of flood-induced bank storage. This work
21 | served as the foundation for later analytical solutions and numerical simulations of floodplain
22 | hydrology. The dynamics of bank storage were later estimated using analytical models (e.g.
23 | Cooper and Rorabaugh, 1963) based on the following simplifications: single flood wave,
24 | homogeneous aquifer and a fully penetrating vertical river bank. During base flow conditions,
25 | hydrological river-aquifer interactions can be described by precipitation-runoff models
26 | (Butturini et al., 2002). However, these often neglect the distributed effects of e.g. unsaturated
27 | zone processes or topography, resulting in residual unexplained variability in bank storage.
28 | Chen and Chen (2003) used a numerical approach to simulate the bank storage response to
29 | changes in river stage and riverbed hydraulic conductivity for a partially penetrating river
30 | with vertical banks in a fully saturated aquifer. They pointed out the importance of subsurface
31 | anisotropy, which governs the directions of hyporheic zone development. Simultaneous
32 | consideration of seepage and a variably saturated aquifer showed that unsaturated zone
33 | processes have a pronounced effect on bank storage (Li et al., 2008; Doble et al., 2012a).
34 | Inclusion of the unsaturated zone in bank storage simulations decreased the modelled storage
35 | and improved the return flows (Doble et al., 2012b). Furthermore, models that consider
36 | vertical river banks for sloping banks under-estimate bank storage (Doble et al., 2012a).
37 | Understanding of bank storage processes can improve hyporheic ecotone and river-aquifer
38 | continuum concepts, with positive implications for e.g. base flow separation techniques
39 | (McCallum et al., 2010) and ecosystem sustainability (Schneider et al., 2011).

40 | For several seasons, hyporheic exchange was studied in the hydropower regulated Lule River,
41 | Northern Sweden (Siergieiev et al., 2014a; Siergieiev et al., 2014c). Low hydraulic

1 conductivity of the riverbed and daily varying river discharge have resulted in depleted
2 hyporheic exchange flows across the river-aquifer interface (Siergieiev et al., 2014a).
3 Deteriorated river water quality as a result of regulation (Smedberg et al., 2009; Siergieiev et
4 al., 2014b) may partly depend on suppression of hyporheic processes due to regulation (Valett
5 et al., 1996). Improved understanding of the major hydrogeological controls of hyporheic
6 exchange has legacy effects on understanding geochemical fluxes between surface water and
7 groundwater (Fritz and Arntzen, 2007) and can provide a platform for implementation of
8 environmental flows and improved management and ecological status of regulated rivers.

9 The aim of this study was therefore to provide a set of scenarios with variable river discharge
10 schemes (wave duration and amplitude), riverbed slope and riverbed hydraulic conductivity,
11 in order to investigate the effects of these parameters on fluxes across the river-aquifer
12 interface, bank storage volume, fill/return time ratio and residence time. A realistic case study
13 was used to setup a conceptual and a numerical model and justify the use of the method for
14 further scenario simulations.

15 **2 Site description**

16 The measurement profile orthogonal to the river included an observation station in the river
17 and two groundwater wells (Fig. 1) with hourly registration of water level during 2010-2011
18 and every 15 min during 2012. The riverbed at the site slopes gently towards the middle of the
19 channel and is composed of silty-clayey material with ~~vertical~~ layers of highly conductive
20 stratum and laterally spread sparse patches of sand and gravel. The former floodplain is
21 limited by an earth embankment from the river side and stretches for over 150 m where it
22 meets the foot of the hillslope. The depth to the bedrock varies from none to 60 m with an
23 average depth of around 10 m and thinnest overburden towards the hillslope. The flat
24 topography of the floodplain and the hillslope orientation which is normal to the river suggest
25 orthogonal groundwater flow towards the channel. Orientation of the groundwater well profile
26 was chosen accordingly.

27 **3 Materials and methods**

28 **3.1 Data collection**

29 The soil was visually inspected during installation of groundwater wells. Samples were
30 collected at 0.3 m interval and sieving analysis was performed on three samples from each of
31 the two locations. Unsaturated flow parameters were estimated on these six selected samples
32 using pressure pot experiments ~~that to demonstrated obtain~~ water-holding characteristics
33 (Ehlert, 2014). To assess saturated hydraulic conductivity, repeatable slug tests (three in each
34 well) using both falling and rising hydraulic head were carried out in the wells, while a direct
35 push piezometer (two repeatable tests at two locations: 1 and 2.5 m from the shoreline) using
36 a falling head was ~~applied performed~~ at the riverbed at 0.1 m depth interval down to 0.7 m
37 depth (Siergieiev et al., 2014a). A harmonic mean over the top 0.3 m least permeable layers
38 was used as hydraulic conductivity of the clogging layer in the model.

39 **3.2 Conceptual model**

1 The data collected at the site did not allow development of a highly distributed model.
2 Parameter values obtained in the field and in the laboratory were therefore averaged for the
3 saturated and unsaturated zone and the clogging layer. The vertical 2-D conceptual model
4 considered a homogeneous aquifer, partially penetrating river, sloping banks, unsaturated
5 zone and clogging layer that covered the entire river-aquifer interface (Fig. 2). The Dirichlet
6 boundary condition at the riverside was varied according to the measured water level time
7 series. The top of the model was represented by a constant flux boundary to consider
8 recharge, which was assumed to be 50% of annual precipitation (Lemmelä, 1990) of 470 mm
9 yr^{-1} , resulting in $6.5 \times 10^{-4} \text{ m d}^{-1}$ recharge. No flow boundaries were assigned to the remaining
10 borders. The distance to the right boundary was set to ensure that influences of river stage
11 fluctuations did not reach this boundary for the longest fluctuation period. A first
12 approximation of the distance of influence based on an analytical solution (Sawyer et al.,
13 2009) assumed negligible riverbed resistance and is therefore questionable in the present case
14 (e.g. Singh, 2004). This estimate of the maximum extent (180 m for a one-month fluctuation
15 period) was further tested using the numerical model.

16 The following assumptions were used in the model:

- 17 - Two-dimensional model space
- 18 - Simplified geometry, neglecting microtopography
- 19 - Constant recharge, representing both groundwater recharge and regional gradient
- 20 - Isotropic and homogeneous aquifer and clogging layer
- 21 - The same unsaturated parameters for the entire model domain
- 22 - Surface flow resistance and hydraulic effects of the river processes due to variable
23 discharge were neglected
- 24 - Viscosity effects (temperature and solute concentration differences between the river
25 and the aquifer) were neglected.

26 **3.3 Numerical model**

27 The numerical modelling code FEFLOW 6.2 (Diersch, 2014) was used to simulate variably
28 saturated flow during river-aquifer interaction by solving Richards' equation using the
29 PARDISO solver (Schenk and Gärtner, 2004). The model domain was discretised using the
30 triangle mesh generator. The time step was set to 30-min intervals to keep the computational
31 time within reasonable limits and was increased to one hour during the validation run. To
32 enable inverse parameter estimation for the van Genuchten model, a plug-in for coupling
33 FePEST (graphical user interface for PEST by Doherty et al., 2011) with FEFLOW was
34 developed (Ehlert, 2014).

35 **3.4 Model calibration**

36 The model was sequentially calibrated against measurements in L5 and L25 collected during
37 June-October 2012. First, only hydraulic conductivity, specific storage and porosity were
38 calibrated, followed by unsaturated van Genuchten parameters and maximum and residual
39 saturation. Finally, all parameters were calibrated together. To track improvement of the fit
40 between modelled and observed hydraulic head, the regression coefficient (R^2), Nash-

1 Sutcliffe index (NS) and root mean square error (RMSE) were calculated for each calibration
2 run.

3 The conceptual understanding of hydrogeological processes is often erroneous in terms of
4 boundary and initial conditions (Bredehoeft, 2005). While the initial state of models is often
5 calibrated, the importance of other conceptualisation aspects seems to be rarely verified, even
6 though their effects have been debated by different authors (e.g. Refsgaard et al., 2006). To
7 test our assumptions on hydrogeological conditions in the area, a sensitivity analysis for
8 model boundary conditions was carried out. The distance from the river to the aquifer
9 boundary was varied from the reference distance (200 m) to 50, 500, 1000 and 5000 m,
10 keeping the recharge at 50% of annual precipitation. Afterwards, the recharge rate was
11 changed from the assumed 50% to 30% and 70%, keeping the distance from the river to the
12 aquifer boundary at 200 m.

13 3.5 Modelling scenarios

14 Based on the calibrated model domain, the effect of multiple hydrogeological parameters on
15 hyporheic exchange was evaluated, varying one parameter at a time. Artificial river stage
16 variations were applied according to the distribution of commonly observed amplitudes and
17 durations during 2012 (Fig. 3). The head boundary on the river side was varied as a cosine-
18 shaped wave between $t = 0$ and $t = t'$, with amplitude $h_{max} - h_0$ (McCallum et al., 2010):

$$h(t) = h_0 + \frac{(h_{max} - h_0)}{2} \left(1 - \cos \left(2\pi \frac{t}{t'} \right) \right) \quad (1)$$

19 where h is the hydraulic head (m), t is time (h), t' is the duration of the stage oscillation (h), h_0
20 is the head at $t = 0$, and h_{max} is the maximum head (at $t = t'/2$). All scenarios used a single
21 wave event and were terminated after steady-state conditions were reached.

22 The sensitivity of the model to various scenarios was evaluated using the flux across the river-
23 aquifer interface, bank storage and the ratio between the time to fill and to empty the bank
24 storage. A reference simulation was based on the calibrated model (hydraulic conductivity of
25 the aquifer 2.14 m d^{-1} and of the clogging layer 0.01 m d^{-1}), the actual conditions at the site
26 (bank slope 10°) and an input wave with 3 h duration and 0.4 m amplitude. Other scenarios
27 included varying hydraulic conductivity of the clogging layer (0.001, 0.01, 0.1, 1 m d^{-1}), river
28 bank slope (5, 10, 15, 30, 45°), wave amplitude (0.03, 0.1, 0.4, 0.7, 1.0 m) and duration (3, 6,
29 12, 24, 168 h) (Fig. 4). The initial conditions were generated by running a transient simulation
30 with constant hydraulic head at the riverside, no flow at the aquifer side and constant
31 distributed diffuse recharge rate at the top for the time sufficient to recreate steady-state
32 conditions. The flux into the aquifer was considered to be positive. The exchange flux per
33 metre riverbed width ($\text{m}^2 \text{ d}^{-1}$) was identified as the difference between inflow and outflow
34 rates across the river boundary. The bank storage (m^2) was the cumulative exchange flux
35 multiplied by the time step assuming the flux being either positive or negative for in- and
36 outflow from the model, respectively, while bank storage always positive. The fill time was
37 the time required for the bank storage to reach its maximum, whereas the return time was the
38 time between the maximum and zero bank storage on the falling limb. Residence time was
39 calculated as the sum of the fill and return times.

1 **4 Results**

2 **4.1 Model calibration**

3 Sequential calibration using FePEST yielded minor under-estimation of hydraulic head in the
4 beginning of the simulation and minor over-estimation in later parts (Fig. 5). Generally, the fit
5 was better for the well closer to the river, i.e. L5. The resulting R^2 , NS, RMSE were 0.95,
6 0.89, 0.06 m, respectively, for observation well L5 and 0.89, 0.77, 0.08 m, respectively, for
7 well L25. The calibrated model (Table 1) was validated using observation data from 2010
8 with R^2 , NS, RMSE of 0.85, 0.97, 0.09 m, respectively, for well L5 and 0.74, -17.85, 0.43 m,
9 respectively, for well L25 (data not shown).

10 Numerical simulation results verified that an aquifer boundary located 200 m away from the
11 river was out of reach of influences induced by the river stage fluctuations used here.
12 Sensitivity analysis of the boundary conditions resulted in substantial over-estimation of
13 measured hydraulic head for a model domain size of 5000 m. The fit improved slightly for a
14 domain size of 500 m and 1000 m compared with the 200 m long domain. However, the
15 larger model domains were discarded to keep the computation time reasonable. A model
16 domain of 50 m tended to under-estimate the measured hydraulic head for both L5 and L25.
17 The highest and lowest recharge rates over- and under-estimated the observed data,
18 respectively. Therefore, the recharge rate taken as 50% of precipitation was the most suitable
19 solution for the present case. Overall, observation well L25 was more affected by changes in
20 the recharge than well L5, indicating a strong influence of groundwater gradient on L25 and
21 of the river boundary on L5. However, it was recognised here that the final influence of
22 recharge and distance to the boundary on calibration results is a combined effect rather than a
23 single effect of one of these.

24 **4.2 Scenarios**

25 The simulated scenarios of varying river bank slope, clogging layer hydraulic conductivity
26 and input wave amplitude and duration were compared based on their effect on the resulting
27 exchange fluxes, bank storage and residence time.

28 **4.2.1 Exchange fluxes**

29 There were variations in exchange fluxes across the river-aquifer interface as a result of
30 varying forcing parameters (Fig. 6). The maximum exchange flux decreased with river bank
31 slope (Fig. 6a). A change in the bank slope from 5° to 10° caused a similar decrease in the
32 exchange flux as a change from 10° to 45° . The exchange flux increased with hydraulic
33 conductivity of the interface (Fig. 6b). However, for the scenario with $K = 0.001 \text{ m d}^{-1}$, the
34 fluxes were always directed towards the river, indicating no bank storage and thus no
35 hyporheic exchange. The increase in exchange flux was ~~not-logarithmically~~ proportional to
36 the change in hydraulic conductivity, e.g. a 100-fold increase in hydraulic conductivity
37 generated only a seven-fold rise in exchange flux. The wave amplitude was related positively
38 to the exchange flux across the river-aquifer interface (Fig. 6c). There were no positive
39 (towards the aquifer) fluxes for amplitude 0.1 m and lower. Every further 0.3 m increase in

1 amplitude resulted in a $0.4 \text{ m}^2 \text{ d}^{-1}$ increase in the maximum simulated fluxes. The maximum
2 exchange flux was higher for shorter wave duration times, e.g. $0.40 \text{ m}^2 \text{ d}^{-1}$ for a 3-h wave and
3 $0.09 \text{ m}^2 \text{ d}^{-1}$ for a 168-h wave (Fig. 6d).

4 **4.2.2 Bank storage**

5 The effects of bank slope, hydraulic conductivity of the clogging layer and input wave
6 amplitude and duration on bank storage volume were plotted as a function of time (Fig. 7).
7 Bank storage increased with lower bank slope (Fig. 7a). For example, an almost five-fold
8 increase in bank slope from approx. 10° to 45° reduced bank storage by less than 50%.
9 Meanwhile, a decrease in bank slope from 10° to 5° doubled bank storage (Fig. 7b), indicating
10 the importance of small slope for river-aquifer exchange. Overall, the bank storage for 5° was
11 five-fold higher than that for 45° . A 10-fold increase in hydraulic conductivity of the riverbed
12 from the reference scenario (bank storage = 0.02 m^2) improved bank storage by 500% (0.10
13 m^2), which further increased to 0.16 m^2 with another 10-fold increase in hydraulic
14 conductivity. As was the case for exchange flux, virtually no bank storage occurred for the
15 scenarios with 0.03 m and 0.1 m wave amplitude (Fig. 7c), which were the most common
16 amplitudes at the observation site (Fig. 3). An approximately 60% rise in wave amplitude
17 (from 0.4 m to 0.7 m) resulted in a 30% increase in maximum bank storage and a 50%
18 increase in maximum exchange flux. Duration of the river stage oscillation also positively
19 affected bank storage (Fig. 7d). A 56-fold rise in duration (from 3 h to 168 h) resulted in a
20 seven-fold increase in maximum bank storage (from 0.02 m^2 to 0.14 m^2).

21 **4.2.3 Residence time**

22 The timing of bank storage (residence time and return/fill ratio) was examined under different
23 modelling scenarios (Fig. 8). The residence time and return/fill time ratio decreased with
24 increasing bank slope (Fig. 8a). The return time always exceeded the fill time except for the
25 slopes above 30° , which indicated $t_R/t_F = 1$. Increased hydraulic conductivity of the river-
26 aquifer interface increased the return time, which positively affected the overall residence
27 time of river water in the subsurface (Fig. 8b). Nonetheless, the residence time was highest
28 (6.5 h) for the scenario with hydraulic conductivity of the interface of 0.1 m d^{-1} , marginally
29 exceeding the scenario with the highest hydraulic conductivity (6 h). Return time increased
30 with rising wave amplitude, as did the residence time, ranging from 0 h for the smallest wave
31 to 9.6 h for the largest (Fig. 8c). The ratio between the change in residence time and the
32 change in amplitude varied between 0.3 and 0.5 and was higher at lower amplitudes. The
33 return time of bank storage was longer than the fill time for the waves with duration below 24
34 h and decreased with wave duration (Fig. 8d). The return/fill time ratio decreased by almost
35 two-thirds from the shortest wave duration (1.7) to the longest (0.6), whereas the residence
36 time increased by more than one order of magnitude from the shortest wave to the longest
37 (from 4.8 to 96 h).

38 **5 Discussion**

39 **5.1 Conceptual and numerical models and calibration**

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1 The model calibration resulted in hydraulic conductivity one order of magnitude higher than
2 estimated via field tests. This difference is attributable to the limitations of slug tests, which
3 provide point data for saturated hydraulic conductivity around the well filter. According to the
4 observed soil profile (Fig. 1), grain size decreased with depth and can be a reason for lower
5 measured and higher calibrated hydraulic conductivity of the aquifer. It was beyond the scope
6 of this work to analyse whether the calibrated parameter set converged around a local or
7 global optimum. However, there was a tendency for over-estimation of both saturated
8 hydraulic conductivity and effective porosity (Table 1). Hydraulic conductivity and porosity
9 are related through the hydraulic diffusivity term that controls the connectivity of a high
10 permeability flowpath (Knudby and Carrera, 2006). Hydraulic diffusivity remains virtually
11 the same with proportional change in both hydraulic conductivity and porosity. Therefore,
12 even if over-estimation of both parameters by model calibration took place, this would have
13 had a limited effect on the results.

14 Although the validation results showed a good fit for hydraulic head at observation well L5,
15 major under-prediction of hydraulic head at L25 was observed. This could be related to the
16 substantially different precipitation patterns during the two years (2012 was used for
17 calibration and 2010 for validation) and the fact that L5 is more influenced by the river and
18 L25 by the aquifer. Recharge on top of the model, used for both calibration and validation
19 runs, was assumed to be constant and equal to 50% of mean annual precipitation. In the
20 following sections, the implications of the modelling results for hyporheic exchange and the
21 limitations due to the assumptions used are discussed.

22 **5.42 Implications for hyporheic exchange**

23 **5.2.1 General reflections**

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24 The implications of sloping banks for numerical modelling have been discussed previously,
25 e.g. by Doble et al., (2012a). In terms of functioning of the river-aquifer interface, ~~more~~
26 ~~steeply sloping banks~~ low bank slope angles increase the contact area between river and
27 aquifer and therefore result in enlarged volume of bank storage and size of the hyporheic
28 zone. ~~More steeply sloping banks~~ This also positively affects the residence time, which is
29 primarily governed by the penetration distance of river water and the return time necessary for
30 it to discharge back into the river. Enhanced hyporheic exchange due to river bank slope
31 would be true for many regulated rivers, as construction of reservoirs is associated with river
32 floodplain inundation, and therefore with the formation of gently sloping banks along the
33 channel. Consequently, ~~less lower sloping banks~~ bank slope angles have the potential to
34 improve e.g. spawning conditions and species richness (Hanrahan, 2008). However, this
35 appears not to be the case in several regulated temperate and boreal rivers due to the effect of
36 colmation (Brunke and Gonser, 1997; Blaschke et al., 2003; Calles et al., 2007; Siergieiev et
37 al., 2014a). In the present study, bank storage decreased with decreasing hydraulic
38 conductivity of the river-aquifer interface (Fig. 7b). Therefore, in rivers with clear interstices,
39 flat river banks contribute greatly to an increase in bank storage and hyporheic exchange, as
40 demonstrated by the simulations, whereas this effect is hampered in rivers with a clogged
41 riverbed.

1 River wave duration and amplitude were positively related to bank storage (Fig. 7c; d), but
2 only amplitude positively affected exchange flux (Fig. 6c). As opposed to bank storage,
3 exchange flux is more dependent on soil properties than on input wave configuration. This is
4 supported by the fact that the peak exchange flux decreased with the prolonged wave duration
5 (Fig. 6d), due to smaller hydraulic gradients at the river-aquifer interface, whereas the
6 maximum bank storage increased (Fig. 7d).

7 A linear relationship between maximum bank storage and the product of wave amplitude and
8 period has been reported previously by Todd (1955). However, this was only valid for a fully
9 saturated homogeneous aquifer adjacent to a fully penetrating river. Using the results of the
10 modelling scenarios in the present study, it was possible to show that there is a relationship
11 between the ratio of wave duration/amplitude and bank storage or residence time for waves
12 with amplitude exceeding 0.1 m (Fig. 9). This indicates that there is an optimal wave
13 configuration (duration and amplitude) for every specific set of hydrogeological conditions
14 that accounts for the highest bank storage and can potentially improve hyporheic exchange
15 and minimise energy losses in hydropower regulated rivers.

16 The hysteresis patterns observed for the t_R/t_F ratio for different modelling scenarios illustrate
17 that the process of filling the pores of an aquifer is different from that of draining them and
18 depends on hydraulic gradient and river-aquifer contact area (Fig. 8). The former is a function
19 of the river wave configuration, while the latter depends on the river bank slope. The
20 contribution of bank storage to river runoff is complex and of high importance in catchment
21 hydrology (Harr, 1977; Turton et al., 1992; McGlynn et al., 2004). The results presented here
22 suggest that with decreasing bank slope, the contribution of bank storage to the river extends
23 in time, prolonging the falling limb of the river hydrograph. The same effect occurs with
24 rising amplitude, which generates steeper hydraulic gradients across the river-aquifer
25 interface. However, it requires less time to return the bank storage to the river with prolonged
26 wave duration. A wave duration exceeding 24 h indicates faster return than the time required
27 to fill the soil moisture deficit. These modelling results were obtained for a one-time wave
28 event and no repeated wetting process was simulated. However, it is known that the hysteresis
29 pattern can change direction over time (McGuire and McDonnell, 2010), which implies that
30 the patterns observed here may differ for initially wet soil.

31 **5.2.2 Site specific implications**

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32 Hyporheic exchange at the observation site was mainly characterised by hindered water flow
33 across the river-aquifer interface and had residence time sufficient to establish suboxic
34 conditions in the subsurface due to a number of reasons. The observed hydrograph for the
35 Lule River was mainly dominated by short-term regulation with daily discharge peaks during
36 July-early August (first 40 days of the simulation) and by long-term regulation with extended
37 discharge waves during late August-October. A geochemical investigation of the hyporheic
38 zone at the site revealed a basically suboxic environment, with elevated dissolved
39 concentrations of Fe, Mn, NH_4 and organic carbon (Siergieiev et al., 2014c). These conditions
40 suggest that the area along the banks and below the bed at the site experiences deficiencies in
41 river water intrusion, which are primarily caused by the river discharge and the clogging

1 layer. Using nitrogen as an example, the hyporheic zone is a nitrate source at low residence
2 time and a nitrate sink at high residence time (Zarnetske et al., 2011). Consequently,
3 biogeochemical activity in the hyporheic zone is controlled by exchange fluxes and bank
4 storage (Gu et al., 2012). The combination of a rapidly rising discharge limb with long
5 duration time favours intensive intrusion of river water into the subsurface, transfer of oxygen
6 and dissolved organic carbon, and therefore promotes nitrification. Note that wave amplitude
7 had a higher influence on the maximum flux across the river-aquifer interface, whereas wave
8 duration affected total bank storage, i.e. the subsurface volume available for hyporheic
9 exchange. This is explained by a steep hydraulic gradient across the river-aquifer interface
10 and thus increased exchange flows due to a rapid rise in river discharge. To provide stable
11 conditions for these ecologically important flows, an extended discharge wave with a sharp
12 rising limb is required. At the Lule River study site, bank hyporheic exchange was triggered
13 by 40% of the wave events during 2012. These conditions had a potential to reset pore water
14 geochemistry but and only part of it might satisfy sufficient residence time for reactions to
15 occur and thus guarantee an effective biogeochemical exchange. The validity of this
16 relationship requires further testing by e.g. a sediment transport survey, among other
17 techniques, which can form the basis for implementation of environmental flows in
18 restoration programmes (Schneider et al., 2011).

19 It is not only the hyporheic zone intimately connected to the river that can be affected by
20 fluctuating river water stages, but also the distant groundwater. The simulation results
21 indicated that groundwater head was affected by pressure propagation beyond observation
22 well L25 (25 m distance to the river). This can have an impact on oxidation-reduction
23 conditions in the aquifer due to changes in the redox potential during wetting and drying
24 cycles of the soil (Reddy and Patrick, 1975; Cavanaugh et al., 2006). The relationship
25 between the depth to the groundwater and groundwater composition in observation well L25,
26 sampled during the period May-October 2011, was investigated. Based on nine water quality
27 samples and principal component analysis, the first two significant components explained
28 77% of the data variance, indicating a positive correlation between depth to groundwater and
29 NO₃ concentration and a negative correlation between depth to groundwater and Fe and Al
30 concentration. A positive correlation between Mn and alkalinity and depth to groundwater
31 explained only 13% of the data variance and P showed no relationship. For the significant
32 correlations, a rising groundwater level promoted a more reduced environment and was
33 associated with higher Fe and lower NO₃ concentrations. This suggests that transient changes
34 in river water stages in response to hydropower management can force time-dependent
35 alterations in groundwater quality, with further potential impacts on riparian soils.

36 **5.2.3 Limitations**

37 The assumptions made in this modelling study resulted in the following limitations:

- 38 • Because of the vertical 2-D conceptualisation perpendicular to the river, longitudinal fluxes
39 parallel to the river were neglected. A 3-D model is required for proper consideration of
40 these processes. Bates et al. (2000) argued that the contribution of the longitudinal
41 component is most important at the beginning and end of an event, implying confidence

1 about timing but not about the absolute value of computed peak bank storage and fluxes
2 using a 2-D approach.

- 3 • In aquifers with a clogging layer, hydraulic pressure propagation will always be ahead of
4 water flow that follows oscillations at the river-aquifer interface (Welch et al., 2014).
5 Assuming homogeneous subsurface media, solute travel time may exceed that of the
6 pressure, resulting in over-estimated bank storage. In addition, it was assumed that all
7 return flow came from bank storage, even though it contains a mixture of old water from
8 the unsaturated zone and groundwater (Burt and Pinay, 2005). This is crucial for chemical
9 fluxes through the hyporheic zone (McDonnell, 1990) and for chemical hydrograph
10 separation (McCallum et al., 2010). Because the response of solute fluxes to bank storage
11 is dependent on heterogeneity, verification of the fluxes obtained by pressure propagation
12 using measurements of solute concentrations, e.g. electrical conductivity (Welch et al.,
13 2014), or measurements of temperature (Anibas et al., 2012) may be required.
- 14 • Lateral variability in riverbed hydraulic conductivity at the site was simplified by
15 implementing a continuous low hydraulic conductivity layer. In field settings, however, a
16 riverbed with variable sediment composition is much more likely (Hancock and Boulton,
17 2005; Siergieiev et al., 2014a), which suggests that hyporheic exchange seeks more
18 conductive patches. This assumption is likely to result in under-estimated hyporheic
19 exchange (Kalbus et al., 2009) and partially compensate for using homogeneous media
20 (see above).
- 21 • Hydraulic conductivity of the riverbed affected the initial distribution of hydraulic
22 gradients. However, the difference in the initial conditions was negligible (4% between the
23 extreme scenarios) compared with the differences in bank storage caused by the presence
24 of a clogging layer with variable hydraulic conductivity. Therefore, possible effects on
25 bank storage can be ignored.
- 26 • In order to avoid over-parameterisation of the model and to limit the effect of sparse
27 information availability regarding regional groundwater gradients, precipitation and
28 evapotranspiration were approximated using a constant recharge flux term.
- 29 • The hydraulic effects of the river flow were not included in the model.
- 30 • Viscosity effects on hydraulic conductivity (Ma and Zheng, 2010) were excluded due to
31 the small temperature difference between river and groundwater (max. 10°C) at the site.
32 Ehlert (2014) has shown that a 24% increase in hydraulic conductivity is possible due to
33 this temperature difference. However, field measurements indicated solely conductive heat
34 transport (Siergieiev et al., 2014a), due to attenuation of the advective-dispersive heat
35 transfer by the clogging layer.

36 **6 Conclusions**

37 Bank hyporheic exchange was simulated using a field case scenario in an alluvial aquifer
38 adjacent to the hydropower-regulated Lule River. The modeling showed that ecosystem
39 requirements in terms of river-aquifer exchange flux are satisfied during 40% of all wave
40 events during the studied year. Discharge waves with longer duration and increased amplitude
41 are essential for this site to improve hydrological exchange across the river-aquifer interface
42 and bank hyporheic water quality. The combination of realistic and theoretical models

1 improved current process understanding of hyporheic exchange in free-flowing and regulated
2 rivers. Hypothetical scenarios included variable river discharge wave (duration and
3 amplitude), river bank slope and hydraulic conductivity of the river-aquifer interface. ~~The~~
4 ~~combination of realistic and theoretical models improved current process understanding of~~
5 ~~hyporheic exchange in free-flowing and regulated rivers.~~ ~~From theoretical scenarios, b~~Bank
6 storage increased with lower bank slope, indicating the necessity of correct data on geometry
7 of the river-aquifer interface when modelling surface water-groundwater interactions.
8 Hydraulic conductivity of the riverbed positively affected bank storage. However, the
9 influence on the residence time was not always consistent. Higher amplitude and longer wave
10 duration increased bank storage, although larger maximum fluxes were observed for shorter
11 waves at a given amplitude. There will be always a unique relationship between bank storage
12 or residence time and the duration/amplitude wave ratio, which depends on the
13 hydrogeological conditions. Hence, hyporheic exchange suppressed by colmation processes or
14 flow manipulation can be improved by periodically releasing river discharge waves that are
15 optimised for the specific river reach.

16 **Author contribution**

17 Dmytro Siergieiev and Ludwig Ehlert collected the data, developed the model and performed
18 the simulations. Thomas Reimann contributed to model development. Dmytro Siergieiev
19 prepared the manuscript with contributions from all co-authors.

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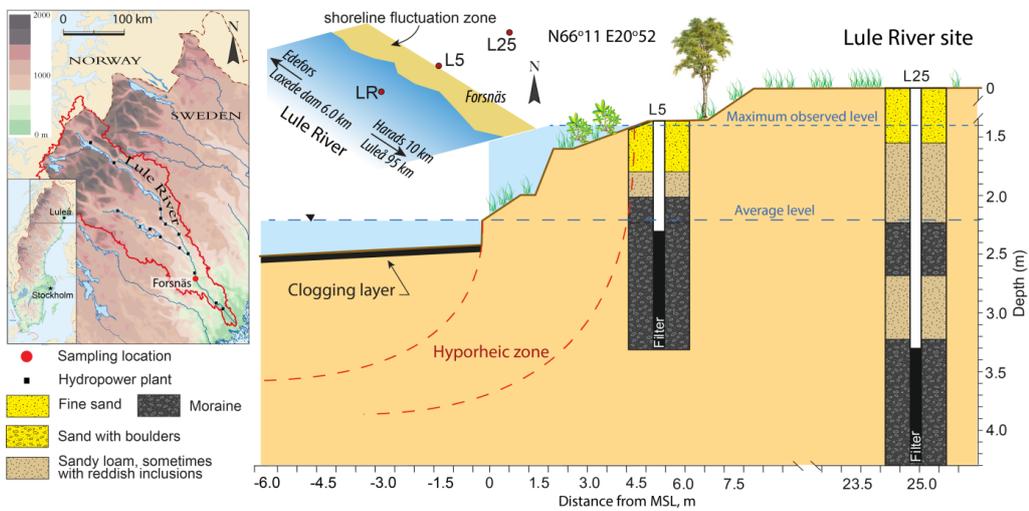
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1 Table 1. Model parameters measured and calibrated using FePEST

Parameter	Symbol	Units	Measured ^a	Calibrated
Saturated K aquifer	K_{aq}	m d ⁻¹	0.13±0.08	2.14
Saturated K clogging layer	K_{cl}	m d ⁻¹	0.04±0.01	0.01
Specific storage	S	m ⁻¹	-	0.001
Effective porosity	n	-	-	0.56
Maximum saturation	θ_s	-	0.92±0.04	0.95
Residual saturation	θ_r	-	0.14±0.08	0.14
Anisotropy ratio	K_v/K_h	-	-	1
van Genuchten parameters	α	m ⁻¹	0.003±0.001	0.015
	n	-	2.1±0.4	2.1

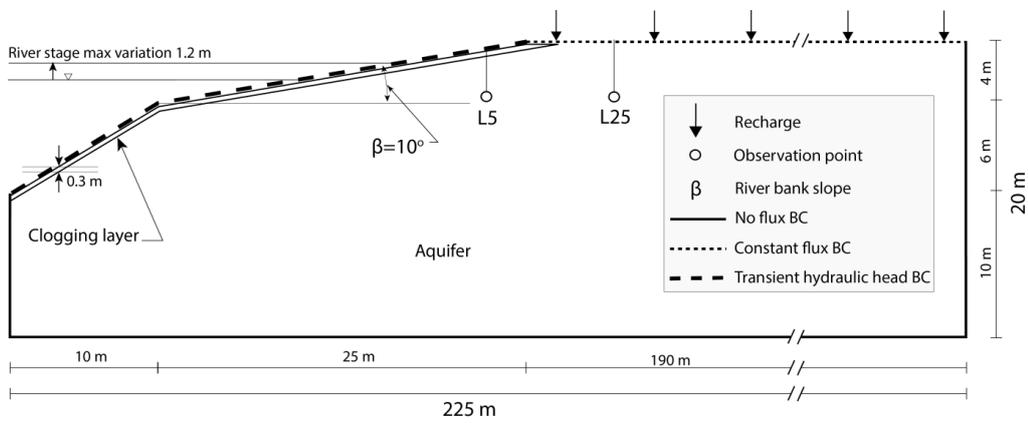
2 ^aMean of all measurements ± standard deviation

3



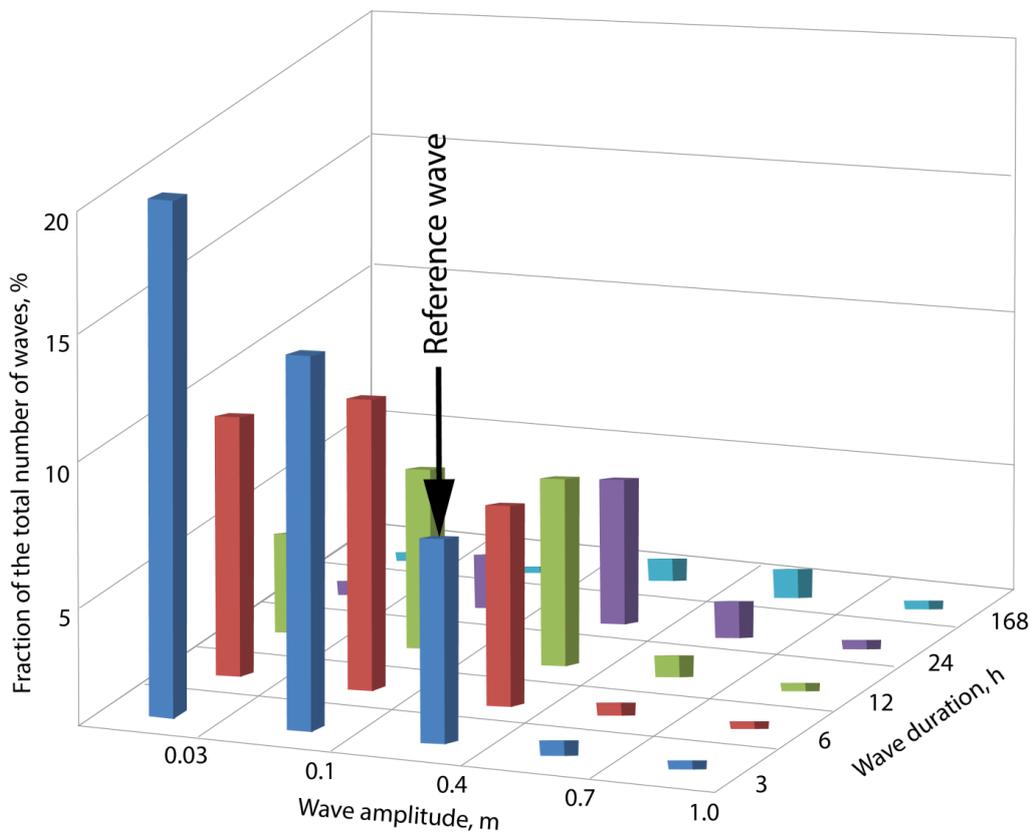
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 2 Figure 1. Location of the observation site and cross-section of the aquifer with groundwater
 3 wells, soil depth profiles and extent of the hyporheic zone. Numbers in well names indicate
 4 distance to the mean shoreline in metres.

5

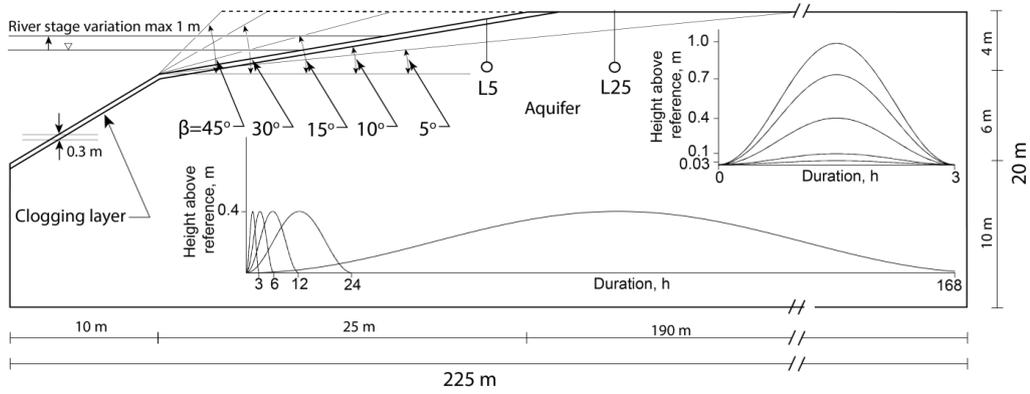


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 2 Figure 2. Conceptual model of the site showing boundary conditions (BC), clogging layer and
 3 observation points.

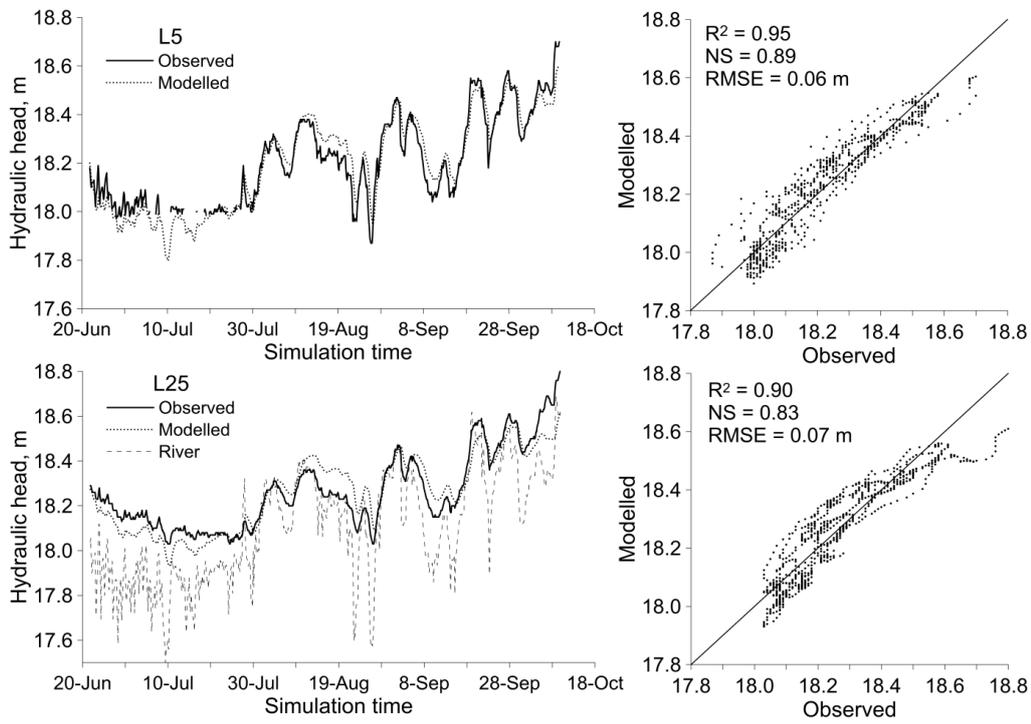
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1
 2 Figure 3. Summary matrix of wave amplitude (m) and duration (h) as a fraction of all
 3 observed wave events.
 4

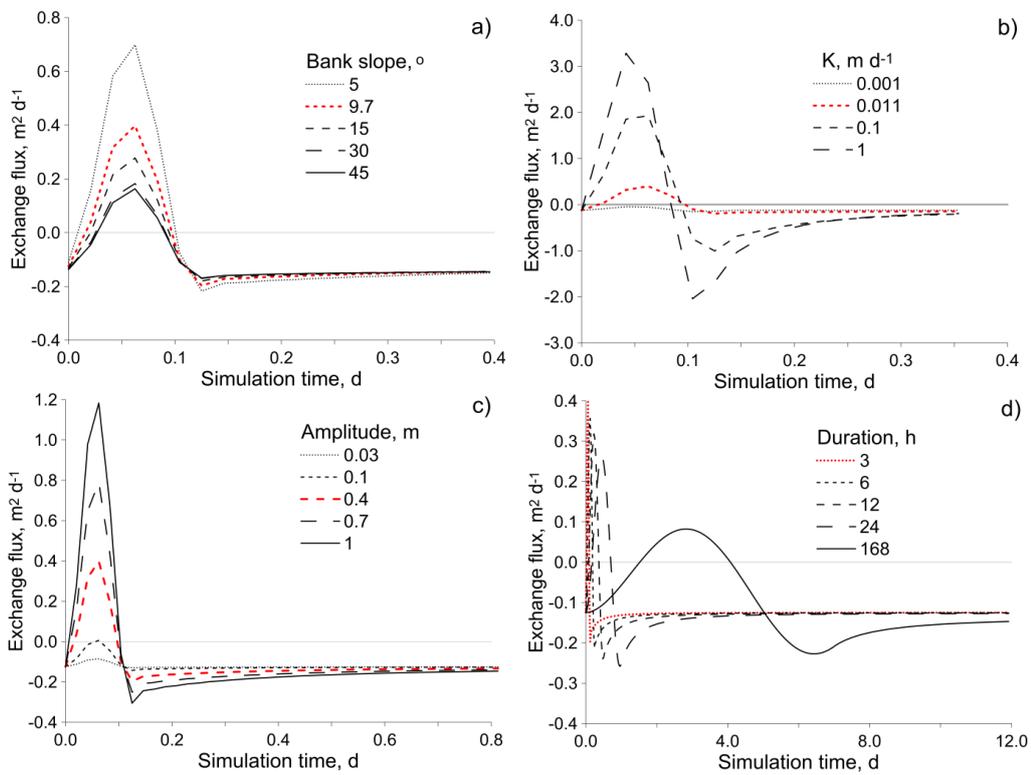


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 2 Figure 4. Graphical summary of modelling scenarios for varying river bank slope, wave
 3 duration and amplitude. For hydraulic conductivity scenarios, see text.
 4

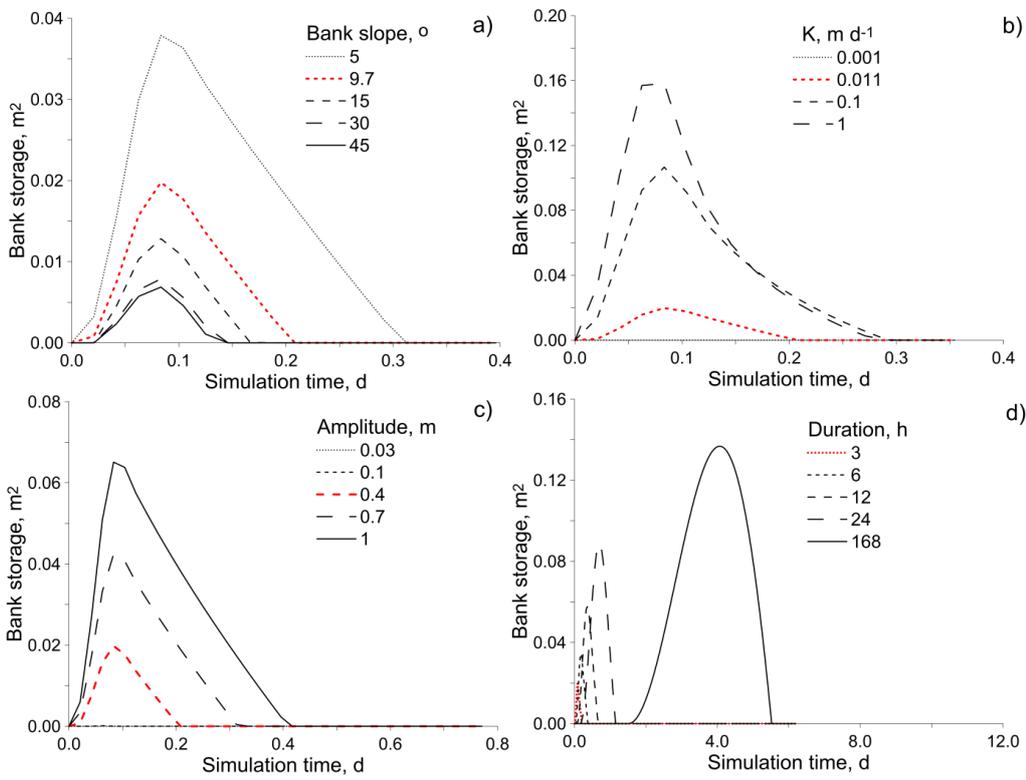


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 2 Figure 5. Hydraulic head at observation wells L5 (above) and L25 and the river (below)
 3 compared with the simulated results using calibrated parameters (see Table 1).

4

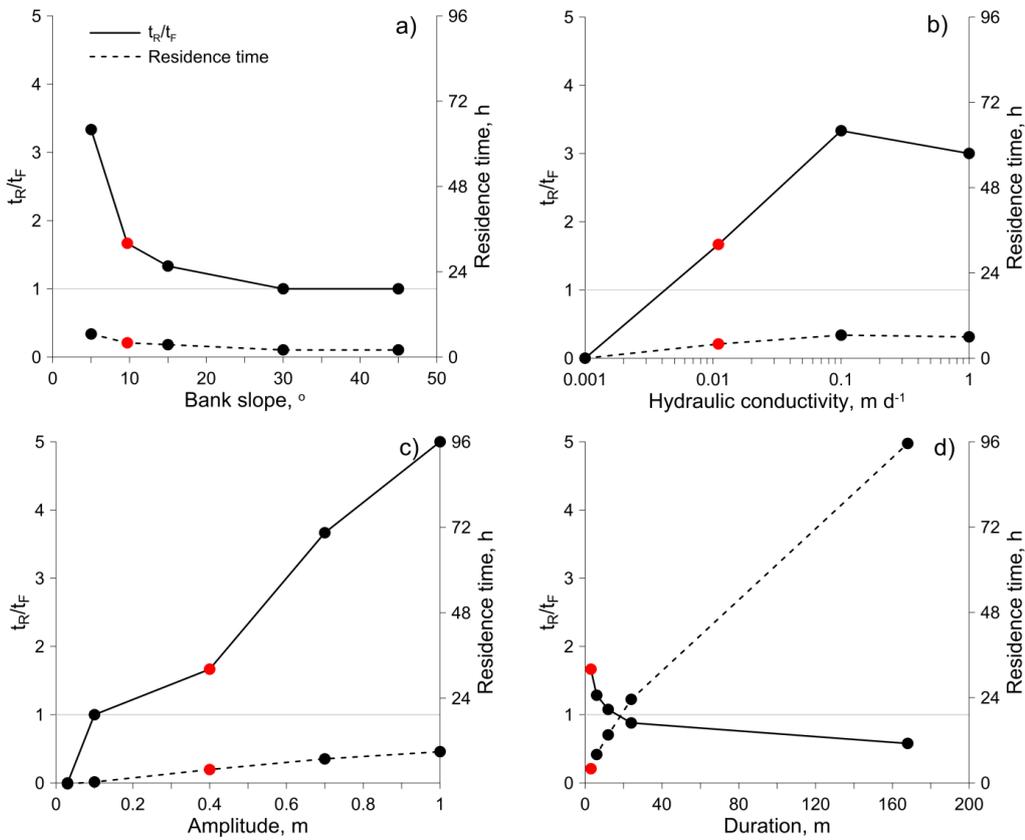


1
 2 Figure 6. Exchange flux for different bank slope (a), hydraulic conductivity (b), wave
 3 amplitude (c) and duration (d) scenarios, with the reference case (red) and insets of the input
 4 wave pulse. Note different scale.



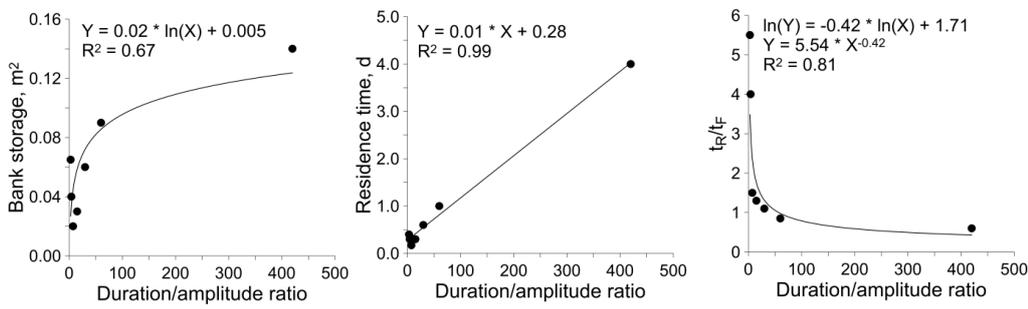
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 2 Figure 7. Bank storage for different bank slope (a), hydraulic conductivity (b), wave
 3 amplitude (c) and duration (d) scenarios, with the reference case (red) and insets of the input
 4 wave pulse. Note different scale.

5



1
 2 Figure 8. Return/fill time ratio (t_R/t_F) and residence time for different bank slope (a), hydraulic
 3 conductivity (b), wave amplitude (c) and duration (d) scenarios, with the reference case (red)
 4 and insets of the input wave pulse.

5



1
 2 Figure 9. Duration/amplitude ratio in relation to bank storage, residence time and return/fill
 3 time ratio (t_R/t_F) for all waves exceeding 0.1 m amplitude.

4